# WORKING GROUP ON WIDELY DISTRIBUTED STOCKS (WGWIDE) 

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# WORKING GROUP ON WIDELY DISTRIBUTED STOCKS (WGWIDE) 

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## Editors

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#### Abstract

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## i Executive summary

WGWIDE reports on the status and considerations for management of the Northeast Atlantic mackerel, blue whiting, Western and North Sea horse mackerel, Northeast Atlantic boarfish, Norwegian spring-spawning herring, striped red mullet (Subareas 6, 8 and Divisions 7.a-c, e-k and 9.a), and red gurnard (Subareas 3, 4, 5, 6, 7, and 8) stocks.

2023 catch advice was drafted for mackerel, Western horse mackerel, blue whiting and herring. For the remainder, multi-annual catch advice was previously published. Benchmark assessments are proposed for 2024 for mackerel, both horse mackerel stocks, herring and striped red mullet with boarfish is scheduled to be benchmarked in 2023. Prior to any benchmark assessment for mackerel, WGWIDE recommends that a workshop to review the latest knowledge with regard to the stock component structure takes place.

Northeast Atlantic Mackerel. This migratory stock is widely distributed throughout the Northeast Atlantic with significant fisheries in several ICES subareas. The assessment conducted in 2022 is an update assessment, based on the configuration agreed during the 2019 interbenchmark and incorporates updates to the commercial catch, tagging, swept area and egg survey (preliminary) data series. No recruitment index is available for the 2021 year-class as survey coverage was inadequate. Advice is given based on stock reference points which were updated during a management strategy evaluation carried out in 2020. Following a decline since 2014, SSB has been stable (above MSY Btrigger) since 2019. Fishing mortality has been increasing since 2016 and is above FMSY since 2020.

Blue Whiting. This pelagic gadoid is widely distributed in the eastern part of the North Atlantic. The current assessment configuration (inter-benchmark in 2016) uses preliminary catch and sampling data along with the acoustic survey data from the current year. The 2022 update assessment indicates that SSB is increasing following strong recent recruitment and is well above MSY Btrigger. Fishing mortality has been above FMSY since 2014 but is falling.

Norwegian Spring Spawning Herring. This stock is migratory, spawning along the Norwegian coast and feeding throughout much of the Norwegian Sea. The 2022 update assessment is based on an implementation of the XSAM assessment model introduced following a benchmark in 2016 and is consistent with the 2021 assessment. Following a period of decline since 2009, SSB has been relatively stable just above MSY Btrigger in recent years, due to the strong 2016 year-class. However, recruitment since 2016 is estimated to be below average and the stock size is forecast to fall below MSY Btrigger in 2024.
Western Horse Mackerel. The western stock of horse mackerel is distributed throughout ICES subareas $4,6,7,8$ and 9 . Following a benchmark in 2017, the stock is assessed using the Stock Synthesis integrated assessment model. Stock reference points were revised in 2019. Following a period of declining SSB, there has been a modest rise since 2017, albeit from a low level. The 2022 assessment indicates that SSB is below Blim and will remain so in 2024, even under the scenario of zero catch in 2023. Based on the MSY approach, advice for 2024 is therefore for zero catch. The assessment continues to display significant retrospective issues which should be investigated by a benchmark assessment.

North Sea Horse Mackerel. Catch advice for this stock is issued biennially on the basis of an assessment based on a combined index from groundfish surveys in the North Sea and the Channel. No survey index was available in 2020 due to restricted survey coverage, and the 2021 value is a reduction on the 2019 value for the exploitable stock. A length based indicator continues to indicate that fishing mortality remains above Fmsy.

Northeast Atlantic Boarfish. Boarfish is a small, pelagic, planktivorous, shoaling species, found over much of the Northeast Atlantic shelf but primarily in ICES subareas 4,6,7 and 8. The directed fishery occurs primarily in the Celtic Sea and developed during the early 2000s, initially unregulated before the introduction of a TAC in 2011. The stock is assessed using an exploratory Bayesian surplus production model with catch and survey data from groundfish surveys and an acoustic survey. The current assessment indicates that, following a sharp decline after 2012, biomass has been increasing in recent years. The most recent acoustic surveys indicate a period of above average recruitment from 2018-2020.

Northeast-Atlantic Red Gurnard. This stock was first considered by WGWIDE in 2016 with advice issued biennially. The assessment was benchmarked in 2021 and a survey-based relative biomass indicator was developed. The 2022 update assessment continues to show the indicator fluctuating without trend since 2010. However, large uncertainties remain with regard to landings data due to poor resolution at the species level and reported discarding levels vary widely.

Striped Red Mullet in Bay of Biscay, Southern Celtic Seas, Atlantic Iberian Waters. No assessment is available for this stock and information on abundance and exploitation level is limited with advice given triennially on the basis of the precautionary approach. However, there are a number of research projects underway which will inform a future benchmark and potential upgrade of the assessment category.

## ii Expert group information

| Expert group name | Working Group on Widely Distributed Stocks (WGWIDE) |
| :--- | :--- |
| Expert group cycle | Annual |
| Year cycle started | 2022 |
| Reporting year in cycle | $1 / 1$ |
| Chair(s) | Andrew Campbell, Ireland |
| Meeting venue(s) and dates | $14-30$ August 2022, Copenhagen, Denmark and online (40 participants) |

## 1 Introduction

### 1.1 Terms of References (ToRs)

The Working Group on Widely Distributed Stocks (WGWIDE), chaired by Andrew Campbell, Ireland, met in ICES, Copenhagen in hybrid format from 24-30 August 2022. The terms of reference for the meeting were the generic ToRs for Regional and Species Working Groups:
a) Consider and comment on Ecosystem and Fisheries overviews where available;
b) For the aim of providing input for the Fisheries Overviews, consider and comment on the following for the fisheries relevant to the working group:
i) descriptions of ecosystem impacts on fisheries
ii) descriptions of developments and recent changes to the fisheries
iii) mixed fisheries considerations, and
iv) emerging issues of relevance for management of the fisheries;
c) Conduct an assessment on the stock(s) to be addressed in 2022 using the method (assessment, forecast or trends indicators) as described in the stock annex; - complete and document an audit of the calculations and results; and produce a brief report of the work carried out regarding the stock, providing summaries of the following where relevant:
i) Input data and examination of data quality; in the event of missing or inconsistent survey or catch information refer to the ACOM document for dealing with COVID19 pandemic disruption and the linked template that formulates how deviations from the stock annex are to be reported.
ii) Where misreporting of catches is significant, provide qualitative and where possible quantitative information and describe the methods used to obtain the information;
iii) For relevant stocks (i.e., all stocks with catches in the NEAFC Regulatory Area), estimate the percentage of the total catch that has been taken in the NEAFC Regulatory Area in 2021.
iv) For category 3 and 4 stocks requiring new advice in 2022, implement the methods recommended by WKLIFE $X$ (e.g. SPiCT, rfb, chr, rb rules) to replace the former 2 over 3 advice rule ( 2 over 5 for elasmobranchs). MSY reference points or proxies for the category 3 and 4 stocks
v) Evaluate spawning stock biomass, total stock biomass, fishing mortality, catches (projected landings and discards) using the method described in the stock annex;

1) for category 1 and 2 stocks, in addition to the other relevant model diagnostics, the recommendations and decision tree formulated by WKFORBIAS (see Annex 2 of https://www.ices.dk/sites/pub/Publication\ Reports/Ex-pert\ Group\ Report/Fisheries\ Resources\ Steering\ Group/2020/WKFORBIAS_2019.pdf) should be considered as guidance to determine whether an assessment remains sufficiently robust for providing advice.
2) If the assessment is deemed no longer suitable as basis for advice, consider whether it is possible and feasible to resolve the issue through an interbenchmark. If this is not possible, consider providing advice using an appropriate Category 2 to 5 approach.;
vi) The state of the stocks against relevant reference points;

Consistent with ACOM's 2020 decision, the basis for Fpa should be Fp. 05.

1) Where Fp. 05 for the current set of reference points is reported in the relevant benchmark report, replace the value and basis of Fpa with the information relevant for Fp. 05
2) Where Fp. 05 for the current set of reference points is not reported in the relevant benchmark report, compute the Fp. 05 that is consistent with the current set of reference points and use as Fpa. A review/audit of the computations will be organized.
3) Where Fp. 05 for the current set of reference points is not reported and cannot be computed, retain the existing basis for Fpa.
vii) Catch scenarios for the year(s) beyond the terminal year of the data for the stocks for which ICES has been requested to provide advice on fishing opportunities;
viii)Historical and analytical performance of the assessment and catch options with a succinct description of associated quality issues. For the analytical performance of category 1 and 2 age-structured assessments, report the mean Mohn's rho (assessment retrospective bias analysis) values for time series of recruitment, spawning stock biomass, and fishing mortality rate. The WG report should include a plot of this retrospective analysis. The values should be calculated in accordance with the "Guidance for completing ToR viii) of the Generic ToRs for Regional and Species Working Groups - Retrospective bias in assessment" and reported using the ICES application for this purpose.
d) Produce a first draft of the advice on the stocks under considerations according to ACOM guidelines.
i. In the section 'Basis for the assessment' Table 3 under input data align the survey names with the ICES survey naming convention
e) Review progress on benchmark issues and processes of relevance to the Expert Group. i) update the benchmark issues lists for the individual stocks in SID;
ii) review progress on benchmark issues and identify potential benchmarks to be initiated in 2023 for conclusion in 2024;
iii) determine the prioritization score for benchmarks proposed for 2023-2024;
iv) as necessary, document generic issues to be addressed by the Benchmark Oversight Group (BOG)
f) Prepare the data calls for the next year's update assessment and for planned data evaluation workshops;
g) Identify research needs of relevance to the work of the Expert Group.
h) Review and update information regarding operational issues and research priorities on the Fisheries Resources Steering Group SharePoint site.
i) If not completed in 2020, complete the audit spread sheet 'Monitor and alert for changes in ecosystem/fisheries productivity' for the new assessments and data used for the stocks.

Also note in the benchmark report how productivity, species interactions, habitat and distributional changes, including those related to climate-change, could be considered in the advice.

### 1.1.1 The WG work 2022 in relation to the ToRs

The WG considered updates for all eight stocks within its remit. Based upon these assessments and associated short term forecasts, the group produced draft advice sheets for Northeast Atlantic mackerel, Blue Whiting, Norwegian spring spawning herring and Western horse mackerel. 2021-22 catch advice for Boarfish, North Sea horse mackerel and red gurnard and 2021-23 catch advice for striped red mullet were issued in 2020. All draft advice sheets were agreed in plenary. Advice sheets, report sections and assessments were audited with 2-4 working group members assigned to each stock. In addition, the stock annexes for mackerel and blue whiting were updated.

### 1.2 Participants at the meeting

WGWIDE 2022 was attended by 40 delegates ( 5 online) from the Netherlands, Ireland, Spain, Norway, Germany, Portugal, Iceland, UK (England and Scotland), Faroe Islands, France, Denmark, Greenland and Sweden. The full list of participants, all of whom are authors of this report is given in Annex 1.

All the participants were made aware of ICES Code of Conduct, which all abided by and none had Conflicts of Interest that prevented them from acting with scientific independence, integrity, and impartiality.

### 1.3 Overview of stocks within the WG

Eight stocks are assessed by WGWIDE. In 2022, the group drafted 2023 advice sheets for 4 stocks. 2022 advice for North Sea horse mackerel, boarfish, red gurnard and striped red mullet was issued in 2020 the relevant data series and stock assessments were updated and considered at WGWIDE 2022. A summary of the WGWIDE stocks, current data category and assessment method and advice frequency is given in the table below:

| Stock | ICES <br> code | Data <br> Cate- <br> gory | Assessment method | Assessment <br> Frequency | Last <br> Assess- <br> ment |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Boarfish | boc.27.6-8 | 3.2 | Bayesian Schafer surplus production model | 2 | 2021 |
| Red gurnard | gur.27.3-8 | 3.2 | Survey trends based | 2 | 2021 |
| Norwegian spring-sp. Herring | her.27.1-24a514a | 1 | XSAM | 1 | 2021 |
| Western horse mackerel | hom.27.2a4a5b6a7a-ce-k8 | 1 | Stock Synthesis | 1 | 2021 |
| North Sea horse mackerel | hom.27.3a4bc7d | 3.2 | Survey trends based | 2 | 2021 |


| Stock | ICES <br> code | Data <br> Cate- <br> gory | Assessment method | Assessment <br> Frequency | Last <br> Assess- <br> ment |
| :---: | :---: | :---: | :---: | :---: | :---: |
| NE-Atlantic mackerel | mac.27.nea | 1 | SAM | 1 | 2021 |
| Striped red mullet | mur.27.67a-ce-k89a | 5 | No assessment | 3 | 2020 |
| Blue whiting | whb.27.1-91214 | 1 | SAM | 1 | 2021 |

### 1.4 Quality and Adequacy of fishery and sampling data

### 1.4.1 Sampling Data from Commercial Fishery

Each year, the working group reviews available sampling data and the level of sampling on the commercial fisheries. Details are given in the relevant stock-specific sections of this report.

Generally, the amount and quality of available data to the WG has been unchanged in the most recent years. However, this year no Russian data submissions were available (for 2021). Russia has significant catches of NEA Mackerel, Blue Whiting and Norwegian Spring Spawning Herring and usually provides sampling data for these fisheries. Information on total catch for 2021 by ICES division is available from the ICES preliminary catch database. Historically, this matches final estimates closely and was therefore used as an estimate of Russian catch by ICES division in 2021. Catch proportion by quarter in 2021 was assumed to be equal to the recent average (20182020). Samples available from other national fisheries operating in the same area and quarter were used to estimate the age structure of Russian catch in 2021.

The WG identified issues associated with the formatting and availability of data from commercial catch sampling programmes such as the requirement for length frequency and age-length key data for the assessment of Western horse mackerel and the availability of data arising from the sampling of catches of North Sea horse mackerel from foreign flagged vessels. The issues have been included on the individual stock issue lists and the ICES data call has been updated such that future data submissions should provide data in the appropriate format.

### 1.4.2 Catch Data

The WG has on a number of occasions discussed the accuracy of the catch statistics and the possibility of large scale under reporting or species and area misreporting. The working group considers that the best estimates of catch it can produce are likely to be underestimates.

In the case of red gurnard catch data, the available information is of poor quality. Prior to 1977, red gurnard catches were not reported. Since this time, landings of gurnards have often been reported as mixed gurnards, or using the incorrect species code. With the exception of Portugal, there is no detail provided to the WG on the methodology used to estimate the proportion of red gurnards in mixed landings.

### 1.4.3 Discards

In 2015, the European Union introduced a landing obligation for fisheries directed on small pelagic fish including mackerel, horse mackerel, blue whiting and herring. The obligation was
expanded over the following years in a stepwise fashion such that discarding of small pelagic species could still legally occur in other fisheries. From 2019 onwards the landing obligation is generally effective. A general discard ban is already in place for Norwegian, Faroese and Icelandic fisheries.

Historically, discarding in pelagic fisheries is more sporadic than in demersal fisheries. This is because the nature of pelagic fishing is to pursue schooling fish, creating hauls with low diversity of species and sizes. Consequently, discard rates typically show extreme fluctuation ( $100 \%$ or zero discards). High discard rates occurred especially during 'slippage' events, when the entire catch is released. The main reasons for 'slipping' are daily or total quota limitations, illegal size and mixture with unmarketable bycatch. Quantifying such discards at a population level is extremely difficult as they vary considerably between years, seasons, species targeted and geographical region.

Discard estimates of pelagic species from pelagic and demersal fisheries have been published by several authors. Discard percentages of pelagic species from demersal fisheries were estimated between $3 \%$ to $7 \%$ (Borges et al., 2005) of the total catch in weight, while from pelagic fisheries were estimated between $1 \%$ to $17 \%$ (Pierce et al. 2002; Hofstede and Dickey-Collas 2006, DickeyCollas and van Helmond 2007, Ulleweit and Panten 2007, Borges et al. 2008, van Helmond and van Overzee 2009, 2010, van Overzee and van Helmond 2011, Ulleweit et al. 2016, van Overzee et al. 2013, 2020). Slipping estimates have been published for the Dutch freezer trawler fleet only, with values at around $10 \%$ by number (Borges et al. 2008) and around $2 \%$ in weight (van Helmond et al. 2009, 2010 and 2011) over the period 2003-2010. In Iberian waters the discard composition of pelagic species, mainly blue whiting, in demersal fisheries were estimated between $20 \%$ and $30 \%$ of the total catch in weight (Fernandes et al. 2015). Nevertheless, the majority of these estimates were associated with very large variances and composition estimates of 'slippages' are liable to strong biases and are therefore open to criticism.

Because of the potential importance of significant discarding levels on pelagic species assessments, the Working Group again recommends that observers should be placed on board vessels in those areas in which discarding occurs, and existing observer programmes should be continued. Furthermore, agreement should be made on sampling methods and raising procedures to allow comparisons and merging of dataset for assessment purposes. The newest update on discards for the different stocks assessed by the WG is provided in the sections for each of the stocks.

### 1.4.4 Age-reading

Reliable age data are an important prerequisite in the stock assessment process. The accuracy and precision of these data, for the various species, is kept under constant review by the Working Group. The most recent updates on this aspect for the different stocks are addressed below.

### 1.4.4.1 Mackerel

The most recent age calibration exercise for this stock was carried out in 2021 using the SmartDots platform under the remit of WGBIOP. The full exercise was completed by 37 readers from 12 countries across Europe. Otolith images ( $\mathrm{n}=237$ ) were provided by 12 of the participating laboratories with the aim to provide a set of images representative of the temporal and spatial coverage of otoliths read for stock assessment purposes (including the southern component, western component, North Sea component and the northern distribution).

Results show a slightly lower percentage of agreement and higher CV in the analysis taking all readers in account than the previous workshop (2018) and exchange (2014) which might be related to an increased number of readers (23 in WK 2018; 37 in Ex 2021 with 10 new (basic)
readers). However, lower agreement (and higher CV) was also found for advanced readers. Here, numbers of readers also increased from 2018 to 2021 (15 to 21).

The overall conclusion was that the slightly worse results than in the prior workshops might be related to the increased number of readers. The image quality of otoliths from different areas was also discussed. However, the problem shown in previous workshops and exchanges persists: Agreement for otoliths with modal age 6 and older remains quite low. A new workshop was recommended.

At the NEA mackerel Inter-benchmark in 2019, concerns related to the quality of age reading of commercial catch were discussed. WGWIDE concludes that additional investigation on the impact of ageing error on stock assessment outputs are required. This includes the development of standardized sensitivity analyses for this purpose, which would be applicable to the different stocks.

### 1.4.4.2 Horse mackerel

The most recent workshop on the age reading of Trachurus trachurus (also T. mediterraneus and T. picturatus) was carried out in November 2018 and involved 15 age readers from 9 countries.

The objectives of this workshop were to review the current methods of ageing Trachurus species, to evaluate the new precision of ageing data of Trachurus species and to update guidelines, common ageing criteria and reference collections of otoliths. The exchange results showed a low value of percentage of agreement from $45.1 \%$ to $59.1 \%$ for the three Trachurus species. The Coefficient of Variation was lower for T. trachurus (17.3-32.2) than for the other Trachurus species (60.1-73.4) because the sampled specimens were older for this species than for the two other species. With feedback from the readers present at the exchange and the discussion during the WKARHOM3 meeting, the main cause of age determination error for T. trachurus was identified as otolith preparation techniques (whole/slice).

However, for the three Trachurus species, there are several difficulties in age determination: identification of the first growth annulus, presence of many false rings (mainly in the first and second annuli) and the interpretation and identification of the edge characteristics (opaque/ translucent). The second reading was performed during the workshop with 50 images per each species. Each reader read only the images of the species that is read in their laboratory. The percentage of agreement between readers increased to $70.6 \%$ with a CV of 18.4 for T. trachurus and to $67.8 \%$ with a CV of 31.7 for $T$. mediterraneus. Finally, the group reached an agreement on defining an ageing guideline and a reference collection presented in this report and the aim is to employ these tools for all laboratories.

The next workshop (WKARHOM4) and exchange is planned for November 2022 using the SmartDots platform.

### 1.4.4.3 Norwegian Spring-spawning Herring

For some years, there have been issues with age reading of herring. These issues were raised around 2010, and since then two scale/otolith exchanges and a workshop have been held; and a final workshop was planned after the second exchange. There were, however, concerns with the second scale/otolith exchange and the final workshop was postponed indefinitely. It is therefore recommended to organise a new scale/otolith exchange and a follow up workshop.

There are several topics to cover in the recommended work.
Firstly, age-error matrices are needed as input to the stock-assessment, to evaluate sensitivity to ageing errors, and such age-error matrices are an output of age-reading inter-calibrations.

Secondly, stock mixing is an issue. There are several herring stocks surrounding the distribution area of Norwegian spring spawning (NSS) herring, e.g. North Sea herring, Icelandic summer spawning herring, local autumn-spawning herring in the Norwegian fjords, and Faroese autumn spawning herring. Mixing with these other stocks in the fringe areas of the NSS herring distribution area leads to confounding effects on the survey indices of NSS herring in the ecosystem surveys and potentially also in the catch data. Methods to separate the NSS herring stock from the other herring stocks are needed - both with regards to obtain more accurate age-readings as well as to reduce confounding effects on the survey indices.

Finally, the experience from earlier exchanges is that age of older fish is more prone to be underestimated when aged is read from otoliths as compared to being read from scales. Some of the institutes mainly sample and read scales, whereas other institutes use the otoliths.
Last year, WGWIDE recommended to organise a scale/otolith exchange and workshop. This work appears to be in progress in WGIPS, WGBIOP and nationally at the institutes, and a workshop is planned for April 2023.

### 1.4.4.4 Blue Whiting

The most recent workshop on the age reading of blue whiting (WKARBLUE3) took place in 2021 ( 31 May-4 June). The workshop was preceded by an inter-calibration age reading exchange, which was undertaken in 2020 using the SmartDots platform. In the exchange, the otolith collection included 407 otoliths from the entire stock distribution area, from which 190 otoliths where from the northern areas and 217 where from the southern areas of distribution. The otolith dataset enables a good coverage of samples by area and sex and took into account the differences in growth patterns by areas (northern and southern), and by sex due to the sexual dimorphism in blue whiting (Gonçalves et al. 2017).

The overall agreement of the pre-workshop exercise was $66 \%$ considering all readers and $70 \%$ for the assessment readers (advanced readers). Considering only the otoliths samples from the northern areas and the readers from the northern that usually read the otoliths from those areas for the assessment, $69 \%$ of agreement was achieved. Otherwise, considering only the otoliths samples from the southern areas and the readers from the southern that usually read the otoliths from those areas for the assessment, $79 \%$ of agreement was achieved. During the workshop, a small exchange was also conducted with 55 otoliths in which $73 \%$ agreement between the advanced readers was achieved.

The main issues identified on blue whiting age reading are still: the fact that the otoliths from some areas revealed to be more difficult to read (e.g. 27.2.a, 27.5.b); the first ring identification; edge type interpretation and false or double rings identification (Gonçalves, 2021).

During the workshop some of the otoliths from the exercise were polished, to help readers in the cases were the first age ring were not so evident, completely absent, or showing a growth pattern different from the expected. The polishing results revealed to be useful on the ring interpretation and to help in cases here the visible first ring size presents a size higher than the expected and the readers have doubts if an inner first ring are there. The hypothesis of the existence of a nonvisible first ring has been described in the otoliths from the adult fish as the otolith becomes thicker and wider.

Although, during the WKARBLUE3 progresses have been made and objective and more clear age reading guidelines had been constructed. The recurrent age reading issues still remain the same, e.g. the identification of the position of the first annual growth ring, false rings and interpretation of the edge. In order to overcome those problems and increase the accuracy on age classifications, age validation studies on blue whiting otoliths to solve growth rings interpretation, were further recommended and should be conducted.

### 1.4.4.5 Boarfish

Sampling of the commercial catch of boarfish has been included within the EU data collection framework since 2017. An age length key was produced in 2012 following increased sampling of a developing fishery. The age reading was conducted by DTU Aqua on samples from the three main fishery participants: Ireland, Denmark and UK (Scotland). No ageing has been carried out since 2012 although otoliths continue to be collected from the Irish fishery during routine catch sampling. In preparation for a benchmark assessment in 2023, an ageing exchange has been initiated via SmartDots.

### 1.4.4.6 Striped red mullet

In 2011, an otolith exchange was carried out, the second such exercise for the striped red mullet. For details see section 10.5.

### 1.4.4.7 Red gurnard

Age data are available for red gurnard from the EVHOE and IGFS groundfish surveys. Improvements in the understanding of the age structure of this stock would be improved by reading otoliths from other surveys in the assessment area (e.g. NS-IBTS, SCO-WCS, CGFS) which also contribute information on stock status in term of their CPUE series.

### 1.4.5 Current methods of compiling fisheries assessment data

Information on official, area misreported, unallocated, discarded and sampled catches have again this year been recorded by the national laboratories on the WG-data exchange sheet (MS Excel; for definitions see text table below) and sent to the stock co-ordinators and uploaded through InterCatch. Co-ordinators collate data using the either the sallocl (Patterson, 1998) application which produces a standard output file (sam.out) or InterCatch.

There are at present no specified criteria on the selection of samples for allocation to unsampled catches. The following general process is implemented by the species co-ordinators. A search is made for appropriate samples by gear (fleet), area, and quarter. If an exact match is not available the search will extend to adjacent areas, should the fishery extend to this area in the same quarter. Should multiple samples be available, more than one sample may be allocated to the unsampled catch. A straight mean or weighted mean (by number of samples, aged or measured fish) of the observations may be used. If there are no samples available the search will move to the closest non-adjacent area by gear (fleet) and quarter, but not in all cases.

It is not possible to formulate a generic method for the allocation of samples to unsampled catches for all stocks considered by WGWIDE. However full documentation of any allocations made are stored each year in the data archives (see below). It should be noted that when samples are allocated the quality of the samples may not be examined (i.e. numbers aged) and that allocations may be made notwithstanding this. The Working Group again encourages national data submitters to provide an indication of what data could be used as representative of their unsampled catches.

Following the introduction of the landings obligations for EU fisheries new catch categories had to be introduced from 2015 onwards. The catch categories used by the WGWIDE are detailed below:

| Official Catch | Catches as reported by the official statistics to ICES |
| :---: | :---: |
| Area misreported Catch | To be used only to adjust official catches which have been reported from the wrong area (can be negative). For any country the sum of all the area misreported catches should be zero. |
| BMS landing | Landings of fish below minimum landing size according to landing obligation |
| Logbook registered discards | Discards which are registered in the logbooks according to landing obligation |
| Discarded Catch | Catch which is discarded |
| WG Catch | The sum of the 6 categories above |
| Sampled Catch | The catch corresponding to the age distribution |

### 1.4.6 Quality of the Input data

Primary responsibility for the accuracy of national biological data lies with the national laboratories that submit such data. Each stock co-ordinator is responsible for combining, collating, and interpolating the national data where necessary to produce the input data for the assessments. A number of validation checks are already incorporated in the data submission spreadsheet currently in use, and these are checked by the co-ordinators who in the first instance report anomalies to the laboratory which provided the data.

Overall, data quality has improved and sampling deficiencies have been reduced compared to earlier years, partly due to the implementation of the EU sampling regulation for commercial catch data. However, some nations still have no (or inadequate) aged samples. Occasionally, no data are submitted such that only catch data from EuroStat is available, which are not aggregated quarterly but are yearly catch data per area.

The Working Group documents sampling coverage of the catches in two ways. National sampling effort is tabulated against official catches of the corresponding country (see stock specific sections). Furthermore, tables showing total catch in relation to numbers of aged and measured fish by area give a picture of the quality of the overall sampling programme in relation to where the fisheries are taking place. These tables are contained in the species sections of this report.

The national data on the amount and the structure of catches and effort are archived in the ICES InterCatch database. The data are provided directly by the individual countries and are highly aggregated for the use of stock assessments.

There exist gaps in some data series, in particular for historical periods. The WG has requested members to provide any national data reported to previous working groups (official catches, working group catches, catch-at-age and biological sampling data) not currently available to the WG. Furthermore, the WG recommends that national institutes increase national efforts to collate historic data.

A number of stock data problems relevant to data collections have been brought forward to the contact person in preceding years. Those that still apply are listed in table below for the information of ICES-Working Groups and RCMs as specified.

| Stock | Data Problem | How to be addressed in | By who |
| :--- | :--- | :--- | :--- |
| Northeast <br> lantic Mackerel | Submission of data | Data submissions must include all the data <br> outlined in the data call and be submitted by | National labor- <br> atories |


| Stock | Data Problem | How to be addressed in <br> the deadline. Data should include length distributions split by area and quarter. <br> Should the data submitter be unavailable after the data has been submitted (e.g. vacation) an alternative contact should be available who can be contacted in the event of any queries. | By who |
| :---: | :---: | :---: | :---: |
| Northeast Atlantic Mackerel | Discard and slippage information | Discard and slippage information is incomplete. All fleets, including demersal fleets should be monitored and sampled for discards and slipping. Data should be supplied to the coordinator by the submission deadline, accompanied by documentation describing the sampling protocol. | National laboratories, RCG NA, RCG NS\&EA |
| Northeast Atlantic Mackerel | Sampling deficienciesgeneral | All countries involved should provide sampling information. Increased cooperation between countries would help reduce redundancy and increase coverage. | National laboratories, RCG NA, RCG NS\&EA |
| Northeast Atlantic Mackerel | Sampling of foreign vessels | Any information available from the sampling of foreign vessels should be forwarded to the appropriate person in the national laboratory in order that they may use this information when compiling the data submission. | National laboratories; RCG NA, RCG NS\&EA |
| Horse Mackerel <br> - Western Stock | Missing sampling data for some parts of the distribution area (e.g. 27.2a, 7e) | Fishing nations to Sample age and length Distributions from commercial fleets | National Institutes |
| Horse Mackerel <br> - North Sea <br> Stock | Incomplete report of discards by non-pelagic fleet. | Reporting of discards by national institutes. | National Institutes |
| Horse Mackerel <br> - North Sea <br> Stock | Lack of maturity ogive both by age or length | Collection of information about maturity stage during regular biological sampling (otoliths) in commercial and survey fleets | National institutes |
| Horse Mackerel <br> - North Sea <br> Stock | Lack of length distributions in the discarded component | Sampling of length distribution of discarded individuals | National institutes |
| Horse Mackerel <br> - North Sea <br> Stock | Low contribution of countries to the estimation of the age and length distribution of catches | To ensure the sampling of age and length information from all catch fractions and all areas and within all quarters from all commercial fleets with a distribution of sampling effort over the year and areas in the North Sea | National institutes |
| Norwegian Spring-spawning Herring | Low sampling effort on some nations | Sampling effort should be increased by nations with little or no samples. | National labor- <br> atories; RCG <br> NS\&EA |
| Red gurnard | Species level catch reporting and sampling | Red gurnard catches should be reported to species level and with the appropriate codification. Where reported as mixed gurnards, this should be accompanied by documented procedures for estimating the proportion of red gurnard. | National laboratories |


| Stock | Data Problem | How to be addressed in | By who |
| :--- | :--- | :--- | :--- |
| Red gurnard | Discard and slippage <br> information | Discard rates for this species can be very high <br> (up to $100 \%$ of catch at a trip level). Alterna- <br> tive data sources and methods for estimation <br> (e.g. CCTV systems) should be investigated. | National labor- <br> atories |
| Red gurnard | Stock area | Red gurnard is found all along the Iberian <br> continental shelf. There are no records of <br> catches of red gurnards in SA5, and this area <br> could be removed from the data call. |  |
| Northeast At- | Submission of data | Data submissions must include all the data <br> outlined in the data call and be submitted <br> by the deadline. | National labor- <br> atories |
| lantic |  | Should the data submitter be unavailable af- <br> ter the data has been submitted (e.g. vaca- |  |

### 1.4.7 Quality control of data and assessments, auditing

As a quality control of the data and the assessment, WG participants were appointed as auditors for each stock. The primary aim of the auditing process is to check that the assessment and forecast has been conducted as detailed in the relevant stock annex. Auditors conducted checks of the assessment input data, assessment code (time permitting), draft WG report and draft advice sheet. Auditors completed an audit report upon completion (annex 4). Issues identified in the audit reports were followed up by the appropriate stock coordinator/assessor with updates made where appropriate.

### 1.4.8 Information from stakeholders

The procedure for the submission of inputs from stakeholders into the scientific advice changed in 2020. Instead of contributing information directly into the Advice Drafting Groups, information from stakeholders is now submitted directly to the expert group for consideration and inclusion into the draft advice, if applicable.

For WGWIDE stocks there are several instances of strong cooperation between research institutes and fishing industry stakeholder in the collection of data that is used in the assessments, e.g. the acoustic survey for Norwegian Spring Spawning herring, the extension of the IESSNS survey into the North Sea and several cases where industry vessels are collecting samples for catch monitoring. In these cases, the research institutes are coordinating the activities and bringing the results directly to the expert group(s).

A recent development that started around 2014 involves fishing industry organizations taking initiatives on their own, to collect additional information that is contributed to the expert groups. In many cases these research activities are undertaken in close cooperation with research institutes. During WGWIDE 2022, the following contributions from fishing industry research activities were reported to the working group:

1. PFA self-sampling report 2015-2021
2. Gonad sampling for mackerel and horse mackerel in support of the 2022 egg survey
3. Horse mackerel genetics
4. Using acoustics from commercial trawlers as potential indicators of abundance

### 1.4.8.1 PFA self-sampling report (WD02)

The Pelagic Freezer-trawler Association (PFA) initiated a self-sampling programme in 2015, aimed at expanding and standardizing ongoing fish monitoring programmes by the vessel quality managers on board of the vessels. An overview of the self-sampling in widely distributed pelagic fisheries from 2016 onwards is presented in the text table below.

| Year | Number Vessels | Number Trips | Number Days | Number Hauls | Catch (t) |  | Number Length Meas urements |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 | 9 | 45 | 591 | 1,307 | 113,900 | 193 | 65,212 |
| 2017 | 12 | 62 | 840 | 1,781 | 177,887 | 212 | 91,357 |
| 2018 | 16 | 86 | 1,219 | 2,677 | 253,237 | 208 | 170,306 |
| 2019 | 16 | 97 | 1,226 | 2,658 | 224,886 | 183 | 124,288 |
| 2020 | 17 | 112 | 1,424 | 3,038 | 305,282 | 214 | 163,955 |
| 2021 | 19 | 119 | 1,398 | 2,874 | 282,097 | 202 | 138,481 |
| 2022* | 18 | 62 | 733 | 1,694 | 144,718 | 197 | 65,457 |
| (all) | 583 | 7,431 | 16,029 | 1,502,007 | 819,056 | 9,490 | 819,056 |

*incomplete

A description of the different fisheries is included in the report. In 2022, a substantial blue whiting fishery was carried out south of the Porcupine back, an area that had hardly been fished in previous years.

In the 2022 self-sampling report, a standardized CPUE calculation has been included for the first time for most of the stocks. The standardized CPUE is based on a GLM model with a negative binomial distribution. The response variable is catch by week and vessel, with an offset of the log effort (number of fishing days per week) and explanatory variables year, GT category, month, division and depth category. An assumed technical efficiency increase of $2.5 \%$ per year has been included in the fitting of the model (Rousseau et al 2019)

### 1.4.8.2 Gonad sampling for mackerel and horse mackerel

During 2022, a dedicated PFA industry researcher carried out three sampling trips on-board of commercial trawlers with the aim to collect fresh and frozen gonad samples of mackerel and horse mackerel to aid the WGMEGS in determining the potential fecundity of mackerel and horse mackerel. In order to determine potential fecundity, it is necessary to collect the gonad samples just prior to spawning Using a commercial vessel for that sampling proved to be an efficient way of collecting the samples as the vessels were targeting mackerel and horse mackerel during the period that the start of spawning could be anticipated.

During 2021 and 2022 DTU Aqua and the Danish Pelagic Producers Organization collected gonads from mackerel in the North Sea, that were fished as bycatch in other fisheries. The gonads have been stored in formalin for accurate maturity staging and egg counting. The sampling has been conducted throughout the year to get more insight into the spawning cycle in the North

Sea. The sampling has been coordinated with the 2021 North Sea mackerel egg survey. Results of the sampling are expected to be available in 2023.

In addition, the PFA is continuing the collection of mackerel gonads throughout the year, as a means of following the maturity development of mackerel (and to a limited extent horse mackerel).


Figure 1.4.8.2 Overview of PFA gonad sampling for mackerel and horse mackerel during 2021 and 2022.

### 1.4.8.3 Horse mackerel genetic stock identification

ICES has long considered horse mackerel in the northeast Atlantic to consist of three stocks, the separation of which was based on a variety of factors including the temporal and spatial distribution of the fisheries, the observed egg and larval distributions, information from acoustic and trawl surveys and from parasite infestation rates (see ICES, 2015). Further refinements of the definitions of stock units were based on the results from the EU-funded HOMSIR project (20002003), which utilised a multidisciplinary approach including various genetic approaches (allozymes, mitochondrial DNA and microsatellites), the use of parasites as biological tags, body morphometrics, otolith shape analysis and the comparative study of life history traits (growth, reproduction and distribution) (Abaunza et al., 2008). However, there remained unresolved issues particularly in areas where mixing between stocks was likely to occur, e.g. between divisions 7 e and 7 d and in division 4 a , and also no reliable method to continue ongoing monitoring.

In response to this, the Pelagic Freezer Trawler Association (PFA) contracted the Wageningen University and Research (WUR) in 2015 to undertake a study on North Sea horse mackerel (Brunel et al., 2016). The primary aim of the study was to improve the data quality used for an analytical stock assessment model of North Sea horse mackerel.

The management boundary between the Western and North Sea stocks in the English Channel (corresponding to the separation between divisions 7 e , western Channel and 7d, eastern Channel) does not correspond to a real biological boundary, as mixing of the two stocks is known to occur in division 7 d in autumn and winter (Brunel et al., 2016). The catches taken in 7 d are officially considered as being North Sea horse mackerel and represent c. $80 \%$ of the catches from this stock. An unknown proportion of this catch is likely from the western stock, which interferes with the cohort signal in the catch at age matrix, hampering the development of an age-
structured assessment model for the North Sea stock. Developing methods to separate catches from the western stock from catches from the North Sea stock in division 7d are therefore necessary to improve the quality of the catch information for the North Sea stock. Within the project, two pilot studies, based on chemical fingerprint analyses and genetics, were conducted to investigate new methods to determine stock structure and to develop techniques to identify the stock origin of the catches taken in the eastern English Channel.

As part of the project, WUR contracted University College Dublin, Ireland (UCD) to undertake a pilot study to develop a method of genetic stock identification for discriminating North Sea and Western horse mackerel (Brunel et al., 2016). The aims of the pilot study were to firstly develop and validate at least 24 polymorphic microsatellites markers in horse mackerel and secondly to screen spawning fish collected in 2015 from the Western and North Sea stocks to establish a genetic baseline of the spawning stocks and test the presence of population structure. Recently developed Next Generation Sequencing (NGS) and Genotyping by Sequencing (GBS) based approaches, which were developed on cod (Gadus morhua Linnaeus, 1758), boarfish (Capros aper Lacépède, 1802) and 6a, 7b-c herring were used for marker development and screening of spawning samples (Farrell et al., 2016; Vartia et al., $2014 \& 2016$ ). The pilot study successfully identified a large number of novel microsatellites, however initial data analyses were confounded by a poor-quality sequencing run and as such the discrimination power between the western and North Sea sample was low. This resulted in the pilot study being unable to separate the two stocks conclusively and unequivocally.

In an effort to resolve these uncertainties, the Northern Pelagic Working Group (NPWG) of the European Association of Fish Producers Organisations (EAPO) contracted EDF Scientific Limited, Ireland and Jens Carlsson (UCD) to undertake a comprehensive genetic stock identification study on Horse Mackerel (Farrell \& Carlsson, 2018). Sampling was conducted over three consecutive years and three spawning seasons and covered a large area of the distribution of the species including the Western, North Sea and Southern stock areas and also West African waters. In total 33 population samples, comprising 2,295 individual fish were collected from 2015 to 2017 across the study area (figure 1.4.8.2). Spawning samples were analysed with a panel of 37 novel, putatively neutral microsatellite markers and statistical analyses ( $F_{S T}$, structure, assignment testing, mixed stock analyses and FCA analyses) indicated that horse mackerel in the northeast Atlantic region does not represent a single biological unit. A high level of species misidentification in the West African samples was also observed. On the highest level there are mixed species catches in African waters, a clear separation of the southern North Sea from other regions and further, less pronounced, structure along the northeast Atlantic continental shelf. Exploratory assignment testing and mixed stock analysis of the western and North Sea baselines indicated a success rate of c.60-65\% for self-assignment. This was considered relatively low and is due to the relatively low genetic differentiation between the populations at putatively neutral loci. Despite this, further exploratory assignment testing and mixed stock analysis of the fish caught outside spawning time in the northern North Sea and western English Channel indicated that a large component of these fish belonged to the Western stock. No samples from the eastern English Channel (7d) were available for testing.


Figure 1.4.8.2. (left panel) The horse mackerel samples collected from 2015 to 2017 and (right panel) those included in the baseline dataset.

The results showed that the genetic information could be used for mixed stock analyses and that the information could be used to delineate the range of the North Sea stock - information that could be taken into account by fisheries management. However, it was suggested in the project report that further genetic analyses were warranted to increase the numbers and types of genetic markers available for this species. This would improve stock discrimination, mixed stock analyses and individual assignment capacity.

In 2019 the NPWG contracted Uppsala University, Sweden and EDF Scientific Limited to apply the same Whole Genome Sequencing (WGS) and pooled population sequencing (Pool-Seq) approaches, that had successfully been developed for herring (see Han et al., 2020), on the horse mackerel samples (Fuentes-Pardo et al., 2020; in review). The aims of the study were to identify informative genetic markers for the stock identification of horse mackerel and to estimate the extent of genetic differentiation among the sampled populations. The samples included in the genome study (figure 1.4.8.3) were primarily a subset of the baseline samples analysed in Farrell and Carlsson (2018). One additional sample from the Alboran Sea in the Mediterranean Sea was provided by the ATLAS Project (https://www.eu-atlas.org/). Samples were aggregated into 12 pools based on spatial and temporal proximity, thus broadly representing most of the geographical range of the species in the northeast Atlantic and the western part of the Mediterranean Sea. Each pooled sample was sequenced and mapped to the newly developed horse mackerel genome (Genner and Collins, 2022).


Figure 1.4.8.3. Sampling locations of the Atlantic horse mackerel included in this study. (Left) Sample batches collected at each location.

The results indicated that while the populations only differed in a small fraction of their DNA (< $1.5 \%)$, these genetic differences were significant as they likely represented natural selection and local adaptation of populations. A small panel of the highly differentiated genetic variants were validated by genotyping individuals from each population ( $n=24$ per pop), which demonstrated that the variants could be used as informative molecular markers for the genetic identification of the main stock divisions of horse mackerel. The results, based on the analysed samples, indicated that the North Sea horse mackerel are a separate and distinct population. The samples from the Western stock, west of Ireland and the northern Spanish shelf, and the northern part of the Southern stock, northern Portugal, appear to form a genetically close group. There was significant genetic differentiation between the northern Portuguese samples and those collected in Southern Portuguese waters, with those in the south representing a separate population. The North African and Alboran Sea samples were distinct from each other and from all other samples.
These results indicated that a further large-scale analysis of samples, with a greater temporal and spatial coverage, with the newly identified molecular markers was required to test and reassess the current stock delineations. To this end a new genetic tool has been developed to enable higher throughput of samples and also to standardise the genotyping approach. The DNA TRACEBACK® Fisheries Array (IdentiGEN, Dublin, Ireland) contains c. 4,000 markers that represent informative regions in the 24 chromosomes. The NPWG has agreed to fund the next part of the analysis and it expected that results will be available for presentation at WGWIDE 2023.

### 1.4.8.4 Use of acoustic data from commercial trawlers as an indicator of abundance

For many years already, acoustic data has been recorded on commercial fishing vessels of the Pelagic Freezer-trawler Association. Many terabytes of data are now available. The equipment is sophisticated, the echo sounders are calibrated and the high fish density regions are visited during the fishing trips with extensive spatio-temporal coverage. But how can we derive meaningful metrics from the acoustic data collected by fishing vessels?

Currently a method is being developed and tested at Wageningen Marine Research to utilize acoustic data collected during commercial fishing operations for biomass estimation. The case study that is explored is the blue whiting stock during the spawning season in March-April. The International Blue Whiting Spawning Stock Survey west of the British Isles (IBWSS) is carried out annually during the spawning season. At the same time, the commercial fishery is taking place in that area.

The acoustic observations during both, the scientific survey and the fishing trips have been processed using the same methods: cleaning noise, removing unwanted regions (e.g. surface and bottom reflections), manually drawn polygons that confine the backscatter regions that can be attributed to the blue whiting, and results exported as integrated acoustic backscatter per nautical mile.

The main difference between the survey and the fishing vessel observations comes from the patterns in the acoustic tracks. The fishing vessels observations comes from localised recordings from high density spots during the actual fishery. The biased property of the acoustic tracks of the fishing vessels makes it difficult to fit them into a statistically meaningful survey design. The method now being developed at Wageningen Marine Research is taking advantage of the good overlap between the commercial fishery and the scientific survey to develop a method to transform the targeted fishing vessel data into unbiased 'survey-like' estimates of abundance.

All commercial acoustic data is broken down into weeks for each fishing trip and polygons were generated around these weekly tracks. Next, synthetic transects were generated in a similar fashion to the survey transects with 1 nautical mile differences between the sampling units and with a predefined inter-transect distance. Acoustic values are assigned to the synthetic transects by taking the average of the acoustic observations within the search radius around each point on the transect.

The use of synthetic transects gives the possibility of interpreting the data from fishing vessels in a similar way as the survey procedures. However, there are two important parameters that need to be determined to generate these synthetic transects: the distance between the transects, and the search radius around the points in the synthetic transects. We looked at the correlation between the survey data and the synthetic transect data with different transect spacing and search radius. The spacing of 0.2 degrees and search radius of 1.2 nautical miles gives a coefficient of determination of 0.94 . This promising correlation encourages us to generate time series that can be used to generate trends independent from the survey data.

### 1.5 Comment on update and benchmark assessments

Updates were presented to the WG for all the eight stocks in the group.
Western and North Sea horse mackerel were assessed on basis of a benchmark that took place in January 2017 (ICES, 2017) and NEA mackerel on an inter-benchmark that took place in 2019 (ICES 2019a). Norwegian spring spawning herring was assessed using the XSAM
implementation benchmarked in 2016. The Blue whiting SAM assessment was introduced following a benchmark in 2012. Since this time, an inter-benchmark in 2016 incorporated the use of preliminary in-year catch data with the stock weights in the assessment year estimated from catch sampling incorporated in 2019 (previously the average of the most recent three years was used). The acoustic survey time series was updated in 2020 following recalculation by the StoX platform with minor updates to the historic index. The red gurnard assessment conducted at WGWIDE 2022 followed a benchmark in February 2021 (WKWEST) during which an index of abundance based on a number of bottom trawl surveys was developed.
The remaining two stocks addressed by the WG (boarfish and striped red mullet) have not been benchmarked recently but were still assessed by the WG.

### 1.6 Planning future benchmarks

Two of the WGWIDE stocks are yet to be benchmarked; Boarfish for which an exploratory surplus production model is used and Striped red mullet for which there is no assessment in place. Boarfish is scheduled to be benchmarked in 2023. Ongoing sampling of the commercial catch, an expanded acoustic survey time series and advances in modelling techniques e.g. VAST will be explored with a view to improving the current assessment and exploring alternative assessment models. Research projects underway for Striped red mullet are due to be completed in the near future and will inform the proposed benchmark for 2024.

The current implementation of the Stock Synthesis model for the assessment of Western horse mackerel has been used since the benchmark in 2017. A number of issues with the assessment and opportunities for improvement were identified at WGWIDE 2021 and a benchmark was proposed and scheduled for 2023. Unfortunately, this could not be achieved and the benchmark had to be postponed. The working group considers that the justification for a benchmark remains strong and it should now take place in 2024 along with North Sea Horse mackerel, which is currently a category 3 assessment with opportunities to improve based on both new data sources and models. Genetic studies (see section 1.4.8.3) have shown that Western and North Sea horse mackerel are genetically distinct. Currently, catches are assigned to stocks on the basis of ICES division and quarter although it is suspected that catches occur on mixed stocks.

WGWIDE 2022 is also proposing benchmark workshops take place for Northeast Atlantic Mackerel and Norwegian Spring Spawning Herring. The benchmark for NEA Mackerel should be precede by a workshop to review the current assumptions with regard to stock structure (components). Terms of reference for the workshop (WKMACEVAL) were drafted by WGWIDE 2022 and will inform a recommendation to ACOM for the WK. Exploratory work is already underway or is planned on a number of issues related to the mackerel assessment including dealing with individual high catch rates in the swept area survey (to be considered by WGISDAA), DEPM vs AEPM methodologies for the egg survey time series, inclusion of additional ages from the tagging dataset, increasing the assessment recruitment age and updating the SAM configuration. The proposed benchmark of Norwegian Spring Spawning Herring will explore issues such as the splitting of exiting survey indices, inclusion of additional surveys, assumptions on maturity in the most recent years and implementation in the mainstream SAM model, which has recently been developed to offer the functionality of the current XSAM model.

Issue lists and benchmark scoring sheets for each of the stocks proposed for benchmarking by WGWIDE 2022 were reviewed and updated during the meeting.

The current status of the WGWIDE stocks with respect to benchmarking is summarised below:

| Stock | Benchmark History | WGWIDE 2022 Proposal |
| :--- | :--- | :--- |
| Boarfish | Benchmark scheduled for 2023 |  |
| Red gurnard | Full benchmark 2021 |  |
| Norwegian Spring | Full benchmark 2016 | Full benchmark |
| Spawning herring | Full benchmark 2017 |  |
| Weference point inter-benchmark 2019 | Full benchmark |  |
| mackerel | 2022 scheduled benchmark postponed | Full benchmark |
| North Sea | Full benchmark 2017 |  |
| horse mackerel | Full benchmark 2014 | Full benchmark |
| Northeast Atlantic | Full benchmark 2017 |  |
| Inter-benchmark 2019 | Full benchmark |  |
| Slue whiting | Never benchmarked | Inter-benchmark 2016 |

### 1.7 Scientific advice and management of widely distributed and migratory pelagic fish

### 1.7.1 General overview of management system

The North East Atlantic Fisheries Commission (NEAFC) is the Regional Fisheries Management Organisation (RFMO) for the North East Atlantic. NEAFC is an end user of ICES advice and provides a forum for its contracting parties (Coastal States and fishing parties) to manage the exploitation of straddling stocks that occur in several EEZs and international waters such as WGWIDE stocks North East Atlantic Mackerel, Blue Whiting and Norwegian Spring Spawning herring (also known as Atlanto-Scandian herring). There are 6 contracting parties to NEAFC: Denmark (in respect of the Faroe Islands and Greenland), European Union, Iceland, Norway, Russian Federation and the UK. The management of Western horse mackerel is not considered by NEAFC with sharing subject of separate agreements between EU, Norway and the UK.

### 1.7.2 Management plans

Catch advice in recent years for two stocks considered by WGWIDE has been given on the basis of an agreed long term management strategy:

- A long term management strategy for Norwegian spring spawning herring was agreed by the European Union, the Faroe Islands, Iceland, Norway and Russian Federation in

2018 following an evaluation by ICES (WKNSSHMSE, ICES, 2018a) which found it to be precautionary. The plan is based on a target fishing mortality of 0.14 when the stock is above $\mathrm{B}_{\mathrm{pa}}$. Should SSB fall below $\mathrm{B}_{\mathrm{pa}}$, the target fishing mortality is linearly reduced to 0.05 at and below Blim. The plan incorporates TAC change limits of $-20 \%$ and $+25 \%$ which are suspended when below $\mathrm{B}_{\mathrm{pa}}$ and $10 \%$ interannual transfer which is suspended when below Blim. The plan is scheduled for review no later than 2023. Although the plan is agreed by the parties involved in the fishery and ICES advice is based on application of the management strategy, there has been no agreement on the relative catch share since 2013 with the total unilaterally declared quotas exceeding the management plan based catch advice since this time.

- A long term management strategy for Blue Whiting was agreed by the European Union, the Faroe Islands, Iceland and Norway in 2016 following an evaluation by ICES (WKBWMS, ICES, 2016) in 2016 which found it to be precautionary. The plan is based on a target fishing mortality equivalent to $\mathrm{F}_{\mathrm{msy}}(0.32)$ when the stock is above $B_{\text {pa }}$. Should SSB fall below $\mathrm{B}_{\mathrm{pa}}$, the target fishing mortality is linearly reduced to 0.05 at and below Blim. The plan incorporates TAC change limits of $+/-20 \%$ which are suspended when below $\mathrm{B}_{\mathrm{pa}}$ and $10 \%$ interannual transfer. No agreement on quota shares has been reached since 2015 and catches have exceeded advice since this time. At WGWIDE 2022, the assessment and forecast indicate a strong increase in SSB and catch advice for 2023 is an $81 \%$ increase on that for 2022 . It should be noted that the management plan clause permitting such an increase (paragraph 6b) was not tested in the 2016 evaluation. Since the management plan target fishing mortality is equivalent to FMSY, the MSY approach results in the same advice as the LTMS.

There is no currently agreed management strategy for either Northeast Atlantic Mackerel or Western horse mackerel. Strategies have been proposed and evaluated but agreement has not yet been reached on their implementation such that catch advice has been given on the basis of the MSY approach.

### 1.7.3 Comparison of advice, TAC and catches

This section presents an overview of the time-series (2010 to present) of ICES catch advice, TAC (either agreed between all fishing parties or a sum of unilaterally declared quotas) and ICES estimates of total catch for Norwegian spring spawning herring, Western horse mackerel, Northeast Atlantic mackerel and blue whiting. The overviews are based on the history of advice, management and catch as reported in the ICES single stock advice documents. The information is summarised in tables 1.7.3.1-4 and figure 1.7.3.1. Figures 1.7.3.2-5 compare the TAC and advice, catch and advice and catch and TAC and catch and the sum of unilateral quotas respectively, each expressed as a percentage difference e.g. (TAC-advice)/advice.
For Norwegian spring-spawning herring some deviations between TAC and advice occurred between 2010-2013, but from 2014 on the sum of unilateral quotas has been in excess of the scientific catch advice which was based on the agreed management plan. Catches have likewise been in excess of the scientific advice and close to the sum of unilateral quotas..

Western horse mackerel: some deviations between TAC and advice have been occurring during the time-series presented, but there does not appear to be a clear trend. No management plan is applicable for western horse mackerel. Catches have generally been at or below the agreed TAC.
Northeast Atlantic mackerel has not had agreed TACs during the period presented. The sum of unilateral quota has always been higher than the scientific advice. Catches have on average been $41 \%$ above the scientific advice and close to the sum of unilateral quota.

Blue whiting: up to 2013, the agreed management plan has been followed. From 2014 onwards, the sum of unilateral quota has been in excess of the scientific advice and the agreed management plan. Catches have likewise been in excess of the scientific advice and close to the sum of unilateral quota.

In summary: although long term management plans exist for Norwegian spring-spawning herring, Northeast Atlantic mackerel and Blue whiting, they have not been instrumental in limiting the TACs to the pre-agreed values. While the Coastal States may have agreed on the TACs for these stocks, there was no agreement on the distribution of quota between Coastal States. As a consequence, the sum of unilateral quota and the catches have been in excess of the scientific advice and the rules of the management plans.

Table 1.7.3.1. Overview of scientific advice, agreed TAC, sum of unilateral quotas and catch for Norwegian Spring Spawning Herring.

| Yr | Advice Basis | Ad- <br> vised <br> Catch <br> (t) | TAC (t) | Unilateral Quotas (t) | Catch <br> (t) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | Do not exceed HCR | 1483 | 1483 |  | 1457 |
|  |  | 000 | 000 |  | 000 |
| 2011 | Scenarios | 1170 | 988000 |  | 993 |
|  |  | 000 |  |  | 000 |
| 2012 | Follow management plan | 833 | 833000 |  | 826 |
|  |  | 000 |  |  | 000 |
| 2013 | Follow management plan | 619 | 619000 | 692 | 685 |
|  |  | 000 |  | 000 | 000 |
| 2014 | Follow management plan | 418 | 418487 | 436 | 461 |
|  |  | 000 |  | 000 | 000 |
| 2015 | Follow management plan | 283 |  | 328 | 329 |
|  |  | 000 |  | 000 | 000 |
| 2016 | Follow management plan | 317 |  | 377 | 383 |
|  |  | 000 |  | 000 | 174 |
| 2017 | Follow management plan | 646 |  | 805 | 721 |
|  |  | 075 |  | 142 | 566 |
| 2018 | Follow management plan | 384 |  | 546 | 592 |
|  |  | 197 |  | 448 | 899 |
| 2019 | Follow management strategy ( $\mathrm{F}_{\mathrm{mgt}}=0.14, \mathrm{~B}_{\mathrm{mgt}}=3.184$ | 588 | 588562 | 773 | 777 |
|  | $\mathrm{Mt})$ | 562 |  | 750 | 165 |
| 2020 | Follow management strategy ( $\mathrm{F}_{\mathrm{mgt}}=0.14, \mathrm{~B}_{\mathrm{mgt}}=3.184$ | 525 | 525594 | 693 | 720 |
|  | $\mathrm{Mt})$ | 594 |  | 915 | 937 |
| 2021 | Follow management strategy ( $\mathrm{F}_{\mathrm{mgt}}=0.14, \mathrm{~B}_{\mathrm{mgt}}=3.184$ | 651 | 561033 | 881 | 851 |
|  | $\mathrm{Mt})$ | 033 |  | 097 | 813 |
| 2022 | Follow management strategy ( $\mathrm{F}_{\mathrm{mgt}}=0.14, \mathrm{~B}_{\mathrm{mgt}}=3.184$ | 598 | 598588 | 827 |  |
|  | Mt ) | 588 |  | 963 |  |
| 2023 | Follow management strategy ( $\mathrm{F}_{\mathrm{mgt}}=0.14, \mathrm{~B}_{\mathrm{mgt}}=3.184$ | 511 |  |  |  |
|  | $\mathrm{Mt})$ | 171 |  |  |  |

Table 1.7.3.2. Overview of scientific advice, agreed TAC, sum of unilateral quotas and catch for Western Horse Mackerel.

| Yr | Advice Basis | Ad- <br> vised <br> Catch <br> (t) | TAC (t) | Unilateral Quotas (t) | Catch <br> (t) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | Follow proposed management plan | 180 | 185000 |  | 203 |
|  |  | 000 |  |  | 112 |
| 2011 | Scenarios | 229 | 184000 |  | 193 |
|  |  | 000 |  |  | 698 |
| 2012 | MSY framework | 211 | 183000 |  | 169 |
|  |  | 000 |  |  | 858 |
| 2013 | MSY framework | 126 | 183000 |  | 165 |
|  |  | 000 |  |  | 258 |
| 2014 | MSY approach | 110 | 135000 |  | 136 |
|  |  | 546 |  |  | 360 |
| 2015 | MSY approach | 99 | 99300 |  | 98419 |
|  |  | 304 |  |  |  |
| 2016 | MSY approach | 126 | 126000 |  | 98811 |
|  |  | 000 |  |  |  |
| 2017 | MSY approach | 69 | 95500 |  | 82961 |
|  |  | 186 |  |  |  |
| 2018 | MSY approach | 117 | 115470 |  | 101 |
|  |  | 070 |  |  | 682 |
| 2019 | MSY approach | 145 | 136376 |  | 124 |
|  |  | 237 |  |  | 947 |
| 2020 | MSY approach | 83 | 81796 |  | 76422 |
|  |  | 954 |  |  |  |
| 2021 | MSY approach | 81 | 81375 |  | 81557 |
|  |  | 376 |  |  |  |
| 2022 | MSY approach | 71 | 71138 |  |  |
|  |  | 138 |  |  |  |
| 2023 | MSY approach | 0 |  |  |  |

Table 1.7.3.3. Overview of scientific advice, agreed TAC, sum of unilateral quotas and catch for Northeast Atlantic Mackerel.

Y Advice Basis
r

Harvest control rule

| Ad- | TAC (t) | Unilat- | Catch |
| :--- | :--- | :--- | :--- |
| vised |  | eral | (t) |
| Catch |  | Quotas |  |
| (t) |  | $(\mathbf{t})$ |  |

572691305
875515

| Scenarios | 672 | 929943 |  | 946661 |
| :---: | :---: | :---: | :---: | :---: |
|  | 000 |  |  |  |
| Follow the management plan | 639 | 938410 |  | 892353 |
|  | 000 |  |  |  |
| Follow the management plan | 542 | 857319 |  | 931732 |
|  | 000 |  |  |  |
| Follow the management plan | 1011 |  | $1400$ | 1393 |
|  | 000 |  |  | 000 |
| Follow the management plan | 906 | $\begin{array}{r} 1054 \\ 000 \end{array}$ | $\begin{array}{r} 1208 \\ 719 \end{array}$ | $\begin{array}{r} 1208 \\ 990 \end{array}$ |
|  | 000 |  |  |  |
| MSY approach | 773 | 895900 | $\begin{array}{r} 1047 \\ 432 \end{array}$ | $\begin{array}{r} 1094 \\ 066 \end{array}$ |
|  | 840 |  |  |  |
| MSY approach | 857 | $\begin{array}{r} 1020 \\ 996 \end{array}$ | $\begin{array}{r} 1191 \\ 970 \end{array}$ | 1155944 |
|  | 000 |  |  |  |
| MSY approach | $\begin{aligned} & 550 \\ & 948 \end{aligned}$ | 816797 | 999929 | 1026437 |
|  |  |  |  |  |
| MSY approach | 770 | 653438 | 864000 | 840021 |
|  | 358 |  |  |  |
| MSY approach | 922 | 922064 | $\begin{array}{r} 1090 \\ 879 \end{array}$ | $\begin{array}{r} 1039 \\ 513 \end{array}$ |
|  | 064 |  |  |  |
| MSY approach | 852 | 852284 | $\begin{array}{r} 1119 \\ 103 \end{array}$ | $\begin{array}{r} 1081 \\ 540 \end{array}$ |
|  | 284 |  |  |  |
| MSY approach | 794 | 794920 | $\begin{array}{r} 1188 \\ 227 \end{array}$ |  |
|  | 920 |  |  |  |
| MSY approach | 782 |  |  |  |
|  | 066 |  |  |  |

Table 1.7.3.4. Overview of scientific advice, agreed TAC, sum of unilateral quotas and catch for Blue Whiting.

| Yr | Advice Basis | Ad- <br> vised <br> Catch <br> (t) | TAC (t) | Unilateral Quotas (t) | Catch (t) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | Follow the agreed management plan | 540 | 548000 |  | 540 |
|  |  | 000 |  |  | 000 |
| 2011 | Scenarios | 40 | 40100 |  | 105 |
|  |  | 000 |  |  | 000 |
| 2012 | Follow the agreed management plan | 391 | 391000 |  | 384 |
|  |  | 000 |  |  | 000 |
| 2013 | Follow the agreed management plan | 643 | 643000 |  | 626 |
|  |  | 000 |  |  | 000 |
| 2014 | Follow the agreed management plan | 948 | 1200 |  | 1155 |
|  |  | 950 | 000 |  | 000 |
| 2015 | Follow the agreed management plan | 839 | 1260 |  | 1396 |
|  |  | 886 | 000 |  | 244 |
| 2016 | MSY approach | 776 | 776000 | 1147 | 1183 |
|  |  | 000 |  | 000 | 187 |
| 2017 | MSY approach | 1342 | 1342 | 1675 | 1558 |
|  |  | 330 | 330 | 400 | 061 |
| 2018 | Long-term management strategy | 1387 | 1387 | 1727 | 1711 |
|  |  | 872 | 872 | 964 | 477 |
| 2019 | Long-term management strategy | 1143 | 1143 | 1483 | 1515 |
|  |  | 629 | 629 | 208 | 527 |
| 2020 | Long-term management strategy | 1161 | 1161 | 1478 | 1495 |
|  |  | 615 | 615 | 358 | 248 |
| 2021 | Long-term management strategy | 929 | 929292 | 1157 | 1143 |
|  |  | 292 |  | 604 | 450 |
| 2022 | Long-term management strategy | 752 | 752736 | 1107 |  |
|  |  | 736 |  | 529 |  |
| 2023 | Long-term management strategy | 1359 |  |  |  |
|  |  | 629 |  |  |  |

## Advice, TAC and catch



Figure 1.7.3.1: Overview of scientific advice, agreed TAC (or sum of unilateral quota) and catch

## TAC over advice



Figure 1.7.3.2: Relative deviations of TAC over advice. Red line indicates average relative deviation over the time series shown.


Figure 1.7.3.3: Overview of catch over advice

## Catch over TAC



Figure 1.7.3.4: Overview of catch over TAC

## Catch over Summed unilateral Quota



Figure 1.7.3.5: Overview of catch over sum of unilateral quotas.

### 1.8 General stock trends for widely distributed and migratory pelagic fish

WGWIDE 2022 has carried out the stock assessments of the following widely distributed and migratory pelagic species: boarfish, red gurnard, Norwegian spring spawning herring, Western horse mackerel, North Sea horse mackerel, Northeast Atlantic mackerel, Striped red mullet and Blue whiting.

Analytical (category 1) assessments are available for the four species that make up the bulk of the biomass of pelagic species in the Northeast Atlantic:

- Northeast Atlantic mackerel
- Norwegian spring spawning herring
- Blue whiting
- Western horse mackerel

The time series of the combined catch of these four stocks since 1988 is shown in figure 1.8.1. The highest combined catch (approx. 4 million tonnes) for these four species was been taken in 2004 and 2005. In the most recent 6 years the total catch has been composed of $\sim 45 \%$ blue whiting, $\sim 33 \%$ mackerel, $\sim 18 \%$ herring and $\sim 3 \%$ horse mackerel.

stock her.27.1-24a514a \| hom.27.2a4a5b6a7a-ce-k8 \| mac.27.nea\| whb.27.1-91214

Figure 1.8.1: Catch of blue whiting, mackerel, western horse mackerel and Norwegian spring spawning herring

An overview of the key variables for each of the stocks (SSB, fishing mortality and recruitment), is shown in Figure 1.8.2. Stock sizes of herring, mackerel and blue whiting have been declining from historical highs in the recent years, but remain above their respective MSY B trigger reference
point values with the exception of Western Horse Mackerel which has been increasing from a historic low in 2017 but is considered to be below Blim. The Blue Whiting SSB has increased in the most recent year following strong recent recruitment.

Fishing mortality for herring, horse mackerel and mackerel has been around FmSy in the most recent period. Fishing mortality for blue whiting has been above FMSY for much of the time series.

Recruitment estimates for blue whiting and herring are on a comparable scale (billions) and are substantially higher and more variable than those for horse mackerel (with the exception of the 1982 year-class) and mackerel.


Figure 1.8.2: top - SSB (million tons), middle - fishing mortality and bottom - recruitment (billions) of Norwegian spring spawning herring, western horse mackerel, Northeast Atlantic mackerel and blue whiting from the WGWIDE 2022 update assessments.

An overview of stock weight-at-age for mackerel and blue whiting is shown in figures 1.8.3 and 1.8.4.

For mackerel, a decline in weight at age started around 2005 for most ages. In more recent years, this has ceased with increases for younger fish noted since 2012.

Weight-at-age of blue whiting shows substantial fluctuations over time. For most ages, a decline in weight at age has been observed from 2010 although this appears to have ceased and, for some ages reversed in the most recent years.


Figure 1.8.3: Stock weight-at-age of NEA mackerel

WHB.27.1-91214 stock weight at age


Figure 1.8.4: Stock weight at age of blue whiting
WGWIDE (and its precursors WGMHSA and WGNPBW) have been publishing catch per statistical rectangle plots in their reports for many years. Catch by rectangle has been compiled by WG members and generally provide an estimate of total catch per rectangle (although catch by rectangle data do not represent the official catches and cannot be used for management purposes). In general, the total annual catches by rectangle are within $10 \%$ from the official catches. In the individual stock report sections, the catch by rectangle is been presented by quarter for the most recent year. For this overview, WGWIDE has collated all the catch by rectangle data that is available for herring, blue whiting, mackerel and horse mackerel. For horse mackerel and mackerel, a long time series is available, starting in 2001 (horse mackerel) and 1998 (mackerel). The time series for herring and blue whiting are shorter (from 2011) although additional information could still be derived from earlier WG reports.


Figure 1.8.5: Catch of mackerel (tonnes) by year and rectangle. Catch by rectangle data do not represent the official catches and cannot be used for management purposes. In general, the total annual catches by rectangle are within 10\% from the official catches.


Figure 1.8.6: Catch of horse mackerel (all stocks, tonnes) by year and rectangle. Catch by rectangle data do not represent the official catches and cannot be used for management purposes. In general, the total annual catches by rectangle are within $10 \%$ from the official catches.


Figure 1.8.7: Catch of blue whiting (tonnes) by year and rectangle. Catch by rectangle data do not represent the official catches and cannot be used for management purposes. In general, the total annual catches by rectangle are within $10 \%$ from the official catches.


Figure 1.8.8: Catch of Norwegian spring-spawning (Atlanto-scandian) herring (tonnes) by year and rectangle. Catch by rectangle data do not represent the official catches and cannot be used for management purposes. In general, the total annual catches by rectangle are within $10 \%$ from the official catches.

### 1.9 Ecosystem considerations for widely distributed and migratory pelagic fish species

A number of studies demonstrate that environmental conditions (physical, chemical and biological) can significantly influence stock productivity by changing the level of recruitment, growth rates, survival rates, or inducing variations in their geographical distribution (e.g. Skjoldal et al., 2004, Sherman and Skjoldal 2002). It has been acknowledged that future lines of work in stock assessment should take ecosystem considerations into account in order to reduce the levels of uncertainty regarding the present and future status of commercial stocks. Hence, WGWIDE encourages further work to be carried out on ecosystem considerations linked to widely distributed fish stocks including NEA mackerel, Norwegian spring-spawning herring, blue whiting and horse mackerel. A close collaboration with the Working Group on Integrated Assessment of Norwegian Sea (WGINOR; ICES 2018b; 2022), and hopefully other relevant Integrated Assessment groups within ICES in the near future, will help in operationalizing the ecosystem approach for the widely distributed pelagic stocks assessed by WGWIDE. The text below was largely provided by WGINOR (ICES 2022). The updated text and figures below include summary of Norwegian Sea ecosystem status on climate variability, circulation pattern, recent trends in oceanography, phytoplankton production, zooplankton biomass, pelagic fish biomass and pelagic fish spatial distribution in the Norwegian Sea. The ecosystem status summary shown below is intended for a wide audience, including scientists, teachers, students, decision-makers, and the public interested in the Norwegian Sea ecosystem and marine environmental issues in general. It is prepared by the ICES working group on integrated ecosystem assessment for the Norwegian Sea (WGINOR). It is a summary of the scientific information prepared by the group and does not constitute ICES advice.

## Highlights

- The recent 3-4 year trend of colder and fresher Atlantic inflow into the Norwegian Sea has ceased; however, the extent of Arctic Water is still increasing.
- Annual primary production was higher and spring blooms lasted longer for the period 2013-2020 compared to earlier years of time series which begins in 2003. Possible cause is increased inflow of cold and fresh Arctic water.
- Zooplankton biomass declined from around mid-2000's and has since remained at a lower level.
- The biomasses of Norwegian spring-spawning herring increased in the last year, following the recruitment of a strong year class. Mackerel and blue whiting biomasses continued to decline as in recent years. Recruitment of blue whiting is estimated to be higher in 2020 and 2021 than during the three previous years


## Graphical summary

|  | Topic | Overall trend | Situation in 2021 | Certainty | Possible implications |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ocean cli- mate | General warm and saline conditions prevailed from the early 2000s until 2015-2016. The recent 2017-2019 trend of colder and fresher Atlantic inflow into the Norwegian Sea has ceased. <br> However, the extent of Arctic Water is still increasing. | The recent 3-4 year trend of colder and fresher Atlantic Inflow into the Norwegian Sea has ceased. The extent of Arctic Water continues to increase. | Highly certain: dedicated monitoring with good spatial coverage exists. | The recent increase of Arctic Water may lead to increased new production due to relative high winter nutrient concentration. |
|  | Primary production | Annual primary production was on average $30 \%$ higher and length of spring bloom on average 17 days longer for the period 2013-2020 compared to 20032012. Start of spring bloom varied from April 25 to June 13 with no temporal trend. | Comparable to the 7 preceding years | Highly certain: the phytoplankton estimates are based on satellite data covering the whole productive season with high geographic resolution. | Increased primary production may have led to increased food resources for herbivores 2013-2020. |
| $\underbrace{}_{\substack{0 \\ 0}}$ | Zooplankton <br> biomass | The spring biomass of mesozooplankton was at a higher level from 1995 to mid-2000s and has been at a lower level afterwards. Summer biomass shows an increasing trend during the last 10 years, except for the last year(s). | Biomass in 2021 was at the same level or decreasing compared to the last years. Summer biomass showed the larger decrease. | Moderately certain: plankton is patchily distributed, which leads to uncertain estimates. | Reduced zooplankton biomass may have caused reduced food resources for planktivorous feeders, including pelagic fish in the recent decade. |
| T | Zooplankton spatial distribution | The spring distribution of zooplankton has changed from higher biomasses in Arctic water | In 2021 the zooplankton was evenly distributed both in spring and summer, but with some | Moderately certain: The spatial distribution reflects and is affected by the timing of the survey | Changes in the spatial distribution of plankton can affect the spatial distribution of planktivorous fish |


|  |  | in the west to become evenly distributed in the Norwegian Sea. | confined high-concentration areas. | and the timing of the zooplankton seasonal development. |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| ${ }_{66}$ | Pelagic fish biomass | The spawning biomass of Norwegian spring-spawning herring increased in the last year after a decade of decline. Spawning biomass of mackerel and blue whiting continue declining as in recent years. | Herring spawning biomass increased by $12 \%$ whereas mackerel spawning biomass declined by $11 \%$ and blue whiting by $17 \%$ compared to previous year. Fishing remains above scientific advice in all stocks. | Highly certain for herring and blue whiting, moderately certain for mackerel: estimates are based on quantitative stock assessments. | Changes in pelagic fish biomass have direct implications for fisheries opportunities. |
| O | Pelagic fish spatial distribution | In the mid-2000's mackerel distribution began expanding westward, into Icelandic and Greenlandic waters but has retracted since 2015 resulting in majority of the mackerel stock feeding in the Norwegian Sea. | No mackerel in Greenlandic waters and low levels in the south-eastern part of Icelandic waters in 2021, as observed in 2020. | Highly certain: based on ecosystem surveys in the Nordic Seas in spring (May) and summer (July) | Changes in pelagic fish spatial distribution have direct implications for fisheries opportunities. |



Figure 1.9.1. A subset of climate indicators for the Norwegian Sea: a) Relative heat content (RHC) and b) Relative Freshwater Content (RFC); Svinøy section Atlantic Water core c) temperature and d) salinity; e) Arctic Water amount in the Norwegian Sea, f) The North Atlantic Oscillation (NAO) winter index, and g) the Sub-polar Gyre (SPG) index (note that strong gyre is represented by negative values and weak gyre with positive values)

## Pelagic Fish

## Current status

Three fish stocks dominate the pelagic ecosystem of the Norwegian Sea: Norwegian springspawning herring (NSS, Clupea harengus), Northeast Atlantic (NEA) mackerel (Scomber scombrus), and blue whiting (Micromesistius poutassou). In 2021, estimated spawning stock biomass (SSB) was similar for all three stocks, ranging from 3.4 to 3.8 million tonnes. Combined SSB for all three stocks was 10.7 million tonnes (figure 1.9.2).

Combined catch of the three stocks was 3.2 million tonnes in 2020, of which approximately 1.5 million tonnes was blue whiting, 1 million tonnes was mackerel, and 0.7 million tonnes was herring. Current exploitation level, relative to biological reference points, show that fishing pressure on herring and blue whiting is above management plan targets and above maximum sustainable yield. Mackerel exploitation is within limits for maximum sustainable yield, however the upper boundary of the $95 \%$ confidence interval for fishing mortality is higher than maximum sustainable yield fishing mortality. Stock status, for all three stocks, is good since SSB is above all biological reference points related to the risk of impaired reproductive capacity. However, herring SSB is very close to biological reference limits, as the $95 \%$ SSB confidence limits include the reference limits.

Recent changes

The 2021 stock assessment results show an estimated $12 \%$ increase in herring SSB in 2021 compared to 2020, after a decade on continuous decline with an overall estimated decline of $52 \%$. Mackerel SSB continue declining in 2021 and has declined by an estimated $37 \%$ from peak stock size in 2014-2015. Blue whiting SSB also declined in 2021 compared to previous years and was estimated to be $43 \%$ lower than at the last peak size in 2017.


Figure 1.9.2. Estimated spawning stock biomass (lines) including 95\% confidence intervals (shaded areas) for Norwegian spring-spawning herring (red filled circles), mackerel (purple filled triangles) and blue whiting (blue filled rectangles) from 1980 to 2021.

Mackerel distribution in the Nordic Seas in summer 2021 was similar to observed distribution in summer 2020 and the western boundary of the distribution was limited to the east coast of Iceland. The distribution of blue whiting in 2021 was similar to the most recent years. The distribution area of herring in May was similar to the most recent period. The large 2016 year-class is now largely distributed throughout the geographical distribution range of the mature herring stock. In July, however, the herring had shifted farther east and north; particularly five-year-old herring was distributed north-easterly.
Possible reasons for recent changes
Herring SSB is dominated by recruitment of large year-classes at irregular intervals with many years of small year-classes in between (figure 1.9.3). After the large 2002- and 2004-year classes, the recruitment has been below average. Since 2018, surveys have indicated an incoming strong 2016 year-class. The magnitude will be known when the year class is fully recruited at around age seven (i.e., in 2023). Fishing above advised level has accelerated the stock decline during a period of low recruitment. Since 2013, when sharing arrangements in fisheries were no longer agreed upon, annual commercial catch has on average been $31 \%$ higher than the advised total allowable catch (TAC). The increase in SBB in 2021 is due to increase in maturity of the large 2016
year-class from $10 \%$ mature at age 4 in 2020 to $60 \%$ at age 5 in 2020, and a small upward revision of this year-class.


Figure 1.9.3. Estimated year-class size at recruitment for Norwegian spring-spawning herring (age 2; red filled circle) and blue whiting (age 1; blue filled triangle) from 1981 to 2021.

The 2021 assessment of the mackerel stock included an upward revision of SSB and a downward revision of fishing mortality which reduced the perception of stock decline. Changes in assessment perception of the stock is due to changes in relative weights of data sources in the assessment model. Estimates of mackerel recruitment at age 0 are highly uncertain and are thus not presented here. Mackerel year-class strength appears to be established when mackerel enter the fishery at age 2-3 years.

Since mackerel abundance peaked in 2015, the annual commercial catches have on average been $37 \%$ higher than the scientific advice. Fishing above advised TAC repeatedly over years contributes to the observed decline in spawning stock size.

Blue whiting's sharp decline in SSB since 2017 is caused by excessive fishing, with catches exceeding the advised TAC by $25 \%$ since 2017, in combination with low recruitment in 2017-2019. However, improved recruitment in 2020 and 2021 are estimated to be higher than the three previous years, and these recruits will mature and contribute to the SSB already in 2022.

The blue whiting fishery mostly targets ages 3-5 years. Hence the stock can sharply decline when several years of poor recruitment coincide with excessive fishing. The stock also has the capacity to recover quickly when recruitment is high as stock fluctuations in early 2000's and late 2010's show.

The reasons why mackerel has retracted from the western area from 2015 onwards remain poorly understood. During this period, estimated mackerel stock size has declined by approximately a
third, zooplankton abundance has remained within the range observed during period of mackerel presence, and the western area remains warm enough for mackerel presence ( $>8-9{ }^{\circ} \mathrm{C}$ ).

### 1.10 Future Research and Development Priorities

As part of the planning towards future benchmark assessments, the working group maintains, for each stock, a list of research and development priorities on topics including proposed research projects, improved sampling and data collection and development of stock assessment techniques. In addition to these individual stock issues, increased consideration should be given to integrated ecosystem assessments for the stocks within WGWIDE. A number of WGWIDE members are also participants in the work of the Working Group on Integrated Assessment for Norwegian Sea (WGINOR). Improving linkages with other regional Integrated Ecosystem Assessment groups within ICES would be beneficial and should be considered in future.

### 1.10.1 NEA Mackerel

In 2019, the ICES Workshop on a Research Roadmap for Mackerel (WKRRMAC, (ICES, 2019b)) met to discuss the research needs for the provision of advice for the management of NEA Mackerel. The workshop involved a diverse range of stakeholders including industry representatives, managers and scientists and identified a number of priorities (see report of WGWIDE 2019 (ICES, 2019c) for details).

In 2020, WGWIDE discussed and proposed the establishment of a workshop to review information on the stock structure of NEA Mackerel and subsequent implications for the current (component based) regional management measures (minimum landing size, area and seasonal closures). The current basis, whereby the stock is considered to consist of 3 separate components (North Sea, Western and Southern) derives from research conducted several decades ago. Since this time, there have been advances in several stock identification methods (e.g. genetics, simulation approaches). WGWIDE 2022 recommended the establishment of WKEVALMAC (A Workshop on the Evaluation of NEA Mackerel stock components and regional management measures) to review available information from appropriate methods to infer the stock structure of NEA Mackerel. WGWIDE 2022 also identified chairs and drafted terms of reference for this workshop and propose convening this workshop in 2023.

### 1.10.2 Blue Whiting

Numerous scientific studies have suggested that blue whiting in the North Atlantic consists of multiple stock units. The ICES Stock Identification Methods Working Group (SIMWG) reviewed this evidence in 2014 (ICES, 2014) and concluded that the perception of blue whiting in the NE Atlantic as a single-stock unit is not supported by the best available science. SIMWG further recommended that blue whiting be considered as two units. There is currently no information available that can be used as the basis for generating advice on the status of the individual stocks. However, there are some studies going on and more data being collected to allow clarify the stock definition for this species. In the future, the newly collected information on stock composition should be evaluated on the behalf of a benchmark of this stock.

### 1.10.3 NSS Herring

The Norwegian spawning ground survey was reintroduced in 2015 as part of the tuning series (fleet 1). However, changes were made to the survey compared to the older part of the series. At the 2016 assessment benchmark, the inclusion of the surveys from 2015 was accepted as an extension to the tuning series. It is now considered appropriate to investigate the splitting of this survey series, particularly since 2020 has provided the sixth estimate from the survey since it was reintroduced. and the time series is now long enough to do this exercise. An inter-benchmark exercise to explore this was proposed during WGWIDE 2020, but it was later decided to postpone such exploration for the next benchmark. Some exploratory work was presented in WGWIDE 2021.

Consider the inclusion of a new tuning series (IESSNS) in the assessment.
Consider the inclusion of a new tuning series (tagging data based on RFID) in the assessment.
Consider the inclusion of a new Norwegian recruitment index into the assessment.
Request and incorporate within the assessment information on the uncertainty in catches from all countries submitting catch data (currently only available from Norway).

The maturity ogive for NSSH is back-calculated but with a delay of 6 years, i.e. the 5 last years use one of two fixed maturity ogives scales (one for small cohort and the other for large cohort). The benchmark report has no objective criteria when to recognize a cohort as strong, and the current model is not optimal for medium-sized cohorts. This may result in deviation in SSB in intermediate year.

There is clear indication of a density dependent effect on maturity at age. A more proper estimate of the maturity for the last 5 years (and for the forecast) should be made using the estimated cohort strength directly, and this should be evaluated through a peer-review process.

The model XSAM is used for the assessment. The SAM model infrastructure now supports the XSAM model as an optional model. A switch from the currently used code to the SAM platform should be done in order to make the model more publicly available and to ensure further development of the infrastructure. The possibility to use the predicting the observation variance in SAM can then be used instead of including external variance from surveys.

### 1.10.4 Western Horse Mackerel

Considering the potential of mixing between Western and North Sea horse mackerel occurring in division 7d and 7e, improved insight into the origin of catches from that area will be a major benefit for improvement of the quality of future scientific advice and thus management of the North Sea and Western horse mackerel stocks. A project addressing stock structure and boundaries of horse mackerel was initiated by the Northern Pelagic Working Group in collaboration with University College Dublin and Wageningen Marine Research. In 2018, the results of the genetic analysis have been published (Farrell et al 2018) which concluded that the spawners of North Sea and Western horse mackerel can be genetically identified as two distinct stocks. However, at that stage it was not yet possible to separate the two stocks when they occur in mixed samples. Subsequently, a full genome sequencing on horse mackerel has been carried out (Fuentes-Pardo et al 2020), which confirmed the earlier results on separating western, North Sea and southern horse mackerel (see also text below on North Sea horse mackerel). In addition, this study concluded that it would also be possible to distinguish horse mackerel from different spawning populations in mixed samples.

The most recent results indicate that a further large-scale analysis of samples, with a greater temporal and spatial coverage, with the newly identified molecular markers was required to test and reassess the current stock delineations. This is currently underway and it expected that results will be available for presentation at WGWIDE 2023.

The 2020 study also concluded that further analysis on the mixing between the Western stock and the Southern stock in area 8c should be carried out: the fishery in the area targets mainly juveniles, would be therefore be very important to understand the impact of this fishery on each of the two stocks.

### 1.10.5 North Sea horse mackerel

Firstly, studies on stock identity and the degree of connection and migrations between the North Sea and the Western Stock are considered particularly relevant. On behalf of the Pelagic Advisory Council and the EAPO Northern Pelagic Working Group, a research project on genetic composition of horse mackerel stocks was initiated. Genetic samples have been taken over the whole distribution area of horse mackerel during the years 2015-2017. The full genome of horse mackerel was sequenced and results indicated that the western horse mackerel stock is clearly genetically different from the North Sea stock (Farrell and Carlsson, 2019; Fuentes-Pardo et al., 2020). Markers were identified that are be able to reveal the stock identity of individual horse mackerel caught in potential mixing areas. Horse mackerel samples from division 7 d and 7 e have been collected by the PFA on board of commercial vessels in the Autumn of 2020, while horse mackerel from division 4a have been collected during the NS-IBTS in Q3. With the genetic markers developed, the stock identity of the individual horse mackerel caught can be identified, which will shed light on mixing in the sampled areas during Q3. Additionally, the Institute of Marine Research in Norway sampled horse mackerel in coastal waters within 4a during all quarters in 2019. Preliminary results presented at WGWIDE 2021 showed that the genetic profile of individuals caught in all quarters matched well with the genetic profile of the Western HOM stock, with just one or two individuals matching better with North Sea HOM profile (Florian Berg, pers. comm.). More samples and research is needed to confirm these results.

Efforts are required to upload historic age and length data to the InterCatch database. The current stock assessment method is based on length data and, with only data from 2016 onwards currently available in InterCatch, it is impossible to compare the F/F msy proxy and the lengthbased indicators that the proxy is based on with information from earlier years. Furthermore, length data are only submitted by accessions to stock coordinators directly, and not through InterCatch. This makes the process of combining the data from different countries prone to error and lack transparency. Since 2020, national data submitters were requested to submit data both via the accessions as well as through InterCatch. A comparative analysis has to be carried out to evaluate the feasibility of using length data from InterCatch only in the future. Moreover, it was discovered that several hundred Dutch age readings coming from foreign vessels (mainly UK) have not been uploaded to InterCatch in the past. Efforts will be made to ensure this historic information will be uploaded in order to increase (the currently low) confidence in the estimates of catch-at-age. In 2021, it was the first time that Dutch age samples from 2020 were used in the raising procedure of UK and uploaded to InterCatch.

Future work on the exploitable biomass index will focus on including a spatial component when modelling the joint FR-CGFS and NS-IBTS survey index, and on the missing survey data in 2020. Additionally, application of the SPiCT model to the stock will be evaluated.

### 1.10.6 Boarfish

From 2017, this stock has been included on the list of stocks sampled under the data collection framework (DCMAP). This permitted sampling of commercial catch for both length and age. However, age reading is difficult and expertise is limited. An increase in the number of age readers would help develop a time-series of commercial catch-at-age which would in turn enable the development of an age-based assessment methodology. The current ALK is static and is based on a limited number of age readings.

Improvements in the survey data can be realized through a change in sampling protocol on groundfish surveys to ensure boarfish are measured to the 0.5 cm . The acoustic time-series should continue to be developed. The current survey does not contain the stock. The use of information from other acoustic surveys, for example, the Pélagiques GAScogne (PELGAS) survey should also be explored.

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# 2 Blue whiting (Micromesistius poutassou) in subareas 27.1-9, 12, and 14 (Northeast Atlantic) 


#### Abstract

Blue whiting (Micromesistius poutassou) is a small pelagic gadoid that is widely distributed in the eastern part of the North Atlantic. The highest concentrations are found along the edge of the continental shelf in areas west of the British Isles and on the Rockall Bank plateau, where it occurs in large schools at depths ranging between 300 and 600 metres, and is also present in almost all other management areas between the Barents Sea and the Strait of Gibraltar and west to the Irminger Sea. Blue whiting reaches maturity at $2-7$ years of age. Adults undertake long annual migrations between the feeding and spawning grounds. Most of the spawning takes place between March and April, along the shelf edge and banks west of the British Isles. Juveniles are abundant in many areas, with the main nursery area believed to be the Norwegian Sea. See the Stock Annex for further details on stock biology.

Russian catches for 2021 and preliminary data for 2022 were not reported to ICES for use by WGWIDE 2022, which changed the default data compilation of international catch data, and have probably increased the uncertainty of the assessment.


### 2.1 ICES advice in 2021

Fishing mortality ( F ) is estimated to be above Fmsy since 2014. Spawning-stock biomass (SSB) has been decreasing since 2018; however, it is estimated to remain above MSY $\mathrm{B}_{\text {trigger. }}$ Recruitment (R) from 2017 to 2019 is estimated to be low, followed by a slight increase. ICES advises that when the long-term management strategy, agreed by the European Union, the Faroe Islands, Iceland, and Norway, is applied, catches in 2022 should be no more than 752736 tonnes.

### 2.2 The fishery in 2021

Total catch of blue whiting by Russia in 2021 was not reported to ICES for consideration by WGWIDE 2022. Preliminary catch data for quarters 1 and 2 of 2021 (submitted by Russia to WGWIDE 2021) has therefore been used in compiling the total catch in 2021 ( 1.143 million tonnes, Table 2.3.1.1 and Section 2.3). The description of the 2021 fishery below does not include the Russian fishery.

As in previous years, the main fisheries on blue whiting were targeting spawning and postspawning fish (Figures 2.2 .1 and 2.2.2). Most of the catches (87.3\%) were taken in the first two quarters of the year and the largest part of this was taken along the slopes of the Western European shelf and around the Faroes. Smaller quantities were taken in the southern part of the Norwegian Sea, in the Norwegian Trench, in the Rockall Trough and along the coast of Spain and Portugal.

The fishery in the second half of the year was mainly east of the Faroes and in the central Norwegian Sea, with smaller amounts in the Norwegian Trench and along the coast of Portugal and Spain.

The multinational fleet targeting blue whiting in 2021 consisted of several types of vessels from 17 countries. The bulk of the catch is caught by large pelagic trawlers, some with capacity to process or freeze on board. The remainder is caught by RSW vessels.

### 2.3 Input to the assessment

At the Inter-Benchmark Protocol on Blue Whiting, IBPBLW (ICES, 2016a), it was decided to use preliminary within year, quarter 1 and quarter 2 , catch-at-age data in the assessment to get additional information to the within year IBWSS survey estimates. In recent years, $85-90 \%$ of the total annual catches of the age 3+ fish have been taken in the first half of the year, which makes it reasonable to estimate the total annual catch-at-age from reported first semester (Q1 \& Q2) data and expected total catches for the remainder of the year. The catch data sections in this report contain a comprehensive description of the 2021 data as reported to ICES and a brief description of the 2022 preliminary catch data. A section describing the procedure adopted to estimate the catch-at-age and the catch-at-weight corresponding to the missing data from Russia was also included (2.3.1.3)

### 2.3.1 Officially reported catch data

Official catches in 2021 were estimated as 1143450 tonnes based on data provided by WGWIDE members (Table 2.3.1.1). Data provided as catch by rectangle represented $86 \%$ of the total WG catch in 2021.

In 2021, the majority of catches were caught on the spawning grounds with largest contribution from ICES divisions 27.7.c, 27.7.k and 27.5.b, 27.6.a (Figure 2.3.1.1; Tables 2.3.1.2, 2.3.1.3), caught respectively in quarter 1 and quarter 2 (Figure 2.3.1.6). In the first two quarters, catches are taken over a broad area, with the highest catches in 27.6.a, 27.5.b, 27.7.c and 27.7.k, while later in the year catches are mainly taken further north in division 27.2.a and in the North Sea (27.4.a) (Figures 2.3.1.6 and 2.3.1.7 and Table 2.3.1.3). The spatial and temporal distribution of catches in 2021 are similar to previous years (Figures 2.3.1.2, 2.3.1.3, 2.3.1.4; Table 2.3.1.4 and Figure 1.10 .7 in Section 1). The majority of the blue whiting catch was caught by four nations - Norway, Faroe Islands, Iceland, and Russia, respectively (Figure 2.3.1.5).

Discards of blue whiting are small. Most of the blue whiting caught in directed fisheries are used for reduction to fish meal and fish oil. However, some discarding occurs in the fisheries for human consumption and as bycatch in fisheries targeting other species.

Reports on discarding from fisheries which catch blue whiting were available from the Netherlands for the years 2002-2007 and 2012-2014. A study carried out to examine discarding in the Dutch fleet found that blue whiting made a minor contribution to the total pelagic discards.

The blue whiting discards data provided by Portuguese vessels operating with bottom otter trawl within the Portuguese portions of ICES Division 27.9.a are available since 2004. The discards data are from two fisheries: the crustacean fishery and the demersal fishery. The blue whiting estimates of discards in the crustacean fishery for the period of 2004-2011 ranged between $23 \%$ and $40 \%$ (in weight). For the same period the frequency of occurrence in the demersal fishery was around zero for the most of the years, in the years where it was significant $(2004,2006,2010)$ discards ranged between $43 \%$ and $38 \%$ (in weight). In 2021, discards were $44 \%$ of the total catches for blue whiting along the Portuguese coast (Table 2.3.1.5). The total catch from Portugal is less than of one percent the total international catches.

Information on discards was available for Spanish fleets since 2006. Blue whiting is a bycatch in several bottom-trawl mixed fisheries. The estimates of discards in these mixed fisheries in 2006 ranged between $23 \%$ and $99 \%$ (in weight) as most of the catch is discarded and only the catch of the last day may be retained for marketing fresh. The catch rates of blue whiting in these fisheries are, however, low. In the directed fishery for blue whiting for human consumption with pair
trawls, discards were estimated to be 8\% (in weight) in 2021 (Table 2.3.1.5). Spanish catches are around $2 \%$ of the international catches.

In general, discards are assumed to be small in the blue whiting directed fishery. Discards data contributed to final catches of the following countries: Denmark, Ireland, Portugal, Spain, UK (England and Wales) and UK (Scotland). The total discards constituted $0.34 \%$ of the total catches, 3936 tonnes. The largest fishing nations, Norway, Faroe Islands, Russia and Iceland do not have discards on blue whiting.

The total estimated catches (tonnes) inside and outside the NEAFC regulatory area by country were reported on Table 2.3.1.6. The catches inside the NEAFC RA represent $16 \%$ of the total catches of blue whiting in 2021.

### 2.3.1.1 Sampling intensity

In $2021,81 \%$ of catches were covered by the sampling program. In 2021, 1676 length samples and 1588 age samples were collected from the fisheries with 129317 fish measured and 15215 aged. Sampling intensity for blue whiting with detailed information on catch, proportion of catch covered by the sampling program, the number of samples, number of fish measured, and number of fish aged per year from 2000 to 2021 is given in Table 2.3.1.1.1. Sampling intensity per country, quarter and ICES division for 2021 is listed in Tables 2.3.1.1.2, 2.3.1.1.3 and 2.3.1.1.4. The most intensive sampling, considering the age samples and the number of aged fish, took place in areas 27.2.a, 27.5.b, 27.6.b, 27.7.b, 27.7.c, 27.7.k, 27.8.c and 27.9.a. No sampling was carried out by Greenland, Lithuania, Poland and Sweden, which together represent $6 \%$ of the total catches. The sampled and estimated catch-at-age data are shown on Figure 2.3.1.1.1.

Sampling intensity for age and weight of blue whiting are made in proportion to landings according to CR 1639/2001 and apply to EU member states. The Fisheries Regulation 1639/2001, requires EU Member States to take a minimum of one sample for every 1000 tonnes landed in their country. Various national sampling programs are in force.

### 2.3.1.2 Age compositions

As an example of an age-length key from sampled catches in 2021, data from ICES area 27.6.a is presented by quarter and country (Figure 2.3.1.2.1). The mean length (mm) by age reveals that age classifications do present some differences between countries. A difference in mean length-at-age was observed in age 1 . Although, the differences in mean length-at-age increase in older ages, higher than age 7 .

The ICES InterCatch program was used to calculate the total international catch-at-age, and to document how it was done.

### 2.3.1.3 Missing data

ICES estimated missing data from Russia using the 2021 ICES preliminary catch statistics reported by the Russian governmental statistical office and the 2021 preliminary available catch-at-age and catch-at-weight data for quarters 1 and 2 submitted to WGWIDE in 2021. A comparison between the ICES preliminary catch statistics reported by the Russian governmental statistical office with the final data submitted to WGWIDE for the most recent years (2018 to 2020) revealed no differences between the two data sources. Also, the comparison between the submitted data to WGWIDE, i.e. between the preliminary available catch-at-age and catch-at-weight data for quarters 1 and 2 and the final data, was performed. From the comparison between the preliminary data with the finalized data, no differences were found for quarter 1, but for quarter 2 a difference in average of around $8 \%$ in total catch was found. For the period between 2018 and 2020, $89 \%$ of the Russian total catches were from quarters 1 and 2 . The allocation of the total catch by ICES area and quarter was based on the spatial and temporal pattern distribution observed
in the period 2018 until 2020. For ICES areas 27.6.b and 27.7, Russian catches were taken during quarter 1 and were included in the data submitted to WGWIDE in 2021. For the other ICES areas (27.2.a, 27.4.a, 27.5.b and 27.6.a), the approach for the ICES estimates by quarter was based on the average catch distribution from the period 2018 to 2020 . Russian data on age composition of the catch in 2021 for quarters 3 and 4 were not available, however, samples available from other fishing nations operating in the same areas were used to estimate catch and weight at age.

For the 2022 preliminary catch data, the approach to complete the preliminary ICES estimated catches was based on the assumption that the missing Russian data correspond to $13 \%$ of the 2022 ICES estimated total preliminary catch in weight for this stock. This assumed percentage was based on data analysis for the most recent 3 years (2019-2021).

### 2.3.2 Preliminary 2022 catch data (Quarters 1 and 2)

The preliminary catches for 2022 as reported by the WGWIDE members are presented in Table 2.3.2.1.

The spatial distribution of these 2022 preliminary catches is similar to the distribution in 2021 with majority of catches taken in division 27.6.a, 27.5.b, 27.7.c and 27.7.k (Figure 2.3.2.1 and Table 2.3.2.2).

Sampling intensity for blue whiting from the preliminary catches by area with detailed information on the number of samples, number of fish measured, and number of fish aged is presented in Table 2.3.2.2.

WGWIDE estimated the expected total catch for 2022 from the sum of declared national quotas, corrected for expected national uptake and transfer of these quotas (Table 2.3.2.3).

For the period 2016 to 2021, preliminary and final catch estimates are similar with maximum deviation in 2021 when the final catch was $8.3 \%$ lower than the preliminary catch (Table 2.3.2.4). Age compositions (Figure 2.3.2.2) are also similar between preliminary and final catch data with the exception of an increase in age 1 in the final data from 2021 compared to the preliminary data. There is no clear pattern in the deviations; it is both the catch at age for young and older fish that change between preliminary and final data.
The estimation of catch at age and mean weight at age followed the method described in the Stock Annex.

### 2.3.3 Catch-at-age

The catch in numbers-at-age from 1981 to 2022 are presented in Table 2.3.3.1 and catch proportions at age shown in Figure 2.3.3.1. Strong year classes that dominated the catches can be clearly seen in the early 1980s, 1990, the late 1990s and early 2000's. More recently, the propagation of the large 2014 year class is also evident. In 2021 there is also an indication of a stronger year class in the catch data.

Catch curves for the international catch-at-age dataset (Figure 2.3.3.2), indicate a consistent decline in catch number by cohort in years with rather high landings (and probably similar high effort). The catch curves for year classes 2010-2015 show a consistent decline in the stock numbers with an estimated total mortality $(\mathrm{Z}=\mathrm{F}+\mathrm{M})$ around $0.6-0.7$ for the ages fully recruited to the fisheries. With an assumed natural mortality $(\mathrm{M}=0.2)$, the assessment F around $0.4-0.5$ fits well to the $Z$ values estimated from the catch curves.

### 2.3.4 Weight at age

Table 2.3.4.1 and Figure 2.3.4.1 show the mean weight-at-age for the total catch during 1981-2022 used in the stock assessment. Mean weight at ages 3-9 has generally decreased in the period 20102018, followed by an increase in the most recent years, for the most abundant ages in the catches. In 2021 and 2022, a decrease in mean weight in almost all ages was observed.

The weight-at-age for the stock is assumed the same as the weight-at-age for the catch.

### 2.3.5 Maturity and natural mortality

Blue whiting natural mortality and proportion of maturation-at-age are shown in Table 2.3.5.1. See the Stock Annex for further details.

### 2.3.6 Information from the fishing industry

No new information available.

### 2.3.7 Fisheries independent data

Data from the International Blue Whiting spawning stock survey are used by the stock assessment model, while recruitment indices from several other surveys are used to qualitatively adjust the most recent recruitment estimate by the assessment model and to guide the recruitments used in the forecast.

### 2.3.7.1 International Blue Whiting spawning stock survey

The Stock Annex gives an overview of the surveys available for the blue whiting. The International Blue Whiting Spawning Stock Survey (IBWSS) is the only survey used as input to the assessment model.

The full time series of IBWSS was recalculated in summer 2020, using the same software (StoX; Johnsen et al., 2019) and method as previously applied. The values are presented in Table 2.3.7.1.1 and Figure 2.3.7.1.1 A.

The survey time-series (2004-2022) show variable internal consistency ranging from 0.26 to 0.84 (Figure 2.3.7.1.1 B) The overall internal consistency for age-disaggregated year classes was slightly reduced compared to last year. There is a high internal consistency for the younger ages (1-5 years) and older ages (7-9 years) with correlation between 0.68 and 0.84 , but poor ( $0.2<\mathrm{r}<$ 0.3 ) between ages 5 to 7 . This may indicate age readings problems for this group of ages.

The distribution of acoustic backscattering densities for blue whiting for the period 2019-2022 is shown in Figure 2.3.7.1.2. The abundance estimate of blue whiting for IBWSS are presented in Table 2.3.7.1.1.

Length and age distributions for the period 2018 to 2022 are given in Figure 2.3.7.1.3.
Survey indices, (ages 1-8 years 2004-2022) as applied in the stock assessment are shown in Table 2.3.7.1.1.

### 2.3.7.2 Other surveys

The Stock Annex provides information and time-series from surveys covering parts of the stock area. A brief survey description and survey results are provided below.

The International ecosystem survey in the Nordic Seas (IESNS) in May which is aimed at observing the pelagic ecosystem with particular focus on Norwegian spring-spawning herring and blue whiting (mainly immature fish) in the Norwegian Sea (Table 2.3.7.2.1).

Norwegian bottom-trawl survey in the Barents Sea (BS-NoRu-Q1(Btr)) in February-March where blue whiting are regularly caught as a bycatch species. This survey gives the first reliable indication of year class strength of blue whiting. The 1-group in this survey is defined as less than 19 cm (Table 2.3.7.2.2).

Icelandic bottom-trawl surveys on the shelf and slope area around Iceland. Blue whiting is caught as bycatch species and 1-group is defined as less than 22 cm in March (Table 2.3.7.2.3).

Faroese bottom-trawl survey on the Faroe plateau in spring where blue whiting is caught as bycatch species. The 1-group in this survey is defined as equal or less than 23 cm in March (Table 2.3.7.2.4).

The International Survey in Nordic Seas and adjacent waters in July-August (IESSNS). Blue whiting have been considered as a main target species in this survey since 2016 and as such methods were changed to ensure there was sampling for blue whiting. This was a recommendation from WGWIDE 2015 to try to have one more time-series for blue whiting. Data for the survey are not used yet, due to the short time series.

### 2.4 Stock assessment

The IBWSS survey is the only survey used by the SAM assessment. The survey was cancelled in 2020 due to the COVID-19 pandemic, but conducted in 2021 and 2022.

The presented assessment in this report follows the recommendations from the Inter-Benchmark Protocol of Blue (ICES, 2016a) to use the SAM model. The configuration of the SAM model was kept unchanged in this year's assessment.

At WGWIDE in 2021 the time period for estimating recruitment for the short term forecast was changed from the full time series (minus terminal year) to the more recent period since 1996 (minus terminal year). This approach was again followed by WGWIDE 2022.

### 2.4.1 2022 stock assessment

For a model such as SAM, Berg and Nielsen (2016) pointed out that the so-called "One Step Ahead" (OSA) residuals should be used for diagnostic purposes. The OSA residuals (Figure 2.4.1.1) show a quite random distribution of residuals. There may be an indication of a "year effect" (too low index values) for the IBWSS 2015 observations which has also be seen in previous assessment.

The estimated parameters from the SAM model from this year's assessment and those from assessments conducted since 2018 are shown in Table 2.4.1.1. There are no abrupt changes in the estimated parameters over the time-series presented. The lowest observation noises, and therefore the largest weight in the assessment model, have in all years been from catches at ages 3-8, which constitute the largest proportion of the catch.

The process error residuals ("Joint sample residuals") (Figure 2.4.1.2) are reasonably well randomly distributed. Process noise within SAM is implemented as a "process mortality, Z"; these deviations in mortalities are shown in Figure 2.4.1.3. The deviations in mortality (plus or minus mortality) seems fairly randomly distributed without very pronounced clusters as also seen in Figure 2.4.1.2).

The correlation matrix between ages for the catches and survey indices (Figure 2.4.1.4) shows a modest observation correlation for the younger ages and a stronger correlation for the older ages. This difference is more distinct for catches, probably because it includes older ages (1-10+) than the survey data (ages 1-8).

Figure 2.4.1.5 presents the exploitation pattern for the whole time-series. There are no abrupt changes in the exploitation pattern from 2010 to 2021, even though the landings in 2011 were just $19 \%$ of the landings in 2010, which might have given a change in exploitation pattern. The plateau in selection at age 6 and older seen since mid-2000s seems more realistic than the more linear selection estimated for the beginning of the time series. The estimated stable exploitation pattern might be influenced by the use of correlated random walks for $F$ at age with a high estimated correlation coefficient ( $\mathrm{Rho}=0.93$, Table 2.4.1.1).

The retrospective analysis (Figure 2.4.1.6) shows a reasonably stable assessment for the last 5 years, with the previous years within the $95 \%$ CI for the current assessment. Mohn's rho by year and as the average value over the last five years are presented in (Table 2.4.1.2). The annual values are rather high (and negative) for recruitment such that the average Mohn's rho for recruitment becomes -0.257 . Last year this value was -0.051 due to a large positive value in the first year (which not is used anymore) but also lower absolute values for the negative values in the remaining 4 years. The average Mohn's rho for F and SSB indicates no bias.

Stock summary results with added $95 \%$ confidence limits (Figure 2.4.1.7 and Table 2.4.1.5) show a decrease in fishing mortality in the period 2004-2011, followed by a steep increase in F up to 2015 after which F has decreased to around 0.35 (above Fmsy at 0.32 ). Recruitment (age 1) was high in 2015, followed by a lower recruitment in 2016 and much lower recruitments in 2017-2020. The recruitment in 2021 is estimated to be a historical high. SSB has increased since 2021 with a huge increase from 2022 to 2023 when $40 \%$ of the large 2021 recruitment is assumed to be mature.

A comparison of the assessments in 2021 and 2022 (Figure 2.4.1.8) shows a substantial revision of the historical values of F, SSB and recruitment for the most recent years of the assessment. The 2021 recruitment is now estimated to be at a historical high ( 71.6 billion) while last year's estimate for the same year class was 22.8 billion. $F$ for 2021 is now estimated at 0.36 while the same value in last year's assessment was 0.51 . Likewise SSB for 2022 is now estimated to 4.96 million tonnes while last year's value was 3.40 million tonnes.

The reasons for this revision is linked to 1) an historical high survey index for the age 2 in 2022 (the 2020 year class) corroborated by high commercial catch at age of the same year class in 2021 and 2022, and 2) the use of (uncertain) preliminary catch data for 2021 in the 2021 assessment.

With respect to point 1, while the IBWSS index for age 2 in 2022 is a historical high, the index for age 1 in 2021 was not especially high such that last year's estimate of year class strength was not especially high. Preliminary catch corroborate the high age 2 index with high age 2 catch numbers in 2022 ( $6^{\text {th }}$ highest in the time series back to 1981) and high age 1 catches in 2021 ( $8^{\text {th }}$ highest in the time series). Data for other surveys confirm the large 2020 year class (see section 2.3.7 for further discussion).

With respect to point 2, the final numbers at age in the catch are higher than the preliminary catch for age 1, while the final catch data are lower than the preliminary data for age 2-10+ (Figure 2.3.2.2). The final total catch weight is $8.3 \%$ lower than the preliminary values for 2021 (Table 2.3.2.4). Figure 2.4.1.9 shows the results from the default assessment configuration, a configuration without preliminary catches for 2022, and a configuration with preliminary catches for 2021 (last year's data not updated) and preliminary data for 2022. When the preliminary catch data for 2021 are maintained (without updating to "final" data), F becomes higher and SSB lower in the final year compared to the default run, as the total catch weigh for the 2021 preliminary catch is higher than the final. Recruitment in 2021 is however estimated lower when the 2021
preliminary data are applied as the catch at age number for age 1 is lower in the preliminary data set. The exclusion of the preliminary catch data provides a similar result for F and SSB as the default configuration. Recruitment in 2021 is also similar but recruitment for 2022 is estimated higher in the run without 2022 catches, as the historically high age 1 index from the IBWSS 2022 data is not corroborated by high catch numbers in the preliminary catch data. If the preliminary catch data for 2021 (applied in last year's assessment) had been a more accurate estimate of the final data, the revision of the historical F and SSB between the 2021 and 2022 assessments would have been smaller. This is seen for the retrospective analysis (Figure 2.4.1.6). F in 2021 was estimated to 0.43 when 2021 was the terminal year (with final 2021 catch at age) whereas the $F$ in 2021 in previous year's assessment was 0.51 , as it used the preliminary 2021 catch data.

If the preliminary 2022 catch are not used, estimates of SSB(2022) and F(2021) becomes very similar to the results from the default run compared with the default assessment (Figure 2.4.1.9). Recruitment in 2022 is however estimated considerably higher, as the survey index is at a record high for age 1 in 2022 IBWSSS and there are no additional catch data.

### 2.4.2 Alternative model runs

The working document WD08 "Blue whiting, an updated alternative assessment including more surveys" (Hølleland et al., 2022) describes an alternative assessment presented to the WGWIDE in 2021. The assessment is a SAM assessment, and makes use of two (IESNS and IESSNS) additional survey indices for blue whiting. The time series for IESSNS is still relatively short (7 years), while the IESNS has been running for 15 years. The alternative assessment gave similar results with a slightly lower SSB and higher F point estimate compared to the presently used SAM (Figure 2.4.2.1). The estimated recruitment in 2021 and 2022 was however larger in the alternative assessment, due to high abundance of age 1 in 2021 and 2022 in both additional surveys.

The WGWIDE assessment for 2021 estimated an F of 0.508 , while in the 2022 assessment there was a large correction for this year to 0.356 . This could be related to lack of information from the cancelled IBWSS in 2020 or an overestimate of the catches for 2021. The estimated F from the alternative assessment was quite consistent between the 2021 and 2022 assessments (see Figure 2.4.2.2).

### 2.5 Final assessment

Following the recommendations from Inter-Benchmark Protocol on Blue Whiting (ICES, 2016a) the SAM model is used for the final assessment. The model settings can be found in the Stock Annex.

Input data are catch numbers-at-age (Table 2.3.3.1), mean weight-at-age in the stock and in the catch (Table 2.3.4.1) and natural mortality and proportion mature in Table 2.3.5.1. Applied survey data are presented in Table 2.3.7.1.1.

The model was run for the period 1981-2022, with catch data up to 2021 and preliminary catch data for the first half-year (Q1 and Q2) of 2022 raised to expected annual catches, and survey data from March-April, 2004-2022. SSB 1st January in 2022 is estimated from survivors and estimated recruits (for 2022 estimated outside the model, see short-term forecast section). $11 \%$ of age group 1 is assumed mature, thus recruitment influences the size of SSB. The key results are presented in Tables 2.4.1.3-2.4.1.4 and summarized in Table 2.4.1.5 and Figure 2.4.1.7. Residuals of the model fit are shown in Figures 2.4.1.1 and 2.4.1.2.

### 2.6 State of the Stock

Fishing pressure (2022) on the stock is above $\mathrm{F}_{\mathrm{msy}}$ and between $\mathrm{F}_{\mathrm{pa}}$ and $\mathrm{F}_{\text {lim; }}$; spawning-stock size (2023) is above MSY $B_{\text {trigger, }} \mathrm{B}_{\mathrm{pa}}$ and $\mathrm{Blim}_{\mathrm{lim}}$.

F increased from a historic low at 0.052 in 2011 to around 0.50 in 2015 followed by a decrease in F to 0.37 in 2022. F has been above Fmsy and Fpa 0.32 since 2015. SSB has increased from 2020 ( 4.48 million tonnes) to an almost historical high in 2023 ( 6.66 million tonnes). SSB has been above MSY Btrigger since 1998.

Recruitment (age 1) in 2021 is estimated to be at a historical high. Survey data indicates that the 2022 recruitment is also above average, but this estimate has a high uncertainty.

### 2.7 Biological reference points

In spring of 2016, the Inter-Benchmark Protocol on Blue Whiting (IBPBLW) (ICES, 2016a) delegated the task of re-evaluating biological reference points of the stock to the ICES Workshop on Blue Whiting Long Term Management Strategy Evaluation (WKBWMSE) (ICES 2016b). During the WGWIDE meeting 2017, WKBWMSE concluded to keep $\mathrm{B}_{\lim }$ and $\mathrm{B}_{\mathrm{pa}}$ unchanged but revised $F_{\text {lim, }} \mathrm{F}_{\mathrm{pa}}$, and $\mathrm{F}_{\mathrm{ms}}$.

ICES made in 2021 the decision to use $\mathrm{F}_{\mathrm{p} 05}$ as the value for $\mathrm{F}_{\mathrm{pa}} . \mathrm{F}_{\mathrm{p} 05}$ was estimated by WKBWMSE (ICES 2016b), where it was concluded that the EQSIM simulations showed that $\mathrm{Fp}_{0.05}(0.32)$ is less than the FMSY in the constant $F$ simulations, so FMSY was set to this lower value.

The table below summarises the currently used reference points.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | $2.25 \text { mil- }$ <br> lion t | $\mathrm{B}_{\mathrm{pa}}$ | ICES (2013a, 2013b, 2016b) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.32 | Stochastic simulations with segmented regression stock-recruitment relationship | ICES (2016b) |
| Precautionary approach | $\mathrm{Blim}^{\text {im }}$ | $1.50 \text { mil- }$ <br> lion $t$ | Approximately $\mathrm{B}_{\text {loss }}$ | ICES (2013a, 2013b, 2016b) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | $2.25 \text { mil- }$ <br> lion t | $\mathrm{B}_{\text {lim }} \exp (1.645 \times \sigma)$, with $\sigma=0.246$ | ICES (2013a, 2013b, 2016b) |
|  | $F_{\text {lim }}$ | 0.88 | Equilibrium scenarios with stochastic recruitment: F value corresponding to $50 \%$ probability of (SSB< $\mathrm{B}_{\text {lim }}$ ) | ICES (2016b) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.32 | Fp05; the F that leads to SSB $\geq$ Blim with 95\% probability | ICES (2016b) and WGWIDE 2021 |

### 2.8 Short-term forecast

### 2.8.1 Recruitment estimates

The benchmark WKPELA in February 2012 concluded that the available survey indices should be used in a qualitative way to estimate recruitment, rather than using them in a strict
quantitative model framework. The WGWIDE has followed this recommendation and investigated several survey time-series indices with the potential to give quantitative or semi-quantitative information of blue whiting recruitment. The investigated survey series were standardized by dividing with their mean and are shown in Figure 2.8.1.1.

The International Ecosystem Survey in the Nordic Seas (IESNS) only partially covers the known distribution of recruitment from this stock. The 1-group (2021 year class) and the 2 -group ( 2020 year class) indices from the survey in 2022 were both well above the median of the historical range.

The 1-group (2021 year class) and the 2-group (2020 year class) indices from The International Blue Whiting Spawning Stock Survey (IBWSS) were both the highest in the time series (Table 2.3.7.1.1).

The Norwegian bottom-trawl survey in the Barents Sea (BS-NoRu-Q1(Btr)) in February-March 2022, showed that 1-group blue whiting was above the median in the time series (Table 2.3.7.2.2). This index should be used as a presence/absence index, in the way that when blue whiting is present in the Barents Sea, this is usually a sign of a strong year class, as all known strong year classes have been strong also in the Barents Sea.

The 1-group estimate in 2022 (2021 year class) from the Icelandic bottom-trawl survey showed an increase compared to 2021 and was the highest in the time-series.

The 1-group estimate in 2022 (2021 year class) from the Faroese Plateau spring bottom-trawl survey showed a decrease compared to 2021 and was below the median in the time-series. This is the only survey which doesn't pick up a strong signal from the 2020 and 2021 year classes.

In conclusion, the indices from available survey time-series indicate that the 2020 year class is among the strongest in the time series, which corresponds to the SAM assessment results. The 2021 year class estimated from surveys are also above the median, also seen in the SAM assessment. It was therefore decided not to change the SAM estimates of the 2020 and 2021 year classes for the purposes of the short term forecast.

No information is available for the 2022 and 2023 year classes and the geometric mean of the time-series from 1996-2021) was used for these year classes (22.5 billion at age 1 in 2023) (Table 2.8.1.1).

As described in the Stock Annex, WGWIDE decided in 2021 to change from using the geometric mean of the full time-series (since 1981) to use a shorter time-series (since 1996) for the calculations recruitment.

### 2.8.2 Short-term forecast

As decided at WGWIDE 2014, a deterministic version of the SAM forecast was applied. Details about specific implementation can be found in the Stock Annex.

### 2.8.2.1 Input

Table 2.8.2.1.1 lists the input data for the short-term predictions. Mean weight at age in the stock and mean weight in the catch are the same, and are calculated as three year averages (20202022) in accordance with the 2019 updated Stock Annex. Selection (exploitation pattern) is based on F in the most recent year. The proportion mature for this stock is assumed constant over the years and values are as used by the assessment.

Recruitment (age 1) in 2021 and 2022 are assumed as estimated by the SAM model, as additional survey information was not conflicting this result. Recruitment in 2023 and 2024 are assumed as
the long-term average from the period with both high and low recruitments (geometric mean of the time-series since 1996, minus the terminal year, 1996-2021).

As the assessment uses preliminary catches for 2022 an estimate of stock size is available for the $1^{\text {st }}$ of January 2023. The normal use of an "intermediate year" calculation is not relevant in this case and F in the "intermediate year" (2022) is as calculated by the assessment model. Catches in 2022 are based on the preliminary catches based on declared national quotas and expected national uptake for 2022. Intermediate year assumptions are summarised in Table 2.8.2.1.2.

### 2.8.2.2 Output

A range of predicted catch and SSB options from the deterministic short-term forecast used for advice are presented in Table 2.8.2.2.1.

Following the ICES MSY framework for the target F from the LTMS implies fishing mortality to be at $\mathrm{F}_{\mathrm{msy}}=0.32$ which will give a TAC in 2022 at 1359629 tonnes. This corresponds to a $80.6 \%$ increase compared to the ICES advice last year, and a $22.8 \%$ increase compared to the preliminary estimate of catches in 2022.

The LTMS specifies a default TAC constraint at $+25 /-20 \%$. However, it states that the TAC constraint shall not be applied when the TAC advice deviates more than $40 \%$ from the TAC of the preceding year (paragraph 6b in the LTMS). With an increase of $80.6 \%$ in catches in relation to the ICES advice last year (LTMS advice), the TAC constraint is not applied.

SSB in 2024 is predicted to increase by $17.5 \%$ to 7781444 tonnes, if the advised catches are taken. The high recruitment estimated for 2021 and 2022 contributes to this increase in SSB.

### 2.9 Comparison with previous assessment and forecast

Comparison of the assessment made in 2021 and 2022 (Figure 2.4.1.8) shows a substantial revision of the historical values of $\mathrm{F}, \mathrm{SSB}$ and recruitments. The 2021 recruitment is now estimated to be historical high ( 71.6 billion) while last year's estimate was 22.8 billion. F for 2021 is now estimated to 0.36 while the same value in last year's assessment was 0.51 . Likewise SSB for 2022 is now estimated to 4.96 million tonnes while last year's value was 3.40 million tonnes. See section 2.4.1 for further discussion.

### 2.10 Quality considerations

Based on the confidence interval produced by the assessment model SAM there is a moderate to high uncertainty of the absolute estimate of F and SSB and the recruiting year classes (Figure 2.4.1.7). The retrospective analysis (Figure 2.4.1.6) shows a tendency to underestimate recruitment, but unbiased estimates of F and SSB. An alternative run (Figure 2.4.2.1) with the SAM model using two additional surveys (IESNS and IESSNS) not covering the full distribution area shows results consistent with the default configuration of the assessment.

There are several sources of uncertainty: age reading, stock identity, survey indices and the use of preliminary catch data. As there is only one survey (IBWSS) that covers the spawning stock, the quality of the survey influences the assessment result considerably. The Inter-Benchmark Protocol on Blue Whiting (IBPBLW 2016) introduced a configuration of the SAM model that includes the use of estimated correlation for catch and survey observations. This handles the "year effects" in the survey observation in a better way than assuming an uncorrelated variance structure as usually applied in assessment models. However, biased survey indices will still give a biased stock estimate with the new SAM configuration. The estimated correlation for catch at age observations might correspond to the age reading discrepancy as also estimated from inter-
calibration exercise. The use of additional survey data may be beneficial, especially in years without IBWSSS data, however the length of the time series is still short (7 years) for the survey (IESSNS) with a low observation variance for age 1 and 2 .

Utilization of preliminary catch data provides the assessment with information for the most recent year in addition to the survey information. This should give a less biased assessment, as potentially biased survey data in the final year are supplemented by additional catch data. The preliminary catch weight was however $9 \%$ higher than the final data for 2021, although the differences are smaller for the year 2016-2020.

### 2.11 Management considerations

The assessment this year estimates a lower $\mathrm{F}(2021)$, a higher $\operatorname{SSB}(2022)$ and a much larger 2020 year class size than estimated last year. The 2020 year class will be fully recruited to the fishery in 2023 and contribute considerably to the SSB ( $82 \%$ mature at age 3 ). SSB in 2023 is estimated to be well above MSY $B_{\text {trigger, }}$ but F in 2021 and 2022 remains above $\mathrm{F}_{\text {msy. }}$

### 2.12 Ecosystem considerations

Blue whiting is one of the most abundant pelagic and mesopelagic fish stocks in the Northeast Atlantic, SSB estimated from 1.4-6.9 million tonnes during the period from 1981 to 2020 (ICES, 2020). The stock is widely distributed and highly migratory. It's distribution range is approximately from latitude $30^{\circ} \mathrm{N}$ to $80^{\circ} \mathrm{N}$ and from the coast of Europe to Greenland, into Barents Sea and the Mediterranean Sea (Trenkel et al., 2014). Spawning is in the spring and mostly occurs on the shelf and banks west of Ireland and Scotland and major summer feeding area is in the Norwegian Sea. Blue whiting is most frequently observed at $100-600 \mathrm{~m}$ depth (Heino and Godo, 2002). Their most important prey are euphausiids, amphipods and copepods (Pinnegar et al., 2015, Bachiller et al., 2016) and they are prey for piscivorous fish (Dolgov et al., 2010) and cetaceans (Hátún et al., 2009a). Blue whiting is an important species in the NE Atlantic and it's best documented ecosystem interactions are listed below:
(a) Stock productivity - recruitment: blue whiting population dynamic is driven by large annual variability in recruitment (at age 1 in the assessment model) which is not linked to spawning stock size (ICES, 2020). Changes in recruitment have been correlated to changes in the North Atlantic subpolar gyre between strong and weak states (Hátún et al., 2009a,b). Two hypotheses have been suggested to explain a causal relationship between low gyre index and high recruitment (Payne et al., 2012). One suggests changes in marine climate where weak gyre results in increased flow of warm subtropical waters and increased abundance of important prey for juvenile blue whiting on their nursing grounds west of Ireland and Scotland. The other suggests increasing predation of mackerel on blue whiting larvae during years of weak index, but neither has been proven right (Payne et al., 2012).
(b) Changes in distribution: blue whiting spawning distribution varies between years. It has been linked to the North Atlantic subpolar gyre as a strong gyre (cold and fresh water masses on the Rockall Plateau) shrinks the spawning area compared to a weak gyre (increasing saline and warm waters at Rockall) which expands the spawning area northward and westward into Rockall Plateau (Hátún et al., 2009a,b; Miesner and Payne, 2018). Salinity appears specifically to impact spawning location of blue whiting (Miesner and Payne, 2018).
(c) It is still disputed whether there are one or two blue whiting populations in the Northeast Atlantic (Keating et al., 2014; Pointin and Payne, 2014; ICES, 2016c; Mahé et al., 2016). Currently blue whiting is considered a single population for management purpose.
(d) Trophic interactions in the Norwegian Sea: there appears to be limited prey competition between blue whiting and the two other abundant pelagic species, Norwegian spring-spawning herring and Atlantic mackerel, as studies show limited dietary overlap between blue whiting and the two other species (Bachiller et al., 2016; Pinnegar et al., 2015). Limited prey competition between blue whiting and mackerel can be explained by limited vertical spatial overlap, mackerel mostly feed in the surface layer and blue whiting deeper in the water column (Utne et al., 2012). Where distribution of blue whiting and herring overlap (Utne et al., 2012) they appear to feed on different species, herring mainly feed on copepods and blue whiting mainly on euphausiids and amphipods, although juvenile blue whiting feed on copepods (Bachiller et al., 2016; Pinnegar et al., 2015).

An extensive overview of ecosystem considerations relevant for blue whiting can be found in the Stock Annex.

### 2.13 Regulations and their effects

There is a long-term management strategy agreed by the European Union, the Faroe Islands, Iceland and Norway. However there is no agreement between the Coastal States, i.e. EU, Norway, Iceland and the Faroe Island on the share of the blue whiting TAC. The catch advice does not take into account consistent deviations from the long-term management strategy as evident from the sum of unilateral quotas since 2018. During the evaluation of the management strategy (ICES, 2016b), the implementation error in the form of a consistent overshoot of the TAC was not included. Therefore, the current implementation of the long-term management strategy may no longer be precautionary. See section 1.8 for a comparison of historic advice, TAC and catch.

WGWIDE estimates the total expected catch for 2022 to be 1107529 tonnes, whereas ICES advised that when the long-term management strategy agreed by the European Union, the Faroe Islands, Iceland, and Norway is applied, catches in 2021 should be no more than 752736 tonnes. This advice was followed by the Coastal States by setting a TAC at the ICES advice, however there was no agreement on the split of TAC between nations. The sum of unilateral quotas for 2022 exceeds the agreed TAC.

### 2.13.1 Management plans and evaluations

A response to a NEAFC request to ICES to evaluate a long-term management strategy for the fisheries on the blue whiting ICES WKBWMSE was established in the fall of 2015. The ICES Advice September 2016, "NEAFC request to ICES to evaluate a long-term management strategy for the fisheries on the blue whiting (Micromesistius poutassou) stock" concluded:

- $\quad$ That the harvest control rule (HCR) proposed for the Long-Term Management Strategy (LTMS) for blue whiting, as described in the request, is precautionary given the ICES estimates of $B_{\lim }(1.5$ million $t), B_{p a}\left(2.25\right.$ million $t$ ), and $F_{\text {mSY }}(0.32)$.
- The HCR was found to be precautionary both with and without the $20 \%$ TAC change limits above $B_{\text {pa. }}$. However, the $20 \%$ TAC change limits can lead to the TAC being lowered significantly if the stock is estimated to be below $\mathrm{B}_{\mathrm{pa}}$, while also limiting how quickly the TAC can increase once the stock is estimated to have recovered above $\mathrm{B}_{\mathrm{pa}}$.
- The evaluation found that including a $10 \%$ interannual quota flexibility ('banking and borrowing') in the LTMS had an insignificant effect on the performance of the HCR.

The management strategy evaluation did not take into account consistent deviations from the long-term management strategy as evident from the sum of unilateral quotas in recent years. During the evaluation of the management strategy (ICES, 2016b), the implementation error in
the form of a consistent overshoot of the TAC was not included. Therefore, the current implementation of the long-term management strategy may no longer be precautionary.

The Agreed Records by the Coastal States (25-26 October 2021) states a TAC for blue whiting at 752736 tonnes for 2022, as advised by ICES (on the basis of the LTMS from 2016). Annex 1 of the Agreed Records "Arrangement for the long-term management of the blue-whiting stock" is similar to the managing plan evaluated by ICES in 2016, but the present version of the LTMS includes a paragraph 6: The TAC constraint described in Paragraph 5 shall not apply if:" and a paragraph 6b: "The rules in paragraph 4 [TAC from $\mathrm{F}=\mathrm{FmSY}$, when SSB is above $\mathrm{B}_{\text {trigger] }}$ would lead to a TAC that deviates by more than $40 \%$ from the TAC of the preceding year.". The management plan evaluated by ICES in 2015-2016, described in the WKBWMSE (ICES 2016b) report, did not include the deviation from the default $-20 \% / 25 \%$ TAC constraint as described in paragraph 6 of the presently used LTMS. Therefore, ICES has not evaluated the presently used plan.

### 2.14 Recommendations

No recommendations.

### 2.15 Deviations from stock annex caused by missing information from Covid-19 disruption.

The one and only survey used for the SAM assessment, the International Blue Whiting Spawning Stock Survey (IBWSS) was not conducted in 2020, but resumed in 2021 and 2022. The stock assessment this year followed the approach outlined in the Stock Annex.

### 2.16 References

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### 2.17 Tables

Table 2.3.1.1. Blue whiting. ICES estimated catches (tonnes) by country for the period 1988-2021.

| Country | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 18941 | 26630 | 27052 | 15538 | 34356 | 41053 | 20456 | 12439 | 52101 | 26270 | 61523 | 82935 |
| Estonia |  |  |  |  | 6156 | 1033 | 4342 | 7754 | 10982 | 5678 | 6320 |  |
| Faroe Islands | 79831 | 75083 | 48686 | 10563 | 13436 | 16506 | 24342 | 26009 | 24671 | 28546 | 71218 | 329895 |
| France |  | 2191 |  |  |  | 1195 |  | 720 | 6442 | 12446 | 7984 | 14149 |
| Germany | 5546 | 5417 | 1699 | 349 | 1332 | 100 | 2 | 6313 | 6876 | 4724 | 17969 | 22803 |
| Iceland |  | 4977 |  |  |  |  |  | 369 | 302 | 10464 | 68681 | 501493 |
| Ireland | 4646 | 2014 |  |  | 781 |  | 3 | 222 | 1709 | 25785 | 45635 | 22580 |
| Japan |  |  |  |  | 918 | 1742 | 2574 |  |  |  |  |  |
| Latvia |  |  |  |  | 10742 | 10626 | 2582 |  |  |  |  |  |
| Lithuania |  |  |  |  |  | 2046 |  |  |  |  |  |  |
| Netherlands | 800 | 2078 | 7750 | 17369 | 11036 | 18482 | 21076 | 26775 | 17669 | 24469 | 27957 | 48303 |
| Norway | 233314 | 301342 | 310938 | 137610 | 181622 | 211489 | 229643 | 339837 | 394950 | 347311 | 560568 | 834540 |
| Poland | 10 |  |  |  |  |  |  |  |  |  |  |  |
| Portugal | 5979 | 3557 | 2864 | 2813 | 4928 | 1236 | 1350 | 2285 | 3561 | 2439 | 1900 | 2651 |
| Spain | 24847 | 30108 | 29490 | 29180 | 23794 | 31020 | 28118 | 25379 | 21538 | 27683 | 27490 | 13825 |
| Sweden ** | 1229 | 3062 | 1503 | 1000 | 2058 | 2867 | 3675 | 13000 | 4000 | 4568 | 9299 | 65532 |
| UK (England + Wales)*** |  |  |  |  |  |  |  |  |  |  |  |  |
| UK (Northern Ireland) |  |  |  |  |  |  |  |  |  |  |  |  |
| UK (Scotland) | 5183 | 8056 | 6019 | 3876 | 6867 | 2284 | 4470 | 10583 | 14326 | 33398 | 92383 | 27382 |
| USSR / Russia * | 177521 | 162932 | 125609 | 151226 | 177000 | 139000 | 116781 | 107220 | 86855 | 118656 | 130042 | 355319 |
| Greenland** |  |  |  |  |  |  |  |  |  |  |  |  |
| Unallocated |  |  |  |  |  |  |  |  |  |  |  |  |
| TOTAL | 557847 | 627447 | 561610 | 369524 | 475026 | 480679 | 459414 | 578905 | 645982 | 672437 | 1128969 | 2321406 |

* From 1992 only Russia.
** Estimates from Sweden and Greenland: are not included in the Catch at Age Number.
*** From 2012.

Table 2.3.1.1. (continued). Blue whiting. ICES estimated catches (tonnes) by country for the period 1988-2021.

| Country | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 20102 | 20112 | 20122 | 2013 | 20142 | 2015 | 20162 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 89500 | 41450 | 54663 | 48659 | 18134 | 248 | 140 | 165 | 340 | 2167 | 35256 | 45178 | 39395 | 60868 | 87348 | 68716 | 58997 | 40321 |
| Estonia | * |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 |  |  |
| Faroe Islands | 322322 | 266799 | 321013 | 317859 | 225003 | 58354 |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  | 49979 | 16405 | 43290 | 85768 | 224700 | 282502 | 282416 | 356501 | 349838 | 336569 | 343372 | 202415 |
| France |  | 8046 | 18009 | 16638 | 11723 | 8831 | 7839 | 4337 | 9799 | 8978 | 10410 | 9659 | 10345 | 13369 | 16784 | 16095 | 13769 | 14612 |
| Germany | 15293 | 22823 | 36437 | 34404 | 25259 | 5044 | 9108 | 278 | 6239 | 11418 | 24487 | 24107 | 20025 | 45555 | 47708 | 38244 | 42362 | 35327 |
| Iceland | 379643 | 265516 | 309508 | 236538 | 159307 | 120202 | 87942 | 5887 | 63056 | 104918 | 182879 | 214870 | 186914 | 228934 | 292944 | 268356 | 243725 | 190146 |
| Ireland | 75393 | 73488 | 54910 | 31132 | 22852 | 8776 | 8324 | 1195 | 7557 | 13205 | 21466 | 24785 | 27657 | 43238 | 49903 | 38836 | 40135 | 39514 |
| Lithuania |  |  | 4635 | 9812 | 5338 |  |  |  |  |  | 4717 |  | 1129 | 5300 |  |  | 9543 | 21183 |
| Netherlands | 95311 | 147783 | 102711 | 79875 | 78684 | 35686 | 33762 | 4595 | 26526 | 51635 | 38524 | 56397 | 58148 | 81156 | 121864 | 75020 | 62309 | 62017 |
| Norway | 957684 | 738490 | 642451 | 539587 | 418289 | 225995 | 194317 | 20539 | 118832 | 196246 | 399520 | 489439 | 310412 | 399363 | 438426 | 351429 | 354033 | 233968 |
| Poland |  |  |  |  |  |  |  |  |  |  |  |  |  | 15889 | 12152 | 27185 | 47616 | 26077 |
| Portugal | 3937 | 5190 | 5323 | 3897 | 4220 | 2043 | 1482 | 603 | 1955 | 2056 | 2150 | 2547 | 2586 | 2046 | 2497 | 3481 | 2819 | 2522 |
| Spain | 15612 | 17643 | 15173 | 13557 | 14342 | 20637 | 12891 | 2416 | 6726 | 15274 | 32065 | 29206 | 31952 | 28920 | 24718 | 22782 | 23676 | 25509 |
| Sweden | 19083 | 2960 | 101 | 464 | 4 | 3 | 50 | 1 | 4 | 199 | 2 | 32 | 42 | 90 | $16^{* *}$ | 54 | 25 | 40 |
| UK (England + Wales) | 2593 | 7356 | 10035 | 12926 | 14147 | 6176 | 2475 | 27 | 1590 | 4100 | 11 | 131 | 1374+ | 3447 | 1864 | 4062 | 7458 | 8783 |
| UK (Northern Ireland) |  |  |  |  |  |  |  |  |  | 1232 | 2205 | 1119 |  |  | 4508 | 2899 | 2958 |  |
|  |  |  |  |  |  | 173 |  |  |  |  |  |  |  |  |  |  |  |  |
| UK (Scotland) | 57028 | 104539 | 72106 | 43540 | 38150 |  | 5496 | 1331 | 6305 | 8166 | 24630 | 30508 | 37173 | 64724 | 66682 | 54040 | 41344 | 65085 |
| Russia | 346762 | 332226 | 329100 | 236369 | 225163 | 149650 | 112553 | 45841 | 88303 | 120674 | 152256 | 185763 | 173655 | 188449 | 170892 | 188006 | 181496 | 133605^ |
| Greenland |  |  |  |  |  |  |  |  |  | 2133 |  |  |  | 20212 | 23333 | 19753 | 19611 | 20190 |
| Unallocated |  |  |  |  |  |  |  |  | 3499 |  |  |  |  |  |  |  |  | 22137 |
| TOTAL | 2380161 | 2034309 | \|1976176| | 1625255\| | 1260615 | [ 641818 \| | 526357 | 103620 | 384021 \| | 628169 | 1155279 | 1396244 | \| 1181850 | 1558061 | 1711461 | 1515527 | 1495248 | 1143450 |
| * Reported to the EU but not to the ICES WGNPBW. (Landings of 19,467 tonnes). |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ** only landings (2018). |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| + data updated in 2018. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ^ Russia 2021 preliminary data (Q1+Q2) submitted to WGWIDE 2021. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 2.3.1.2. Blue whiting. ICES estimated catches (tonnes) by country and ICES division for 2021.

| ICES Division | Denmark | Faroe Islands | France | Germany | Greenland | Iceland | Ireland | Lithuania | Netherlands | Norway | Poland | Portugal | Russia ^ | Spain | Sweden | UK (England) | UK(Scotland) | Unallocated | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 27.4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 100 |  | 100 |
| 27.2.a | 20 | 27550 | 296 | 21 | 2933 | 26450 |  |  | 7.62 | 7918 | 121 |  | 7214 |  | 6 |  |  | 19071 | 91606 |
| 27.3.a | 98 |  |  |  |  |  |  |  |  |  |  |  |  |  | 8 |  |  |  | 106 |
| 27.4.a | 63 | 903 | 467 | 2095 | 344 | 6861 |  | 170 | 457.01 | 26578 | 243 |  | 34 |  | 25 |  | 22 | 1029 | 39290 |
| 27.4.b | 4 |  |  |  |  |  |  |  |  | 8 |  |  |  |  | 1 |  |  |  | 13 |
| 27.5.a |  | 1192 |  |  |  | 19284 |  |  |  |  |  |  |  |  |  |  |  |  | 20476 |
| 27.5.b | 98 | 127617 | 53 | 636 | 12750 | 101488 |  | 127 | 132.25 | 973 | 4053 |  | 48792 |  |  |  |  | 1920 | 298640 |
| 27.6.a | 14476 | 19565 | 7296 | 23309 | 4163 | 32826 | 18577.45 | 15865 | 30921.13 | 52402 | 14975 |  | 24239 | 3 |  | 6251 | 41400 | 117 | 306385 |
| 27.6.b |  | 4805 |  |  |  | 206 |  |  |  | 1176 |  |  | 2568 | 20 |  |  | 29 |  | 8804 |
| 27.7.b | 8 |  | 455 | 483 |  |  | 2102.687 |  | 5693.73 | 1092 |  |  |  | 32 |  | 58 | 2807 |  | 12732 |
| 27.7.c | 16450 | 2940 | 3748 | 8770 |  | 1533 | 14904.99 | 5021 | 10909.26 | 64982 |  |  | 22809 | 137 |  | 2276 | 20728 |  | 175208 |
| 27.7.e |  |  | 62 |  |  |  |  |  | 0.05 |  |  |  |  |  |  | 1 |  |  | 64 |
| 27.7.f |  |  |  | 1 |  |  |  |  |  |  |  |  |  |  |  | 19 |  |  | 20 |
| 27.7.g |  |  |  | 0 |  |  | 554.4149 |  |  |  |  |  |  | 14 |  | 1 |  |  | 569 |
| 27.7.h | 2 |  | 31 |  |  |  |  |  | 260.37 |  |  |  |  | 40 |  | 4 |  |  | 336 |
| 27.7.j | 2 |  | 2 | 13 |  |  | 109.572 |  | 557.52 |  |  |  |  | 368 |  | 174 |  |  | 1225 |
| 27.7.k | 9098 | 17843 | 27 |  |  | 1498 | 3264.397 |  | 9731.67 | 78841 | 6686 |  | 27949 | 0 |  |  |  |  | 154937 |
| 27.8.a | 3 |  | 889 |  |  |  |  |  | 2669.33 |  |  |  |  | 9 |  |  |  |  | 3570 |
| 27.8.b |  |  | 3 |  |  |  |  |  |  |  |  |  |  | 160 |  |  |  |  | 164 |
| 27.8.c |  |  | 0 |  |  |  |  |  |  |  |  | 197 |  | 16563 |  |  |  |  | 16760 |
| 27.8.d |  |  | 1282 |  |  |  |  |  | 676.94 |  |  |  |  |  |  |  |  |  | 1959 |
| 27.9.a |  |  |  |  |  |  |  |  |  |  |  | 2325 |  | 8162 |  |  |  |  | 10487 |
| Total | 40321 | 202415 | 14612 | 35327 | 20190 | 190146 | 39513.51 | 21183 | 62016.88 | 233968 | 26077 | 2522 | 133605 | 25509 | 40 | 8783 | 65085 | 22137 | 1143450 |

^ Russia 2021 preliminary data (Q1+Q2) submitted to WGWIDE 2021.

Table 2.3.1.3. Blue whiting. ICES estimated catches (tonnes) by quarter and ICES division for 2021.

| ICES <br> Division | Quarter 1 | Quarter 2 | Quarter 3 | Quarter 4 | 2021* | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 27.4 |  |  |  |  | 100 | 100 |
| 27.2.a | 1094 | 56453 | 14557 | 19502 |  | 91606 |
| 27.3.a | 0 | 11 | 93 | 2 |  | 106 |
| 27.4.a | 1084 | 15793 | 8037 | 14376 |  | 39290 |
| 27.4.b | 0 | 7 | 1 | 4 |  | 13 |
| 27.5.a | 1 | 262 | 5 | 20208 |  | 20476 |
| 27.5.b | 52819 | 227524 | 16 | 18281 |  | 298640 |
| 27.6.a | 86866 | 188481 | 2 | 31003 | 33 | 306385 |
| 27.6.b | 8756 | 19 | 0 | 0 | 29 | 8804 |
| 27.7.b | 6443 | 6263 | 21 | 4 |  | 12732 |
| 27.7.c | 162839 | 12325 | 16 | 28 |  | 175208 |
| 27.7.e | 0 | 1 |  | 62 |  | 64 |
| 27.7.f | 20 |  |  |  |  | 20 |
| 27.7.g | 0 | 13 | 554 | 1 |  | 569 |
| 27.7.h | 6 | 35 | 4 | 291 |  | 336 |
| 27.7.j | 151 | 316 | 547 | 212 |  | 1225 |
| 27.7.k | 154911 | 27 | 0 |  |  | 154937 |
| 27.8.a | 0 | 10 | 0 | 3560 |  | 3570 |
| 27.8.b | 80 | 67 | 7 | 10 |  | 164 |
| 27.8.c | 4188 | 4179 | 5627 | 2766 |  | 16760 |
| 27.8.d | 0 |  |  | 1959 |  | 1959 |
| 27.9.a | 1806 | 4041 | 2318 | 2323 |  | 10487 |
| Total | 481063 | 515826 | 31806 | 114591 | 162 | 1143450 |

*Discards data from UK(Scotland) were provided by year, due to sampling intensity.

Table 2.3.1.4. Blue whiting. ICES estimated catches (tonnes) from the main fisheries 1988-2021 by area.

| Year | Norwegian Sea fishery $\begin{aligned} & \text { (SAs1+2;Divs. } 5 \\ & . a, 14 a-b) \end{aligned}$ | Fishery in the spawning area (SA 12.; Divs. 5.b, 6.ab, 7.a-c) | Directedand mixed fisheries in the North Sea (SA4; Div.3.a) | Total northern areas | Total southern areas (SAs8+9;Div s.7.d-k) | Grand total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 55829 | 426037 | 45143 | 527009 | 30838 | 557847 |
| 1989 | 42615 | 475179 | 75958 | 593752 | 33695 | 627447 |
| 1990 | 2106 | 463495 | 63192 | 528793 | 32817 | 561610 |
| 1991 | 78703 | 218946 | 39872 | 337521 | 32003 | 369524 |
| 1992 | 62312 | 318018 | 65974 | 446367 | 28722 | 475026 |
| 1993 | 43240 | 347101 | 58082 | 448423 | 32256 | 480679 |
| 1994 | 22674 | 378704 | 28563 | 429941 | 29473 | 459414 |
| 1995 | 23733 | 423504 | 104004 | 551241 | 27664 | 578905 |
| 1996 | 23447 | 478077 | 119359 | 620883 | 25099 | 645982 |
| 1997 | 62570 | 514654 | 65091 | 642315 | 30122 | 672437 |
| 1998 | 177494 | 827194 | 94881 | 1099569 | 29400 | 1128969 |
| 1999 | 179639 | 943578 | 106609 | 1229826 | 26402 | 1256228 |
| 2000 | 284666 | 989131 | 114477 | 1388274 | 24654 | 1412928 |
| 2001 | 591583 | 1045100 | 118523 | 1755206 | 24964 | 1780170 |
| 2002 | 541467 | 846602 | 145652 | 1533721 | 23071 | 1556792 |
| 2003 | 931508 | 1211621 | 158180 | 2301309 | 20097 | 2321406 |
| 2004 | 921349 | 1232534 | 138593 | 2292476 | 85093 | 2377569 |
| 2005 | 405577 | 1465735 | 128033 | 1999345 | 27608 | 2026953 |
| 2006 | 404362 | 1428208 | 105239 | 1937809 | 28331 | 1966140 |
| 2007 | 172709 | 1360882 | 61105 | 1594695 | 17634 | 1612330 |
| 2008 | 68352 | 1111292 | 36061 | 1215704 | 30761 | 1246465 |
| 2009 | 46629 | 533996 | 22387 | 603012 | 32627 | 635639 |
| 2010 | 36214 | 441521 | 17545 | 495280 | 28552 | 523832 |
| 2011 | 20599 | 72279 | 7524 | 100401 | 3191 | 103592 |
| 2012 | 24391 | 324545 | 5678 | 354614 | 29402 | 384016* |
| 2013 | 31759 | 481356 | 8749 | 521864 | 103973 | 625837** |
| 2014 | 45580 | 885483 | 28596 | 959659 | 195620 | 1155279 |
| 2015 | 150828 | 895684 | 44661 | 1091173 | 305071 | 1396244 |
| 2016 | 59744 | 905087 | 55774 | 1020604 | 162583 | 1183187*** |
| 2017 | 136565 | 1284105 | 45474 | 1466144 | 91917 | 1558061 |
| 2018 | 143204 | 1445957 | 43484 | 1632646 | 78831 | 1711477 |
| 2019 | 68593 | 1271883 | 44856 | 1385333 | 130194 | 1515527 |
| 2020 | 92084 | 1059197 | 64327 | 1215608 | 279640 | 1495248 |
| 2021 | 112082 | 801768 | 39509 | 953359 | 190091 | 1143450 |

[^1]Table 2.3.1.5. Blue whiting. ICES estimates (tonnes) of catches, landings and discards by country for 2021.

| Country | Catches | Landings | Discards | \% discards |
| :---: | :---: | :---: | :---: | :---: |
| Denmark | 40321 | 40269 | 52 | 0.13 |
| Faroe Islands | 202415 | 202415 | 0 | 0.00 |
| France | 14612 | 14612 | 0 | 0.00 |
| Germany | 35327 | 35327 | 0 | 0.00 |
| Greenland | 20190 | 20190 | 0 | 0.00 |
| Iceland | 190146 | 190146 | 0 | 0.00 |
| Ireland | 39514 | 38959 | 554 | 1.40 |
| Lithuania | 21183 | 21183 | 0 | 0.00 |
| Netherlands | 62017 | 62017 | 0 | 0.00 |
| Norway | 233968 | 233968 | 0 | 0.00 |
| Poland | 26077 | 26077 | 0 | 0.00 |
| Portugal | 2522 | 1417 | 1105 | 43.81 |
| Russia ^ | 133605 | 133605 | 0 | 0.00 |
| Spain | 25509 | 23471 | 2038 | 7.99 |
| Sweden | 40 | 40 | 0 | 0.00 |
| UK (England) | 8783 | 8758 | 24 | 0.28 |
| UK(Scotland) | 65085 | 64923 | 162 | 0.25 |
| Unallocated | 22137 | 22137 | 0 | 0.00 |
| Total | 1143450 | 1139514 | 3936 | 0.34 |

^ Russia 2021 preliminary data (Q1+Q2) submitted to WGWIDE 2021.

Table 2.3.1.6. Blue whiting. ICES estimated catches (tonnes) inside and outside NEAFC regulatory area for 2021 by country.

| Country | Catches inside NEAFC RA | Catches outside NEAFC RA | Total catches |
| :--- | ---: | ---: | ---: | ---: |
| Denmark | 1965 | 38356 | 40321 |
| Faroe Islands | 27809 | 174606 | 202415 |
| France* | 0 | 14612 | 14612 |
| Germany | 57 | 35270 | 35327 |
| Greenland | 11211 | 8979 | 20190 |
| Iceland | 4737 | 185409 | 190146 |
| lreland | 1202 | 38312 | 39514 |
| Lithuania | 127 | 21055 | 21183 |
| Netherlands | 5 | 62012 | 62017 |
| Norway* | 100017 | 133952 | 233968 |
| Poland | 7531 | 18546 | 26077 |
| Portugal | 0 | 2522 | 2522 |
| Spain | 0 | 25509 | 25509 |
| Sweden | 0 | 40 | 40 |
| UK (England) | 0 | 8783 | 8783 |
| UK(Scotland) | 0 | 65085 | 65085 |
| Total in 2021** | $\mathbf{1 5 4 6 6 1}$ | $\mathbf{8 3 3 0 4 7}$ | $\mathbf{9 8 7 7 0 8}$ |

[^2]Table 2.3.1.1.1. Blue whiting. ICES estimated catches (tonnes), the percentage of catch covered by the sampling programme, No. of age samples, No. of fish measured and No. of fish aged for 2000-2021.

| Year | Catch (tonnes) | \% catch covered by sampling programme | No. Age samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 1412928 | * | 1136 | 125162 | 13685 |
| 2001 | 1780170 | * | 985 | 173553 | 17995 |
| 2002 | 1556792 | * | 1037 | 116895 | 19202 |
| 2003 | 2321406 | * | 1596 | 188770 | 26207 |
| 2004 | 2377569 | * | 1774 | 181235 | 27835 |
| 2005 | 2026953 | * | 1833 | 217937 | 32184 |
| 2006 | 1966140 | * | 1715 | 190533 | 27014 |
| 2007 | 1610090 | 87 | 1399 | 167652 | 23495 |
| 2008 | 1246465 | 90 | 927 | 113749 | 21844 |
| 2009 | 635639 | 88 | 705 | 79500 | 18142 |
| 2010 | 524751 | 87 | 584 | 82851 | 16323 |
| 2011 | 103591 | 85 | 697 | 84651 | 12614 |
| 2012 | 373937 | 80 | 1143 | 173206 | 15745 |
| 2013 | 625837 | 96 | 915 | 111079 | 14633 |
| 2014 | 1155279 | 89 | 912 | 111316 | 39738 |
| 2015 | 1396244 | 94 | 1570 | 102367 | 29821 |
| 2016 | 1183187 | 89 | 1092 | 120329 | 13793 |
| 2017 | 1558061 | 91 | 1779 | 147297 | 15828 |
| 2018 | 1711477 | 87 | 1565 | 131779 | 16426 |
| 2019 | 1515527 | 84 | 1253 | 136604 | 17869 |
| 2020 | 1495248 | 81 | 672 | 89110 | 16641 |
| 2021 | 1143450 | 81 | 1424 | 129317 | 15379 |

Table 2.3.1.1.2. Blue whiting. ICES estimated catches (tonnes), the percentage of catch covered by the sampling programme (catch-at-age numbers), No. of length samples, No. of age samples, No. of fish measured, No. of fish aged, No. of fish aged by 1000 tonnes and No. of fish measured by 1000 tonnes by country for 2021.

| Country | Catch (ton) | \% catch covered by sampling programme | No. Length samples | No. Age samples | No. <br> Measured | No. <br> Aged | No Aged/ 1000 tonnes | No Measured/ 1000 tonnes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 40321 | 94 | 19 | 19 | 817 | 815 | 20 | 20 |
| Faroe Islands | 202415 | 93 | 17 | 17 | 1796 | 1651 | 8 | 9 |
| France | 14612 | 15 | 21 | 0 | 3613 | 0 | 0 | 247 |
| Germany | 35327 | 27 | 4 | 4 | 2492 | 691 | 20 | 71 |
| Greenland | 20190 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Iceland | 190146 | 97 | 64 | 64 | 5639 | 1725 | 9 | 30 |
| Ireland | 39514 | 98 | 53 | 29 | 10548 | 1800 | 46 | 267 |
| Lithuania | 21183 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Netherlands | 62017 | 82 | 55 | 55 | 11483 | 1350 | 22 | 185 |
| Norway | 233968 | 90 | 68 | 68 | 2020 | 2020 | 9 | 9 |
| Poland | 26077 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Portugal | 2522 | 83 | 34 | 34 | 3129 | 1003 | 398 | 1241 |
| Russia^ | 133605 | 88 | 133 | 133 | 40413 | 1382 | 10 | 302 |
| Spain | 25509 | 98 | 984 | 984 | 41289 | 2259 | 89 | 1619 |
| Sweden | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK (England) | 8783 | 0 | 13 | 0 | 95 | 0 | 0 | 11 |
| UK(Scotland) | 65085 | 95 | 47 | 17 | 5983 | 683 | 10 | 92 |
| Unallocated | 22137 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total Geral | 1143450 | 81 | 1512 | 1424 | 129317 | 15379 | 13 | 113 |

[^3]Table 2.3.1.1.3. Blue whiting. ICES estimated catches (tonnes), No. of Age samples, No. of fish measured and No. of fish aged by country and quarter for 2021.

| Country | Catches (ton) | No. of Length Samples | No. of Length Measured | No. Age Readings |
| :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  |
| Quarter 1 | 27685 | 12 | 541 | 539 |
| Quarter 2 | 10331 | 7 | 276 | 276 |
| Quarter 3 | 102 | 0 | 0 | 0 |
| Quarter 4 | 2204 | 0 | 0 | 0 |
| Total | 40321 | 19 | 817 | 815 |
| Faroe Islands |  |  |  |  |
| Quarter 1 | 64013 | 9 | 1025 | 894 |
| Quarter 2 | 124078 | 7 | 702 | 688 |
| Quarter 3 | 2541 | 0 | 0 | 0 |
| Quarter 4 | 11783 | 1 | 69 | 69 |
| Total | 202415 | 17 | 1796 | 1651 |
| France |  |  |  |  |
| Quarter 1 | 237 | 0 | 0 | 0 |
| Quarter 2 | 12110 | 5 | 302 | 0 |
| Quarter 3 | 1 | 0 | 0 | 0 |
| Quarter 4 | 2263 | 16 | 3311 | 0 |
| Total | 14612 | 21 | 3613 | 0 |
| Germany |  |  |  |  |
| Quarter 1 | 21899 | 1 | 327 | 162 |
| Quarter 2 | 11979 | 0 | 0 | 0 |
| Quarter 3 | 14 | 0 | 0 | 0 |
| Quarter 4 | 1434 | 3 | 2165 | 529 |
| Total | 35327 | 4 | 2492 | 691 |
| Greenland |  |  |  |  |
| Quarter 2 | 17737 | 0 | 0 | 0 |
| Quarter 3 | 79 | 0 | 0 | 0 |
| Quarter 4 | 2374 | 0 | 0 | 0 |
| Total | 20190 | 0 | 0 | 0 |
| Iceland |  |  |  |  |
| Quarter 1 | 23123 | 10 | 739 | 244 |
| Quarter 2 | 129192 | 37 | 3650 | 1116 |
| Quarter 3 | 1861 | 0 | 0 | 0 |
| Quarter 4 | 35970 | 17 | 1250 | 365 |
| Total | 190146 | 64 | 5639 | 1725 |
| Ireland |  |  |  |  |
| Quarter 1 | 22817 | 20 | 5266 | 1300 |
| Quarter 2 | 16131 | 9 | 2368 | 500 |
| Quarter 3 | 554 | 24 | 2914 | 0 |
| Quarter 4 | 11 | 0 | 0 | 0 |
| Total | 39514 | 53 | 10548 | 1800 |

Table 2.3.1.1.3. (continued) Blue whiting. ICES estimated catches (tonnes), No. of Age samples, No. of fish measured and No. of fish aged by country and quarter for 2021.

| Lithuania |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Quarter 1 | 9117 | 0 | 0 | 0 |
| Quarter 4 | 12066 | 0 | 0 | 0 |
| Total | 21183 | 0 | 0 | 0 |
| Netherlands |  |  |  |  |
| Quarter 1 | 22908 | 30 | 5544 | 737 |
| Quarter 2 | 32039 | 22 | 5219 | 538 |
| Quarter 3 | 375 | 0 | 0 | 0 |
| Quarter 4 | 6694 | 3 | 720 | 75 |
| Total | 62017 | 55 | 11483 | 1350 |
| Norway |  |  |  |  |
| Quarter 1 | 174903 | 59 | 1753 | 1753 |
| Quarter 2 | 41332 | 8 | 237 | 237 |
| Quarter 3 | 8130 | 0 | 0 | 0 |
| Quarter 4 | 9604 | 1 | 30 | 30 |
| Total | 233968 | 68 | 2020 | 2020 |
| Poland |  |  |  |  |
| Quarter 1 | 12445 | 0 | 0 | 0 |
| Quarter 4 | 13633 | 0 | 0 | 0 |
| Total | 26077 | 0 | 0 | 0 |
| Portugal |  |  |  |  |
| Quarter 1 | 646 | 6 | 434 | 100 |
| Quarter 2 | 529 | 12 | 1092 | 269 |
| Quarter 3 | 631 | 10 | 825 | 333 |
| Quarter 4 | 716 | 6 | 778 | 301 |
| Total | 2522 | 34 | 3129 | 1003 |
| Russia ${ }^{\wedge}$ |  |  |  |  |
| Quarter 1 | 61551 | 84 | 25439 | 1092 |
| Quarter 2 | 72054 | 49 | 14974 | 290 |
| Total | 133605 | 133 | 40413 | 1382 |
| Spain |  |  |  |  |
| Quarter 1 | 5502 | 191 | 8051 | 200 |
| Quarter 2 | 8254 | 268 | 9292 | 709 |
| Quarter 3 | 7340 | 313 | 12849 | 593 |
| Quarter 4 | 4414 | 212 | 11097 | 757 |
| Total | 25509 | 984 | 41289 | 2259 |
| Sweden |  |  |  |  |
| Quarter 1 | 0 | 0 | 0 | 0 |
| Quarter 2 | 0 | 0 | 0 | 0 |
| Quarter 3 | 9 | 0 | 0 | 0 |
| Quarter 4 | 31 | 0 | 0 | 0 |
| Total | 40 | 0 | 0 | 0 |

[^4]Table 2.3.1.1.3. (continued) Blue whiting. ICES estimated catches (tonnes), No. of Age samples, No. of fish measured and No. of fish aged by country and quarter for 2021.

| UK (England) |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: |
| Quarter 1 | 20 | 0 | 0 | 0 |  |
| Quarter 2 | 7136 | 13 | 95 | 0 |  |
| Quarter 3 | 180 | 0 | 0 | 0 |  |
| Quarter 4 | 1447 | 0 | 0 | 0 |  |
| UK(Scotland) | 8783 | 13 | $\mathbf{9 5}$ | $\mathbf{0}$ |  |
| Quarter 1 |  |  |  |  |  |
| Quarter 2 | 34198 | 10 | 2248 | 456 |  |
| Quarter 4 | 30703 | 7 | 872 | 227 |  |
| 2021* | 22 | 0 | 0 | 0 |  |
|  | 162 | 30 | 2863 | 0 |  |
| Unallocated | $\mathbf{T o t a l}$ |  | $\mathbf{4 7}$ | $\mathbf{5 9 8 3}$ | $\mathbf{6 8 3}$ |
| Quarter 2 | 2223 |  |  |  |  |
| Quarter 3 | 9988 | 0 | 0 | 0 |  |
| Quarter 4 | 9925 | 0 | 0 | 0 |  |
|  | Total | 22137 | 0 | 0 | 0 |
|  | $\mathbf{1 1 4 3 4 5 0}$ | $\mathbf{0}$ | 0 | 0 |  |
| Total Geral |  | $\mathbf{1 5 1 2}$ | $\mathbf{1 2 9 3 1 7}$ | $\mathbf{1 5 3 7 9}$ |  |

* discards data not raised by quarter due to sampling intensity.

Table 2.3.1.1.4. Blue whiting. ICES estimated catches (tonnes), the percentage of catch covered by the sampling programme, No. of length samples, No. of age samples, No. of fish measured, No. of fish aged, No. of fish aged by 1000 tonnes and No. of fish measured by 1000 tonnes by ICES division for 2021.

| ICES Division | Catch (tonnes) | No. Length samples | No. Age samples | No. Measured | No. Aged | No Aged/ 1000 tonnes | No Measured/ 1000 tonnes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 27.4 | 100 | 18 | 0 | 1508 | 0 | 0 | 15050 |
| 27.2.a | 91606 | 19 | 14 | 1936 | 540 | 6 | 21 |
| 27.3.a | 106 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.4.a | 39290 | 2 | 2 | 149 | 58 | 1 | 4 |
| 27.4.b | 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.5.a | 20476 | 13 | 13 | 932 | 287 | 14 | 46 |
| 27.5.b | 298640 | 54 | 54 | 11644 | 2179 | 7 | 39 |
| 27.6.a | 306385 | 133 | 124 | 21812 | 3501 | 11 | 71 |
| 27.6.b | 8804 | 7 | 4 | 1670 | 226 | 26 | 190 |
| 27.7.b | 12732 | 6 | 6 | 1210 | 174 | 14 | 95 |
| 27.7.c | 175208 | 179 | 179 | 26304 | 3270 | 19 | 150 |
| 27.7.e | 64 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.7.f | 20 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.7.g | 569 | 32 | 0 | 2950 | 0 | 0 | 5182 |
| 27.7.h | 336 | 30 | 25 | 376 | 0 | 0 | 1118 |
| 27.7.j | 1225 | 109 | 109 | 525 | 24 | 20 | 428 |
| 27.7.k | 154937 | 74 | 74 | 12187 | 1835 | 12 | 79 |
| 27.8.a | 3570 | 7 | 1 | 744 | 23 | 6 | 208 |
| 27.8.b | 164 | 111 | 111 | 1465 | 0 | 0 | 8959 |
| 27.8.c | 16760 | 383 | 383 | 25706 | 1066 | 64 | 1534 |
| 27.8.d | 1959 | 10 | 0 | 2890 | 0 | 0 | 1475 |
| 27.9.a | 10487 | 325 | 325 | 15309 | 2196 | 209 | 1460 |
| TOTAL | 1143450 | 1512 | 1424 | 129317 | 15379 | 13 | 113 |

Table 2.3.2.1. Blue whiting. ICES estimated preliminary landings (tonnes) in 2022 by quarter and ICES division. Data submitted to InterCatch.

|  | Landings |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| ICES div. | Quarter 1 | Quarter 2 | Quarter 3 | Total |
| 27.2.a | 1487 | 14702 | 65 | 16253 |
| 27.3.a |  | 4 | 6 | 11 |
| 27.4.a | 6026 | 22379 |  | 28405 |
| 27.4.b |  | 0.09 |  | 0.09 |
| 27.5.a |  | 15 |  | 15 |
| 27.5.b | 54091 | 242073 | 1 | 296165 |
| 27.6.a | 31769 | 113921 |  | 145691 |
| 27.6.b | 5860 | 680 |  | 6540 |
| 27.7 | 4 | 28 |  | 32 |
| 27.7.b | 1639 |  |  | 1639 |
| 27.7.c | 116209 | 855 |  | 117065 |
| 27.7.f | 0.38 |  |  | 0.38 |
| 27.7.g | 1 |  |  | 1 |
| 27.7.j | 2165 |  |  | 2165 |
| 27.7.k | 69116 |  |  | 69116 |
| 27.8.a | 5 | 18 |  | 23 |
| 27.8.c | 4574 | 7183 |  | 11757 |
| 27.8.d | 2700 |  |  | 2700 |
| 27.9.a | 362 | 524 |  | 886 |
| Total | 296008 | 402383 | 72 | 698463 |

Table 2.3.2.2. Blue whiting. ICES estimated preliminary catches (tonnes), the percentage of catch covered by the sampling programme, No. of samples, No. of fish measured, No. of fish aged, No. of fish aged by 1000 tonnes and No. of fish measured by 1000 tonnes by ICES division for 2022 preliminary data (quarters 1 and 2). Data submitted to InterCatch.

| ICES Division | Catch (tonnes) | No. samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: |
| 27.2.a | 16253 | 1 | 116 | 100 |
| 27.3.a | 11 | 0 | 0 | 0 |
| 27.4.a | 28405 | 1 | 100 | 100 |
| 27.4.b | 0 | 0 | 0 | 0 |
| 27.5.a | 15 | 0 | 0 | 0 |
| 27.5.b | 296165 | 6 | 494 | 459 |
| 27.6.a | 145691 | 31 | 4645 | 1458 |
| 27.6.b | 6540 | 1 | 30 | 30 |
| 27.7 | 32 | 0 | 0 | 0 |
| 27.7.b | 1639 | 0 | 0 | 0 |
| 27.7.c | 117065 | 27 | 3964 | 1349 |
| 27.7.f | 0 | 0 | 0 | 0 |
| 27.7.g | 1 | 0 | 0 | 0 |
| 27.7.j | 2165 | 0 | 0 | 0 |
| 27.7.k | 69116 | 11 | 888 | 516 |
| 27.8.a | 23 | 0 | 0 | 0 |
| 27.8.c | 11757 | 0 | 0 | 0 |
| 27.8.d | 2700 | 0 | 0 | 0 |
| 27.9.a | 886 | 18 | 1526 | 369 |
| Total | 698463 | 96 | 11763 | 4381 |

Table 2.3.2.3. Blue whiting. ICES estimates of catches (tonnes) in 2022, based on (initial) declared quotas and expected uptake estimated by WGWIDE.

| Country | Q1 | Q2 | Q3 | Preliminary Catch | Expected Catch remained catch or Total year catch | Total Catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 25639 | 19782 | 6 | 45427 | 10 | 45437 |
| Faroe Islands | 65427 | 139620 | 66 | 205113 | 66218 | 271331 |
| France |  |  |  |  | 9128 | 9128 |
| Germany | 15594 | 601 |  | 16195 | 13988 | 30183 |
| Greenland |  |  |  |  | 22878 | 22878 |
| Iceland | 0 | 157563 |  | 157563 | 12707 | 174557 |
| Ireland | 17232 | 11221 |  | 28453 | 0 | 28453 |
| Lithuania |  |  |  |  | 6467 | 6467 |
| Netherlands | 4799 |  |  | 4799 | 74948 | 79747 |
| Norway | 111166 |  |  | 164650 | 25350 | 190000 |
| Poland | 18024 | 2924 |  | 20948 | 5865 | 26814 |
| Portugal | 224 | 412 |  | 636 | 1364 | 2000 |
| Spain | 4716 | 7323 |  | 12039 | 10961 | 23000 |
| UK(Scotland) | 35264 | 16945 | 1277 | 53486 | 0 | 53486 |
| Sweden |  |  |  | 0 | 70 | 70 |
| Total reporting countries |  |  |  |  |  | 963550 |
| Russia (assumed $13 \%$ of total) |  |  |  |  |  | 143979 |
| Total |  |  |  |  |  | 1107529 |

Table 2.3.2.4. Blue whiting. Comparison of preliminary and final catches (in tonnes) calculated from sum of product of catch number and mean weight at age used in the assessment).

|  | Final | Preliminary | Change in \% * |
| :--- | :--- | :--- | :--- |
| 2016 | 1180786 | 1147000 | 2.9 |
| 2017 | 1555069 | 1559437 | -0.3 |
| 2018 | 1709856 | 1712874 | -0.2 |
| 2019 | 1460507 | 1474301 | 4.7 |
| 2020 | 1139531 | 1242727 | -1.2 |
| 2021 |  | -8.3 |  |

* (final-preliminary)/preliminary*100

Table 2.3.3.1. Blue whiting. Catch-at-age numbers (thousands) by year. Discards included since 2014. Values for 2022 are preliminary.

| Year <br> Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1981 | 258000 | 348000 | 681000 | 334000 | 548000 | 559000 | 466000 | 634000 | 578000 | 1460000 |
| 1982 | 148000 | 274000 | 326000 | 548000 | 264000 | 276000 | 266000 | 272000 | 284000 | 673000 |
| 1983 | 2283000 | 567000 | 270000 | 286000 | 299000 | 304000 | 287000 | 286000 | 225000 | 334000 |
| 1984 | 2291000 | 2331000 | 455000 | 260000 | 285000 | 445000 | 262000 | 193000 | 154000 | 255000 |
| 1985 | 1305000 | 2044000 | 1933000 | 303000 | 188000 | 321000 | 257000 | 174000 | 93000 | 259000 |
| 1986 | 650000 | 816000 | 1862000 | 1717000 | 393000 | 187000 | 201000 | 198000 | 174000 | 398000 |
| 1987 | 838000 | 578000 | 728000 | 1897000 | 726000 | 137000 | 105000 | 123000 | 103000 | 195000 |


| $\begin{aligned} & \text { Year } \\ & \text { Age } \end{aligned}$ | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 425000 | 721000 | 614000 | 683000 | 1303000 | 618000 | 84000 | 53000 | 33000 | 50000 |
| 1989 | 865000 | 718000 | 1340000 | 791000 | 837000 | 708000 | 139000 | 50000 | 25000 | 38000 |
| 1990 | 1611000 | 703000 | 672000 | 753000 | 520000 | 577000 | 299000 | 78000 | 27000 | 95000 |
| 1991 | 266686 | 1024468 | 513959 | 301627 | 363204 | 258038 | 159153 | 49431 | 5060 | 9570 |
| 1992 | 407730 | 653838 | 1641714 | 569094 | 217386 | 154044 | 109580 | 79663 | 31987 | 11706 |
| 1993 | 263184 | 305180 | 621085 | 1571236 | 411367 | 191241 | 107005 | 64769 | 38118 | 17476 |
| 1994 | 306951 | 107935 | 367962 | 389264 | 1221919 | 281120 | 174256 | 90429 | 79014 | 30614 |
| 1995 | 296100 | 353949 | 421560 | 465358 | 615994 | 800201 | 253818 | 159797 | 59670 | 41811 |
| 1996 | 1893453 | 534221 | 632361 | 537280 | 323324 | 497458 | 663133 | 232420 | 98415 | 82521 |
| 1997 | 2131494 | 1519327 | 904074 | 577676 | 295671 | 251642 | 282056 | 406910 | 104320 | 169235 |
| 1998 | 1656926 | 4181175 | 3541231 | 1044897 | 383658 | 322777 | 303058 | 264105 | 212452 | 85513 |
| 1999 | 788200 | 1549100 | 5820800 | 3460600 | 412800 | 207200 | 151200 | 153100 | 68800 | 140500 |
| 2000 | 1814851 | 1192657 | 3465739 | 5014862 | 1550063 | 513663 | 213057 | 151429 | 58277 | 139791 |
| 2001 | 4363690 | 4486315 | 2962163 | 3806520 | 2592933 | 585666 | 170020 | 97032 | 76624 | 66410 |
| 2002 | 1821053 | 3232244 | 3291844 | 2242722 | 1824047 | 1647122 | 344403 | 168848 | 102576 | 142743 |
| 2003 | 3742841 | 4073497 | 8378955 | 4824590 | 2035096 | 1117179 | 400022 | 121280 | 19701 | 27493 |
| 2004 | 2156261 | 4426323 | 6723748 | 6697923 | 3044943 | 1276412 | 649885 | 249097 | 75415 | 36805 |
| 2005 | 1427277 | 1518938 | 5083550 | 5871414 | 4450171 | 1419089 | 518304 | 249443 | 100374 | 55226 |
| 2006 | 412961 | 939865 | 4206005 | 6150696 | 3833536 | 1718775 | 506198 | 181181 | 67573 | 36688 |
| 2007 | 167027 | 306898 | 1795021 | 4210891 | 3867367 | 2353478 | 935541 | 320529 | 130202 | 88573 |
| 2008 | 408790 | 179211 | 545429 | 2917190 | 3262956 | 1919264 | 736051 | 315671 | 113086 | 126637 |
| 2009 | 61125 | 156156 | 231958 | 594624 | 1596095 | 1156999 | 592090 | 251529 | 88615 | 48908 |
| 2010 | 349637 | 222975 | 160101 | 208279 | 646380 | 992214 | 702569 | 256604 | 70487 | 43693 |
| 2011 | 162997 | 101810 | 63954 | 53863 | 69717 | 116396 | 120359 | 55470 | 25943 | 12542 |
| 2012 | 239667 | 351845 | 663155 | 141854 | 106883 | 203419 | 363779 | 356785 | 212492 | 157947 |
| 2013 | 228175 | 508122 | 848597 | 896966 | 462714 | 224066 | 321310 | 397536 | 344285 | 383601 |
| 2014 | 588717 | 584084 | 2312953 | 2019373 | 1272862 | 416523 | 386396 | 462339 | 526141 | 662747 |
| 2015 | 2944849 | 2852384 | 2427329 | 2465286 | 1518235 | 707533 | 329882 | 258743 | 239164 | 450046 |
| 2016 | 1239331 | 3518677 | 2933271 | 1874011 | 1367844 | 756824 | 339851 | 185368 | 131039 | 288635 |


| Year <br> Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $\mathbf{2 0 1 7}$ | 401947 | 1999011 | 7864694 | 4063916 | 1509651 | 777185 | 263007 | 110351 | 63945 | 149369 |
| $\mathbf{2 0 1 8}$ | 418781 | 541041 | 3572357 | 7340084 | 2983975 | 1022883 | 424206 | 150753 | 90387 | 163289 |
| 2019 | 249923 | 433573 | 1288871 | 3778379 | 5037323 | 1645999 | 431925 | 145916 | 50622 | 81357 |
| 2020 | 1135859 | 834162 | 1106838 | 1797157 | 3072708 | 3041983 | 923392 | 235330 | 80440 | 64535 |
| 2021 | 2069387 | 830692 | 1266077 | 1214790 | 1438769 | 1404443 | 1360104 | 304891 | 100993 | 59441 |
| 2022 | 906699 | 3344062 | 1873517 | 1778289 | 1092800 | 814544 | 753595 | 795714 | 130995 | 95271 |

Table 2.3.4.1. Blue whiting. Individual mean weight (kg) at age in the catch. Preliminary values for 2022.

| Year Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 0.052 | 0.065 | 0.103 | 0.125 | 0.141 | 0.155 | 0.170 | 0.178 | 0.187 | 0.213 |
| 1982 | 0.045 | 0.072 | 0.111 | 0.143 | 0.156 | 0.177 | 0.195 | 0.200 | 0.204 | 0.231 |
| 1983 | 0.046 | 0.074 | 0.118 | 0.140 | 0.153 | 0.176 | 0.195 | 0.200 | 0.204 | 0.228 |
| 1984 | 0.035 | 0.078 | 0.089 | 0.132 | 0.153 | 0.161 | 0.175 | 0.189 | 0.186 | 0.206 |
| 1985 | 0.038 | 0.074 | 0.097 | 0.114 | 0.157 | 0.177 | 0.199 | 0.208 | 0.218 | 0.237 |
| 1986 | 0.040 | 0.073 | 0.108 | 0.130 | 0.165 | 0.199 | 0.209 | 0.243 | 0.246 | 0.257 |
| 1987 | 0.048 | 0.086 | 0.106 | 0.124 | 0.147 | 0.177 | 0.208 | 0.221 | 0.222 | 0.254 |
| 1988 | 0.053 | 0.076 | 0.097 | 0.128 | 0.142 | 0.157 | 0.179 | 0.199 | 0.222 | 0.260 |
| 1989 | 0.059 | 0.079 | 0.103 | 0.126 | 0.148 | 0.158 | 0.171 | 0.203 | 0.224 | 0.253 |
| 1990 | 0.045 | 0.070 | 0.106 | 0.123 | 0.147 | 0.168 | 0.175 | 0.214 | 0.217 | 0.256 |
| 1991 | 0.055 | 0.091 | 0.107 | 0.136 | 0.174 | 0.190 | 0.206 | 0.230 | 0.232 | 0.266 |
| 1992 | 0.057 | 0.083 | 0.119 | 0.140 | 0.167 | 0.193 | 0.226 | 0.235 | 0.284 | 0.294 |
| 1993 | 0.066 | 0.082 | 0.109 | 0.137 | 0.163 | 0.177 | 0.200 | 0.217 | 0.225 | 0.281 |
| 1994 | 0.061 | 0.087 | 0.108 | 0.137 | 0.164 | 0.189 | 0.207 | 0.217 | 0.247 | 0.254 |
| 1995 | 0.064 | 0.091 | 0.118 | 0.143 | 0.154 | 0.167 | 0.203 | 0.206 | 0.236 | 0.256 |
| 1996 | 0.041 | 0.080 | 0.102 | 0.116 | 0.147 | 0.170 | 0.214 | 0.230 | 0.238 | 0.279 |
| 1997 | 0.047 | 0.072 | 0.102 | 0.121 | 0.140 | 0.166 | 0.177 | 0.183 | 0.203 | 0.232 |
| 1998 | 0.048 | 0.072 | 0.094 | 0.125 | 0.149 | 0.178 | 0.183 | 0.188 | 0.221 | 0.248 |
| 1999 | 0.063 | 0.078 | 0.088 | 0.109 | 0.142 | 0.170 | 0.199 | 0.193 | 0.192 | 0.245 |
| 2000 | 0.057 | 0.075 | 0.086 | 0.104 | 0.133 | 0.156 | 0.179 | 0.187 | 0.232 | 0.241 |
| 2001 | 0.050 | 0.078 | 0.094 | 0.108 | 0.129 | 0.163 | 0.186 | 0.193 | 0.231 | 0.243 |


| Year Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | 0.054 | 0.074 | 0.093 | 0.115 | 0.132 | 0.155 | 0.173 | 0.233 | 0.224 | 0.262 |
| 2003 | 0.049 | 0.075 | 0.098 | 0.108 | 0.131 | 0.148 | 0.168 | 0.193 | 0.232 | 0.258 |
| 2004 | 0.042 | 0.066 | 0.089 | 0.102 | 0.123 | 0.146 | 0.160 | 0.173 | 0.209 | 0.347 |
| 2005 | 0.039 | 0.068 | 0.084 | 0.099 | 0.113 | 0.137 | 0.156 | 0.166 | 0.195 | 0.217 |
| 2006 | 0.049 | 0.072 | 0.089 | 0.105 | 0.122 | 0.138 | 0.163 | 0.190 | 0.212 | 0.328 |
| 2007 | 0.050 | 0.064 | 0.091 | 0.103 | 0.115 | 0.130 | 0.146 | 0.169 | 0.182 | 0.249 |
| 2008 | 0.055 | 0.075 | 0.100 | 0.106 | 0.120 | 0.133 | 0.146 | 0.160 | 0.193 | 0.209 |
| 2009 | 0.056 | 0.085 | 0.105 | 0.119 | 0.124 | 0.138 | 0.149 | 0.179 | 0.214 | 0.251 |
| 2010 | 0.052 | 0.064 | 0.110 | 0.154 | 0.154 | 0.163 | 0.175 | 0.187 | 0.200 | 0.272 |
| 2011 | 0.055 | 0.079 | 0.107 | 0.136 | 0.169 | 0.169 | 0.179 | 0.189 | 0.214 | 0.270 |
| 2012 | 0.041 | 0.072 | 0.098 | 0.141 | 0.158 | 0.172 | 0.180 | 0.185 | 0.189 | 0.203 |
| 2013 | 0.051 | 0.077 | 0.094 | 0.117 | 0.139 | 0.162 | 0.185 | 0.188 | 0.198 | 0.197 |
| 2014 | 0.049 | 0.078 | 0.093 | 0.112 | 0.128 | 0.155 | 0.178 | 0.190 | 0.202 | 0.217 |
| 2015 | 0.039 | 0.070 | 0.094 | 0.117 | 0.137 | 0.155 | 0.174 | 0.183 | 0.193 | 0.201 |
| 2016 | 0.047 | 0.066 | 0.084 | 0.107 | 0.125 | 0.142 | 0.152 | 0.167 | 0.184 | 0.206 |
| 2017 | 0.056 | 0.072 | 0.080 | 0.094 | 0.113 | 0.131 | 0.148 | 0.172 | 0.190 | 0.212 |
| 2018 | 0.055 | 0.080 | 0.091 | 0.098 | 0.111 | 0.129 | 0.142 | 0.165 | 0.175 | 0.216 |
| 2019 | 0.068 | 0.085 | 0.099 | 0.109 | 0.118 | 0.130 | 0.144 | 0.167 | 0.167 | 0.228 |
| 2020 | 0.063 | 0.084 | 0.099 | 0.115 | 0.127 | 0.135 | 0.144 | 0.161 | 0.176 | 0.207 |
| 2021 | 0.058 | 0.086 | 0.099 | 0.119 | 0.133 | 0.143 | 0.150 | 0.166 | 0.181 | 0.209 |
| 2022 | 0.042 | 0.065 | 0.084 | 0.104 | 0.119 | 0.136 | 0.139 | 0.158 | 0.154 | 0.199 |

$\qquad$

Table 2.3.5.1. Blue whiting. Natural mortality and proportion mature.

| AGE | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 - 1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Proportion mature | 0.00 | 0.11 | 0.40 | 0.82 | 0.86 | 0.91 | 0.94 | 1.00 |
| Natural mortality | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |

Table 2.3.7.1.1. Blue whiting. Time-series of StoX abundance estimates of blue whiting (millions) by age in the IBWSS. Total biomass in last column (1000 t). Shaded values (ages 1-8; years 2004-2022) are used as input to the assessment

| Year | Age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | TSB |
| 2004 | 1097 | 5538 | 13062 | 15134 | 5119 | 1086 | 994 | 593 | 164 | 0 | 3505 |
| 2005 | 2129 | 1413 | 5601 | 7780 | 8500 | 2925 | 632 | 280 | 129 | 23 | 2513 |
| 2006 | 2512 | 2224 | 10881 | 11695 | 4717 | 2719 | 923 | 352 | 198 | 39 | 3517 |
| 2007 | 468 | 706 | 5241 | 11244 | 8437 | 3155 | 1110 | 456 | 123 | 65 | 3274 |
| 2008 | 337 | 524 | 1455 | 6661 | 6747 | 3882 | 1719 | 1029 | 269 | 296 | 2647 |
| 2009 | 275 | 329 | 360 | 1292 | 3739 | 3458 | 1636 | 587 | 250 | 194 | 1599 |
| 2010* |  |  |  |  |  |  |  |  |  |  |  |
| 2011 | 312 | 1361 | 1135 | 930 | 1043 | 1713 | 2171 | 2423 | 1298 | 272 | 1827 |
| 2012 | 1140 | 1816 | 6454 | 1021 | 595 | 1415 | 2220 | 1777 | 1249 | 1085 | 2347 |
| 2013 | 582 | 1337 | 6175 | 7211 | 2938 | 1282 | 1308 | 1398 | 929 | 1807 | 3110 |
| 2014 | 4183 | 1491 | 5239 | 8420 | 10202 | 2754 | 772 | 577 | 899 | 2251 | 3761 |
| 2015 | 3255 | 4570 | 1891 | 3641 | 1797 | 466 | 174 | 108 | 206 | 365 | 1405 |
| 2016 | 2745 | 7893 | 10164 | 6274 | 4687 | 1539 | 413 | 133 | 235 | 361 | 2873 |
| 2017 | 262 | 2248 | 15682 | 10176 | 3762 | 1793 | 921 | 76 | 84 | 173 | 3135 |
| 2018 | 836 | 628 | 6615 | 21490 | 7692 | 2187 | 755 | 188 | 72 | 138 | 4035 |
| 2019 | 1129 | 1169 | 3468 | 9590 | 16979 | 3434 | 484 | 513 | 99 | 43 | 4198 |
| 2020** |  |  |  |  |  |  |  |  |  |  |  |
| 2021 | 1948 | 2095 | 2545 | 2275 | 3914 | 3197 | 3379 | 463 | 189 | 114 | 2357 |
| 2022 | 4461 | 9313 | 4830 | 5460 | 2587 | 1880 | 898 | 1764 | 71 | 178 | 2707 |

[^5]Table 2.3.7.2.1. Blue whiting. Estimated abundance of 1 and 2 -year old blue whiting from the International Ecosystem Survey in Nordic Seas (IESNS), 2003-2022.

| Year\Age | Age 1 | Age 2 |
| :---: | :---: | :---: |
| 2003* | 16127 | 9317 |
| 2004* | 17792 | 11020 |
| 2005* | 19933 | 7908 |
| 2006* | 2512 | 5504 |
| 2007* | 592 | 213 |
| 2008 | 25 | 17 |
| 2009 | 7 | 8 |
| 2010 | 0 | 280 |
| 2011 | 1613 | 0 |
| 2012 | 9476 | 3265 |
| 2013 | 454 | 6544 |
| 2014 | 3937 | 2030 |
| 2015 | 8563 | 2796 |
| 2016 | 4223 | 8089 |
| 2017 | 1236 | 2087 |
| 2018 | 441 | 1491 |
| 2019 | 3157 | 215 |
| 2020 | 2822 | 481 |
| 2021 | 10264 | 1500 |
| 2022 | 17169 | 10575 |

*Using the old TS-value. To compare the results all values were divided by approximately 3.1.

Table 2.3.7.2.2. Blue whiting. 1-group indices of blue whiting from the Norwegian winter survey (late January-early March) in the Barents Sea. (Blue whiting < 19 cm in total body length which most likely belong to 1-group.)

| Catch Rate |  |  |
| :---: | :---: | :---: |
| Year | All | < 19 cm |
| 1981 | 0.13 | 0 |
| 1982 | 0.17 | 0.01 |
| 1983 | 4.46 | 0.46 |
| 1984 | 6.97 | 2.47 |
| 1985 | 32.51 | 0.77 |
| 1986 | 17.51 | 0.89 |
| 1987 | 8.32 | 0.02 |
| 1988 | 6.38 | 0.97 |
| 1989 | 1.65 | 0.18 |
| 1990 | 17.81 | 16.37 |
| 1991 | 48.87 | 2.11 |
| 1992 | 30.05 | 0.06 |
| 1993 | 5.80 | 0.01 |
| 1994 | 3.02 | 0 |
| 1995 | 1.65 | 0.10 |
| 1996 | 9.88 | 5.81 |
| 1997 | 187.24 | 175.26 |
| 1998 | 7.14 | 0.21 |
| 1999 | 5.98 | 0.71 |
| 2000 | 129.23 | 120.90 |
| 2001 | 329.04 | 233.76 |
| 2002 | 102.63 | 9.69 |
| 2003 | 75.25 | 15.15 |
| 2004 | 124.01 | 36.74 |
| 2005 | 206.18 | 90.23 |
| 2006 | 269.2 | 3.52 |
| 2007 | 80.38 | 0.16 |


| Catch Rate |  |  |
| :---: | :---: | :---: |
| Year | All | < 19 cm |
| 2008 | 17.97 | 0.04 |
| 2009 | 4.50 | 0.01 |
| 2010 | 3.30 | 0.08 |
| 2011 | 1.48 | 0.01 |
| 2012 | 127.71 | 125.93 |
| 2013 | 39.54 | 2.33 |
| 2014 | 31.48 | 24.97 |
| 2015 | 148.4 | 128.34 |
| 2016 | 86.99 | 11.31 |
| 2017 | 167.16 | 0.71 |
| 2018 | 9.19 | 0.03 |
| 2019 | 12.66 | 6.00 |
| 2020 | 26.42 | 19.33 |
| 2021 | 182.86 | 161.04 |
| 2022 | 79.19 | 41.55 |

Table 2.3.7.2.3. Blue whiting. 1-group indices of blue whiting from the Icelandic bottom-trawl surveys, 1-group (<22 cm in March).

| Catch Rate |  |
| :---: | :---: |
| Year | < 22 cm |
| 1996 | 6.5 |
| 1997 | 3.4 |
| 1998 | 1.1 |
| 1999 | 6.3 |
| 2000 | 9 |
| 2001 | 5.2 |
| 2002 | 14.2 |
| 2003 | 15.4 |
| 2004 | 8.9 |
| 2005 | 8.3 |
| 2006 | 30.4 |
| 2007 | 3.9 |
| 2008 | 0.1 |
| 2009 | 1.6 |
| 2010 | 0.2 |
| 2011 | 10.8 |
| 2012 | 29.9 |
| 2013 | 11.7 |
| 2014 | 66.3 |
| 2015 | 43.8 |
| 2016 | 6.3 |
| 2017 | 1.8 |
| 2018 | 0.4 |
| 2019 | 0.1 |
| 2020 | 9.8 |
| 2021 | 79.6 |
| 2022 | 91.2 |

Table 2.3.7.2.4. Blue whiting. 1-group indices of blue whiting from Faroese bottom-trawl surveys, $\mathbf{1}$-group (<= $\mathbf{2 3} \mathbf{~ c m ~ i n ~}$ March).

| Catch Rate |  |
| :---: | :---: |
| Year | $<=23 \mathrm{~cm}$ |
| 1994 | 1401 |
| 1995 | 1162 |
| 1996 | 4821 |
| 1997 | 2307 |
| 1998 | 463 |
| 1999 | 1717 |
| 2000 | 863 |
| 2001 | 4424 |
| 2002 | 4480 |
| 2003 | 1038 |
| 2004 | 15749 |
| 2005 | 35159 |
| 2006 | 23105 |
| 2007 | 11568 |
| 2008 | 1268 |
| 2009 | 4362 |
| 2010 | 855 |
| 2011 | 23323 |
| 2012 | 8366 |
| 2013 | 13254 |
| 2014 | 70139 |
| 2015 | 34806 |
| 2016 | 21316 |
| 2017 | 4446 |
| 2018 | 1890 |
| 2019 | 286 |
| 2020 | 141 |
| 2021 | 2224 |
| 2022 | 1781 |

Table 2.4.1.1. Blue whiting. Parameter estimates, from final assessment (2022) and retrospective analysis (20182021).

| Parameter Year | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Random walk variance |  |  |  |  |  |
| -F Age 1-10 | 0.38 | 0.37 | 0.37 | 0.36 | 0.36 |
| Process error |  |  |  |  |  |
| $-\log (\mathrm{N})$ Age 1 | 0.62 | 0.61 | 0.60 | 0.61 | 0.62 |
| --- Age 2-10 | 0.18 | 0.18 | 0.18 | 0.18 | 0.18 |
| Observation variance |  |  |  |  |  |
| -Catch Age 1 | 0.44 | 0.43 | 0.44 | 0.44 | 0.43 |
| --- Age 2 | 0.28 | 0.28 | 0.28 | 0.28 | 0.27 |
| --- Age 3-8 | 0.19 | 0.19 | 0.19 | 0.19 | 0.18 |
| --- Age 9-10 | 0.40 | 0.39 | 0.38 | 0.38 | 0.37 |
| -IBWSS Age 1 | 0.74 | 0.75 | 0.74 | 0.72 | 0.74 |
| --- Age 2 | 0.31 | 0.33 | 0.33 | 0.33 | 0.33 |
| --- Age 3 | 0.42 | 0.41 | 0.41 | 0.39 | 0.39 |
| --- Age 4-6 | 0.39 | 0.37 | 0.37 | 0.36 | 0.35 |
| --- Age 7-8 | 0.50 | 0.54 | 0.54 | 0.53 | 0.53 |
| Survey catchability |  |  |  |  |  |
| -IBWSS Age 1 | 0.07 | 0.07 | 0.06 | 0.06 | 0.06 |
| --- Age 2 | 0.12 | 0.11 | 0.11 | 0.12 | 0.11 |
| --- Age 3 | 0.37 | 0.37 | 0.37 | 0.37 | 0.36 |
| --- Age 4 | 0.69 | 0.68 | 0.68 | 0.68 | 0.67 |
| --- Age 5-8 | 0.87 | 0.87 | 0.88 | 0.89 | 0.88 |
| Rho |  |  |  |  |  |
| -- | 0.93 | 0.93 | 0.93 | 0.94 | 0.93 |

Table 2.4.1.2. Blue whiting. Mohn's rho by year and average over the last five years ( $\mathrm{n}=5$ ).

| Last data year | R(age 1) | SSB | Fbar(3-7) |
| :--- | :--- | :--- | :--- |
| 2017 | -0.109 | -0.107 | 0.142 |
| 2018 | -0.218 | -0.102 | 0.074 |
| 2019 | -0.335 | -0.001 | -0.024 |
| 2020 | -0.216 | -0.076 | 0.077 |
| 2021 | -0.406 | -0.172 | 0.180 |
| Rho mean | -0.257 | -0.091 | 0.090 |

Table 2.4.1.3. Blue whiting. Estimated fishing mortalities. Catch data for 2022 are preliminary.

| Year Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 0.078 | 0.119 | 0.172 | 0.212 | 0.245 | 0.318 | 0.347 | 0.444 | 0.486 | 0.486 |
| 1982 | 0.067 | 0.102 | 0.149 | 0.183 | 0.209 | 0.271 | 0.294 | 0.372 | 0.405 | 0.405 |
| 1983 | 0.078 | 0.118 | 0.171 | 0.211 | 0.240 | 0.314 | 0.338 | 0.420 | 0.446 | 0.446 |
| 1984 | 0.095 | 0.143 | 0.212 | 0.265 | 0.305 | 0.397 | 0.418 | 0.509 | 0.530 | 0.530 |
| 1985 | 0.101 | 0.150 | 0.229 | 0.294 | 0.346 | 0.447 | 0.465 | 0.560 | 0.575 | 0.575 |
| 1986 | 0.113 | 0.168 | 0.268 | 0.357 | 0.431 | 0.552 | 0.573 | 0.692 | 0.705 | 0.705 |
| 1987 | 0.100 | 0.150 | 0.247 | 0.337 | 0.414 | 0.537 | 0.560 | 0.674 | 0.676 | 0.676 |
| 1988 | 0.098 | 0.148 | 0.253 | 0.349 | 0.439 | 0.575 | 0.589 | 0.694 | 0.678 | 0.678 |
| 1989 | 0.113 | 0.171 | 0.304 | 0.420 | 0.526 | 0.686 | 0.712 | 0.842 | 0.806 | 0.806 |
| 1990 | 0.105 | 0.159 | 0.292 | 0.408 | 0.511 | 0.665 | 0.713 | 0.850 | 0.817 | 0.817 |
| 1991 | 0.059 | 0.089 | 0.168 | 0.235 | 0.290 | 0.368 | 0.396 | 0.466 | 0.451 | 0.451 |
| 1992 | 0.048 | 0.073 | 0.140 | 0.196 | 0.233 | 0.286 | 0.311 | 0.370 | 0.363 | 0.363 |
| 1993 | 0.042 | 0.063 | 0.125 | 0.176 | 0.206 | 0.246 | 0.268 | 0.319 | 0.314 | 0.314 |
| 1994 | 0.036 | 0.054 | 0.113 | 0.160 | 0.186 | 0.219 | 0.241 | 0.292 | 0.286 | 0.286 |
| 1995 | 0.046 | 0.070 | 0.150 | 0.216 | 0.244 | 0.285 | 0.314 | 0.383 | 0.369 | 0.369 |
| 1996 | 0.055 | 0.085 | 0.185 | 0.271 | 0.297 | 0.348 | 0.383 | 0.473 | 0.451 | 0.451 |
| 1997 | 0.054 | 0.084 | 0.188 | 0.279 | 0.300 | 0.349 | 0.382 | 0.474 | 0.453 | 0.453 |
| 1998 | 0.070 | 0.110 | 0.251 | 0.382 | 0.408 | 0.473 | 0.510 | 0.630 | 0.593 | 0.593 |
| 1999 | 0.064 | 0.101 | 0.237 | 0.369 | 0.397 | 0.458 | 0.482 | 0.592 | 0.558 | 0.558 |


| Year Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 0.073 | 0.117 | 0.279 | 0.446 | 0.498 | 0.576 | 0.589 | 0.705 | 0.665 | 0.665 |
| 2001 | 0.069 | 0.111 | 0.265 | 0.430 | 0.494 | 0.572 | 0.574 | 0.679 | 0.644 | 0.644 |
| 2002 | 0.065 | 0.104 | 0.250 | 0.418 | 0.504 | 0.595 | 0.597 | 0.702 | 0.667 | 0.667 |
| 2003 | 0.067 | 0.107 | 0.261 | 0.440 | 0.545 | 0.634 | 0.629 | 0.710 | 0.670 | 0.670 |
| 2004 | 0.068 | 0.109 | 0.269 | 0.461 | 0.592 | 0.690 | 0.688 | 0.753 | 0.711 | 0.711 |
| 2005 | 0.059 | 0.095 | 0.238 | 0.419 | 0.557 | 0.650 | 0.656 | 0.704 | 0.667 | 0.667 |
| 2006 | 0.051 | 0.082 | 0.208 | 0.372 | 0.509 | 0.596 | 0.606 | 0.640 | 0.606 | 0.606 |
| 2007 | 0.048 | 0.077 | 0.197 | 0.357 | 0.506 | 0.604 | 0.629 | 0.661 | 0.629 | 0.629 |
| 2008 | 0.041 | 0.068 | 0.170 | 0.309 | 0.444 | 0.529 | 0.563 | 0.590 | 0.569 | 0.569 |
| 2009 | 0.027 | 0.045 | 0.112 | 0.198 | 0.288 | 0.342 | 0.371 | 0.386 | 0.375 | 0.375 |
| 2010 | 0.019 | 0.032 | 0.080 | 0.138 | 0.201 | 0.236 | 0.259 | 0.264 | 0.257 | 0.257 |
| 2011 | 0.006 | 0.010 | 0.024 | 0.040 | 0.057 | 0.066 | 0.073 | 0.075 | 0.074 | 0.074 |
| 2012 | 0.012 | 0.020 | 0.051 | 0.085 | 0.121 | 0.141 | 0.159 | 0.166 | 0.165 | 0.165 |
| 2013 | 0.019 | 0.034 | 0.090 | 0.150 | 0.214 | 0.244 | 0.278 | 0.292 | 0.292 | 0.292 |
| 2014 | 0.036 | 0.066 | 0.175 | 0.295 | 0.416 | 0.472 | 0.538 | 0.569 | 0.566 | 0.566 |
| 2015 | 0.046 | 0.085 | 0.230 | 0.389 | 0.546 | 0.624 | 0.698 | 0.734 | 0.727 | 0.727 |
| 2016 | 0.039 | 0.073 | 0.197 | 0.339 | 0.478 | 0.554 | 0.616 | 0.644 | 0.637 | 0.637 |
| 2017 | 0.037 | 0.069 | 0.188 | 0.326 | 0.457 | 0.527 | 0.577 | 0.595 | 0.590 | 0.590 |
| 2018 | 0.036 | 0.068 | 0.187 | 0.328 | 0.463 | 0.534 | 0.584 | 0.597 | 0.594 | 0.594 |
| 2019 | 0.032 | 0.060 | 0.168 | 0.298 | 0.421 | 0.482 | 0.527 | 0.531 | 0.528 | 0.528 |
| 2020 | 0.035 | 0.066 | 0.182 | 0.326 | 0.463 | 0.527 | 0.577 | 0.578 | 0.572 | 0.572 |
| 2021 | 0.030 | 0.056 | 0.156 | 0.281 | 0.399 | 0.453 | 0.494 | 0.497 | 0.491 | 0.491 |
| 2022 | 0.031 | 0.058 | 0.161 | 0.292 | 0.417 | 0.470 | 0.516 | 0.518 | 0.511 | 0.511 |

Table 2.4.1.4. Blue whiting. Estimated stock numbers-at-age (thousands). Preliminary catch data for 2022 have been used

| Year Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 3948198 | 3483134 | 4853128 | 2063381 | 2615682 | 2138821 | 1642831 | 1743162 | 1227396 | 2980519 |
| 1982 | 4696698 | 2963754 | 2517522 | 3288737 | 1583208 | 1494152 | 1291796 | 1012284 | 889662 | 1940890 |
| 1983 | 18293953 | 3809208 | 1877680 | 1818746 | 1897666 | 1218555 | 1014568 | 854628 | 627992 | 1252407 |
| 1984 | 18077398 | 14562962 | 2450003 | 1233914 | 1261344 | 1397475 | 815593 | 549720 | 481649 | 922153 |
| 1985 | 9550303 | 13545071 | 9807290 | 1453542 | 749637 | 914081 | 747079 | 458534 | 265294 | 722100 |
| 1986 | 7206799 | 6372536 | 9406612 | 5561299 | 948825 | 452061 | 469404 | 376188 | 231523 | 499543 |
| 1987 | 9113538 | 5032490 | 4072187 | 6882719 | 2568177 | 394492 | 253898 | 238137 | 156693 | 293640 |
| 1988 | 6409993 | 6861313 | 3511774 | 2870746 | 3727659 | 1275834 | 199445 | 125573 | 99146 | 170141 |
| 1989 | 8492388 | 4620755 | 4994215 | 2424184 | 2130302 | 1684904 | 350755 | 103076 | 60787 | 115044 |
| 1990 | 18840757 | 5973402 | 3093924 | 2728803 | 1480249 | 1186791 | 560115 | 120836 | 33119 | 85695 |
| 1991 | 9049081 | 15643199 | 4259229 | 1785365 | 1491130 | 874989 | 562933 | 188115 | 32126 | 45442 |
| 1992 | 6698667 | 7458376 | 12503692 | 3310414 | 1257597 | 788315 | 485873 | 287464 | 101477 | 39068 |
| 1993 | 4942776 | 5123283 | 5296885 | 9738212 | 2263702 | 976646 | 516955 | 281742 | 156944 | 74164 |
| 1994 | 8113390 | 3379619 | 4070890 | 3390595 | 6950171 | 1439175 | 766071 | 328519 | 207427 | 115713 |
| 1995 | 9322644 | 5867025 | 3129586 | 2563312 | 2856905 | 3741838 | 1041446 | 545237 | 221227 | 184619 |
| 1996 | 28090039 | 7109483 | 4071822 | 2392659 | 1544304 | 1861760 | 2238167 | 645981 | 306922 | 249290 |
| 1997 | 45080996 | 21344009 | 5501906 | 2566564 | 1414602 | 1064387 | 1059818 | 1211979 | 288024 | 336938 |


| Year Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 26696161 | 37985412 | 16454475 | 3496892 | 1370190 | 925363 | 782485 | 604708 | 616206 | 291964 |
| 1999 | 20303064 | 20503747 | 27713908 | 10599808 | 1703960 | 770894 | 518846 | 410948 | 236101 | 427358 |
| 2000 | 39448363 | 15270043 | 16580561 | 15810425 | 4338169 | 1111059 | 472766 | 323676 | 153198 | 313508 |
| 2001 | 55947436 | 31755217 | 12078805 | 10736413 | 7446097 | 1691153 | 488651 | 227410 | 163649 | 177810 |
| 2002 | 49106855 | 45338753 | 20408977 | 8304748 | 5441968 | 3392286 | 687547 | 255845 | 103097 | 154656 |
| 2003 | 52947963 | 39192643 | 35075975 | 13575701 | 5077023 | 2966716 | 1200986 | 344891 | 88649 | 106704 |
| 2004 | 28714768 | 42109802 | 30137706 | 20847617 | 7248626 | 2464658 | 1312509 | 500226 | 151135 | 80113 |
| 2005 | 22271163 | 21718324 | 28499051 | 18145779 | 10750885 | 3223618 | 1108963 | 512665 | 191117 | 98337 |
| 2006 | 9009113 | 15413359 | 22321756 | 19293696 | 9481605 | 4454118 | 1354232 | 481870 | 217274 | 119414 |
| 2007 | 4913368 | 5967442 | 13108559 | 15933357 | 10301052 | 4703381 | 1836143 | 608879 | 228295 | 162451 |
| 2008 | 5883847 | 3465837 | 4331922 | 11056361 | 9160778 | 4912569 | 1855148 | 753418 | 234404 | 199862 |
| 2009 | 5813482 | 4042546 | 2418910 | 3703135 | 6937430 | 4708636 | 2192796 | 855782 | 323765 | 188156 |
| 2010 | 15534228 | 5108247 | 2365111 | 1853628 | 3372665 | 4350095 | 2847425 | 1199051 | 411438 | 264795 |
| 2011 | 19713440 | 13522085 | 3355946 | 1661963 | 1613842 | 2613131 | 2697703 | 1346774 | 811348 | 389208 |
| 2012 | 19509226 | 15674187 | 12781899 | 2296933 | 1185528 | 1619913 | 2344919 | 2125776 | 1077665 | 896461 |
| 2013 | 16196200 | 16181915 | 11820392 | 7461025 | 2245751 | 1097606 | 1384976 | 1645697 | 1349228 | 1380746 |
| 2014 | 37769539 | 12759745 | 14030071 | 8116097 | 4403727 | 1351921 | 942553 | 1011060 | 1026225 | 1492297 |
| 2015 | 64728113 | 33248750 | 10910324 | 8533489 | 4211336 | 1736364 | 738558 | 523308 | 487940 | 1058255 |


| Year Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 | 35567778 | 57871493 | 21657974 | 7737276 | 4336329 | 1809839 | 707922 | 354459 | 223199 | 594127 |
| 2017 | 12172965 | 28851639 | 46306625 | 15275299 | 4577931 | 2166987 | 738670 | 285246 | 162222 | 375993 |
| 2018 | 13119298 | 9414030 | 22960923 | 30015459 | 8861786 | 2498004 | 952440 | 316812 | 143809 | 266442 |
| 2019 | 15254049 | 9759745 | 9037146 | 15384432 | 16762022 | 4668475 | 1155173 | 412610 | 140328 | 199051 |
| 2020 | 26772174 | 12754712 | 7294693 | 7026232 | 8955708 | 8254837 | 2249921 | 583463 | 201500 | 163929 |
| 2021 | 71562826 | 19445326 | 9940991 | 5134725 | 4644904 | 4294899 | 4043115 | 878041 | 288811 | 172055 |
| 2022 | 43220294 | 65372551 | 13900332 | 7564442 | 3356125 | 2474668 | 2082227 | 2211891 | 397414 | 249106 |
| 2023 | 22537250* | 34317672 | 50514113 | 9683829 | 4624746 | 1811531 | 1265829 | 1018083 | 1078587 | 317446 |

*assuming GM(1996-2021) recruitment in 2023.

Table 2.4.1.5. Blue whiting. Estimated recruitment (R) in thousands, spawning-stock biomass (SSB) in tonnes, average fishing mortality for ages 3 to 7 (Fbar 3-7) and total-stock biomass (TSB) in tonnes. Preliminary catch data for 2022 are included. Low and High refer to the $\mathbf{9 5 \%}$ confidence limits

| Year | R(age 1) | Low | High | SSB | Low | High | Fbar | Low | High |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| (3-7) |  |  |  |  |  |  |  |  |  |


| Year | R(age 1) | Low | High | SSB | Low | High | Fbar <br> (3-7) | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 9113538 | 6047121 | 13734894 | 1933155 | 1603976 | 2329890 | 0.419 | 0.324 | 0.541 | 2817509 | 2341994 | 3389573 |
| 1988 | 6409993 | 4251396 | 9664593 | 1638205 | 1370653 | 1957983 | 0.441 | 0.342 | 0.569 | 2425857 | 2024649 | 2906570 |
| 1989 | 8492388 | 5609692 | 12856438 | 1546404 | 1297863 | 1842541 | 0.529 | 0.412 | 0.680 | 2391067 | 1985764 | 2879096 |
| 1990 | 18840757 | 12270809 | 28928337 | 1356528 | 1128145 | 1631147 | 0.518 | 0.396 | 0.676 | 2499552 | 2000264 | 3123468 |
| 1991 | 9049081 | 5826489 | 14054068 | 1777830 | 1429325 | 2211309 | 0.291 | 0.216 | 0.393 | 3224253 | 2529151 | 4110395 |
| 1992 | 6698667 | 4365051 | 10279865 | 2460102 | 1951887 | 3100643 | 0.233 | 0.173 | 0.315 | 3532097 | 2804852 | 4447902 |
| 1993 | 4942776 | 3184796 | 7671146 | 2543671 | 2027313 | 3191545 | 0.204 | 0.151 | 0.275 | 3420359 | 2744348 | 4262893 |
| 1994 | 8113390 | 5277115 | 12474069 | 2535468 | 2042807 | 3146943 | 0.184 | 0.136 | 0.248 | 3415435 | 2776614 | 4201231 |
| 1995 | 9322644 | 6125353 | 14188844 | 2308211 | 1901995 | 2801183 | 0.242 | 0.183 | 0.320 | 3354448 | 2764391 | 4070452 |
| 1996 | 28090039 | 18496998 | 42658289 | 2207682 | 1836837 | 2653398 | 0.297 | 0.226 | 0.390 | 3726979 | 3035570 | 4575870 |
| 1997 | 45080996 | 29748090 | 68316864 | 2467358 | 2048273 | 2972189 | 0.300 | 0.229 | 0.392 | 5448075 | 4288051 | 6921914 |
| 1998 | 26696161 | 17717886 | 40224044 | 3685247 | 3017327 | 4501018 | 0.405 | 0.313 | 0.524 | 6834539 | 5472054 | 8536268 |
| 1999 | 20303064 | 13419599 | 30717343 | 4449851 | 3630259 | 5454480 | 0.389 | 0.300 | 0.504 | 7178200 | 5844370 | 8816443 |
| 2000 | 39448363 | 26002313 | 59847495 | 4233012 | 3521792 | 5087862 | 0.477 | 0.372 | 0.613 | 7470574 | 6096978 | 9153629 |
| 2001 | 55947436 | 37187292 | 84171646 | 4575712 | 3821738 | 5478433 | 0.467 | 0.363 | 0.600 | 9021213 | 7286025 | 11169642 |
| 2002 | 49106855 | 32631938 | 73899478 | 5401566 | 4504809 | 6476838 | 0.473 | 0.367 | 0.609 | 10346234 | 8388365 | 12761075 |
| 2003 | 52947963 | 35655989 | 78625974 | 6875173 | 5716742 | 8268346 | 0.502 | 0.395 | 0.638 | 11858110 | 9739057 | 14438233 |
| 2004 | 28714768 | 19302031 | 42717673 | 6778079 | 5701925 | 8057341 | 0.540 | 0.428 | 0.682 | 10401328 | 8701897 | 12432648 |


| Year | R (age 1) | Low | High | SSB | Low | High | Fbar (3-7) | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 22271163 | 14998683 | 33069881 | 6031586 | 5083744 | 7156150 | 0.504 | 0.396 | 0.641 | 8508967 | 7156398 | 10117173 |
| 2006 | 9009113 | 6005043 | 13515995 | 5891687 | 4948048 | 7015288 | 0.458 | 0.358 | 0.587 | 7732632 | 6496994 | 9203272 |
| 2007 | 4913368 | 3264414 | 7395256 | 4673775 | 3915203 | 5579322 | 0.459 | 0.355 | 0.593 | 5709625 | 4792945 | 6801625 |
| 2008 | 5883847 | 3862867 | 8962167 | 3593824 | 2971047 | 4347144 | 0.403 | 0.303 | 0.535 | 4414990 | 3665381 | 5317902 |
| 2009 | 5813482 | 3702391 | 9128311 | 2754203 | 2222003 | 3413874 | 0.262 | 0.192 | 0.357 | 3474921 | 2822816 | 4277670 |
| 2010 | 15534228 | 10141289 | 23795027 | 2690487 | 2129607 | 3399087 | 0.183 | 0.131 | 0.255 | 3776135 | 3014174 | 4730713 |
| 2011 | 19713440 | 12986631 | 29924598 | 2717639 | 2166277 | 3409333 | 0.052 | 0.036 | 0.075 | 4477901 | 3566528 | 5622163 |
| 2012 | 19509226 | 13074672 | 29110475 | 3476370 | 2840438 | 4254677 | 0.112 | 0.084 | 0.149 | 5176194 | 4222226 | 6345701 |
| 2013 | 16196200 | 10896851 | 24072726 | 3803134 | 3168694 | 4564602 | 0.195 | 0.149 | 0.255 | 5642058 | 4679857 | 6802092 |
| 2014 | 37769539 | 25189811 | 56631551 | 4045006 | 3409524 | 4798933 | 0.379 | 0.292 | 0.492 | 6710627 | 5543292 | 8123787 |
| 2015 | 64728113 | 43428474 | 96474232 | 4218197 | 3548784 | 5013883 | 0.497 | 0.389 | 0.635 | 8265505 | 6686427 | 10217500 |
| 2016 | 35567778 | 23913789 | 52901145 | 4974200 | 4114630 | 6013339 | 0.437 | 0.339 | 0.562 | 9254134 | 7480583 | 11448171 |
| 2017 | 12172965 | 8062125 | 18379902 | 6199504 | 5085700 | 7557239 | 0.415 | 0.322 | 0.536 | 8988316 | 7357708 | 10980297 |
| 2018 | 13119298 | 8655930 | 19884169 | 6090429 | 5006887 | 7408460 | 0.419 | 0.322 | 0.546 | 8082094 | 6653535 | 9817373 |
| 2019 | 15254049 | 9638265 | 24141898 | 5284355 | 4319920 | 6464102 | 0.379 | 0.285 | 0.505 | 7305759 | 5945587 | 8977098 |
| 2020 | 26772174 | 16299607 | 43973408 | 4480563 | 3571612 | 5620836 | 0.415 | 0.302 | 0.571 | 7042419 | 5456145 | 9089872 |
| 2021 | 71562826 | 40213851 | 127350103 | 4440379 | 3320100 | 5938666 | 0.356 | 0.243 | 0.523 | 9511299 | 6526756 | 13860608 |
| 2022 | 43220294 | 19011128 | 98257917 | 4955777 | 3341999 | 7348815 | 0.371 | 0.228 | 0.605 | 9506755 | 6004619 | 15051478 |


| Year | R(age 1) | Low | High | SSB | Low | High | Fbar (3-7) | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2023 | 22537250* |  |  | 6621 |  |  |  |  |  |  |  |  |

*assuming GM(1996-2021) recruitment in 2023.
$\wedge$ SSB calculated from the survivors age 2-10 and GM(1996-2021) recruitment in 2023

Table 2.4.1.5. Blue whiting. Estimated recruitment (R) in thousands, spawning-stock biomass (SSB) in tonnes, average fishing mortality for ages 3 to 7 (Fbar 3-7) and total-stock biomass (TSB) in tonnes. Preliminary catch data for 2022 are included. Low and High refer to the $\mathbf{9 5 \%}$ confidence limits

| Year | R(age 1) | Low | High | SSB | Low | High | Fbar | Low |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| (3-7) |  |  |  |  |  |  |  |  |


| Year | R(age 1) | Low | High | SSB | Low | High | Fbar <br> (3-7) | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 4942776 | 3184796 | 7671146 | 2543671 | 2027313 | 3191545 | 0.204 | 0.151 | 0.275 | 3420359 | 2744348 | 4262893 |
| 1994 | 8113390 | 5277115 | 12474069 | 2535468 | 2042807 | 3146943 | 0.184 | 0.136 | 0.248 | 3415435 | 2776614 | 4201231 |
| 1995 | 9322644 | 6125353 | 14188844 | 2308211 | 1901995 | 2801183 | 0.242 | 0.183 | 0.320 | 3354448 | 2764391 | 4070452 |
| 1996 | 28090039 | 18496998 | 42658289 | 2207682 | 1836837 | 2653398 | 0.297 | 0.226 | 0.390 | 3726979 | 3035570 | 4575870 |
| 1997 | 45080996 | 29748090 | 68316864 | 2467358 | 2048273 | 2972189 | 0.300 | 0.229 | 0.392 | 5448075 | 4288051 | 6921914 |
| 1998 | 26696161 | 17717886 | 40224044 | 3685247 | 3017327 | 4501018 | 0.405 | 0.313 | 0.524 | 6834539 | 5472054 | 8536268 |
| 1999 | 20303064 | 13419599 | 30717343 | 4449851 | 3630259 | 5454480 | 0.389 | 0.300 | 0.504 | 7178200 | 5844370 | 8816443 |
| 2000 | 39448363 | 26002313 | 59847495 | 4233012 | 3521792 | 5087862 | 0.477 | 0.372 | 0.613 | 7470574 | 6096978 | 9153629 |
| 2001 | 55947436 | 37187292 | 84171646 | 4575712 | 3821738 | 5478433 | 0.467 | 0.363 | 0.600 | 9021213 | 7286025 | 11169642 |
| 2002 | 49106855 | 32631938 | 73899478 | 5401566 | 4504809 | 6476838 | 0.473 | 0.367 | 0.609 | 10346234 | 8388365 | 12761075 |
| 2003 | 52947963 | 35655989 | 78625974 | 6875173 | 5716742 | 8268346 | 0.502 | 0.395 | 0.638 | 11858110 | 9739057 | 14438233 |
| 2004 | 28714768 | 19302031 | 42717673 | 6778079 | 5701925 | 8057341 | 0.540 | 0.428 | 0.682 | 10401328 | 8701897 | 12432648 |
| 2005 | 22271163 | 14998683 | 33069881 | 6031586 | 5083744 | 7156150 | 0.504 | 0.396 | 0.641 | 8508967 | 7156398 | 10117173 |
| 2006 | 9009113 | 6005043 | 13515995 | 5891687 | 4948048 | 7015288 | 0.458 | 0.358 | 0.587 | 7732632 | 6496994 | 9203272 |
| 2007 | 4913368 | 3264414 | 7395256 | 4673775 | 3915203 | 5579322 | 0.459 | 0.355 | 0.593 | 5709625 | 4792945 | 6801625 |
| 2008 | 5883847 | 3862867 | 8962167 | 3593824 | 2971047 | 4347144 | 0.403 | 0.303 | 0.535 | 4414990 | 3665381 | 5317902 |
| 2009 | 5813482 | 3702391 | 9128311 | 2754203 | 2222003 | 3413874 | 0.262 | 0.192 | 0.357 | 3474921 | 2822816 | 4277670 |
| 2010 | 15534228 | 10141289 | 23795027 | 2690487 | 2129607 | 3399087 | 0.183 | 0.131 | 0.255 | 3776135 | 3014174 | 4730713 |


| Year | R(age 1) | Low | High | SSB | Low | High | $\begin{aligned} & \text { Fbar } \\ & (3-7) \end{aligned}$ | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 19713440 | 12986631 | 29924598 | 2717639 | 2166277 | 3409333 | 0.052 | 0.036 | 0.075 | 4477901 | 3566528 | 5622163 |
| 2012 | 19509226 | 13074672 | 29110475 | 3476370 | 2840438 | 4254677 | 0.112 | 0.084 | 0.149 | 5176194 | 4222226 | 6345701 |
| 2013 | 16196200 | 10896851 | 24072726 | 3803134 | 3168694 | 4564602 | 0.195 | 0.149 | 0.255 | 5642058 | 4679857 | 6802092 |
| 2014 | 37769539 | 25189811 | 56631551 | 4045006 | 3409524 | 4798933 | 0.379 | 0.292 | 0.492 | 6710627 | 5543292 | 8123787 |
| 2015 | 64728113 | 43428474 | 96474232 | 4218197 | 3548784 | 5013883 | 0.497 | 0.389 | 0.635 | 8265505 | 6686427 | 10217500 |
| 2016 | 35567778 | 23913789 | 52901145 | 4974200 | 4114630 | 6013339 | 0.437 | 0.339 | 0.562 | 9254134 | 7480583 | 11448171 |
| 2017 | 12172965 | 8062125 | 18379902 | 6199504 | 5085700 | 7557239 | 0.415 | 0.322 | 0.536 | 8988316 | 7357708 | 10980297 |
| 2018 | 13119298 | 8655930 | 19884169 | 6090429 | 5006887 | 7408460 | 0.419 | 0.322 | 0.546 | 8082094 | 6653535 | 9817373 |
| 2019 | 15254049 | 9638265 | 24141898 | 5284355 | 4319920 | 6464102 | 0.379 | 0.285 | 0.505 | 7305759 | 5945587 | 8977098 |
| 2020 | 26772174 | 16299607 | 43973408 | 4480563 | 3571612 | 5620836 | 0.415 | 0.302 | 0.571 | 7042419 | 5456145 | 9089872 |
| 2021 | 71562826 | 40213851 | 127350103 | 4440379 | 3320100 | 5938666 | 0.356 | 0.243 | 0.523 | 9511299 | 6526756 | 13860608 |
| 2022 | 43220294 | 19011128 | 98257917 | 4955777 | 3341999 | 7348815 | 0.371 | 0.228 | 0.605 | 9506755 | 6004619 | 15051478 |
| 2023 | 22537250* |  |  | 6621207^ |  |  |  |  |  |  |  |  |

*assuming GM(1996-2021) recruitment in 2023.
^ SSB calculated from the survivors age 2-10 and GM(1996-2021) recruitment in 2023

Table 2.4.6. Blue whiting. Model estimate of total catch weight (in tonnes) and Sum of Product of catch number and mean weight at age for ages 1-10+ (Observed catch). Preliminary catch data for 2022 are included.

| Year | Estimate | Low | High | SOP <br> catch |
| :---: | :---: | :---: | :---: | :---: |
| 1981 | 788899 | 568869 | 1094034 | 922980 |
| 1982 | 544462 | 415928 | 712717 | 550643 |
| 1983 | 511429 | 397097 | 658681 | 553344 |
| 1984 | 562849 | 436560 | 725671 | 615569 |
| 1985 | 638357 | 503485 | 809358 | 678214 |
| 1986 | 760657 | 600423 | 963653 | 847145 |
| 1987 | 638031 | 503867 | 807920 | 654718 |
| 1988 | 569671 | 450572 | 720253 | 552264 |
| 1989 | 618818 | 492656 | 777287 | 630316 |
| 1990 | 553618 | 437776 | 700115 | 558128 |
| 1991 | 407963 | 318674 | 522271 | 364008 |
| 1992 | 438562 | 347202 | 553963 | 474592 |
| 1993 | 440392 | 346971 | 558966 | 475198 |
| 1994 | 424782 | 332919 | 541992 | 457696 |
| 1995 | 508660 | 405109 | 638679 | 505176 |
| 1996 | 597489 | 475874 | 750185 | 621104 |
| 1997 | 640338 | 505643 | 810912 | 639681 |
| 1998 | 1080858 | 848979 | 1376068 | 1131955 |
| 1999 | 1248067 | 975731 | 1596416 | 1261033 |
| 2000 | 1503388 | 1184000 | 1908931 | 1412449 |
| 2001 | 1560119 | 1228373 | 1981459 | 1771805 |
| 2002 | 1711364 | 1347754 | 2173071 | 1556955 |
| 2003 | 2202490 | 1742780 | 2783462 | 2365319 |
| 2004 | 2319262 | 1842881 | 2918785 | 2400795 |
| 2005 | 1998969 | 1590932 | 2511659 | 2018344 |
| 2006 | 1854141 | 1475771 | 2329521 | 1956239 |
| 2007 | 1558186 | 1238596 | 1960237 | 1612269 |


| Year | Estimate | Low | High | SOP <br> catch |
| :---: | :---: | :---: | :---: | :---: |
| 2008 | 1167284 | 920948 | 1479510 | 1251851 |
| 2009 | 656865 | 517289 | 834101 | 634978 |
| 2010 | 478520 | 370983 | 617227 | 539539 |
| 2011 | 135746 | 100837 | 182742 | 103771 |
| 2012 | 326794 | 260010 | 410731 | 375692 |
| 2013 | 591003 | 469650 | 743713 | 613863 |
| 2014 | 1113687 | 879716 | 1409885 | 1147650 |
| 2015 | 1348361 | 1074286 | 1692359 | 1390656 |
| 2016 | 1243182 | 987399 | 1565224 | 1180786 |
| 2017 | 1481707 | 1175774 | 1867243 | 1555069 |
| 2018 | 1706765 | 1348261 | 2160595 | 1709856 |
| 2019 | 1535424 | 1211096 | 1946607 | 1512026 |
| 2020 | 1428866 | 1132542 | 1802721 | 1460507 |
| 2021 | 1147426 | 915250 | 1438500 | 1139531 |
| 2022 | 1121336 | 884135 | 1422175 | 1107529 |

Table 2.8.2.1.1. Blue whiting. Input to short-term projection (median values for exploitation pattern and stock numbers).

| Age | Mean weight in <br> the stock and catch <br> (kg) in 2022 | Mean weight in the <br> stock and catch (kg) <br> in 2023+ | Proportion <br> mature | Natural <br> mortality | Exploitation <br> pattern | Stock number <br> (2023) <br> (thousands) |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 0.042 | 0.055 | 0.11 | 0.20 | 0.083 | 22537250 |
| 2 | 0.065 | 0.079 | 0.40 | 0.20 | 0.156 | 34317672 |
| 3 | 0.084 | 0.094 | 0.82 | 0.20 | 0.435 | 50514113 |
| 4 | 0.104 | 0.113 | 0.91 | 0.20 | 0.787 | 9683829 |
| 5 | 0.119 | 0.136 | 0.138 | 1.94 | 0.20 | 1.1267 |
| 7 | 0.139 | 0.145 | 1.00 | 0.20 | 1.389 | 4624746 |
| 8 | 0.158 | 0.154 | 0.170 | 1.00 | 0.20 | 1.396 |

Table 2.8.2.1.2. Blue whiting. Deterministic forecast, intermediate year assumptions and recruitments.

| Variable | Value | Notes |
| :---: | :---: | :---: |
| $\mathrm{F}_{\text {ages 3-7 }}$ (2022) | 0.37 | From the assessment (based on assumed 2022 catches) |
| SSB (2023) | 6621207 | From the forecast; in tonnes |
| Rage 1 (2022) $^{\text {(2) }}$ | 43220294 | From the assessment; in thousands |
| $\mathrm{R}_{\text {age } 1}$ (2023-2024) | 22537250 | GM (1996-2021); in thousands |
| Total catch (2022) | 1107529 | As estimated by ICES, based on declared national quotas and expected uptake; in tonnes |

Table 2.8.2.2.1. Blue whiting. Deterministic forecast (weights in tonnes).

| Basis | Catch(2023) | F(2023) | SSB(2024) | \% SSB <br> change* | \% Catch change** | \% Advice change*** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Long-term management strategy ( $\mathrm{F}=\mathrm{FMSY}$ ) | 1359629 | 0.320 | 7781444 | 17.5 | 22.8 | 80.6 |
| MSY approach: FMSY | 1359629 | 0.320 | 7781444 | 17.5 | 22.8 | 80.6 |
| $\mathrm{F}=0$ | 5 | 0.000 | 9039585 | 36.5 | -100.0 | -100.0 |
| Fpa | 1359629 | 0.320 | 7781444 | 17.5 | 22.8 | 80.6 |
| Flim | 3146002 | 0.880 | 6157129 | -7.0 | 184.1 | 317.9 |
| SSB (2024) = Blim | 8696303 | 6.503 | 1499996 | -77.3 | 685.2 | 1055.3 |
| SSB (2024 = Bpa | 7715688 | 4.401 | 2249993 | -66.0 | 596.7 | 925.0 |
| SSB (2024) = MSY Btrigger | 7715688 | 4.401 | 2249993 | -66.0 | 596.7 | 925.0 |
| $F=F(2022)$ | 1550784 | 0.371 | 7605942 | 14.9 | 40.0 | 106.0 |
| SSB (2024) = SSB (2023) | 2631402 | 0.698 | 6621196 | -0.0 | 137.6 | 249.6 |
| Catch (2023) = Catch (2022) | 1107553 | 0.255 | 8013430 | 21.0 | 0.0 | 47.1 |
| Catch (2023) = Catch (2022) -20\% | 886105 | 0.200 | 8217731 | 24.1 | -20.0 | 17.7 |
| Catch (2023) = Catch (2022) +25\% | 1384385 | 0.327 | 7758694 | 17.2 | 25.0 | 83.9 |
| Catch (2023) = Advice (2022) -20\% | 602183 | 0.133 | 8480325 | 28.1 | -45.6 | -20.0 |
| Catch (2023) = Advice (2022) +25\% | 940871 | 0.214 | 8167163 | 23.3 | -15.0 | 25.0 |
| $F=0.05$ | 233147 | 0.050 | 8822699 | 33.2 | -78.9 | -69.0 |
| $F=0.10$ | 458089 | 0.100 | 8613869 | 30.1 | -58.6 | -39.1 |
| $F=0.15$ | 675211 | 0.150 | 8412714 | 27.1 | -39.0 | -10.3 |
| $F=0.16$ | 717729 | 0.160 | 8373372 | 26.5 | -35.2 | -4.7 |
| $F=0.17$ | 759951 | 0.170 | 8334319 | 25.9 | -31.4 | 1.0 |


| Basis | Catch(2023) | F(2023) | SSB(2024) | $\% \text { SSB }$ <br> change* | \% Catch change** | \% Advice change*** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $F=0.18$ | 801880 | 0.180 | 8295552 | 25.3 | -27.6 | 6.5 |
| $F=0.19$ | 843520 | 0.190 | 8257070 | 24.7 | -23.8 | 12.1 |
| $F=0.20$ | 884872 | 0.200 | 8218869 | 24.1 | -20.1 | 17.6 |
| $F=0.21$ | 925939 | 0.210 | 8180947 | 23.6 | -16.4 | 23.0 |
| $F=0.22$ | 966725 | 0.220 | 8143300 | 23.0 | -12.7 | 28.4 |
| $F=0.23$ | 1007231 | 0.230 | 8105927 | 22.4 | -9.1 | 33.8 |
| $F=0.24$ | 1047460 | 0.240 | 8068825 | 21.9 | -5.4 | 39.2 |
| $F=0.25$ | 1087415 | 0.250 | 8031990 | 21.3 | -1.8 | 44.5 |
| $F=0.26$ | 1127098 | 0.260 | 7995421 | 20.8 | 1.8 | 49.7 |
| $F=0.27$ | 1166512 | 0.270 | 7959115 | 20.2 | 5.3 | 55.0 |
| $F=0.28$ | 1205659 | 0.280 | 7923070 | 19.7 | 8.9 | 60.2 |
| $F=0.29$ | 1244542 | 0.290 | 7887282 | 19.1 | 12.4 | 65.3 |
| $F=0.30$ | 1283163 | 0.300 | 7851750 | 18.6 | 15.9 | 70.5 |
| $F=0.31$ | 1321525 | 0.310 | 7816471 | 18.1 | 19.3 | 75.6 |
| $F=0.32$ | 1359629 | 0.320 | 7781444 | 17.5 | 22.8 | 80.6 |
| $F=0.33$ | 1397479 | 0.330 | 7746664 | 17.0 | 26.2 | 85.7 |
| $F=0.34$ | 1435076 | 0.340 | 7712131 | 16.5 | 29.6 | 90.6 |
| $F=0.35$ | 1472423 | 0.350 | 7677841 | 16.0 | 32.9 | 95.6 |
| $F=0.45$ | 1832624 | 0.450 | 7347865 | 11.0 | 65.5 | 143.5 |
| $F=0.50$ | 2004100 | 0.500 | 7191261 | 8.6 | 81.0 | 166.2 |

* SSB 2024 relative to SSB 2023.
** Catch 2023 relative to expected catch in 2022 (1 107529 tonnes).
*** Catch 2023 relative to advice for 2022 ( 752736 tonnes).


### 2.18 Figures



Figure 2.2.1. Blue whiting catches in 2021. Catch data from Russia are not available and the catches on the map constitute $\mathbf{8 6} \%$ of the ICES estimated catches. The $\mathbf{2 0 0} \mathrm{m}$ and 1000 m depth contours are indicated in blue.


Figure 2.2.2. Blue whiting catches per quarter 2021. Catch data from Russia are not available and the catches on the maps constitute $86 \%$ of the ICES estimated catches and thus, the total catches and percentages shown on each panel might deviate slightly from the ICES estimated catches pr. quarter. The $\mathbf{2 0 0} \mathbf{~ m}$ and 1000 m depth contours are indicated in blue.


Figure 2.3.1.1. Blue whiting. ICES estimated catches (' 1000 tonnes) in 2021 by ICES division and country. Note: Russia 2021 catch data is preliminary and only for quarters 1 and 2, submitted to WGWIDE 2021.


B


Figure 2.3.1.2. Blue whiting.(A) ICES estimated catches (tonnes) of blue whiting by fishery subareas from 1988-2021 and $(B)$ the percentage contribution to the overall catch by fishery subarea over the same period.


Figure 2.3.1.3. Blue whiting. Distribution of 2021 ICES estimated catches (in percentage) by quarter.


Figure 2.3.1.4. Blue whiting. Distribution of 2021 ICES estimated catches (in percentage) by ICES division area.


Figure 2.3.1.5. Blue whiting. ICES estimated catches (' 1000 tonnes) in 2021 by country. Note: Russia 2021 catch data is preliminary and only for quarters 1 and 2, submitted to WGWIDE 2021.


Figure 2.3.1.6. Blue whiting. Distribution of 2021 ICES estimated catches (' 1000 tonnes) by ICES division and by quarter.


Figure 2.3.1.7. Blue whiting. Catch-at-age numbers (CANUM) distribution by quarter and ICES division for 2021.


Figure 2.3.1.1.1. Blue whiting. 2021 ICES catches (' 1000 tonnes) based on sampled or estimated distribution by ICES division.


Figure 2.3.1.2.1. Blue whiting. Mean length ( mm ) by age ( $0-10$ year), by quarter ( $\mathbf{1 , 2 , 4 \text { ), by country for ICES division area }}$ 27.6.a. These data only comprises the 2021 ICES catch-at-age sampled estimates for ICES division 27.6.a.


Figure 2.3.2.1. Blue whiting. 2022 ICES preliminary catches (' 1000 tonnes) (Quarter 1 + Quarter 2) based on sampled or estimated distribution by ICES division.


Figure 2.3.2.2 Preliminary and final estimates of catch at age number by age and year.


Figure 2.3.3.1. Blue whiting. Catch proportion at age, 1981-2021. Preliminary values for 2022 have been used.


Figure 2.3.3.2. Blue whiting. Age disaggregated catch (numbers) plotted on log scale. The labels for each panel indicate year classes. The grey dotted lines correspond to $Z=0.6$. Preliminary catch-at-age data for 2022 have been used.


Figure 2.3.4.1. Blue whiting. Mean catch (and stock) weight ( kg ) at age by year. Preliminary values for 2022 have been used


Figure 2.3.7.1.1. Blue whiting. (A) Estimate of total biomass from the International blue whiting spawning stock survey. The black dots and error bands are StoX estimates with $90 \%$ confidence intervals. (B) Internal consistency within the International blue whiting spawning stock survey. The upper left part of the plots shows the relationship between log index-at-age within a cohort. Linear regression line shows the best fit to the log-transformed indices. The lower-right part of the plots shows the correlation coefficient ( $r$ ) for the two ages plotted in that panel. The background colour of each panel is determined by the $r$ value, where red equates to $r=1$ and white to $r<0$.


## NO SURVEY

2019
2020


2021

Figure 2.3.7.1.2. Blue whiting. Distribution of the blue whiting stock in the area to the west of the British Isles, spring 2019 (upper panel) to 2022 (lower panel).

| 2022 |   |
| :---: | :---: |
| 2021 |   |
| 2020 | NO SURVEY |
| 2019 |  |
| 2018 |  |

Figure 2.3.7.1.3. Blue whiting. Length (line) and age (bars) distribution of the blue whiting stock in the area to the west of the British Isles, spring 2018 (lower panel) to 2022 (upper panel). Spawning-stock biomass and numbers are given.


Figure 2.4.1.1. Blue Whiting. OSA (One Step Ahead) residuals (see Berg and Nielsen, 2016) from catch-at-age and the IBWSS survey 2004-2022 (no survey in 2020). Red (lighter) bubbles show that the observed value is less than the expected value. Preliminary catch data for 2022 have been used.


Figure 2.4.1.2 Blue whiting. Joint sample residuals (Process errors) for stock number and F at age. Red (lighter) bubbles show that the observed value is less than the expected value. Preliminary catch data for 2022 have been used.


Figure 2.4.1.3. Blue whiting. Process errors expressed as deviation in instantaneous mortality at age by age and year.

## Residual catch



IBWSS


Figure 2.4.1.4. Blue whiting. The correlation matrix between ages for the catches and survey indices. Each ellipse represents the level curve of a bivariate normal distribution with the corresponding correlation. Hence, the sign of a correlation corresponds to the sign of the slope of the major ellipse axis. Increasingly darker shading is used for increasingly larger absolute correlations, while uncorrelated pairs of ages are depicted as circles with no shading. Preliminary catch data for 2022 have been used.


Figure 2.4.1.5. Blue whiting. Exploitation pattern by 5 -years' time blocks. Preliminary catch data for 2022 have been used.


Figure 2.4.1.6. Blue whiting. Retrospective analysis of recruitment (age 1), SSB (tonnes), F and total catch using the SAM model. The $95 \%$ confidence interval is shown for the most recent assessment.


Figure 2.4.1.7. Blue whiting. SAM final run: Stock summary, total catches, recruitment (age 1), F and SSB. The graphs show the median value and the $95 \%$ confidence interval. Catches for 2022 are preliminary.


Figure 2.4.1.8. Blue whiting. SAM final run: Comparison of the 2021 and 2022 stock assessments, shown with $95 \%$ confidence intervals. Catches for 2022 are preliminary.


Figure 2.4.1.9. Blue whiting. Comparison of assessment runs with the default configuration (wgwide2020: final 2021 data and preliminary 2022 catch data), a run with no catch information for 2022 ("no_prelim_2022") and a run with preliminary catch data for both 2021 and 2022 ("prelim_2020_2022").


Figure 2.4.2.1. Blue whiting. Comparison of SSB, F and recruitment estimated by the final WGWIDE 2022 SAM model and an alternative version including the two surveys IESNS and IESSNS. Catch values for $\mathbf{2 0 2 2}$ are preliminary.


Figure 2.4.2.2. Blue whiting. Historical retrospective F estimated by the final WGWIDE SAM model and the alternative version including the two surveys IESNS and IESSNS for 2021 and 2022 showing only the last 5 years. Catch values for both years and assessments are preliminary. The confidence intervals are from the respective final WGWIDE assessments.


Figure 2.8.1.1. Blue whiting young fish indices from five different surveys and recruitment index from the assessment, standardized by dividing each series by their mean. BarSea - Norwegian bottom-trawl survey in the Barents Sea, IESNS: International Ecosystem Survey in the Nordic Seas in May (1 and 2 is the age groups), IBWSS (Not updated in 2020): International Blue Whiting Spawning Stock survey (1 and $\mathbf{2}$ is the age groups), FO: the Faroese bottom-trawl surveys in spring, IS: the Icelandic bottom-trawl survey in spring, SAM: recruits from the assessment.

SSB (million t)


F (ages 3-7)


Rec (age 1; Billions)


Figure 2.9.1. Blue whiting. Comparison of the 2018-2022 assessments (historical retrospective).

## 3 Northeast Atlantic boarfish (Capros aper)

The boarfish (Capros aper, Linnaeus) is a deep bodied, laterally compressed, pelagic shoaling species distributed from Norway to Senegal, including the Mediterranean, Azores, Canaries, Madeira and Great Meteor Seamount (Blanchard \& Vandermeirsch 2005).
Boarfish is targeted in a pelagic trawl fishery for fish meal, to the south and southwest of Ireland and Northern Biscay. The boarfish fishery is conducted in shelf waters with the first landings reported in 2001. Landings were at very low levels from 2001-2005. The main expansion period of the fishery took place between 2006 and 2010 when unrestricted landings increased from 2772 t to 137503 t . A restrictive TAC of 33000 t was implemented in 2011. In 2011, ICES was asked by the European Commission to provide catch advice for 2012 for the first time.

An analysis of bottom trawl survey data suggests a continuity of distribution spanning ICES Subareas 27.4, 6, 7, 8 and 9 (Figure 3.1). Isolated occurrences appear in the North Sea (ICES Subarea 27.4) in some years indicating spill-over into this region. A hiatus in distribution was suggested between ICES Divisions 27.8.c and 9.a as boarfish were considered very rare in northern Portuguese waters but abundant further south (Cardador \& Chaves 2010). Results from a dedicated genetic study on the stock structure of boarfish within the Northeast Atlantic and Mediterranean Sea suggests that this hiatus represents a true stock separation (Farrell et al. (2016); see section 3.12). Based on these data, a single stock is considered to exist in ICES Subareas 27.4, 6, 7, 8 and the northern part of 9.a. This distribution is slightly broader than the current EC TAC area (27.6, 7 and 8 ) and for the purposes of assessment in 2022 only data from these areas were utilized.

### 3.1 The fishery

### 3.1.1 Advice and management applicable from 2011 to 2021

In 2011 a TAC was set for this species for the first time, covering ICES Subareas 6, 7 and 8 . This TAC was set at 33000 t . Before 2010, the fishery was unregulated. In October 2010, the European Commission notified national authorities that under the terms of Annex 1 of Regulation 850/1998, industrial fisheries for this species should not proceed with mesh sizes of less than 100 mm. In 2011, the European Parliament voted to change Regulation 850/1998 allowing the fishery to use mesh sizes ranging from 32 to 54 mm .

For 2012, ICES advised that catches of boarfish should not increase, based on precautionary considerations. As supporting information, ICES noted that it would be cautious that landings did not increase above 82000 t , the average over the period 2008-2010, during which the stock did not appear to be overexploited. In 2012 the TAC was set at 82000 t by the Council of the European Union.

For 2013, ICES advised that catches of boarfish should not be more than 82000 t . This was based on applying a harvest ratio of $12.2 \%$ (F0.1, as an FMSY proxy). For 2013, the TAC was set at 82000 $t$ by the Council of the European Union.
For 2014, ICES advised that, based on $\mathrm{Fmsy}_{\text {( }}$ (0.23), catches of boarfish should not be more than 133 957 t , or 127509 t when the average discard rate of the previous ten years is taken into account. For 2014 the TAC was set at 133957 t by the Council of the European Union. This advice was based on a Schaefer state space surplus production model (see section 3.6.3 for further details).

In 2014 there was concern about the use of the production model (see stock annex). ICES considered that the model was no longer suitable for providing category 1 advice and further model development was required. The model was still considered suitable for category 3 advice. The advised catch for 2015 of $53296 t$ was based on the data limited stock HCR and an index calculated (method 3.1; ICES, 2012) using the total stock biomass trends from the model.

For 2016 and 2017 ICES advised based on the precautionary approach that catches should be no more than 42637 t and 27288 t respectively. In 2017, the acoustic survey suggested that the stock abundance was at an historic low. The Advice Drafting Group decided the advice of 21830 proposed ( $20 \%$ reduction) would stand for 2 years. The update assessments in 2018 and 2019 confirmed that the biomass was stable and at a low level. In 2019, advice of 19152 t was issued for each of 2020 and 2021, once again on the basis of the precautionary approach. In 2021, with indications of an increase in recruitment, advice of 22791 t was given for 2022 and 2023 based on the precautionary approach.

Since 2011, there has been a provision for bycatch of boarfish (also whiting, haddock and mackerel) to be taken from the Western and North Sea horse mackerel EC quotas. These provisions are shown in the table below. The effect of this is that a quantity not exceeding the value of these 4 species combined may be landed legally and subtracted from quotas for horse mackerel.

| Year | North Sea (t) | Western (t) |
| :---: | :---: | :---: |
| 2011 | 2031 | 7779 |
| 2012 | 2148 | 7829 |
| 2013 | 1702 | 7799 |
| 2014 | 1392 | 5736 |
| 2015 | 583 | 4202 |
| 2016 | 760 | 5443 |
| 2017 | 912 | 4191 |
| 2018 | 759 | 5053 |
| 2019 | 759 | 5956 |
| 2020 | 688 | 3531 |
| 2021 | 701 | 3513 |
| 2022 | 173 | 2459 |

In 2010, an interim management plan was proposed by Ireland, which included a number of measures to mitigate potential bycatch of other TAC species in the boarfish fishery. A closed season from the 15th March to 31st August was proposed, as anecdotal evidence suggested that mackerel and boarfish are caught in mixed aggregations during this period. A closed season was proposed in ICES Division 7.g from 1st September to 31st October, in order to prevent catches of Celtic Sea herring, which is known to form feeding aggregations in this region at these times. Additionally, if catches of a species covered by a TAC, other than boarfish, amount to more than $5 \%$ of the total catch by day by ICES statistical rectangle, the management plan stipulates that fishing must cease in that rectangle for 5 days.

In August 2012 the Pelagic RAC proposed a long term management plan for boarfish. The management plan was not fully evaluated by ICES; however, in 2013 ICES advised that Tier 1 of the plan could be considered precautionary if a Category 1 assessment was available.

A revised draft management strategy was proposed by the Pelagic AC in July 2015. This management strategy aimed to achieve exploitation of boarfish in line with the precautionary approach to fisheries management, FAO guidelines for new and developing fisheries, and the ICES form of advice. ICES evaluated the plan and considered it to be precautionary, in that it followed the rationale for TAC setting enshrined in the ICES advice, but with additional caution.

The closed season, as detailed in the interim and revised management plans, has been enacted in legislation in Ireland, but not in other countries.

### 3.1.2 The fishery in recent years

In 2017 a total of 17388 t of boarfish were caught. Ireland was the main participant landing 15 484 t although landings were almost $20 \%$ below the 18858 t quota. Denmark landed only 548 t , (less than $10 \%$ of its national quota of 6696 t ). UK reported almost zero boarfish landings. Total reported discards accounted for 1173 t . Approximately $90 \%$ of the Irish landings were taken in ICES divisions-7.h and 8.a (Tables 3.1.2.5 and 3.1.2.7). 35 Irish registered fishing vessels reported catches with almost entirely from in Q1 (8570 t) and Q4 (6 270 t ).

In 2018 a total of 11286 t of boarfish were caught. This represented $55 \%$ of the 2018 quota of 20 380 t . Ireland continued to be the main participant landing 9513 t ( $68 \%$ of its national quota). The Irish catch represented $85 \%$ of the total boarfish catch in 2018. Other countries reporting boarfish in 2018 were Denmark ( 94 t ), The Netherlands (172 t), Spain (148t), UK England (0.085 $\mathrm{t})$ and UK Scotland $(0.229 \mathrm{t})$. Discards accounted for 1359 t . Tables 3.1.2.5 and 3.1.2.7 show that $82 \%$ of the Irish landings were taken in ICES divisions 7.h and 8.a.

A total of 11312 t of boarfish was caught in 2019. This represents $52 \%$ of the 2019 quota of 21830 t . The main participant in the fishery, Ireland, landed 9910 t ( $75 \%$ of its national quota). The Irish catch represents $88 \%$ of the total boarfish catch in 2019. Other countries reporting boarfish catches in 2019 were Denmark ( 757 t ), the Netherlands ( 317 t ), England ( 19 t ) and Spain ( 2.5 t ). Discards accounted for 306 t overall. Tables 3.1.2.5 and 3.1.2.7 shows that about $87 \%$ of Irish landings were taken in ICES divisions 7.h and 8.a.

In 2020, the total catch was 15649 t which represented $82 \%$ of the quota (19 152 t ). Ireland was the main participant in the fishery ( 14666 t ) and landed more than its national quota ( 13234 t ) for the first time since TAC and quota regulations were established. The Irish landings accounted for $94 \%$ of the total catch. The other countries reporting catches are Denmark ( 196 t ), the Netherlands (416 t), England ( 62 t ), Poland ( 109 t ) and Spain ( 1 t ). The total discards for this year were 198 t . The majority of landings were taken in ICES divisions 7.b and 7.h (Tables 3.1.2.4 and 3.1.2.5).

### 3.1.3 The fishery in 2021

In 2021, 17693 t of boarfish was caught, representing $92 \%$ of the total allowable catch (19 152 t ) for the year (Table 3.1.2.1). Ireland was the main contributor to the fishery landing $11830 \mathrm{t}, 89 \%$ of their quota ( 13234 t ). Other countries reporting landings for 2021 were Denmark ( 4322 t ), the Netherlands (781 t), England (45 t), Poland (45 t), Spain (11 t) and Scotland (9t). Total discards were 651 t . ICES divisions 7.j and 7.b had the highest landings of 10466 t and 3984 t respectively.

### 3.1.4 Regulations and their effects

In 2010, the fishery finished early when the European Commission notified member states that mesh sizes of less than 100 mm were illegal. However, in 2011, the European Parliament voted to change Regulation 850/1998 to allow fishing for boarfish using mesh sizes ranging from 32 to 54 mm . The TAC ( 33000 t ) that was introduced in 2011 significantly reduced landings.

### 3.1.5 Changes in fishing technology and fishing patterns

The expansion of the fishery in the mid-2000s was associated with developments in the pumping and processing technology for boarfish catches. In the past (2009-2012), the majority of boarfish landings by Danish, Irish and Scottish vessels have been into Skagen, Denmark and Fuglafjorour, Faroe Islands for fishmeal. In recent years, most landings are made into Ireland, although vessels monitor the price in different countries before making a decision on where to land.

In the past two years, the fishery has noticed an increase in the abundance of boarfish, in the southern Celtic Sea, where boarfish had been sparse for many years. There has also been a shift in distribution, with fishable marks being observed along the western Irish coast even up to $57^{\circ} \mathrm{N}$. In 2021, with the improvement in the stock and more vessels targeting boarfish, allocations were re-established with 85-90\% allocated for Q1 and the remainder for Q3 and Q4. The division in allocations may initiate a change in fishing pattern as boats may wait until Q3/4 to utilise their full allocation, especially if the quota is low (E. Farrell, pers comm).

### 3.1.6 Discards

It is to be expected that discarding occurred before 2003, particularly in demersal fisheries, however it is difficult to determine what the levels may have been.

Since 2003, the major sources of discard estimates are the Dutch pelagic freezer trawlers and both the Irish and Spanish demersal fleets. More sporadic discards are observed in German pelagic freezer trawlers and the UK demersal fleet. In 2016, Lithuania declared discards for the first time but has not provided estimates since 2018. Denmark has only declared discards in 2017 and 2018. Discard estimates are not obtained from French freezer trawlers, though discard patterns in these fleets are likely to be similar to the Dutch fleet. Discard data from the Portuguese bottom otter trawl fleet in ICES Division 9a are also available but are not included in the assessment as they are outside the TAC area. Presently Ireland, Spain and the UK are the only countries to declare discards from the fishery since 2019. Table 3.1.2.2 shows the total annual discards and estimates from the demersal and non-target fisheries respectively.

Discard data were included in the calculation of catch numbers at age. All discards were raised as a single metier using the same age length keys and sampling information as for the landed catches. In the absence of more comprehensive sampling information on discards, this was considered the best approach. This placed the stock in Category A2 for the ICES Advice in October 2013: Discards 'topped up' onto landings calculations. With the introduction of the discard ban in 2015 this stock was placed in A4: Discards known, with discard ban in place in year +1 . As such the advice will be given for catch in ICES Advice October 2014 and onwards.

### 3.2 Biological composition of the catch

### 3.2.1 Catches in numbers-at-age

Catch numbers-at-age were prepared from Irish, Danish, Dutch, Spanish, Polish and English landings using the ALK in Table 3.2.1.1 together with available samples from the fishery. This general ALK was constructed based on 814 aged fish from Irish, Danish and Scottish caught samples from 2012 (see the stock annex for a description of ALKs prior to 2012). In 2021, 12 samples, comprising of 564 fish, were collected and measured for length from the catch (Table 3.2.1.3). These samples covered the most heavily fished areas of 7 j and 7 b and equated to one sample per 1474 t landed.

The results of the application of the ALK to commercial length-frequency data (available for the years 2007-2021) produced proxy catch numbers-at-age values which are available in Table 3.2.1.4. In the most recent years, there has been the appearance of strong year classes in the catch numbers. A high number of 3-5 year olds were present in the 2021 data. The modal age from 2007-2011 was 6 and in 2012-2018 it was 7 . The modal age for 2021 is 4 . It should be noted that in WGWIDE 2011 and 2012 the plus group for boarfish was 20+. This was reduced to $15+$ in WGWIDE 2013 due to potential inaccuracy of the age readings of older fish. Ageing was based on the method that has been validated for ages $0-7$ by Hüssy et al. (2012a; b). The age range is similar to the published growth information presented by White et al. (2011).

### 3.2.2 Quality of catch and biological data

Length-frequencies of the international commercial landings by year are presented in Table 3.2.2.1. Sampling in the early years of the fishery (2006-2009) was sparse as there was no dedicated sampling programme in place. The sampling programme was initiated in 2010 and good coverage of the landings has been achieved since then. Full details of the sampling programme in the earlier years are presented in the stock annex. Until 2017, boarfish was not included on the DCF list of species for sampling. Irish sampling comprises only samples from Irish registered vessels. Samples are collected on-board directly from the fish pump during fishing operations and are frozen until the vessel returns to port, which ensures high quality samples. Each sample consists of approximately 6 kg of boarfish. This equates to approximately 150 fish which, given the limited size range of boarfish, is sufficient for determining a representative length frequency. The established sampling target is one sample per 1000 t of landings per ICES Division, which is also standard in other pelagic fisheries. Since 2017, all fish in each sample should be measured to the 0.5 cm below for length frequency. Following standard protocols 5 fish per 0.5 cm length class should be randomly selected from each sample for biological data collection i.e. otolith extraction, measurement to the 1 mm below and sex and maturity determination. There is no sampling programme in place for Scottish catches.

The current surplus production model used to assess boarfish is considered an interim measure prior to the development of an aged-based assessment. In 2017, boarfish was included in the list of species to be sampled by the Data Collection Multi Annual Programme (DCMAP) which should provide estimates of catch at age and facilitate the future development of an age-based stock assessment method.

### 3.3 Fishery Independent Information

### 3.3.1 Acoustic Surveys

The Boarfish Acoustic Survey (BFAS) was first conducted in July 2011. The 2022 survey was carried out by the RV Celtic Explorer and run in conjunction with the Malin Shelf herring survey as the WESPAS survey (Western European Shelf Pelagic Acoustic Survey). The survey was carried out over a 42-day period beginning on the 14 June in the south $\left(47^{\circ} 30 \mathrm{~N}\right)$ and working northwards to $59^{\circ} 30 \mathrm{~N}$ ending on 24 July.

Calculation of acoustic abundance
The StoX software package (Johnsen et. al., 2019) was used to calculate acoustic abundance from survey data (StoX V3.4.0 and R-StoX V3.4.0). Aggregated survey data are available for download from the ICES Trawl Acoustic database. Survey design and analysis procedures adhere to guidelines laid out in the Manual for International Pelagic Surveys (ICES, 2015).

## Survey results 2022

The estimate of boarfish biomass is presented in Table 3.3.1.1 and the spatial distribution of the echotraces attributed to boarfish in 2022 are presented in Figure 6.3.1.1b. Overall, the WESPAS survey provided continuous synoptic coverage from south to north over 42 days, relating to an area coverage of over $49,988 \mathrm{nmi}^{2}$ (boarfish strata) and transect mileage of over $5,084 \mathrm{nmi}$. In total, 40 trawl stations were undertaken with 26 hauls containing boarfish providing 6,575 individual lengths, 2,498 length and weight measurements and 1,270 otoliths for use during the analysis.

The 2022 estimate of total stock biomass (TSB) was comparable to that observed in 2021 (443,777 t in 2021 and 451,415 t in 2022). Survey effort, in terms of survey miles and area coverage, saw a reduction of $10 \%$ and $17 \%$ respectively when compared to 2021 , as a consequence of poor weather. Over $61.5 \%$ of the standing stock biomass was observed in the Celtic Sea stratum followed by $32.5 \%$ along the Irish west coast. The southern Celtic Sea/Northern Biscay area contained a high abundance of fish, dominated by newly recruited fish (2-year-olds) and first and second year spawning fish ( $3-4$ yrs old). Older age classes dominated further north along the Irish and Scottish west coasts and on the Porcupine Bank, with the exception of a discreet cluster of aggregations of immature and newly recruited fish south of the Minch.
Overall, immature boarfish represented only $5 \%$ of total stock abundance compared to $61 \%$ observed in 2021. In both years, the highest proportions of immature fish were observed in the Celtic Sea stratum. Preliminary results from the PELGAS survey in the Bay of Biscay, saw for the second successive year a high abundance of boarfish in the northern and mid-Biscay region, surpassing the abundance reported in 2021 (M. Doray, pers comm.). Therefore, it is feasible to suggest that the southern boundary of the stock was not contained within the survey area during WESPAS 2022.

The 3-year age class dominated the 2022 estimate contributing over $32 \%$ of TSB and $40.5 \%$ of total abundance (TSN), followed by the 4-year-old ( $27.2 \%$ TSB \& $22.9 \%$ TSN) and 2-year-old ( $9.3 \%$ TSB \& $17.4 \% \mathrm{TSN}$ ) fish respectively. Combined, these three successive year classes have tracked well through the index and represent a strong growth period for the stock, and is consistent with the previous observations. Older year classes (5 to 15+-year-old) are still evident and well represented within the stock.

The 2022 biomass saw an increase in TSB of $2 \%$ compared to 2021 and an increase of $26 \%$ in SSB. The increase in SSB has largely driven new recruitment (over $98 \%$ of 2 yr old fish reported as mature) and the lower numbers of immature fish present in the survey area this year. Survey
effort was reduced in the core Celtic Sea stratum due to poor weather, and combined with observations from the PELGAS survey would indicate that containment was not fully achieved. That said, comprehensive trawl and biological sampling was undertaken.

### 3.3.2 International bottom trawl survey (IBTS) Indices Investigation

The western IBTS data and CEFAS English Celtic Sea Groundfish Survey were investigated for their use as abundance indices for boarfish for the first time in 2012. An index of abundance was constructed from the following surveys:

- EVHOE, French Celtic Sea and Biscay Survey, (Q4) 1997 to 2021
- IGFS, Irish Groundfish Survey, (Q4) 2003 to 2021
- WCSGFS, West of Scotland, (Q1 and Q4) 1986 to 2009 (survey design changed in 2010)
- SPPGFS, Spanish Porcupine Bank Survey, (Q3) 2001 to 2021
- SPNGFS, Spanish North Coast Survey, (Q3/Q4) 1991 to 2021
- ECSGFS, CEFAS English Celtic Sea Groundfish Survey, (Q4) 1982 to 2003

From the IBTS data, CPUE was computed as the number of boarfish per 30 min haul. The abundance of boarfish per year per ICES statistical rectangle (used for visualisation only) was then calculated by summing the boarfish in a given rectangle and dividing by the total number of hauls in that rectangle. Length frequencies are presented in Table 3.3.2.1 for each survey. These surveys cover the majority of the observed range of boarfish in the ICES Area (Figure 3.1). Figure 3.3.2.1 shows the haul positions for each of the 6 surveys analysed.

A detailed analysis of the IBTS data was carried out in 2012 to investigate the main areas of abundance of boarfish in these surveys. This analysis included GAM modelling based on the probability of occurrence of boarfish. The full details of this work are presented in the stock annex. The IBTS appears to give a relative index of abundance, with good resolution between periods of high and low abundance. The main centres of abundance in the survey (Figure 3.3.2.2) correspond to main fishing grounds (Figure 3.1.2.1). Figures 3.3.2.3a and b shows the signal in abundance and biomass, increasing gradually in the 1990s, slowly declining in the early 2000s, before increasing again with a strong increase in the most recent period. Much of this increase which is stronger in terms of abundance is due to increased recruitment since 2017. The low estimates for the 2017 survey are partly explained by issues with the execution of the EVHOE survey. Due to mechanical breakdown, the majority of the survey stations could not be completed. The missed stations would have covered the area in North Biscay typically associated with the highest catch rates of boarfish.

For subsequent surplus production modelling (see Section 3.6.3), biomass indices were extracted from each of the IBTS surveys using a delta-lognormal model (Stefánsson 1996). Many of the surveys exhibited a large proportion of zero tows with occasionally very large tows, hence the decision to explicitly model the probability of a non-zero tow and the mean of the positive tows. A delta-lognormal fit comprises fitting two generalized linear models (GLMs). The first model (binomial GLM) is used to obtain the proportion of non-zero tows and is fit to the data coded as 1 or 0 if the tow contained a positive or zero CPUE, respectively. The second model is fit to the positive only CPUE data using a lognormal GLM. Both GLMs were fit using ICES statistical rectangle and year as explanatory factor variables. Where the number of tows per rectangle was less than 5 over the entire series, they are grouped into an "others" rectangle. An index per rectangle and year is constructed, according to Stefánsson (1996), by the product of the estimated probability of a positive tow times the mean of the positive tows. The station indices are aggregated by taking the estimated average across all rectangles within a year. To propagate the uncertainty, all survey index analyses were conducted in a Bayesian framework using Markov chain Monte

Carlo (MCMC) sampling (Kery 2010). The analyses were performed in WinBUGS from R with the R2WinBUGS package.

When the indices were recalculated in 2021, (following a refresh of the input data from DATRAS and national data submitters), the following issues were encountered

- An error with the coding of the EVHOE 2018 data in DATRAS was corrected, revising upwards the estimates from 2018 for this survey
- The truncated EVHOE 2017 dataset was removed from the analysis. In previous years, this data was retained but, because the available data only corresponds to a small fraction of the total survey area (where boarfish are not usually encountered in significant quantities) a very low survey estimate resulted. It was considered appropriate to remove this data from the analysis. In future, explicit modelling of spatial and temporal correlations may permit this data to be considered again.
- An error in the analysis was discovered whereby hauls with more than one catch category were underrepresented as only a single catch category was included during the model fitting. Multiple catch categories are usually the result of splitting the catch into adult and juvenile portions and using an appropriate subsampling strategy for each. This issue is particularly relevant for the IGFS which, over the most recent 4 years has 2 catch categories for boarfish recorded for approximately $20 \%$ of hauls. The outcome is an increase in CPUE for these hauls and a subsequent increase in the survey index for the IGFS in recent years (2016 onwards).


### 3.4 Mean weights- at-age, maturity-at-age and natural mortality

Mean weight-at-age was obtained from the ageing studies of Hüssy et al. (2012b). These mean weights are presented in the text table below. The variation in weight-at-age is due to the small sample size and the seasonal variation in weight and maturity stage.

| Age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Mean <br> Weight (g) | 0.84 | 6.65 | 14.6 | 19.5 | 23.7 | 26.8 | 33.3 | 37.7 | 40 | 47.1 |


| Age | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Mean <br> Weight (g) | 50.2 | 51.2 | 62.8 | 56.4 | 62.2 | 68.9 | 50.5 | 86.7 | 77.9 | 64.6 |


| Age | $\mathbf{2 0}$ | $\mathbf{2 1}$ | $\mathbf{2 2}$ | $\mathbf{2 3}$ | $\mathbf{2 4}$ | $\mathbf{2 5}$ | $\mathbf{2 6}$ | $\mathbf{2 7}$ | $\mathbf{2 8}$ | $\mathbf{2 9}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Mean Weight <br> (g) | 63.5 | 75 | 86 | 71 | 77 | 84.4 | 79.4 | - | 67.6 | 52.8 |

Maturity-at-age was obtained from the ageing studies of Hüssy et al. (2012a; b) and the reproductive study by Farrell et al. (2012).

| Age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Prop mature | 0 | 0 | 0.07 | 0.25 | 0.81 | 0.97 | 1 |

Natural mortality (M) was estimated over the life span of the stock using the method described by King (1995). This method assumed that M was the mortality that would reduce a population to $1 \%$ of its initial size over the lifespan of the stock. Based on a maximum age of $31, \mathrm{M}$ was calculated as follows

$$
M=-\ln (0.01) / 31
$$

Following this procedure, $M=0.16$ year ${ }^{-1}$ was considered a good estimate of natural mortality over the life span of the boarfish stock, as it was similar to the total mortality estimate from 2007, ( $Z=0.18$, see Section 3.6.5). Given that catches in 2007 were relatively low, this estimate of total mortality was considered a good estimate of natural mortality, assuming negligible fishing mortality in previous years.

Similarly, total mortality was estimated from age-structured IBTS data from 2003 to 2006 (years from which data was available for all areas). The total mortality was considered a good estimate of natural mortality as fishing mortality was assumed to be negligible during this period. Total mortality ranged from 0.09-0.2 with a mean of 0.16 .

The special review in 2012 questioned the validity of a single estimate of M across the entire age range. If an age based assessment is possible in the future, age specific estimates of natural mortality will be required. However, the current estimate of M , which covers the whole age range, is considered appropriate in the context of the current situation where age data are used as an indicator approach, rather than as a full assessment method. Given that Z and F are also calculated over the entire (fully selected) range (Section 3.6.5) a single value of M was considered appropriate.

### 3.5 Recruitment

The common ALK (Table 3.2.1.1.) was applied to the IBTS number-at-length data. The lengthfrequency is presented in Table 3.3.2.1. and the age-structured index in Table 3.6.1.1. and Figure 3.6.1.1.

A cohort effect can be seen with those cohorts from the early 2000s appearing weak. This coincides with a decline in overall abundance in the early 2000s. From the mid-2000s onwards recruitment improved as observed in the abundance of 1-5 year olds in the EVHOE and Spanish northern shelf surveys (It should be noted however that the IBTS data is measured to the 1.0 cm not the 0.5 cm until 2015. Therefore, application of the common ALK to this data must be viewed with caution).

The EVHOE, IGFS and SPNGFS surveys provide the best indices of recruitment as this is where the juveniles appear to be most abundant (Table 3.3.2.1) For example, in the EVHOE survey, particularly high recruitment has been noted between the years 2018 and 2021 for ages 1-3. And also, in the IGFS survey, signs of high recruitment could be observed as early as 2018, peaking in 2020. In 2021, the progression of the cohort can be seen as 3-5 years old.

### 3.6 Exploratory assessment

In 2012, a new stock assessment method for Boarfish was tested. In 2013 this Bayesian state space surplus production model (BSP; Meyer \& Millar (1999)) was further developed following reviewers' recommendations in 2012. Different applications of a Bayesian biomass dynamic model were run in 2013 incorporating combinations of catch data, abundance data from the groundfish surveys, and estimates of biomass (and associated uncertainty) from the acoustic surveys (see stock annex for more details of the sensitivity runs). The model and settings from the final accepted
run in 2013 were used as the basis of ICES category 1 advice for catch in 2014. However, in 2014 there was concern about the use of the production model for a number of reasons and ICES considered this model as no longer suitable for providing category 1 advice. Since 2014, the assessment model has been used as a basis for trends for providing DLS advice (ICES category 3). ICES considers the current basis for the advice on this stock to be an interim measure prior to development of an age-based assessment.

### 3.6.1 IBTS data

Some of the IBTS CPUE indices display marked variability with a large proportion of zero tows and occasionally very large tows (e.g. West of Scotland survey, Figure B.4.7 stock annex). More southern surveys display a consistently higher proportion of positive tows. The variability of the data is reflected in the estimated mean CPUE indices (Figure 3.6.1.2). The West of Scotland survey index had been increasing between 2000 and 2009 but is uncertain and was stopped soon after. The English Celtic Sea survey showed an upward trend in the last couple of years before the survey ended in 2003. Of the four current bottom trawl surveys, the French, Irish and Spanish Porcupine groundfish surveys experienced an increase in CPUE, particularly the French survey. The spatial extent of each survey is shown in Figure 3.3.2.1.

Diagnostics from the positive component of the delta-lognormal fits indicate relatively good agreement with a normal distribution on the natural logarithmic scale (Figure 3.6.1.4). There is an indication of longer tails in some of the surveys (e.g. WCSGFS, SPPGFS).
Pair-wise correlations between the annual mean survey indices is variable. The updates described above with respect to data and analysis code corrections have resulted in increased correlation between the surveys most affected i.e. IGFS and EVHOE (Figure 3.6.1.5). The WCSGFS displayed positive correlations with all five surveys except the Spanish north coast survey (SPNGFS). The SPPGFS displayed a negative correlation with EVHOE and IGFS. Weighting the correlations by the sum of the pair-wise variances resulted in a largely similar correlation structure (Figure 3.6.1.6). Note that though some surveys displayed weak or no correlation, no surveys were excluded a-priori from the assessment. Sensitivity tests were conducted in 2013, which led to the exclusion of the surveys mentioned previously (see the stock annex).

### 3.6.2 Biomass estimates from acoustic surveys

The Boarfish Acoustic Survey (BFAS) series was initiated in 2011 in partnership with industry. The 2011 survey collected data over 24 hours. In 2012, the protocol was changed to exclude the hours between 00:00 and 04:00 as aggregations break up during the hours of darkness. The 2011 data was reworked in 2015 to exclude the data between 00:00 and 04:00. An acoustic target strength model of (-66.2dB) was developed in 2013 (Fässler et al. (2013)) and is applied to all surveys in the time series. Over the time series of the survey total biomass has been estimated in the range 863 kt (in 2012) to 70 kt (2016) with CV estimates ranging 0.11 to 0.31 . Total biomass estimates declined sharply between 2012 and 2016 after which an increasing trend is seen. In the most recent surveys, the contribution of immature boarfish to the total estimate has been increasing such that the increase seen between 2020 and 2021 is largely due to juveniles. In 2022, the abundance of juvenile boarfish decreased; however, this year, the survey was dominated by a high abundance of 2-4 year olds. No substantial evidence exists for removing any of the survey points from the time series although 2016 may be considered an outlier (Table 3.3.1.1).

The PELACUS survey is conducted annually in waters to the south of the boarfish (WESPAS) survey. For the second successive year, PELACUS recorded an increase in biomass on its northern and mid-Biscay transects in 2022, (immediately south of the WESPAS southern limit), in
broad agreement with increases noted on WESPAS. The PELACUS survey takes place approximately 1 month prior to the boarfish survey.

### 3.6.3 Biomass dynamic model

In 2012, an exploratory biomass dynamic model was developed for the assessment of boarfish. The model is a Bayesian state space surplus production model (Meyer \& Millar 1999), incorporating the catch data, IBTS data, and acoustic biomass data. Following the initial development of the model, the assessment was peer-reviewed by two independent experts on behalf of ICES. In 2013 a new assessment was provided, which was based on the previous year's work and the reviewers' comments and formed the basis of a category 1 assessment. Details of the review and the associated changes can be found in the stock annex.

In 2014 the Bayesian state space surplus production model was fit using the catch data, deltalognormal estimated IBTS survey indices, and the acoustic survey estimates. However, the inclusion of the low 2014 acoustic biomass estimate changed the perception on the stock, which raised concerns over the sensitivity and process error of the model and the stock assessment was moved from ICES category 1 to category 3 with the results of the surplus production model being used to calculate an index for the data limited stock approach.
Since 2014, the procedure used to run the model has not changed with annual updates to the input data only.

In the Bayesian state space surplus production model the biomass dynamics are given by a difference form of a Schaefer biomass dynamic model:

$$
B_{t}=B_{t-1}+r B_{t-1}\left(1-\frac{B_{t-1}}{K}\right)-C_{t-1}
$$

where $B_{t}$ is the biomass at time $\mathrm{t}, \mathrm{r}$ is the intrinsic rate of population growth, $K$ is the carrying capacity, and $C_{t}$ is the catch, assumed known exactly. To assist estimation, the biomass is scaled by the carrying capacity, denoting the scaled biomass $P_{t}=B_{t} / K$. A lognormal error structure is assumed giving the scaled biomass dynamics (process) model:

$$
P_{t}=\left(P_{t-1}+r P_{t-1}\left(1-P_{t-1}\right)+\frac{C_{t-1}}{K}\right) e^{\mu_{t}}
$$

where the logarithm of process deviations are assumed normal $u_{t}=N\left(0, \sigma_{2}^{\mu}\right)$ with $\sigma_{2}^{\mu}$ the process error variance.

The starting year biomass is given by $a K$, where a is the proportion of the carrying capacity in the first year. The biomass dynamics process is related to the observations on the indices through the measurement error equation:

$$
I_{j, t}=q_{j} P_{t} K e^{\varepsilon_{j, t}}
$$

where $I_{j, t}$ is the value of abundance index $j$ in year $t, q_{j}$ is survey-specific catchability, $B_{t}=P_{t} K$, and the measurement errors are assumed log-normally distributed with $u_{t}=N\left(0, \varepsilon_{e, j, t}^{2}\right)$ where $\varepsilon_{e, j, t}^{2}$ is the index-specific measurement error variance. $\operatorname{Var}\left(I_{j, t}\right)$ is obtained from the delta-lognormal survey fits. That is, the variance of the mean annual estimate per survey is input directly from the delta-lognormal fits (Figure 3.6.1.2) as opposed to estimating a measurement error within the assessment. The measurement error is obtained from:

$$
\sigma_{e, j, t}^{2}=\ln \left(1+\frac{\operatorname{Var}\left(I_{j, t}\right)}{\left(I_{j, t}\right)^{2}}\right)
$$

For the acoustic survey, the CV of the survey was transformed into a lognormal variance via

$$
\sigma_{\varepsilon, a c o u s t i c, t}^{2}=\ln \left(C V_{a c o u s t i c, t}^{2}+1\right)
$$

Prior assumptions on the parameter distributions are:

- Intrinsic rate of population growth: $r \sim U(0.001,2)$
- Natural logarithm of the carrying capacity: $\ln (K) \sim U(\ln (\max (C), \ln (10 . \operatorname{sum}(C))=$ $U(\ln (144047), \ln (4450407))$
- $\quad$ Proportion of carrying capacity in first year of assessment: $a \sim$ U[0.001, 1.0]
- Natural logarithm of the survey-specific catchabilities $\ln \left(q_{i}\right) \sim U(-16,0)$ (for IBTS only). The acoustic survey prior is discussed below.
- $\quad$ Process error precision $\frac{1}{\sigma_{u}^{2}} \sim \operatorname{gamma}(0.001,0.001)$


## Specification

During the 2013 WGWIDE meeting a number of different iterations of the model were run to discern the best parameters for the assessment. After four initial runs and four sensitivity runs the settings for the final run (run 2.2) were chosen. These settings are shown below and were used for the assessment model since 2014. (More details of the trial runs in 2013 can be found in the stock annex).

The specifications for the final boarfish assessment model runs are:

## Acoustic survey

Years: 2011-2022
Index value (Iacoustic,y): 'total' in tonnes (i.e. Definitely Boarfish + Probably Boarfish + Boarfish in a Mix)

Catchability ( $q_{\text {acoustic }}$ ): A free, but strong prior (i.e. the acoustic survey is treated as a relative index but is strongly informed, this allows the survey to cover $<100 \%$ of the stock).

## IBTS surveys

6 delta log normal indices (WCSGFS, SPPGFS, IGFS, ECSGFS, SPNGFS, EVHOE)
First 5 and last 7 (since 2017, because of change in survey design) years omitted from WCSGFS
First 9 years omitted from ECSGFS
Following discussion of the sensitivity runs in 2013, it was decided that the final run be based on a run that includes all surveys with the omission of the first 5 years of the WCSGFS and first 9 years of the ECSGFS as it was unclear whether boarfish were consistently recorded in the early part of the ECSGFS. The WCSGFS is thought to be at the northern extreme of the distribution and may not be an appropriate index for the whole stock. The initial data year was set at 1991 when 3 groundfish survey indices are available (SPNGFS, ECSGFS and WCSGFS). The survey indices are weighted such that highly uncertain values receive lower weight in the fitting.

## Catches

2003-2021 time-series

## Priors

The final run assumes a strong prior for the acoustic survey catchability with $\ln \left(q_{\text {acoustic }}\right) \sim N(1$, $1 / 4$ ) (mean 1 , standard deviation 0.25 ), which has $95 \%$ of the density between 0.5 and 2 . Given the relatively short acoustic series it is not possible to estimate this parameter freely (i.e. using an uninformative prior). The prescription of a strong prior removes the assumption of an absolute index from the acoustic survey. This assumption will be continually updated as additional data accrue.

## Run convergence

Parameters for the 2022 model run converged with good mixing of the chains and Rhat values lower than 1.1 indicating convergence and acceptable autocorrelation (Figures 3.6.3.1-3).

Diagnostic plots are provided in Figure 3.6.3.4 showing residuals about the model fit. A fairly balanced residual pattern is evident. In some cases, outliers are apparent, for instance in the English survey in the final year (2003). However, these points are down weighted according to the inverse of their variance and hence do not contribute much to the model fit. For the early years of the acoustic survey (BFAS), it overestimates the stock in the first 3 years, then underestimates it for the next 4 years before again overestimating it slightly in 2020. This suggests that this index is perhaps not representative of the whole stock. For the last two years, the residuals have been well behaved. Figure 3.6.3.5 shows the prior and posterior distributions of the parameters of the biomass dynamic model. The estimate of $q$ is less than 1.0, leading to a higher estimate of final stock biomass than the acoustic survey result.

## Results

Trajectories of observed and expected indices are shown in Figure 3.6.3.6, along with the stock size over time and a harvest ratio (total catch divided by estimated biomass). Parameter estimates from the model run are summarized in Table 3.6.3.1. TSB in 2022 is estimated to be 565 kt , continuing the increasing trend in stock size since 2016. The extremely low biomass estimate from the 2016 acoustic survey appears to be largely considered as an outlier by the model. This is also the case for the high survey estimate in 2012 although the drop in biomass between these points is seen in a number of the input data series. Retrospective plots of TSB and F, presented in Figure 3.6.3.7, show that the perception of the stock is stable over the most recent 5 years.

### 3.6.4 State of the stock

The most recent assessment indicates that total stock biomass increased from a low to average level from the early to mid-1990s (Figure 3.6.3.6). The stock fluctuated around this level until 2009, before increasing until 2012. A sharp decline is seen between 2013 and 2014. Since 2014, the abundance has increased although it remains below that from the previous high period. There was concern in 2014 that this decline was exaggerated by an unusually low acoustic biomass estimate that led to a downward revision in stock trajectory. However, the 2014 survey is considered satisfactory in terms of containment. The comparably low 2014 biomass estimate was supported by results of the 2015 survey. The 2016 biomass estimate, the lowest of the time series is considered likely an outlier and has little influence on stock abundance estimates. The $95 \%$ uncertainty bounds are relatively large reflecting the uncertainty in the survey indices, and short exploitation history of the stock and the treatment of the acoustic survey as a relative biomass index.

Catch data are available from 2001, the first year of commercial landings, and reasonably comprehensive discard data are available from 2003. Peak catches were recorded in 2010, when over 140000 t were taken. Elevated fishing mortality was observed, associated with the highest recorded catch in 2010. Fishing mortality, expressed as a harvest ratio (catch divided by total biomass), was first recorded in 2003. Before that time, it is to be expected that some discarding took place, and there were some commercial landings. Fishing mortality increased measurably from 2006, reaching a peak in 2009-2010. F declined in 2011 as catches became regulated by the precautionary TAC but increased year on year until 2015 when reduced catches resulted in a reduction in F. The considerable catches in recent years do not appear to have significantly truncated the size or age structure of the stock and $15+$ group fish are still abundant (Figure 3.2.1.1).

MSY reference points can be estimated from the production model assessment parameter values. In 2021, $\mathrm{F}_{\text {MSY }}(\mathrm{r} / 2)$ is estimated to be 0.17 and MSY $\mathrm{B}_{\text {trigger }}(\mathrm{K} / 4) 138 \mathrm{kt}$. Throughout the history of the fishery, estimates of total biomass have remained above MSY $\mathrm{B}_{\text {trigger. }}$. Fishing mortality (F) was briefly larger than the estimate of FMSY between 2009 and 2010 and again in 2014, but has decreased since. In 2021, the stock is in the green area of the Kobe plot (Figure 3.6.6.1).

Estimates of recruitment are not available from the stock assessment. However, all available data sources (catch, acoustic survey and IBTS surveys) indicate above average recruitment since 2017. The large juvenile biomass observed in the 2021 acoustic survey is tracking well through the index and is present in the 2022 survey as newly recruited fish of the 2-4 year classes.

### 3.7 Short Term Projections

As the assessment is exploratory, no short term projections were conducted.

### 3.8 Long term simulations

No long term simulations were conducted.

### 3.9 Candidate precautionary and yield based reference points

### 3.9.1 Yield per Recruit

A yield per recruit analysis was conducted in 2011 (Minto et al. 2011) and F0.1 was estimated to be 0.13 whilst $\mathrm{Fmax}_{\text {was }}$ wastimated in the range 0.23 to 0.33 (Figure 3.9.1.1). F0.1 was considered to be well estimated (Figure 3.9.1.2). No new yield per recruit analyses were performed in subsequent years.

### 3.9.2 Precautionary reference points

No reference points have been defined for boarfish.

### 3.9.3 Other yield based reference points

Yield per recruit analysis, following the method of Beverton \& Holt (1957), found F0.1 to be robustly estimated at 0.13 (ICES 2011; Minto et al. 2011).

### 3.10 Quality of the assessment

ICES considers the current basis for the advice on this stock to be an interim measure prior to development of an age-based assessment. The acoustic survey has undergone several developments to improve its suitability with updates to methodology in 2012, a change in direction in 2017 and extension of transects at the boundaries to improve containment. The assessment was downgraded from Category 1 to Category 3 in 2014, and it has remained in this category since. The model is still considered suitable for category 3 advice, because it provides the best means of combining the available survey series. The assessment is sensitive to the acoustic series. In addition, a substantial part of the year to year variations in the stock abundance is linked to the
process error. The use of some priors (like ratio to virgin biomass in the first year of the assessment) and survey (e.g. WCSGFS for instance) may require revision.

The bottom trawl survey data are considered to be a good index of abundance given that boarfish aggregate near the bottom at this time of year. The trawl surveys record high abundances of the species, but with many zero hauls. The delta-lognormal error structure used in the analyses is considered to be an appropriate means of dealing with such data. The biomass dynamic model used in the stock assessment is based on the assessment of megrim in Sub-divisions 4 and 6 with the model further developed by including acoustic survey biomass estimates. A drawback of the current assessment model is that it does not provide estimates of recruitment although estimates of recruitment strength are available from the Spanish and French bottom trawl surveys.

### 3.11 Management considerations

As this stock is placed in category 3, the advice is based on harvest control rules for data limited stocks (ICES 2012). Since the biomass estimate from the Bayesian model is considered reliable for trends based assessment, an index can be calculated according to Method 3.1 of ICES (2012). The advice is based on a comparison of the average of the two most recent index values with the average of the three preceding values multiplied by the most recent catch. Table 3.6.5.1 shows the biomass estimates from the model from which the index was calculated. Although not currently accepted as the basis for an analytic assessment, the surplus production model still provides the best unified view of this stock (Figure 3.6.3.6).

### 3.12 Stock structure

A dedicated study on the stock structure of boarfish within the Northeast Atlantic and Mediterranean Sea commenced in October 2013 in order to resolve outstanding questions regarding the stock structure of boarfish and the suitability of assessment data. Results (Farrell et al. 2016) indicated strong population structure across the distribution range of boarfish with 7-8 genetic populations identified (Figure 3.12.1).

The eastern Mediterranean (MED) samples comprised a single population and were distinct from all other samples. Similarly, the Azorean (AZA), Western Saharan (MOR) and Alboran (ALM) samples were distinct from all others. Of particular relevance to the assessment and management of the boarfish fishery is the identification and delineation of the population structure between southern Portuguese waters (PTN2B-PTS) and waters to the geographic north. A distinct and temporally stable mixing zone was evident in the waters around Cabo da Roca. The PTN2A sample appeared to be significantly different from all other samples however this sample was relatively small and was considered to represent a mixed sample rather than a true population.

No significant spatial or temporal population structure was found within the samples comprising the NEA population (Figure 3.12.1). A statistically significant but comparatively low level of genetic differentiation was found between this population and the northern Spanish shelf/northern Portuguese samples (NSA-PTN1). However, a high level of migration was revealed between these two populations and no barriers to gene flow were detected between them. Therefore, for the purposes of assessment and management these areas can be considered as one unit.

Analyses indicated a lack of significant immigration into this northeast Atlantic boarfish stock from populations to the south or from insular elements and the strong genetic differentiation among these regions indicate that the purported increases in abundance in the northeast Atlantic area are not the result of a recent influx from other regions. The increase in abundance is most
likely the result of demographic processes within the northeast Atlantic stock (Blanchard \& Vandermeirsch 2005; Coad et al. 2014).

Whilst the current assessment and management area constitutes the majority of the most northern population it should be extended into Northern Portuguese waters and repeated genetic monitoring of the stock in this region should be conducted to ensure the validity of this delineation. Based on analyses of IBTS data the biomass in this area is suspected to be small relative to the overall biomass in the TAC area.

### 3.13 Ecosystem considerations

The ecological role and significance of boarfish in the NE Atlantic is largely unknown. However, in the southeast North Atlantic, in Portuguese waters, they are considered to have an important position in the marine food web (Lopes et al. 2006). The diet has been investigated in the eastern Mediterranean, Portuguese waters and at Great Meteor Seamount and consists primarily of copepods, specifically Calanus helgolandicus, with some mysid shrimp and euphausiids (Macpherson 1979; Fock et al. 2002; Lopes et al. 2006). This contrasted with the morphologically similar species, the slender snipefish, Macroramphosus gracilis and the longspine snipefish, M. scolopax, whose diet comprised Temora spp., copepods and mysid shrimps, respectively (Lopes et al. 2006). Despite the obvious potential for these species to feed on fish eggs and larvae, there was no evidence to support this conclusion in Portuguese waters and they were not considered predators of commercial fishes and thus their increase in abundance was unlikely to affect recruitment of commercial fish species. If the NE Atlantic population of boarfish is sufficiently large then there exists, the possibility of competition for food with other widely distributed planktivorous species.

Both seasonal and diurnal variations were observed in the diet of boarfish in all three regions. In the eastern Mediterranean and Portuguese waters, mysids become an important component of the diet in autumn, which correlates with their increased abundance in these regions at this time (Macpherson 1979; Lopes et al. 2006). Fock et al. (2002) found that boarfish at Great Meteor Seamount fed mainly on copepods and euphausiids diurnally and on decapods nocturnally, indicating habitat dependent resource utilization.

Boarfish appear an unlikely target of predation given their array of strong dorsal and anal fin spines and covering of ctenoid scales. However, there is evidence to suggest that they may be an important component of some species' diets. Most studies have focused in the Azores and few have mentioned the NE Atlantic, probably due to the relatively low abundance in the region until recent years. In the Azores, boarfish was found to be one of the most important prey items for tope (Galeorhinus galeus), thornback ray (Raja clavata), conger eel (Conger conger), forkbeard (Phycis phycis), bigeye tuna (Thunnus obesus), yellowmouth barracuda (Sphyraena viridensis), swordfish (Xiphias gladius), blackspot seabream (Pagellus bogaraveo), axillary seabream (Pagellus acarne) and blacktail comber (Serranus atricauda) (Clarke et al. 1995; Morato et al. 1999, 2000, 2001, 2003; Arrizabalaga et al. 2008). Many of these species also occur in the NE Atlantic shelf waters although it is unknown whether boarfish represent a significant component of the diet in this region.

In the NE Atlantic boarfish have not previously been recorded in the diets of tope or thornback ray (Holden \& Tucker 1974; Ellis et al. 1996). However, this does not prove that they are currently not a prey item. A study of conger eel diet in Irish waters from 1998-1999 failed to find boarfish in the diet (O'Sullivan et al. 2004). However, in Portuguese waters a recent study has found boarfish to be the most numerous species in the diet of conger eels (Xavier et al. 2010). It has been suggested that boarfish are an important component of the diet of hake (Merluccius merluccius), as they are sometimes caught together. However, a recent study of the diet of hake in the Celtic

Sea and Bay of Biscay did not report any boarfish in the stomachs of hake caught during the 2001 EVHOE survey (Mahe et al. 2007).

The conspicuous presence of boarfish in the diet of so many fish species in the Azores is perhaps more related to the lack of other available food sources than to the palatability of boarfish themselves. Given the large abundance in NE Atlantic shelf waters it is likely that they would have been recorded more frequently if they were a significant and important prey item.

Boarfish are also an important component of the diet a number of sea birds in the Azores, most notably the common tern (Sterna hirundo) (Granadeiro et al. 2002) and Cory's shearwater (Calonectris diomedea) (Granadeiro et al. 1998). This is surprising given that in the Mediterranean discarded boarfish were rejected by seabirds whereas in the Azores they were actively preyed on (Oro \& Ruiz 1997). Cory's shearwaters are capable of diving up to 15 m whilst the common tern is a plunge-diver and may only reach $2-3 \mathrm{~m}$. It is therefore surprising that boarfish are such a significant component of their diet given that it is generally considered a deeper water fish. In the Azores boarfish shoals are sometimes driven to the surface by horse mackerel and barracuda where they are also attacked by diving sea birds (J. Hart, CW Azores, pers. comm.). Anecdotal reports from the Irish fishery indicate that boarfish are rarely found in waters shallower than 40 m . This may suggest that they are outside the range of shearwaters and gannets, the latter having a mean diving depth of $19.7 \pm 7.5 \mathrm{~m}$ (Brierley \& Fernandes 2001). However, the upper depth range of boarfish is within maximum diving depth recorded for auks ( 50 m ) as recorded by Barrett \& Furness (1990). Given their frequency in the diets of marine and bird life in the Azores, boarfish appear to be an important component of the marine ecosystem in that region. There is currently insufficient evidence to draw similar conclusions in the NE Atlantic.

The length-frequency distribution of boarfish may be important to consider. IBTS data shows an increase in mean total length with latitude (Table 3.3.2.1) and perhaps the smaller boarfish in the southern regions are more easily preyed upon. Length data of boarfish from stomach contents studies of both fish and sea birds in the Azores indicate that the boarfish found are generally $<$ 10 cm (Granadeiro et al. 1998, 2002).

### 3.14 Proposed management plan

In 2015 the Pelagic Advisory Council submitted a revised draft management strategy for Northeast Atlantic boarfish. The EU has requested ICES to evaluate the following management plan:

This management strategy aims to achieve sustainable exploitation of boarfish in line with the precautionary approach to fisheries management, FAO guidelines for new and developing fisheries, and the ICES form of advice.

1 ) The TAC shall be set in accordance with the following procedure, depending on the ICES advice
a) If category 1 advice (stocks with quantitative assessments) is given based on a benchmarked assessment, the TAC shall be set following that advice.
b) If category 1 or 2 (qualitative assessments and forecasts) advice is given based on a non-benchmarked assessment the TAC shall be set following this advice.
c) Categories 3-6 are described below as follows:
i) Category 3: stocks for which survey-based assessments indicate trends. This category includes stocks with quantitative assessments and forecasts which for a variety of reasons are considered indicative of trends in fishing mortality, recruitment, and biomass.
ii ) Category 4: stocks for which only reliable catch data are available. This category included stocks for which a time series of catch can be used to approximate MSY.
iii ) Category 5: landings only stocks. This category includes stocks for which only landings data are available.
iv ) Category 6: negligible landings stocks and stocks caught in minor amounts as bycatch.
2 ) Notwithstanding paragraph 1, if, in the opinion of ICES, the stock is at risk of recruitment impairment, a TAC may be set a lower level.
3 ) If the stock, estimated in either of the 2 years before the TAC is to be set, is at or below Blim or any suitable proxy thereof, the TAC shall be set at 0 t .
4 ) The TAC shall not exceed $75,000 t$ in any year.
5 ) The TAC shall not be allowed to increase by more than $25 \%$ per year. However, there shall be no limit on the decrease in TAC.

6 ) Closed seasons, closed areas, and moving on procedures shall apply to all directed boarfish fisheries as follows:
i) A closed season shall operate from $31^{\text {st }}$ March to $31^{\text {st }}$ August. This is because it is known that herring and mackerel are present in these areas and may be caught with boarfish.
ii A closed area shall be implemented inside the Irish 12-miles limit south of $52^{\circ} 30$ from $12^{\text {th }}$ February to $31^{\text {st }}$ October, in order to prevent catches of Celtic Sea herring, known to form aggregations at these times.
iii ) If catches of other species covered by a TAC amount to more than $5 \%$ of the total catch by day by ICES statistical rectangle, then all fishing must cease in that rectangle for 5 consecutive days.

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### 3.16 Tables

Table 3.1.2.1. Boarfish in ICES Subareas 27.6, 7, 8. Landings by country, total discards and TAC by year (t), 2001-2021. (Data provided by Working Group members)

|  | Denmark | Germany | Ire- <br> land | Netherlands | Eng- <br> land | Po- <br> land | Scot- <br> land | Spain | Discards | Total | TAC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 |  |  | 120 |  |  |  |  |  |  | 120 |  |
| 2002 |  |  | 91 |  |  |  |  |  |  | 91 |  |
| 2003 |  |  | 458 |  |  |  |  |  | 10929 | 11387 |  |
| 2004 |  |  | 675 |  |  |  |  |  | 4476 | 5151 |  |
| 2005 |  |  | 165 |  |  |  |  |  | 5795 | 5959 |  |
| 2006 |  |  | 2772 |  |  |  |  |  | 4365 | 7137 |  |
| 2007 |  |  | 17615 |  |  |  | 772 |  | 3189 | 21576 |  |
| 2008 | 3098 |  | 21585 |  |  |  | 0 |  | 10068 | 34751 |  |
| 2009 | 15059 |  | 68629 |  |  |  |  |  | 6682 | 90370 |  |
| 2010 | 39805 |  | 88457 |  |  |  | 9241 |  | 6544 | 144047 |  |
| 2011 | 7797 |  | 20685 |  |  |  | 2813 |  | 5802 | 37096 | 33000 |
| 2012 | 19888 |  | 55949 |  |  |  | 4884 |  | 6634 | 87355 | 82000 |
| 2013 | 13182 |  | 52250 |  |  |  | 4380 |  | 5598 | 75409 | 82000 |
| 2014 | 8758 |  | 34622 |  |  |  | 38 |  | 1813 | 45231 | 133957 |
| 2015 | 29 | 4 | 16325 | 375 | 104 |  |  |  | 929 | 17766 | 53296 |
| 2016 | 337 | 7 | 17496 | 171 | 21 |  |  |  | 1283 | 19315 | 47637 |
| 2017 | 548 |  | 15485 | 182 | 0 |  |  |  | 1173 | 17388 | 27288 |
| 2018 | 94 |  | 9513 | 172 | 0 |  | 0 | 148 | 1359 | 11286 | 21830 |
| 2019 | 757 |  | 9910 | 318 | 19 |  |  | 3 | 306 | 11312 | 21830 |
| 2020 | 196 |  | 14666 | 416 | 62 | 109 |  | 1 | 198 | 15649 | 19152 |
| 2021 | 4322 |  | 11830 | 781 | 45 | 45 | 9 | 11 | 651 | 17693 | 19152 |
| $0=<0.5 \mathrm{t}$ |  |  |  |  |  |  |  |  |  |  |  |

Table 3.1.2.2. Boarfish in ICES Subareas 27.6, 7, 8. Discards in demersal and non-target pelagic fisheries by year (data provided by Working Group members)

| Year | Denmark | Germany | Ireland | Netherlands | Spain | UK | Lithuania |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 |  |  | 119 | 1998 | 8812 |  |  |
| 2004 |  |  | 60 | 837 | 3579 |  |  |
| 2005 |  |  | 55 | 733 | 5007 |  |  |
| 2006 |  |  | 22 | 411 | 3933 |  |  |
| 2007 |  |  | 549 | 23 | 2617 |  |  |
| 2008 |  |  | 920 | 738 | 8410 |  |  |
| 2009 |  |  | 377 | 1258 | 5047 |  |  |
| 2010 |  |  | 85 | 512 | 5947 |  |  |
| 2011 |  | 49 | 107 | 185 | 5461 |  |  |
| 2012 |  |  | 181 | 88 | 6365 |  |  |
| 2013 |  | 22 | 47 | 11 | 5518 |  |  |
| 2014 |  | 117 | 50 | 477 | 1119 | 50 |  |
| 2015 |  |  | 7 |  | 921 | 1 |  |
| 2016 |  | 869 | 20 | 41 | 348 | 4 | 1 |
| 2017 | 386 |  | 640 | 146 |  |  | 1 |
| 2018 | 744 |  | 525 | 89 |  |  | 1 |
| 2019 |  |  | 57 |  | 240 | 8 |  |
| 2020 |  |  | 64 |  | 133 | 1 |  |
| 2021 |  |  | 11 |  | 594 | 46 |  |
| $0=<0.5 \mathrm{t}$ |  |  |  |  |  |  |  |

Table 3.1.2.3. Boarfish in ICES Subareas 27.6

| Country | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 37 | 67 | 172 | 10 | 23 |
| England |  |  |  |  |  |  |  |  |  |  |  |  | 9 |  |  |  | 9 | 7 |  |
| Ireland | 65 | 292 | 10 | 21 | 99* | 28 | 45 | 1356 | 26 | 125 | 538 | 182 | 116 | 377 | 907 | 269 | 568 | 1214 | 378 |
| Netherlands |  |  |  |  |  |  |  |  |  |  |  |  | 128 | 45 | 34 | 78 | 79 | 108 | 52 |
| Scotland |  |  |  |  |  |  |  | 10 |  |  | 15 | 30 |  |  |  |  |  |  | 6 |

* 6 t in $5 \mathrm{~b}, 0=0-0.5 \mathrm{t}$

Table 3.1.2.4 Boarfish in ICES Subareas 27.7bc

| Country | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  |  |  |  |  |  |  | 80 | 12 | 8 | 21 |  |  |  | 85 | 13 |
| England |  |  |  |  |  |  |  |  |  |  |  |  | 85 | 1 |  |  | 0 | 32 | 10 |
| Germany |  |  |  |  |  |  |  |  |  |  |  |  | 4 | 5 |  |  |  |  |  |
| Ireland | 214 | 224 | 105 | 15 | 1259 | 3 | 74 | 2293 | 283 | 4609 | 10405 | 3262 | 2829 | 1198 | 124 | 163 | 241 | 6818 | 3732 |
| Netherlands |  |  |  |  |  |  |  |  |  |  |  |  | 33* | 35 | 138 | 10 | 150 | 212 | 228 |
| Scotland |  |  |  |  |  |  |  | 4 |  | 1745 | 100 |  |  |  |  |  |  |  | 2 |

## Table 3.1.2.5 Boarfish in ICES Subareas 27.7h-k

| Country | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  |  |  |  | 39132 | 7779 | 18203 | 11828 | 8747 | 5 | 330 | 239 | 6 | 268 | 101 | 4151 |
| England |  |  |  |  |  |  |  |  |  |  |  |  | 10 | 16 | 0 | 0 | 3 | 23 | 23 |
| Ireland | 179 | 122 | 12 | 2360 | 16131 | 21370 | 63597 | 81160 | 19565 | 50507 | 38358 | 30925 | 12152 | 8623 | 2994 | 3745 | 6222 | 6365 | 6956 |
| Netherlands |  |  |  |  |  |  |  |  |  |  |  |  |  | 90 | 9 | 68 | 80 | 79 | 325 |
| Poland |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 109 | 12 |
| Scotland |  |  |  |  | 772 |  |  | 9227 | 2813 | 3139 | 3381 | 8 |  |  |  | 0 |  |  |  |
| Spain |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 0 |  |
| $0=0-0.5 \mathrm{t}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 3.1.2.6 Boarfish in ICES Divisions 7e-g

| Country | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  |  |  |  | 674 |  |  |  |  |  |  | 1 |  | 1 | 0 | 23 |
| England |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 |  | 6 |  | 12 |
| Ireland |  |  |  | 375 | 120 | 184 | 4912 | 3649 | 811 | 616 | 1808 | 135 | 547 |  | 1 | 2 |  | 1 | 764 |
| Netherlands |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 0 | 3 | 7 | 1 | 126 |
| Scotland |  |  |  |  |  |  |  |  |  |  | 883 |  |  |  |  |  |  |  |  |
| $0=0-0.5 \mathrm{t}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Country | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  |  |  |  |  | 18 |  | 1354 |  | 6 | 7 | 271 |  | 315 |  | 111 |
| England |  |  |  |  |  |  |  |  |  |  |  |  |  | 5 |  |  |  |  |  |
| Germany |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 1 |  |  |  |  |  |
| Ireland |  | 38 | 38 | 1 | 5 |  |  |  |  | 93 | 1140 | 119 | 682 | 7297 | 11458 | 5336 | 2876 | 269 |  |
| Netherlands |  |  |  |  |  |  |  |  |  |  |  |  | 2014 |  |  | 14 | 0 | 17 | 48 |
| Spain |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 148* | 2 | 1 | 11 |

*94t in $9 \mathrm{a}, 0=0-0.5 \mathrm{t}$

Table 3.2.1.1. Boarfish in ICES Subareas 27.6, 7, 8. General boarfish age length key produced from 2012 commercial samples. Figures highlighted in grey are estimated

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7.25 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 7.75 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 8.25 |  | 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 8.75 |  | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 9.25 |  | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 9.75 |  |  | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 10.25 |  |  | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 10.75 |  |  | 2 | 10 | 3 |  |  |  |  |  |  |  |  |  |  |
| 11.25 |  |  | 1 | 29 | 14 | 2 | 2 |  |  |  |  |  |  |  |  |
| 11.75 |  |  |  | 9 | 21 | 21 | 18 | 2 | 2 | 1 |  |  |  |  |  |
| 12.25 |  |  |  | 4 | 17 | 22 | 38 | 12 | 8 |  |  |  |  |  | 1 |
| 12.75 |  |  |  |  | 5 | 9 | 42 | 37 | 14 | 6 | 2 |  | 1 | 1 | 1 |
| 13.25 |  |  |  |  | 2 | 4 | 31 | 28 | 24 | 12 | 6 | 2 | 3 | 1 | 5 |
| 13.75 |  |  |  |  | 1 | 3 | 25 | 22 | 21 | 14 | 6 | 5 | 4 | 2 | 11 |
| 14.25 |  |  |  |  |  |  | 6 | 8 | 18 | 22 | 8 | 3 | 7 | 1 | 20 |
| 14.75 |  |  |  |  |  | 1 | 1 | 2 | 3 | 8 | 1 | 6 | 6 | 6 | 30 |
| 15.25 |  |  |  |  |  |  | 1 | 1 |  | 2 | 2 | 2 | 5 | 2 | 19 |
| 15.75 |  |  |  |  |  |  |  |  |  | 2 |  |  |  | 2 | 19 |
| 16.25 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 8 |
| 16.75 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |
| 17.25 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |
| 17.75 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |
| 18.25 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |
| 18.75 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |

Table 3.2.1.2. Boarfish in ICES Subareas 27.6, 7, 8. Number of samples collected from the catch per year

| Year | Landings | Percent landings covered by sampling | No. samples | No. measured | No. aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 120 | 0 | 0 | 0 | 0 |
| 2002 | 91 | 0 | 0 | 0 | 0 |
| 2003 | 458 | 0 | 0 | 0 | 0 |
| 2004 | 675 | 0 | 0 | 0 | 0 |
| 2005 | 165 | 0 | 0 | 0 | 0 |
| 2006 | 2772 | 0 | 0 | 0 | 0 |
| 2007 | 18387 | NA | 3 | 217 | 0 |
| 2008 | 24683 | NA | 1 | 152 | 0 |
| 2009 | 83688 | NA | 9 | 1475 | 0 |
| 2010 | 137503 | NA | 95 | 10675 | 403* |
| 2011 | 31295 | NA | 27 | 4066 | 704 |
| 2012 | 80720 | NA | 80(68)*** | 9656(8565)*** | 814** |
| 2013 | 69812 | NA | 76 | 9392 | 0**** |
| 2014 | 43418 | NA | 54 | 7008 | 0**** |
| 2015 | 16837 | NA | 32 | 3356 | 0**** |
| 2016 | 18031 | NA | 27 | 3861 | 0**** |
| 2017 | 16215 | NA | 18 | 1140 | 0**** |
| 2018 | 9927 | NA | 12 | 556 | 0**** |
| 2019 | 11006 | NA | 8 | 371 | 0**** |
| 2020 | 15451 | NA | 10 | 534 | 0**** |
| 2021 | 17042 | NA | 12 | 564 | 0**** |

* A common ALK was developed from fish collected from both commercial and survey samples. This comprehensive ALK was used to produce catch numbers at age data for pseudo-cohort analyses.
** A common ALK was developed from fish collected from Danish, Irish and Scottish commercial landings. This comprehensive ALK was used for all métiers to produce catch numbers-at-age for the pseudo-cohort analysis.

Only aged fish measured to the 0.5 cm were included in the ALK.
*** Only Irish collected samples were used for the length frequency, see stock annex.
**** 2012 ALK was used.

Table 3.2.1.3. Boarfish in ICES Subareas 27.6, 7, 8. Catch per country and corresponding number of samples collected in 2021

| Official catch | Country | No. samples | No. measured | No. aged |
| :--- | :--- | :--- | :--- | :--- |
| 4322 | DK | 0 | 0 | 0 |
| 11 | ES | 0 | 0 | 0 |
| 11830 | IE | 12 | 564 | 0 |
| 781 | PL | 0 | 0 | 0 |
| 45 | UKE | 0 | 0 | 0 |
| 45 | UKS | 0 | 0 | 0 |
| 9 |  |  | 0 | 0 |

Table 3.2.1.4. Boarfish in ICES Subareas 27.6, 7, 8. Proxy catch numbers-at-age of the international catches (raised numbers in ‘000s) for the years 2007-2021

| Age | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0 | 0 | 1575 | 2415 | 0 | 28 | 301 | 0 | 5556 | 218 | 1862 | 314 | 17427 | 40397 | 4147 |
| 2 | 352 | 5488 | 15043 | 11229 | 2894 | 893 | 7148 | 695 | 116135 | 2385 | 4387 | 1736 | 37620 | 57719 | 21195 |
| 3 | 2114 | 21140 | 65744 | 72709 | 41913 | 5467 | 156680 | 49503 | 32248 | 10737 | 8830 | 2628 | 9737 | 37192 | 56256 |
| 4 | 40851 | 105575 | 338931 | 294382 | 28148 | 41278 | 58522 | 127520 | 16588 | 25114 | 34448 | 13610 | 9944 | 26433 | 78892 |
| 5 | 48915 | 141300 | 475619 | 567689 | 30116 | 110272 | 59797 | 93705 | 24564 | 20263 | 27266 | 15570 | 12682 | 10162 | 41988 |
| 6 | 62713 | 195339 | 543707 | 878363 | 175696 | 146582 | 68949 | 67275 | 26566 | 18025 | 21103 | 14731 | 12716 | 2583 | 16995 |
| 7 | 26132 | 104031 | 307333 | 522703 | 143967 | 492078 | 302967 | 193061 | 74115 | 61229 | 55189 | 38686 | 29513 | 9113 | 22437 |
| 8 | 29766 | 66570 | 172783 | 293719 | 107126 | 365840 | 250341 | 139124 | 52052 | 47573 | 38229 | 26821 | 18819 | 7487 | 8077 |
| 9 | 56075 | 53159 | 155477 | 276672 | 77861 | 271916 | 212318 | 121042 | 44615 | 42478 | 32258 | 23670 | 15875 | 7897 | 7021 |
| 10 | 44875 | 46893 | 130148 | 232122 | 60022 | 173486 | 160137 | 94225 | 34264 | 35150 | 25716 | 19395 | 11359 | 8164 | 5266 |
| 11 | 14019 | 15289 | 42521 | 78588 | 46079 | 69396 | 63025 | 36078 | 12999 | 13297 | 9560 | 7148 | 4272 | 3049 | 1818 |
| 12 | 32359 | 21178 | 61350 | 114600 | 40468 | 40968 | 41490 | 24895 | 9114 | 9132 | 7564 | 5846 | 2937 | 2786 | 1532 |
| 13 | 4848 | 11854 | 39609 | 59932 | 24352 | 58888 | 59380 | 36309 | 13362 | 13774 | 10922 | 8183 | 4256 | 4152 | 2316 |
| 14 | 16837 | 13570 | 31569 | 59060 | 19724 | 30277 | 30355 | 19064 | 7152 | 6682 | 5924 | 4554 | 2156 | 2333 | 1314 |
| 15+ | 109481 | 112947 | 196967 | 349320 | 157707 | 217260 | 239366 | 150688 | 59139 | 49589 | 40797 | 32130 | 14864 | 17663 | 10006 |

Table 3.2.2.1. Boarfish in ICES Subareas 27.6, 7, 8. Length-frequency distributions of the international catches (raised numbers in ‘000s) for the years 2007-2021

| Length | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4.5 |  |  |  |  |  |  |  |  | 14 |  |  |  |  |  |  |
| 5.0 |  |  |  |  |  |  |  |  | 878 |  |  |  |  |  |  |
| 5.5 |  |  |  |  |  |  |  |  | 515 |  |  |  |  | 2746 |  |
| 6.0 |  |  |  | 156 |  |  |  |  | 810 |  | 765 |  | 15868 | 37073 | 537 |
| 6.5 |  |  |  | 439 |  |  |  |  | 14 |  | 4607 | 203 | 70362 | 150810 | 2147 |
| 7.0 |  |  |  | 1090 | 522 | 56 | 52 |  | 513 | 417 | 5250 | 405 | 80160 | 233347 | 13936 |
| 7.5 |  |  | 1354 | 1574 |  |  | 551 |  | 10598 | 1684 | 12616 | 2635 | 85420 | 147915 | 25740 |
| 8.0 |  |  | 677 | 375 | 1345 | 185 | 1419 |  | 80716 | 8685 | 11473 | 4703 | 115154 | 38949 | 30699 |
| 8.5 |  |  |  | 1082 |  | 555 | 3592 | 1064 | 49508 | 6412 | 10115 | 3559 | 67471 | 43556 | 45234 |
| 9.0 |  |  | 677 | 5382 | 851 | 555 | 7263 | 327 | 10219 | 7104 | 3874 | 6554 | 16504 | 101918 | 107121 |
| 9.5 |  | 7473 | 17367 | 7883 | 7012 | 641 | 47509 | 4916 | 213 | 23065 | 14047 | 6196 | 3147 | 115103 | 191656 |
| 10.0 | 9609 | 11209 | 54130 | 29410 | 33243 | 2791 | 94702 | 31649 | 1211 | 46010 | 32346 | 5559 | 9173 | 100550 | 177751 |
| 10.5 |  | 52308 | 174796 | 130889 | 15848 | 6132 | 59833 | 71344 | 3865 | 39071 | 36242 | 4450 | 10144 | 55049 | 98863 |
| 11.0 | 84555 | 63517 | 343283 | 361774 | 70615 | 24571 | 18359 | 108261 | 12226 | 14181 | 32445 | 17658 | 5796 | 9475 | 72207 |
| 11.5 |  | 59781 | 321637 | 655875 | 93487 | 81928 | 20938 | 82470 | 28142 | 18249 | 31589 | 22826 | 22722 | 3172 | 44227 |
| 12.0 | 44199 | 119561 | 297737 | 739025 | 189434 | 264888 | 98564 | 84288 | 41613 | 30975 | 33618 | 24070 | 22353 | 2396 | 14710 |
| 12.5 |  | 70990 | 207739 | 564347 | 114904 | 398772 | 204868 | 112826 | 42461 | 51110 | 41650 | 24514 | 17521 | 3251 | 5711 |


| Length | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 13.0 | 82633 | 52308 | 147965 | 353484 | 133539 | 419060 | 315063 | 172416 | 59990 | 57000 | 46495 | 30665 | 28815 | 9494 | 6738 |
| 13.5 |  | 29890 | 149314 | 246146 | 51235 | 307533 | 285688 | 153742 | 52625 | 58696 | 43121 | 38698 | 16688 | 13707 | 8599 |
| 14.0 | 117224 | 22418 | 105782 | 224611 | 50857 | 176710 | 210137 | 138549 | 50139 | 76872 | 45353 | 34080 | 20053 | 16381 | 8468 |
| 14.5 |  | 14945 | 71273 | 127711 | 25309 | 89726 | 105571 | 74059 | 28771 | 37755 | 39524 | 29908 | 13809 | 14913 | 7389 |
| 15.0 | 65338 | 33627 | 47816 | 125463 | 25569 | 52791 | 62175 | 43347 | 16087 | 23137 | 21854 | 15561 | 5710 | 12563 | 7222 |
| 15.5 |  | 11209 | 13082 | 81386 | 5473 | 25065 | 31122 | 22629 | 8572 | 7841 | 4932 | 5778 | 1513 | 4304 | 2880 |
| 16.0 | 13452 | 11209 | 19397 | 24256 | 4181 | 13149 | 14990 | 7672 | 4331 | 625 | 1020 | 1948 | 143 | 1041 | 633 |
| 16.5 |  | 3736 | 4061 | 6209 | 2280 | 2738 | 4918 | 2134 | 2081 | 128 |  | 54 | 143 | 353 | 457 |
| 17.0 |  | 3736 | 677 | 1913 | 456 | 827 | 1109 | 1361 | 289 |  |  |  |  |  |  |
| 17.5 |  |  |  |  |  |  | 407 |  | 23 |  |  |  |  | 353 |  |
| 18.0 |  |  |  | 283 |  |  | 296 |  |  |  |  |  |  |  |  |
| 18.5 |  |  |  |  |  |  |  |  | 592 |  |  |  |  |  |  |

Table 3.3.1.1. Boarfish in ICES Subareas 27.6. 7, 8. Acoustic survey abundance and biomass estimates

| Age | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  | 1084 | 259 |  |
| 1 | 5 | 22 |  |  | 199 | 5 | 111 | 77 | 782 | 897 | 9523 | 587 |
| 2 | 12 | 11 | 78 |  | 319 | 36 | 127 | 31 | 389 | 1157 | 3392 | 3234 |
| 3 | 58 | 174 | 1843 | 15 | 17 | 46 | 345 | 115 | 97 | 967 | 2955 | 7537 |
| 4 | 187 | 65 | 696 | 98 | 34 | 44 | 367 | 68 | 93 | 113 | 1315 | 4259 |
| 5 | 437 | 95 | 382 | 102 | 80 | 6 | 156 | 107 | 88 | 157 | 463 | 619 |
| 6 | 1166 | 736 | 254 | 105 | 112 | 10 | 209 | 166 | 106 | 183 | 150 | 509 |
| 7 | 1184 | 974 | 1057 | 415 | 437 | 169 | 493 | 321 | 446 | 913 | 953 | 752 |
| 8 | 704 | 759 | 879 | 344 | 363 | 113 | 463 | 198 | 183 | 885 | 207 | 266 |
| 9 | 1095 | 849 | 801 | 342 | 354 | 118 | 397 | 293 | 288 | 721 | 378 | 302 |
| 10 | 1032 | 956 | 704 | 332 | 360 | 97 | 286 | 625 | 290 | 331 | 249 | 122 |
| 11 | 333 | 651 | 264 | 130 | 132 | 17 | 121 | 339 | 50 | 81 | 151 | 41 |
| 12 | 653 | 1100 | 203 | 105 | 113 | 32 | 82 | 264 | 192 | 195 | 188 | 23 |
| 13 | 336 | 857 | 297 | 166 | 174 | 49 | 74 | 198 | 79 | 299 | 81 | 127 |
| 14 | 385 | 656 | 170 | 89 | 108 | 18 | 220 | 117 | 57 | 267 | 327 | 90 |
| 15+ | 3519 | 6354 | 1464 | 855 | 1195 | 400 | 931 | 302 | 759 | 1641 | 1213 | 148 |
| $\begin{aligned} & \text { TS } \\ & \mathrm{N} \end{aligned}$ | 11104 | 14257 | 9091 | 3098 | 3996 | 1157 | 4387 | 3221 | 3899 | 9888 | 21805 | 18614 |
| TSB | $\begin{aligned} & 67017 \\ & 6 \end{aligned}$ | $\begin{aligned} & 86344 \\ & 6 \end{aligned}$ | $\begin{aligned} & 43989 \\ & 0 \end{aligned}$ | $\begin{aligned} & 18777 \\ & 9 \end{aligned}$ | $\begin{aligned} & 23263 \\ & 4 \end{aligned}$ | $\begin{aligned} & 6969 \\ & 0 \end{aligned}$ | $\begin{aligned} & 23006 \\ & 2 \end{aligned}$ | $\begin{aligned} & 18625 \\ & 2 \end{aligned}$ | $\begin{aligned} & 17915 \\ & 6 \end{aligned}$ | $\begin{aligned} & 39987 \\ & 2 \end{aligned}$ | $\begin{aligned} & 44377 \\ & 7 \end{aligned}$ | $\begin{aligned} & 45141 \\ & 5 \end{aligned}$ |
| SSB | $\begin{aligned} & 66939 \\ & 2 \end{aligned}$ | $\begin{aligned} & 86154 \\ & 4 \end{aligned}$ | $\begin{aligned} & 42315 \\ & 8 \end{aligned}$ | $\begin{aligned} & 18765 \\ & 4 \end{aligned}$ | $\begin{aligned} & 22665 \\ & 9 \end{aligned}$ | $\begin{aligned} & 6910 \\ & 3 \end{aligned}$ | $\begin{aligned} & 21881 \\ & 0 \end{aligned}$ | $\begin{aligned} & 18462 \\ & 4 \end{aligned}$ | $\begin{aligned} & 16921 \\ & 3 \end{aligned}$ | $\begin{aligned} & 35787 \\ & 1 \end{aligned}$ | $\begin{aligned} & 35195 \\ & 5 \end{aligned}$ | $\begin{aligned} & 42272 \\ & 2 \end{aligned}$ |
| CV | 21.2 | 10.6 | 17.5 | 15.1 | 17.0 | 19 | 21.9 | 19.9 | 25.4 | 34.8 | 31.0 | 24.0 |

Table 3.3.2.1. Boarfish in ICES Subareas 27.6, 7, 8. IBTS length-frequency data
EVHOE

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 0 | 5 | 12 | 7 | 17 | 195 | 2645 | 5006 | 3691 | 3570 | 4422 | 12054 | 16633 | 7200 | 3472 | 503 | 18 | 1 | 0 | 0 |
| 1998 | 0 | 1 | 4 | 25 | 70 | 2083 | 18263 | 8566 | 6117 | 5961 | 7082 | 11828 | 14363 | 9600 | 5261 | 971 | 8 | 0 | 0 | 1 |
| 1999 | 0 | 0 | 13 | 52 | 33 | 245 | 10949 | 25911 | 23235 | 6484 | 2818 | 4632 | 7780 | 6151 | 1357 | 268 | 8 | 0 | 0 | 0 |
| 2000 | 0 | 17 | 79 | 120 | 8 | 1508 | 26901 | 17725 | 9864 | 22076 | 16424 | 29584 | 36849 | 16508 | 5399 | 988 | 76 | 0 | 0 | 0 |
| 2001 | 0 | 1 | 45 | 687 | 490 | 916 | 21328 | 37173 | 13322 | 28492 | 31640 | 18378 | 12315 | 6507 | 3193 | 1272 | 81 | 4 | 0 | 0 |
| 2002 | 0 | 2 | 18 | 23 | 11 | 547 | 9634 | 29844 | 17728 | 13175 | 9280 | 9513 | 9615 | 6185 | 2458 | 642 | 37 | 1 | 1 | 0 |
| 2003 | 0 | 0 | 17 | 47 | 17 | 57 | 426 | 1663 | 7155 | 20073 | 24977 | 21358 | 21939 | 15004 | 7355 | 1599 | 35 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 33 | 534 | 397 | 123 | 1248 | 1420 | 1308 | 1083 | 3102 | 7308 | 7224 | 6353 | 7866 | 3630 | 241 | 5 | 0 | 0 |
| 2005 | 0 | 2 | 94 | 964 | 1264 | 146 | 1097 | 2302 | 1225 | 1551 | 3182 | 13394 | 15782 | 9879 | 6012 | 1658 | 117 | 70 | 0 | 0 |
| 2006 | 1 | 26 | 111 | 77 | 74 | 15506 | 37545 | 10729 | 3611 | 2128 | 1518 | 1960 | 4165 | 4024 | 2601 | 940 | 93 | 2 | 12 | 0 |
| 2007 | 0 | 7 | 188 | 473 | 234 | 1511 | 22812 | 127331 | 65589 | 6442 | 6823 | 5477 | 6110 | 6003 | 4268 | 1411 | 118 | 11 | 0 | 0 |
| 2008 | 0 | 3 | 432 | 2795 | 823 | 5487 | 54355 | 256210 | 169633 | 163128 | 69199 | 38406 | 18310 | 17213 | 9157 | 3486 | 745 | 6 | 1 | 0 |
| 2009 | 0 | 6 | 128 | 194 | 69 | 1482 | 19663 | 35649 | 5260 | 3906 | 9562 | 12271 | 9402 | 10835 | 6722 | 775 | 39 | 1 | 0 | 0 |
| 2010 | 0 | 21 | 529 | 116 | 154 | 5774 | 46490 | 74999 | 27177 | 12168 | 37971 | 59369 | 38501 | 37683 | 15699 | 1555 | 248 | 8 | 1 | 0 |
| 2011 | 0 | 61 | 95 | 214 | 5 | 536 | 2232 | 8210 | 14905 | 32671 | 29788 | 50316 | 56963 | 36588 | 11723 | 3058 | 572 | 159 | 47 | 0 |
| 2012 | 0 | 9 | 146 | 594 | 142 | 2913 | 28823 | 26800 | 6124 | 11739 | 13607 | 22370 | 37138 | 44084 | 19963 | 4893 | 127 | 1 | 0 | 0 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2013 | 0 | 3 | 48 | 92 | 10 | 305 | 2187 | 2141 | 2558 | 13769 | 9938 | 15006 | 37563 | 40266 | 20130 | 6888 | 686 | 0 | 3 | 0 |
| 2014 | 0 | 2 | 693 | 1386 | 508 | 84 | 1440 | 885 | 3074 | 8732 | 28586 | 39397 | 74122 | 69736 | 26871 | 3908 | 59 | 433 | 0 | 0 |
| 2015 | 0 | 5 | 183 | 5898 | 4143 | 607 | 19075 | 179269 | 119004 | 15765 | 18014 | 61575 | 62024 | 59904 | 21525 | 5487 | 541 | 429 | 8 | 0 |
| 2016 | 5 | 31 | 379 | 846 | 115 | 733 | 10284 | 14280 | 17251 | 42132 | 25304 | 68583 | 130633 | 131220 | 48538 | 11611 | 1358 | 26 | 0 | 0 |
| 2018 | 0 | 14 | 4957 | 193861 | 173779 | 210 | 10910 | 76288 | 48343 | 29096 | 45773 | 85164 | 132174 | 157883 | 48603 | 14951 | 592 | 18 | 0 | 0 |
| 2019 | 2 | 997 | 6467 | 589 | 10688 | 531908 | 561517 | 329850 | 59733 | 4505 | 3418 | 8451 | 32547 | 61582 | 30031 | 7468 | 962 | 204 | 0 | 0 |
| 2020 | 3 | 283 | 1280 | 657 | 21381 | 408706 | 595107 | 142947 | 218153 | 421028 | 220190 | 54726 | 70612 | 97364 | 74415 | 30606 | 4736 | 1 | 0 | 0 |
| 2021 | 0 | 35 | 166 | 27 | 32861 | 954046 | 852223 | 313053 | 640456 | 208802 | 106995 | 57674 | 96633 | 65504 | 12047 | 3416 | 387 | 53 | 0 | 0 |

IGFS

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 0 | 1 | 33 | 22 | 7 | 22 | 129 | 172 | 879 | 2942 | 2322 | 1325 | 3823 | 4629 | 2898 | 896 | 163 | 38 | 0 | 0 |
| 2004 | 0 | 23 | 63 | 34 | 8 | 117 | 628 | 1444 | 423 | 397 | 464 | 2276 | 4325 | 4709 | 3972 | 1019 | 90 | 5 | 1 | 0 |
| 2005 | 0 | 8 | 59 | 52 | 20 | 203 | 1024 | 585 | 288 | 636 | 341 | 3463 | 11457 | 11348 | 7955 | 1744 | 382 | 2 | 1 | 0 |
| 2006 | 5 | 60 | 68 | 48 | 35 | 212 | 969 | 621 | 2046 | 4190 | 8044 | 7946 | 24208 | 42119 | 32168 | 12296 | 2454 | 532 | 0 | 0 |
| 2007 | 1 | 6 | 44 | 18 | 31 | 501 | 923 | 1251 | 1638 | 1166 | 2510 | 3581 | 8275 | 10740 | 7093 | 1934 | 92 | 0 | 0 | 0 |
| 2008 | 0 | 0 | 26 | 18 | 23 | 127 | 672 | 531 | 2095 | 13780 | 17664 | 19268 | 16980 | 19484 | 15953 | 8789 | 1747 | 76 | 1 | 0 |
| 2009 | 0 | 3 | 80 | 76 | 25 | 94 | 228 | 486 | 1000 | 1139 | 9081 | 7749 | 5138 | 6921 | 5592 | 1084 | 68 | 1 | 0 | 0 |
| 2010 | 0 | 6 | 42 | 3 | 18 | 199 | 272 | 463 | 920 | 393 | 7914 | 34236 | 28611 | 16063 | 8161 | 1974 | 433 | 0 | 0 | 0 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 0 | 7 | 17 | 5 | 4 | 189 | 772 | 592 | 556 | 669 | 2600 | 20246 | 22121 | 10851 | 5319 | 2218 | 269 | 9 | 6 | 0 |
| 2012 | 0 | 7 | 36 | 20 | 10 | 130 | 271 | 378 | 702 | 2143 | 1183 | 11104 | 34005 | 22731 | 10905 | 3901 | 525 | 4 | 0 | 0 |
| 2013 | 1 | 3 | 9 | 9 | 20 | 127 | 352 | 340 | 1320 | 2833 | 3971 | 15572 | 51637 | 52868 | 20485 | 6560 | 492 | 20 | 0 | 0 |
| 2014 | 0 | 10 | 68 | 54 | 4 | 18 | 13 | 25 | 60 | 130 | 1127 | 3251 | 19125 | 23016 | 10355 | 2988 | 284 | 18 | 0 | 0 |
| 2015 | 0 | 3 | 11 | 16 | 24 | 193 | 1008 | 3708 | 848 | 105 | 713 | 6315 | 29727 | 48220 | 33024 | 17350 | 1885 | 531 | 0 | 0 |
| 2016 | 4 | 31 | 121 | 63 | 7 | 67 | 187 | 1515 | 4057 | 2891 | 1349 | 4111 | 32753 | 57753 | 40907 | 15527 | 3670 | 85 | 0 | 0 |
| 2017 | 0 | 0 | 37 | 131 | 48 | 132 | 460 | 652 | 11411 | 20321 | 5909 | 5520 | 16426 | 33117 | 29972 | 15815 | 3194 | 369 | 0 | 0 |
| 2018 | 4 | 51 | 247 | 139 | 32 | 45 | 286 | 585 | 1194 | 6107 | 17005 | 15168 | 48895 | 61833 | 36519 | 10722 | 2030 | 63 | 0 | 0 |
| 2019 | 4 | 19 | 117 | 47 | 52 | 262 | 583 | 173 | 106 | 487 | 2677 | 4967 | 6863 | 12080 | 10480 | 5125 | 772 | 71 | 4 | 0 |
| 2020 | 9 | 388 | 233 | 21 | 16 | 1772 | 2052 | 13941 | 65121 | 24505 | 7709 | 17859 | 12157 | 17223 | 9125 | 2499 | 110 | 2 | 0 | 0 |
| 2021 | 2 | 7 | 98 | 36 | 293 | 16275 | 125036 | 87742 | 210710 | 171970 | 67893 | 20086 | 16044 | 22040 | 23112 | 4589 | 816 | 7 | 1 | 0 |

## SPNGFS

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 0 | 0 | 8 | 0 | 16 | 317 | 1817 | 2496 | 260 | 141 | 154 | 314 | 632 | 613 | 689 | 97 | 0 | 0 | 0 | 0 |
| 1991 | 0 | 1 | 0 | 0 | 31 | 690 | 1311 | 313 | 49 | 9 | 6 | 7 | 7 | 4 | 0 | 0 | 0 | 6 | 0 | 0 |
| 1992 | 0 | 57 | 38 | 9 | 178 | 3290 | 2743 | 282 | 48 | 10 | 8 | 69 | 162 | 390 | 779 | 246 | 95 | 0 | 0 | 0 |
| 1993 | 0 | 57 | 1206 | 488 | 97 | 3730 | 3753 | 421 | 105 | 54 | 7 | 4 | 8 | 3 | 2 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 1 | 40 | 33 | 0 | 342 | 4789 | 10162 | 8920 | 3195 | 53 | 106 | 20 | 9 | 12 | 1 | 0 | 0 | 0 | 0 | 0 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 0 | 84 | 108 | 4 | 342 | 3063 | 2157 | 220 | 84 | 65 | 58 | 105 | 105 | 90 | 20 | 4 | 0 | 0 | 0 | 0 |
| 1996 | 0 | 218 | 537 | 143 | 245 | 4457 | 4449 | 267 | 820 | 722 | 82 | 145 | 126 | 219 | 96 | 39 | 2 | 0 | 0 | 0 |
| 1997 | 2 | 102 | 809 | 441 | 235 | 3458 | 6824 | 2189 | 1923 | 534 | 156 | 353 | 161 | 88 | 3 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 3 | 2 | 7 | 4 | 49 | 1920 | 4685 | 2217 | 337 | 153 | 125 | 88 | 147 | 135 | 86 | 13 | 2 | 3 | 0 | 0 |
| 1999 | 0 | 6 | 59 | 13 | 134 | 2736 | 3010 | 193 | 106 | 83 | 109 | 143 | 390 | 645 | 402 | 69 | 0 | 0 | 0 | 0 |
| 2000 | 0 | 7 | 3729 | 2046 | 17 | 554 | 1947 | 489 | 277 | 486 | 756 | 1252 | 999 | 1021 | 199 | 34 | 13 | 0 | 0 | 0 |
| 2001 | 0 | 68 | 4 | 1 | 153 | 3241 | 5085 | 659 | 225 | 206 | 205 | 236 | 692 | 407 | 120 | 22 | 9 | 0 | 0 | 0 |
| 2002 | 0 | 4 | 20 | 0 | 133 | 2333 | 2013 | 284 | 50 | 58 | 54 | 60 | 231 | 314 | 72 | 9 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 4 | 950 | 567 | 4 | 77 | 221 | 57 | 39 | 28 | 16 | 22 | 17 | 23 | 16 | 5 | 1 | 0 | 0 | 0 |
| 2004 | 0 | 6 | 22 | 4 | 43 | 2289 | 3808 | 443 | 110 | 83 | 58 | 219 | 931 | 776 | 303 | 2 | 1 | 0 | 0 | 0 |
| 2005 | 0 | 16 | 451 | 25 | 9 | 754 | 1007 | 207 | 85 | 102 | 30 | 54 | 257 | 218 | 90 | 44 | 2 | 0 | 0 | 0 |
| 2006 | 0 | 14 | 156 | 160 | 50 | 2238 | 8913 | 4507 | 175 | 94 | 9 | 36 | 229 | 419 | 169 | 9 | 2 | 0 | 0 | 0 |
| 2007 | 0 | 49 | 40 | 1 | 111 | 3025 | 6620 | 1099 | 129 | 260 | 81 | 7 | 93 | 215 | 89 | 21 | 3 | 0 | 0 | 0 |
| 2008 | 7 | 4 | 92 | 247 | 1 | 936 | 1561 | 1326 | 234 | 1483 | 304 | 537 | 11 | 833 | 201 | 186 | 11 | 0 | 0 | 0 |
| 2009 | 1 | 17 | 62 | 119 | 11 | 2587 | 3893 | 4070 | 119 | 250 | 45 | 142 | 59 | 819 | 120 | 17 | 1 | 1 | 0 | 0 |
| 2010 | 0 | 55 | 102 | 5 | 232 | 13090 | 22032 | 3169 | 1160 | 1056 | 89 | 82 | 179 | 1007 | 1981 | 518 | 9 | 0 | 0 | 0 |
| 2011 | 0 | 29 | 260 | 105 | 46 | 2805 | 5511 | 1278 | 148 | 340 | 145 | 100 | 144 | 591 | 724 | 134 | 3 | 1 | 0 | 0 |
| 2012 | 0 | 29 | 132 | 35 | 556 | 7550 | 7844 | 1364 | 88 | 53 | 59 | 170 | 1051 | 2394 | 1553 | 432 | 21 | 0 | 0 | 0 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2013 | 0 | 0 | 2 | 11 | 126 | 2163 | 4664 | 854 | 302 | 609 | 251 | 61 | 113 | 134 | 156 | 81 | 8 | 0 | 0 | 0 |
| 2014 | 0 | 75 | 117 | 6 | 12 | 263 | 465 | 79 | 1083 | 1175 | 1174 | 1266 | 998 | 2444 | 3623 | 817 | 31 | 1 | 0 | 0 |
| 2015 | 0 | 13 | 67 | 3 | 58 | 1889 | 4248 | 534 | 75 | 465 | 750 | 970 | 695 | 1173 | 1473 | 453 | 70 | 1 | 0 | 0 |
| 2016 | 0 | 17 | 99 | 5 | 41 | 922 | 2423 | 473 | 925 | 746 | 346 | 548 | 452 | 561 | 169 | 22 | 4 | 0 | 0 | 0 |
| 2017 | 1 | 23 | 20 | 1 | 16 | 641 | 1947 | 755 | 134 | 165 | 285 | 405 | 579 | 967 | 936 | 177 | 13 | 3 | 0 | 0 |
| 2018 | 0 | 0 | 2 | 0 | 45 | 708 | 1635 | 258 | 43 | 99 | 230 | 605 | 1370 | 3324 | 3865 | 949 | 3 | 0 | 0 | 2 |
| 2019 | 0 | 12 | 2 | 1 | 259 | 4128 | 3887 | 379 | 18 | 83 | 273 | 329 | 717 | 4200 | 8402 | 2215 | 202 | 0 | 0 | 0 |
| 2020 | 0 | 8 | 33 | 2 | 33 | 1218 | 2123 | 525 | 387 | 314 | 75 | 225 | 705 | 2518 | 4751 | 1603 | 10 | 0 | 0 | 0 |
| 2021 | 1 | 10 | 11 | 0 | 42 | 803 | 2654 | 562 | 127 | 1367 | 3149 | 1102 | 2200 | 4773 | 6485 | 1175 | 118 | 1 | 0 | 0 |

SPPGFS

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 0 | 1 | 0 | 1 | 1 | 2 | 0 | 44 | 5 | 52 | 133 | 162 | 667 | 1129 | 230 | 40 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 90 | 212 | 791 | 843 | 313 | 60 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 15 | 22 | 21 | 62 | 268 | 426 | 249 | 51 | 2 | 1 | 0 | 0 |
| 2004 | 0 | 1 | 0 | 0 | 0 | 6 | 3 | 0 | 5 | 6 | 23 | 124 | 385 | 592 | 390 | 52 | 1 | 0 | 0 | 0 |
| 2005 | 0 | 1 | 0 | 1 | 8 | 1 | 20 | 11 | 10 | 16 | 8 | 118 | 628 | 1118 | 833 | 272 | 23 | 0 | 0 | 0 |
| 2006 | 0 | 0 | 1 | 1 | 8 | 120 | 118 | 26 | 43 | 95 | 34 | 58 | 431 | 863 | 716 | 252 | 13 | 1 | 0 | 0 |
| 2007 | 0 | 0 | 0 | 0 | 4 | 5 | 12 | 20 | 16 | 12 | 37 | 34 | 96 | 202 | 191 | 34 | 5 | 0 | 0 | 0 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 0 | 1 | 0 | 0 | 0 | 1 | 17 | 10 | 23 | 19 | 79 | 156 | 349 | 666 | 442 | 113 | 7 | 0 | 0 | 0 |
| 2009 | 0 | 8 | 7 | 0 | 3 | 10 | 11 | 1 | 0 | 2 | 220 | 457 | 1333 | 1746 | 1698 | 474 | 11 | 0 | 0 | 0 |
| 2010 | 2 | 0 | 0 | 1 | 6 | 17 | 4 | 1 | 6 | 3 | 43 | 390 | 710 | 976 | 620 | 164 | 13 | 0 | 0 | 0 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 20 | 22 | 6 | 180 | 815 | 960 | 522 | 151 | 17 | 0 | 2 | 0 |
| 2012 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 2 | 2 | 1 | 10 | 87 | 456 | 570 | 267 | 79 | 4 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 0 | 1 | 0 | 8 | 24 | 7 | 10 | 0 | 1 | 48 | 500 | 1032 | 564 | 163 | 15 | 1 | 0 | 0 |
| 2014 | 0 | 10 | 9 | 0 | 1 | 0 | 3 | 17 | 62 | 11 | 6 | 85 | 2453 | 6703 | 3168 | 2115 | 162 | 82 | 0 | 0 |
| 2015 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 32 | 300 | 471 | 316 | 151 | 43 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 3 | 0 | 0 | 0 | 1 | 0 | 13 | 7 | 0 | 9 | 157 | 336 | 220 | 84 | 19 | 0 | 0 | 0 |
| 2017 | 0 | 67 | 19 | 0 | 0 | 0 | 10 | 0 | 0 | 1 | 18 | 26 | 148 | 498 | 529 | 268 | 17 | 0 | 0 | 0 |
| 2018 | 0 | 2 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 37 | 1159 | 3574 | 2449 | 1131 | 159 | 0 | 0 | 0 |
| 2019 | 5 | 36 | 4 | 0 | 0 | 0 | 0 | 0 | 3 | 4 | 0 | 15 | 426 | 952 | 796 | 192 | 15 | 0 | 0 | 0 |
| 2020 | 0 | 5 | 1 | 0 | 0 | 4 | 1 | 1 | 2 | 4 | 0 | 26 | 250 | 616 | 851 | 661 | 111 | 0 | 0 | 1 |
| 2021 | 1 | 20 | 0 | 0 | 5 | 12 | 0 | 5 | 34 | 38 | 24 | 39 | 129 | 916 | 768 | 357 | 147 | 3 | 0 | 0 |

WCSGFS

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 0 | 0 | 3 | 2 | 0 | 3 | 24 | 42 | 62 | 172 | 210 | 1286 | 856 | 450 | 52 | 17 | 0 | 0 | 0 | 0 |
| 1991 | 0 | 0 | 0 | 2 | 0 | 31 | 138 | 80 | 183 | 644 | 683 | 848 | 226 | 89 | 12 | 1 | 2 | 4 | 0 | 0 |
| 1992 | 0 | 0 | 0 | 1 | 0 | 8 | 12 | 14 | 44 | 478 | 1160 | 4028 | 1674 | 502 | 5 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 0 | 0 | 0 | 0 | 0 | 1 | 109 | 2 | 670 | 2078 | 1074 | 4904 | 2753 | 2882 | 28 | 2 | 0 | 0 | 0 | 0 |
| 1994 | 0 | 0 | 2 | 0 | 0 | 0 | 15 | 30 | 30 | 205 | 283 | 312 | 454 | 388 | 147 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 8 | 12 | 18 | 4 | 2 | 10 | 40 | 30 | 94 | 162 | 640 | 1485 | 1770 | 1139 | 318 | 14 | 2 | 4 | 6 | 0 |
| 1996 | 0 | 0 | 0 | 4 | 0 | 10 | 48 | 27 | 49 | 48 | 64 | 188 | 920 | 1888 | 416 | 18 | 1 | 0 | 0 | 0 |
| 1997 | 0 | 0 | 4 | 0 | 0 | 1 | 17 | 42 | 120 | 64 | 116 | 249 | 436 | 301 | 91 | 8 | 4 | 0 | 0 | 0 |
| 1998 | 0 | 0 | 0 | 1 | 0 | 1 | 7 | 6 | 7 | 16 | 47 | 69 | 105 | 171 | 78 | 8 | 2 | 0 | 0 | 0 |
| 1999 | 0 | 0 | 1 | 0 | 0 | 2 | 6 | 8 | 189 | 221 | 312 | 458 | 346 | 221 | 69 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 3 | 42 | 118 | 230 | 303 | 206 | 108 | 54 | 8 | 0 | 0 | 0 | 0 |
| 2001 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 12 | 27 | 54 | 90 | 233 | 414 | 242 | 80 | 15 | 1 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 1 | 8 | 2 | 1 | 82 | 759 | 3243 | 5711 | 5896 | 1558 | 189 | 1 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 1 | 0 | 0 | 0 | 3 | 52 | 9 | 107 | 326 | 1536 | 3294 | 5409 | 3553 | 413 | 37 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 1 | 0 | 0 | 6 | 2 | 45 | 83 | 744 | 4576 | 8611 | 9526 | 5698 | 954 | 84 | 0 | 0 | 0 |
| 2005 | 0 | 2 | 0 | 0 | 0 | 9 | 38 | 15 | 30 | 31 | 113 | 442 | 1115 | 1747 | 818 | 141 | 9 | 3 | 2 | 0 |


| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5}$ | $\mathbf{1 6}$ | $\mathbf{1 7}$ | $\mathbf{1 8}$ | $\mathbf{1 9}$ | $\mathbf{2 0}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2006 | 0 | 1 | 2 | 1 | 0 | 2 | 9 | 4 | 22 | 256 | 311 | 508 | 1524 | 2964 | 2104 | 449 | 73 | 2 | 0 | 0 |
| 2007 | 0 | 0 | 3 | 2 | 0 | 8 | 14 | 65 | 118 | 182 | 795 | 2938 | 5220 | 6953 | 5332 | 1538 | 116 | 0 | 0 | 0 |
| 2008 | 0 | 1 | 3 | 0 | 0 | 16 | 37 | 38 | 200 | 482 | 1406 | 3218 | 9904 | 22777 | 18407 | 6293 | 575 | 71 | 0 | 0 |
| 2009 | 0 | 0 | 1 | 0 | 1 | 1 | 4 | 6 | 64 | 2460 | 2246 | 694 | 505 | 416 | 338 | 136 | 12 | 0 | 0 | 0 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 530 | 1443 | 1384 | 1357 | 828 | 149 | 29 | 0 | 0 | 0 |

Table 3.6.1.1. Boarfish in ICES Subareas 27.6, 7, 8. IBTS length-frequency data converted to age-structured indices by application of the 2012 common ALK rounded down to 1 cm length classes

EVHOE

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 1323 | 5891 | 4835 | 3829 | 3369 | 3053 | 9614 | 6955 | 5556 | 3779 | 1521 | 973 | 1456 | 828 | 6235 |
| 1998 | 9132 | 16881 | 8109 | 6147 | 4527 | 3452 | 9545 | 6632 | 5452 | 4058 | 1597 | 1312 | 1733 | 1022 | 8419 |
| 1999 | 5474 | 30494 | 25366 | 5015 | 2592 | 1427 | 4373 | 3215 | 2887 | 2276 | 855 | 564 | 888 | 491 | 3675 |
| 2000 | 13450 | 28555 | 16758 | 19454 | 12310 | 8420 | 23424 | 16159 | 12783 | 8538 | 3354 | 1885 | 3099 | 1722 | 12485 |
| 2001 | 10664 | 39887 | 26874 | 27998 | 16428 | 8946 | 15285 | 7816 | 5688 | 3538 | 1301 | 863 | 1271 | 750 | 6396 |
| 2002 | 4817 | 30622 | 24313 | 11299 | 6215 | 3393 | 7688 | 4838 | 3852 | 2716 | 1035 | 726 | 1060 | 611 | 4928 |
| 2003 | 213 | 3707 | 9293 | 20716 | 13365 | 8409 | 18107 | 11109 | 8937 | 6448 | 2467 | 1932 | 2635 | 1547 | 12700 |
| 2004 | 624 | 2006 | 1574 | 1777 | 1923 | 1842 | 5376 | 3816 | 3078 | 2541 | 1075 | 1423 | 1434 | 932 | 11369 |
| 2005 | 549 | 2492 | 1901 | 2205 | 2758 | 2983 | 9853 | 7261 | 5865 | 4310 | 1727 | 1437 | 1869 | 1110 | 9951 |
| 2006 | 18772 | 27129 | 6395 | 1838 | 1086 | 692 | 2217 | 1683 | 1593 | 1407 | 557 | 586 | 688 | 416 | 4256 |
| 2007 | 11406 | 118156 | 87434 | 6252 | 3796 | 2250 | 4968 | 3140 | 2686 | 2208 | 861 | 923 | 1067 | 657 | 6591 |
| 2008 | 27177 | 254528 | 229646 | 124210 | 54539 | 19047 | 30818 | 15021 | 10954 | 7348 | 2618 | 2251 | 2934 | 1795 | 16959 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 9832 | 35351 | 16200 | 5643 | 4832 | 3830 | 8969 | 5783 | 4721 | 3809 | 1459 | 1524 | 1806 | 1110 | 9216 |
| 2010 | 23245 | 82303 | 45710 | 20517 | 19648 | 16749 | 39369 | 25075 | 19324 | 14156 | 5280 | 4343 | 5906 | 3511 | 26732 |
| $2011$ | 1116 | 11557 | 19043 | 30617 | 20479 | 14495 | 39161 | 26846 | 21792 | 15613 | 5980 | 3928 | 6016 | 3404 | 27139 |
| 2012 | 14412 | 34320 | 15329 | 11984 | 8843 | 6877 | 21882 | 16580 | $15805$ | 14165 | 5382 | 5221 | 6581 | 3893 | 34397 |
| $2013$ | $1093$ | 3373 | 5082 | 11975 | $7436$ | $5156$ | 18526 | 14722 | 14572 | 13248 | 5121 | 5049 | 6254 | 3703 | 35819 |
| 2014 | 720 | 2334 | 4216 | 15081 | 14776 | 13252 | 40953 | 30549 | 28568 | 24182 | 9208 | 7776 | 10517 | 6071 | 49039 |
| $2015$ | 9537 | 168718 | 142196 | 16589 | 15129 | 14025 | 43805 | 31952 | 26892 | 21239 | 8025 | 6461 | 8982 | 5218 | 43843 |
| 2016 | 5142 | 20412 | 24368 | 35467 | 23775 | 18507 | 68150 | 53795 | 50979 | 44038 | 16743 | 14289 | 19326 | 11149 | 95082 |
| 2018 | 5455 | 72428 | 63489 | 33998 | 28889 | 24760 | 79148 | 59901 | 56898 | 49999 | 18526 | 15688 | 21690 | 12453 | 106474 |
| 2019 | 280759 | 520569 | 150645 | 4035 | 3104 | 2844 | 14950 | 13581 | 15700 | 16891 | 6358 | 7404 | 8669 | 5219 | 49538 |
| 2020 | 297553 | 465569 | 273832 | 332726 | 148543 | 51435 | 79125 | 38909 | 36296 | 32676 | 12326 | 15407 | 16693 | 10460 | 118335 |
| 2021 | 426111 | 848299 | 571349 | 164881 | 76916 | 31315 | 65603 | 40367 | 35579 | 26598 | 9833 | 5812 | 9725 | 5289 | 39566 |

IGFS

| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 64 | 472 | 1214 | 2586 | 1401 | 743 | 2065 | 1523 | 1556 | 1484 | 578 | 653 | 750 | 456 | 4672 |
| 2004 | 314 | 1418 | 842 | 434 | 493 | 543 | 2252 | 1838 | 1732 | 1603 | 653 | 802 | 864 | 541 | 5422 |
| 2005 | 512 | 998 | 509 | 567 | 717 | 908 | 4790 | 4166 | 4162 | 3867 | 1557 | 1730 | 1973 | 1201 | 11568 |
| 2006 | 484 | 1580 | 2423 | 5269 | 4211 | 3388 | 12623 | 10487 | 11436 | 12263 | 4853 | 6606 | 6952 | 4368 | 50651 |
| 2007 | 462 | 1842 | 1748 | 1576 | 1408 | 1235 | 4362 | 3474 | 3496 | 3378 | 1326 | 1557 | 1754 |  |  |
| 2008 | 336 | 1388 | 4302 | 14466 | 9811 | 6581 | 15265 | 9859 | 8231 | 6912 | 2728 | 3247 | 3553 | 2238 | 28119 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 114 | 772 | 1117 | 3682 | 3665 | 2967 | 5991 | 3553 | 2883 | 2398 | 928 | 1136 | 1233 | 783 | 7266 |
| 2010 | 136 | 752 | 906 | 3336 | 6161 | 7220 | 21721 | 15262 | 11417 | 7656 | 3025 | 2151 | 3055 | 1795 | 14845 |
| 2011 | 386 | 966 | 715 | 1598 | 3198 | 4038 | 13856 | 10232 | 7932 | 5384 | 2159 | 1453 | 2121 | 1224 | 10962 |
| 2012 | 136 | 622 | 1006 | 1911 | 2306 | 2843 | 13844 | 11639 | 10956 | 8966 | 3576 | 2903 | 3900 | 2242 | 21003 |
| 2013 | 176 | 843 | 1557 | 3292 | 3917 | 4545 | 21801 | 18670 | 19029 | 17278 | 6613 | 5870 | 7777 | 4484 | 40599 |
| 2014 | 6 | 43 | 82 | 492 | 927 | 1262 | 7300 | 6613 | 7255 | 7083 | 2717 | 2714 | 3384 | 1986 | 18529 |
| 2015 | 504 | 3259 | 1827 | 403 | 1251 | 1945 | 12476 | 11625 | 13072 | 13999 | 5512 | 7082 | 7697 | 4765 | 58017 |
| 2016 | 93 | 2456 | 3763 | 2302 | 1775 | 1846 | 13082 | 12553 | 14753 | 16394 | 6464 | 8634 | 9226 | 5742 | 65723 |
| 2017 | 230 | 4468 | 11683 | 14642 | 6277 | 2402 | 9024 | 7578 | 8395 | 9474 | 3824 | 5785 | 5766 | 3703 | 49915 |
| 2018 | 143 | 930 | 2275 | 9391 | 8194 | 6861 | 23782 | 19030 | 19873 | 19320 | 7511 | 8412 | 9756 | 5903 | 59025 |
| 2019 | 292 | 442 | 242 | 1229 | 1449 | 1419 | 4664 | 3618 | 3540 | 3626 | 1453 | 2058 | 2107 | 1346 | 16899 |
| 2020 | 1026 | 32027 | 52719 | 18043 | 8761 | 4356 | 11714 | 8061 | 6664 | 5578 | 2105 | 2193 | 2649 | 1618 | 14790 |
| 2021 | 62518 | 191249 | 202522 | 128995 | 53951 | 16137 | 23800 | 10942 | 9297 | 7968 | 3069 | 4310 | 4329 | 2815 | 28141 |

SPNGFS

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 909 | 2660 | 1033 | 142 | 110 | 93 | 335 | 263 | 243 | 224 | 95 | 128 | 129 | 83 | 770 |
| 1991 | 656 | 880 | 138 | 8 | 4 | 2 | 6 | 3 | 3 | 2 | 1 | 0 | 1 | 0 | 8 |
| 1992 | 1371 | 1575 | 128 | 10 | 13 | 16 | 97 | 89 | 92 | 122 | 57 | 124 | 102 | 71 | 965 |
| 1993 | 1877 | 2192 | 220 | 36 | 13 | 2 | 5 | 3 | 2 | 2 | 1 | 0 | 1 | 0 | 3 |
| 1994 | 5081 | 12093 | 5114 | 66 | 43 | 23 | 28 | 9 | 7 | 5 | 1 | 1 | 1 | 1 | 5 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 1079 | 1254 | 142 | 61 | 41 | 29 | 78 | 54 | 44 | 33 | 12 | 8 | 13 | 7 | 53 |
| 1996 | 2225 | 2676 | 772 | 479 | 175 | 40 | 109 | 77 | 70 | 65 | 24 | 25 | 31 | 18 | 181 |
| 1997 | 3412 | 5512 | 2113 | 389 | 183 | 84 | 198 | 123 | 82 | 47 | 17 | 6 | 14 | 8 | 43 |
| 1998 | 2343 | 3933 | 993 | 137 | 76 | 41 | 96 | 64 | 58 | 49 | 19 | 19 | 23 | 14 | 125 |
| 1999 | 1505 | 1669 | 151 | 88 | 66 | 53 | 202 | 168 | 181 | 188 | 73 | 89 | 100 | 61 | 556 |
| 2000 | 973 | 1392 | 445 | 562 | 447 | 351 | 877 | 582 | 475 | 359 | 130 | 88 | 138 | 78 | 577 |
| 2001 | 2542 | 3057 | 410 | 197 | 130 | 93 | 311 | 237 | 219 | 170 | 66 | 43 | 66 | 36 | 286 |
| 2002 | 1006 | 1212 | 139 | 54 | 35 | 26 | 103 | 87 | 95 | 92 | 33 | 28 | 40 | 22 | 172 |
| 2003 | 110 | 162 | 50 | 23 | 12 | 7 | 16 | 11 | 9 | 8 | 3 | 3 | 4 | 2 | 25 |
| 2004 | 1904 | 2236 | 237 | 74 | 66 | 71 | 359 | 310 | 313 | 273 | 106 | 88 | 120 | 68 | 508 |
| 2005 | 504 | 670 | 145 | 74 | 36 | 21 | 99 | 85 | 86 | 76 | 30 | 25 | 34 | 19 | 191 |
| 2006 | 4457 | 7519 | 1636 | 62 | 27 | 14 | 93 | 89 | 106 | 114 | 42 | 46 | 56 | 33 | 268 |
| 2007 | 3310 | 4086 | 502 | 187 | 74 | 19 | 50 | 39 | 50 | 56 | 20 | 24 | 28 | 17 | 155 |
| 2008 | 781 | 1743 | 878 | 1031 | 419 | 134 | 290 | 185 | 174 | 186 | 60 | 69 | 89 | 53 | 594 |
| 2009 | 1947 | 4700 | 1483 | 173 | 75 | 31 | 113 | 100 | 138 | 174 | 56 | 59 | 81 | 46 | 363 |
| 2010 | 11016 | 13516 | 2029 | 689 | 234 | 34 | 167 | 157 | 182 | 283 | 134 | 313 | 253 | 178 | 2099 |
| 2011 | 2756 | 3657 | 590 | 260 | 117 | 46 | 134 | 106 | 121 | 158 | 67 | 127 | 114 | 77 | 791 |
| 2012 | 3922 | 4860 | 523 | 54 | 58 | 68 | 465 | 450 | 551 | 640 | 247 | 337 | 361 | 225 | 2268 |
| 2013 | 2332 | 3002 | 602 | 460 | 194 | 59 | 100 | 54 | 51 | 48 | 19 | 28 | 28 | 18 | 238 |
| 2014 | 232 | 646 | 978 | 1123 | 697 | 431 | 1071 | 739 | 675 | 751 | 325 | 610 | 539 | 367 | 3971 |
| 2015 | 2124 | 2505 | 322 | 542 | 409 | 300 | 726 | 482 | 406 | 388 | 162 | 260 | 245 | 163 | 1874 |


| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 3}$ |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2016 | 1211 | 1835 | 917 | 584 | 300 | 157 | 397 | 267 | 226 | 184 | 67 | 55 | 77 |  |  |
| 2017 | 974 | 1522 | 374 | 199 | 161 | 129 | 397 | 301 | 291 | 298 | 121 | 178 | 178 | 115 | 1130 |
| 2018 | 817 | 1004 | 135 | 145 | 163 | 171 | 810 | 719 | 786 | 945 | 398 | 690 | 641 | 424 | 4531 |
| 2019 | 1943 | 2202 | 156 | 143 | 137 | 120 | 669 | 645 | 749 | 1182 | 560 | 1325 | 1065 | 752 | 9058 |
| 2020 | 1062 | 1540 | 492 | 224 | 113 | 68 | 460 | 447 | 505 | 731 | 341 | 759 | 623 | 436 | 5435 |
| 2021 | 1327 | 1744 | 554 | 1855 | 1300 | 818 | 1784 | 1197 | 1245 | 1445 | 616 | 1116 | 1005 | 675 | 7033 |

SPPGFS

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 0 | 31 | 29 | 77 | 73 | 68 | 300 | 262 | 304 | 308 | 110 | 94 | 135 | 76 | 596 |
| 2002 | 0 | 0 | 2 | 34 | 58 | 71 | 330 | 283 | 294 | 270 | 103 | 92 | 122 | 70 | 584 |
| 2003 | 0 | 7 | 15 | 21 | 20 | 21 | 115 | 105 | 117 | 123 | 48 | 57 | 65 | 39 | 366 |
| 2004 | 1 | 3 | 5 | 13 | 25 | 34 | 177 | 158 | 169 | 175 | 69 | 85 | 94 | 58 | 515 |
| 2005 | 10 | 21 | 14 | 14 | 25 | 38 | 264 | 251 | 288 | 319 | 126 | 172 | 182 | 114 | 1218 |
| 2006 | 59 | 91 | 56 | 71 | 39 | 28 | 184 | 176 | 209 | 242 | 97 | 142 | 145 | 92 | 1021 |
| 2007 | 6 | 25 | 20 | 20 | 18 | 15 | 54 | 46 | 50 | 58 | 23 | 36 | 36 | 23 | 230 |
| 2008 | 8 | 23 | 23 | 40 | 47 | 48 | 193 | 163 | 176 | 188 | 73 | 95 | 104 | 64 | 636 |
| 2009 | 6 | 7 | 3 | 78 | 127 | 147 | 639 | 540 | 550 | 561 | 232 | 325 | 329 | 210 | 2203 |
| 2010 | 2 | 5 | 5 | 22 | 61 | 85 | 379 | 317 | 313 | 301 | 118 | 138 | 156 | 96 | 930 |
| 2011 | 0 | 9 | 19 | 19 | 35 | 52 | 320 | 290 | 310 | 301 | 118 | 125 | 149 | 89 | 861 |
| 2012 | 0 | 2 | 3 | 5 | 18 | 28 | 176 | 161 | 177 | 174 | 67 | 68 | 84 | 50 | 466 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2013 | 12 | 20 | 9 | 1 | 12 | 22 | 197 | 197 | 244 | 277 | 105 | 132 | 148 | 90 | 899 |
| 2014 | 2 | 33 | 49 | 11 | 45 | 89 | 992 | 1044 | 1403 | 1685 | 624 | 783 | 898 | 543 | 6669 |
| 2015 | 0 | 1 | 1 | 1 | 7 | 14 | 112 | 109 | 126 | 137 | 54 | 68 | 75 | 46 | 564 |
| 2016 | 1 | 5 | 10 | 5 | 4 | 6 | 61 | 62 | 78 | 91 | 35 | 48 | 51 | 32 | 360 |
| 2017 | 5 | 5 | 0 | 7 | 10 | 12 | 80 | 80 | 100 | 132 | 54 | 96 | 90 | 59 | 786 |
| 2018 | 0 | 0 | 0 | 1 | 19 | 41 | 501 | 534 | 718 | 906 | 349 | 516 | 536 | 337 | 4050 |
| 2019 | 0 | 1 | 3 | 3 | 8 | 15 | 167 | 172 | 215 | 260 | 104 | 157 | 158 | 101 | 1040 |
| 2020 | 0 | 2 | 2 | 3 | 7 | 11 | 113 | 115 | 136 | 177 | 77 | 146 | 129 | 87 | 1519 |
| 2021 | 0 | 15 | 32 | 32 | 20 | 14 | 104 | 109 | 154 | 219 | 86 | 149 | 144 | 94 | 1290 |

WCSGFS

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 12 | 61 | 90 | 197 | 233 | 248 | 736 | 509 | 363 | 224 | 85 | 38 | 74 | 41 | 261 |
| 1991 | 69 | 184 | 275 | 631 | 405 | 256 | 482 | 257 | 153 | 72 | 25 | 8 | 19 | 12 | 63 |
| 1992 | 6 | 30 | 133 | 733 | 849 | 840 | 2097 | 1321 | 823 | 409 | 155 | 41 | 112 | 63 | 301 |
| 1993 | 54 | 279 | 846 | 1723 | 1227 | 981 | 2777 | 1908 | 1446 | 1017 | 359 | 177 | 351 | 191 | 1165 |
| 1994 | 8 | 38 | 71 | 222 | 157 | 112 | 292 | 202 | 179 | 143 | 54 | 43 | 60 | 35 | 250 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 20 | 71 | 109 | 328 | 387 | 385 | 1141 | 811 | 665 | 480 | 184 | 116 | 183 | 102 | 718 |
| 1996 | 24 | 59 | 51 | 53 | 58 | 67 | 398 | 375 | 458 | 490 | 174 | 160 | 222 | 126 | 953 |
| 1997 | 8 | 76 | 107 | 81 | 76 | 71 | 233 | 174 | 154 | 119 | 46 | 31 | 47 | 26 | 197 |
| 1998 | 4 | 10 | 10 | 26 | 25 | 22 | 68 | 52 | 52 | 50 | 19 | 20 | 24 | 15 | 121 |
| 1999 | 3 | 71 | 173 | 244 | 182 | 134 | 315 | 199 | 150 | 100 | 38 | 24 | 37 | 21 | 141 |
| 2000 | 2 | 18 | 53 | 151 | 122 | 93 | 205 | 125 | 90 | 56 | 22 | 14 | 21 | 12 | 92 |
| 2001 | 0 | 5 | 14 | 35 | 33 | 30 | 122 | 103 | 112 | 118 | 45 | 55 | 62 | 38 | 397 |
| 2002 | 4 | 6 | 23 | 347 | 634 | 778 | 3010 | 2402 | 2269 | 1942 | 725 | 559 | 813 | 459 | 3480 |
| 2003 | 2 | 39 | 46 | 196 | 311 | 380 | 1730 | 1482 | 1545 | 1585 | 619 | 774 | 853 | 528 | 4647 |
| 2004 | 3 | 19 | 52 | 367 | 802 | 1054 | 4442 | 3641 | 3470 | 3148 | 1237 | 1315 | 1553 | 939 | 8289 |
| 2005 | 19 | 39 | 32 | 63 | 97 | 118 | 547 | 472 | 504 | 506 | 191 | 207 | 250 | 149 | 1307 |
| 2006 | 4 | 15 | 67 | 266 | 208 | 177 | 781 | 680 | 760 | 834 | 326 | 442 | 470 | 294 | 2900 |
| 2007 | 7 | 90 | 141 | 415 | 626 | 727 | 2893 | 2356 | 2285 | 2205 | 881 | 1104 | 1195 | 746 | 7600 |
| 2008 | 18 | 110 | 248 | 798 | 948 | 1026 | 5180 | 4696 | 5396 | 6246 | 2479 | 3677 | 3739 | 2381 | 26466 |
| 2009 | 2 | 27 | 524 | 2249 | 1182 | 537 | 771 | 336 | 263 | 187 | 68 | 70 | 81 | 51 | 531 |
| 2010 | 0 | 0 | 4 | 191 | 315 | 347 | 1030 | 738 | 612 | 492 | 192 | 191 | 231 | 140 | 1236 |

Table 3.6.3.1. Boarfish in ICES Subareas 27.6, 7, 8. Key parameter estimates from the exploratory Schaeffer state space surplus production model. Posterior parameter distributions are provided in Figure 3.6.3.5

| Parameter | Mean | SD | 2.5 | 25 | 50 | 75 | 97.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| r | 0.34 | 0.17 | 0.06 | 0.22 | 0.33 | 0.45 | 0.70 |
| K | 685102 | 511831 | 310500 | 443300 | 550600 | 730600 | 2116000 |
| $\mathrm{F}_{\text {MSY }}$ | 0.17 | 0.08 | 0.03 | 0.11 | 0.17 | 0.23 | 0.35 |
| $\mathrm{B}_{\text {MSY }}$ | 171276 | 127958 | 77625 | 110825 | 137650 | 182650 | 529000 |
| TSB | 628184 | 293425 | 306800 | 450800 | 565050 | 723900 | 1314975 |

Table 3.6.5.1. Boarfish in ICES Subareas 27.6, 7, 8. Estimates of total stock biomass and $F$

| Year | TSB.2.5 | TSB. 50 | TSB. 97.5 | F2.5 | F. 50 | F.97.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 96640 | 184200 | 440800 |  |  |  |
| 1992 | 159300 | 286400 | 666697 |  |  |  |
| 1993 | 194200 | 351200 | 815395 |  |  |  |
| 1994 | 229100 | 416300 | 977795 |  |  |  |
| 1995 | 198000 | 358800 | 833400 |  |  |  |
| 1996 | 199000 | 360000 | 835500 |  |  |  |
| 1997 | 168900 | 300800 | 694097 |  |  |  |
| 1998 | 222900 | 394750 | 910187 |  |  |  |
| 1999 | 168000 | 297300 | 686197 |  |  |  |
| 2000 | 145700 | 258600 | 598497 |  |  |  |
| 2001 | 162400 | 281900 | 642600 |  |  |  |
| 2002 | 140400 | 244100 | 559400 |  |  |  |
| 2003 | 128700 | 222500 | 506200 | 0.02 | 0.05 | 0.09 |
| 2004 | 179300 | 310100 | 706997 | 0.01 | 0.02 | 0.03 |
| 2005 | 175400 | 303900 | 690000 | 0.01 | 0.02 | 0.03 |
| 2006 | 220700 | 377500 | 855997 | 0.01 | 0.02 | 0.03 |
| 2007 | 198000 | 340400 | 769000 | 0.03 | 0.06 | 0.11 |
| 2008 | 242900 | 412500 | 935397 | 0.04 | 0.08 | 0.14 |
| 2009 | 248400 | 418500 | 932997 | 0.10 | 0.22 | 0.36 |
| 2010 | 367200 | 617600 | 1388000 | 0.10 | 0.23 | 0.39 |
| 2011 | 323000 | 546700 | 1226000 | 0.03 | 0.07 | 0.11 |
| 2012 | 464500 | 761200 | 1690000 | 0.05 | 0.11 | 0.19 |
| 2013 | 313300 | 523400 | 1178000 | 0.06 | 0.14 | 0.24 |
| 2014 | 147100 | 245800 | 550997 | 0.08 | 0.18 | 0.31 |
| 2015 | 176500 | 296500 | 668197 | 0.03 | 0.06 | 0.10 |
| 2016 | 130200 | 220150 | 497697 | 0.04 | 0.09 | 0.15 |
| 2017 | 230100 | 389450 | 880500 | 0.02 | 0.04 | 0.08 |
| 2018 | 246200 | 414400 | 933292 | 0.01 | 0.03 | 0.05 |


| Year | TSB.2.5 | TSB.50 | TSB.97.5 | F2.5 | F.50 | F.97.5 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2019 | 207800 | 350600 | 790090 | 0.01 | 0.03 | 0.05 |
| 2020 | 242300 | 414300 | 938790 | 0.02 | 0.04 | 0.06 |
| 2021 | 323800 | 547300 | 1231000 | 0.01 | 0.03 | 0.05 |
| 2022 | 306800 | 565050 | 1314975 |  |  |  |

### 3.17 Figures



Figure 3.1. Boarfish in ICES Subareas 4, 6, 7, 8 and 9. Distribution of boarfish in the NE Atlantic area based on presence and absence in IBTS surveys (all years).


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Figure 3.3.1.2. Boarfish in ICES Subareas 27.6, 7, 8. Boarfish acoustic survey biomass estimate by stratum, 2022.


Figure 3.3.1.3. Boarfish in ICES Subareas 27.6, 7, 8. Boarfish acoustic survey time series of acoustic estimates of abundance at age, 2011-2022.


Figure 3.3.2.1. Boarfish in ICES Subareas 27.6, 7, 8. The haul positions of bottom trawl surveys analysed as an index for boarfish abundance.


Figure 3.3.2.2. Boarfish in ICES Subareas 27.6, 7, 8. Distribution of boarfish in the NE Atlantic from the 6 IBTS surveys.


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Figure 3.6.1.5. Boarfish in ICES Subareas 27.6, 7, 8. Pair-wise correlation between the annual mean survey indices.


Figure 3.6.1.6. Boarfish in ICES Subareas 27.6, 7, 8 . Weighted correlation between the annual mean survey indices. Correlations are weighted by the sum of the pair-wise variances.


Figure 3.6.3.1. Boarfish in ICES Subareas 27.6, 7, 8. MCMC chain autocorrelation for final run.


Figure 3.6.3.2. Boarfish in ICES Subareas 27.6, 7, 8. Rhat values lower than 1.01 indicating convergence.


Figure 3.6.3.3. Boarfish in ICES Subareas 27.6, 7, 8. Parameters for final run converged with good mixing of the chains.


Figure 3.6.3.4. Boarfish in ICES Subareas 27.6, 7, 8. Residuals around the model fit for the final assessment run.


Figure 3.6.3.5. Boarfish in ICES Subareas 27.6, 7, 8. Prior (red) and posterior (black) distributions of the parameters of the biomass dynamic model.


Figure 3.6.3.6. Boarfish in ICES Subareas 27.6, 7, 8. Trajectories of observed and expected indices for the final assessment run. The stock size over time and a harvest ratio (total catch divided by estimated biomass) are also shown.


Figure 3.6.3.7. Boarfish in ICES Subareas 27.6, 7, 8. Retrospective plot of total stock biomass (above) and fishing mortality (below) from the surplus production model in 2013-2021.


Figure 3.6.6.1. Boarfish in ICES Subareas 27.6, 7, 8. Ratios 'B / MSYBtrigger' and 'F / FMSY' through time and corresponding Kobe plot. Confidence intervals ( 50 and $95 \%$ ) are given for the first two panels, the third displays median estimates only with the small dark blue point representing the first point of the time series and the large light blue point the last.


Figure 3.9.1.1. Boarfish in ICES Subareas 27.6, 7, 8. Results of exploratory yield per recruit analysis. Beverton and Holt model applied to various fits of the VBGF and for comparison with the VBGF parameters provided by White et al. 2011.


Figure 3.9.1.2. Boarfish in ICES Subareas 27.6, 7, 8. Sensitivity of estimation of F0.1.


Figure 3.12.1. Boarfish in ICES Subareas 27.6, 7, 8. Boarfish samples included in the genetic stock identification study are indicated in green. Population clusters identified by the STRUCTURE analyses are indicated by colour coded circles.

## 4 Herring (Clupea harengus) in subareas 1, 2, and 5, and in divisions 4.a and 14.a, (Northeast Atlantic) (Norwegian Spring Spawning)

### 4.1 ICES advice in 2021

ICES advised that when the long-term management strategy agreed by the European Union, the Faroe Islands, Iceland, Norway, and the Russian Federation is applied, catches in 2022 should be no more than 598588 tonnes. The advice for 2022 was $8 \%$ lower than that for 2021.

### 4.2 The fishery in 2021

### 4.2.1 Description and development of the fisheries

The distribution of the 2021 Norwegian spring-spawning herring (NSSH) fishery for all countries by ICES rectangles is shown in Figure 4.2.1.1. The catches by ICES statistical rectangle and quarter are seen in Figure 4.2.1.2. The 2021 herring fishing pattern was similar to recent years. The fishery began in January on the Norwegian shelf and focused on overwintering, pre-spawning, spawning and post-spawning fish (Figure 4.2.1.2, quarter 1). In the second quarter, the fishery was insignificant (Figure 4.2.1.2, quarter 2). In summer, the fishery moved into mainly Icelandic and International waters and in early autumn commenced in the overwintering area off Lofoten (Figure 4.2.1.2, quarter 3). In autumn and winter, the fishery continued in Icelandic and Faroese waters but also in the overwintering area in the fjords and oceanic areas off Lofoten (Figure 4.2.1.2, quarter 4). $60 \%$ of the catches were taken in the fourth quarter. Catches of Norwegian spring-spawning herring inside the NEAFC regulatory area was estimated by the working group to be 20347 tonnes in 2021, which represents $2 \%$ of the total catch. Note though that this does not include catches from the Russian Federation.

Since spatial and seasonal data were not available from the Russian fleet, Russian landings were not included in the description of the fisheries in 2021.

### 4.3 Stock Description and management units

### 4.3.1 Stock description

A description of the stock is given in the Stock Annex.

### 4.3.2 Changes in migration

Generally, it is not clear what drives the variability in migration of the stock, but the biomass and production of zooplankton are likely factors, as well as feeding competition with other pelagic fish species (e.g. mackerel and to a lesser extent blue whiting) and oceanographic conditions (e.g. limitations due to cold areas). Besides environmental factors, the age distribution in the stock will also influence the migration. Changes in the migration pattern of NSSH, as well as that of other herring stocks, are often linked to large year classes entering the stock initiating a different migration pattern, which subsequent year classes will follow. The large 2016 year-class has now
entered the adult stock. The distribution of the 2016 year-class in the feeding area in 2022 as observed in the ecosystem survey in May appeared to be distributed throughout the survey area. In 2017/2018 there was a shift in wintering areas. While wintering has been observed in fjords west of Tromsø (Norway) for several years, the 2013 year-class wintered in fjords farther north (Kvænangen) since 2017/2018 while the older fish seemed to have had an oceanic wintering area. A similar pattern was observed during the winter 2021/2022. The old fish wintered in the Norwegian Sea while part of the 2016 year-class wintered in Kvænangen. From Norwegian catches during winter, it was, however observed that a large fraction of the 2016 year-class wintered in the ocean further north (north of $70^{\circ} \mathrm{N}$ ). The oldest and largest fish move farthest south and west during feeding, and the older year classes were in May-July 2022 concentrated in the south-western areas during the feeding season.

### 4.4 Input data

### 4.4.1 Catch data

Catches in tonnes by ICES division, ICES rectangle and quarter in 2021 were available from Denmark, Faroe Islands, Germany, Greenland, Iceland, Ireland, The Netherlands, Norway, Poland and Sweden. This year the only information available from Russia was total catch by ICES division from the ICES preliminary catches data base. The total working group catch in 2021 was 851 813 tonnes (Table 4.4.1.1) compared to the ICES-recommended catch of a maximum of 651033 tonnes. The majority of the catches (around $85 \%$ ) were taken in division 2.a as in previous years. Samples were not provided by Russia, Greenland, Ireland, Poland or Sweden. Sampled catches accounted for $88 \%$ of the total catches, which is somewhat lower compared to previous years. The sampling levels of catches in 2021 in total, by country and by ICES division are shown in Tables 4.4.1.2, 4.4.1.3 and 4.4.1.4. Catch by nation, ICES division and quarter are shown in Table 4.4.1.5. Regarding the Russian catch, some assumptions were made in order to make it possible to handle these data using the existing method: see next paragraph. The software SALLOC (ICES, 1998) was used to calculate total catches in numbers-at-age and mean weight at age representing the total catch. Samples allocated (termed fill-in in SALLOC) to cells (nation, ICES division and quarter) without sampling information are shown in Table 4.4.1.5. Note that the cells with Russian catches were assumed unsampled, so sample information from other countries in the same cells were allocated.

### 4.4.1.1 Missing catch data in 2021

No Russian catch data or samples by ICES sub-division and quarter were delivered for 2021. The only information available on Russian catches in 2021 is from the ICES preliminary catches database for the entire year: 92840 tonnes in sub-division 2 .a and 1 ton in $5 . b$, which corresponds to about $11 \%$ of total catches. Some assumptions regarding the spatial and temporal distribution of the Russian catches had to be made in order to make it possible to estimate numbers and weights at age for the total international catch, and it was decided to base these assumptions on data from the period 2018-2020. Figure 4.4.1.1.1 shows the proportion by quarter of the Russian catch within each of the years 2018-2020, and the proportions are quite constant between years with most of the catches taken in the last two quarters. Table 4.4.1.1.1 shows the proportion of Russian catches by ICES sub-division for the years 2018-2020 and in practice all the catch was taken in area 2.a. Figures 4.4.1.1.2 and 4.4.1.1.3 show the Russian catch by ICES rectangle compared with the corresponding total international catch; the Russian fishery has been conducted in the same areas as the other nation's fishery in 2018-2020. Based on these results it was decided to assume that the Russian catch in 2021 (taken from the ICES preliminary catch database) was taken in area 2.a and that the distribution by quarter corresponded to the average proportions in 20182020. The Russian catches in the different quarter-area cells were treated as unsampled and
sample information from other nations was allocated to these according to the standard SALLOC procedure. Two additional figures are shown here that are relevant for the assumptions in the forecast: Figure 4.4.1.1.4 shows the Russian proportions of the total international catch per year in the period 2001-2020 and Figure 4.4.1.1.5 shows the Russian landings as a function of ICES advice for the period 2001-2020. The Russian proportions have been quite constant in recent years and there is a strong linear relationship between ICES advice and Russian landings.

### 4.4.2 Discards

In 2008, the Working Group noted that in this fishery an unaccounted mortality caused by fishing operations and underreporting probably exists (ICES, 2008). It has not been possible to assess the magnitude of these extra removals from the stock, and considering the large catches taken after the recovery of the stock, the relative importance of such additional mortality is probably low. Therefore, no extra mortality to account for these factors has been added since 1994. In previous years, when the stock and the quotas were much smaller, an estimated amount of fish was added to the catches.

The Working Group has not had access to comprehensive data to estimate discards of herring. Although discarding may occur on this stock, it is considered to be low and a minor problem for the assessment. This is confirmed by estimates from sampling programmes carried out by some EU countries in the Data Collection Framework. Estimates of discarding in 2008 and 2009 of about $2 \%$ in weight were provided for the trawl fishery carried out by the Netherlands. In 2010 and 2012, this métier was sampled by Germany. No discarding of herring was observed ( $0 \%$ ) in either of the two years. An investigation on fisheries induced mortality carried out by IMR with EU partners on fisheries induced and unreported mortality in mackerel and herring fisheries in the North Sea concluded with an estimated level of discarding at around 3\%.

In order to provide information on unaccounted mortality caused by fishing operations in the Norwegian fishery, Ipsos Public Affairs, in cooperation with IMR and the fishing industry, conducted a survey in January/February 2016. The survey was done by phoning skippers and interviewing them. A total of 146 herring skippers participated in the survey, 31 skippers representing the bigger vessel group and 115 skippers representing the smaller vessel group. The data provided an indication that there have been periods of increased occurrence of net bursting. This was seen especially in the period 2007-2010. There was, however, no trend in the size of catches where bursting has occurred.

When it comes to slipping, the data showed a steady increase in the percentage that has slipped herring from 2004-2012, and then a significant decline in recent years. The variations in the proportion that have slipped herring were largely driven by the skippers on smaller coastal purseseiners. Average size of purse-seine hauls slipped seems to be relatively steady over the period. However, the average size of net hauls slipped was lowest in the recent period.

### 4.4.3 Age composition of the catch

The estimated catch-at-age in numbers by year are shown in Table 4.4.3.1. The numbers are calculated using the SALLOC software. In 2021, catches (in numbers) were dominated by the 2016 year-class which comprised around $50 \%$ of the catch. Catch curves were made on the basis of the international catch-at-age (Figure 4.4.3.1). For comparison, lines corresponding to $\mathrm{Z}=0.3$ are drawn in the background. The big year classes, in the periods of relatively constant effort, show a consistent decline in catch number by cohort, indicating a reasonably good quality of the catch-at-age data. Catch curves for year classes 2005 onwards show a flatter curve than for previous year classes indicating a lower F or a changed exploitation pattern.

### 4.4.4 Weight-at-age in catch and in the stock

The weight-at-age in the catches in 2021 was computed from the sampled catches using SALLOC. Trends in weight-at-age in the catch are presented in Figure 4.4.4.1 and Table 4.4.4.1. The mean weights at age for most of the age groups have generally been increasing in 2010-2013 but levelled off around 2014. In the most recent years the weight-at-age seems to have decreased slightly for most ages - earlier for the younger ages than for the older. The decrease from 2020 to 2021 was generally larger than the preceding years. A similar pattern is observed in weight-at-age in the stock which is presented in Figure 4.4.4.2 and Table 4.4.4.2. The mean weight-at-age in the stock was based on the survey in the wintering area until 2008. Since then the mean weight-atage in the stock was derived from samples taken in the fishery in the same area and at the same time as the wintering surveys were conducted in.

### 4.4.5 Maturity-at-age

In 2010 the method for estimating maturity-at-age in the stock assessment of NSSH was changed based on work done by the "workshop on estimation of maturity ogive in Norwegian springspawning herring" (WKHERMAT; ICES, 2010a). The method which was adopted by WGWIDE in 2010 (ICES, 2010b) is based on work by Engelhard et al. (2003) and Engelhard and Heino (2004). They developed a method to back-calculate age-at-maturity for individual herring based on scale measurements, and used this to construct maturity ogives for the year classes 1930-1992.

The NSSH has irregular recruitment pattern with a few large year classes dominating in the stock when it is on a high level. Most of the year classes are, however, relatively small and referred to as "normal" year classes. The back-calculation dataset indicates that maturation of the large year classes is slower than for "normal" year classes.

WKHERMAT and WGWIDE considered the dataset derived by back calculation as a suitable candidate for use in the assessment because it is conceived in a consistent way over the whole period and can meet standards required in a quality-controlled process. However, the back-calculation estimates cannot be used for the most recent years since all year classes have to be fully matured before the calculation can be made. Therefore, assumptions have to be made for the recent year classes. For recent year classes, WGWIDE (ICES, 2010b) decided to use average backcalculated maturity for "normal" and "big"year classes thereby reducing maturity-at-age for ages 4,5 and 6 when strong year classes enter the spawning stock. The default maturity ogives used for "normal" and "big" year-classes are given in the text table below.

| age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| normal year class | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| strong | 0 | 0 | 0 | 0 | 0.1 | 0.6 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

year class
Assumed values should be replaced by back-calculated values in the annual assessments for each year where updated values are available. In 2022 the year 2017 was updated with back-calculated values used in the present assessment. Assumed and updated values are shown in figure 4.4.5.1. The 2016 year-class was considered a strong year-class by the working group based on the assessment where several survey indices of this year-class are included, and maturity at age 6 was set to 0.9 for this year-class in the 2022 assessment according to the table above. The maturity ogives used in the present assessment are presented in Table 4.4.5.1.

### 4.4.6 Natural mortality

In this year's assessment, the natural mortality $\mathrm{M}=0.15$ was used for ages 3 and older and $\mathrm{M}=0.9$ was used for ages $0-2$. These levels of natural mortality are in accordance to previous years and their justification is provided in the stock annex. Information about deviations from these levels in the time-series, e.g. due to diseases, are also provided in the stock annex.

### 4.4.7 Survey data

The surveys available for the assessment are described in the stock annex. Only two of the available surveys are used in the final assessment and will therefore be dealt with in this section:

1) The International Ecosystem Survey in the Nordic Seas (IESNS) in May. This survey covers the entire stock during its migration on the feeding grounds, the adults in the Norwegian Sea and adjacent waters ("Fleet 5") and the juveniles in the Barents Sea ("Fleet 4")
2 ) The Norwegian acoustic survey on the spawning grounds in February ("Fleet 1")
The cruise reports from the IESNS (WD05) and spawning survey (WD04) in 2022 are available as working documents to this report. The spawning survey and IESNS in the Norwegian Sea were both carried out successfully in 2022, however, the IESNS in the Barents Sea was not carried out by Russia this year.

The abundance estimates from "Fleet 1" are shown in Table 4.4.7.1 and Figure 4.4.7.2; from "Fleet 4" in Table 4.4.7.2 and Figure 4.4.7.1 and "Fleet 5" in Table 4.4.7.3 and Figure 4.4.7.1. In 2020 it was decided to use the bootstrap mean values as point estimates of abundance instead of the baseline estimates. This applies to the years were the software Stox is used to estimate abundance. Variance estimates from the bootstrap runs were already being used in the assessment, thus it is more logical to also use point estimates from the bootstrap. A comparison using point estimates for both bootstrap and baseline was made, and the effect on the assessment was negligible.

Catch curves were made on the basis of the abundance estimates from the surveys "Fleet 1 " (Figure 4.4.7.3) and "Fleet5" (Figure 4.4.7.4). The same arguments are valid for the interpretation of the catch curves from the surveys as from the catches. In 2010, the numbers of all age groups decreased suddenly in "Fleet 5" and this is seen as a drop in the catch curves that year. This drop has continued for some of the year classes and the year classes 1998 and 1999 are disappearing faster from the stock than expected. This observed fast reduction in these age classes may also be influenced by the changes in "Fleet 5" catchability, with seemingly higher catchability in years 2006-2009. Like the catch curves from commercial landings, the corresponding curves from "Fleet 5" are also quite flat for year classes 2005 onwards. As "Fleet 1 " was not conducted in the years 2009-2014, there is a gap in the catch curves, making it difficult to interpret them.

### 4.4.8 Sampling error in catches and surveys

Sampling errors for Norwegian catch-at-age for the years 2010-2021 is estimated using ECA (Salthaug and Aanes 2015, Hirst et al. 2012). Using the Taylor function (Aanes 2016a) to model the sampling variance of the catches yields a very good fit $\left(R_{a d j}^{2}=0.94\right)$ and using this function to impute missing sampling variances for catch-at-age yields relative standard errors shown in Table 4.4.8.1. It is assumed that the relative standard errors in the total catches are equal to the Norwegian catches (which comprise $\sim 60 \%$ of the total catches). Sampling errors for survey indices are estimated using StoX (http://www.imr.no/forskning/prosjekter/stox/nb-no) and Johnsen et al. (2019). For Fleet 1, estimates are available for the years 1988-1989, 1994-1996, 1998-2000,

2005-2008, and 2015-2021, for Fleet 4 estimates of sampling errors are available for 2009-2019 and 2021, and for Fleet 5 for 2008-2021. Missing values for sampling variances are imputed using the Taylor function which provides good fits ( $R_{\text {adj }}^{2}$ 's are $0.95,0.98$ and 0.96 respectively). The resultant relative standard errors are given in Tables 4.4.8.2-4.4.8.4. Due to the very good fits of the Taylor functions, estimates of relative standard where empirical estimates are available, are also replaced by the model predicted values to reduce potential effects of imprecise estimates of errors.

### 4.4.9 Information from the fishing industry

No information was made available to the working group.

### 4.5 Stock assessment

The first benchmark of the NSSH assessment took place in 2008 with the assessment tool TASACS selected as the standard assessment tool for the stock. A second benchmark took place in 2016 (WKPELA - ICES, 2016) where three assessment models were explored - TASACS, XSAM and one separable model. WKPELA accepted XSAM as the standard assessment tool for the NSSH.

### 4.5.1 XSAM final assessment 2022

The XSAM model is documented in Aanes 2016a and 2016b. XSAM includes the option to utilize the prediction of total catch in the assessment year (typically the sum of national quotas) along with the precision of the prediction. This approach was changed in 2017 when it was found that the model estimated a highly variable and significantly lower catch compared to the working group's prediction (sum of national quotas). In addition, this caused an abrupt change in the selection pattern from 2017 and onwards. The abrupt change in the selection pattern was not fully understood by the working group, but the effect was less pronounced if not using the catch prediction from the model for 2017. Therefore, it was decided to not utilize the prediction of total catches in 2017 when fitting the model to data (i.e. the assessment) and consequently in the shortterm forecast. The same approach is taken in the 2022 assessment, i.e. the catch prediction for 2022 is not included when fitting the model to data. The resulting estimated selection pattern is gradual (Figure 4.5.1.1) and in line with the current knowledge about the fishery. It is important to note that this has marginal effect on the assessment, but larger effects on the prediction and short-term forecast.

The 2022 XSAM assessment was performed with the same model options as in 2017. In summary, this means that the model was fit with time varying selectivity and effort according to $\operatorname{AR}(1)$ models in the model for fishing mortality; the recruitment was modelled as a process with constant mean and variance; the standard errors for all input data were predetermined using sample data (Tables 4.4.8.1-4.4.8.4), and a scaling constant common for all input data to allow additional variability in the input data that is not controlled by sampling is estimated. Additional details on the assessment settings are given in the Stock Annex.

The same input data over the same age ranges was used as in 2017. At the 2016 benchmark, data from 1988 and onwards was used from ages 3-12+ with input data catch-at-age, Fleet 1 and Fleet 5, At WGWIDE 2016, it was decided to start the model at age 2 to enable short-term predictions with reasonable levels of variability. To achieve this, age 2 from Fleet 4, and age 2 in catch-at-age was included in input data. Evaluation of diagnostics including lower ages than 2 and/or other fleets resulted in excluding lower ages than 2 and other fleets for the final assessment.

The parameter estimates from the 2022 assessment are shown in Table 4.5.1.1 and in Figure 4.5.1.10. For a precise definition of the parameters, refer to Aanes 2016a in ICES (2016). Note that the variance components $\sigma_{1}^{2}$ (variability in the separable model for F ) and $\sigma_{R}^{2}$ (variability in recruitment) are rather imprecise. The estimate of the scaling constant $h$ is larger than 1, indicating that the model adds additional variability on the observation errors than explained by the sampling errors alone.

The catchabilities for all the fleets are on average positively correlated indicating some uncertainty due to a common scaling of all surveys to the total abundances although the correlations in general are small (Figure 4.5.1.2). There is a slight negative correlation between $\sigma_{1}^{2}$ (variability in the separable model for $\mathrm{F}, \operatorname{logs} 2 \_1$ in figure) and $\sigma_{2}^{2}$ (variability in the AR process for time varying selectivity, logs2_2 in figure) indicating little contrast in data for separating variability in the separable model from variability due to changes in selection pattern. The slopes in the multivariate AR model for time-varying selectivity gradually changes from negative to positive, but is expected as it is imposed due to the sum to zero constraint for the selection (see Aanes 2016a for details).

The weights each datum is given in the model fit (inverse of the sampling variance) is proportional to the empirical weights derived from sampling variances (Tables 4.4.8.1-4.4.8.4) which shows that the strong year classes in general are given larger weight to the model than weaker year classes, and the ordering of the average weights (from high to low) is Catch-at-age, Fleet 5, Fleet 1 and Fleet 4 (Figure 4.5.1.3).

Two types of residuals are considered for this model. The first type is the model prediction (based on all data) vs. the data. In such time-series models, the residuals based on the prediction which uses all data points will be serially correlated although useful as they explain the unexplained part of the model (cf Harvey 1990 p 258). This means that patterns in residuals over time is to be expected and questions the use of e.g. qq-plots as an additional diagnostic tool to assess distributional assumptions. To obtain residuals which follow the assumptions about the data in the observation models (e.g. serially uncorrelated) single joint sample residuals are extracted (ICES, 2017). In short these are obtained by sampling predicted values from the conditional distribution of values given the observations. This sample corresponds to a sample from the joint distribution of latent variables and observations. A third approach could have been to extract the one step ahead observation residuals which are standard for diagnostics for regular state-space models (cf Harvey 1990). This is not done here.

The negative residuals tracing the 1983 year-class for catch-at-age represents low fishing mortalities examining the type 1 residuals (Figure 4.5.1.4). This effect is less pronounced considering the type 2 residuals. The type 2 residuals are qualitatively comparable with the type 1 residuals but generally display more mixed residuals as predicted by the theory. Otherwise the residuals for catch-at-age appears fairly mixed apart for some serial correlation for age 2 and 3 (which are very low), and some negative residuals for the plus group the most recent years. The residuals for Fleet 1 in year 1994, 1999, 2006 for young and old ages are all of the same signs and may appear as year effects. Also note that the residuals for Fleet 1 for ages 12+ from 2015 are all positive (Figure 4.5.1.4) which shows that the abundance indices from Fleet 1 displays a larger stock size over these ages and years compared to the assessment using all input data. Some serial correlation for residuals for ages 3 and 4 in Fleet 1 can also be detected, but is down weighted as these is found to be uncertain. Serial correlation in residuals for age 2 in Fleet 4 can also be detected indicating trends over time in mismatch between estimates and observations of abundance at age 2. Residuals for Fleet 5 appears adequate compared to previous years although some serial correlations can be detected also here.

The residuals for small values are bigger than residuals for the larger values since smaller values in general have higher variances than larger values (Tables 4.4.8.1-4.4.8.4) (Figure 4.5.1.5). The
qq-plots for the standardized residuals show that the distributional assumptions on the observation errors are adequate, except for the smallest and largest values of catch-at-age and indices from Fleet 1 and the smallest indices for Fleet 4. As qq-plots for residuals of type 1 may be questioned (see above) it is noted that qq-plots for residuals of type 2 is more relevant and generally shows a significantly better fit based on a visual inspection compared to using type 1.

The marginal likelihood and the components for each data source (see Aanes 2016b for details) are profiled over a range of the common scaling factor $h$ for all input data (Figure 4.5.1.6). It is apparent that the optimum of the marginal likelihood is clearly defined. The catch component is decreasing with decreasing values of $h$ indicating that the model puts more weight on the catch component than indicated by the comparison of sampling errors for all input data. This is in line with the findings in Aanes (2016a and 2016b) who showed that these types of models tend to put too much weight on the catch data if the weighting is not constrained. However, the likelihood component for the catch is overruled by the information in Fleets 1,4 and 5 such that the optimum for the marginal likelihood is clearly defined. The point estimates of SSB and F is insensitive to different values of $h$.

The retrospective runs for this model shows estimates within the estimated levels of precision (Figure 4.5.1.7), and has a reasonably low Mohn's rho value for SSB of -0.04 and -0.13 for F (Mohn, 1999; Brooks and Legault, 2016). Note that the retrospective patterns are remarkably stable.

Figure 4.5.1.8 illustrates the conflict in data and increased uncertainty in estimates for the most recent years. The spawning-stock biomass shown for each survey index is calculated using the stock weights at age and proportion mature at age, with the abundance indices are scaled to the absolute abundance by the estimated catchabilities. A fairly good temporal match between the model estimate of SSB and the survey SSBs is seen, except for the years 2015 for Fleet 1, which displays a significantly faster reduction in the stock compared to Fleet 5 which shows a flatter trend in the same years. Both Fleet 1 and Fleet 5 indicate an increase in SSB from 2007 to 2009, then a decrease in 2020 before an increase in 2021. It is worth noting that, although the point estimate of SSB based on Fleet 1 appears very much higher than Fleet 5 in 2015, the uncertainty in the estimates are very high, such that the respective estimates do not appear as significantly different. Since 2016 the conflict between Fleet 1 and Fleet 5 has become less.

The results of leave-one-out runs are presented in Figure 4.5.1.11 and can be used to assess the influence of individual data sources on the assessment. Removing Fleet 1 leads to a downward revision of SSB and an upward revision of F. The overall assessment uncertainty is similar to the base run which includes all data sources. Removing Fleet 5 results in an upward revision of SSB and a downward revision of F, with an increase in uncertainty. Removing Fleet 4 does not influence the SSB nor F.

The final 2022 assessment results are shown in Figure 4.5.1.9. The estimate of fishing mortality for 2019 to 2021 is rather high, as a response to the high catch in both years with a point estimates from $\sim 0.17$ to $\sim 0.19$. In 2018 the fishing mortality is estimated to be lower than in 2017 and 2019 ( $\mathrm{F}=0.129$ ). The spawning stock shows a declining trend since 2009 but an increase in 2021 and then a small decrease in 2022, and the $95 \%$ confidence interval of the stock level in 2022 ranges from $\sim 3.134$ to $\sim 4.6$ million tonnes with a point estimate of 3.867 which is above $B_{m p}=3.184$ million tonnes, such that the probability of the stock being above $\mathrm{B}_{\mathrm{lim}}=2.5$ million tonnes is high.

The final results of the assessment are also presented in Tables 4.5.1.2 (stock in numbers), 4.5.1.3 (fishing mortality) and Table 4.5.1.4 is the summary table of the assessment.

### 4.5.2 Exploratory assessments

### 4.5.2.1 TASACS

TASACS was run according to the benchmark in 2008 using the VPA population model in the TASACS toolbox with the same model options as the benchmark (see Stock Annex). The information used in the TASACS run is catch data and survey data from eight surveys. The analysis was restricted to the years 1988-2022. The model was run with catch data from 1988 to 2021, and projected forwards through 2022 assuming Fs in 2022 equal to those in 2021, to include survey data from 2022. The larval survey (SSB fleet) was discontinued in 2017 and no new information is therefore available from this survey. Additionally, no new index was provided for fleet 7 since 2019 (0-group from the autumn survey in the Barents Sea) since this index was not updated by the survey group. This time series ( 0 -group) is presently being re-calculated.

Residuals of the tuning series are shown in Figure 4.5.2.1.1. Particularly survey 8 (larval survey) seems to have a poor fit. This is seen as a block of positive residuals for this survey in later years. The residual plot for survey 5 (IESNS) also shows some pattern with consecutive series of negative and positive residuals indicating year-effects.

The results from TASACS are compared to those from XSAM and SAM in Figure 4.5.2.1.2. The time-series of SSB show similar trends for XSAM, SAM (configured as XSAM) and TASACS, although SSB in recent years are higher in TASACS due to an upward revision in the 2021 TASACS assessment. For most of the years, the estimates from TASACS are within the confidence limits estimated by XSAM except for the assessment year 2022 where the SSB from TASACS is slightly above the upper confidence limit. The SSB on 1 January 2022 is estimated by TASACS to be 4.63 million tonnes.

### 4.6 NSSH reference points

ICES last reviewed the reference points of Norwegian spring spawning herring in April 2018 during WKNSSHREF (ICES, 2018a). ICES concluded that Blim should remain unchanged at 2.5 million tonnes and MSY $\mathrm{B}_{\text {trigger }}=\mathrm{B}_{\mathrm{pa}}$ was estimated at 3.184 million tonnes. FmSy was estimated at the reference point workshop, but during the subsequent Management Strategy Evaluation WKNSSHMSE (ICES, 2018b) the fishing mortality reference points were revisited as issues were found with numerical instability and settings during the reference point workshop. FmsY was reestimated to be 0.157 .

### 4.6.1 PA reference points

The PA reference points for the stock were last estimated by WKNSSHREF and WKNSSHMSE in 2018. The WKNSSHREF group concluded that Blim should be kept at 2.5 million tonnes and $\mathrm{B}_{\mathrm{pa}}$ was estimated at 3.184 million tonnes. WKNSSHMSE estimated $\mathrm{F}_{\mathrm{pa}}=0.227$. However, following recent ICES guidelines $\mathrm{F}_{\mathrm{pa}}$ is now based on Fp05 which was estimated at 0.157 by WKNSSHMSE in 2018.

### 4.6.2 MSY reference points

The MSY reference points were evaluated by WKNSSHREF and WKNSSHMSE in 2018. In the ICES MSY framework $B_{p a}$ is proposed/adopted as the default trigger biomass $B_{\text {trigger }}$ and was estimated by WKNSSHREF at 3.184 million tonnes. FMSY was estimated by WKNSSHMSE at 0.157.

### 4.6.3 Management reference points

In the current management strategy, which was agreed upon in October 2018, the Coastal States have agreed a target reference point defined at $\mathrm{F}_{\text {target }}=0.14$ when the stock is above $\mathrm{B}_{\mathrm{pa}}$. If the SSB is below $\mathrm{B}_{\mathrm{pa}}$, a linear reduction in the fishing mortality rate will be applied from 0.14 at $\mathrm{B}_{\mathrm{pa}}$ to 0.05 at Blim.

### 4.7 State of the stock

The SSB on 1 January 2022 is estimated by XSAM to be 3.87 million tonnes which is above $B_{p a}$ ( 3.184 million $t$ ). The spawning stock has been declining since 2009 but increased in 2021 followed by a decrease again in 2022. The SSB time-series from the 2022 assessment is consistent with the SSB time-series from the 2021 assessment. In the last 20 years, several large year classes have been produced (1998, 1999, 2002, and 2004). The year classes 2005-2015 and 2017-2019 are estimated to be average or small, while the 2016 year-class is estimated to be above average in the 2022 assessment. Since there was no recruitment survey in 2022, the size of the 2020 yearclass at age 2 was defined as the stochastic median recruitment in the time series. Fishing mortality in 2021 is estimated to be 0.168 which is above the management strategy $\mathrm{F}(0.140)$ that was used to give advice for 2021. A new management strategy was implemented for the 2019 advisory year.

### 4.8 NSSH Catch predictions for 2023

### 4.8.1 Input data for the forecast

Forecasting was conducted using XSAM according to the method described in the Stock Annex and by Aanes (2016c). WGWIDE 2016 decided to use the point estimates from this forecast as basis for the advice. In short, the forecast is made by applying the point estimates of the stock status as input to set TAC, then based on the TAC a stochastic forecast was performed to determine levels of precision in the forecast. Table 4.8.1.1 lists the point estimates of the starting values for the forecast. The input stock numbers-at-age 2 and older were taken from the final assessment. The catch weight-at-age, used in the forecast, is the average of the observed catch weights over the last 3 years (2019-2021).

For the weight-at-age in the stock, the values for 2022 were obtained from the commercial fisheries in the wintering areas in January. For the years 2023 and 2024 the average of the last 3 years (2020-2022) was used.

Standard values for natural mortality were used. Maturity-at-age was based on the information presented in Section 4.4.5.
The exploitation pattern used in the forecast is taken from the predictions made by the model (see Aanes 2016c for details). The resultant mean annual exploitation pattern is shown in Figure 4.8.1.1 and displays a shift towards older fish in the recent years and further in the prediction. Prediction of recruitment at age 2 is obtained by the model with a mean that in practice represents the long term (1988-2021) estimated mean recruitment (back-transformed mean at log scale) and variance the corresponding recruitment variability over the period. Forecasted values of recruits are highly imprecise but have little influence on the short-term forecast of SSB as the herring starts to mature at age 4 . Note that the 2016 year-class is regarded as large; hence, the maturity is set to be lower than for smaller year-classes. This results in the contribution of the 2016 year-class to the SSB being delayed.

The average fishing mortality is defined as the average over the ages 5 to $12+$, weighted over the population numbers in the relevant year

$$
\bar{F}_{y}=\sum_{a=5}^{12} N_{a, y} F_{a, y} / \sum_{a=5}^{12} N_{a, y}
$$

where $F_{a, y}$ and $N_{a, y}$ are fishing mortalities and numbers by age and year. This procedure is in accordance with that used in previous years for this stock although the age range was shifted from 5-11 to 5-12+ from 2018.

There was no agreement between the fishing parties on the sharing of the TAC for 2021. Therefore, to obtain an estimate of the total catch to be used as input for the catch-constraint projections for 2022, the sum of the unilateral quotas was used. In total, the expected outtake from the stock in 2021 amounts to 827963 tonnes. F in 2022 is estimated by XSAM based on this catch.

### 4.8.2 Results of the forecast

The Management Options Table with the results of the forecast is presented in Table 4.8.2.1. Assuming a total catch 827963 tonnes is taken in 2022, it is expected that the SSB will decrease from 3.867 million tonnes on 1 January in 2022 to 3.532 million tonnes in 2023. The weighted F over ages $5-12+$ is 0.192 . The model predicts that the catch in 2023 to be dominated by three age groups, age 7 ( $46 \%$ ), age 10 ( $11 \%$ ), and age $12+(12 \%)$.

### 4.9 Comparison with previous assessment

A comparison between the assessments 2011-2022 is shown in Figure 4.9.1. In the years 20112015 the assessments were made with TASACS, whereas since 2016 XSAM has been applied, as accepted by WKPELA 2016. With the change of the assessment tool in 2016 the age of the recruitment changed from 0 to 2 and the age span in the reference $F$ changed from 5-14 to $5-11$. In WKNSSHREF (ICES, 2018a) this was further changed to 5-12+.

The table below shows the SSB (thousand tonnes) on 1 January in 2021 and weighted F in 2020 as estimated in 2021 and 2022.

|  | ICES 2021 | WG 2022 | \%difference |
| :--- | :--- | :--- | :--- |
| SSB (2021) | 3765 | 3930 | $4.4 \%$ |
| Weighted F (2020) | 0.188 | 0.19 | $1.1 \%$ |

### 4.10 Management plans and evaluations

The current management strategy for the Norwegian spring spawning herring fishery was agreed by the Coastal States in October 2018.

The implemented long-term management strategy of Norwegian spring spawning herring is consistent with the precautionary approach and the MSY approach (WKNSSHREF, ICES, 2018a; WKNSSHMSE, ICES, 2018b) and aims at ensuring harvest rates within safe biological limits. The management strategy in use contains the following elements:

As a priority, the long-term management strategy shall ensure with high probability that the size of the spawning stock is maintained above Blim.

In the case that the spawning biomass is forecast to be above or equal to $B_{\text {trigger }}\left(=B_{p a}\right)$ on 1 January of the year for which the TAC (i.e. the TAC agreed by Coastal States) is to be set, the TAC shall be fixed to a fishing mortality of $\mathrm{F}_{\mathrm{mgt}}=0.14$.

If $\mathrm{F}_{\text {mgt }}(0.14)$ would lead to a TAC, that deviates by more than $20 \%$ below or $25 \%$ above the TAC of the preceding year, the Parties shall fix a TAC that is respectively no more than $20 \%$ less or $25 \%$ more than the TAC of the preceding year. The TAC constraint shall not apply if the spawning biomass at 1 January in the year for which the TAC is to be set is less than $B_{\text {trigger }}$.

If SSB is forecast to be lower than $B_{\text {trigger }}$ but above $B_{\text {lim }}$ on the 1 January of the TAC-year, TAC is to be set using F, which decreases linearly from $F_{\text {mgt }}$ to $F=0.05$ over the biomass range from $B_{\text {trigger }}$ to Blim.

The Coastal States Parties may transfer 10\% of quotas between neighbouring years, except when SSB is less than Blim; those years the management plan does not allow fishing of next year's quota.
The Coastal States Parties, on the basis of ICES advice, shall review the long-term management strategy at intervals not exceeding five years. The first such review shall take place no later than 2023.

A brief history of management strategies is in the stock annex. In general, the stock has been managed in compliance with the management strategy. There has, however, been no agreement on sharing of the TAC since 2013, resulting in the total catch being higher than the advised catch.

### 4.11 Management considerations

Perception of the stock has not changed since last year's assessment (estimated SSB in 2021 is 5\% higher in this year's assessment).
Historically, the size of the stock has shown large variations and dependency on the irregular occurrence of very strong year classes. Between 1998 and 2004 the stock produced several strong year classes which lead to an increase in SSB until 2009. Since then, SSB has declined due to absence of strong year classes in 2005-2015. The 2016 year-class was however, estimated to be well above average in the 2021 assessment and resulted in an increase in SSB from 2020 to 2021. SSB, however, declined in 2022 and is predicted to be below $B_{m g t}$ in 2024 even if the management strategy ( $\mathrm{F}=0.14$ ) is applied in 2023.

Between 1999 and 2013, catches were regulated through an agreed management. However, since 2013, a lack of agreement by the Coastal States on their share in the TAC has led to unilaterally set quotas which together are higher than the TAC indicated by the management strategy resulting in steeper reduction in the SSB than otherwise.

A new management strategy was implemented for the advisory year 2019.

### 4.12 Ecosystem considerations

NSS herring juveniles and adults are an important part of the ecosystems in the Barents Sea, along the Norwegian coast, in the Norwegian Sea and in adjacent waters. This refers both to predation on zooplankton by herring and herring being a food resource to higher trophic levels (e.g. cod, saithe, seabirds, and marine mammals). The predation intensity of and on herring have seasonal, spatial and temporal variation as a consequence of variation in migration pattern, prey density, stock size, size of year classes and stock sizes of competing stocks for resources and predators. Recent features of some of these ecosystem factors of relevance for the stock are summarized below.

- Following a maximum in zooplankton biomass in May during the early 2000s the biomass declined with a minimum in 2006. From 2010, the trend turned to an increase and the last five years the zooplankton biomass has fluctuated around the long-term mean in the Lofoten and Norwegian Basins (IESNS survey report - ICES, 2022a), but is still low compared to the early years in East Iceland waters and the Jan Mayen front. Interestingly, all the areas, excluding east of Iceland and on few occasions Jan Mayen, show co-varying changes in zooplankton biomass.
- The Atlantic water mass in the Norwegian Sea was warmer and saltier over the period 2000-2016 than the long-term mean (WGINOR - ICES, 2022b). However, during the period, 2017-2020 the temperature remained relatively warm while the salinity had a marked decrease. Two different mechanisms can explain this, increased fraction of subpolar water (fresh and cold) and low heat loss to the atmosphere in the Norwegian Atlantic flow. The recent trend of colder and fresher Atlantic Inflow into the Norwegian Sea has ceased. The extent of Arctic water continues to increase (ICES, 2022b).
- The sea temperature in 2022 was generally below the long-term mean (1995-2021) in the Norwegian Sea, but the pattern was more fragmented below 50 m depth. The Arctic front in the southern Norwegian Sea was more southerly and easterly located in 2022 compared to the long-term mean.
- In general, the herring stock has had a more westerly feeding distribution (ICES, 2022a; IESSNS survey report - 2022c) in the recent years than what was previously observed. In May 2022, the herring in west was more northerly distributed than in recent years. The large 2016 year -lass was now widely distributed into the southwestern feeding area. The westerly distribution might be due to either better feeding opportunities there or a response to feeding competition with mackerel but the consequence is a less spatial overlap of herring and mackerel in Norwegian Sea and adjoining waters since around 2014 (ICES, 2022c).
- Where herring and mackerel overlap spatially they compete for food to some extent (Bachiller et al., 2016, 2018; Debes et al., 2012; Langøy et al., 2012; Óskarsson et al., 2016). There are studies showing mackerel being more effective feeder, which might indicate that the herring is forced to the south western and north eastern fringe of Norwegian Sea (ICES, 2021b). Alternatively, the higher zooplankton biomass in the southwest could also attract the herring into this location, since zooplankton biomass is much lower in the north east (ICES, 2022b).
- Results of stomach analyses of mackerel on the Norwegian coastal shelf (between about $66^{\circ} \mathrm{N}$ and $69^{\circ} \mathrm{N}$ ) suggest that mackerel fed opportunistically on herring larvae, and that predation pressure therefore largely depends on the degree of overlap in time and space (Skaret et al., 2015). Sampling in June 2017 and 2018, specifically studying mackerel predation on herring larvae, found significant numbers of herring larvae in mackerel stomachs in the area just south of Lofoten (Allan et al., 2021).
- The 2016 year-class of herring was the strongest since the 2004 year class in the Norwegian Sea as 4 year olds but as expected abundance is now beginning to visibly decrease (see the IESNS survey 2022 (Table 4.4.7.3).
- In the winter 2017/2018, the overwintering grounds shifted northward along the coast of Norway with older individuals occurring in oceanic areas. Such changes previously coincided with large year classes entering the spawning stock, however this recent change did not. Also, the onset of the overwintering period has been later in the year since the end of the 2000s.
Around spawning time of 2022 most of the spawning stock was found outside Lofoten and Vesterålen, further north and more concentrated than usual. The observed maturity indicated a later spawning compared to the previous year (WGWIDE WD04).


### 4.13 Changes in fishing patterns

The fishery for Norwegian spring spawning herring has previously (before 2013) been described as progressing clockwise in the Nordic Seas during the year. However, the last 5-8 years the annual progression of the fishery has changed into a pendular behaviour, starting in the winter along the Norwegian coast, moving gradually to the west towards Iceland in the summer, and then east again into the central Norwegian Sea in the last quarter of the year.

The fishery reached its lowest catches since the mid-nineties in 2015, after which the catches increased again, reaching a maximum in 2021 of 850000 tonnes (Table 4.4.1.1). It is mainly the fishery in the fourth quarter that has increased since 2015, with up to $2 / 3$ of the catches taken in this quarter. The fishery in quarter four in the last few years has partly been north of Lofoten and partly in the central Norwegian Sea, whereas before 2015 it used to be stretched out towards the coast of Norway and north towards the Bear Island.
The change in migration pattern since 2017/18, where the part of the stock (old fish) overwinters in the central Norwegian Sea, has caused the fishery in this area to be extended to later in the winter, and in 2021 there was fishery in the central Norwegian Sea in the first quarter as well as the fourth.

Annual fishing pattern 2011-2020 is shown in Section 1.8.

### 4.14 Recommendations

For some years there have been issues with age reading of herring. WGWIDE has recommended organising a scale/otolith exchange and workshop. This workshop is now scheduled for April 2023 with a preceding exchange in winter 2022/2023.

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### 4.16 Tables

Table 4.4.1.1 Total landings (ICES estimate) of Norwegian spring-spawning herring (tons) since 1972. Data provided by Working Group members.

| Year | Norway | USSR/ <br> Russia | Denmark | Faroes | Iceland | Ireland | Netherlands | Greenland | UK | Germany | France | Poland | Sweden | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 13161 | - | - | - | - | - | - | - | - | - | - | - | - | 13161 |
| 1973 | 7017 | - | - | - | - | - | - | - | - | - | - | - | - | 7017 |
| 1974 | 7619 | - | - | - | - | - | - | - | - | - | - | - | - | 7619 |
| 1975 | 13713 | - | - | - | - | - | - | - | - | - | - | - | - | 13713 |
| 1976 | 10436 | - | - | - | - | - | - | - | - | - | - | - | - | 10436 |
| 1977 | 22706 | - | - | - | - | - | - | - | - | - | - | - | - | 22706 |
| 1978 | 19824 | - | - | - | - | - | - | - | - | - | - | - | - | 19824 |
| 1979 | 12864 | - | - | - | - | - | - | - | - | - | - | - | - | 12864 |
| 1980 | 18577 | - | - | - | - | - | - | - | - | - | - | - | - | 18577 |
| 1981 | 13736 | - | - | - | - | - | - | - | - | - | - | - | - | 13736 |
| 1982 | 16655 | - | - | - | - | - | - | - | - | - | - | - | - | 16655 |
| 1983 | 23054 | - | - | - | - | - | - | - | - | - | - | - | - | 23054 |
| 1984 | 53532 | - | - | - | - | - | - | - | - | - | - | - | - | 53532 |
| 1985 | 167272 | 2600 | - | - | - | - | - | - | - | - | - | - | - | 169872 |
| 1986 | 199256 | 26000 | - | - | - | - | - | - | - | - | - | - | - | 225256 |


| Year | Norway | USSR/ <br> Russia | Denmark | Faroes | Iceland | Ireland | Netherlands | Greenland | UK | Germany | France | Poland | Sweden | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 108417 | 18889 | - | - | - | - | - | - | - | - | - | - | - | 127306 |
| 1988 | 115076 | 20225 | - | - | - | - | - | - | - | - | - | - | - | 135301 |
| 1989 | 88707 | 15123 | - | - | - | - | - | - | - | - | - | - | - | 103830 |
| 1990 | 74604 | 11807 | - | - | - | - | - | - | - | - | - | - | - | 86411 |
| 1991 | 73683 | 11000 | - | - | - | - | - | - | - | - | - | - | - | 84683 |
| 1992 | 91111 | 13337 | - | - | - | - | - | - | - | - | - | - | - | 104448 |
| 1993 | 199771 | 32645 | - | - | - | - | - | - | - | - | - | - | - | 232457 |
| 1994 | 380771 | 74400 | - | 2911 | 21146 | - | - | - | - | - | - | - | - | 479228 |
| 1995 | 529838 | 101987 | 30577 | 57084 | 174109 | - | 7969 | 2500 | 881 | 556 | - | - | - | 905501 |
| 1996 | 699161 | 119290 | 60681 | 52788 | 164957 | 19541 | 19664 | - | 46131 | 11978 | - | - | 22424 | 1220283 |
| 1997 | 860963 | 168900 | 44292 | 59987 | 220154 | 11179 | 8694 | - | 25149 | 6190 | 1500 | - | 19499 | 1426507 |
| 1998 | 743925 | 124049 | 35519 | 68136 | 197789 | 2437 | 12827 | - | 15971 | 7003 | 605 | - | 14863 | 1223131 |
| 1999 | 740640 | 157328 | 37010 | 55527 | 203381 | 2412 | 5871 | - | 19207 | - | - | - | 14057 | 1235433 |
| 2000 | 713500 | 163261 | 34968 | 68625 | 186035 | 8939 | - | - | 14096 | 3298 | - | - | 14749 | 1207201 |
| 2001 | 495036 | 109054 | 24038 | 34170 | 77693 | 6070 | 6439 | - | 12230 | 1588 | - | - | 9818 | 766136 |
| 2002 | 487233 | 113763 | 18998 | 32302 | 127197 | 1699 | 9392 | - | 3482 | 3017 | - | 1226 | 9486 | 807795 |
| 2003* | 477573 | 122846 | 14144 | 27943 | 117910 | 1400 | 8678 | - | 9214 | 3371 | - | - | 6431 | 789510 |
| 2004 | 477076 | 115876 | 23111 | 42771 | 102787 | 11 | 17369 | - | 1869 | 4810 | 400 | - | 7986 | 794066 |


| Year | Norway | USSR/ <br> Russia | Denmark | Faroes | Iceland | Ireland | Netherlands | Greenland | UK | Germany | France | Poland | Sweden | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 580804 | 132099 | 28368 | 65071 | 156467 | - | 21517 | - | - | 17676 | 0 | 561 | 680 | 1003243 |
| 2006 | 567237 | 120836 | 18449 | 63137 | 157474 | 4693 | 11625 | - | 12523 | 9958 | 80 | - | 2946 | 968958 |
| 2007 | 779089 | 162434 | 22911 | 64251 | 173621 | 6411 | 29764 | 4897 | 13244 | 6038 | 0 | 4333 | 0 | 1266993 |
| 2008 | 961603 | 193119 | 31128 | 74261 | 217602 | 7903 | 28155 | 3810 | 19737 | 8338 | 0 | 0 | 0 | 1545656 |
| 2009 | 1016675 | 210105 | 32320 | 85098 | 265479 | 10014 | 24021 | 3730 | 25477 | 14452 | 0 | 0 | 0 | 1687371 |
| 2010 | 871113 | 199472 | 26792 | 80281 | 205864 | 8061 | 26695 | 3453 | 24151 | 11133 | 0 | 0 | 0 | 1457015 |
| 2011 | 572641 | 144428 | 26740 | 53271 | 151074 | 5727 | 8348 | 3426 | 14045 | 13296 | 0 | 0 | 0 | 992997 |
| 2012 | 491005 | 118595 | 21754 | 36190 | 120956 | 4813 | 6237 | 1490 | 12310 | 11945 | 0 | 0 | 705 | 826000 |
| 2013 | 359458 | 78521 | 17160 | 105038 | 90729 | 3815 | 5626 | 11788 | 8342 | 4244 | 0 | 0 | 23 | 684743 |
| 2014 | 263253 | 60292 | 12513 | 38529 | 58828 | 706 | 9175 | 13108 | 4233 | 669 | 0 | 0 | 0 | 461306 |
| 2015 | 176321 | 45853 | 9105 | 33031 | 42625 | 1400 | 5255 | 12434 | 55 | 2660 | 0 | 0 | 0 | 328740 |
| 2016 | 197501 | 50455 | 10384 | 44727 | 50418 | 2048 | 3519 | 17508 | 4031 | 2582 | 0 | 0 | 0 | 383174 |
| 2017 | 389383 | 91118 | 19037 | 98170 | 90400 | 3495 | 6679 | 12569 | 4358 | 5201 | 0 | 1 | 1155 | 721566 |
| 2018 | 332028 | 64185 | 17052 | 82062 | 83393 | 2428 | 4290 | 2465 | 2582 | 1989 | 0 | 0 | 425 | 592899 |
| 2019 | 430507 | 84364 | 21207 | 113945 | 108045 | 2775 | 5111 | 3190 | 1801 | 4188 | 0 | 1327 | 705 | 777165 |
| 2020 | 409436 | 74936 | 16523 | 103029 | 98173 | 2704 | 5060 | 3546 | 143 | 2969 | 0 | 1352 | 3065 | 720937 |
| 2021** | 489632 | 92841 | 15854 | 114291 | 114299 | 1793 | 10939 | 6456 | 0 | 3365 | 0 | 1242 | 1101 | 851813 |

*In 2003 the Norwegian catches were raised of 39433 to account for changes in percentages of water content.
**The Russian catch for 2021 was taken from the ICES preliminary catches database

Table 4.4.1.1.1 Proportion (\%) of Russian catches by ICES sub-division for the years 2018-2020.

|  | $\mathbf{1}$ | 2.a | $\mathbf{5 . b}$ |
| :--- | :--- | :--- | :--- |
| 2018 | 0 | 100 | 0 |
| 2019 | 0.04 | 98.33 | 1.64 |
| 2020 | 0 | 99.93 | 0.07 |

Table 4.4.1.2 Norwegian spring-spawning herring. Sampling coverage by year.

| Year | TOTAL CATCH | \% catch covered by sampling programme | No. samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 1207201 | 86 | 389 | 55956 | 10901 |
| 2001 | 766136 | 86 | 442 | 70005 | 11234 |
| 2002 | 807795 | 88 | 184 | 39332 | 5405 |
| 2003 | 789510 | 71 | 380 | 34711 | 11352 |
| 2004 | 794066 | 79 | 503 | 48784 | 13169 |
| 2005 | 1003243 | 86 | 459 | 49273 | 14112 |
| 2006 | 968958 | 93 | 631 | 94574 | 9862 |
| 2007 | 1266993 | 94 | 476 | 56383 | 14661 |
| 2008 | 1545656 | 94 | 722 | 81609 | 31438 |
| 2009 | 1686928 | 94 | 663 | 65536 | 12265 |
| 2010 | 1457015 | 91 | 1258 | 124071 | 12377 |
| 2011 | 992.997 | 95 | 766 | 79360 | 10744 |
| 2012 | 825.999 | 93 | 649 | 59327 | 14768 |
| 2013 | 684.743 | 91 | 402 | 33169 | 11431 |
| 2014 | 461.306 | 89 | 229 | 18370 | 5813 |
| 2015 | 328.739 | 92 | 177 | 25156 | 5039 |
| 2016 | 383.174 | 91 | 203 | 39120 | 5892 |
| 2017 | 721566 | 95 | 335 | 31755 | 7241 |
| 2018 | 592899 | 97 | 253 | 22106 | 6047 |
| 2019 | 777165 | 97 | 361 | 29856 | 7421 |
| 2020 | 720937 | 98 | 232 | 34232 | 6742 |
| 2021 | 851813 | 88 | 207 | 18830 | 5975 |

Table 4.4.1.3 Norwegian spring-spawning herring. Sampling coverage by country in 2021.

| COUNTRY | OFFICIAL CATCH | \% catch covered by sampling programme | NO. SAMPLES | NO. MEASURED | NO. <br> AGED |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 15854 | 100 | 11 | 1129 | 292 |
| Faroe Islands | 114291 | 98 | 17 | 958 | 861 |
| Germany | 3365 | 100 | 32 | 10555 | 337 |
| Greenland | 6456 | 0 | 0 | 0 | 0 |
| Iceland | 114299 | 100 | 55 | 2446 | 1958 |
| Ireland | 1793 | 0 | 0 | 0 | 0 |
| The Netherlands | 10939 | 100 | 12 | 1514 | 299 |
| Norway | 489632 | 100 | 80 | 2228 | 2228 |
| Poland | 1242 | 0 | 0 | 0 | 0 |
| UK | 0 | 0 | 0 | 0 | 0 |
| Sweden | 1101 | 0 | 0 | 0 | 0 |
| Russia | 92841 | 0 | 0 | 0 | 0 |
| Total for Stock | 851814 | 88 | 207 | 18830 | 5975 |

Table 4.4.1.4 Norwegian spring-spawning herring. Sampling coverage by ICES Division in 2021.

| Area | Official Catch | No Sam- <br> ples | No Aged | No Meas- <br> ured | No Aged/ 1000 <br> tonnes | No Measured/ 1000 <br> tonnes |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2.1 | 725400 | 161 | 4038 | 16763 | 6 | 23 |
| $4 . a$ | 113 | 0 | 0 | 0 | 0 | 0 |
| $5 . a$ | 126279 | 46 | 1937 | 2067 | 15 | 16 |
| $5 . b$ | 21 | 0 | 0 | 0 | 0 | 0 |
| Total | 851813 | 207 | 5975 | 18830 | 7 | 22 |

Table 4.4.1.5 Norwegian spring-spawning herring. Catch data provided by working group members and samples allocated to unsampled catches in SALLOC.

| Line | Country | Quarter | Div. | Catch (T) | Samples allocated (line) |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | Norway | 1 | $2 a$ | 203631.8 |  |
| 2 | Norway | 2 | $2 a$ | 168.2 | 1 |
| 3 | Norway | 3 | $2 a$ | 22706.5 |  |
| 4 | Norway | 4 | $2 a$ | 263110 |  |


| Line | Country | Quarter | Div. | Catch ( T ) | Samples allocated (line) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | Norway | 3 | 4 a | 15.9 | 3 |
| 6 | Iceland | 2 | 2a | 516 |  |
| 7 | Iceland | 3 | 2 a | 6302 |  |
| 8 | Iceland | 3 | 5a | 57001 |  |
| 9 | Iceland | 4 | 5a | 50480 |  |
| 10 | Faroe Islands | 1 | 2 a | 10574 |  |
| 11 | Faroe Islands | 2 | 2a | 163 | 6 |
| 12 | Faroe Islands | 3 | 2a | 1800 | 3,7 |
| 13 | Faroe Islands | 4 | 2a | 82935 |  |
| 14 | Faroe Islands | 3 | 5a | 49 | 8 |
| 15 | Faroe Islands | 4 | 5a | 18749 |  |
| 16 | Faroe Islands | 3 | 5b | 21 | 3,7,8 |
| 17 | Russia | 1 | 2a | 236.3 | 1,10 |
| 18 | Russia | 2 | 2 a | 58.4 | 6 |
| 19 | Russia | 3 | 2a | 13868.4 | 3,7 |
| 20 | Russia | 4 | 2a | 78677.9 | 4,13,21,23,24 |
| 21 | Denmark | 4 | 2 a | 15854.5 |  |
| 22 | Germany | 3 | 2 a | 0.5 | 3,7 |
| 23 | Germany | 4 | 2 a | 3364.9 |  |
| 24 | Netherlands | 4 | 2 a | 10939.1 |  |
| 25 | Greenland | 3 | 2 a | 71.9 | 3,7 |
| 26 | Greenland | 4 | 2a | 6384 | 4,13,21,23,24 |
| 27 | Ireland | 4 | 2a | 1792.6 | 4,13,21,23,24 |
| 28 | Poland | 4 | 2a | 1144.3 | 4,13,21,23,24 |
| 29 | Poland | 4 | $4 a$ | 97.4 | 4,13,21,23,24 |
| 30 | Sweden | 4 | 2a | 1101 | 4,13,21,23,24 |

## Table 4.4.3.1. Norwegian spring spawning herring. Catch in numbers (thousands).

|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1950 | 5112600 | 2000000 | 600000 | 276200 | 184800 | 185500 | 547000 | 628600 | 79500 | 88600 | 109500 | 86900 | 194500 | 368300 | 66400 | 344300 |
| 1951 | 1635500 | 7607700 | 400000 | 6600 | 383800 | 172400 | 164400 | 515600 | 602000 | 77100 | 82700 | 103100 | 107600 | 253500 | 348000 | 352500 |
| 1952 | 13721600 | 9149700 | 1232900 | 39300 | 60500 | 602300 | 136300 | 204500 | 380200 | 377900 | 79200 | 85700 | 107700 | 106800 | 186500 | 564400 |
| 1953 | 5697200 | 5055000 | 581300 | 740100 | 46600 | 100900 | 355600 | 81900 | 110900 | 314100 | 394900 | 61700 | 91200 | 94100 | 98800 | 730400 |
| 1954 | 10675990 | 7071090 | 855400 | 266300 | 1435500 | 142900 | 236000 | 490300 | 128100 | 199800 | 440400 | 460700 | 88400 | 100600 | 133000 | 803200 |
| 1955 | 5175600 | 2871100 | 510100 | 93000 | 276400 | 2045100 | 114300 | 189600 | 274700 | 85300 | 193400 | 295600 | 203200 | 58700 | 84600 | 580600 |
| 1956 | 5363900 | 2023700 | 627100 | 116500 | 251600 | 314200 | 2555100 | 110000 | 203900 | 264200 | 130700 | 198300 | 272800 | 163300 | 63000 | 565100 |
| 1957 | 5001900 | 3290800 | 219500 | 23300 | 373300 | 153800 | 228500 | 1985300 | 72000 | 127300 | 182500 | 88400 | 121200 | 149300 | 131600 | 281400 |
| 1958 | 9666990 | 2798100 | 666400 | 17500 | 17900 | 110900 | 89300 | 194400 | 973500 | 70700 | 123000 | 200900 | 98700 | 77400 | 70900 | 255600 |
| 1959 | 17896280 | 198530 | 325500 | 15100 | 26800 | 25900 | 146600 | 114800 | 240700 | 1103800 | 88600 | 124300 | 198000 | 88500 | 77400 | 235900 |
| 1960 | 12884310 | 13580790 | 392500 | 121700 | 18200 | 28100 | 24400 | 96200 | 73300 | 203900 | 1163000 | 85200 | 129700 | 153500 | 56700 | 168900 |
| 1961 | 6207500 | 16075600 | 2884800 | 31200 | 8100 | 4100 | 15000 | 19400 | 61600 | 49200 | 136100 | 728100 | 49700 | 45000 | 63000 | 60100 |
| 1962 | 3693200 | 4081100 | 1041300 | 1843800 | 8000 | 3100 | 7200 | 20200 | 11900 | 59100 | 52600 | 117000 | 813500 | 44200 | 54700 | 152300 |
| 1963 | 4807000 | 2119200 | 2045300 | 760400 | 835800 | 5300 | 1800 | 3600 | 18300 | 9300 | 107700 | 92500 | 174100 | 923700 | 79600 | 185300 |
| 1964 | 3613000 | 2728300 | 220300 | 114600 | 399000 | 2045800 | 13700 | 1500 | 3000 | 24900 | 29300 | 95600 | 82400 | 153000 | 772800 | 336800 |
| 1965 | 2303000 | 3780900 | 2853600 | 89900 | 256200 | 571100 | 2199700 | 19500 | 14900 | 7400 | 19100 | 40000 | 100500 | 107800 | 138700 | 883100 |


|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1966 | 3926500 | 662800 | 1678000 | 2048700 | 26900 | 466600 | 1306000 | 2884500 | 37900 | 14300 | 17400 | 26200 | 11000 | 69100 | 72100 | 556700 |
| 1967 | 426800 | 9877100 | 70400 | 1392300 | 3254000 | 26600 | 421300 | 1132000 | 1720800 | 8900 | 5700 | 3500 | 8500 | 8900 | 17500 | 104400 |
| 1968 | 1783600 | 437000 | 388300 | 99100 | 1880500 | 1387400 | 14220 | 94000 | 134100 | 345100 | 2000 | 1100 | 830 | 2500 | 2600 | 17000 |
| 1969 | 561200 | 507100 | 141900 | 188200 | 800 | 8800 | 4700 | 700 | 11700 | 33600 | 36000 | 300 | 200 | 200 | 200 | 2400 |
| 1970 | 119300 | 529400 | 33200 | 6300 | 18600 | 600 | 3300 | 3300 | 1000 | 13400 | 26200 | 28100 | 300 | 100 | 200 | 2000 |
| 1971 | 30500 | 42900 | 85100 | 1820 | 1020 | 1240 | 360 | 1110 | 1130 | 360 | 4410 | 6910 | 5450 | 0 | 20 | 120 |
| 1972 | 347100 | 41000 | 20400 | 35376 | 3476 | 3583 | 2481 | 694 | 1486 | 198 | 0 | 494 | 593 | 593 | 0 | 0 |
| 1973 | 29300 | 3500 | 1700 | 2389 | 25200 | 651 | 1506 | 278 | 178 | 0 | 0 | 0 | 0 | 0 | 180 | 0 |
| 1974 | 65900 | 7800 | 3900 | 100 | 241 | 24505 | 257 | 196 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 30600 | 3600 | 1800 | 3268 | 132 | 910 | 30667 | 5 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | . 20100 | 2400 | 1200 | 23248 | 5436 | 0 | 0 | 13086 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 43000 | 6200 | 3100 | 22103 | 23595 | 336 | 0 | 419 | 10766 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 20100 | 2400 | 1200 | 3019 | 12164 | 20315 | 870 | 0 | 620 | 5027 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 32600 | 3800 | 1900 | 6352 | 1866 | 6865 | 11216 | 326 | 0 | 0 | 2534 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 6900 | 800 | 400 | 6407 | 5814 | 2278 | 8165 | 15838 | 441 | 8 | 0 | 2688 | 0 | 0 | 0 | 0 |
| 1981 | 8300 | 1100 | 11900 | 4166 | 4591 | 8596 | 2200 | 4512 | 8280 | 345 | 103 | 114 | 964 | 0 | 0 | 0 |
| 1982 | 22600 | 1100 | 200 | 13817 | 7892 | 4507 | 6258 | 1960 | 5075 | 6047 | 121 | 37 | 37 | 121 | 0 | 0 |


| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1983 | 127000 | 4680 | 1670 | 3183 | 21191 | 9521 | 6181 | 6823 | 1293 | 4598 | 7329 | 143 | 40 | 143 | 860 | 0 |
| 1984 | 33860 | 1700 | 2490 | 4483 | 5388 | 61543 | 18202 | 12638 | 15608 | 7215 | 16338 | 6478 | 0 | 0 | 0 | 1650 |
| 1985 | 28570 | 13150 | 207220 | 21500 | 15500 | 16500 | 130000 | 59000 | 55000 | 63000 | 10000 | 31000 | 50000 | 0 | 0 | 2640 |
| 1986 | 13810 | 1380 | 3090 | 539785 | 17594 | 14500 | 15500 | 105000 | 75000 | 42000 | 77000 | 19469 | 66000 | 80000 | 0 | 2470 |
| 1987 | 13850 | 6330 | 35770 | 19776 | 501393 | 18672 | 3502 | 7058 | 28000 | 12000 | 9500 | 4500 | 7834 | 6500 | 7000 | 450 |
| 1988 | 15490 | 2790 | 9110 | 62923 | 25059 | 550367 | 9452 | 3679 | 5964 | 14583 | 8872 | 2818 | 3356 | 2682 | 1560 | 540 |
| 1989 | 7120 | 1930 | 25200 | 2890 | 3623 | 5650 | 324290 | 3469 | 800 | 679 | 3297 | 1375 | 679 | 321 | 260 | 0 |
| 1990 | 1020 | 400 | 15540 | 18633 | 2658 | 11875 | 10854 | 226280 | 1289 | 1519 | 2036 | 2415 | 646 | 179 | 590 | 480 |
| 1991 | 100 | 3370 | 3330 | 8438 | 2780 | 1410 | 14698 | 8867 | 218851 | 2499 | 461 | 87 | 690 | 103 | 260 | 540 |
| 1992 | 1630 | 150 | 1340 | 12586 | 33100 | 4980 | 1193 | 11981 | 5748 | 225677 | 2483 | 639 | 247 | 1236 | 0 | 0 |
| 1993 | 6570 | 130 | 7240 | 28408 | 106866 | 87269 | 8625 | 3648 | 29603 | 18631 | 410110 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 430 | 20 | 8100 | 32500 | 110090 | 363920 | 164800 | 15580 | 8140 | 37330 | 35660 | 645410 | 2830 | 460 | 100 | 2070 |
| 1995 | 0 | 0 | 1130 | 57590 | 346460 | 622810 | 637840 | 231090 | 15510 | 15850 | 69750 | 83740 | 911880 | 4070 | 250 | 450 |
| 1996 | 0 | 0 | 30140 | 34360 | 713620 | 1571000 | 940580 | 406280 | 103410 | 5680 | 7370 | 66090 | 17570 | 836550 | 0 | 0 |
| 1997 | 0 | 0 | 21820 | 130450 | 270950 | 1795780 | 1993620 | 761210 | 326490 | 60870 | 20020 | 32400 | 90520 | 19120 | 370330 | 300 |
| 1998 | 0 | 0 | 82891 | 70323 | 242365 | 368310 | 1760319 | 1263750 | 381482 | 129971 | 42502 | 25343 | 3478 | 112604 | 5633 | 108514 |
| 1999 | 0 | 0 | 5029 | 137626 | 35820 | 134813 | 429433 | 1604959 | 1164263 | 291394 | 106005 | 14524 | 40040 | 7202 | 88598 | 63983 |


| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 2000 | 0 | 0 | 14395 | 84016 | 560379 | 34933 | 110719 | 404460 | 1299253 | 1045001 | 216980 | 71589 | 16260 | 22701 | 23321 | 71811 |
| 2001 | 0 | 0 | 2076 | 102293 | 160678 | 426822 | 38749 | 95991 | 296460 | 839136 | 507106 | 73673 | 23722 | 3505 | 3356 | 22164 |
| 2002 | 0 | 0 | 62031 | 198360 | 643161 | 255516 | 326495 | 29843 | 93530 | 264675 | 663059 | 339326 | 52922 | 12437 | 7000 | 10087 |
| 2003 | 0 | 3461 | 4524 | 75243 | 323958 | 730468 | 175878 | 167776 | 22866 | 74494 | 217108 | 567253 | 219097 | 38555 | 8111 | 6192 |
| 2004 | 125 | 1846 | 43800 | 24299 | 92300 | 429510 | 714433 | 111022 | 137940 | 26656 | 52467 | 169196 | 401564 | 210547 | 28028 | 11883 |
| 2005 | 0 | 442 | 20411 | 447788 | 94206 | 170547 | 643600 | 930309 | 121856 | 123291 | 37967 | 65289 | 139331 | 344822 | 126879 | 15697 |
| 2006 | 0 | 1968 | 45438 | 75824 | 729898 | 82107 | 171370 | 726041 | 772217 | 88701 | 77115 | 30339 | 57882 | 133665 | 142240 | 49128 |
| 2007 | 0 | 4475 | 8450 | 224636 | 366983 | 1804495 | 152916 | 242923 | 728836 | 511664 | 47215 | 25384 | 15316 | 24488 | 64755 | 58465 |
| 2008 | 0 | 39898 | 123949 | 36630 | 550274 | 670681 | 2295912 | 199592 | 256132 | 586583 | 369620 | 29633 | 36025 | 23775 | 25195 | 63176 |
| 2009 | 0 | 3468 | 113424 | 192641 | 149075 | 1193781 | 914748 | 1929631 | 142931 | 262037 | 423972 | 238174 | 45519 | 9337 | 10153 | 70538 |
| 2010 | 0 | 75981 | 61673 | 101948 | 209295 | 189784 | 1064866 | 711951 | 1421939 | 175010 | 180164 | 340781 | 179039 | 12558 | 11602 | 49773 |
| 2011 | 0 | 126972 | 249809 | 61706 | 104634 | 234330 | 210165 | 755382 | 543212 | 642787 | 90515 | 117230 | 136509 | 45082 | 6628 | 11638 |
| 2012 | 0 | 2680 | 13083 | 211630 | 49999 | 119627 | 281908 | 263330 | 747839 | 314694 | 357902 | 53109 | 44982 | 64273 | 12420 | 3604 |
| 2013 | 0 | 1 | 20715 | 60364 | 276901 | 71287 | 112558 | 283658 | 242243 | 591912 | 169525 | 145318 | 24936 | 10614 | 9725 | 2299 |
| 2014 | 0 | 265 | 1441 | 28301 | 57838 | 257529 | 50424 | 71721 | 194814 | 147083 | 381317 | 83050 | 57315 | 12746 | 1809 | 7501 |
| 2015 | 0 | 647 | 3244 | 16139 | 55749 | 52369 | 152347 | 34046 | 65728 | 156075 | 103393 | 201141 | 24310 | 49373 | 3369 | 6397 |
| 2016 | 0 | 197 | 2351 | 45483 | 43416 | 112147 | 85937 | 164454 | 52267 | 73576 | 174655 | 96476 | 179051 | 38546 | 32880 | 8379 |


| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 2017 | 0 | 618 | 16390 | 64275 | 305483 | 114976 | 248192 | 162566 | 289931 | 98836 | 133145 | 276874 | 107473 | 220368 | 22357 | 49442 |
| 2018 | 0 | 1261 | 22414 | 25638 | 59802 | 264182 | 150759 | 179628 | 109121 | 180968 | 85954 | 99061 | 212052 | 113841 | 136096 | 39249 |
| 2019 | 0 | 769 | 2205 | 148669 | 64237 | 185336 | 557804 | 146597 | 217346 | 119855 | 167569 | 133910 | 104730 | 220400 | 91773 | 121229 |
| 2020 | 0 | 1299 | 8252 | 49455 | 544337 | 70633 | 150932 | 412498 | 118081 | 156696 | 94975 | 188852 | 100408 | 96557 | 132619 | 103350 |
| 2021 | 204 | 3644 | 2368 | 25015 | 110359 | 1432164 | 162903 | 203923 | 345729 | 117846 | 127846 | 73558 | 68834 | 60477 | 40165 | 113929 |

Table 4.4.4.1. Norwegian spring spawning herring. Weight at age in the catch (kg).

| age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1950 | 0.007 | 0.025 | 0.058 | 0.110 | 0.188 | 0.211 | 0.234 | 0.253 | 0.266 | 0.280 | 0.294 | 0.303 | 0.312 | 0.32 | 0.323 | 0.334 |
| 1951 | 0.009 | 0.029 | 0.068 | 0.130 | 0.222 | 0.249 | 0.276 | 0.298 | 0.314 | 0.330 | 0.346 | 0.357 | 0.368 | 0.377 | 0.381 | 0.394 |
| 1952 | 0.008 | 0.026 | 0.061 | 0.115 | 0.197 | 0.221 | 0.245 | 0.265 | 0.279 | 0.293 | 0.308 | 0.317 | 0.327 | 0.335 | 0.339 | 0.349 |
| 1953 | 0.008 | 0.027 | 0.063 | 0.120 | 0.205 | 0.230 | 0.255 | 0.275 | 0.290 | 0.305 | 0.320 | 0.330 | 0.34 | 0.347 | 0.351 | 0.363 |
| 1954 | 0.008 | 0.026 | 0.062 | 0.117 | 0.201 | 0.225 | 0.250 | 0.269 | 0.284 | 0.299 | 0.313 | 0.323 | 0.333 | 0.341 | 0.345 | 0.356 |
| 1955 | 0.008 | 0.027 | 0.063 | 0.119 | 0.204 | 0.229 | 0.254 | 0.274 | 0.289 | 0.304 | 0.318 | 0.328 | 0.338 | 0.346 | 0.350 | 0.362 |
| 1956 | 0.008 | 0.028 | 0.066 | 0.126 | 0.215 | 0.241 | 0.268 | 0.289 | 0.304 | 0.320 | 0.336 | 0.346 | 0.357 | 0.365 | 0.369 | 0.382 |
| 1957 | 0.008 | 0.028 | 0.066 | 0.127 | 0.216 | 0.243 | 0.269 | 0.290 | 0.306 | 0.322 | 0.338 | 0.348 | 0.359 | 0.367 | 0.371 | 0.384 |
| 1958 | 0.009 | 0.030 | 0.070 | 0.133 | 0.227 | 0.255 | 0.283 | 0.305 | 0.321 | 0.338 | 0.355 | 0.366 | 0.377 | 0.386 | 0.390 | 0.403 |


|  | age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1959 | 0.009 | 0.030 | 0.071 | 0.135 | 0.231 | 0.259 | 0.287 | 0.310 | 0.327 | 0.344 | 0.360 | 0.372 | 0.383 | 0.392 | 0.397 | 0.409 |
| 1960 | 0.006 | 0.011 | 0.074 | 0.119 | 0.188 | 0.277 | 0.337 | 0.318 | 0.363 | 0.379 | 0.360 | 0.420 | 0.411 | 0.439 | 0.450 | 0.447 |
| 1961 | 0.006 | 0.010 | 0.045 | 0.087 | 0.159 | 0.276 | 0.322 | 0.372 | 0.363 | 0.393 | 0.407 | 0.397 | 0.422 | 0.447 | 0.465 | 0.452 |
| 1962 | 0.009 | 0.023 | 0.055 | 0.085 | 0.148 | 0.288 | 0.333 | 0.360 | 0.352 | 0.350 | 0.374 | 0.384 | 0.374 | 0.394 | 0.399 | 0.414 |
| 1963 | 0.008 | 0.026 | 0.047 | 0.098 | 0.171 | 0.275 | 0.268 | 0.323 | 0.329 | 0.336 | 0.341 | 0.358 | 0.385 | 0.353 | 0.381 | 0.386 |
| 1964 | 0.009 | 0.024 | 0.059 | 0.139 | 0.219 | 0.239 | 0.298 | 0.295 | 0.339 | 0.350 | 0.358 | 0.351 | 0.367 | 0.375 | 0.372 | 0.433 |
| 1965 | 0.009 | 0.016 | 0.048 | 0.089 | 0.217 | 0.234 | 0.262 | 0.331 | 0.360 | 0.367 | 0.386 | 0.395 | 0.393 | 0.404 | 0.401 | 0.431 |
| 1966 | 0.008 | 0.017 | 0.040 | 0.063 | 0.246 | 0.260 | 0.265 | 0.301 | 0.410 | 0.425 | 0.456 | 0.460 | 0.467 | 0.446 | 0.459 | 0.472 |
| 1967 | 0.009 | 0.015 | 0.036 | 0.066 | 0.093 | 0.305 | 0.305 | 0.310 | 0.333 | 0.359 | 0.413 | 0.446 | 0.401 | 0.408 | 0.439 | 0.430 |
| 1968 | 0.010 | 0.027 | 0.049 | 0.075 | 0.108 | 0.158 | 0.375 | 0.383 | 0.364 | 0.382 | 0.441 | 0.410 |  | 0.517 | 0.491 | 0.485 |
| 1969 | 0.009 | 0.021 | 0.047 | 0.072 |  | 0.152 | 0.296 |  | 0.329 | 0.329 | 0.341 |  |  |  |  | 0.429 |
| 1970 | 0.008 | 0.058 | 0.085 | 0.105 | 0.171 |  | 0.216 | 0.277 | 0.298 | 0.304 | 0.305 | 0.309 |  |  |  | 0.376 |
| 1971 | 0.011 | 0.053 | 0.121 | 0.177 | 0.216 | 0.250 |  | 0.305 | 0.333 |  | 0.366 | 0.377 | 0.388 |  |  |  |
| 1972 | 0.011 | 0.029 | 0.062 | 0.103 | 0.154 | 0.215 | 0.258 |  | 0.322 |  |  |  |  |  |  |  |
| 1973 | 0.006 | 0.053 | 0.106 | 0.161 | 0.213 |  | 0.255 |  |  |  |  |  |  |  |  |  |
| 1974 | 0.006 | 0.055 | 0.117 |  |  | 0.249 |  |  |  |  |  |  |  |  |  |  |
| 1975 | 0.009 | 0.079 | 0.169 | 0.241 |  |  | 0.381 |  |  |  |  |  |  |  |  |  |


| Year | age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1976 | 0.007 | 0.062 | 0.132 | 0.189 | 0.250 |  |  | 0.323 |  |  |  |  |  |  |  |  |
| 1977 | 0.011 | 0.091 | 0.193 | 0.316 | 0.350 |  |  |  | 0.511 |  |  |  |  |  |  |  |
| 1978 | 0.012 | 0.100 | 0.210 | 0.274 | 0.424 | 0.454 |  |  |  | 0.613 |  |  |  |  |  |  |
| 1979 | 0.010 | 0.088 | 0.181 | 0.293 | 0.359 | 0.416 | 0.436 |  |  |  | 0.553 |  |  |  |  |  |
| 1980 | 0.012 |  |  | 0.266 | 0.399 | 0.449 | 0.460 | 0.485 |  |  |  | 0.608 |  |  |  |  |
| 1981 | 0.010 | 0.082 | 0.163 | 0.196 | 0.291 | 0.341 | 0.368 | 0.380 | 0.397 |  |  |  |  |  |  |  |
| 1982 | 0.010 | 0.087 | 0.159 | 0.256 | 0.312 | 0.378 | 0.415 | 0.435 | 0.449 | 0.448 |  |  |  |  |  |  |
| 1983 | 0.011 | 0.090 | 0.165 | 0.217 | 0.265 | 0.337 | 0.378 | 0.410 | 0.426 | 0.435 | 0.444 |  |  |  |  |  |
| 1984 | 0.009 | 0.047 | 0.145 | 0.218 | 0.262 | 0.325 | 0.346 | 0.381 | 0.400 | 0.413 | 0.405 | 0.426 |  |  |  | 0.415 |
| 1985 | 0.009 | 0.022 | 0.022 | 0.214 | 0.277 | 0.295 | 0.338 | 0.360 | 0.381 | 0.397 | 0.409 | 0.417 | 0.435 |  |  | 0.435 |
| 1986 | 0.007 | 0.077 | 0.097 | 0.055 | 0.249 | 0.294 | 0.312 | 0.352 | 0.374 | 0.398 | 0.402 | 0.401 | 0.410 | 0.410 |  | 0.410 |
| 1987 | 0.010 | 0.075 | 0.091 | 0.124 | 0.173 | 0.253 | 0.232 | 0.312 | 0.328 | 0.349 | 0.353 | 0.370 | 0.385 | 0.385 | 0.385 |  |
| 1988 | 0.008 | 0.062 | 0.075 | 0.124 | 0.154 | 0.194 | 0.241 | 0.265 | 0.304 | 0.305 | 0.317 | 0.308 | 0.334 | 0.334 | 0.334 |  |
| 1989 | 0.010 | 0.060 | 0.204 | 0.188 | 0.264 | 0.260 | 0.282 | 0.306 |  |  | 0.422 | 0.364 |  |  |  |  |
| 1990 | 0.007 |  | 0.102 | 0.230 | 0.239 | 0.266 | 0.305 | 0.308 | 0.376 | 0.407 | 0.412 | 0.424 |  |  |  |  |
| 1991 |  | 0.015 | 0.104 | 0.208 | 0.250 | 0.288 | 0.312 | 0.316 | 0.330 | 0.344 |  |  |  |  |  |  |
| 1992 | 0.007 |  | 0.103 | 0.191 | 0.233 | 0.304 | 0.337 | 0.365 | 0.361 | 0.371 | 0.403 |  |  | 0.404 |  |  |


| Year | age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1993 | 0.007 |  | 0.106 | 0.153 | 0.243 | 0.282 | 0.320 | 0.330 | 0.365 | 0.373 | 0.379 |  |  |  |  |  |
| 1994 |  |  | 0.102 | 0.194 | 0.239 | 0.280 | 0.317 | 0.328 | 0.356 | 0.372 | 0.390 | 0.379 | 0.399 | 0.403 |  |  |
| 1995 |  |  | 0.102 | 0.153 | 0.192 | 0.234 | 0.283 | 0.328 | 0.349 | 0.356 | 0.374 | 0.366 | 0.393 | 0.387 |  |  |
| 1996 |  |  | 0.136 | 0.136 | 0.168 | 0.206 | 0.262 | 0.309 | 0.337 | 0.366 | 0.360 | 0.361 | 0.367 | 0.379 |  |  |
| 1997 |  |  | 0.089 | 0.167 | 0.184 | 0.207 | 0.232 | 0.277 | 0.305 | 0.331 | 0.328 | 0.344 | 0.343 | 0.397 | 0.357 |  |
| 1998 |  |  | 0.111 | 0.150 | 0.216 | 0.221 | 0.249 | 0.277 | 0.316 | 0.338 | 0.374 | 0.372 | 0.366 | 0.396 | 0.377 | 0.406 |
| 1999 |  |  | 0.096 | 0.173 | 0.228 | 0.262 | 0.274 | 0.292 | 0.307 | 0.335 | 0.362 | 0.371 | 0.399 | 0.396 | 0.400 | 0.404 |
| 2000 |  |  | 0.124 | 0.175 | 0.222 | 0.242 | 0.289 | 0.303 | 0.310 | 0.328 | 0.349 | 0.383 | 0.411 | 0.410 | 0.419 | 0.409 |
| 2001 |  |  | 0.105 | 0.166 | 0.214 | 0.252 | 0.268 | 0.305 | 0.308 | 0.322 | 0.337 | 0.363 | 0.353 | 0.378 | 0.400 | 0.427 |
| 2002 |  |  | 0.056 | 0.128 | 0.198 | 0.255 | 0.281 | 0.303 | 0.322 | 0.323 | 0.334 | 0.345 | 0.369 | 0.407 | 0.410 | 0.435 |
| 2003 |  | 0.062 | 0.068 | 0.169 | 0.218 | 0.257 | 0.288 | 0.316 | 0.323 | 0.348 | 0.354 | 0.351 | 0.363 | 0.372 | 0.376 | 0.429 |
| 2004 | 0.022 | 0.066 | 0.143 | 0.18 | 0.227 | 0.26 | 0.29 | 0.323 | 0.355 | 0.375 | 0.383 | 0.399 | 0.395 | 0.405 | 0.429 | 0.439 |
| 2005 |  | 0.092 | 0.106 | 0.181 | 0.235 | 0.266 | 0.290 | 0.315 | 0.344 | 0.367 | 0.384 | 0.372 | 0.384 | 0.398 | 0.402 | 0.413 |
| 2006 |  | 0.055 | 0.102 | 0.171 | 0.238 | 0.268 | 0.292 | 0.311 | 0.330 | 0.365 | 0.374 | 0.376 | 0.388 | 0.396 | 0.398 | 0.407 |
| 2007 | 0.000 | 0.074 | 0.137 | 0.162 | 0.228 | 0.271 | 0.316 | 0.332 | 0.342 | 0.358 | 0.361 | 0.381 | 0.390 | 0.400 | 0.405 | 0.399 |
| 2008 | 0.000 | 0.026 | 0.106 | 0.145 | 0.209 | 0.254 | 0.296 | 0.318 | 0.341 | 0.353 | 0.363 | 0.367 | 0.395 | 0.396 | 0.386 | 0.413 |
| 2009 |  | 0.040 | 0.156 | 0.184 | 0.220 | 0.251 | 0.291 | 0.311 | 0.338 | 0.347 | 0.363 | 0.375 | 0.382 | 0.375 | 0.375 | 0.387 |


| Year | age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 2010 |  | 0.059 | 0.107 | 0.177 | 0.218 | 0.261 | 0.279 | 0.311 | 0.325 | 0.343 | 0.362 | 0.370 | 0.388 | 0.391 | 0.376 | 0.441 |
| 2011 |  | 0.011 | 0.098 | 0.200 | 0.257 | 0.273 | 0.300 | 0.316 | 0.340 | 0.348 | 0.365 | 0.371 | 0.387 | 0.374 | 0.403 | 0.401 |
| 2012 |  | 0.034 | 0.126 | 0.211 | 0.272 | 0.301 | 0.308 | 0.331 | 0.335 | 0.351 | 0.354 | 0.370 | 0.389 | 0.389 | 0.382 | 0.388 |
| 2013 |  | 0.048 | 0.163 | 0.237 | 0.276 | 0.300 | 0.331 | 0.339 | 0.351 | 0.357 | 0.370 | 0.373 | 0.394 | 0.391 | 0.389 | 0.367 |
| 2014 |  | 0.057 | 0.179 | 0.233 | 0.271 | 0.293 | 0.322 | 0.342 | 0.353 | 0.367 | 0.365 | 0.374 | 0.375 | 0.378 | 0.418 | 0.371 |
| 2015 |  | 0.059 | 0.146 | 0.203 | 0.272 | 0.323 | 0.331 | 0.358 | 0.370 | 0.372 | 0.383 | 0.382 | 0.392 | 0.386 | 0.383 | 0.391 |
| 2016 |  | 0.048 | 0.111 | 0.212 | 0.255 | 0.290 | 0.333 | 0.339 | 0.361 | 0.367 | 0.370 | 0.381 | 0.378 | 0.388 | 0.383 | 0.395 |
| 2017 |  | 0.092 | 0.143 | 0.205 | 0.241 | 0.292 | 0.322 | 0.350 | 0.360 | 0.382 | 0.392 | 0.391 | 0.396 | 0.399 | 0.407 | 0.394 |
| 2018 |  | 0.068 | 0.127 | 0.207 | 0.240 | 0.276 | 0.321 | 0.348 | 0.371 | 0.380 | 0.399 | 0.404 | 0.400 | 0.407 | 0.408 | 0.418 |
| 2019 |  | 0.135 | 0.186 | 0.209 | 0.235 | 0.269 | 0.298 | 0.327 | 0.345 | 0.376 | 0.387 | 0.403 | 0.409 | 0.423 | 0.417 | 0.449 |
| 2020 |  | 0.131 | 0.170 | 0.204 | 0.236 | 0.274 | 0.306 | 0.317 | 0.342 | 0.358 | 0.374 | 0.395 | 0.402 | 0.408 | 0.415 | 0.444 |
| 2021 | 0.050 | 0.122 | 0.130 | 0.195 | 0.229 | 0.256 | 0.278 | 0.319 | 0.325 | 0.363 | 0.364 | 0.384 | 0.386 | 0.397 | 0.412 | 0.431 |

Table 4.4.4.2. Norwegian spring spawning herring. Weight at age in the stock (kg).

| Year | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1950 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.230 | 0.255 | 0.275 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1951 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.230 | 0.255 | 0.275 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1952 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.230 | 0.255 | 0.275 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1953 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.230 | 0.255 | 0.275 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1954 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.230 | 0.255 | 0.275 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1955 | 0.001 | 0.008 | 0.047 | 0.100 | 0.195 | 0.213 | 0.260 | 0.275 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1956 | 0.001 | 0.008 | 0.047 | 0.100 | 0.205 | 0.230 | 0.249 | 0.275 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1957 | 0.001 | 0.008 | 0.047 | 0.100 | 0.136 | 0.228 | 0.255 | 0.262 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1958 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.242 | 0.292 | 0.295 | 0.293 | 0.305 | 0.315 | 0.330 | 0.340 | 0.345 | 0.352 | 0.363 |
| 1959 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.252 | 0.260 | 0.290 | 0.300 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.358 |
| 1960 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.270 | 0.291 | 0.293 | 0.321 | 0.318 | 0.320 | 0.344 | 0.349 | 0.370 | 0.379 | 0.378 |
| 1961 | 0.001 | 0.008 | 0.047 | 0.100 | 0.232 | 0.250 | 0.292 | 0.302 | 0.304 | 0.323 | 0.322 | 0.321 | 0.344 | 0.357 | 0.363 | 0.368 |
| 1962 | 0.001 | 0.008 | 0.047 | 0.100 | 0.219 | 0.291 | 0.300 | 0.316 | 0.324 | 0.326 | 0.335 | 0.338 | 0.334 | 0.347 | 0.354 | 0.358 |
| 1963 | 0.001 | 0.008 | 0.047 | 0.100 | 0.185 | 0.253 | 0.294 | 0.312 | 0.329 | 0.327 | 0.334 | 0.341 | 0.349 | 0.341 | 0.358 | 0.375 |
| 1964 | 0.001 | 0.008 | 0.047 | 0.100 | 0.194 | 0.213 | 0.264 | 0.317 | 0.363 | 0.353 | 0.349 | 0.354 | 0.357 | 0.359 | 0.365 | 0.402 |
| 1965 | 0.001 | 0.008 | 0.047 | 0.100 | 0.186 | 0.199 | 0.236 | 0.260 | 0.363 | 0.350 | 0.370 | 0.360 | 0.378 | 0.387 | 0.390 | 0.394 |
| 1966 | 0.001 | 0.008 | 0.047 | 0.100 | 0.185 | 0.219 | 0.222 | 0.249 | 0.306 | 0.354 | 0.377 | 0.391 | 0.379 | 0.378 | 0.361 | 0.383 |


| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 |  | 2 |  | 3 |  | 4 |  | 5 |  | 6 |  | 7 |  | 8 |  | 9 |  | 10 |  | 11 |  | 12 |  | 13 |  | 14 |  | 15+ |
| 1967 | 0.001 | 0.008 |  | 0.047 |  | 0.100 |  | 0.180 |  | 0.228 |  | 0.269 |  | 0.270 |  | 0.294 |  | 0.324 |  | 0.420 |  | 0.430 |  | 0.366 |  | 0.368 |  | 0.433 |  | 0.414 |
| 1968 | 0.001 | 0.008 |  | 0.047 |  | 0.100 |  | 0.115 |  | 0.206 |  | 0.266 |  | 0.275 |  | 0.274 |  | 0.285 |  | 0.350 |  | 0.325 |  | 0.363 |  | 0.408 |  | 0.388 |  | 0.378 |
| 1969 | 0.001 | 0.008 |  | 0.047 |  | 0.100 |  | 0.115 |  | 0.145 |  | 0.270 |  | 0.300 |  | 0.306 |  | 0.308 |  | 0.318 |  | 0.340 |  | 0.368 |  | 0.360 |  | 0.393 |  | 0.397 |
| 1970 | 0.001 | 0.008 |  | 0.047 |  | 0.100 |  | 0.209 |  | 0.272 |  | 0.230 |  | 0.295 |  | 0.317 |  | 0.323 |  | 0.325 |  | 0.329 |  | 0.380 |  | 0.370 |  | 0.380 |  | 0.391 |
| 1971 | 0.001 | 0.015 |  | 0.080 |  | 0.100 |  | 0.190 |  | 0.225 |  | 0.250 |  | 0.275 |  | 0.290 |  | 0.310 |  | 0.325 |  | 0.335 |  | 0.345 |  | 0.355 |  | 0.365 |  | 0.390 |
| 1972 | 0.001 | 0.010 |  | 0.070 |  | 0.150 |  | 0.150 |  | 0.140 |  | 0.210 |  | 0.240 |  | 0.270 |  | 0.300 |  | 0.325 |  | 0.335 |  | 0.345 |  | 0.355 |  | 0.365 |  | 0.390 |
| 1973 | 0.001 | 0.010 |  | 0.085 |  | 0.170 |  | 0.259 |  | 0.342 |  | 0.384 |  | 0.409 |  | 0.404 |  | 0.461 |  | 0.520 |  | 0.534 |  | 0.500 |  | 0.500 |  | 0.500 |  | 0.500 |
| 1974 | 0.001 | 0.010 |  | 0.085 |  | 0.170 |  | 0.259 |  | 0.342 |  | 0.384 |  | 0.409 |  | 0.444 |  | 0.461 |  | 0.520 |  | 0.543 |  | 0.482 |  | 0.482 |  | 0.482 |  | 0.482 |
| 1975 | 0.001 | 0.010 |  | 0.085 |  | 0.181 |  | 0.259 |  | 0.342 |  | 0.384 |  | 0.409 |  | 0.444 |  | 0.461 |  | 0.520 |  | 0.543 |  | 0.482 |  | 0.482 |  | 0.482 |  | 0.482 |
| 1976 | 0.001 | 0.010 |  | 0.085 |  | 0.181 |  | 0.259 |  | 0.342 |  | 0.384 |  | 0.409 |  | 0.444 |  | 0.461 |  | 0.520 |  | 0.543 |  | 0.482 |  | 0.482 |  | 0.482 |  | 0.482 |
| 1977 | 0.001 | 0.010 |  | 0.085 |  | 0.181 |  | 0.259 |  | 0.343 |  | 0.384 |  | 0.409 |  | 0.444 |  | 0.461 |  | 0.520 |  | 0.543 |  | 0.482 |  | 0.482 |  | 0.482 |  | 0.482 |
| 1978 | 0.001 | 0.010 |  | 0.085 |  | 0.180 |  | 0.294 |  | 0.326 |  | 0.371 |  | 0.409 |  | 0.461 |  | 0.476 |  | 0.520 |  | 0.543 |  | 0.500 |  | 0.500 |  | 0.500 |  | 0.500 |
| 1979 | 0.001 | 0.010 |  | 0.085 |  | 0.178 |  | 0.232 |  | 0.359 |  | 0.385 |  | 0.420 |  | 0.444 |  | 0.505 |  | 0.520 |  | 0.551 |  | 0.500 |  | 0.500 |  | 0.500 |  | 0.500 |
| 1980 | 0.001 | 0.010 |  | 0.085 |  | 0.175 |  | 0.283 |  | 0.347 |  | 0.402 |  | 0.421 |  | 0.465 |  | 0.465 |  | 0.520 |  | 0.534 |  | 0.500 |  | 0.500 |  | 0.500 |  | 0.500 |
| 1981 | 0.001 | 0.010 |  | 0.085 |  | 0.170 |  | 0.224 |  | 0.336 |  | 0.378 |  | 0.387 |  | 0.408 |  | 0.397 |  | 0.520 |  | 0.543 |  | 0.512 |  | 0.512 |  | 0.512 |  | 0.512 |
| 1982 | 0.001 | 0.010 |  | 0.085 |  | 0.170 |  | 0.204 |  | 0.303 |  | 0.355 |  | 0.383 |  | 0.395 |  | 0.413 |  | 0.453 |  | 0.468 |  | 0.506 |  | 0.506 |  | 0.506 |  | 0.506 |
| 1983 |  |  | 0.010 |  | 0.085 |  | 0.155 |  | 0.249 |  | 0.304 |  | 0.368 |  | 0.404 |  | 0.424 |  | 0.437 |  | 0.436 |  | 0.493 |  | 0.495 |  | 0.495 |  | 0.495 | 0.495 |


| Year | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |  | 15+ |
| 1984 |  | 0.001 | 0.010 | 0.085 | 0.140 | 0.204 | 0.295 | 0.338 | 0.376 | 0.395 | 0.407 | 0.413 | 0.422 | 0.437 | 0.437 | 0.437 | 0.437 |
| 1985 |  | 0.001 | 0.010 | 0.085 | 0.148 | 0.234 | 0.265 | 0.312 | 0.346 | 0.370 | 0.395 | 0.397 | 0.428 | 0.428 | 0.428 | 0.428 | 0.428 |
| 1986 |  | 0.001 | 0.010 | 0.085 | 0.054 | 0.206 | 0.265 | 0.289 | 0.339 | 0.368 | 0.391 | 0.382 | 0.388 | 0.395 | 0.395 | 0.395 | 0.395 |
| 1987 |  | 0.001 | 0.010 | 0.055 | 0.090 | 0.143 | 0.241 | 0.279 | 0.299 | 0.316 | 0.342 | 0.343 | 0.362 | 0.376 | 0.376 | 0.376 | 0.376 |
| 1988 |  | 0.001 | 0.015 | 0.050 | 0.098 | 0.135 | 0.197 | 0.277 | 0.315 | 0.339 | 0.343 | 0.359 | 0.365 | 0.376 | 0.376 | 0.376 | 0.376 |
| 1989 |  | 0.001 | 0.015 | 0.100 | 0.154 | 0.175 | 0.209 | 0.252 | 0.305 | 0.367 | 0.377 | 0.359 | 0.395 | 0.396 | 0.396 | 0.396 | 0.396 |
| 1990 |  | 0.001 | 0.008 | 0.048 | 0.219 | 0.198 | 0.258 | 0.288 | 0.309 | 0.428 | 0.370 | 0.403 | 0.387 | 0.440 | 0.440 | 0.440 | 0.44 |
| 1991 |  | 0.001 | 0.011 | 0.037 | 0.147 | 0.210 | 0.244 | 0.300 | 0.324 | 0.336 | 0.343 | 0.382 | 0.366 | 0.425 | 0.425 | 0.425 | 0.425 |
| 1992 |  | 0.001 | 0.007 | 0.030 | 0.128 | 0.224 | 0.296 | 0.327 | 0.355 | 0.345 | 0.367 | 0.341 | 0.361 | 0.430 | 0.470 | 0.470 | 0.46 |
| 1993 |  | 0.001 | 0.008 | 0.025 | 0.081 | 0.201 | 0.265 | 0.323 | 0.354 | 0.358 | 0.381 | 0.369 | 0.396 | 0.393 | 0.374 | 0.403 | 0.4 |
| 1994 |  | 0.001 | 0.010 | 0.025 | 0.075 | 0.151 | 0.254 | 0.318 | 0.371 | 0.347 | 0.412 | 0.382 | 0.407 | 0.410 | 0.410 | 0.410 | 0.41 |
| 1995 |  | 0.001 | 0.018 | 0.025 | 0.066 | 0.138 | 0.230 | 0.296 | 0.346 | 0.388 | 0.363 | 0.409 | 0.414 | 0.422 | 0.410 | 0.410 | 0.426 |
| 1996 |  | 0.001 | 0.018 | 0.025 | 0.076 | 0.118 | 0.188 | 0.261 | 0.316 | 0.346 | 0.374 | 0.390 | 0.390 | 0.384 | 0.398 | 0.398 | 0.398 |
| 1997 |  | 0.001 | 0.018 | 0.025 | 0.096 | 0.118 | 0.174 | 0.229 | 0.286 | 0.323 | 0.370 | 0.378 | 0.386 | 0.360 | 0.393 | 0.391 | 0.391 |
| 1998 |  | 0.001 | 0.018 | 0.025 | 0.074 | 0.147 | 0.174 | 0.217 | 0.242 | 0.278 | 0.304 | 0.310 | 0.359 | 0.340 | 0.344 | 0.385 | 0.369 |
| 1999 |  | 0.001 | 0.018 | 0.025 | 0.102 | 0.150 | 0.223 | 0.240 | 0.264 | 0.283 | 0.315 | 0.345 | 0.386 | 0.386 | 0.386 | 0.382 | 0.395 |
| 2000 |  | 0.001 | 0.018 | 0.025 | 0.119 | 0.178 | 0.225 | 0.271 | 0.285 | 0.298 | 0.311 | 0.339 | 0.390 | 0.398 | 0.406 | 0.414 | 0.427 |


| Year | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $0 \quad 1$ | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |  | 15+ |
| 2001 | 0.001 | 0.018 | 0.025 | 0.075 | 0.178 | 0.238 | 0.247 | 0.296 | 0.307 | 0.314 | 0.328 | 0.351 | 0.376 | 0.406 | 0.414 | 0.425 |
| 2002 | 0.001 | 0.010 | 0.023 | 0.057 | 0.177 | 0.241 | 0.275 | 0.302 | 0.311 | 0.314 | 0.328 | 0.341 | 0.372 | 0.405 | 0.415 | 0.438 |
| 2003 | 0.001 | 0.010 | 0.055 | 0.098 | 0.159 | 0.211 | 0.272 | 0.305 | 0.292 | 0.331 | 0.337 | 0.347 | 0.356 | 0.381 | 0.414 | 0.433 |
| 2004 | 0.001 | 0.010 | 0.055 | 0.106 | 0.149 | 0.212 | 0.241 | 0.279 | 0.302 | 0.337 | 0.354 | 0.355 | 0.360 | 0.371 | 0.400 | 0.429 |
| 2005 | 0.001 | 0.010 | 0.046 | 0.112 | 0.156 | 0.234 | 0.267 | 0.295 | 0.330 | 0.363 | 0.377 | 0.414 | 0.406 | 0.308 | 0.420 | 0.452 |
| 2006 | 0.001 | 0.010 | 0.042 | 0.107 | 0.179 | 0.232 | 0.272 | 0.297 | 0.318 | 0.371 | 0.365 | 0.393 | 0.395 | 0.399 | 0.415 | 0.428 |
| 2007 | 0.001 | 0.010 | 0.036 | 0.086 | 0.155 | 0.226 | 0.265 | 0.312 | 0.310 | 0.364 | 0.384 | 0.352 | 0.386 | 0.304 | 0.420 | 0.412 |
| 2008** | 0.001 | 0.010 | 0.044 | 0.077 | 0.146 | 0.212 | 0.269 | 0.289 | 0.327 | 0.351 | 0.358 | 0.372 | 0.411 | 0.353 | 0.389 | 0.393 |
| 2009*** | 0.001 | 0.010 | 0.044 | 0.077 | 0.141 | 0.215 | 0.270 | 0.306 | 0.336 | 0.346 | 0.364 | 0.369 | 0.411 | 0.353 | 0.389 | 0.393 |
| 2010**** | 0.001 | 0.01 | 0.044 | 0.077 | 0.188 | 0.22 | 0.251 | 0.286 | 0.308 | 0.333 | 0.344 | 0.354 | 0.373 | 0.353 | 0.389 | 0.393 |
| 2011 | 0.001 | 0.01 | 0.044 | 0.118 | 0.185 | 0.209 | 0.246 | 0.277 | 0.310 | 0.322 | 0.339 | 0.349 | 0.364 | 0.363 | 0.389 | 0.393 |
| 2012 | 0.001 | 0.01 | 0.044 | 0.138 | 0.185 | 0.256 | 0.273 | 0.290 | 0.305 | 0.330 | 0.342 | 0.361 | 0.390 | 0.377 | 0.389 | 0.393 |
| 2013 | 0.001 | 0.01 | 0.044 | 0.138 | 0.204 | 0.267 | 0.305 | 0.309 | 0.320 | 0.328 | 0.346 | 0.350 | 0.390 | 0.377 | 0.389 | 0.393 |
| 2014 | 0.001 | 0.01 | 0.044 | 0.138 | 0.198 | 0.274 | 0.301 | 0.326 | 0.333 | 0.339 | 0.347 | 0.344 | 0.362 | 0.362 | 0.389 | 0.393 |
| 2015 | 0.001 | 0.01 | 0.044 | 0.138 | 0.187 | 0.243 | 0.299 | 0.326 | 0.319 | 0.345 | 0.346 | 0.354 | 0.382 | 0.376 | 0.389 | 0.393 |
| 2016 | 0.001 | 0.01 | 0.054 | 0.115 | 0.186 | 0.247 | 0.293 | 0.320 | 0.334 | 0.353 | 0.354 | 0.352 | 0.361 | 0.370 | 0.380 | 0.388 |
| 2017 | 0.001 | 0.01 | 0.054 | 0.115 | 0.190 | 0.247 | 0.282 | 0.322 | 0.338 | 0.351 | 0.359 | 0.361 | 0.361 | 0.368 | 0.380 | 0.386 |


| Year | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |  | 15+ |
| 2018 |  | 0.001 | 0.01 | 0.054 | 0.115 | 0.149 | 0.225 | 0.260 | 0.289 | 0.312 | 0.343 | 0.359 | 0.361 | 0.369 | 0.368 | 0.377 | 0.386 |
| 2019 |  | 0.001 | 0.01 | 0.054 | 0.104 | 0.151 | 0.203 | 0.277 | 0.311 | 0.331 | 0.355 | 0.353 | 0.363 | 0.381 | 0.376 | 0.385 | 0.382 |
| 2020 |  | 0.001 | 0.01 | 0.054 | 0.104 | 0.150 | 0.203 | 0.266 | 0.301 | 0.328 | 0.343 | 0.358 | 0.366 | 0.374 | 0.367 | 0.384 | 0.391 |
| 2021 |  | 0.001 | 0.01 | 0.054 | 0.104 | 0.160 | 0.209 | 0.266 | 0.284 | 0.302 | 0.325 | 0.352 | 0.366 | 0.384 | 0.376 | 0.404 | 0.391 |
| 2022 |  | 0.001 | 0.01 | 0.054 | 0.104 | 0.125 | 0.168 | 0.243 | 0.287 | 0.303 | 0.323 | 0.352 | 0.366 | 0.384 | 0.376 | 0.404 | 0.391 |

** mean weight at ages 11 and 13 are mean of 5 previous years at the same age. These age groups were not present in the catches of the wintering survey from which the stock weight are derived.
*** derived from catch data from the wintering area north of $69^{\circ} \mathrm{N}$ during December 2008 - January 2009 for age groups 4-11.
${ }^{* * * *}$ derived from catch data from the wintering area north of $69^{\circ} \mathrm{N}$ during January 2010 for age groups 4-12.

Table 4.4.5.1. Norwegian Spring-spawning herring. Maturity at age.

| Year/Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1951 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1952 | 0 | 0 | 0 | 0 | 0.1 | 0.6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1953 | 0 | 0 | 0 | 0 | 0.3 | 0.4 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1954 | 0 | 0 | 0 | 0 | 0.1 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1955 | 0 | 0 | 0 | 0.1 | 0.4 | 0.4 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1956 | 0 | 0 | 0 | 0 | 0.5 | 0.7 | 0.6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1957 | 0 | 0 | 0 | 0 | 0.3 | 0.8 | 0.8 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1958 | 0 | 0 | 0 | 0 | 0.3 | 0.5 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1959 | 0 | 0 | 0 | 0 | 0.7 | 0.8 | 1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1960 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1961 | 0 | 0 | 0 | 0 | 0.1 | 0.8 | 1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1962 | 0 | 0 | 0 | 0 | 0.1 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1963 | 0 | 0 | 0 | 0 | 0.1 | 0.4 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1964 | 0 | 0 | 0 | 0 | 0.1 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1965 | 0 | 0 | 0 | 0 | 0.5 | 0.4 | 0.9 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1966 | 0 | 0 | 0 | 0 | 0.5 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1967 | 0 | 0 | 0 | 0 | 0.3 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1968 | 0 | 0 | 0 | 0 | 0 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1969 | 0 | 0 | 0 | 0.1 | 0.2 | 0.3 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1970 | 0 | 0 | 0 | 0 | 0.4 | 0.3 | 0.4 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1971 | 0 | 0 | 0 | 0 | 0.1 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1972 | 0 | 0 | 0 | 0 | 0.4 | 0.3 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1973 | 0 | 0 | 0 | 0.1 | 0.6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1974 | 0 | 0 | 0 | 0 | 0.6 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1975 | 0 | 0 | 0 | 0.1 | 0.5 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1976 | 0 | 0 | 0 | 0.1 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1977 | 0 | 0 | 0 | 0.3 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1978 | 0 | 0 | 0 | 0.2 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |


| Year/Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 0 | 0 | 0 | 0.1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1980 | 0 | 0 | 0 | 0.1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1981 | 0 | 0 | 0 | 0.1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1982 | 0 | 0 | 0 | 0.1 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1983 | 0 | 0 | 0 | 0.1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1984 | 0 | 0 | 0 | 0.1 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1985 | 0 | 0 | 0 | 0.1 | 0.8 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1986 | 0 | 0 | 0 | 0 | 0.5 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1987 | 0 | 0 | 0 | 0 | 0.1 | 0.8 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1988 | 0 | 0 | 0 | 0 | 0.2 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1989 | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1990 | 0 | 0 | 0 | 0.2 | 0.5 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1991 | 0 | 0 | 0 | 0 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1992 | 0 | 0 | 0 | 0 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1993 | 0 | 0 | 0 | 0 | 0.5 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1994 | 0 | 0 | 0 | 0 | 0.1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0.6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0.5 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1997 | 0 | 0 | 0 | 0.1 | 0 | 0.4 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1998 | 0 | 0 | 0 | 0 | 0.6 | 0.4 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1999 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2000 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2001 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2002 | 0 | 0 | 0 | 0 | 0.1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2003 | 0 | 0 | 0 | 0 | 0.2 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2004 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2005 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2006 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2007 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2008 | 0 | 0 | 0 | 0 | 0.1 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |


| Year/Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 0 | 0 | 0 | 0 | 0.1 | 0.4 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2010 | 0 | 0 | 0 | 0 | 0.2 | 0.4 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2011 | 0 | 0 | 0 | 0 | 0.4 | 0.7 | 0.8 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2012 | 0 | 0 | 0 | 0 | 0.5 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2013 | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2014 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2015 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2016 | 0 | 0 | 0 | 0 | 0.1 | 0.5 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2017 | 0 | 0 | 0 | 0 | 0.2 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2018 | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2019 | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2020 | 0 | 0 | 0 | 0 | 0.1 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2021 | 0 | 0 | 0 | 0 | 0.4 | 0.6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2022 | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

Table 4.4.7.1. Norwegian Spring-spawning herring. Estimated indices (mean of bootstrap with 1000 iterations in StoX) from the acoustic surveys on the spawning grounds in February-March. Numbers in millions. Biomass in thousand tonnes. "Fleet 1"

| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | Total | Biomass |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 0 | 392 | 307 | 8015 | 81 | 33 | 12 | 36 | 22 | 45 | 0 | 0 | 0 | 0 | 8943 | 1621 |
| 1989 | 161 | 16 | 338 | 91 | 3973 | 101 | 12 | 4 | 55 | 0 | 4 | 42 | 0 | 9 | 4813 | 1169 |
| 1990 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1991 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1992 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1993 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1994 | 37 | 100 | 48 | 848 | 483 | 62 | 13 | 144 | 49 | 1836 | 4 | 4 | 0 | 0 | 3665 | 1207 |
| 1995 | 4 | 450 | 4679 | 3211 | 1957 | 299 | 20 | 0 | 106 | 55 | 2327 | 0 | 0 | 0 | 13745 | 2860 |
| 1996 | 119 | 186 | 1976 | 7960 | 2326 | 875 | 301 | 0 | 0 | 136 | 0 | 1760 | 0 | 0 | 15645 | 3366 |
| 1997 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1998 | 51 | 308 | 978 | 2982 | 12859 | 8133 | 1851 | 592 | 163 | 43 | 0 | 329 | 0 | 1400 | 29705 | 6886 |
| 1999 | 114 | 1530 | 369 | 1351 | 2669 | 9334 | 7004 | 1666 | 511 | 130 | 0 | 0 | 353 | 373 | 25438 | 6262 |
| 2000 | 1394 | 691 | 2600 | 109 | 477 | 1144 | 4282 | 2838 | 493 | 50 | 2 | 0 | 7 | 228 | 14315 | 3285 |
| 2001 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2002 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2003 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2004 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | Total | Biomass |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 38 | 238 | 661 | 2128 | 5947 | 8328 | 613 | 503 | 156 | 92 | 576 | 1152 | 587 | 9 | 21026 | 5260 |
| 2006 | 26 | 90 | 6054 | 548 | 882 | 3362 | 3311 | 110 | 86 | 20 | 89 | 58 | 246 | 63 | 14951 | 3431 |
| 2007 | 33 | 367 | 1618 | 12397 | 815 | 655 | 2956 | 3205 | 141 | 228 | 40 | 204 | 284 | 470 | 23427 | 5350 |
| 2008 | 15 | 48 | 2564 | 2824 | 8882 | 522 | 471 | 1566 | 1567 | 161 | 102 | 46 | 128 | 136 | 19090 | 4553 |
| 2009 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2010 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2011 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2012 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2013 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2014 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2015 | 204 | 533 | 2754 | 744 | 3267 | 388 | 692 | 2715 | 784 | 7222 | 367 | 1658 | 51 | 237 | 21662 | 6365 |
| 2016 | 18 | 197 | 237 | 594 | 365 | 2119 | 240 | 514 | 2930 | 652 | 3995 | 199 | 824 | 97 | 12982 | 4182 |
| 2017 | 19 | 110 | 1076 | 641 | 880 | 428 | 1326 | 181 | 206 | 2026 | 303 | 2542 | 80 | 729 | 10550 | 3314 |
| 2018 | 104 | 146 | 1720 | 2771 | 459 | 845 | 639 | 1095 | 444 | 370 | 1159 | 368 | 1538 | 354 | 12013 | 3262 |
| 2019 | 2 | 372 | 310 | 940 | 3778 | 754 | 879 | 660 | 1054 | 736 | 412 | 1807 | 182 | 2161 | 14166 | 4250 |
| 2020 | 6 | 44 | 3502 | 571 | 1212 | 3337 | 530 | 609 | 364 | 650 | 131 | 279 | 677 | 825 | 12750 | 3274 |
| 2021 | 21 | 112 | 293 | 10210 | 733 | 738 | 1932 | 427 | 451 | 312 | 219 | 395 | 208 | 1153 | 17250 | 4021 |
| 2022 | 27 | 72 | 162 | 760 | 6393 | 317 | 563 | 1515 | 301 | 486 | 301 | 255 | 385 | 630 | 12183 | 3302 |

Table 4.4.7.2. Norwegian spring-spawning herring. Acoustic estimates (billion individuals) of immature herring in the Barents Sea in May/June from IESNS. Values in the years 2009-2022 are estimated with StoX (mean of bootstrap with 1000 iterations). "Fleet 4".

| age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 |
| 1991 | 24.3 | 5.2 |  |  |  |
| 1992 | 32.6 | 14 | 5.7 |  |  |
| 1993 | 102.7 | 25.8 | 1.5 |  |  |
| 1994 | 6.6 | 59.2 | 18 | 1.7 |  |
| 1995 | 0.5 | 7.7 | 8 | 1.1 |  |
| 1996* | 0.1 | 0.25 | 1.8 | 0.6 | 0.03 |
| 1997** | 2.6 | 0.04 | 0.4 | 0.35 | 0.05 |
| 1998 | 9.5 | 4.7 | 0.01 | 0.01 | 0 |
| 1999 | 49.5 | 4.9 | 0 | 0 | 0 |
| 2000 | 105.4 | 27.9 | 0 | 0 | 0 |
| 2001 | 0.3 | 7.6 | 8.8 | 0 | 0 |
| 2002 | 0.5 | 3.9 | 0 | 0 | 0 |
| 2003*** |  |  |  |  |  |
| 2004*** |  |  |  |  |  |
| 2005 | 23.3 | 4.5 | 2.5 | 0.4 | 0.3 |
| 2006 | 3.7 | 35.0 | 5.3 | 0.87 | 0 |
| 2007 | 2.1 | 3.7 | 12.5 | 1.9 | 0 |
| 2008^ |  |  |  |  |  |
| 2009 | 0.289 | 0.300 | 0.233 | 0.060 |  |
| 2010 | 5.196 | 1.380 | 0.000 | 0.000 |  |
| 2011 | 1.166 | 3.920 | 0.041 | 0.000 |  |
| 2012 | 0.787 | 0.030 | 0.000 | 0.000 |  |
| 2013 | 0.107 | 2.190 | 0.211 | 0.070 |  |
| 2014 | 4.239 | 3.110 | 1.728 | 0.127 | 0.043 |
| 2015 | 0.345 | 11.760 | 1.183 | 0.206 | 0.000 |
| 2016 | 1.826 | 5.620 | 1.568 | 0.101 | 0.038 |


| age |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Year | 1 | 2 | 3 | 4 | 0.000 |
| 2017 | 14.522 | 3.080 | 0.000 | 0.009 |  |
| 2018 | 7.329 | 17.420 | 0.827 | 0.044 |  |
| 2019 | 0.113 |  |  |  |  |
| $2020^{* * *}$ | 0.021 |  | 0.086 | 0.002 |  |
| 2021 |  |  |  |  |  |
| $2022^{* * *}$ |  |  |  |  |  |

*Average of Norwegian and Russian estimates
**Combination of Norwegian and Russian estimates as described in 1998 WG report, since then only Russian estimates
***No surveys
${ }^{\wedge}$ Not a full survey

Table 4.4.7.3. Norwegian spring-spawning herring. Estimates from the international acoustic survey on the feeding areas in the Norwegian Sea in May (IESNS). Numbers in millions. Biomass in thousands. Values in the years 2008-2022 are estimated indices by StoX (mean of bootstrap with 1000 iterations). "Fleet 5"

| Year | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Total <br> Biomass |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | Total |  |
| 1996 | 0 | 0 | 4114 | 22461 | 13244 | 4916 | 2045 | 424 | 14 | 7 | 155 | 0 | 3134 |  |  | 50514 | 8532 |
| 1997 | 0 | 0 | 1169 | 3599 | 18867 | 13546 | 2473 | 1771 | 178 | 77 | 288 | 190 | 60 | 2697 |  | 44915 | 9435 |
| 1998 | 24 | 1404 | 367 | 1099 | 4410 | 16378 | 10160 | 2059 | 804 | 183 | 0 | 0 | 35 | 0 | 492 | 37415 | 8004 |
| 1999 | 0 | 215 | 2191 | 322 | 965 | 3067 | 11763 | 6077 | 853 | 258 | 5 | 14 | 0 | 158 | 128 | 26016 | 6299 |
| 2000 | 0 | 157 | 1353 | 2783 | 92 | 384 | 1302 | 7194 | 5344 | 1689 | 271 | 0 | 114 | 0 | 75 | 20758 | 6001 |
| 2001 | 0 | 1540 | 8312 | 1430 | 1463 | 179 | 204 | 3215 | 5433 | 1220 | 94 | 178 | 0 | 0 | 6 | 23274 | 3937 |
| 2002 | 0 | 677 | 6343 | 9619 | 1418 | 779 | 375 | 847 | 1941 | 2500 | 1423 | 61 | 78 | 28 | 0 | 26089 | 4628 |
| 2003 | 32073 | 8115 | 6561 | 9985 | 9961 | 1499 | 732 | 146 | 228 | 1865 | 2359 | 1769 |  | 287 | 0 | 75580 | 6653 |
| 2004 | 0 | 13735 | 1543 | 5227 | 12571 | 10710 | 1075 | 580 | 76 | 313 | 362 | 1294 | 1120 | 10 | 88 | 48704 | 7687 |
| 2005 | 0 | 1293 | 19679 | 1353 | 1765 | 6205 | 5371 | 651 | 388 | 139 | 262 | 526 | 1003 | 364 | 115 | 39114 | 5109 |
| 2006 | 0 | 19 | 306 | 14560 | 1396 | 2011 | 6521 | 6978 | 679 | 713 | 173 | 407 | 921 | 618 | 243 | 35545 | 9100 |
| 2007 | 0 | 411 | 2889 | 5877 | 20292 | 1260 | 1992 | 6780 | 5582 | 647 | 488 | 372 | 403 | 1048 | 1010 | 49051 | 12161 |
| 2008 | 0 | 1213 | 655 | 10997 | 8406 | 14798 | 1543 | 2232 | 4890 | 2790 | 511 | 148 | 172 | 244 | 529 | 49187 | 10655 |
| 2009 | 0 | 137 | 1817 | 2280 | 12118 | 8599 | 9735 | 2054 | 1433 | 2608 | 1375 | 237 | 198 | 112 | 248 | 43057 | 9692 |
| 2010 | 231 | 119 | 572 | 2296 | 1828 | 8395 | 5918 | 5676 | 923 | 888 | 1002 | 550 | 89 | 42 | 62 | 28772 | 6649 |
| 2011 | 0 | 1110 | 921 | 1663 | 3592 | 2605 | 9303 | 4390 | 4257 | 771 | 956 | 732 | 269 | 29 | 33 | 30731 | 7336 |


| Year | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Total <br> Biomass |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | Total |  |
| 2012 | 0 | 396 | 2942 | 410 | 668 | 1736 | 2633 | 4328 | 1884 | 2148 | 297 | 604 | 303 | 139 | 41 | 18540 | 4476 |
| 2013 | 0 | 201 | 718 | 3555 | 425 | 1161 | 1859 | 2905 | 4449 | 2772 | 1865 | 678 | 790 | 222 | 102 | 21722 | 5653 |
| 2014 | 13 | 515 | 1258 | 784 | 2788 | 715 | 1118 | 2634 | 2268 | 2806 | 1118 | 703 | 337 | 72 | 212 | 17350 | 4504 |
| 2015 | 0 | 391 | 432 | 1316 | 1132 | 3535 | 1309 | 1191 | 3156 | 2526 | 4457 | 687 | 816 | 290 | 211 | 21450 | 5851 |
| 2016 | 0 | 75 | 3550 | 1538 | 2229 | 1749 | 2631 | 938 | 1092 | 1806 | 1882 | 2853 | 934 | 436 | 130 | 21851 | 5408 |
| 2017 | 10 | 131 | 948 | 4295 | 1198 | 1543 | 826 | 1414 | 317 | 738 | 1008 | 1741 | 2230 | 507 | 237 | 17159 | 4152 |
| 2018 | 0 | 496 | 1004 | 1968 | 5664 | 970 | 1409 | 569 | 1279 | 354 | 675 | 1564 | 1464 | 1498 | 500 | 19412 | 4987 |
| 2019 | 4 | 157 | 2625 | 680 | 2187 | 4656 | 1158 | 1223 | 952 | 1232 | 823 | 655 | 1406 | 917 | 803 | 19487 | 4805 |
| 2020 | 0 | 43 | 472 | 13065 | 513 | 1009 | 2492 | 786 | 629 | 434 | 694 | 324 | 505 | 726 | 902 | 22616 | 4210 |
| 2021 | 15 | 34 | 1109 | 1290 | 11906 | 698 | 1051 | 2039 | 501 | 551 | 476 | 462 | 442 | 615 | 1515 | 22984 | 5096 |
| 2022 | 0 | 507 | 383 | 1207 | 1286 | 9633 | 1151 | 1640 | 2064 | 577 | 339 | 325 | 293 | 115 | 288 | 19817 | 4427 |

Table 4.4.8.1 Norwegian spring-spawning herring. Relative standard error of estimated catch-at-age used by XSAM.

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 0.366 | 0.190 | 0.260 | 0.092 | 0.361 | 0.497 | 0.422 | 0.312 | 0.369 | 0.544 | 0.380 |
| 1989 | 0.259 | 0.539 | 0.499 | 0.430 | 0.109 | 0.507 | 0.832 | 0.879 | 0.515 | 0.693 | 0.713 |
| 1990 | 0.305 | 0.287 | 0.554 | 0.334 | 0.345 | 0.124 | 0.708 | 0.670 | 0.607 | 0.573 | 0.622 |
| 1991 | 0.514 | 0.375 | 0.546 | 0.687 | 0.311 | 0.369 | 0.125 | 0.566 | 1.002 | 1.760 | 0.659 |
| 1992 | 0.699 | 0.328 | 0.237 | 0.448 | 0.727 | 0.333 | 0.427 | 0.124 | 0.567 | 0.897 | 0.675 |
| 1993 | 0.395 | 0.249 | 0.159 | 0.170 | 0.373 | 0.498 | 0.246 | 0.287 | 0.101 | NA | NA |
| 1994 | 0.381 | 0.238 | 0.158 | 0.105 | 0.138 | 0.305 | 0.380 | 0.227 | 0.231 | 0.087 | 0.435 |
| 1995 | 0.740 | 0.196 | 0.107 | 0.088 | 0.087 | 0.123 | 0.306 | 0.303 | 0.184 | 0.173 | 0.077 |
| 1996 | 0.244 | 0.234 | 0.084 | 0.064 | 0.076 | 0.101 | 0.161 | 0.429 | 0.393 | 0.187 | 0.079 |
| 1997 | 0.272 | 0.149 | 0.116 | 0.061 | 0.059 | 0.082 | 0.109 | 0.193 | 0.280 | 0.238 | 0.096 |
| 1998 | 0.173 | 0.183 | 0.121 | 0.105 | 0.062 | 0.069 | 0.104 | 0.149 | 0.217 | 0.259 | 0.123 |
| 1999 | 0.447 | 0.146 | 0.230 | 0.147 | 0.100 | 0.064 | 0.071 | 0.113 | 0.160 | 0.312 | 0.129 |
| 2000 | 0.313 | 0.173 | 0.091 | 0.232 | 0.157 | 0.102 | 0.068 | 0.074 | 0.125 | 0.182 | 0.147 |
| 2001 | 0.603 | 0.162 | 0.139 | 0.100 | 0.224 | 0.165 | 0.113 | 0.079 | 0.094 | 0.181 | 0.202 |
| 2002 | 0.191 | 0.129 | 0.087 | 0.119 | 0.109 | 0.245 | 0.167 | 0.117 | 0.086 | 0.108 | 0.174 |
| 2003 | 0.463 | 0.179 | 0.109 | 0.083 | 0.135 | 0.137 | 0.268 | 0.180 | 0.125 | 0.091 | 0.116 |
| 2004 | 0.215 | 0.263 | 0.167 | 0.100 | 0.084 | 0.157 | 0.146 | 0.254 | 0.202 | 0.136 | 0.086 |
| 2005 | 0.278 | 0.098 | 0.166 | 0.136 | 0.087 | 0.077 | 0.152 | 0.152 | 0.226 | 0.188 | 0.088 |
| 2006 | 0.213 | 0.179 | 0.083 | 0.174 | 0.136 | 0.083 | 0.082 | 0.170 | 0.178 | 0.244 | 0.103 |
| 2007 | 0.375 | 0.124 | 0.105 | 0.061 | 0.141 | 0.121 | 0.083 | 0.094 | 0.210 | 0.259 | 0.138 |
| 2008 | 0.151 | 0.229 | 0.092 | 0.086 | 0.056 | 0.129 | 0.119 | 0.090 | 0.105 | 0.246 | 0.143 |
| 2009 | 0.156 | 0.130 | 0.142 | 0.070 | 0.077 | 0.060 | 0.144 | 0.118 | 0.100 | 0.121 | 0.147 |
| 2010 | 0.192 | 0.162 | 0.127 | 0.131 | 0.073 | 0.084 | 0.066 | 0.135 | 0.133 | 0.108 | 0.119 |
| 2011 | 0.120 | 0.192 | 0.160 | 0.122 | 0.127 | 0.082 | 0.092 | 0.087 | 0.168 | 0.154 | 0.129 |
| 2012 | 0.324 | 0.126 | 0.206 | 0.153 | 0.115 | 0.117 | 0.083 | 0.111 | 0.106 | 0.202 | 0.151 |
| 2013 | 0.277 | 0.193 | 0.115 | 0.183 | 0.156 | 0.114 | 0.121 | 0.089 | 0.136 | 0.144 | 0.209 |
| 2014 | 0.682 | 0.249 | 0.196 | 0.118 | 0.205 | 0.182 | 0.130 | 0.143 | 0.104 | 0.173 | 0.176 |
| 2015 | 0.518 | 0.301 | 0.198 | 0.203 | 0.141 | 0.234 | 0.188 | 0.140 | 0.161 | 0.129 | 0.173 |


| Year/Age | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2016 | 0.578 | 0.212 | 0.216 | 0.157 | 0.171 | 0.138 | 0.203 | 0.181 | 0.135 | 0.165 | 0.118 |
| 2017 | 0.300 | 0.189 | 0.112 | 0.155 | 0.120 | 0.138 | 0.114 | 0.163 | 0.148 | 0.115 | 0.102 |
| 2018 | 0.270 | 0.258 | 0.194 | 0.117 | 0.142 | 0.134 | 0.158 | 0.133 | 0.171 | 0.163 | 0.094 |
| 2019 | 0.591 | 0.142 | 0.189 | 0.132 | 0.091 | 0.143 | 0.125 | 0.153 | 0.137 | 0.148 | 0.092 |
| 2020 | 0.378 | 0.207 | 0.092 | 0.183 | 0.142 | 0.101 | 0.154 | 0.140 | 0.166 | 0.131 | 0.099 |
| 2021 | 0.576 | 0.260 | 0.157 | 0.066 | 0.138 | 0.128 | 0.107 | 0.154 | 0.150 | 0.181 | 0.115 |

Table 4.4.8.2 Norwegian spring-spawning herring. Relative standard error of Fleet 1 used by XSAM.

| Year/Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 0.316 | 0.334 | 0.161 | 0.449 | 0.549 | 0.687 | 0.538 | 0.600 | 0.512 | NA |
| 1989 | 0.645 | 0.327 | 0.438 | 0.189 | 0.427 | 0.687 | 0.878 | 0.489 | NA | 0.489 |
| 1990 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1991 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1992 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1993 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1994 | 0.428 | 0.505 | 0.266 | 0.302 | 0.477 | 0.675 | 0.395 | 0.502 | 0.224 | 0.752 |
| 1995 | 0.306 | 0.182 | 0.198 | 0.221 | 0.336 | 0.613 | NA | 0.423 | 0.489 | 0.212 |
| 1996 | 0.373 | 0.220 | 0.161 | 0.212 | 0.264 | 0.335 | NA | NA | 0.400 | 0.226 |
| 1997 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1998 | 0.333 | 0.258 | 0.201 | 0.145 | 0.161 | 0.224 | 0.288 | 0.384 | 0.517 | 0.227 |
| 1999 | 0.233 | 0.320 | 0.240 | 0.206 | 0.156 | 0.166 | 0.229 | 0.298 | 0.404 | 0.275 |
| 2000 | 0.278 | 0.207 | 0.420 | 0.302 | 0.249 | 0.185 | 0.203 | 0.300 | 0.500 | 0.353 |
| 2001 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2002 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2003 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2004 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2005 | 0.353 | 0.281 | 0.217 | 0.172 | 0.160 | 0.286 | 0.299 | 0.388 | 0.436 | 0.212 |
| 2006 | 0.439 | 0.172 | 0.293 | 0.264 | 0.196 | 0.196 | 0.419 | 0.443 | 0.613 | 0.305 |
| 2007 | 0.321 | 0.230 | 0.146 | 0.268 | 0.282 | 0.201 | 0.198 | 0.397 | 0.357 | 0.257 |
| 2008 | 0.505 | 0.208 | 0.203 | 0.158 | 0.296 | 0.303 | 0.232 | 0.232 | 0.385 | 0.312 |


| Year/Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2010 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2011 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2012 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2013 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2014 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2015 | 0.295 | 0.205 | 0.274 | 0.197 | 0.317 | 0.278 | 0.205 | 0.271 | 0.165 | 0.213 |
| 2016 | 0.368 | 0.353 | 0.288 | 0.321 | 0.217 | 0.352 | 0.297 | 0.202 | 0.282 | 0.178 |
| 2017 | 0.419 | 0.252 | 0.283 | 0.264 | 0.310 | 0.241 | 0.375 | 0.365 | 0.219 | 0.192 |
| 2018 | 0.394 | 0.227 | 0.204 | 0.305 | 0.266 | 0.283 | 0.251 | 0.307 | 0.320 | 0.195 |
| 2019 | 0.320 | 0.333 | 0.260 | 0.191 | 0.273 | 0.264 | 0.281 | 0.253 | 0.275 | 0.183 |
| 2020 | 0.514 | 0.194 | 0.291 | 0.246 | 0.196 | 0.295 | 0.286 | 0.321 | 0.282 | 0.222 |
| 2021 | 0.418 | 0.337 | 0.153 | 0.275 | 0.274 | 0.221 | 0.310 | 0.306 | 0.332 | 0.220 |
| 2022 | 0.461 | 0.385 | 0.273 | 0.170 | 0.331 | 0.291 | 0.234 | 0.335 | 0.301 | 0.232 |

Table 4.4.8.3 Norwegian spring-spawning herring. Relative standard error of Fleet 4 used by XSAM.

| Year/Age | 2 |
| :---: | :---: |
| 1991 | 0.462 |
| 1992 | 0.419 |
| 1993 | 0.395 |
| 1994 | 0.364 |
| 1995 | 0.444 |
| 1996 | 0.620 |
| 1997 | 0.741 |
| 1998 | 0.466 |
| 1999 | 0.464 |
| 2000 | 0.392 |
| 2001 | 0.445 |
| 2002 | 0.475 |
| 2003 | NA |


| Year/Age | 2 |
| :---: | :---: |
| 2004 | NA |
| 2005 | 0.468 |
| 2006 | 0.383 |
| 2007 | 0.477 |
| 2008 | 0.595 |
| 2009 | 0.609 |
| 2010 | 0.525 |
| 2011 | 0.474 |
| 2012 | 0.763 |
| 2013 | 0.502 |
| 2014 | 0.485 |
| 2015 | 0.426 |
| 2016 | 0.458 |
| 2017 | 0.486 |
| 2018 | 0.410 |
| 2019 | 0.498 |
| 2020 | NA |
| 2021 | 1.006 |
| 2022 | NA |

Table 4.4.8.4 Norwegian spring-spawning herring. Relative standard error of Fleet 5 used by XSAM.

| Year/Age | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1996 | 0.203 | 0.136 | 0.154 | 0.195 | 0.239 | 0.347 | 0.774 | 0.912 | 0.44 | 0.216 |
| 1997 | 0.273 | 0.21 | 0.142 | 0.153 | 0.229 | 0.248 | 0.425 | 0.518 | 0.38 | 0.22 |
| 1998 | 0.359 | 0.277 | 0.2 | 0.147 | 0.164 | 0.239 | 0.298 | 0.423 | $N A$ | 0.329 |
| 1999 | 0.236 | 0.37 | 0.286 | 0.218 | 0.159 | 0.185 | 0.294 | 0.39 | 0.987 | 0.376 |
| 2000 | 0.264 | 0.223 | 0.497 | 0.355 | 0.266 | 0.178 | 0.191 | 0.25 | 0.385 | 0.42 |
| 2001 | 0.172 | 0.26 | 0.259 | 0.425 | 0.412 | 0.215 | 0.19 | 0.27 | 0.495 | 0.422 |
| 2002 | 0.183 | 0.166 | 0.261 | 0.301 | 0.357 | 0.295 | 0.242 | 0.228 | 0.261 | 0.432 |
| 2003 | 0.182 | 0.165 | 0.165 | 0.258 | 0.305 | 0.446 | 0.401 | 0.245 | 0.231 | 0.239 |


| Year/Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | 0.256 | 0.192 | 0.156 | 0.162 | 0.279 | 0.322 | 0.52 | 0.373 | 0.36 | 0.228 |
| 2005 | 0.14 | 0.264 | 0.248 | 0.184 | 0.191 | 0.313 | 0.354 | 0.451 | 0.388 | 0.24 |
| 2006 | 0.375 | 0.151 | 0.262 | 0.24 | 0.182 | 0.179 | 0.31 | 0.307 | 0.428 | 0.236 |
| 2007 | 0.221 | 0.187 | 0.139 | 0.268 | 0.241 | 0.181 | 0.189 | 0.314 | 0.336 | 0.222 |
| 2008 | 0.313 | 0.161 | 0.172 | 0.15 | 0.256 | 0.235 | 0.195 | 0.223 | 0.332 | 0.277 |
| 2009 | 0.246 | 0.233 | 0.157 | 0.171 | 0.166 | 0.239 | 0.26 | 0.226 | 0.263 | 0.299 |
| 2010 | 0.323 | 0.233 | 0.246 | 0.172 | 0.186 | 0.188 | 0.289 | 0.291 | 0.283 | 0.304 |
| 2011 | 0.289 | 0.251 | 0.21 | 0.226 | 0.168 | 0.2 | 0.201 | 0.301 | 0.286 | 0.279 |
| 2012 | 0.22 | 0.35 | 0.312 | 0.249 | 0.226 | 0.201 | 0.244 | 0.237 | 0.377 | 0.278 |
| 2013 | 0.306 | 0.21 | 0.347 | 0.274 | 0.245 | 0.22 | 0.199 | 0.223 | 0.245 | 0.247 |
| 2014 | 0.268 | 0.3 | 0.223 | 0.307 | 0.276 | 0.226 | 0.234 | 0.222 | 0.276 | 0.265 |
| 2015 | 0.345 | 0.266 | 0.275 | 0.21 | 0.266 | 0.272 | 0.216 | 0.228 | 0.199 | 0.241 |
| 2016 | 0.21 | 0.256 | 0.235 | 0.248 | 0.226 | 0.288 | 0.278 | 0.247 | 0.244 | 0.2 |
| 2017 | 0.287 | 0.201 | 0.272 | 0.256 | 0.296 | 0.261 | 0.371 | 0.304 | 0.283 | 0.197 |
| 2018 | 0.283 | 0.242 | 0.188 | 0.285 | 0.261 | 0.324 | 0.267 | 0.362 | 0.311 | 0.194 |
| 2019 | 0.226 | 0.31 | 0.236 | 0.197 | 0.274 | 0.27 | 0.287 | 0.27 | 0.297 | 0.207 |
| 2020 | 0.338 | 0.155 | 0.332 | 0.283 | 0.229 | 0.3 | 0.316 | 0.345 | 0.309 | 0.229 |
| 2021 | 0.277 | 0.267 | 0.158 | 0.308 | 0.28 | 0.24 | 0.333 | 0.326 | 0.337 | 0.218 |
| 2022 | 0.355 | 0.271 | 0.267 | 0.166 | 0.274 | 0.252 | 0.239 | 0.323 | 0.366 | 0.282 |

Table 4.5.1.1. Norwegian spring-spawning herring. Parameter estimates of the final XSAM model fit. The estimates from the final 2021 assessment are also shown.

| Parameter | Estimate | Std. Error | CV | Estimate 2021 | Std. Error 2021 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| $\boldsymbol{\operatorname { l o g } ( N _ { \mathbf { 3 , 1 9 8 8 } } )}$ | 7.088 | 0.164 | 0.023 | 7.087 | 0.167 |
| $\boldsymbol{\operatorname { l o g } ( N _ { \mathbf { 4 , 1 9 8 8 } } )}$ | 6.631 | 0.203 | 0.031 | 6.621 | 0.206 |
| $\boldsymbol{\operatorname { l o g } ( \boldsymbol { N } _ { \mathbf { 5 , 1 9 8 8 } } )}$ | 9.584 | 0.066 | 0.007 | 9.584 | 0.069 |
| $\boldsymbol{\operatorname { l o g } ( \boldsymbol { N } _ { \mathbf { 6 , 1 9 8 8 } } )}$ | 4.837 | 0.380 | 0.079 | 4.825 | 0.381 |
| $\boldsymbol{\operatorname { l o g } ( \boldsymbol { N } _ { \mathbf { 7 , 1 9 8 8 } } )}$ | 3.527 | 0.532 | 0.151 | 3.518 | 0.529 |
| $\boldsymbol{\operatorname { l o g } ( \boldsymbol { N } _ { \mathbf { 8 , 1 9 8 8 } } )}$ | 3.079 | 0.594 | 0.193 | 3.087 | 0.591 |
| $\boldsymbol{\operatorname { l o g } ( \boldsymbol { N } _ { \mathbf { 9 , 1 9 8 8 } } )}$ | 4.073 | 0.455 | 0.112 | 4.076 |  |


| Parameter | Estimate | Std. Error | CV | Estimate 2021 | Std. Error 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\log \left(N_{10,1988}\right)$ | 3.282 | 0.669 | 0.204 | 3.286 | 0.667 |
| $\log \left(N_{11,1988}\right)$ | 3.191 | 0.691 | 0.216 | 3.180 | 0.695 |
| $\log \left(N_{12,1988}\right)$ | 3.585 | 0.753 | 0.210 | 3.578 | 0.753 |
| $\log \left(q_{3}^{F 1}\right)$ | -9.657 | 0.173 | 0.018 | -9.669 | 0.179 |
| $\log \left(q_{4}^{F 1}\right)$ | -8.143 | 0.124 | 0.015 | -8.108 | 0.128 |
| $\log \left(q_{5}^{F 1}\right)$ | -7.487 | 0.111 | 0.015 | -7.474 | 0.115 |
| $\log \left(q_{6}^{F 1}\right)$ | -7.283 | 0.110 | 0.015 | -7.296 | 0.117 |
| $\log \left(q_{7}^{F 1}\right)$ | -7.165 | 0.123 | 0.017 | -7.152 | 0.128 |
| $\log \left(q_{8}^{F 1}\right)$ | -6.926 | 0.086 | 0.012 | -6.939 | 0.091 |
| $\log \left(q_{2}^{F 4}\right)$ | -14.525 | 0.189 | 0.013 | -14.515 | 0.193 |
| $\log \left(q_{3}^{F 5}\right)$ | -7.654 | 0.105 | 0.014 | -7.653 | 0.107 |
| $\log \left(q_{4}^{F 5}\right)$ | -7.133 | 0.093 | 0.013 | -7.123 | 0.095 |
| $\log \left(q_{5}^{F 5}\right)$ | -6.913 | 0.091 | 0.013 | -6.904 | 0.093 |
| $\log \left(q_{6}^{F 5}\right)$ | -6.796 | 0.094 | 0.014 | -6.805 | 0.097 |
| $\log \left(q_{7}^{F 5}\right)$ | -6.721 | 0.101 | 0.015 | -6.734 | 0.103 |
| $\log \left(q_{8}^{F 5}\right)$ | -6.541 | 0.106 | 0.016 | -6.557 | 0.109 |
| $\log \left(q_{9}^{F 5}\right)$ | -6.537 | 0.118 | 0.018 | -6.543 | 0.121 |
| $\log \left(q_{10}^{F 5}\right)$ | -6.474 | 0.132 | 0.020 | -6.490 | 0.135 |
| $\log \left(q_{11}^{F 5}\right)$ | -6.433 | 0.126 | 0.020 | -6.433 | 0.131 |
| $\log \left(\sigma_{1}^{2}\right)$ | -5.000 | 1.409 | 0.282 | -5.000 | 1.441 |
| $\log \left(\sigma_{2}^{2}\right)$ | $-2.777$ | 0.243 | 0.088 | -2.769 | 0.256 |
| $\log \left(\sigma_{4}^{2}\right)$ | -2.281 | 0.299 | 0.131 | -2.250 | 0.303 |
| $\log \left(\sigma_{R}^{2}\right)$ | -0.022 | 0.255 | 11.598 | -0.008 | 0.275 |
| $\log (h)$ | 1.565 | 0.063 | 0.040 | 1.595 | 0.065 |
| $\mu_{R}$ | 9.275 | 0.176 | 0.019 | 9.275 | 0.180 |
| $\alpha_{Y}$ | -0.492 | 0.294 | 0.596 | $-0.513$ | 0.300 |
| $\boldsymbol{\beta}_{Y}$ | 0.816 | 0.107 | 0.131 | 0.810 | 0.108 |
| $\alpha_{2 U}$ | -1.239 | 0.164 | 0.133 | -1.242 | 0.167 |


| Parameter | Estimate | Std. Error | CV | Estimate 2021 | Std. Error 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\boldsymbol{\alpha}_{\boldsymbol{3} \boldsymbol{U}}$ | -0.629 | 0.095 | 0.151 | -0.620 | 0.096 |
| $\boldsymbol{\alpha}_{\boldsymbol{4 U}}$ | -0.215 | 0.059 | 0.273 | -0.214 | 0.060 |
| $\boldsymbol{\alpha}_{\mathbf{5} \boldsymbol{U}}$ | 0.054 | 0.049 | 0.916 | 0.043 | 0.051 |
| $\boldsymbol{\alpha}_{\boldsymbol{6} \boldsymbol{U}}$ | 0.199 | 0.054 | 0.271 | 0.196 | 0.055 |
| $\boldsymbol{\alpha}_{\mathbf{7 U}}$ | 0.263 | 0.058 | 0.222 | 0.264 | 0.060 |
| $\boldsymbol{\alpha}_{\boldsymbol{B} \boldsymbol{U}}$ | 0.319 | 0.065 | 0.203 | 0.327 | 0.066 |
| $\boldsymbol{\alpha}_{\boldsymbol{9} \boldsymbol{U}}$ | 0.368 | 0.070 | 0.191 | 0.368 | 0.072 |
| $\boldsymbol{\alpha}_{\mathbf{1 0 U}}$ | 0.419 | 0.076 | 0.182 | 0.420 | 0.078 |
| $\boldsymbol{\beta}_{\boldsymbol{U}}$ | 0.603 | 0.052 | 0.086 | 0.603 | 0.053 |

Table 4.5.1.2 Norwegian spring-spawning herring. Point estimates of Stock in numbers (millions).

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 672 | 1197 | 759 | 14529 | 126 | 34 | 22 | 59 | 27 | 24 | 36 |
| 1989 | 1173 | 260 | 967 | 635 | 12012 | 104 | 28 | 16 | 40 | 16 | 44 |
| 1990 | 4339 | 471 | 219 | 818 | 531 | 10012 | 86 | 22 | 13 | 30 | 47 |
| 1991 | 11466 | 1758 | 401 | 186 | 687 | 444 | 8370 | 71 | 18 | 10 | 62 |
| 1992 | 18683 | 4656 | 1505 | 341 | 157 | 577 | 372 | 6977 | 58 | 15 | 59 |
| 1993 | 50101 | 7589 | 3992 | 1278 | 286 | 131 | 482 | 310 | 5769 | 47 | 59 |
| 1994 | 59953 | 20347 | 6502 | 3367 | 1041 | 231 | 106 | 389 | 249 | 4571 | 83 |
| 1995 | 15751 | 24339 | 17426 | 5476 | 2632 | 778 | 177 | 83 | 301 | 187 | 3436 |
| 1996 | 5722 | 6386 | 20794 | 14597 | 4174 | 1750 | 507 | 128 | 60 | 207 | 2243 |
| 1997 | 2152 | 2315 | 5420 | 17205 | 11160 | 2798 | 1122 | 333 | 89 | 41 | 1353 |
| 1998 | 10941 | 868 | 1919 | 4366 | 13115 | 7770 | 1742 | 659 | 207 | 54 | 754 |
| 1999 | 6461 | 4417 | 715 | 1480 | 3371 | 9610 | 5447 | 1117 | 410 | 122 | 458 |
| 2000 | 32626 | 2615 | 3680 | 557 | 1130 | 2504 | 6817 | 3652 | 698 | 243 | 299 |
| 2001 | 28927 | 13217 | 2189 | 2739 | 416 | 830 | 1788 | 4661 | 2248 | 406 | 267 |
| 2002 | 11339 | 11726 | 11211 | 1742 | 2003 | 311 | 615 | 1286 | 3236 | 1483 | 446 |
| 2003 | 6678 | 4590 | 9908 | 9050 | 1282 | 1405 | 226 | 432 | 873 | 2155 | 1284 |
| 2004 | 57658 | 2707 | 3888 | 8191 | 7103 | 944 | 1027 | 164 | 303 | 587 | 2247 |
| 2005 | 24428 | 23396 | 2301 | 3240 | 6622 | 5467 | 703 | 745 | 119 | 213 | 1753 |


| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 43044 | 9907 | 19779 | 1895 | 2590 | 5071 | 3862 | 479 | 504 | 78 | 1122 |
| 2007 | 12127 | 17457 | 8425 | 16363 | 1524 | 2027 | 3714 | 2639 | 331 | 349 | 695 |
| 2008 | 17706 | 4911 | 14809 | 6936 | 12530 | 1154 | 1484 | 2526 | 1747 | 222 | 705 |
| 2009 | 7109 | 7141 | 4155 | 12207 | 5361 | 8717 | 815 | 1022 | 1613 | 1102 | 614 |
| 2010 | 5074 | 2851 | 5979 | 3412 | 9431 | 3809 | 5662 | 546 | 635 | 963 | 1055 |
| 2011 | 15315 | 2035 | 2376 | 4915 | 2719 | 7112 | 2651 | 3524 | 341 | 390 | 1095 |
| 2012 | 5658 | 6144 | 1700 | 1950 | 3964 | 2124 | 5363 | 1798 | 2352 | 222 | 939 |
| 2013 | 8319 | 2287 | 5142 | 1403 | 1569 | 3140 | 1623 | 3936 | 1265 | 1644 | 813 |
| 2014 | 5491 | 3370 | 1928 | 4213 | 1130 | 1244 | 2446 | 1211 | 2875 | 913 | 1917 |
| 2015 | 17709 | 2228 | 2868 | 1609 | 3420 | 916 | 997 | 1924 | 927 | 2162 | 2259 |
| 2016 | 7341 | 7190 | 1902 | 2414 | 1330 | 2791 | 746 | 799 | 1521 | 717 | 3525 |
| 2017 | 4432 | 2980 | 6134 | 1595 | 1979 | 1068 | 2229 | 590 | 622 | 1158 | 3288 |
| 2018 | 39850 | 1797 | 2521 | 5028 | 1264 | 1483 | 789 | 1622 | 427 | 428 | 3167 |
| 2019 | 5149 | 16169 | 1527 | 2097 | 4059 | 965 | 1120 | 588 | 1204 | 308 | 2521 |
| 2020 | 4358 | 2088 | 13732 | 1257 | 1649 | 3012 | 706 | 791 | 413 | 847 | 1786 |
| 2021 | 1958 | 1766 | 1767 | 11303 | 990 | 1245 | 2203 | 504 | 545 | 282 | 1693 |
| 2022 | 10671 | 793 | 1491 | 1431 | 8497 | 716 | 882 | 1547 | 331 | 353 | 1346 |

Table 4.5.1.3 Norwegian spring-spawning herring. Point estimates of Fishing mortality.

| Year/Age | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1988 | 0.050 | 0.064 | 0.029 | 0.040 | 0.042 | 0.042 | 0.142 | 0.224 | 0.334 | 0.170 | 0.170 |
| 1989 | 0.012 | 0.021 | 0.018 | 0.028 | 0.032 | 0.035 | 0.077 | 0.110 | 0.152 | 0.092 | 0.092 |
| 1990 | 0.004 | 0.012 | 0.015 | 0.024 | 0.030 | 0.029 | 0.052 | 0.073 | 0.098 | 0.071 | 0.071 |
| 1991 | 0.001 | 0.005 | 0.011 | 0.019 | 0.025 | 0.025 | 0.032 | 0.044 | 0.057 | 0.050 | 0.050 |
| 1992 | 0.001 | 0.004 | 0.013 | 0.025 | 0.031 | 0.031 | 0.035 | 0.040 | 0.055 | 0.058 | 0.058 |
| 1993 | 0.001 | 0.005 | 0.020 | 0.055 | 0.063 | 0.059 | 0.063 | 0.068 | 0.083 | 0.105 | 0.105 |
| 1994 | 0.002 | 0.005 | 0.022 | 0.096 | 0.142 | 0.116 | 0.099 | 0.107 | 0.135 | 0.153 | 0.153 |
| 1995 | 0.003 | 0.007 | 0.027 | 0.121 | 0.258 | 0.278 | 0.175 | 0.171 | 0.223 | 0.330 | 0.330 |
| 1996 | 0.005 | 0.014 | 0.039 | 0.118 | 0.250 | 0.295 | 0.271 | 0.213 | 0.244 | 0.444 | 0.444 |
| 1997 | 0.008 | 0.038 | 0.066 | 0.121 | 0.212 | 0.324 | 0.382 | 0.325 | 0.352 | 0.464 | 0.464 |


| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 0.007 | 0.044 | 0.109 | 0.109 | 0.161 | 0.205 | 0.294 | 0.326 | 0.378 | 0.417 | 0.417 |
| 1999 | 0.004 | 0.032 | 0.100 | 0.120 | 0.147 | 0.193 | 0.250 | 0.321 | 0.373 | 0.514 | 0.514 |
| 2000 | 0.004 | 0.028 | 0.145 | 0.141 | 0.158 | 0.187 | 0.230 | 0.335 | 0.391 | 0.557 | 0.557 |
| 2001 | 0.003 | 0.015 | 0.078 | 0.163 | 0.141 | 0.150 | 0.180 | 0.215 | 0.266 | 0.262 | 0.262 |
| 2002 | 0.004 | 0.018 | 0.064 | 0.157 | 0.205 | 0.171 | 0.204 | 0.237 | 0.256 | 0.257 | 0.257 |
| 2003 | 0.003 | 0.016 | 0.040 | 0.092 | 0.156 | 0.163 | 0.170 | 0.203 | 0.248 | 0.276 | 0.276 |
| 2004 | 0.002 | 0.013 | 0.032 | 0.063 | 0.112 | 0.145 | 0.172 | 0.174 | 0.204 | 0.330 | 0.330 |
| 2005 | 0.002 | 0.018 | 0.044 | 0.074 | 0.117 | 0.197 | 0.234 | 0.240 | 0.266 | 0.411 | 0.411 |
| 2006 | 0.002 | 0.012 | 0.040 | 0.068 | 0.095 | 0.161 | 0.231 | 0.220 | 0.219 | 0.396 | 0.396 |
| 2007 | 0.004 | 0.015 | 0.045 | 0.117 | 0.128 | 0.162 | 0.236 | 0.262 | 0.248 | 0.242 | 0.242 |
| 2008 | 0.008 | 0.017 | 0.043 | 0.108 | 0.213 | 0.199 | 0.224 | 0.298 | 0.311 | 0.262 | 0.262 |
| 2009 | 0.014 | 0.028 | 0.047 | 0.108 | 0.192 | 0.281 | 0.250 | 0.326 | 0.366 | 0.336 | 0.336 |
| 2010 | 0.013 | 0.032 | 0.046 | 0.077 | 0.132 | 0.213 | 0.324 | 0.321 | 0.337 | 0.462 | 0.462 |
| 2011 | 0.013 | 0.030 | 0.048 | 0.065 | 0.097 | 0.132 | 0.238 | 0.254 | 0.281 | 0.308 | 0.308 |
| 2012 | 0.006 | 0.028 | 0.043 | 0.067 | 0.083 | 0.119 | 0.159 | 0.202 | 0.208 | 0.205 | 0.205 |
| 2013 | 0.004 | 0.021 | 0.049 | 0.066 | 0.082 | 0.100 | 0.143 | 0.164 | 0.177 | 0.098 | 0.098 |
| 2014 | 0.002 | 0.011 | 0.031 | 0.059 | 0.060 | 0.071 | 0.090 | 0.117 | 0.135 | 0.075 | 0.075 |
| 2015 | 0.001 | 0.008 | 0.022 | 0.040 | 0.053 | 0.055 | 0.072 | 0.085 | 0.107 | 0.077 | 0.077 |
| 2016 | 0.002 | 0.009 | 0.026 | 0.049 | 0.069 | 0.075 | 0.084 | 0.100 | 0.123 | 0.105 | 0.105 |
| 2017 | 0.003 | 0.017 | 0.049 | 0.082 | 0.138 | 0.152 | 0.168 | 0.173 | 0.225 | 0.189 | 0.189 |
| 2018 | 0.002 | 0.013 | 0.034 | 0.064 | 0.120 | 0.130 | 0.145 | 0.148 | 0.175 | 0.205 | 0.205 |
| 2019 | 0.003 | 0.013 | 0.044 | 0.090 | 0.148 | 0.163 | 0.198 | 0.203 | 0.202 | 0.310 | 0.310 |
| 2020 | 0.003 | 0.017 | 0.045 | 0.089 | 0.131 | 0.163 | 0.187 | 0.222 | 0.231 | 0.292 | 0.292 |
| 2021 | 0.004 | 0.019 | 0.061 | 0.135 | 0.174 | 0.195 | 0.203 | 0.269 | 0.284 | 0.233 | 0.233 |
| 2022 | 0.004 | 0.018 | 0.054 | 0.113 | 0.152 | 0.175 | 0.191 | 0.236 | 0.257 | 0.269 | 0.269 |

Table 4.5.1.4 Norwegian spring spawning herring. Final stock summary table. High and low represent approximate $95 \%$ confidence limits.

| Year | Recruitment (Age 2) | High | Low | Stock Size: SSB | High | Low | Catches | Fishing Pressure: $F$ | High | Low |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | millions |  |  | thousnd tonnes |  |  | thousand tonnes | Ages 5-12 |  |  |
| 1988 | 672 | 996 | 349 | 2126 | 2389 | 1863 | 135.301 | 0.042 | 0.058 | 0.026 |
| 1989 | 1173 | 1653 | 693 | 3287 | 3694 | 2881 | 103.830 | 0.033 | 0.046 | 0.019 |
| 1990 | 4339 | 5358 | 3320 | 3562 | 3993 | 3131 | 86.411 | 0.029 | 0.042 | 0.017 |
| 1991 | 11466 | 13340 | 9592 | 3340 | 3743 | 2937 | 84.683 | 0.031 | 0.044 | 0.018 |
| 1992 | 18683 | 21308 | 16058 | 3368 | 3753 | 2982 | 104.448 | 0.039 | 0.054 | 0.023 |
| 1993 | 50101 | 55342 | 44859 | 3340 | 3687 | 2993 | 232.457 | 0.076 | 0.100 | 0.053 |
| 1994 | 59953 | 65811 | 54096 | 3471 | 3817 | 3125 | 479.228 | 0.129 | 0.160 | 0.099 |
| 1995 | 15751 | 18082 | 13419 | 3536 | 3868 | 3205 | 905.501 | 0.219 | 0.259 | 0.179 |
| 1996 | 5722 | 6843 | 4601 | 4118 | 4450 | 3787 | 1220.283 | 0.192 | 0.223 | 0.162 |
| 1997 | 2152 | 2716 | 1587 | 5374 | 5765 | 4984 | 1426.507 | 0.193 | 0.220 | 0.166 |
| 1998 | 10941 | 12719 | 9163 | 5954 | 6383 | 5526 | 1223.131 | 0.186 | 0.214 | 0.158 |
| 1999 | 6461 | 7677 | 5246 | 5854 | 6304 | 5403 | 1235.433 | 0.214 | 0.247 | 0.180 |
| 2000 | 32626 | 36501 | 28751 | 4873 | 5287 | 4458 | 1207.201 | 0.257 | 0.300 | 0.215 |
| 2001 | 28927 | 32527 | 25328 | 4043 | 4416 | 3669 | 766.136 | 0.204 | 0.241 | 0.167 |
| 2002 | 11339 | 13212 | 9465 | 3565 | 3913 | 3218 | 807.795 | 0.224 | 0.265 | 0.183 |
| 2003 | 6678 | 7956 | 5399 | 4189 | 4571 | 3806 | 789.510 | 0.153 | 0.181 | 0.125 |
| 2004 | 57658 | 63732 | 51584 | 5269 | 5734 | 4805 | 794.066 | 0.129 | 0.152 | 0.105 |
| 2005 | 24428 | 27784 | 21072 | 5389 | 5880 | 4898 | 1003.243 | 0.174 | 0.204 | 0.143 |
| 2006 | 43044 | 48247 | 37840 | 5350 | 5832 | 4868 | 968.958 | 0.178 | 0.211 | 0.145 |
| 2007 | 12127 | 14277 | 9976 | 6882 | 7471 | 6294 | 1266.993 | 0.157 | 0.184 | 0.130 |
| 2008 | 17706 | 20549 | 14863 | 6965 | 7584 | 6346 | 1545.656 | 0.202 | 0.235 | 0.169 |
| 2009 | 7109 | 8536 | 5681 | 6937 | 7588 | 6285 | 1687.373 | 0.207 | 0.239 | 0.174 |
| 2010 | 5074 | 6171 | 3977 | 6154 | 6775 | 5533 | 1457.014 | 0.215 | 0.251 | 0.178 |
| 2011 | 15315 | 17846 | 12785 | 5824 | 6450 | 5198 | 992.998 | 0.158 | 0.188 | 0.129 |
| 2012 | 5658 | 6812 | 4504 | 5673 | 6312 | 5034 | 825.999 | 0.142 | 0.169 | 0.115 |
| 2013 | 8319 | 9879 | 6760 | 5307 | 5926 | 4687 | 684.743 | 0.122 | 0.147 | 0.097 |


| Year | Recruitment <br> (Age 2) | High Low | Stock Size: <br> SSB | High | Low | Catches | Fishing Pres- <br> sure: F | High | Low |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | millons |  |  |  |  |  | thousnd <br> tonnes |  | Ages 5-12 |

## Table 4.8.1.1 Norwegian Spring-spawning herring. Input to short-term prediction. Stock size is in millions and weight in kg.

| Input for | 2022 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Stockno. | Natural | Maturity | Proportion of M | Proportion of F | Weight | Exploitation | Weight |
| age | 1-Jan. | mortality | ogive | before spawning | before spawning | in stock | pattern | in catch |
| 2 | 10671 | 0.90 | 0.0 | 0 | 0 | 0.054 | 0.004 | 0.162 |
| 3 | 793 | 0.15 | 0.0 | 0 | 0 | 0.104 | 0.020 | 0.202 |
| 4 | 1491 | 0.15 | 0.4 | 0 | 0 | 0.125 | 0.059 | 0.233 |
| 5 | 1431 | 0.15 | 0.8 | 0 | 0 | 0.168 | 0.124 | 0.266 |
| 6 | 8497 | 0.15 | 0.9 | 0 | 0 | 0.243 | 0.167 | 0.294 |
| 7 | 716 | 0.15 | 1.0 | 0 | 0 | 0.287 | 0.191 | 0.321 |
| 8 | 882 | 0.15 | 1.0 | 0 | 0 | 0.303 | 0.209 | 0.338 |
| 9 | 1547 | 0.15 | 1.0 | 0 | 0 | 0.323 | 0.258 | 0.365 |
| 10 | 331 | 0.15 | 1.0 | 0 | 0 | 0.352 | 0.281 | 0.375 |
| 11 | 353 | 0.15 | 1.0 | 0 | 0 | 0.366 | 0.294 | 0.394 |
| 12 | 1346 | 0.15 | 1.0 | 0 | 0 | 0.389 | 0.294 | 0.418 |
| 2 | 10671 | 0.90 | 0.0 | 0 | 0 | 0.054 | 0.014 | 0.162 |
| 3 |  | 0.15 | 0.0 | 0 | 0 | 0.104 | 0.066 | 0.202 |
| 4 |  | 0.15 | 0.4 | 0 | 0 | 0.145 | 0.192 | 0.233 |
| 5 |  | 0.15 | 0.8 | 0 | 0 | 0.193 | 0.395 | 0.266 |


| Input for | 2022 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Stockno. | Natural | Maturity | Proportion of M | Proportion of F | Weight | Exploitation | Weight |
| age | 1-Jan. | mortality | ogive | before spawning | before spawning | in stock | pattern | in catch |
| 6 |  | 0.15 | 1.0 | 0 | 0 | 0.258 | 0.547 | 0.294 |
| 7 |  | 0.15 | 1.0 | 0 | 0 | 0.291 | 0.633 | 0.321 |
| 8 |  | 0.15 | 1.0 | 0 | 0 | 0.311 | 0.705 | 0.338 |
| 9 |  | 0.15 | 1.0 | 0 | 0 | 0.330 | 0.842 | 0.365 |
| 10 |  | 0.15 | 1.0 | 0 | 0 | 0.354 | 0.933 | 0.375 |
| 11 |  | 0.15 | 1.0 | 0 | 0 | 0.366 | 1.000 | 0.394 |
| 12 |  | 0.15 | 1.0 | 0 | 0 | 0.386 | 1.000 | 0.418 |

Table 4.8.2.1 Norwegian spring spawning herring. Short-term prediction.

| Basis: |  |
| :--- | :--- |
| SSB (2022): | 3.867 million $t$ |
| Landings(2022): | 827963 t (sum of national quotas) |
| SSB(2023): | 3.532 million $t$ |
| Fw5-12+(2022) | 0.192 |
| Recruitment(2022-2024): | $10.671,10.671,10.671$ |

$\qquad$
The catch options:

| Rationale | Catches <br> (2023) | Basis | FW |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| (2023) |  |  |  |

*95\% confidence interval

### 4.17 Figures



Figure 4.2.1.1. Total reported landings (ICES estimates) of Norwegian spring-spawning herring in 2021 by ICES rectangle. Landings below 10 tonnes per statistical rectangle are not included. Catch data by ICES rectangle from Russia are not available. The landings with information on statistical rectangle constitute $89 \%$ of the reported landings.


Figure 4.2.1.2. Total reported landings (ICES estimates) of Norwegian spring-spawning herring in 2021 by quarter and ICES rectangle. Landings below 10 tonnes per statistical rectangle are not included. Catch data by ICES rectangle from Russia are not available. The landings with information on statistical rectangle constitute $89 \%$ of the reported landings.


Figure 4.4.1.1.1. Proportion of Russian catches by quarter for the years 2018-2020.


Figure 4.4.1.1.2. Russian and international catch per ICES rectangle in quarter 3 for the years 2018-2020. Lines in the map are limits for ICES sub-divisions.

## Russian

## Total

international


Figure 4.4.1.1.3. Russian and international catch per ICES rectangle in quarter 4 for the years 2018-2020. Lines in the map are limits for ICES sub-divisions.


Figure 4.4.1.1.4. Russian proportion of the total international catch for the period 2001-2020.


Figure 4.4.1.1.5. Russian landings and ICES advice for the period 2001-2020.


Figure 4.4.3.1. Norwegian spring spawning herring. Age disaggregated landings in numbers plotted on a log scale. Age is on x -axis. The labels indicate year classes and grey lines correspond to $\mathrm{Z}=\mathbf{0 . 3}$.


Figure 4.4.4.1. Norwegian spring spawning herring. Mean weight at age by age groups 3-14 in the years 1981-2021 in the landings.


Figure 4.4.4.2. Norwegian spring-spawning herring. Mean weight at age in the stock by age groups 3-14 for the years 1981-2022.


Figure 4.4.5.1. Assumed (blue line) and back-calculated (orange line) maturity-at-age for the year 2017.


Figure 4.4.7.1. Distribution of Norwegian spring-spawning herring as measured during the IESNS survey in April-June 2022 in terms of NASC values $\left(\mathrm{m}^{2} / \mathrm{nm}^{2}\right)$.


Figure 4.4.7.2. Norwegian acoustic survey on the NSSH spawning grounds. Distribution and acoustic density of herring recorded in 2022.


Figure 4.4.7.3. Norwegian spring spawning herring. Age disaggregated abundance indices (millions) from the acoustic survey on the spawning area in February-March (Fleet 1) plotted on a log scale. The labels indicate year classes and grey lines correspond to $Z=0.3$. Age is on $x$-axis.


Figure 4.4.7.4. Norwegian spring spawning herring. Age disaggregated abundance indices (millions) from the acoustic survey in the feeding area in the Norwegian Sea in May (Fleet 5) plotted on a log scale. The labels indicate year classes and grey lines correspond to $\mathrm{Z}=\mathbf{0 . 3}$.


Figure 4.5.1.1. Estimated exploitation pattern for the years 1988-2022 by the XSAM model fit. All panels show the same data, but depicted at different angles to improve visibility at different time periods


Figure 4.5.1.2. Norwegian spring spawning herring. Correlation between estimated parameters in the final XSAM model fit.

Catch at age


Fleet4


Fleet1



Figure 4.5.1.3. Norwegian spring spawning herring. Weights (inverse of variance) of data-input of the final XSAM model fit.


Fig-
ure 4.5.1.4. Norwegian spring spawning herring. Standardized residuals type 1 (left) and type 2 (right) (see text) of datainput of the final XSAM model fit. Red is negative and blue is positive residuals.







Figure 4.5.1.5. Norwegian spring spawning herring. Observed vs. predicted values (left column) and qq-plot based on type 1 (middle) and type 2 (right) residuals (see text) based on the final XSAM model fit.


Figure 4.5.1.6. Norwegian spring spawning herring. Profiles of marginal $\log$-likelihood $l_{M}$, the catch component $l_{C}$, Fleet 1 component $l_{F 1}$, Fleet 4 component $l_{\text {F4 }}$, Fleet 5 component $l_{F 5}$, point estimate of SSB and average $F$ (ages 5-12+) in 2022 over the common scaling factor for variance in data $h$ for the final XSAM fit. The red dots indicate the value of the respective scaling factors for which the log-likelihood is maximized.


Figure 4.5.1.7. Norwegian spring spawning herring. Retrospective XSAM model fits of SSB and weighted average of fishing mortality ages 5-12 for the years 2017-2022. Mohn's rho computed to be -0.04 for SSB and -0.12 for F.


Figure 4.5.1.8. Norwegian spring spawning herring. Point estimates of Spawning-stock biomass by years 1988-2022 from model (black lines) and by survey indices from Fleet 1 (blue) and Fleet 5 (red). Shaded area is approximate to standard deviation.


Figure 4.5.1.9. Total reported landings 1988-2021, estimated recruitment, weighted average of fishing mortality (ages 512 ) and spawning-stock biomass for the years 1988-2022 based on the final XSAM model fit.


Fig-
ure 4.5.1.10. Norwegian spring-spawning herring. A visual representation of parameter estimates of the final XSAM model fit (see table 4.5.1.1). The estimates from the 2021 assessment are also shown (red).


Figure 4.5.1.11. Norwegian spring-spawning herring. Alternative runs showing the effect of leaving one fleet out. The F is shown to the left and SSB to the right. The base run is shown as purple, leaving out Fleet 1 is red, leaving out Fleet 4 is green and leaving out Fleet 5 is shown as blue. Shaded regions show the standard deviation.


Figure 4.5.2.1.1. Norwegian spring-spawning herring. Residual sum of squares in the surveys separately from TASACS. First row starts with survey 1 and the last one in row four is larval survey.


Figure 4.5.2.1.2. Comparison of SSB time-series from the final assessment from XSAM (blue) and exploratory runs from TASACS (green) following the 2008 benchmark procedure) and SAM (red) with XSAM configurations.


Figure 4.8.1.1. XSAM estimated selection pattern; selected years (estimates for 2016-2021 and predictions for 20222023) are shown in colours as indicated in the legend.


Spawning Stock Biomass

Recruitment at age 2


Figure 4.9.1. Norwegian spring spawning herring. Comparisons of spawning stock; weighted fishing mortality F(5-11/512+); and recruitment at age 2 with previous assessments. In 2016 the proportion mature in the years 2006-2011 was changed; recruitment age changed from 0 to 2 and fishing mortality is calculated over ages 5 to 11 . In 2018 (WKNSSHREF) the age range for the fishing mortality changed to ages 5 to $12+$. The vertical dotted lines indicate the benchmark years 2008 and 2016.

## 7 Horse Mackerel in the Northeast Atlantic

### 7.1 Fisheries in 2021

The total international catches of horse mackerel in the North East Atlantic are shown in Table 5.1.1. Since 2011, the southern horse mackerel stock is assessed by ICES WGHANSA. The total catch from all areas in 2021 for the Western and North Sea stocks was $92,639 t$ which is $3,630 t$ more than in 2020 and the 3rd lowest in the time series.

France, Germany and the Netherlands have a directed trawl fishery and Norway and France a directed purse-seine fishery for horse mackerel. Spain has directed as well as mixed trawl and purse-seine fisheries targeting horse mackerel. In earlier years, most of the catches were used for meal and oil while in later years most of the catches have been used for human consumption.
The quarterly catches of North Sea and Western horse mackerel by Division and Subdivision in 2021 are given in Table 5.1.2 and the distributions of the fisheries are given in Figures 5.1.1.a-d. Note that the figures also include catches of southern horse mackerel. The maps are based on data provided by Belgium, Denmark, France, Germany, Ireland, Netherlands, Norway, Poland, Portugal, Spain, Sweden and the UK and represent $99 \%$ of the total catches. The distribution of the fishery is similar to recent years with the highest catches taken in the $1^{\text {st }}$ and $4^{\text {th }}$ quarter. The historic catch by rectangle and year is also shown in section 1.10 of this report.

The Dutch, Danish, Irish and German fleets operated mainly in the North and West of Ireland and the Western waters off Scotland. The French fleet were in the Bay of Biscay and West Scotland whereas the Norwegian fleet fished in the North-eastern part of the North Sea. The Spanish fleet operated mainly in waters of Cantabrian Sea and Bay of Biscay.

First quarter: The fishing season with most of the catches $43,160 t(46 \%$ of the total catch of the combined Western and North Sea horse mackerel catch). The fishery was mainly carried out west of Scotland and West and North of Ireland and along the Spanish coast (Figure 5.1.1.a).

Second quarter: 8,046 t. As usual, catches were significantly lower than in the first quarter as the second quarter is the main spawning period. Most of the catch were taken West of Ireland and along the Spanish coast. (Figure 5.1.1.b)

Third quarter: $13,517 \mathrm{t}$. Most of the catch were taken in Spanish waters, West of Ireland, in the Channel area and at the Norwegian coast (Figure 5.1.1.c).

Fourth quarter: Catches were $27,073 \mathrm{t}$ ( $29 \%$ of the total catch). The catches were distributed in five main areas (Figure 5.1.1.d):

- Spanish waters,
- Western and Northern Irish waters and West of Scotland
- Norwegian coast
- Eastern part of the Channel
- Along the shelf edge of the Celtic Sea


### 7.2 Stock Units

For many years the Working Group has considered the horse mackerel in the Northeast Atlantic as consisting of three separate stocks: the North Sea, the Southern and the Western stocks (ICES 1990, ICES 1991). For further information, see the Western Horse Mackerel Stock Annex and the

WD document on horse mackerel stock structure (WD Brunel et al., 2016). The boundaries for the different stocks are given in Figure 5.2.1.

### 7.3 WG Catch Estimates

In 2017, a review of catch statistics for North Sea and Western horse mackerel stocks was carried out. The results of this report have been reported in previous Working Groups (Costas, 2017a).

As a result of this review, catches and catch-at-age of reported historical data of both North Sea and Western stocks of horse mackerel were updated (Figures 5.3.1 and 5.3.2). Catch statistics were reviewed since 1990 onward for the Western stock and since 2000 onward for the North Sea stock. The main mismatches between the catch statistics in working group reports and these reviewed data were due to several reasons such as late availability of some data for the report or the availability of official catch data only.

### 7.4 Allocation of Catches to Stocks

The distribution areas for the three stocks are given in the Stock Annex for the Western Horse Mackerel. The catches in 2021 were allocated to the three stocks as follows:

Western stock: $3^{\text {rd }}$ and $4^{\text {th }}$ quarters: Divisions 3.a and 4.a. Quarters 1-4: 2.a, 5.b, 6.a, 7.a-c, e-k and 8.a-e.

North Sea stock: $1^{\text {st }}$ and $2^{\text {nd }}$ quarters: Divisions 3.a and 4.a Quarters 1-4: Divisions 4.b, 4.c and 7.d.

Southern stock: Division 9.a. All catches from these areas were allocated to the southern stock. This stock is now dealt with by another working group (ICES WGHANSA).
The catches by stock are given in Table 5.4.1 and Figure 5.4.1. The catches by ICES sub-area and division for the Western and North Sea stocks for period 1982-2021 are shown in Figures 5.4.2-3. The catches by stock and countries for the period 1997-2021 are given in Table 5.4.2-5.4.3.
Recent genetic investigations show that the current boundaries might need to be newly evaluated in future (see section 1.4.8.3).

### 7.5 Estimates of discards

Only the Netherlands have provided data on discards over an extended period with occasional estimates from Germany and Spain. Since 2017 however, additional countries have provided estimates of discards with 6 countries reporting in 2021. Following the introduction of the European landing obligation for the pelagic fisheries targeting horse mackerel in large areas of the overall fishing area and for Norwegian waters there is a general discard ban in place and discards in recent years have decreased substantially. The discard rate is estimated to be $1.8 \%$ in weight for the combined horse mackerel stocks. The discard rate for the North Sea stock is estimated to be $0.5 \%$ and for the Western stock $2.0 \%$ in 2021.

### 7.6 Trachurus Species Mixing

Three species of genus Trachurus: T. trachurus, T. mediterraneus and T. picturatus are found together and are commercially exploited in NE Atlantic waters. Following the Working Group recommendation (ICES 2002/ACFM: 06) special care was taken to ensure that catch and length
distributions and numbers-at-age of T. trachurus supplied to the Working Group did not include T. mediterraneus and/or T. picturatus.

The T. mediterraneus fishery mainly takes place in the eastern part of ICES division 8.c. There is no clear trend in T. mediterraneus catches in this area although the most recent catch is the second lowest in the time series (Table 5.6.1). Information on the T. picturatus fishery is available in the WGHANSA Report (Working Group on Horse Mackerel, Anchovy and Sardine).

Taking into account that the WGWIDE horse mackerel assessments are only made for T. trachurus, the Working Group recommends that the TACs and any other management regulations which might be established in the future should be related only to T. trachurus and not to Trachurus spp. More information is needed about the Trachurus spp. before the fishery and the stock can be evaluated.

### 7.7 Length Distribution by Fleet and by Country:

Germany, Ireland, the Netherlands, France, UK (England and Scotland) and Spain provided length distributions for their catches in 2021. The length distributions cover approximately $79 \%$ of the total landings of the Western and North Sea horse mackerel catches and are shown in Figure 5.7.1.

### 7.8 Comparing trends between areas and stocks

Horse mackerel (Trachurus trachurus) in the northeast Atlantic is assumed to consist of three separate stocks:

- North Sea (4a part of the year, $4 \mathrm{~b}, 4 \mathrm{c}$ and 7 d )
- Western (4a part of the year, 5b, 6a, 7a-c,e-k, 8a-d)
- Southern (9a)

Catches between 2000 and 2021 are shown in figure 5.4.1 and indicate an overall decline in the catches of horse mackerel since 2009.

A detailed analysis on the development of the catch by age data was presented to the 2017 working group (Pastoors, 2017). In this analysis it was indicated that there is an increase in the catches of juveniles in the Western and North Sea stocks in recent years. This could be an indication of a stronger recruitment of horse mackerel which has been reported by surveys and fishermen. However, it is also an alarming signal if a larger proportion of the catch consists of juveniles. These catches could be seen mostly in division 7.d and to a lesser extent, 7.e.

### 7.9 Quality and Adequacy of fishery and sampling data

Table 5.9 .1 shows a summary of the overall sampling intensity on horse mackerel catches in recent years based on the InterCatch input. Since 2011 the Southern horse mackerel is dealt with by ICES WGHANSA.
Countries that historically sample are Ireland, the Netherlands, Germany, Norway and Spain, covering $42-100 \%$ of their respective catches. In 2020, due to the Covid pandemic, sampling activities in some countries were hampered which lead to an overall lower sampling coverage for 2020. However, due to the fact that for the first time it was possible to upload age samples taken from English vessels in the Netherlands for North Sea horse mackerel the proportion of sampling increased in comparison to last year for this stock. Overall, sampling improved again in 2021.

Table 5.9.2 shows the sampling intensity for the Western stock in 2021 and table 5.9.3 shows the sampling intensity for the North Sea stock in 2021 by country.

In 2021, France, Germany, Ireland, the Netherlands, UK (England), UK Scotland, and Spain provided samples and length distributions and Germany, Ireland, the Netherlands, Norway, UK (England), and Spain provided also age distributions. However, the lack of age and length distribution data for relatively large portions of the horse mackerel catches continues to have a serious effect on the accuracy and reliability of the assessment and the Working Group remain especially concerned about the low number of fish which are aged.

An analysis of the sampling intensity was carried out for the period 2000-2019 for both the North Sea and the Western stock. Sampling intensity in fisheries can be defined as the ratio of sampled catch to the total catch. The precision and accuracy of sampled catch are of considerable importance to obtain a reliable estimate of the commercial catch. Sampled catch is used to extrapolate to total catch in order to obtain a catch-at-age (or at-length) and weight-at-age which are often used as inputs for the stock assessment models. In addition, in the case of horse mackerel the impact of temporal (quarter) and spatial (area by ICES division) factors have to be taken in account in order to obtain a reliable estimate of the commercial catches.

Figure 5.9 .1 shows the proportion of sampled catches by division for the North Sea stock. In general, all ICES divisions show low levels of sampling, especially in recent years. The sampling intensity in relation to the length composition of catch was $>60 \%$. In relation to age composition sampling level are dramatically lower in recent years (Figure 5.9.2) but due to the inclusion of samples of English vessels sampled in the Netherlands this situation improved substantially in the last two years. However, divisions that are usually not sampled can affect the precision and accuracy of total catch-at-age and weight-at-age. For the North Sea stock, samples were only available for area 4.c for all quarters and 7.d from the $3^{\text {rd }}$ and $4^{\text {th }}$ quarters. Therefore, these estimates can be biased, especially, since samples are usually less than the recommended 100 fish per sample. (Table 5.9.1)

The proportion of the sampled catches by region for the Western stock are shown in figure 5.9.3. The general index of sampling intensity increased in 2021 to $78 \%$ in comparison to $51 \%$ last year which was mainly due to the impact of the pandemic on sampling. Divisions (regions) that are not sampled can affect the precision and accuracy of total catch-at-age and weight at age (Figure 9.5.4).

Length distributions were supplied by a number of countries. However, as some countries only deliver catch-at-age distributions and others only length distributions of the catch, the obtained catch-at-age and length distributions do not reflect the total catch especially in case of North Sea horse mackerel. Furthermore, some of the length distributions are only taken from discards of non-horse mackerel targeting fleets and omit the horse mackerel target fleet. This lack of coverage may also affect the accuracy and reliability of the assessment and is a matter of concern for the Working Group.

### 7.10 References

Brunel, T., 2016. Revision of the Maturity Ogive for the Western Spawning Component of NEA Mackerel. Working document to WKWIDE, 6pp.

Costas, G. 2017a. Review of Horse Mackerel catch data. North Sea and Western Stocks. WD to WGWIDE 2017. 11 pp.

Costas, G. 2017b. Sampling coverage for Horse Mackerel Stocks. Presentation to WGWIDE 2017.
ICES, 1990. Report of the Working Group on the Assessment of the Stocks of Sardine, Horse Mackerel and Anchovy. ICES, C.M. 1990/Assess: 24.

ICES, 1991. Working group on the Assessment of the Stocks of Sardine, Horse Mackerel, and Anchovy. ICES CM 1991/Assess: 22. 138 pp.

Pastoors, M. (2017). A look at all the horse mackerel. WD to WGWIDE 2017.

### 7.11 Tables

Table 5.1.1 HORSE MACKEREL general. Catches ( t ) by Sub-area. Data as submitted by Working Group members. Data of limited discard information are only available for some years.

| Sub-area | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 2 | - | + | - | 412 | 23 | 79 | 214 |
| $4+3 . a$ | 1,412 | 2,151 | 7,245 | 2,788 | 4,420 | 25,987 | 24,238 | 20,746 |
| 6 | 7,791 | 8,724 | 11,134 | 6,283 | 24,881 | 31,716 | 33,025 | 20,455 |
| 7 | 43,525 | 45,697 | 34,749 | 33,478 | 40,526 | 42,952 | 39,034 | 77,628 |
| 8 | 47,155 | 37,495 | 40,073 | 22,683 | 28,223 | 25,629 | 27,740 | 43,405 |
| 9 | 37,619 | 36,903 | 35,873 | 39,726 | 48,733 | 23,178 | 20,237 | 31,159 |
| Total | 137,504 | 130,970 | 129,074 | 104,958 | 147,195 | 149,485 | 144,353 | 193,607 |


| Sub-area | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 3,311 | 6,818 | 4,809 | 11,414 | 3200 | 13457 | 0 | 759 |
| $4+3 . a$ | 20,895 | 62,892 | 112,047 | 145,062 | 71,195 | 120,054 | 145,965 | 111,899 |
| 6 | 35,157 | 45,842 | 34,870 | 20,904 | 29,726 | 39,061 | 65,397 | 69,616 |
| 7 | 100,734 | 90,253 | 138,890 | 192,196 | 150,575 | 183,458 | 202,083 | 196,192 |
| 8 | 37,703 | 34,177 | 38,686 | 46,302 | 42,840 | 54,172 | 44,726 | 35,501 |
| 9 | 24,540 | 29,763 | 29,231 | 24,023 | 34,992 | 27,858 | 31,521 | 28,442 |
| Disc | 222,340 | 269,745 | 358,533 | 439,901 | 337,968 | 440,280 | 499,222 | 446,974 |
| Total |  |  |  |  | 5,440 | 2,220 | 9,530 | 4,565 |


| Sub-area | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 13151 | 3366 | 2601 | 2544 | 2557 | 919 | 310 | 1324 |
| $4+3 . a$ | 100,916 | 25,998 | 79,761 | 34,917 | 58,745 | 31,435 | 18,513 | 52,337 |
| 6 | 83,568 | 81,311 | 40,145 | 35,073 | 40,381 | 20,735 | 24,839 | 14,843 |
| 7 | 328,995 | 263,465 | 326,469 | 300,723 | 186,622 | 140,190 | 138,428 | 98,677 |
| 8 | 28,707 | 48,360 | 40,806 | 38,571 | 48,350 | 54,197 | 75,067 | 55,897 |
| 9 | 25,147 | 20,400 | 29,491 | 41,574 | 27,733 | 26,160 | 24,912 | 23,665 |
| Disc | 2,076 | 17,082 | 168 | 996 | 0 | 385 | 254 | 307 |
| Total | 582,560 | 459,982 | 519,441 | 454,398 | 364,388 | 274,022 | 282,323 | 247,049 |


| Sub-area | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 36 | 42 | 176 | 27 | 366.34 | 572 | 1847 | 1667 |
| 4 + $3 . a$ | 34,095 | 30,736 | 40,594 | 37,583 | 16,226 | 15,628 | 78,064 | 13,600 |
| 6 | 23,772 | 22,177 | 22,053 | 15,722 | 25,949 | 25,867 | 17,775 | 23,199 |
| 7 | 123,428 | 115,739 | 106,671 | 101,183 | 93,013 | 102,755 | 96,915 | 148,701 |
| 8 | 41,711 | 24,126 | 41,491 | 34,121 | 28,396 | 33,756 | 33,580 | 39,659 |
| 9 | 19,570 | 23,581 | 23,111 | 24,557 | 23,423 | 23,596 | 26,496 | 27,217 |
| Disc | 842 | 2,356 | 1,864 | 1,431 | 509 | 474 | 1,483 | 434 |
| Total | 243,455 | 218,758 | 235,961 | 214,624 | 187,882 | 202,649 | 256,161 | 254,478 |
| Sub-area | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| 2 | 647.588 | 66.02912 | 30 | 424.291 | 10 | 45.276 | 5 | 718 |
| $4+3 . a$ | 25,158 | 5,234 | 8,183 | 17,270 | 10,560 | 11,565 | 12,609 | 11,758 |
| 6 | 39,496 | 44,971 | 43,266 | 32,444 | 24,153 | 32,186 | 28,170 | 38,896 |
| 7 | 120,340 | 120,476 | 100,859 | 66,853 | 49,644 | 46,901 | 33,297 | 38,816 |
| 8 | 35,245 | 17,209 | 26,983 | 30,844 | 19,822 | 17,511 | 18,307 | 23,393 |
| $9^{1}$ | 22,575 | 25,316 | 29,382 | 29,205 | 33,179 | 41,081 | 37,080 | 31,920 |
| Disc | 430 | 3,279 | 4,582 | 1,904 | 6,232 | 5,944 | 5,488 | 2,873 |
| Total | 243,892 | 216,552 | 213,285 | 178,945 | 143,600 | 155,232 | 134,956 | 148,374 |


| Sub-area | 2019 | 2020 | 2021 |
| :--- | :--- | :--- | :--- |
| 2 | 867 | 290 | 12 |
| $4+3 . a$ | 12,593 | 13,792 | 7,672 |
| 6 | 47,351 | 19,037 | 13,727 |
| 7 | 42,973 | 33,310 | 49,934 |
| 8 | 29,640 | 19,639 | 19,602 |
| $9^{1}$ | 3,080 | 31,344 | 26,745 |
| Disc | 170,829 | 120,347 | 119,384 |
| Total | 1,692 |  |  |

[^6]Table 5.1.2 HORSE MACKEREL Western and North Sea Stock combined.
Quarterly catches ( t ) by Division and Subdivision in 2021.

| Division | 1Q | 2Q | 3Q | 4Q | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2.a+5.b | 5 | 1 | 0 | 6 | 12 |
| 3 | 1 | 0 | 7 | 6 | 14 |
| 4.a | 2080 | 190 | 1897 | 1790 | 6180 x |
| 4.bc | 371 | 257 | 487 | 524 | 1639 |
| 7.d | 888 | 203 | 140 | 5922 | 7172* |
| 6.a,b | 10691 | 2 | 8 | 2568 | 13819** |
| 7.a-c,e-k | 26735 | 590 | 5481 | 10592 | 43490*** |
| 8.a-e | 2389 | 6805 | 5496 | 5624 | 20314 |
| Sum | 43160 | 8048 | 13516 | 27032 | 92460**** |

* for the total 221t were added which were only declared as yearly catch
** for the total 19t were added which were only declared as yearly catch
*** for the total 550t were added which were only declared as yearly catch
**** for the total 92t were added which were only declared as yearly catch
***** for the total 882t were added which were only declared as yearly catch
X includes 222t declared as yearly catch

Table 5.4.1 HORSE MACKEREL General. Landings and discards ( $\mathbf{t}$ ) by year and ICES Division, for the North Sea, Western, and Southern horse mackerel stocks. (Data submitted by Working Group members.)

| Year | 3.9 | 4.a | 4.b,c | 7.d | Disc | NS Stock | 2.a 5.b | 3.a | 4.a | 6.a,b | $\begin{aligned} & \text { 7.a-c, e- } \\ & \text { k } \end{aligned}$ | 8.a-e | Disc | Western Stock | $\begin{aligned} & \text { W + NS } \\ & \text { Stock } \end{aligned}$ | Southern Stock(9.a) ${ }^{\mathrm{x}}$ | All Stocks |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 2,788* |  | - | 1,247 |  | 4,035 | - |  | - | 6,283 | 32,231 | 3,073 | - | 61,197 | 65,232 | 39,726 | 104,958 |
| 1983 | 4,420* |  | - | 3,600 |  | 8,020 | 412 |  | - | 24,881 | 36,926 | 28,223 | - | 90,442 | 98,462 | 48,733 | 147,195 |
| 1984 | 25,893* |  | - | 3,585 |  | 29,478 | 23 |  | 94 | 31,716 | 38,782 | 25,629 | 500 | 96,744 | 126,222 | 23,178 | 149,400 |
| 1985 | - |  | 22,897 | 2,715 |  | 26,750 | 79 |  | 203 | 33,025 | 35,296 | 27,740 | 7,500 | 103,843 | 129,455 | 20,237 | 150,830 |
| 1986 | - |  | 19,496 | 4,756 |  | 24,648 | 214 |  | 776 | 20,343 | 72,761 | 43,405 | 8,500 | 145,999 | 170,251 | 31,159 | 201,806 |
| 1987 | 1,138 |  | 9,477 | 1,721 |  | 11,634 | 3,311 |  | 11,185 | 35,197 | 99,942 | 37,703 | - | 187,338 | 199,674 | 24,540 | 223,512 |
| 1988 | 396 |  | 18,290 | 3,120 |  | 23,671 | 6,818 |  | 42,174 | 45,842 | 81,978 | 34,177 | 3,740 | 214,729 | 236,535 | 29,763 | 268,163 |
| 1989 | 436 |  | 25,830 | 6,522 |  | 33,265 | 4,809 |  | 85304** | 34,870 | 131,218 | 38,686 | 1,150 | 296,037 | 328,825 | 29,231 | 358,533 |
| 1990 | 2,261 |  | 17,437 | 1,325 |  | 18,762 | 11,414 | 14,878 | 112753** | 20,794 | 182,580 | 46,302 | 9,930 | 398,645 | 419,668 | 24,023 | 441,430 |
| 1991 | 913 | 0 | 11,400 | 600 | 0 | 12,913 | 3,200 | 2,725 | 56,157 | 29,726 | 149,975 | 42,840 | 5,440 | 290,063 | 302,976 | 34,992 | 337,968 |
| 1992 | 0 | 0 | 13,955 | 688 | 400 | 15,043 | 13,457 | 2,374 | 103,725 | 39,061 | 182,770 | 54,172 | 1,820 | 397,379 | 412,422 | 27,858 | 440,280 |


| Year | 3.a | $4 . a$ | 4.b,c | 7.d | Disc | NS Stock | 2.a 5.b | 3.a | 4.a | 6.a,b | $\begin{aligned} & \text { 7.a-c, e- } \\ & \text { k } \end{aligned}$ | 8.a-e | Disc | Western Stock | W + NS <br> Stock | Southern <br> Stock(9.a) ${ }^{\mathrm{x}}$ | All Stocks |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0 | 0 | 3,895 | 8,792 | 930 | 13,617 | 0 | 850 | 141,220 | 65,397 | 193,291 | 44,726 | 8,600 | 454,084 | 467,701 | 31,521 | 499,222 |
| 1994 | 0 | 0 | 2,496 | 2,503 | 630 | 5,629 | 759 | 2,492 | 106,911 | 69,616 | 193,689 | 35,501 | 3,935 | 412,903 | 418,532 | 28,442 | 446,974 |
| 1995 | 112 | 0 | 7,948 | 8,666 | 30 | 16,756 | 13,151 | 128 | 92,728 | 83,568 | 320,329 | 28,707 | 2,046 | 540,657 | 557,413 | 25,147 | 582,560 |
| 1996 | 1,657 | 0 | 7,558 | 9,416 | 212 | 18,843 | 3,366 | 0 | 16,783 | 81,311 | 254,049 | 48,360 | 16,870 | 420,739 | 439,582 | 20,400 | 459,982 |
| 1997 | 0 | 0 | 14,078 | 5,452 | 10 | 19,540 | 2,601 | 2,037 | 63,646 | 40,145 | 321,017 | 40,806 | 158 | 470,410 | 489,950 | 29,491 | 519,441 |
| 1998 | 3,693 | 0 | 10,530 | 16,194 | 83 | 30,500 | 2,544 | 3,693 | 17,001 | 35,073 | 284,529 | 38,571 | 913 | 382,324 | 412,824 | 41,574 | 454,398 |
| 1999 | 0 | 0 | 9,335 | 27,889 | 0 | 37,224 | 2,557 | 2,095 | 47,315 | 40,381 | 158,733 | 48,350 | 0 | 299,431 | 336,655 | 27,733 | 364,388 |
| 2000 | 0 | 176 | 25,931 | 19,019 | 4 | 45,130 | 919 | 1,014 | 4,314 | 20,735 | 121,171 | 54,197 | 382 | 202,732 | 247,862 | 26,160 | 274,022 |
| 2001 | 43 | 212 | 6,686 | 21,390 | 0 | 28,331 | 310 | 134 | 11,438 | 24,839 | 117,038 | 75,067 | 254 | 229,081 | 257,411 | 24,912 | 282,323 |
| 2002 | 0 | 639 | 15,303 | 11,323 | 0 | 27,264 | 1,324 | 174 | 36,221 | 14,843 | 87,354 | 55,897 | 307 | 196,120 | 223,384 | 23,665 | 247,049 |
| 2003 | 49 | 622 | 10,309 | 21,049 | 0 | 32,028 | 36 | 1,843 | 21,272 | 23,772 | 102,379 | 41,711 | 842 | 191,856 | 223,885 | 19,570 | 243,455 |
| 2004 | 303 | 133 | 18,544 | 16,455 | 0 | 35,435 | 42 | 48 | 11,708 | 22,177 | 99,284 | 24,126 | 2,356 | 159,742 | 195,177 | 23,581 | 218,758 |


| Year | 3.a | 4.a | 4.b,c | 7.d | Disc | NS Stock | 2.a 5.b | 3.a | 4.a | 6.a,b | $\begin{aligned} & \text { 7.a-c, e- } \\ & \text { k } \end{aligned}$ | 8.a-e | Disc | Western Stock | $\begin{aligned} & \text { W + NS } \\ & \text { Stock } \end{aligned}$ | Southern Stock(9.a) ${ }^{\mathrm{x}}$ | All Stocks |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 0 | 1,331 | 13,995 | 15,460 | 62 | 30,848 | 176 | 284 | 24,983 | 22,053 | 91,211 | 41,491 | 1,802 | 182,001 | 212,850 | 23,111 | 235,961 |
| 2006 | 185 | 2,192 | 7,996 | 23,789 | 78 | 34,240 | 27 | 58 | 27,152 | 15,722 | 77,394 | 34,121 | 1,353 | 155,827 | 190,067 | 24,557 | 214,624 |
| 2007 | 11 | 2,051 | 9,114 | 29,789 | 139 | 41,103 | 366 | 110 | 4,940 | 25,949 | 63,224 | 28,396 | 370 | 123,356 | 164,459 | 23,423 | 187,882 |
| 2008 | 27 | 910 | 2,582 | 32,185 | 0 | 35,704 | 572 | 3 | 12,107 | 25,867 | 70,570 | 33,756 | 474 | 143,349 | 179,053 | 23,596 | 202,649 |
| 2009 | 21 | 314 | 18,975 | 25,537 | 1,036 | 45,883 | 1,847 | 17 | 58,738 | 17,775 | 71,378 | 33,580 | 447 | 183,782 | 229,665 | 26,496 | 256,161 |
| 2010 | 0 | 100 | 1,969 | 22,077 | 2 | 24,149 | 1,667 | 88 | 11,442 | 23,199 | 126,624 | 39,659 | 432 | 203,112 | 227,261 | 27,217 | 254,478 |
| 2011 | 0 | 0 | 10,435 | 17,184 | 0 | 27,619 | 648 | 0 | 14,723 | 39,496 | 103,156 | 35,245 | 430 | 193,698 | 221,317 | 22,575 | 243,892 |
| 2012 | 0 | 355 | 1,559 | 19,464 | 0 | 21,378 | 66 | 9 | 3,311 | 44,971 | 101,012 | 17,209 | 3,279 | 169,858 | 191,236 | 25,316 | 216,552 |
| 2013 | 0 | 17 | 1,453 | 17,175 | 0 | 18,645 | 30 | 10 | 6,702 | 43,266 | 83,684 | 26,983 | 4,582 | 165,258 | 183,903 | 29,382 | 213,285 |
| 2014 | 1 | 2 | 2,597 | 10,772 | 7 | 13,380 | 424 | 4,096 | 10,573 | 32,444 | 56,081 | 30,844 | 1,896 | 136,360 | 149,740 | 29,205 | 178,945 |
| 2015 | 3 | 644 | 770 | 8,581 | 2,004 | 12,002 | 10 | 65 | 9,078 | 24,153 | 41,063 | 19,822 | 4,228 | 98,419 | 110,421 | 33,179 | 143,600 |
| 2016 | 2 | 1,628 | 975 | 11,209 | 1,527 | 15,341 | 45 | 0 | 8,960 | 32,186 | 35,692 | 17,511 | 4,417 | 98,811 | 114,151 | 41,081 | 155,232 |


| Year | $3 . a$ | 4.a | 4.b,c | 7.d | Disc | NS <br> Stock | 2.a 5.b | 3.a | 4.a | 6.a,b | $\begin{aligned} & \text { 7.a-c, e- } \\ & \text { k } \end{aligned}$ | 8.a-e | Disc | Western Stock | $\begin{aligned} & \text { W + NS } \\ & \text { Stock } \end{aligned}$ | Southern Stock(9.a) ${ }^{\mathrm{x}}$ | All Stocks |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | 0 | 22 | 2,557 | 10,787 | 1,213 | 14,579 | 5 | 697 | 9,332 | 28,170 | 22,510 | 18,307 | 3,939 | 82,961 | 97,540 | 37,088 | 134,956 |
| 2018 | 0 | 1,418 | 1,413 | 11,677 | 265 | 14,773 | 718 | 380 | 8,547 | 38,896 | 27,140 | 23,393 | 2,609 | 101,683 | 116,456 | 31,920 | 148,376 |
| 2019 | 0.5 | 2,571 | 1,217 | 7,829 | 185 | 11,803 | 867 | 490 | 8,314 | 47,351 | 35,144 | 29,640 | 3,141 | 124,947 | 136,750 | 34,080 | 170,830 |
| 2020 | 0 | 2,211 | 1,099 | 9,077 | 201 | 12,587 | 290 | 96 | 10,387 | 19,037 | 24,232 | 19,359 | 2,741 | 76,422 | 89,009 | 31,344 | 120,347 |
| 2021 | 1 | 2,270 | 1,639 | 7,120 | 52 | 11,082 | 12 | 12 | 3,751 | 13,727 | 42,813 | 19,602 | 1,641 | 81,557 | 92,639 | 26,745 | 119,384 |

*Divisions 3.a and 4.b,c combined•
**Norwegian catches in 4.b included in Western horse mackerel-
${ }^{\mathrm{x}}$ Southern Horse Mackerel is assessed by ICES WGHANSA since 2011

Table 5.4.2 National catches of the Western Horse mackerel stock.

| Country | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 18 | 19 | 21 | 0 | - | - | - | - | - |
| Denmark | 62,897 | 31,023 | 26,040 | 16,385 | 21,254 | 10,147 | 11340 | 11,667 | 10,155 |
| Estonia | 78 | 22 | - | 0 | - | - | - | 3,826 | 3,695 |
| Faroe Islands | 1,095 | 216 | 1,040 | 24 | 800 | 671 | 4 | 8,056 | 10,690 |
| France | 39,188 | 26,667 | 25,141 | 20,457 | 15,145 | 18,951 | 10,381 | 17,744 | 16,364 |
| Germany, Fed.Rep. | 28,533 | 33,716 | 23,549 | 13,014 | 11,491 | 12,658 | 15,696 | 26,432 | 34,607 |
| Ireland | 74,250 | 73,672 | 57,983 | 55,229 | 51,874 | 36,422 | 35,857 | - | - |
| Lithuania | - | - | - | - | - | - | - | 40986 | 41,057 |
| Netherlands | 82,885 | 103,246 | 83,450 | 57,261 | 73,440 | 44,997 | 48,924 | 10729 | 24,909 |
| Norway | 45,058 | 13,363 | 46,648 | 1,982 | 7,956 | 36,164 | 20,371 | 16,272 | 16,636 |
| Russia | 554 | 345 | 121 | 80 | 16 | 3 | 2 | 567 | 216 |
| Spain | 31,087 | 43,829 | 39,831 | 24,204 | 23,537 | 24,763 | 24,599 | 4,617 | 3,560 |
| Sweden | 1,761 | 3411 | 1,957 | 1009 | 68 | 561 | 1,002 | 458 | 210 |
| UK (Engl. + Wales) | 19,778 | 13,068 | 9,268 | 4,554 | 7,096 | 5,970 | 4,438 | 1,522 | 143 |
| UK (N. Ireland) | - | 1,158 | - | 625 | 1140 | 1129 | 914 | 14,506 | 17,962 |
| UK (Scotland) | 32,865 | 18,283 | 11,197 | 10,283 | 8,026 | 2,905 | 721 | 2356 | 1802 |
| Unallocated | 17,158 | 15,262 | 23,763 | -2757 | 6,978 | 472 | 16,765 | 159,737 | 182,006 |
| Discard | 158 | 913 | - | 382 | 254 | 307 | 842 | - | - |
| Total | 437,363 | 378,213 | 350,009 | 202,732 | 229,075 | 196,120 | 191,856 | 11,667 | 10,155 |


| Country | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | - | - | - | - | 19 | 2 | 0.2 | 14 |
| Denmark | 8,411 | 7,617 | 5,261 | 6,027 | 5,940 | 6,108 | 4,002 | 6,820 |
| Faroe Islands | - | 478 | 841 | - | 377 | 349 | - |  |
| France | 11,031 | 12,748 | 12,626 | - | 260 | 8,271 | 1,797 | 3,595 |
| Germany, Fed.Rep. | 10,862 | 5,784 | 11,801 | 15,122 | 17,688 | 21,114 | 17,063 | 24,835 |
| Ireland | 26,779 | 29,759 | 35,332 | 40,754 | 44,488 | 38,466 | 45,239 | 35,791 |
| Lithuania | 6,828 | 5,467 | 5,548 | - | - | - | - |  |
| Netherlands | 37,130 | 29,462 | 43,648 | 39,453 | 61,504 | 55,690 | 66,396 | 53,697 |
| Norway | 27,114 | 4,182 | 12,223 | 59,764 | 11,978 | 13,755 | 3,251 | 6,596 |
| Spain | 13,877 | 14,277 | 19,851 | 21,077 | 38,745 | 34,581 | 13560 | 22,541 |
| Sweden | - | 76 | 8 | 258 | 2 | 90 | - | 1 |
| UK (Engl. + Wales) | 3,574 | 5,482 | 3,365 | 6,482 | 12,714 | 11,716 | 12,122 | 3,959 |
| UK (N. Ireland) | 103 | - | - | - | 59 | 198 | - | 2,325 |
| UK (Scotland) | 468 | 776 | 1,077 | 1,412 | 2,349 | 2,928 | 1,335 | 504 |
| Unallocated | 8,292 | 6,878 | $-8,703$ | -7,014 | 6,556 | - | 1815 | - |
| Discard | 1353 | 370 | 474 | 447 | 432 | 430 | 3,280 | 4,582 |
| Total | 155,822 | 123,356 | 143,352 | 183,782 | 203,111 | 193,698 | 169,860 | 165,260 |


| Country | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium |  |  |  |  |  |  |  |  |
| Denmark | 5,945 | 4,556 | 321 | 4,541 | 6,302 | 7,764 | 5,487 | 6,042 |
| Faroe Islands | 68 | - | - | 180 | - | 26 | - |  |
| France | 3,428 | 3,247 | 2,797 | 3,923 | 3,443 | 4,382 | 2,217 | 2,710 |
| Germany, Fed.Rep. | 17,161 | 9,417 | 11,414 | 7,172 | 4,734 | 9,211 | 954 | 5,530 |
| Ireland | 32,667 | 21,654 | 27,605 | 23,560 | 25,347 | 28,899 | 17,390 | 18,770 |
| Lithuania | - | - | 2,596 | - | - | - | 0 |  |
| Netherlands | 25,053 | 24,958 | 23,792 | 14,269 | 25,942 | 29,656 | 14,240 | 20,786 |
| Norway | 14,353 | 8,897 | 9,438 | 9,885 | 9,319 | 9,021 | 10,666 | 3,663 |
| Poland | - | - | -- | - | - | 127 | 1,002 | 1,605 |
| Spain | 19,442 | 13,071 | 14,235 | 14,901 | 20,362 | 25,776 | 18,582 | 16,191 |
| Sweden | 0 | 10 | - | 41 | 23 | 323 | 83 | 4 |
| UK (Engl. + Wales) | 4,832 | 2,063 | 842 | 549 | 2,443 | 4,036 | 1,496 | 2,651 |
| UK (N. Ireland) | 1,579 | 1,204 | - |  | 1,080 | 1,907 | 1,231 | 1,350 |
| UK (Scotland) | 1,389 | 738 | 970 | - | - | 678 | 333 | 615 |
| Unallocated | 8,545 | 4,377 | 1,010 | 3,994 | 74 | - | - | - |
| Discard | 1,896 | 4,228 | 4,417 | 3,928 | 2,609 | 3,141 | 2,741 | 1,641 |
| Total | 136,360 | 98,419 | 98,810 | 82,950 | 101,682 | 124,947 | 76,422 | 81,557 |

Table 5.4.3. National catches of the North Sea Horse mackerel stock.

| Country | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium | - | 19 | 21 |  | 30 | 5 | 4 | 4 | - |  |
| Denmark | 180 | 1,481 | 3,377 | 4,403 | 885 | 2,315 | 3,301 | 8,690 | 3,987 | 8,353 |
| Faroe Islands | - | - | 135 | - | - | 28 | 804 | 21 | - | - |
| France | 3,246 | 2,399 | - | - |  | 1,246 | 2,326 | 231 | 5,236 | 1,205 |
| Germany, Fed.Rep. | 7,847 | 5,844 | 5,920 | 3,728 | 974 | 6,532 | 2,936 | 5,194 | 2,725 | 11,034 |
| Ireland | - | 2,861 | 27 | 201 | 338 | 61 | - | 1 | 753 | 10,863 |
| Lithuania | - | 10,711 | - | - | - | - | - | - | - | 26,779 |
| Netherlands | 36,855 | - | 8,117 | 8,697 | 13,867 | 12,209 | 24,119 | 26,303 | 27,730 | 6,829 |
| Norway | - | - | 238 | 105 | 36 | 525 | 144 | 22 | 204 | 37,130 |


| Country | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sweden | - | 3,401 | 5 | 40 | 46 | 16 | 72 | 98 | 4 | 27,114 |
| UK (Engl. + Wales) | 269 | 907 | 11 | 1,585 | 3,425 | 2,322 | 1,966 | 5,633 | 3,859 | - |
| UK (Scotland) | 29 | - - | - 4 | 421 | - | 2 | 1 | 2 | - | 13,878 |
| Unallocated | -28,896 | 2,794 | 19,373 | 25,944 | 48,805 | 1,981 | -3,645 | -13,064 | -13,719 | - |
| Discard | 10 | 83 | - 4 | 4 | - |  | - | - | 62 | 3,583 |
| Total | 19,540 | 30,500 | 37,224 | 45,128 | 8 28,376 | 27,267 | 32,029 | 33,135 | 30,845 | 155,094 |
| Country | 2006 | 2007 | 2008 |  | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| Belgium |  |  |  | 4 | 4 | 16 |  | 46 | 51.077 | 74 |
| Denmark | 1,283 | 252 | 57 |  | 72 | 15 | 142 | 1514 | 1,020 | 552 |
| Faroe Islands | - | - | - | - | - - | - | - | 0 |  |  |
| France | 4,380 | 5,349 | 2,247 | 7 - | - 81 | 813 | 273 | 1,047 | 1,010 | 1,742 |
| Germany, Fed.Rep. | 1,125 | 65 | 1,081 |  | 1,539 | 3,794 | 3,461 | 5,356 | 2,941 | 1,619 |
| Ireland | 2,077 |  | 887 |  | 25 | - | - | 0 |  | 0 |
| Lithuania | 1,999 | 297 | - | - | - - | - | - | 0 |  | 0 |
| Netherlands | 27,285 | 5 31,153 | 319,439 |  | 22,546 | 17,093 | 16,289 | 12,157 | 8,725 | 4,925 |
| Norway | 113 | 1,243 | 21 |  | 12,855 | 526 | 7,359 | 129 | 377 | 0 |
| Sweden | 9 | 21 | 36 |  | 401 | - | - | 0 |  | 1 |
| UK (Engl. + Wales) | 595 | 6921 | 1,061 |  | 1,435 | 1,890 |  | 935 | 4,401 | 4,198 |
| UK (Scotland) | 300 | 625 | 7 | 4 | 4 | 111 | 93 | 240 | 172 | 262 |
| Unallocated | -5,004 | -4,960 | 10,869 |  | 5,964 | -116 | 0 | 0 | 0 |  |
| Discard | 78 | 139 | - |  | 1,036 | 2 | 0 | 0 | 0 | 7 |
| Total | 34,240 | 0 41,105 | 35,70 |  | 45,881 | 24,144 | 27,617 | 21,424 | 18,696 | 13,380 |
| Country |  | 2015 | 2016 |  | 2017 | 2018 | 2019 |  |  | 2021 |
| Belgium |  | 63 | 51 |  | 67 | 44 | 18 | 39 |  | 38 |
| Denmark |  | 800 | 268 |  | 294 | 397 | 100 | 17 |  | 72 |
| Faroe Islands |  | 0 | 0 | 4 | 4 | 0 | 10 | 10 |  |  |
| France |  | 934 | 1,322 |  | 1,863 | 1,443 | 935 | 75 |  | 503 |


| Germany, Fed.Rep. | 644 | 1,879 | 949 | 2,766 | 946 | 3 | 87 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ireland | 0 | 0 | 0 | 0 | 0 | 0 | 174 |
| Lithuania | 0 | 0 | 0 | 0 | 1,254 | 0 |  |
| Netherlands | 3,305 | 3,892 | 5,638 | 5,184 | 2,089 | 4,803 | 3,377 |
| Norway | 662 | 1,701 | 5 | 1,423 | 2,543 | 2,090 | 2,091 |
| Sweden | 9 | 0 | 0 | 0 | 0 | 1 | 0 |
| UK (Engl. + Wales) | 3,581 | 4,697 | 4,546 | 3,250 | 3,632 | 4,381 | 4,669 |
| UK (Northern Ireland) | 0 | 0 | 0 | 0 | 53 | 0 |  |
| UK (Scotland) | 0 | 0 | 0 | 0 | 38 | 24 | 19 |
| Unallocated | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Discard | 2,004 | 1,527 | 1,213 | 265 | 185 | 201 | 52 |
| Total | 12,002 | 15,337 | 14,579 | 14,773 | 11,802 | 12,587 | 11,082 |

Table 5.6.1. Catches $(t)$ of Trachurus mediterraneus in Divisions 8.ab, 8.c and Sub-Area 7

|  | 7 | 8.ab | 8.c East | 8.c West | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | 0 | 23 | 3903 |  | 3926 |
| 1990 | 0 | 298 | 2943 |  | 3241 |
| 1991 | 0 | 2122 | 5020 |  | 7142 |
| 1992 | 0 | 1123 | 4804 |  | 5927 |
| 1993 | 0 | 649 | 5576 |  | 6225 |
| 1994 | 0 | 1573 | 3344 |  | 4917 |
| 1995 | 0 | 2271 | 4585 |  | 6856 |
| 1996 | 0 | 1175 | 3443 |  | 4618 |
| 1997 | 0 | 557 | 3264 |  | 3821 |
| 1998 | 0 | 740 | 3755 |  | 4495 |
| 1999 | 0 | 1100 | 1592 |  | 2692 |
| 2000 | 59 | 988 | 808 |  | 1854 |
| 2001 | 1 | 525 | 1293 |  | 1820 |
| 2002 | 1 | 525 | 1198 |  | 1724 |
| 2003 | 0 | 340 | 1699 |  | 2039 |
| 2004 | 0 | 53 | 841 |  | 894 |


|  | 7 | 8.ab | 8.c East | 8.c West | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 1 | 155 | 1005 |  | 1162 |
| 2006 | 1 | 168 | 794 |  | 963 |
| 2007 | 0 | 126 | 326 |  | 452 |
| 2008 | 0 | 82 | 405 |  | 487 |
| 2009 | 0 | 42 | 1082 |  | 1124 |
| 2010 | 0 | 97 | 370 |  | 467 |
| 2011 | 0 | 119 | 1096 |  | 1225 |
| 2012 | 0 | 186 | 667 | 116 | 969 |
| 2013 | 0 | 52 | 238 | 0 | 290 |
| 2014 | 0 | 130 | 1160 | 0 | 1290 |
| 2015 | 0 | 8 | 890 | 0 | 899 |
| 2016 | 0 | 5 | 471 | 0 | 476 |
| 2017 | 0 | 18 | 684 | 0 | 702 |
| 2018 | 0.4 | 38 | 640 | 0 | 678 |
| 2019 | 0.02 | 81 | 384 | 1 | 466 |
| 2020 | 0 | 0 | 558 | 2 | 560 |
| 2021 | 0.9 | 265 | 390 | 0 | 656 |

Table 5.9.1. Summary of the overall sampling intensity on horse mackerel catches in recent years in all areas 1992-2021

| Year | Total Catch (ICES estimate) | \% catch covered by sampling programme* | No. samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 436500 | 45 | 1803 | 158447 | 5797 |
| 1993 | 504190 | 75 | 1178 | 158954 | 7476 |
| 1994 | 447153 | 61 | 1453 | 134269 | 6571 |
| 1995 | 580000 | 48 | 2041 | 177803 | 5885 |
| 1996 | 460200 | 63 | 2498 | 208416 | 4719 |
| 1997 | 518900 | 75 | 2572 | 247207 | 6391 |
| 1998 | 399700 | 62 | 2539 | 245220 | 6416 |
| 1999 | 363033 | 51 | 2158 | 208387 | 7954 |
| 2000 | 247862 | 50 | 378 | 33317 | 4126 |
| 2001 | 257411 | 61 | 467 | 46885 | 7141 |


| Year | Total Catch (ICES estimate) | \% catch covered by sampling programme* | No. samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | 223384 | 68 | 540 | 79103 | 6831 |
| 2003 | 223885 | 77 | 434 | 59241 | 8044 |
| 2004 | 195177 | 62 | 518 | 62720 | 9273 |
| 2005 | 212850 | 76 | 573 | 67898 | 8840 |
| 2006 | 190067 | 75 | 602 | 57701 | 9905 |
| 2007 | 164459 | 58 | 397 | 41046 | 8061 |
| 2008 | 179053 | 72 | 488 | 46768 | 8870 |
| 2009 | 229665 | 84 | 902 | 57505 | 10575 |
| 2010 | 227261 | 82 | 710 | 49307 | 14159 |
| 2011 | 221317 | 71 | 502 | 40492 | 7484 |
| 2012 | 191236 | 69 | 501 | 41148 | 8220 |
| 2013 | 183903 | 75 | 686 | 87300 | 9776 |
| 2014 | 149740 | 83 | 650 | 53945 | 8085 |
| 2015 | 110421 | 68 | 825 | 39415 | 7034 |
| 2016 | 114151 | 76 | 1033 | 93853 | 6675 |
| 2017 | 97539 | 63 | 1113 | 116722 | 8221 |
| 2018 | 116455 | 74 | 1584 | 117768 | 6965 |
| 2019 | 136750 | 64 | 1014 | 77211 | 7476 |
| 2020 | 89,009 | 52 | 516 | 41811 | 5662 |
| 2021 | 92,639 | 77 | 977 | 59222 | 8080 |

*Percentage related to catch (catch at age) according to ICES estimation

Table 5.9.2. Horse mackerel sampling intensity for the Western stock in 2021.

| Country | Catch | \% Catch Sampled* | No. Samples | No. Measured | No. Aged |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Denmark | 6042 | 0 | 0 | 0 | 0 |
| Faroe Islands | - | 0 | 0 | 2034 | 0 |
| France** | 3288 | $-*$ | 106 | 7952 | 0 |
| Germany | 5530 | 99 | 49 | 9609 | 49 |
| Ireland | 18770 | 96 | 59 | 7774 | 59 |
| Netherlands | 20785 | 89 |  |  | 49 |


| Country | Catch | \% Catch Sampled* | No. Samples | No. Measured | No. Aged |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Norway | 3662 | 99 | 8 | 422 | 8 |
| Poland | 1605 | 0 | 0 | 0 | 23750 |
| Spain | 16994 | 97 | 763 | 0 | 3391 |
| Sweden | 6 | 0 | 67 | 0 | 0 |
| UK (England) | 2658 | 100 | 0 | 1344 | 0 |
| UK(Northern Ireland) | 1350 | 0 | 95 | 52702 | 0 |
| UK(Scotland) | 863 | $-*$ | 78 | 01557 |  |

*Percentage based on ICES estimate with regards to age samples
**provided only length distributions
*** age samples processed by the Netherlands

Table 5.9.3. Horse mackerel sampling intensity for the North Sea stock in 2021.

| Country | Catch | \% Catch Sampled* | No. Samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 37 | 0 | 0 | 0 | 0 |
| Denmark | 72 | 0 | 0 | 0 | 0 |
| Faroe Islands | 0 | 0 | 0 | 0 | 0 |
| France** | 549 | -* | 19 | 373 | 0 |
| Germany | 87 | 0 | 0 | 0 | 0 |
| Netherlands | 3377 | 88 | 9 | 1600 | 225 |
| Norway | 2091 | 0 | 0 | 0 | 0 |
| Sweden | 0 | 0 | 0 | 0 | 0 |
| UK (England)**** | 4674 | 100 | 24 | 5137 | 422 |
| UK(Northern Ireland) | 0 | 0 | 0 | 0 | 0 |
| UK(Scotland)*** | 19 | 0 | 0 | 0 | 0 |
| Total | 11082 | 68 | 26 | 6520 | 647 |

[^7]**provided only length distributions
*** provided length distributions not incl. in InterCatch
****age samples processed by the Netherlands

### 7.12 Figures



Figure 5.1.1a. Horse mackerel catches 1st quarter 2021


Figure 5.1.1b. Horse mackerel catches $\mathbf{2}^{\text {nd }}$ quarter 2021.


Figure 5.1.1c. Horse mackerel catches $3^{\text {rd }}$ quarter 2021.


Figure 5.1.1d. Horse mackerel catches $4^{\text {th }}$ quarter 2021.


Figure 5.2.1: Distribution of Horse Mackerel in the Northeast-Atlantic: Stock definitions as used by the 2004 WGMHSA. Note that the "Juvenile Area" is currently only defined for the Western Stock distribution area - juveniles do also occur in other areas (like in Div. 7.d). Map source: GEBCO, polar projection, $\mathbf{2 0 0} \mathbf{m}$ depth contour drawn.


Figure 5.3.1. Total catch for Western Horse Mackerel stock, 1982-2021.
North Sea HOM Catches


Figure 5.3.4. Total catch for North Sea Horse Mackerel stock, 1982-2021


Figure 5.4.1 Horse mackerel general overview. Total catches in the northeast Atlantic during the period 1982-2021. The catches taken from the southern, western and North Sea horse mackerel stocks are shown in relation to the total catches in the northeast Atlantic. Catches from Div. 8.c were transferred from southern stock to western stock from 1982 onwards. Southern horse mackerel is assessed by ICES WGHANSA since 2011.


Figure 5.4.2. North Sea horse mackerel stock. Total catches by Division during the period 1982-2021.

Western Stock 1982-2021 Catches by division


Figure 5.4.3. Western horse mackerel stock. Total catches by Sub-Area during the period 1982-2021.


Figure 5.7.1. Length distributions contributed by country and area of the Western and North Sea horse mackerel 2021.

North Sea HOM \% observed vs. estimated. 2000-21


Figure 5.9.1 North Sea horse mackerel stock. Percentage sampled catch (blue) vs. unsampled catch (red) by division and year, 2000-2021.


Figure 5.9.2. North Sea horse mackerel stock. Sampling intensity index as percentage sampled catch in total catch by year, 2000-2021.


Figure 5.9.5. Western horse mackerel stock. Percentage sampled catch (blue) vs. unsampled catch (red) by division and year. Period 2000-2021. Area of distribution of Western stock was divided into different regions. Chan: (7.e,f,h); WSCO+IRL (7.a-c, 7.j-k and 6.a); BoB (8.a,b,d); CantSea(8.c); N-Nsea (3.a and 4.a); NOR (2.a and 5.a).


Figure 5.9.6. Western horse mackerel stock. Sampling intensity index as percentage sampled catch in total catch by year, 2000-2021.

# 6 North Sea Horse Mackerel: Divisions 27.4.a (Q1 and Q2), 27.3.a (excluding Western Skagerrak Q3 and Q4), 27.4.b, 27.4.c and 27.7.d 

### 6.1 ICES Advice 2022

In 2012, the North Sea horse mackerel (NSHOM) was classified as a category 5 stock, based on the ICES approach to data-limited stocks (DLS). Since then, a progressive reduction in TAC was advised by ICES, from 25500 tonnes in 2013-2014 to 15200 tonnes in 2015-2016. This reduction in the advised catch was supported by the analysis of information from the North Sea International Bottom Trawl Survey (NS-IBTS) traditionally used in the assessment, but also new information from the French Channel Ground Fish Survey (CGFS) since 2014. Additionally, in 2015, information on discards in non-directed fisheries became available that has been taken into account in the advice since 2017.

In 2017, this stock was benchmarked and the NS-IBTS and CGFS survey indices where modelled together. The resulting joint index was considered an appropriate indicator of trend in abundance over time and the NSHOM stock assessment was upgraded to category 3 . The joint index showed an increasing trend in 2014 to 2016, but was followed by a decrease in 2017. In 2018, the index remained at a similar level as in 2017, while the index slightly increased in 2019. In 2020 no index value was calculated due to the absence of UK-stations in the French Ground fish survey. In 2021 the survey index decreased to values similar to 2018. Length-based DLS methods have been applied to data from 2016 onwards. The length-based F/Fmsy ratio has been decreasing since 2016, and F was estimated to be still slightly above Fmsy up till 2021. Stock size relative to reference points is unknown.
Biannual advice for 2022 and 2023 was provided in 2021, based on the data up to 2020 (ICES, 2021a). The uncertainty cap was applied, as the index ratio indicated a decrease of more than $20 \%$ in 2019-2020 compared to 2016-2018. The Lmean/LF=m ratio in 2020 was 0.912 , indicating that the fishing mortality was above Fmsy. Since the precautionary buffer was applied last in 2017 (more than three years ago), the PA buffer was applied again in 2021. This resulted in a catch advice for 2022 and 2023 of 8969 tonnes.

### 6.2 Fishery of North Sea horse mackerel stock

Based on historical catches taken by the Danish industrial fleet for reduction to fishmeal and fish oil in the 1970s and 1980s, approximately $48 \%$ of the EU North Sea horse mackerel TAC was taken by Denmark. Catches were taken in the fourth quarter mainly in divisions 4 b and 7 d . The 1990s saw a drop in the value of industrial fish, limited fishing opportunities and steep increases in fuel costs that affected the Danish quota uptake. In 2001, an individual quota scheme for a number of species was introduced in Denmark, but not for North Sea horse mackerel. This lead to a rapid restructuring and lower capacity of the Danish fleet, which in combination with the above mentioned factors led to a decrease of the Danish North Sea horse mackerel catches.

Since the 1990's, a larger proportion of the catches have been taken in a directed horse mackerel fishery for human consumption by the Dutch-owned freezer-trawler fleet. This is possible because Denmark has traded parts of its quota with the Netherlands for other species. However, due to the structure of the Danish quota management setup only a limited amount of quota can be made available for swaps with other countries. These practical implications of the
management scheme largely explain the consistent underutilisation of the TAC over the period 2010-2014 (approximately 50\%; Figure 6.2.1)). However, following the sharp reduction in TAC in 2015, uptake increased significantly in the years thereafter. In $2020,91 \%$ of the TAC was used, with the highest catches taken by the Netherlands, followed by UK, Norway and France. For 2021 the TAC utilization is $79 \%$ with the highest catches taken by the UK, followed by Netherlands, Norway and France (Figure 6.2.2)

Catches taken in Divisions 3a and 4a during the two first quarters and all year round in Divisions $4 b, 4 c$ and $7 d$ are currently regarded as North Sea horse mackerel (Section 5, Table 5.4.1). The catches were relatively low during the period $1982-1997$ with an average of 18000 t , but increased between 1998 ( 30500 t) and 2000 ( 45130 t). From 2000 to 2010, the catches varied between 24149 and 45883 t . Since 2014, a steep decline in catches is observed, both due to the reduction in the TAC since 2014 but also due to the underutilization of the quota. In 2021 the catch was 11 082 t similar to 2019, with $65 \%$ of the total catch being caught in 7 d , less than in 2020 ( $72 \%$, Figure 6.2.4).

Over the period 1985-2001 most catches were taken in the $4 b$ (Figure 6.2.3). However, since the early 2000s the proportion of catches from 7d increased steadily until 2013, when the $92 \%$ of total catches were fished in this area (Figure 6.2.4). In 2020, the Netherlands accounted for most of the landings, followed by UK, Norway and France (Figure 6.2.5). The majority was still caught in quarter 4 in 7 d , whereas the Norwegian catches were taken during quarters 1 and 2 in 4 a . The relative contribution of the catches in 4 a has been increasing since 2015, with the exception of 2017 when catches in 4a were negligible. In 2014 the contribution from 4a was only $0.02 \%$ but in 2021 the contribution is $20.5 \%$, partially because absolute catches in other areas have gone down.

Most of the discards reported were from 7d by the French bottom-trawl fleet. Discarding in the target pelagic fisheries is considered negligible. New information in 2015 from bottom-trawl fisheries (not directed at horse mackerel) indicated an overall discard rate of $16.7 \%$ for the stock as a whole, while in 2016 this rate was $10 \%$. Complete discard information for earlier years has not been submitted to ICES. Information from national discard reports for the non-directed bottomtrawl fisheries indicates a similar level of discarding in earlier years. In 2017 and 2018 the discard rate was $8.3 \%$ and $1.8 \%$, respectively, while it decreased to $1.6 \%$ in 2019 and 2020. In 2021 the discard rate dropped to $0.5 \%$.

### 6.3 Biological Data

### 6.3.1 Catch in Numbers at Age

For the second year in a row in 2021, it was possible to include samples taken from English vessels in the Netherlands, increasing further the biological sampling coverage. For 2021, the proportion of sampling increased to $67 \%$ in comparison to $56 \%$ in 2020 and for 2019 where only $1 / 3$ of the landings were sampled. Age samples were available from two countries (the Netherlands and UK/England) from all quarters in 4c and for Q1 and Q4 in 7d. Although most landed catch was taken from 7d in Q4, substantial parts of the landings were fished in other areas and quarters (Figure 6.2.5). In order to avoid a biased perception of the age distribution of catches over the year and areas, this partial and uneven sampling effort should be avoided in future years.

Annual catch numbers at age are shown in Table 6.3.1. Catch-at-age for the whole period 19952021 are given in Table 6.3.2 and in Figures 6.3 .1 and 6.3.2. These data show that since 2005 the age distribution of catches has experienced a reduction, with a decrease in the range of ages of importance in total catches. However, this decrease could be due to the low age sampling, in particular in 2018 (maximum age observed 7 years). In parallel to the rejuvenation of catches, the comparison of catch-at-age data after 1998 by area (Figure 6.3.2) shows that since 2010
commercial catches have increased in 7 d in comparison to 3 a and $4 \mathrm{a}, \mathrm{b}$ and c where the opposite pattern was found. Due to the low level of sampling effort in 2018, data for this year are only based on a single sample from 7d in Q4.

Although the 2015 cohort seems to be clear in the catch-at-age distribution, in general, cohort structure is not clearly detectable in the data. In addition to the low sampling levels, this may partly be due to the shifts in the distribution of the fishery. It may also be due to age reading difficulties, which are a known to be encountered (e.g. Bolle et al., 2011). Most clearly detectable is the relatively large 2001 year-class, although it is not clearly present in the catch data in all years. There are indications that environmental conditions may be an important factor (possibly stronger than stock size) contributing to spawning success of horse mackerel. This is, for example, illustrated by the largest year-classes (1982 and 2001) observed in the Western stock which were produced at the lowest observed stock sizes. Since 2001 is considered to have been a relatively strong year class in the Western stock, it is plausible that circumstances in the North Sea were similar to those in Western areas and also allowed for relatively high spawning success in the North Sea.

The potential for mixing of fish from the Western and North Sea stock in 7d and 7e in winter may also confuse the cohort signals. For example, the large recruitment in the Western stock may have led to more of these fish being located in the North Sea stock area as age 1 fish in 2002. On behalf of the Pelagic Advisory Council and the EAPO Northern Pelagic Working Group, a research project on genetic composition of horse mackerel stocks was initiated in 2015 with University College Dublin (Ireland) with the intention of clarifying the mixing among the North Sea and the Western horse mackerel stocks. Genetic samples have been taken over the entire distribution area of horse mackerel during the years 2015, 2016, and 2017, with a specific focus on the separation between horse mackerel in the western waters and horse mackerel in the North Sea. The results of the whole-genome sequencing indicated that the North Sea horse mackerel stock is clearly genetically different from the Western stock (Farrell and Carlsson, 2019; Fuentes-Pardo et al., 2020). Markers were identified that could distinguish with up to $95 \%$ accuracy between individuals collected in the North Sea and Western stocks. Follow-up work on this project is described in Section 6.7.

### 6.3.2 Mean weight at age and mean length at age

The mean annual weight and length over the period 2000-2021 are presented in Table 6.3.2 and Figures 6.3.3 and 6.3.4, respectively.

There do not seem to be strong differences over this period, although for the last two years the values at age 2-5 seem to be in decline and are relatively low.

### 6.3.3 Maturity-at-age

Peak spawning in the North Sea occurs in May and June (Macer, 1974), and spawning occurs in the coastal regions of the southern North Sea along the coasts of Belgium, the Netherlands, Germany, and Denmark.

There is no information available about the maturity-at-age of the North Sea Horse mackerel stock.

### 6.3.4 Natural mortality

There is no specific information available about natural mortality of this stock.

### 6.4 Data Exploration

### 6.4.1 Catch curves

The log-catch numbers were plotted by cohort to calculate the negative slope to get an estimate of total mortality ( $Z$ ). Fully selected ages 3 to $15+$ from the 1992-2009 period provide complete data for the 1992 to 2009 cohorts (Figure 6.4.1). The estimated negative slopes by cohort (Figure 6.4.2) indicate an increasing trend in total mortality up to the late 1990s, after which Z fluctuates from year to year. However, due to the low quality of the signals for some cohorts these Z estimates should be considered with caution.

An analysis of the catch number at age data carried out in 2011 showed that only the 1vs.2, 2vs.3, 7 vs .8 and 9 vs .10 age groups were positively and significantly correlated in the catch. This analysis has not been updated since, but these results suggest limitations in the catch-at-age data.

### 6.4.2 Assessment models and alternative methods to estimate the biomass

In 2002 Rückert et al. estimated the North Sea horse mackerel biomass based on a ratio estimate that related CPUE data from the IBTS to CPUE data of whiting (Merlangius merlangus). The applied method assumes that length specific catchability of whiting and horse mackerel are the same for the IBTS gear. Subsequently, they use the total biomass of whiting derived from an analytical stock assessment (MSVPA) to estimate the relationship between CPUE and biomass.

At the 2014 WGWIDE meeting exploratory model fits were attempted with the JAXass model, a simple statistical catch-at-age model fitted to an age-aggregated index of ( $2+$ ) biomass, total catch data and proportions at age from the catch. JAXass is based on Per Sparre's "separable VPA" model, an ad hoc method tested for the first time at WGWIDE in 2003, and later 2004. A new analysis using this model was also carried out in 2007 using an IBTS index. In 2014 the model has been coded in ADMB (Fournier et al., 2012) and updated with an improved objective function (dnorm), additional years of data and new methods for calculating the index (see above).

Difficulties in fitting an assessment model for this stock include:

- Unclear stock boundaries
- Difficulty aging horse mackerel
- Lack of strong cohort signals in catch-at-age data
- $\quad$ Scientific index derived from a survey not specifically designed for horse mackerel and not covering one of the main fishing grounds for the stock (7d)

Catches taken in 7d are close to the management boundary between the (larger) Western horse mackerel stock and the NS horse mackerel stock. It is quite possible that given changes in oceanographic conditions, or changes in abundance of either of the two stocks, that some proportion of the catches taken in 7d actually originated from the Western horse mackerel stock. Nevertheless, all assessment models used assume that $100 \%$ of fish caught in area 7 d belong to the North Sea horse mackerel stock. This is in agreement with stock and management definitions.

In 2018, the working group explored the Surplus Production model in Continuous Time (SPiCT) model for North Sea horse mackerel. SPiCT is one of the methods in the ICES guidelines to estimate MSY reference points for category 3 and 4 stocks (ICES, 2018). The model was run using the joint survey index as input or with separate survey indices (NS-IBTS and CGFS). The model with the joint survey index led to conflicting results with the perception of the stock, as biomass was estimated to be above BMSY and fishing mortality below FMSY. The model with two separate indices resulted in stock biomass and fishing mortality that were more in line with the perception
of the stock. However, there were strong retrospective patterns and wide confidence intervals in recent years. Furthermore, additional work is necessary on the setting of the priors, and on ensuring that model assumptions are not violated.

### 6.4.3 Survey data

### 6.4.3.1 Egg Surveys

No egg surveys for horse mackerel have been carried out in the North Sea since 1991. Such surveys were carried out during the period 1988-1991. SSB estimates are available historically. However, they were calculated assuming horse mackerel to be a determinate spawner. Horse mackerel is now considered an indeterminate spawner (Gordo et al. 2008). Therefore, egg abundance could only be considered a relative index of SSB. The Mackerel and Horse Mackerel Egg Surveys in the North Sea do not cover the spawning area of the North Sea horse mackerel stock.

### 6.4.3.2 North Sea International Bottom Trawl Survey

Many pelagic species are frequently found close to the bottom during daytime (which is when the North Sea IBTS survey operates) and migrate upwards predominantly during the night when they are susceptible to semi-pelagic fishing gear and to bottom trawls (Barange et al. 1998). Macer (1977) observed that dense shoals are formed close to the bottom during daytime, but the top of the shoals may extend into midwater. Eaton et al. (1983) argued that horse mackerel of 2 years and older are predominantly demersal in habit. Therefore, in the absence of a targeted survey for this stock, the NS-IBTS is considered a reasonable alternative.

NS-IBTS data from quarter 3 were obtained from DATRAS and analysed. Based on a comparison of NS-IBTS data from all 4 quarters in the period 1991 - 1996, Rückert et al. (2002) showed that horse mackerel catches in the NS-IBTS were most abundant in the third quarter of the year. In 2013 WGWIDE considered that using an 'exploitable biomass index' estimated with the abundance by haul of individuals of 20 cm and larger is the most appropriate for the purpose of interpreting trend in the stock.

To create indices, a subset of ICES statistical rectangles was identified. Rectangles that were not covered by the survey more than once during the period 1991 - 2012 were excluded from the index area. In 2012, WGWIDE expressed concern that the previously selected index area did not sufficiently cover the distribution area of the stock, especially in years that the stock would be relatively more abundant and spread out more. Rückert et al. (2002) also identified a larger distribution area of the North Sea stock. Based on the above, WGWIDE 2013 identified 61 rectangles to be included in the index area as shown in Figure 6.4.3.

### 6.4.3.3 French Channel Groundfish Survey

In order to improve data basis for the North Sea horse mackerel assessment, alternative survey indices have been explored. Previous indices only covered the North Sea distribution of the stock, while the majority of catches in recent years come from the eastern English Channel (7d). We evaluated the potential contribution of the French Channel Groundfish Survey (FR-CGFS) in 7d in quarter 4. The FR-CGFS has been carried out since 1990 and has frequent captures of horse mackerel. Although this survey is conducted in a different quarter to the NS-IBTS, the observed seasonal migration patterns of horse mackerel indicate that fish move into the Channel following quarter 3, so the timing is considered appropriate.

In 2015, the RV Gwen Drez was replaced by the RV Thalassa to carry out the FR-CGFS. In 2014 an inter-calibration process was conducted to quantify the differences in catchability for a large number of species. ICES reviewed this inter-calibration exercise and found a number of drawbacks that may undermine the reliability of the estimated conversion factors. The main concerns were:

- The analyses were limited in the number of tows. Considering that a number of these tows could be zeros for one of the two vessels and possibly resulting in highly uncertain estimates.
- Lack of length-specific correction factor.
- At a standardized depth of 50 m and above, wing spread estimates for the R/V Thalassa as measured by the MARPORT sensor were deemed erroneous, which may question the validity of estimated area swept by the net on the R/V Thalassa and the effect it may have on correction factors for species caught at depth at 50 m and greater.
- A number of tow locations including areas outside 7d were excluded. Changing the depth range of a survey can add serious bias in the calibration and the current approach seems to be ignoring this issue.
- Correction coefficients were not measured without error.

However, these limitations were considered by WGWIDE to be of minor importance for the North Sea horse mackerel since:

- Despite being still a low sample size the North Sea horse mackerel was present in all the 32 paired hauls.
- $\quad$ There are no important differences in size distribution (Figure 6.4.4).
- The analysis with and without the areas excluded in the new sampling design did not show important differences (ICES, 2017).
- CPUE of North Sea horse mackerel for hauls deeper than 50 m was relatively low (Figure 6.4.5), and it is expected than the potential problems in determining the conversion factor below that depth range would have a relatively minor impact in the estimated abundance.

For these reasons it was considered appropriate to continue using the FR-CGFS, standardizing the time-series of abundance for the period 1990-2015 with the estimated conversion factor 10.363.

### 6.4.3.4 Impact of Covid-19

Due to the Covid-19 pandemic and the lockdown in place in France at that time there was a delay in submitting the cruise application form for the FR-CGFS in 2020 to the French Foreign Ministry. The result was that no authorisation was provided in time to allow the survey to trawl within UK waters in 2020. Therefore, only French waters were sampled, meaning that only $70 \%$ of the core survey stations were completed (ICES, 2021b).

To assess the potential impact of missing UK stations in the FR-CGFS on the resulting abundance index for the exploitable stock, we tested the impact of
i) removing all UK sampling stations from the 1992-2019 time series,
ii) removing UK sampling stations from 2016-2019, one year at the time, and
iii) removing the FR-CGFS in 2016-2019, one year at the time, when modelling the abundance and calculating the index.

Removing all UK sampling stations from all years did not change the overall trend of the abundance index, but there were quite some deviations for individual years (Figure 6.4.6). Removing UK stations from one year at a time for 2016-2019 resulted in virtually no change for 2017 and 2018, but more apparent changes for 2016 and 2019 (Figure 6.4.7). Both these exercises suggest that basing the abundance index on NS-IBTS and French stations from FR-CGFS only may lead to different index values compared to when UK stations are included. The French sampling stations in the FR-CGFS only are thus not representative for the abundance of adult horse mackerel in the entire eastern Channel. As a further exploration, the abundance index was modelled by
leaving out the FR-CGFS entirely for 2019. However, the hurdle model was not able to run, and therefore a zero-inflated model was run instead. This model was considered to be the secondbest model during the benchmark process in 2017 and performed almost equally well as the hurdle model (ICES, 2017). Removing the FR-CGFS one year at a time for 2016-2019 resulted in minimal change for 2017 and 2018, but more apparent changes for 2016 and 2019 (Figure 6.4.8). Similar to (i) and (ii), leaving out the FR-CGFS may lead to different index values compared to when FR-CGFS is included.

As the investigations suggest that the missing UK stations from the FR-CGFS or leaving out the FR-CGFS entirely may lead to changes in the abundance index, it was decided that no reliable index value for 2020 could be produced. For the 2021 assessment, the approach of previous year was continued and thus no index value for 2020 was modelled. UK stations were visited in 2021 so the index value for 2021 was modelled according to the method with the hurdle model.

### 6.4.4 Length distributions from the surveys

The highest proportion of fish caught in 2021 were around 6-8 cm in the NS-IBTS (Figure 6.4.9). No group of strong year classes from previous years could be observed. In the FR-CGFS, the highest proportion of fish were between 10-11 cm, while in 2019, larger fish were dominating the catches (Figure 6.4.10). Despite that in 2020 the length frequencies are only based on French sampling stations, the length frequencies from 2021 shows a similar pattern A cluster of larger fish can be observed, possibly a remainder of the strong year class from 2020.

### 6.4.5 Length distributions from commercial catches

Currently, length distributions from catch data are available from 2016 to 2021. Future work is needed to retrieve historic length data in order to present a longer time series. The data used for the analysis come from the commercial catch sampling by national sampling programmes. For comparison, the analysis has also been run in the past with length data from the self-sampling programme of the Pelagic-Freezer-trawler Association (PFA), see for instance ICES $(2019,2020)$.

The length distributions based on the commercial catch data from 7d show a consistent distribution in time with a mean length between 22.15 and 22.5 cm each year, although with the exception of 24.7 cm in 2019 (Figure 6.4.11). Lengths in 4 c were on average 21.7 cm in 2019, 22.0 cm in 2020, and 23.8 in 2021 (Figure 6.4.12).

An error was found in the calculation of the length frequency distributions in the previous 2016 to 2020 assessments. Furthermore, the length frequency distribution calculated in 2019 included French data from only quarters 3 and 4, whereas data is also available for quarters 1 and 2 . The length frequency distributions for 2018-2020 were re-calculated using all available data, and upgraded code. No recalculation could be done for 2016-2017 since the original files are not available. Because for the calculation of the $\mathrm{F} / \mathrm{F}_{\text {msy }}$ proxy only the length frequency of the current assessment year is used, these older files are not needed for the calculations of advice anymore.

### 6.5 Stock assessment

### 6.5.1 Modelling the survey data

In January 2017, a benchmark of the North Sea horse mackerel assessment was conducted (ICES, 2017). Based on a capacity to model the over-dispersion and the high proportion of zero values in the survey catch data, a hurdle model was considered the best option of all model alternatives tested. The log-likelihood ratio test, AIC and the evidence ratio statistic supported that the model
that best represented the data was a hurdle model with Year and Survey as explanatory factors (including the interaction term) in the count model (GLM-negative binomial), and Year and Survey (without the interaction) in the zero model (GLM-binomial).

The probability of having a CPUE of zero was modelled by a logistic regression with a GLMbinomial distribution model:

$$
\operatorname{logit}\left(\pi_{i}\right)=\text { Intercept }_{z e r o}+\text { Year }_{i, z e r o}+\text { Survey }_{i, z e r o}
$$

where $\pi_{i}$ is the mean probability of having a CPUE of zero in haul $i$ as a function Year and Survey.
The expected CPUE of North Sea horse mackerel per haul $i$, conditional to not having a zero in hurdle models (not having a false zero in zero-inflated models), was modelled with a GLM-negative binomial distribution model:

$$
\log \left(C P U E_{i}\right)=\text { Intercept }_{\text {count }}+\text { Year }_{i, \text { count }} x \text { Survey }_{i, \text { count }}
$$

This model was used to synthesise the information from both the FR-CGFS and NS-IBTS and predict the average annual CPUE index as an indicator of trends in stock abundance. Separate models were fitted to the juvenile $(<20 \mathrm{~cm})$ and adult exploitable $(\geq 20 \mathrm{~cm})$ sub-stocks. The contribution of the two surveys to the combined index is weighted taken into consideration their respective area coverage as well as the mean wing spread. This index model allowed upgrading of the NSHOM to a category 3 stock within the ICES classification.

Similar to the 2019 assessment (ICES, 2019) ,2020 assessment (ICES, 2020), and 2021 assessment (ICES, 2021a), the model for the adult sub-stock that was run this year returned a warning despite the fact that the model converged. All parameter coefficients were estimated, but not the standard error for the intercept and the parameter $\theta$ of the count model. To check the robustness of the hurdle model with the warning, a zero-inflated model was run with the same set-up as the hurdle model. This zero-inflated model was considered to be the second-best model during the benchmark process in 2017 and performed almost equally well as the hurdle model (ICES, 2017). The fitted values of the zero-inflated model were very similar to that of the hurdle model with warning (Figure 6.5.1). The hurdle model from this year and its resulting index values where thus considered robust. Should the warning continue to occur in future assessments, additional testing and investigation should be conducted.

Due to the exclusion of the 2020 survey for modelling the abundance index, during the 2021 assessment the same time period (1992-2019) was used as during the 2020 assessment (ICES, 2020). In the current 2022 assessment the 2020 survey is still excluded. Since the last the 2021 assessment the updated abundance index resulted in a higher value for 2016 for the exploitable stock compared to last year (ICES 2021; Figure. 6.5.2). For each assessment, survey data from all years are extracted so that any underlying changes in the raw data stored in DATRAS are taken account of. Changes in reported raw HOM catches in 2016 in the NS-IBTS led to a higher mean catch rate of HOM (Figure 6.5.3), resulting in a higher abundance index value for 2016.

### 6.5.2 Summary of index trends and survey length distributions

The survey index for both the juvenile and exploitable sub-stock experienced a marked decline in the early 1990s and fluctuated at relatively low levels thereafter (Figures 6.5.4; Table 6.5.1). This reduction was partly due to the decline of the average abundance per haul over time, but also due to the increase of hauls with zero catch of the adult sub-stock (Figure 6.5.5). The survey index was at its third and second lowest in 2017 and 2018 (lowest in 2009), shows a slight increase again in 2019, but shows an average decline again in 2021, because of the low index of the NSIBTS (Figure 6.5.4).

The index trend for the juvenile sub-stock shows large fluctuations since 2015 (Figure 6.5.4). These are mainly attributed to the fluctuating trend of juveniles in the NS-IBTS (Figure 6.5.6), caused by some hauls with high catches of small horse mackerel in 2016 and 2018 (Figure 6.4.9). Fitted values for juveniles in the FR-CGFS show decreasing trend since 2014, but a slight increase again in 2019 and also in 2021 (Figure 6.5.6). The index of abundance of individuals $<20 \mathrm{~cm}$ could be considered a recruitment index, but future analyses should be carried out to study the correlation between the abundances and survey indices of year classes over time in more detail.

### 6.5.3 Length-based indicator and MSY proxy reference points

As part of the ICES approach to provide advice within the MSY framework for stocks of category 3 and 4, different Data Limited Stock (DLS) methods to estimate MSY proxy reference points (ICES, 2012, 2018) for the North Sea horse mackerel were previously explored (Pérez-Rodríguez, 2017). The Length Based Indicators analysis is the DLS method used in this assessment.

As most length samples and catches originate from 7d, length distributions from this area were used to calculate the MSY proxy. In 2021, the F/FMSY proxy based on the commercial catch samples indicated that fishing mortality was still slightly above $\mathrm{F}_{\mathrm{mSY}}$, with $\mathrm{L}_{\text {mean }} / \mathrm{LfF}_{\mathrm{F}} \mathrm{M}=22.86 \mathrm{~cm} /$ $24.33 \mathrm{~cm}=0.94$ (Figure 6.5.7).

Due to the recalculation of the 2018-2020 length distributions the F/FMSY ratios in those years changed from 0.927 to 0.932 for 2018, from 0.978 to 0.972 for 2019 , and from 0.927 to 0.912 from 2020. These revisions have no effect on the advice given in the 2021 assessment.

### 6.6 Basis for 2022 and 2023 Advice

Stock advice for North Sea horse mackerel is biennial. The NS-IBTS and FR-CGFS were modelled together to produce a joint abundance index for the exploitable part of the stock ( $\geq 20 \mathrm{~cm}$ ). No index value for 2020 could be produced. For this reason, the 2 -over- 3 rule applied to the index could only make use of index values from 2016 to 2019. The resulting index ratio (index value of 2019 over mean index value of 2016-2018) indicated that the adult sub-stock declined by $21 \%$. As the decline was more than $20 \%$, the uncertainty cap of 0.8 was applied to the catch advice. The $L_{m e a n} / \mathrm{Lf}=\mathrm{m}$ ratio in 2020 was 0.912 , indicating that the fishing mortality is above $\mathrm{Fmsy}^{\mathrm{ms}}$. Because the precautionary buffer was last applied in 2017, and thus more than three years ago, the buffer was applied once again in 2021. Under these circumstances, and based on the previous catch advice of 14014 t , ICES advised that catches of North Sea horse mackerel in 2022 and 2023 should be no more than 8969 t .

There are some signs of improved recruitment in some years (e.g. 2016, 2018 and to a lesser extend in 2021), but the trend of the abundance index for the juvenile sub-stock is fluctuating and, when separated, the two surveys, NS-IBTS and FR-CGFS, do not show the same trend. It remains to be seen if the weak signs of improved recruitment result in higher adult abundance. In 2019there was a slight increase in the index of the exploitable sub-stock but that trend has not continued this year.

### 6.7 Ongoing work

On behalf of the Pelagic Advisory Council and the EAPO Northern Pelagic Working Group, a research project on genetic composition of horse mackerel stocks was initiated in 2015 with University College Dublin (Ireland). Genetic samples have been taken over the whole distribution area of horse mackerel during the years 2015, 2016, and 2017, with a specific focus on the separation between horse mackerel in the western waters and horse mackerel in the North Sea. The
result of the research indicated that the western horse mackerel stock is clearly genetically different from the North Sea stock (Farrell and Carlsson, 2019; Fuentes-Pardo et al., 2020). Markers were identified that are able to reveal the stock identity of individual horse mackerel from potential mixing areas, namely Division 7.d, 7.e and 4.a. Following this, the Institute of Marine Research in Norway sampled horse mackerel in coastal waters within 4.a during all quarters in 2019. Preliminary results presented at WGWIDE 2021 showed that the genetic profile of individuals caught in all quarters matched well with the genetic profile of the Western HOM stock, with just one or two individuals matching better with North Sea HOM profile (Florian Berg, pers. comm.). More samples and research is needed to confirm these results. In another research project, horse mackerel from 7d and 7e have been collected by the PFA on board of commercial vessels in the Autumn of 2020, while during the same period horse mackerel from 4a have been collected during the NS-IBTS in Q3. The stock identity of the sampled fish will be investigated. The Norwegian research as well as the ongoing research described here may have large implications for stock delineation.

### 6.8 Management considerations

In the past, Division 7d was included in the management area for Western horse mackerel together with Divisions 2a, 7a-c, 7e-k, 8a, 8b, 8d, 8e, Subarea 6, EU and international waters of Division 5 b, and international waters of Subareas 12 and 14. ICES considers Division 7d now to be part of the North Sea horse mackerel distribution area. Since 2010, the TAC for the North Sea area has included Divisions 4.b, c and 7d. Considering that a majority of the catches are taken in Division 7d, the total North Sea horse mackerel catches are effectively constrained by the TAC since the realignment of the management areas in 2010.

Catches in Divisions 3a (Western Skagerrak) and 4a in quarters 3 and 4 are considered to be from the Western horse mackerel stock, while catches in quarters 1 and 2 are considered to be from the North Sea horse mackerel stock. Catches in area 4a and 3a are variable. In recent years only Norway has had significant catches in this area, but these are only taken in some years. Recent work suggest that all horse mackerel caught in 4 a belong to the Western stock, and ongoing genetic research on samples from 4 a and 7 d will shed more light on the proportions of the two stocks in catches from these areas.

### 6.9 Deviations from stock annex caused by missing information from Covid-19 disruption

1. Stock: hom.27.3a4bc7d
2. Missing or deteriorated survey data:

The assessment is based on two surveys, NS-IBTS and FR-CGFS. Due to the pandemic, trawling authorization in UK EEZ was not delivered in time, consequently FR-CGFS survey was not allowed to sample stations within UK waters in 2020.
3. Missing or deteriorated catch data:

Related to age sampling coverage was $56 \%$ and was covering only Q3, Q4 in areas 27.4.c and 27.7.d. Although most landed catch is taken from 27.7.d in Q 4 , other areas and quarters remain uncovered. Length sampling were impacted by the pandemic as samples were only available by two countries.
4. Missing or deteriorated commercial LPUE/CPUE data:

Not applicable
5. Missing or deteriorated biological data:

## Not applicable

6. Brief description of methods explored to remedy the challenge:

Effects of having only UK stations in FR-CGFS in all years or a single year, and excluding FRCGFS entirely for a single year on the combined survey index were investigated.
7. Suggested solution to the challenge, including reason for this selecting this solution:

Exploration methods suggested that leaving out UK stations or FR-CGFS entirely may affect the survey index and would lead to a survey index value not representative of stock abundance. It was therefore decided to produce no survey index value for 2020.
8. Was there an evaluation of the loss of certainty caused by the solution that was carried out?

The chosen solution affects the 2-over-3 rule by that only four instead of five index values can be used to assess the change in stock abundance. Like this year's assessment for 2022 and 2023, this will also affect the advice given in 2023 for 2024 and 2025.

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### 6.11 Tables

Table 6.3.1. North Sea Horse Mackerel stock. Catch in numbers (1000) by quarter and area in 2021


| 8 | 0.01 | 35.64 | 0.81 | 39.03 | 38.02 | 113.51 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0 | 0 | 0 | 0 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 5.78 | 0.13 | 6.33 | 6.16 | 18.41 |
| Sum | 0.33 | 1679.59 | 38.21 | 2296.48 | 1791.43 | 5806.05 |
| 3Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 49.1 | 206.53 | 141.39 | 397.02 |
| 2 | 0 | 0 | 141.3 | 783.18 | 406.9 | 1331.37 |
| 3 | 0 | 0 | 129.76 | 1256.95 | 373.67 | 1760.38 |
| 4 | 0 | 0 | 76.92 | 411.33 | 221.5 | 709.75 |
| 5 | 0 | 0 | 28.52 | 433.38 | 82.12 | 544.01 |
| 6 | 0 | 0 | 1.89 | 2.81 | 5.45 | 10.14 |
| 7 | 0 | 0 | 15.87 | 299.9 | 45.71 | 361.49 |
| 8 | 0 | 0 | 1.98 | 2.94 | 5.7 | 10.62 |
| 9 | 0 | 0 | 0.26 | 0.39 | 0.76 | 1.41 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0.63 | 0.93 | 1.81 | 3.37 |
| 12 | 0 | 0 | 1.27 | 1.89 | 3.67 | 6.83 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sum | 0 | 0 | 447.5 | 3400.23 | 1288.68 | 5136.39 |
| 4Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 5.45 | 191.78 | 6563.36 | 6760.59 |
| 2 | 0 | 0 | 15.53 | 312.84 | 18928.67 | 19257.04 |
| 3 | 0 | 0 | 13.81 | 1724.58 | 15381.03 | 17119.41 |


| 4 | 0 | 0 | 8.47 | 985.63 | 9504.59 | 10498.68 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 0 | 0 | 2.9 | 644.4 | 2949.8 | 3597.09 |
| 6 | 0 | 0 | 0.21 | 0.99 | 264.68 | 265.88 |
| 7 | 0 | 0 | 1.56 | 7.24 | 1931.81 | 1940.61 |
| 8 | 0 | 0 | 0.22 | 1.04 | 276.95 | 278.21 |
| 9 | 0 | 0 | 0.03 | 0.14 | 36.82 | 36.99 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0.07 | 0.33 | 88.05 | 88.45 |
| 12 | 0 | 0 | 0.14 | 0.67 | 178.29 | 179.1 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sum | 0 | 0 | 48.39 | 3869.64 | 56104.05 | 60022.05 |
| 1-4Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1 | 0 | 10.34 | 54.79 | 464.51 | 6732.43 | 7262.07 |
| 2 | 0.23 | 455.88 | 159.12 | 1596.64 | 19681.29 | 21893.16 |
| 3 | 4.2 | 7672.49 | 157.91 | 4567.27 | 19824.98 | 32226.85 |
| 4 | 1.85 | 3374.45 | 91.77 | 2192.75 | 11447.82 | 17108.63 |
| 5 | 2.7 | 4934.07 | 40.65 | 2655.48 | 5073.05 | 12705.95 |
| 6 | 0.35 | 633.57 | 3.29 | 156.14 | 581.87 | 1375.21 |
| 7 | 1.27 | 2329.5 | 21.87 | 874.27 | 3143.02 | 6369.94 |
| 8 | 0.24 | 443.81 | 3.03 | 171.22 | 440.72 | 1059.02 |
| 9 | 0 | 0 | 0.29 | 0.53 | 37.67 | 38.49 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0.7 | 1.26 | 90.07 | 92.03 |
| 12 | 0 | 0 | 1.42 | 2.56 | 182.39 | 186.37 |
| 13 | 0.04 | 71.97 | 0.13 | 11.02 | 41.62 | 124.78 |
| Sum | 10.88 | 19926.08 | 534.97 | 12693.65 | 67276.93 | 100442.5 |

Table 6.3.2. Numbers at age (millions). weight at age (kg) and length at age (cm) for the North Sea horse mackerel 1995-2021 in the commercial fleet catches (2018 distribution based on one sample only due to low sampling level).

| Catch |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | んૂ | 익 | 욱 | $\stackrel{\infty}{\circ}$ | 익 | Oi | O- | Nion | Ò | O | Nì | O O | Nì | $\underset{\sim}{\circ}$ | Oio | Oin | $\underset{\sim}{\underset{\sim}{7}}$ | Ṅ্N | $\underset{\sim}{\underset{\sim}{2}}$ | $\underset{\sim}{\underset{N}{N}}$ | $\stackrel{\sim}{\sim}$ | $\begin{aligned} & \text { O} \\ & \underset{N}{1} \end{aligned}$ | $\stackrel{\text { Ni}}{\sim}$ | $\underset{\sim}{\boldsymbol{\sim}}$ | 7 <br>  | 잉 | N |
| 0 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.10 |  |
| 1 | 1.8 | 4.6 | 12.6 | 2.3 | 12.4 | 77.2 | 2.8 | 71.4 | 13.5 | 12.8 | 13.8 | 5.8 | 3.1 | 1.8 | 34.2 | 2.9 | 7.3 | 9.5 | 7.6 | 15.4 | 49.7 | 3.6 | 20.7 | 27.4 |  | 4.75 | 7.26 |
| 2 | 3.1 | 13.8 | 27.2 | 22.1 | 31.5 | 75.8 | 15.6 | 21.4 | 41.9 | 14.7 | 73.7 | 26.5 | 15 | 8.9 | 14.2 | 22 | 19.5 | 24.3 | 9.9 | 15.3 | 23.8 | 65.2 | 20.9 | 49.1 | 1 | 58.99 | 21.89 |
| 3 | 7.2 | 11 | 14.1 | 36.7 | 23.1 | 29.2 | 32.5 | 24.5 | 24.2 | 27.8 | 39.8 | 58.8 | 22.3 | 36.2 | 29.1 | 9 | 72.1 | 20.3 | 21.2 | 8.7 | 10.1 | 15.9 | 62.6 | 13.2 | 5.8 | 24.57 | 32.23 |
| 4 | 10.3 | 11.9 | 14.9 | 38.8 | 17.6 | 24.1 | 16.8 | 13.9 | 35.7 | 14.7 | 32.6 | 40.1 | 82.9 | 16.8 | 22.6 | 15.3 | 36.4 | 40.1 | 22.1 | 30.2 | 5.8 | 9.8 | 10.2 | 32.7 | 3.1 | 21.02 | 17.11 |
| 5 | 12.1 | 9.6 | 14.6 | 20.8 | 23.1 | 33.3 | 8.3 | 6.6 | 30.3 | 28.1 | 16.4 | 67.6 | 72.5 | 36.3 | 17.4 | 25.2 | 13.4 | 25.7 | 27.1 | 13.8 | 7.2 | 7.7 | 6 | 4.5 | 12.1 | 2.24 | 12.71 |
| 6 | 13.2 | 12.5 | 12.4 | 12.1 | 26.2 | 21.5 | 10.8 | 6.5 | 11.4 | 21 | 19.8 | 23 | 31.2 | 36.1 | 17.1 | 15.9 | 9.2 | 20.8 | 6 | 7.1 | 3.8 | 5.7 | 3.4 | 0.7 | 2 | 4.00 | 1.38 |
| 7 | 11.4 | 8 | 10.1 | 14 | 20.6 | 16.8 | 15.7 | 6.2 | 7.3 | 10.9 | 4.4 | 12.6 | 24.3 | 27.2 | 22.3 | 29.4 | 3.5 | 3.1 | 7.2 | 2.7 | 3.3 | 2.5 | 2.8 | 0.7 | 1.8 | 0.42 | 6.37 |
| 8 | 12.6 | 6.6 | 8.6 | 10.8 | 21.8 | 8.7 | 12.9 | 11.6 | 10.7 | 4 | 2.4 | 5.6 | 17.6 | 21.9 | 46.8 | 5.1 | 6.8 | 5 | 4.2 | 3.4 | 1.4 | 5.1 | 2.4 |  | 1.3 | 0.09 | 1.06 |
| 9 | 7.3 | 1.5 | 2.5 | 8.3 | 12.9 | 9.5 | 9.9 | 7 | 11.8 | 5.8 | 2.1 | 1.2 | 8.5 | 10.1 | 11.1 | 5.6 | 3 | 4.6 | 4 | 0.9 | 1.6 | 1.2 | 0.9 |  | 4.7 | 0.28 | 0.04 |
| 10 | 5.9 | 5.3 | 0.8 | 4 | 8.2 | 5.6 | 10 | 7.7 | 1.3 | 12 | 3.5 | 1.3 | 1.7 | 7.5 | 9.3 | 8.6 | 8.4 | 1.5 | 5.4 | 1 | 0.9 | 0.1 | 0.3 |  | 2.8 | 0.35 | 0 |
| 11 | 0 | 0.3 | 0.3 | 2.7 | 2.1 | 4.6 | 9.6 | 5.9 | 3.6 | 6.7 | 3.8 | 0.1 | 1.1 | 1.9 | 7.1 | 2.6 | 6.1 | 0.5 | 3.7 | 1.3 | 0.2 | 0.1 | 0.5 |  | 3.5 | 0.27 | 0.09 |
| 12 | 8.8 | 1.3 | 0.3 | 0.7 | 0.4 | 1.7 | 5.4 | 3.4 | 3.1 | 5 | 1.3 | 1.5 | 0.2 | 2.1 | 3.6 | 0.3 | 3.5 | 0.1 | 1 | 0.4 | 0.9 | 0.4 | 0 |  | 0.7 | 0.13 | 0.19 |
| 13 | 0.2 | 8.9 |  | 1.8 | 1.4 | 1.2 | 3.7 | 2.4 | 2.4 | 6.8 | 2.7 | 0.4 | 0.8 | 0.4 | 0.3 | 0.3 | 0.4 | 0 | 0.6 | 0 | 0.2 | 1.4 | 0 |  | 0.7 | 0.03 | 0.12 |
| 14 | 4.4 | 8 | 1.4 | 0.3 | 3.8 | 0 | 2 | 1.4 | 3.4 | 2.5 | 2.1 | 0.8 | 0.8 | 2.4 | 1 | 0.2 | 0.3 | 0.2 | 0 | 0.2 | 0.2 | 0.5 | 0.3 |  | 0.3 | 0.03 |  |
| 15+ |  |  |  | 5.1 | 4 | 6.7 | 5.8 | 3 | 3.7 | 8.8 | 5.5 | 0.7 | 0 | 1.1 | 6.1 | 1.1 | 0.5 | 0 | 0.1 | 0.1 | 0 | 3.1 | 0.3 |  | 7.7 | 0.07 |  |


| $\stackrel{y}{8}$ | 슥 | ஆ육 | 숙 | かっ | 익 | O | O- | No | No | ষ্N | Ni | O O N | 술 | $\underset{\sim}{\circ}$ | Oio | 욱 | $\underset{\sim}{\underset{N}{7}}$ | $\underset{\sim}{\sim}$ | $\underset{\sim}{\underset{\sim}{2}}$ | $\underset{\sim}{\underset{\sim}{2}}$ | $\underset{\sim}{\sim}$ | $\begin{aligned} & 0 \\ & \underset{N}{N} \end{aligned}$ | 구N | $\stackrel{\infty}{\underset{\sim}{\sim}}$ | $\stackrel{7}{\mathrm{~N}}$ | O్N | N |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.04 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 7 |  |  |
| 1 | 0.07 | 0.10 | 0.06 | 0.06 | 0.06 | 0.07 | 0.06 | 0.06 | 0.07 | 0.07 | 0.07 | 0.07 | 0.06 | 0.06 | 0.07 | 0.07 | 0.06 | 0.06 | 0.07 | 0.07 | 0.06 | 0.07 | 0.06 | 0.06 |  | 0.05 | 0.04 |
|  | 6 | 7 | 3 | 3 | 3 | 5 | 7 | 6 | 4 | 6 |  | 4 | 15 | 3 | 4 | 7 |  | 9 | 7 | 8 | 2 |  |  | 1 |  | 7 | 6 |
| 2 | 0.12 | 0.12 | 0.10 | 0.10 | 0.10 | 0.1 | 0.09 | $\begin{aligned} & 0.09 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0.10 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0.10 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0.08 \\ & 7 \end{aligned}$ | $\begin{aligned} & 0.09 \\ & 8 \end{aligned}$ | $\begin{aligned} & 0.08 \\ & 1 \end{aligned}$ | $\begin{aligned} & 0.09 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0.08 \\ & 7 \end{aligned}$ | $\begin{aligned} & 0.10 \\ & 1 \end{aligned}$ | $\begin{array}{ll}0.09 & 0.09 \\ 2 & \end{array}$ |  | $\begin{array}{ll}0.09 & 0.11 \\ 9 & \end{array}$ |  | 0.099 | 0.093 | 0.08 | 0.09 | 0.11 | 0.08 | 0.05 |
|  | 6 | 3 | 2 | 2 | 2 |  |  |  |  |  |  |  |  |  |  |  |  |  | 6 | 3 |  |  | 1 | 6 | 7 |
| 3 | 0.12 | 0.14 | 0.12 | 0.12 | 0.12 | 0.13 | 0.09 | 0.12 | 0.12 | 0.12 | 0.10 | 0.11 | 0.10 | 0.10 | 0.11 | 0.11 | 0.09 | 0.11 |  |  | 0.11 | 0.11 | 0.13 | 0.11 | 0.11 | 0.13 | 0.12 | 0.11 | 0.08 |
|  | 5 | 3 | 6 | 6 | 6 | 7 | 4 | 9 | 2 | 2 | 4 | 6 | 4 | 9 | 3 | 8 | 6 | 8 | 2 | 3 | 5 | 3 |  | 1 | 5 | 9 | 6 |
| 4 | 0.13 | 0.15 | 0.14 | 0.14 | 0.14 | 0.15 | 0.11 | 0.15 | 0.13 | 0.14 | 0.13 | 0.12 | 0.11 | 0.12 | 0.13 | 0.13 | 0.11 | 0.14 | 0.13 | 0.13 | 0.15 | 0.126 | 0.131 | 0.147 | $\begin{aligned} & 0.15 \\ & 5 \end{aligned}$ | 0.132 | 0.119 |
|  | 3 | 6 | 2 | 2 | 2 | 2 | 7 | 5 | 6 | 6 | 3 | 4 | 5 | 5 | 4 | 7 | 5 | 2 | 8 | 5 |  |  |  |  |  |  |  |
| 5 | 0.14 | 0.17 | 0.16 | 0.16 | 0.16 | $\begin{aligned} & 0.16 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0.15 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.17 \\ & 1 \end{aligned}$ | $\begin{aligned} & 0.16 \\ & 4 \end{aligned}$ | $\begin{aligned} & 0.17 \\ & 4 \end{aligned}$ | $\begin{aligned} & 0.15 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.14 \\ & 1 \end{aligned}$ | 0.13 | $\begin{aligned} & 0.14 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0.15 \\ & 1 \end{aligned}$ | $\begin{aligned} & 0.15 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0.14 \\ & 5 \end{aligned}$ | $0.15$ | 0.166 | 0.144 | 0.16 | 0.15 | 0.17 | 0.17 | 0.16 | 0.19 | 0.13 |
|  | 6 | 7 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 9 | 8 | 3 |  | 5 | 1 | 2 |
| 6 | 0.16 | 0.18 | 0.17 | 0.17 | 0.17 | 0.19 | 0.18 | 0.19 | 0.18 | $\begin{aligned} & 0.19 \\ & 8 \end{aligned}$ | $\begin{aligned} & 0.19 \\ & 7 \end{aligned}$ | $\begin{aligned} & 0.17 \\ & 8 \end{aligned}$ | $\begin{aligned} & 0.16 \\ & 3 \end{aligned}$ | 0.16 | $\begin{aligned} & 0.18 \\ & 2 \end{aligned}$ | $\begin{aligned} & 0.18 \\ & 3 \end{aligned}$ | $\begin{aligned} & 0.16 \\ & 6 \end{aligned}$ | 0.17 | 0.18 | 0.177 | 0.196 | 0.155 | 0.189 | 0.189 | 0.202 | $0.17$ | $0.19$ |
|  | 4 | 7 | 5 | 5 | 5 | 2 | 3 | 5 |  |  |  |  |  |  |  |  |  | 2 |  |  |  |  |  |  |  |  |  |
| 7 | 0.16 | 0.20 | 0.19 | 0.19 | 0.19 | 0.19 | 0.19 | 0.21 | 0.19 | 0.22 | 0.23 | 0.21 | 0.19 | 0.19 | 0.19 | 0.20 | 0.19 | 0.18 | 0.2 | 0.18 | 0.26 | 0.16 | 0.17 | 0.20 | 0.26 | 0.24 | 0.17 |
|  | 1 | 3 | 9 | 9 | 9 | 4 | 8 | 6 | 3 | 4 | 8 | 2 | 2 | 3 | 5 | 6 | 3 | 3 |  | 4 |  | 2 | 7 | 1 | 1 | 7 |  |
| 8 | 0.17 | 0.19 | 0.23 | 0.23 | 0.23 | 0.21 | 0.20 | 0.22 | 0.21 | 0.22 | 0.24 | 0.24 | 0.19 | 0.22 | 0.25 | 0.19 | 0.19 | 0.18 | 0.21 | 0.20 | 0.29 | 0.23 | 0.18 |  | $\begin{aligned} & 0.24 \\ & 8 \end{aligned}$ | $\begin{aligned} & 0.30 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0.24 \\ & 7 \end{aligned}$ |
|  | 8 | 5 | 1 | 1 | 1 | 6 | 1 | 7 | 2 | 9 | 8 | 7 | 7 | 1 | 8 | 9 | 3 | 8 | 6 | 1 |  | 5 | 8 |  |  |  |  |
| 9 | 0.16 | 0.21 | 0.25 | 0.25 | 0.25 | 0.24 | 0.23 | 0.22 | 0.24 | $\begin{aligned} & 0.25 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0.25 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.23 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0.25 \\ & 7 \end{aligned}$ | $\begin{aligned} & 0.28 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0.25 \\ & 3 \end{aligned}$ | $\begin{aligned} & 0.24 \\ & 1 \end{aligned}$ | $\begin{aligned} & 0.30 \\ & 5 \end{aligned}$ | 0.212 | 0.223 | 0.222 | 0.265 | 0.24 | 0.22 |  | 0.26 | 0.30 | 0.30 |
|  | 5 | 8 |  |  |  | 4 | 7 | 8 |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.24 6 | 2 |  |  | 8 | 6 |
| 10 | 0.17 | 0.24 | 0.25 | 0.25 | 0.25 | 0.28 | 0.24 | 0.25 | 0.26 | 0.29 | 0.287 | 0.28 | 0.25 | 0.29 | 0.32 | 0.22 | 0.33 | 0.204 | 0.226 | 0.22 | 0.31 | 0.35 | 0.23 |  | 0.30 | 0.31 | 0.30 |
|  | 3 | 1 | 9 | 9 | 9 | 3 | 6 | 3 | 9 |  |  | 6 | 5 | 6 | 2 | 7 | 4 |  |  |  | 2 | 9 | 3 |  | 4 | 3 | 8 |


| $\underset{\sim}{\text { 品 }}$ | 슥 | ْ육 | 숙 | かㅇㄱ | 永 | O O | Oi | Oì | Ò O | O | ㅅN | O O N | ÒN | ষ্N | Oio | 음 | ت묵 | $\underset{\sim}{\sim}$ | $\underset{\sim}{\underset{\sim}{2}}$ | $\underset{\sim}{\underset{N}{N}}$ | $\stackrel{n}{\sim}$ | $\begin{aligned} & 0 \\ & \underset{N}{1} \end{aligned}$ | $\stackrel{\underset{N}{N}}{ }$ | $\stackrel{\infty}{\sim}$ | $\stackrel{7}{\mathrm{~N}}$ | 으N | － |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 11 | 0.31 | 0.30 | 0.3 | 0.3 | 0.3 | 0.28 | 0.26 | 0.30 | 0.24 | 0.3 | 0.33 | 0.23 | 0.51 | 0.27 | 0.42 | 0.28 | 0.34 | 0.27 | 0.24 | 0.26 | 0.26 | 0.36 | 0.25 |  | 0.30 | 0.33 | 0.31 |
|  | 7 | 7 |  |  |  | 6 |  | 3 |  |  | 5 | 7 | 7 | 3 | 2 | 4 | 5 | 5 | 2 | 4 | 2 | 9 | 7 |  | 1 | 9 | 3 |
| 12 | 0.23 | 0.21 | 0.32 | 0.32 | 0.32 | 0.35 | 0.28 | 0.29 | 0.29 | 0.29 | 0.34 | 0.26 | 0.27 | 0.30 | 0.44 | 0.23 | 0.40 | 0.19 | 0.26 | 0.28 | 0.31 | 0.37 |  |  | 0.41 | 0.37 | 0.33 |
|  | 3 | 1 | 9 | 9 | 9 | 4 | 6 | 3 | 8 | 7 | 9 | 1 | 9 | 9 | 7 | 4 | 8 | 5 | 3 | 7 | 8 | 9 |  |  | 1 | 8 | 9 |
| 13 | 0.24 | 0.25 | 0.36 | 0.36 | 0.36 | 0.31 | 0.28 | 0.31 | 0.35 | 0.30 | 0.33 | 0.26 | 0.33 | 0.37 | 0.38 | 0.28 | 0.47 |  | 0.26 | 0.25 | 0.35 | 0.24 |  |  | 0.42 | 0.32 | 0.37 |
|  | 1 | 8 | 7 | 7 | 7 | 6 | 7 | 7 | 6 | 1 | 8 | 7 | 9 | 5 | 3 | 8 | 4 |  | 2 | 2 | 1 | 2 |  |  |  | 5 | 8 |
| 14 | 0.34 | 0.27 | 0.29 | 0.29 | 0.29 |  | 0.29 | 0.32 | 0.31 | 0.33 | 0.37 | 0.30 | 0.41 | 0.27 | 0.36 | 0.31 | 0.41 | 0.18 | 0.55 | 0.40 | 0.23 | 0.39 | 0.21 |  | 0.42 | 0.38 | 0.32 |
|  | 8 | 7 | 9 | 9 | 9 |  | 5 |  | 6 | 8 | 3 | 2 | 4 | 7 | 2 | 5 | 5 | 7 | 9 | 8 | 5 |  | 4 |  | 9 | 9 | 5 |
| 15 | 0.34 | 0.27 | 0.36 | 0.36 | 0.36 | 0.35 | 0.33 | 0.39 | 0.35 | 0.40 | 0.37 | 0.40 | 0 | 0.38 | 0.46 | 0.35 | 0.47 | 0 | 0.33 | 0.27 | 0 | 0.37 | 0.26 |  | 0.43 | 0.37 | 0.38 |
| ＋ | 8 | 7 |  |  |  |  | 6 |  | 3 | 2 | 5 | 4 |  | 9 |  | 1 | 5 |  | 9 | 3 |  | 8 |  |  | 1 |  | 9 |
| cm | length |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\underset{\substack{\infty \\ \multirow{2}{*}{\hline}\\ \hline}}{ }$ | 늑 | 익 | 숙 | か | 익 | Oi | -ì | Ò | Ò | ষ্N | No | O O | 우 | Oi | Oio | Oi | $\underset{\sim}{7}$ | $\begin{gathered} \text { N } \\ \text { N } \end{gathered}$ | $\stackrel{m}{\sim}$ | $\underset{N}{\underset{N}{N}}$ | $\stackrel{n}{\sim}$ | $\begin{aligned} & 0 \\ & \underset{\sim}{1} \end{aligned}$ | $\stackrel{N}{N}$ | $\stackrel{\infty}{\sim}$ | $\underset{\sim}{7}$ | Ni | N |
| 0 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 18 |  |
| 1 | 19.2 | 19.2 | 19.2 | 19.2 | 19.2 | 19.1 | 19.5 | 19.4 | 20.3 | 19.8 | 18.1 | 20.1 | 19.9 | 20 | 20.3 | 20.8 | 19.2 | 19.9 | 20.9 | 20.4 | 19.8 | 20 | 19.1 | 19.51 |  | 19.3 | 20 |
| 2 | 22 | 22 | 22 | 22 | 22 | 21.5 | 21.5 | 21.7 | 22.3 | 22.2 | 21.5 | 22 | 20.8 | 21.6 | 21.6 | 22.6 | 21.7 | 21.7 | 22.4 | 22.9 | 22.9 | 22 | 21.3 | 22.19 | 23.5 | 21.7 | 21.2 |
| 3 | 23.5 | 23.5 | 23.5 | 23.5 | 23.5 | 23.9 | 21.9 | 23.8 | 23.7 | 23.6 | 22.9 | 23.4 | 22.6 | 23.2 | 23.2 | 23.9 | 23 | 23.5 | 23.5 | 23.5 | 24.6 | 23.6 | 23.3 | 24.67 | 24.4 | 24.3 | 22.9 |
| 4 | 24.8 | 24.8 | 24.8 | 24.8 | 24.8 | 24.9 | 23.4 | 25.4 | 24.6 | 25.2 | 24.7 | 24.1 | 23.6 | 24.1 | 24.6 | 25 | 24.5 | 25 | 25.3 | 24.8 | 25.8 | 24.8 | 24.1 | 25.58 | 26.1 | 24.8 | 24.5 |
| 5 | 25.5 | 25.5 | 25.5 | 25.5 | 25.5 | 26 | 26.7 | 26.3 | 26.2 | 26.6 | 25.9 | 25.4 | 24.4 | 25.6 | 25.8 | 25.7 | 25.9 | 25.7 | 27 | 25.4 | 26.6 | 26.4 | 26.7 | 26.78 | 26.6 | 27.5 | 25.4 |
| 6 | 26.4 | 26.4 | 26.4 | 26.4 | 26.4 | 27.6 | 27.5 | 27.4 | 27.3 | 27.5 | 27.7 | 27 | 26.6 | 26.3 | 27.2 | 27.1 | 27.6 | 27 | 27.1 | 27.3 | 28.2 | 26.1 | 27.5 | 27.5 | 28.1 | 27 | 25.4 |
| 7 | 27.2 | 27.2 | 27.2 | 27.2 | 27.2 | 28.1 | 28.1 | 28.6 | 28.2 | 28.8 | 29.8 | 28.6 | 27.9 | 28.1 | 28.1 | 28.3 | 27.7 | 27.1 | 28.3 | 27.5 | 30.4 | 27.5 | 27.5 | 28.04 | 30.6 | 29.5 | 28 |


| 品 | 슥 | ஃ육 | 욱 |  | 익 | O | -i | Oì | ÒN | $\underset{\sim}{O}$ | NiO | O O N | 人̀N | OiN | O | Oì | $\underset{\sim}{\mathrm{N}}$ | $\underset{\sim}{\sim}$ | $\underset{\sim}{\underset{\sim}{2}}$ | $\underset{\sim}{\underset{\sim}{N}}$ | $\stackrel{n}{\sim}$ | $\begin{aligned} & 0 \\ & \underset{\sim}{1} \end{aligned}$ | $\underset{\sim}{\underset{N}{2}}$ | $\underset{\sim}{\infty}$ | $\underset{\sim}{\circ}$ | 잉 | N |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 8 | 29.2 | 29.2 | 29.2 | 29.2 | 29.2 | 28.6 | 28.5 | 29.3 | 29 | 29.2 | 30.4 | 29.8 | 28.1 | 28.8 | 30.6 | 28.4 | 27.8 | 27.1 | 28.9 | 28 | 31.7 | 30.2 | 28 |  | 30 | 31.8 | 27.3 |
| 9 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 | 29.9 | 29.8 | 29.4 | 29.9 | 30.4 | 30.8 | 30.8 | 30.1 | 31.2 | 31.1 | 30.2 | 31.9 | 28.6 | 29.2 | 28.8 | 30.5 | 30.5 | 29.1 |  | 30.6 | 31.6 | 29.5 |
| 10 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 | 31.2 | 30.2 | 30.3 | 30.9 | 31.4 | 31.8 | 31.5 | 31 | 31.8 | 32.5 | 30 | 32.5 | 28 | 29.5 | 29.2 | 32.5 | 34.7 | 29.5 |  | 32.1 | 32.1 |  |
| 11 | 30.6 | 30.6 | 30.6 | 30.6 | 30.6 | 31.5 | 30.7 | 31.4 | 30.7 | 31.9 | 33.8 | 31.2 | 39.5 | 31.6 | 35 | 32.2 | 33.2 | 30.1 | 30 | 30.7 | 31.5 | 35.2 | 31.1 |  | 32.1 | 32.7 | 30.5 |
| 12 | 32.1 | 32.1 | 32.1 | 32.1 | 32.1 | 33.6 | 32 | 31.6 | 31.9 | 31.7 | 35.6 | 30.8 | 31.5 | 32.2 | 35.3 | 30.8 | 34.6 | 27.5 | 30.4 | 30.6 | 32.3 | 35.5 |  |  | 36 | 34.4 | 34.5 |
| 13 | 33.3 | 33.3 | 33.3 | 33.3 | 33.3 | 33.3 | 31.7 | 32.4 | 32.8 | 31.9 | 34 | 32.1 | 33.4 | 33.9 | 34 | 31.8 | 36.4 |  | 32.1 | 30 | 32.5 | 31.5 |  |  | 36.3 | 31.5 | 30.5 |
| 14 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 |  | 32.1 | 32.4 | 32.6 | 33 | 34.4 | 32.5 | 34.5 | 32.3 | 34.2 | 33 | 36 | 27.5 | 38.5 | 36 | 30.5 | 36.1 | 30.5 |  | 36.6 | 35.5 |  |
| 15+ | 32.5 | 32.5 | 32.5 | 32.5 | 32.5 | 33.8 | 33.4 | 34.3 | 33.6 | 34.8 | 35.2 | 35.3 |  | 35.1 | 36.1 | 34.5 | 36.9 |  | 34.2 | 32.5 |  | 36.1 | 31.5 |  | 36.5 | 33 |  |

Table 6.3.3. North Sea Horse Mackerel stock. Mean weight at age (grams) in the catch by area for all quarters in 2020

Q1-Q4

| Ages | 27.3.a (Q1.2) | 27.4.a(Q1.2) | 27.4.b | 27.4.c | 27.7.d | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 46.5 | 46.5 | 46.5 | 46.5 | 46.5 | 46.5 |
| 1 | 57.4 | 57.4 | 57.4 | 60.3 | 57.1 | 57.4 |
| 2 | 86 | 86 | 86 | 87.7 | 85.8 | 86.0 |
| 3 | 119 | 119 | 119 | 112.2 | 119.8 | 119.0 |
| 4 | 132.1 | 132.1 | 132.1 | 130.8 | 132.2 | 132.1 |
| 5 | 191 | 191 | 191 | 175.7 | 193.1 | 191.0 |
| 6 | 173.6 | 173.6 | 173.6 | 165.4 | 174.6 | 173.6 |
| 7 | 246.7 | 246.7 | 246.7 | 208.7 | 253 | 246.8 |
| 8 | 305.7 | 305.7 | 305.7 | 305.7 | 305.7 | 305.7 |
| 9 | 307.8 | 307.8 | 307.8 | 307.8 | 307.8 | 307.8 |
| 10 | 313.3 | 313.3 | 313.3 | 313.3 | 313.3 | 313.3 |
| 11 | 339 | 339 | 339 | 339 | 339 | 339 |
| 12 | 378.1 | 378.1 | 378.1 | 378.1 | 378.1 | 378.1 |
| 13 | 325 | 325 | 325 | 325 | 325 | 325 |
| 14 | 389 | 389 | 389 | 389 | 389 | 389 |
| 15+ | 370 | 370 | 370 | 370 | 370 | 370 |

Table 6.3.4. North Sea Horse Mackerel stock. Mean length (mm) at age in the catch by area for all quarters in 2020

1-4Q

| Ages | 27.3.a (Q1.2) | 27.4.a(Q1.2) | 27.4.b | 27.4.c | 27.7.d | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 180 | 180 | 180 | 180 | 180 | 180 |
| 1 | 192.9 | 192.9 | 192.9 | 193 | 192.9 | 192.9 |
| 2 | 217.4 | 217.4 | 217.4 | 217.1 | 217.5 | 217.4 |
| 3 | 243.5 | 243.5 | 243.5 | 235.6 | 244.4 | 243.5 |
| 4 | 248.3 | 248.3 | 248.3 | 245.9 | 248.5 | 248.2 |
| 5 | 275.4 | 275.4 | 275.4 | 266 | 276.7 | 275.4 |
| 6 | 269.7 | 269.7 | 269.7 | 264.3 | 270.3 | 269.6 |
| 7 | 295 | 295 | 295 | 281.3 | 297.2 | 295.0 |


| Ages | 27.3.a (Q1.2) | 27.4.a(Q1.2) | 27.4.b | 27.4.c | 27.7.d | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 8 | 318.3 | 318.3 | 318.3 | 318.3 | 318.3 | 318.3 |
| 9 | 315.7 | 315.7 | 315.7 | 315.7 | 315.7 | 315.7 |
| 10 | 321 | 321 | 321 | 321 | 321 | 321 |
| 11 | 327.3 | 327.3 | 327.3 | 327.3 | 327.3 | 327.3 |
| 12 | 343.9 | 343.9 | 343.9 | 343.9 | 343.9 | 343.9 |
| 13 | 315 | 315 | 315 | 315 | 315 | 315 |
| 14 | 355 | 355 | 355 | 355 | 355 | 355 |
| 15+ | 329.7 | 329.7 | 329.7 | 329.7 | 329.7 | 329.7 |

Table 6.5.1. North Sea Horse Mackerel. CPUE Indices of abundance (number/hour) for the juvenile (<20cm) and exploitable ( $\geq \mathbf{2 0} \mathrm{cm}$ ) sub-stock. estimated as a combined index for the NS-IBTS Q3 and the FR-CGFS in Q4. The survey indices are derived from the prediction of a hurdle model fit to data over the period 1992-2021 and include a 95\% confidence interval based on a bootstrapping procedure (CI_low = lower bound. CI_high = upper bound). Survey data from 2020 were not included in the modelling procedure as not all sampling stations of FR-CGFS could be visited in 2020, and therefore no reliable index value for $\mathbf{2 0 2 0}$ could be calculated.

|  | Juvenile sub-stock (<20 cm) |  |  | Exploitable sub-stock ( $\mathbf{2 0} \mathbf{~ c m}$ ) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Index | Cl_low | Cl_high | Index | CI_low | Cl_high |
| 1992 | 4268 | 2077 | 9435 | 4268 | 2077 | 9435 |
| 1993 | 1854 | 894 | 3675 | 1854 | 894 | 3675 |
| 1994 | 2589 | 1291 | 5153 | 2589 | 1291 | 5153 |
| 1995 | 2016 | 1049 | 4220 | 2016 | 1049 | 4220 |
| 1996 | 730 | 320 | 1598 | 730 | 320 | 1598 |
| 1997 | 2171 | 928 | 4715 | 2171 | 928 | 4715 |
| 1998 | 650 | 323 | 1270 | 650 | 323 | 1270 |
| 1999 | 1438 | 755 | 2593 | 1438 | 755 | 2593 |
| 2000 | 1566 | 813 | 3110 | 1566 | 813 | 3110 |
| 2001 | 2188 | 1148 | 4925 | 2188 | 1148 | 4925 |
| 2002 | 2398 | 1282 | 4992 | 2398 | 1282 | 4992 |
| 2003 | 1787 | 933 | 3234 | 1787 | 933 | 3234 |
| 2004 | 1002 | 522 | 1837 | 1002 | 522 | 1837 |
| 2005 | 802 | 434 | 1590 | 802 | 434 | 1590 |
| 2006 | 530 | 278 | 918 | 530 | 278 | 918 |
| 2007 | 600 | 324 | 1099 | 600 | 324 | 1099 |


| 2008 | Juvenile sub-stock (<20 cm) |  |  | Exploitable sub-stock ( $\mathbf{2 0} \mathbf{~ c m}$ ) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 532 | 287 | 932 | 532 | 287 | 932 |
| 2009 | 691 | 356 | 1227 | 691 | 356 | 1227 |
| 2010 | 2257 | 1105 | 4387 | 2257 | 1105 | 4387 |
| 2011 | 504 | 263 | 933 | 504 | 263 | 933 |
| 2012 | 315 | 158 | 669 | 315 | 158 | 669 |
| 2013 | 1087 | 573 | 2080 | 1087 | 573 | 2080 |
| 2014 | 1545 | 835 | 2979 | 1545 | 835 | 2979 |
| 2015 | 1471 | 735 | 3035 | 1471 | 735 | 3035 |
| 2016 | 3044 | 1460 | 6036 | 3044 | 1460 | 6036 |
| 2017 | 941 | 427 | 1948 | 941 | 427 | 1948 |
| 2018 | 3171 | 1594 | 7299 | 3171 | 1594 | 7299 |
| 2019 | 812 | 375 | 1628 | 812 | 375 | 1628 |
| 2020 | - | - | - | - | - | - |
| 2021 | 1663 | 785 | 3309 | 1663 | 785 | 3309 |

### 6.12 Figures



Figure 6.2.1. North Sea horse mackerel. Utilisation of quota from 2000 to 2021.


Figure 6.2.2. North Sea horse mackerel. Utilisation of quota by country in 2021.

North Sea Stock: Catch by division


Figure 6.2.3. North Sea horse mackerel. Catch in (1000 t) by division and year from 1982 to 2021.


Figure 6.2.4. North Sea horse mackerel. Proportion of catches by ICES division from 2000 to 2021.


Figure 6.2.5. North Sea Horse Mackerel. Total catch (in tonnes) by ICES division, quarter, catch category and country in 2021.


Figure 6.3.1. North Sea horse mackerel age distribution in the catch for 1995-2021. The size of bubbles is proportional to the catch number. Note that age 15 is a plus $g$

NSHM: catch at age ( N ; observed) 27.7.d


NSHM: catch at age ( N ; observed) out of 27.7.d


Figure 6.3.2. North Sea horse mackerel. Bubble plots of age distribution in the catch by area for 1998-2021 for area 7.d (upper panel) and without 7.d (bottom panel). The size of bubbles is proportional to the catch numbers. Note that age 15 is a plus group.

## Mean weight at age (kg)



Figure 6.3.3. North Sea horse mackerel. Mean weight at age in commercial catches over the period 2000-2021. Note that only age 1-10 are presented and that 10 is not a plus group.

Mean length at age (cm)


Figure 6.3.4. North Sea horse mackerel. Mean length at age in commercial catches over the period 2000-2021. Note that only age 1-10 are presented and that 10 is not a plus group.


Figure 6.4.1. North Sea Horse Mackerel. Catch curves for the 1992 to 2009 cohorts, ages from $\mathbf{3}$ to $\mathbf{1 5 +}$. Values plotted on the vertical axis are the log(catch) values for each cohort in each year. The negative slope of these curves estimates total mortality $(Z)$ in the cohort.

Total mortality by cohort


Figure 6.4.2. North Sea Horse Mackerel. Total mortality by cohort (Z) estimated from the negative gradients of the 19922009 cohort catch curves (Figure 6.4.1).


Figure 6.4.3. North Sea horse mackerel. ICES rectangles selected by WGWIDE in 2013 and currently used by the working group.


Figure 6.4.4. North Sea horse mackerel. Size distribution of North Sea horse mackerel catches during the inter-calibration exercise conducted in 2014 between the RV Gwen Drez (red bars) and Thalassa (blue bars).


Figure 6.4.5. North Sea horse mackerel. CPUE by depth for the CGFS survey from 1992 to 2017.


Figure 6.4.6. North Sea horse mackerel. Modelled abundance index from 1992-2019 including both UK and French stations in the FR-CGFS (blue) and excluding UK stations in the FR-CGFS (red) for the exploitable sub-stock ( $\mathbf{2 0} \mathbf{~ c m}$ ).


Figure 6.4.7. North Sea horse mackerel. Modelled abundance index from 1992-2019 for the exploitable sub-stock ( $\geq 20$ cm ) for when UK sampling stations from FR-CGFS have been excluded for 2016 (top left), 2017 (top right), 2018 (bottom left) and 2019 (bottom right).


Figure 6.4.8. North Sea horse mackerel. Modelled abundance index from 1992-2019 for the exploitable sub-stock ( $\geq 20$ cm ) for when the FR-CGFS has been excluded for 2016 (top left), 2017 (top right), 2018 (bottom left) and 2019 (bottom right).


Figure 6.4.9. North Sea horse mackerel. Relative occurrence by length for the period 2014-2021 in the NS-IBTS.


Figure 6.4.10. North Sea horse mackerel. Relative occurrence by length for the period 2015-2021 in the FR-CGFS. Note that stations in UK waters could not be visited in 2020.

NSHM length frequency catches 27.7.d


Figure 6.4.11. North Sea horse mackerel. Length distributions in proportion to catch numbers from commercial catches in 27.7.d for the period 2016-2021.

NSHM length frequency catches 27.4.c


Figure 6.4.12. North Sea horse mackerel. Length distributions in proportion to catch numbers from commercial catches in 27.4.c in 2019 and 2021.


Figure 6.5.1. North Sea horse mackerel. CPUE per year of the exploitable sub-stock ( $\geq 20 \mathrm{~cm}$ ) from 1992 to 2021 as modelled by the hurdle model (red) that returned a warning when ran, and the zero-inflated model (grey).


Figure 6.5.2. North Sea horse mackerel. CPUE per year of the exploitable sub-stock ( $\geq 20 \mathrm{~cm}$ ) from 1992 to 2019 as modelled by the hurdle model at WGWIDE 2020 (grey) and WGWIDE 2021 (red). In the model the 2020 index value for the exploitable sub-stock ( $\mathbf{2 0} \mathbf{~ c m}$ ) is left out and the index is modelled using a Hurdle model. Complete overlap between last and current year is observed.

Exploitable stock


Figure 6.5.3. North Sea horse mackerel. Mean CPUE across hauls of the exploitable sub-stock ( $\geq 20 \mathrm{~cm}$ ) from 1992 to 2019 for the FR-CGFS (blue WGWIDE 2020 (not visible), grey WGWIDE 2021) and the NS-IBTS (black WGWIDE 2020, red WGWIDE 2021). Small changes in reported catches of NS-IBTS resulted in a higher index value for 2016.


Figure 6.5.4. North Sea Horse Mackerel. Joint CPUE survey index (number/hour) derived from the hurdle model fit to the NS-IBTS survey in the North Sea and the FR-CGFS survey in the Eastern English channel for the period 1991-2021. No index value for 2020 could be produced due to sampling issues in the FR-CGFS. Top: exploitable sub-stock ( $\geq 20 \mathrm{~cm}$ ), bottom: juvenile sub-stock ( $<20 \mathrm{~cm}$ ). Red shaded area represent the $95 \%$ confidence interval, which is determined by bootstrap resampling of Pearson residuals with 999 iterations.


Figure 6.5.5. North Sea horse mackerel. Proportion of hauls with zero catch for the exploitable ( $\mathbf{2 0} \mathbf{c m}$ ) and juvenile (<20 cm ) sub-stocks in the NS-IBTS (blue) and the FR-CGFS (red) from 1992 to 2021. Note that the FR-CGFS 2020 values are based on French stations only, as UK stations could not be sampled.


Figure 6.5.6. North Sea Horse Mackerel. Mean CPUE survey index (number/hour) obtained from the hurdle model fit to the NS-IBTS survey in the North Sea (in red), the FR-CGFS survey in the English channel (in grey) and the joint survey index (in blue). Top: exploitable sub-stock ( $\geq 20 \mathrm{~cm}$ ), bottom: juvenile sub-stock ( $<20 \mathrm{~cm}$ ). No index values for 2020 could be produced due to COVID-19 pandemic impacting the FR-CGFS.


Figure 6.5.7. Length distribution ( cm ), estimated parameters $L_{c}, L_{\text {mean }}, L_{F=M}(c m)$ and $F / F_{M S Y}$ ratio for 2016-2021. Length samples from commercial catches in ICES division 27.7.d. Recalculations for 2018 till 2020 have been performed for the constructions of these plots.

## 7 Western Horse Mackerel -in Subarea 8 and divisions 2.a, 3.a (Western Part), 4.a, 5.b, 6.a, 7.a-c and 7.e-k

### 7.1 TAC and ICES advice applicable to 2021 and 2022

Since 2011, the TACs cover areas in line with the distribution areas of the stock.
For 2021 the TAC was the following (EU 2021/1239):

| Areas | TAC 2021 | Stocks fished in this area |
| :--- | :--- | :--- |
| 2.a, 4.a, 5.b, 6, 7.a-c, 7.e-k, 8.abde, 12, 14 | 70254 t | Western stock \& North Sea stock in 4.a 1-2 <br> quarters |
| 4.b,c, 7.d | 14014 t | North Sea stocks |
| Division 8.c | 11121 t | Western stock |

For 2022 the TAC was the following (EU 2022/109):

| Areas | TAC 2022 | Stocks fished in this area |
| :--- | :--- | :--- |
| 2.a, 4.a, 5.b, 6, 7.a-c, 7.e-k, 8.abde, 12, 14 | 49178 t | Western stock \& North Sea stock in 4.a 1-2 <br> quarters |
| 4.b,c, 7.d | 3461 t | North Sea stocks |
| Division 8.c | 2780 t | Western stock |

The TAC for the western stock should apply to the distribution area of western horse mackerel as follows:

All Quarters: 2.a, 5.b, 6.a, 7.a-c, 7.e-k, 8.a-e
Quarters 3\&4: 3.a (west), 4.a
The TAC for the North Sea stock should apply to the distribution area of North Sea horse mackerel as follows:

All Quarters: 3.a (east), 4.b-c, 7.d
Quarters 1\&2: 3.a (west), 4.a
In 2021, ICES advised on the basis of MSY approach that Western horse mackerel catches in 2022 should be no more than 71138 tonnes. The Western horse mackerel TAC for 2022 is 71138 tonnes. The TAC should apply to the total distribution area of this stock. The horse mackerel catches in Division 3.a are taken outside the horse mackerel TACs.

### 7.1.1 The fishery in 2021

Information on the development of the fisheries by quarter and division is shown in Tables 5.1.1 and 5.1.2 and in Figures 5.1.1.a-d. The total catch allocated to Western horse mackerel in 2021 was 81375 tonnes which is 421 tonnes less than in 2020 and matched the ICES advice. The catches of horse mackerel by country and area are shown in Tables 7.1.1.1-5 while the catches by quarter since 2000 are shown in Figure 7.1.1.1.

### 7.1.2 Estimates of discards

Discard data are available since 2000 for some countries. Prior to 2013, the estimates available are considered to be an underestimate (Figure 7.1.2.1).

In 2021, most countries have submitted discard information. Countries that reported discard estimates for horse mackerel were Spain, France, Ireland, and UK). 2021 discard estimates for Germany, the Netherlands and Norway are considered to be equal to zero. Total discards for Western horse mackerel were 1641 tonnes, equal to $2 \%$ in weight of the total catches, a decrease in comparison to last year.

Discard data are included in the assessment as part of the total catches.
Length frequency distributions of discards were provided by Spain, France, Ireland, and UK but are not included in the assessment length-frequency input data.

### 7.1.3 Stock description and management units

The Western horse mackerel stock is distributed in divisions 2.a, 5.b, 3.a, 4.a, 6.a, 7.a-c, 7.e-k and 8.a-e (for more details see Section 5.3 and Figure 7.1.3.1) and spawns in the Bay of Biscay, and in UK and Irish waters before parts of the stock migrate northwards into the Norwegian Sea and the North Sea where they are fished in the third and fourth quarter (note for area 4.a, only catches taken in quarters 3 and 4 are considered to be from the western stock). The western stock is considered a management unit and advised accordingly with TAC set in accordance with the distribution of the stock (note that catches in division 3.a are taken outside the TAC).

### 7.2 Scientific data

### 7.2.1 Egg survey estimates

In 2022 a new egg survey was carried out in the western and southern spawning areas and a working document with preliminary results of the survey was presented to WGWIDE members (WD07: O'Hea et al. 2022). Final results for 2022 will be available for the WGWIDE meeting 2023. Details of this mackerel and horse mackerel egg survey are also given in section 8.6.1 of this report. The 2022 egg survey results are not used in the 2022 assessment (the SS model is not configured to use in-year survey data and catch data would also be required, which is not available until the end of the year). Data from the 1992-2019 surveys are therefore used in the assessment (O'Hea et al. 2019). The time series of TAEP estimates used in the assessment is shown in table 7.2.1.1.

## 2022 (preliminary) egg survey results

Sampling was undertaken over 6 sampling periods. Egg abundance plots displaying the spatial distribution of stage 1 horse mackerel eggs are presented for periods 3-7 in figures 7.2.1.1-5. Period number and duration are the same as those used to estimate the western mackerel stock, as are the dates defining the start and end of spawning. In general, egg numbers were low with occasional high counts are some stations.
During the 2022survey, peak spawning for horse mackerel occurred during period 6 with high egg numbers recorded from the Celtic Sea to the northern Bay of Biscay, close to the 200 m contour (figure 7.2.1.5).

Egg production by survey period (plotted at mid-period) is shown in figure 7.2.1.6. The results from previous surveys are also shown for comparison. The shape of the egg production curve does not suggest that spawning start and end dates should be altered for the 2022 survey. Total annual egg production for 2022 is estimated to be $5.15 \times 10^{14}$, a threefold increase on the 2019 estimate of $1.78 \times 10^{14}$, which is the lowest estimate in the time series.

## Fecundity parameters

In 2022, only DEPM ovary samples were collected in periods 6 and 7 for horse mackerel, during peak of spawning. In addition, samples were collected from the Irish WESPAS surveys in periods 6 and 7. At the time of writing, no horse mackerel fecundity results are available. All samples will be analysed and results presented at the 2023 WGMEGS meeting.

The Western horse-mackerel egg data of the DEPM survey are still under revision. Data are expected to be analysed and results will be presented at the 2023 WGMEGS.

### 7.2.2 Other surveys for Western horse mackerel

## Bottom-trawl surveys

A bottom-trawl survey index for recruitment was available for 2021. The recruitment index is based on IBTS surveys conducted by Ireland, France and Scotland covering the main distribution of the stock (Bay of Biscay, Celtic Sea, West of Ireland and West of Scotland) from 2003 to 2021. A Bayesian Delta-GLMM is used to calculate an index of juvenile abundance based on catch rates, and the index is updated every year when new data become available (ICES 2017). The updated values are shown in Figure 7.2.2.1 (middle panel) and the indices estimated in 20202022 are given in Table 7.2.2.1. Annual revisions of the index are minor. The 2017 data point was highly uncertain due to very limited coverage of the French survey: the French research vessel had technical issues and could therefore only cover less than $1 / 3$ of the stations usually sampled. Despite this high uncertainty, the 2017 data point suggested a very strong recruitment to be expected the following year. This perception was confirmed by the presence of numerous small fish in the 2017 and 2018 catch data. The overall trend suggests an increase in recruitment from 2013 to 2017 and a decrease back down to 2016 levels in 2018. Recruitment in 2019 and 2020 decreased further and was close to the lowest values of the time series, followed by an increase again in 2021.

## Acoustic surveys

In the Bay of Biscay two coordinated acoustic surveys take place in spring, PELGAS (IfremerFrance) and PELACUS (IEO-Spain). Only the PELACUS survey, which cover the ICES division 8 c , is used in the assessment. There is no biomass estimate for 2020 because the survey was cancelled due to the Covid-19 pandemic, however the survey resumed in 2021. The estimate for 2022
is shown in this report (Figure 7.2.2.1, Table 7.2.2.2.), but it is not part of the assessment this year (no catches available yet for 2022).

The biomass estimated by the PELACUS survey was high in the 1990s, reaching the maximum value in 1998 ( 139395 t ). Biomass values are lower in the $21^{\text {st }}$ century, peaking in $2010(53417 \mathrm{t}$ ) and $2015(67068 \mathrm{t})$. Biomass has fluctuated around 10000 t over the most recent 4 surveys.

### 7.2.3 Effort and catch per unit effort

No new information was presented on effort and catch per unit effort.

### 7.2.4 Catch in numbers

In 2021, the Netherlands (4a, 6a, 7bcefghj, 8a), Ireland (6a, 7bj), Norway (4a), Spain (8bc), UK (England \& Scotland) (4a, 6a, 7efj) and Germany (7efghj) provided catch in numbers-at-age (Figure 7.2.4.1). The catch sampled for age readings in 2021 covered $79 \%$ of the total reported catch which is a considerable increase in comparison to last year. Last year's reduction to $51 \%$ was primarily due to the impact of the Covid pandemic on the national sampling programs. Catch in number-at-length were available from the Netherlands (4a, 6a, 7bcefghj, 8a), Ireland (6a, 7bgj), Spain (6a, 7bcghjk, 8abc) and UK (England \& Scotland) (4a, 6a, 7efj) as well as from France (7e, 8ab), Norway (4a) and Germany 7efghj).

The total annual and quarterly catches in number for western horse mackerel in 2021 are shown in Table 7.2.4.1. The sampling intensity is discussed in Section 5.9.

The catch-at-age matrix is given in Table 7.2.4.2 and illustrated in Figures 7.2.4.2 and 7.2.4.3. The latter shows the dominance of the 1982-year class in the catches since 1984 until it entered the plus group in 1997. Since 2002, the 2001-year class, which entered the plus group in 2016, has been caught in considerable numbers. The 2008-year class can be followed in the catch data suggesting it was stronger than other year classes subsequent to the 2001.

Spain, Ireland, the Netherlands and UK (England) also provided the age length keys (ALK) for 2021.

### 7.2.5 Length and age data

Mean length-at-age and mean weight-at-age in the catches
The mean weight and mean length-at-age in the catches by area, and by quarter in 2021 are shown in Tables 7.2.5.1 and 7.2.5.2. Weight-at-age time-series is shown in Figure 7.2.5.1.

Mean weight at age in the stock
Prior to 2017, estimates of mean weight-at-age in the stock for the assessment were based on catch weight-at-age from Q1 and Q2, (Table 7.2.5.3). At present, the stock weight-at-age used in the forecast is an output of the assessment (presented in Table 7.4.1). Further information can be found in the stock annex.

### 7.2.6 Maturity ogive

Maturity-at-age is presented in Table 7.2.6.1. In the assessment model a constant logistic function was used (Figure 7.2.6.1). Further information can be found in the stock annex.

### 7.2.7 Natural mortality

A fixed natural mortality of 0.15 year $^{-1}$ is assumed for all ages and years in the assessment. Further information can be found in the stock annex.

### 7.2.8 Fecundity data

Potential fecundity data ( $10^{6}$ eggs) per kg spawning females are available for the years 1987, 1992, 1995, 1998, 2000, 2001: the data are presented in Table 7.2.8.1 but were not used in the assessment model. In the assessment the fecundity is modelled as linear eggs $/ \mathrm{kg}$ on body weight. Further information can be found in the stock annex.

### 7.2.9 Information from stakeholders

The EU fishing industry, partly in conjunction with the Pelagic Advisory Council (PELAC), has been working on a number of research projects relevant to Western horse mackerel that are briefly reported here. More details can be found in section 1.4.8 of this report and (WD02: Pastoors, 2022).

The Pelagic Freezer-trawler Association (PFA) provided an annual report on the self-sampling programme that started in 2015. Currently, all members ( 15 vessels in 2021) participate in the programme providing data during the main fishing season (October-March). Overall, the selfsampling activities for the horse mackerel fisheries during the years 2016-2022 (up to 11/08/2022) covered 250 fishing trips with 3316 hauls, a total catch of 128553 tonnes and 130146 individual length measurements. The main fishing areas are ICES division 6.a, division 7.b and division 7.j. Western horse mackerel have a wide range in the length distributions in the catch. Median lengths in divisions 6.a, 7.b and 7.j have fluctuated between 25.2 and 31.9 cm . In ICES division 7.h, median lengths in the catch have been smaller and fluctuated between 20.7 and 24.5 cm (for more details see WD02: Pastoors, 2022).

There is also an industry-science collaboration aimed at improving the knowledge on gonad development of mackerel and horse mackerel. Samples were taken by the fishing industry (PFA vessels) on both targeted and by-catches of mackerel and/or horse mackerel. The overall aim for Western horse mackerel is to identify the spawning period in 2020 and investigate if the current egg survey (MEGS) is covering this period for details see section 1.4.8.3.

Additionally, genetic samples have been also collected from 7.d and 7.e by the PFA on board of commercial vessels in the Autumn of 2020, as well as from 4.a during the NS-IBTS in Q3. The study has shown evidence for separating western, North Sea and southern horse mackerel for details see section 1.4.8.3.

### 7.2.10 Data exploration

The length frequency distributions of the landings for the entire fleet included in the model are shown in Figures 7.2.10.1-2. The length distributions 2015-2020 show a considerable amount of very small fish, mostly from Spanish catches. In 2021 the recent trend of large catches of small fish has changed and the length distribution is a more normal distribution with the most common landed lengths around 30 cm . The main mode of the distribution continuously increased since 2004 to 2017. It has been lower in recent years, but has started to increase, probably due to the growth of the small individuals observed in recent years. The length distribution of discards
has been provided by some countries since 2018. However, this information was not available at the last benchmark (2017) and therefore they are not included in the current assessment.

Within-cohort consistency of the catch-at-age matrix is investigated in Figure 7.2.10.3: this shows that the catch-at-age data contains information on year-class strength that could form the basis for an age-structured model. The numbers at age in the catch by decade show a slight trend towards younger individuals when moving from the beginning of the time-series towards the end (Figure 7.2.10.4).

The indices of abundance used in the assessment cover different areas and therefore represent different parts of the stock. Negative correlations between indices that should represent the same portion of the population may lead to problems in the fitting of the model. The correlation between time-series was therefore estimated and is presented in Figure 7.2.10.5. There was no strong correlation between the IBTS recruitment index and the other two surveys. The egg survey index, which aims to represent the adult portion of the stock was strongly positively correlated with the PELACUS acoustic survey biomass estimate.

### 7.2.11 Assessment model, diagnostics

A one fleet, one sex, one area stock synthesis model (SS; Stock Synthesis v3.30) is used for the assessment of Western horse mackerel stock in the Northeast Atlantic. A description of the model can be found in the stock annex. The assessment presented is an update of the 2021 assessment, with the inclusion of the 2021 estimates for the IBTS recruitment index, the 2021 length frequency distribution of the landings, and the 2021 total catch and conditional ALKs. The biomass estimates and length distribution provided by the PELACUS survey were not available in 2020 because the survey was cancelled due to the Covid pandemic (see section 7.13), but the survey resumed in 2021 and was used in this assessment. As in last year 's assessment, the length and age distributions were tuned using the Francis reweighting approach.

Fits to the available data are given in Figure 7.2.11.1, and model estimates with associated precision in Figure 7.2.11.2. Model estimates and residual patterns are similar to those presented in the benchmark (ICES, 2017b) and remain unchanged from last year's assessment for almost all variables, except for some patterns noted in the 2018 and 2020 ALK, that was not evident in 2019 or 2021. Recruitment estimates are higher than last year's assessment. The model does not fit well to the biomass estimates and length composition provided by the PELACUS survey. The fitting to the most recent length frequency distributions and the conditional ALKs remains suboptimal and it does not capture the small fish observed in recent years.

The 2022 assessment shows strong retrospective patterns, with peels falling outside the confidence intervals of SSB (2 peels) and recruitment (3 peels) estimates (Figure 7.2.11.3). The pattern is very consistent and has led to a rescaling of the SSB (downwards) and F (upwards) in the past years. Further investigation is needed to identify the reason of the pattern and resolve it. The Mohn's rho values are now above the limit of the tolerance threshold with 0.329 for SSB and 0.251 for F .

### 7.3 State of the Stock

### 7.3.1 Stock assessment

The SS model with new length and age data from the commercial fleet, and the 2021 information from the IBTS index is presented as the final assessment model. Stock numbers-at-age and fishing mortality-at-age are given in Tables 7.3.1.1 and 7.3.1.2, and a stock summary is provided in Table 7.3.1.3, and illustrated in Figure 7.2.11.2. SSB peaked in 1988 following the recruitment of
the exceptionally strong 1982 year-class. Subsequently, SSB slowly declined until 2003 and then recovered again following the moderate-to-strong year-class of 2001 (a third of the size of the 1982 year-class). SSB reached the minimum values of the time series in 2017 ( 834480 t ), increasing slightly in recent years. In 2022, SSB is estimated to be below Blim.

The recruitment has been weak since 2001, reaching the lowest values in 2009-2011 and 2013. Recruitment estimates for 2014-2018 are the highest observed since 2008 and are higher than the geometric mean estimated over the years 1983-2021. Recruitment in 2019-2021was low again with an increase to remain above the mean in 2022.

Fishing mortality (ages 1-10) has oscillated over the time series. It increased after 2007 as a result of increasing catches and decreasing biomass as the 2001 year-class was reduced. The fishing mortality decreased between 2013 and 2017 due to a decrease in catches and a reduced proportion of the adult population in the exploited stock. The fishing mortality in 2021 (0.086) is just above $\mathrm{F}_{\text {MSY }}(0.074)$ and slightly higher than the previous year but continues to be close to the lowest value in the time series since 2007.

### 7.4 Short-term forecast

A deterministic short-term forecast was conducted using the 'fwd()' method in FLR (Flash R addon package).

## Input

Table 7.4.1. lists the input data for the short-term predictions. Weight at age in the stock and weight at age in the catch are equal to the year-invariant weight at age function used in the stock synthesis model. Exploitation pattern is based on estimated fishing mortality in 2021 and is the average of ages 1 to 10 . Natural mortality is assumed to be 0.15 across all ages. The proportion mature for this stock has a logistic form with fully mature individuals at age 4 as used in the assessment model.

The expected landings for the intermediate year were set at $100 \%$ of the TAC (71 138 t ) after confirmation from individual institutes that TAC was close to being fully taken by August 2022. Note that although the plus group in the catch was set at $15+$, the true population in SS model is set to arrive up to age 20 (as from literature) and is therefore estimated accordingly.

Output
A range of predicted catch and SSB options from the short-term forecast are presented in Table 7.4.2.

### 7.5 Uncertainties in the assessment and forecast

Despite the increased amount of data used and information available to the stock assessment, the model suffers from a retrospective pattern whenever a new year of data is included. This year rescaling is relatively significant with a pattern over the past 5 years (rescaling biomass down and vice-versa for $\mathrm{F}_{1-10}$ ).

The fitting to the fishery independent indices remains good for two of the three surveys used: IBTS and MEGS. A degradation of the fitting to the IBTS recruitment index was observed the past couple of years, but the estimates remained within the confidence intervals provided. The fit to the PELACUS acoustic index remains poor.

The change in selectivity, which is detected from both the length and the age composition of the catch data, is not entirely picked up from the model. In general, the model tends to overestimate the mean age of the last decade. The selectivity issue should be further investigated and
addressed: for example, it is not clear whether the high presence of small specimens in the landings data is due to the inclusion of BMS individuals in the overall catch instead of having it as discard (the discard ban was implemented in 2015 for pelagic species) or if this is due to an effective change in selectivity (i.e. catchability of the gear and availability of the stock).

The model fixes the realised fecundity with a constant number of eggs $/ \mathrm{kg}$ independently of the individual weight. However, Western horse mackerel is known to be an indeterminate spawner, which implies this relationship may not be appropriate when it comes to the use of an egg survey as index of spawning biomass. During the benchmark an attempt was made to estimate the parameters relative to fecundity, however, the information provided to the model was not sufficient. The inclusion of this feature, whenever appropriate data become available, would help to improve the reliability of the assessment.

The assumed value for natural mortality should be investigated. However, there is no data available (such as tagging) that could assist in estimating natural mortality more accurately. Nevertheless, total mortality appears to be low, given the persistence of the 1982-year class in the catch data.

The assessment, as was developed at the benchmark, has an increased amount of information for providing more robust estimates of recruitment, also informed when occasional strong year classes are observed in the catch. On the contrary, the SSB is informed only by the triennial egg survey and by the acoustic survey (which only covers a small part of the stock distribution and size ranges, has a very low weight in the model and is very noisy): a new index for the spawning biomass would therefore be beneficial for the future stability of this assessment. The development of a combined SSB index estimated from appropriate surveys in the area (e.g., PELACUS, PELGAS, WESPAS) should be pursued.

### 7.6 Comparison with previous assessment and forecast

A comparison of the update assessment with the historic ones (previous 4 years) is shown in Figure 7.2.11.4: the new information created a downward rescaling of the assessment biomass and upward revision of F. Recruitment, on the other hand, remains fairly stable until 2015 but a downward revision is estimated from then on.

### 7.7 Management Options

### 7.7.1 MSY approach

In 2017 stochastic equilibrium analyses were carried out using the EqSim software (WKWIDE 2017) to provide an estimate for $\mathrm{Fmsy}^{2}$ and other biological reference points. During WGWIDE 2017 further investigations were carried out and summarised in a Working Document attached to WGWIDE 2017 report (ICES, 2017a).

Reference points were subsequently revised during an inter-benchmark workshop carried out in July-August 2019 as those derived during the 2017 benchmark were deemed no longer appropriate in light of the retrospective pattern observed in the model. More robust reference points were therefore put forward after a number of alternatives were examined, following ICES guidelines, and based on the 2018 assessment. The detailed rationale can be found in the inter-benchmark report (ICES, 2019a).

SSB in 2003 was adopted as a proxy for $\mathrm{B}_{\mathrm{pa}}$ on the basis that fishing mortality had been relatively low for the data period ( $\mathrm{F}_{\text {bar }}$ mean $\sim 0.11$, natural mortality $=0.15$ ), and there was no indication of impaired recruitment below the associated Blim, despite a continuing decline in SSB. Fmsy was
derived from stochastic simulations as before and evaluated at 0.074 . In $2021, \mathrm{~F}_{\mathrm{pa}}$ was re-defined as $\mathrm{F}_{\mathrm{p} 05}$ (ICES, 2021b). These updated reference points were used in determining the MSY based 2023 catch advice.

### 7.7.2 Management plans and evaluations

An overview of earlier management plans and management plan evaluations was presented at WGWIDE 2017. To date, no agreed management plan is available for this stock despite several attempts to develop such management plans. The Pelagic Advisory Council (PELAC), together with several researchers have carried out an evaluation of potential harvest control rules for Western horse mackerel (for details see Stock Annex). This rebuilding plan has not been currently approved by the European Commission and the UK. The working group no longer considers this management plan appropriate as it is outdated by 2 years.

### 7.8 Management considerations

The 2001 year-class has now entered the plus group but no other detectable very strong yearclasses entering the fishery, even though a higher amount of age 1-2 years old fish have been observed in the catches in the past 4-5 years.

Due to the downward revision of the stock, and SSB falling below Blim, following the MSY approach, the advice for 2023 is catches in 2023 should be zero. It is expected that even with 0 catch there will be some discard landings in 2023 available as with previous years.

Note that subarea 8.c is included in the ICES advice for Western horse mackerel.

### 7.9 Ecosystem considerations

Knowledge about the distribution of the Western horse mackerel stock is mostly gained from the egg surveys and the seasonal changes in the fishery. Based on these observations it is not possible to infer a trend in the distribution of Western horse mackerel. However, from catch data it appears that the stock is concentrated in the southern areas, and it is mostly characterized by small individuals.

### 7.10 Regulations and their effects

There are horse mackerel management agreements between EU and the UK, but not with Norway. The TAC set by EU and the UK therefore only applies to EU and UK waters and the EU and UK fleet in international waters. The minimum landing size of horse mackerel by the EU and UK fleet is 15 cm ( $10 \%$ undersized allowed in the catches). In Norwegian waters there is no quota for horse mackerel but existing regulations on bycatch proportions as well as a general discard prohibition (for all species) apply to horse mackerel.

An overview of the scientific advice, the TACs (or sum of unilateral quota) and the catches is shown in Figure 7.10.1. From 2001 onwards, TACs and catches have fluctuated around the scientific advice, where in some years the TACs were set higher and in other years lower than the scientific advice.

The stock allocations were changed in 2005 following the results of the HOMSIR project (Abaunza et al. 2003) and 8.c is considered to be the western stock. Landings from 7.d are now allocated to the North Sea horse mackerel stock. Results of a recent genetic research project on stock structure of horse mackerel has been reported in sections 1.4.8.3. of this report.

### 7.11 Changes in fishing technology and fishing patterns

The description of the fishery is given in Section 5.1 and no large changes in fishing areas or patterns have taken place.

### 7.12 Changes in the environment

Migrations are closely associated with the slope current, and horse mackerel migrations are known to be modulated by temperature. Continued warming of the slope current is likely to affect the timing and spatial extent of this migration.

It has been reported a good correspondence between the modelled influx of Atlantic water to the North Sea in the first quarter and the horse mackerel catches taken by Norwegian purse-seiners in the Norwegian EEZ later in the year (October-November) since 1987 (Iversen et al. 2002, Iversen WD presented in ICES 2007/ACFM:31).

### 7.13 Deviations from stock annex caused by missing information from historic Covid-19 disruption

1. Stock: hom.27.2a4a5b6a7a-ce-k8
2. Missing or deteriorated survey data:

The length composition and the biomass index annually provided by the PELACUS survey were not available in 2020 because the survey was cancelled due to the Covid pandemic.
3. Missing or deteriorated catch data:

The samples for age readings in 2020 covered only $51 \%$ of the catch, whereas in previous years was $69 \%$. This decrease is due to the impact of the Covid pandemic on the national sampling programs. Spain had to reduce its sampling program and no sampling from Germany and Norway were available.
4. Missing or deteriorated commercial LPUE/CPUE data:

Not applicable
5. Missing or deteriorated biological data:

Not applicable
6. Brief description of methods explored to remedy the challenge:

Not applicable
7. Suggested solution to the challenge, including reason for this selecting this solution:

The assessment was carried out without the 2020 data from PELACUS. No alternative options were found.
8. Was there an evaluation of the loss of certainty caused by the solution that was carried out?

To test the sensitivity of the model to the PELACUS data, the assessment conducted in 2020 was carried out without the PELACUS data for 2019 and the results were compared with the outputs of the actual assessment in 2020. The fishing mortality was slightly higher and the spawning biomass slightly lower in recent years in the model without survey data, although the differences were inside of the confidence intervals of the parameters (Figure 7.13.1).

### 7.14 References

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Pastoors, P. 2022. PFA self-sampling report. Working Document to ICES WGWIDE, 24-30. 2022

### 7.15 Tables

Table 7.1.1.1. Western horse mackerel. Catches ( t ) in Subarea 2 by country (Data as submitted by Working Group members).

| Country | 1980 | 1981 | 1982 | 21983 |  | 1984 |  | 1985 | 1986 | 1987 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | - | - | - | - |  | - |  | - | - | 39 |
| France | - | - | - | - |  | 1 |  | 1 | - 2 | - ${ }^{2}$ |
| Germany, Fed.Rep | - | + | - | - |  | - |  | - | - | - |
| Norway | - | - | - | 412 |  | 22 |  | 78 | 214 | 3,272 |
| USSR | - | - | - | - |  | - |  | - | - | - |
| Total | - | + | - | 412 |  | 23 |  | 79 | 214 | 3,311 |
|  | 1988 | 1989 | 1990 | 1991 |  | 1992 |  | 1993 | 1994 | 1995 |
| Faroe Islands | - | - | 9643 | 1,115 |  | 9,157 ${ }^{3}$ |  | 1,068 | - | 950 |
| Denmark | - | - | - | - |  | - |  | - | - | 200 |
| France | -2 | - | - | - |  | - |  | - | 55 | - |
| Germany, Fed. Rep. | 64 | 12 | + | - |  | - |  | - | - | - |
| Norway | 6,285 | 4,770 | 9,135 | 3,200 |  | 4,300 |  | 2,100 | 4 | 11,300 |
| USSR / Russia (1992-) | 469 | 27 | 1,298 | 172 |  |  |  | - | 700 | 1,633 |
| UK (England + Wales) | - | - | 17 |  |  | - |  | - | - | - |
| Total | 6,818 | 4,809 | 11,414 | 4,487 |  | 13,457 |  | 3,168 | 759 | 14,083 |
|  | 1996 | 1997 |  | 1998 | 1999 | 2 | 2000 | 2001 | 2002 | 2003 |
| Faroe Islands | 1,598 | $799^{3}$ |  | $188^{3}$ | $132^{3}$ |  |  | - | - | - |
| Denmark | - | - |  | 1,755 ${ }^{3}$ | - |  |  | - | - | - |
| France | - | - |  | - | - |  |  | - | - | - |
| Germany | - | - |  | - | - |  |  | - | - | - |
| Norway | 887 | 1,170 |  | 234 | 2,304 |  | 841 | 44 | 1,321 | 22 |
| Russia | 881 | 554 |  | 345 | 121 |  | 78 | 16 | 3 | 2 |
| UK (England + Wales) | - | - |  | - | - |  | - | - | - | - |
| Estonia | - | 78 |  | 22 | - |  | - | - | - | - |
| Total | 3,366 | 2,601 |  | 2,544 | 2557 | 9 | 919 | 60 | 1,324 | 24 |


|  |  | 2004 | 2005 | 2006 | 2007 |  | 2008 |  | 2009 | 2010 |  | 2011 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands |  | - | - | 3 | - |  | - |  | - | 222 |  | 224 |
| Denmark |  | - | - | - | - |  | - |  | - | - |  | - |
| France |  | - | - | - | - |  | - |  | - | - |  | - |
| Germany |  | - | - | - | - |  | - |  | - | - |  | - |
| Ireland |  | - | - | - | - |  | - |  | - | - |  | - |
| Netherlands |  | - | - | - | - |  | - |  | - | - |  | 1 |
| Norway |  | 42 | 176 | 27 | - |  | 572 |  | 1,847 | 7 1,364 |  | 298 |
| Russia |  | - | - | - | - |  | - |  | - | - |  | - |
| UK (England + Wales) |  | - | - | - | - |  | - |  | - | - |  | - |
| Estonia |  | - | - | - | - |  | - |  | - | - |  | - |
| Total |  | 42 | 176 | 27 | 0 |  | 572 |  | 1,847 | 7 1,586 |  | - |
|  | 2012 | 2013 | 2014 |  | 2015 | 2016 | 2017 | 2018 |  | 2019 | 2020 | 2021 |
| Faroe Islands | - | - | - |  | - | - | - | - | - | - | - | - |
| Denmark | - | - | - |  | - | - | - | - |  | - | - | - |
| France | + | - | - |  | - | - | - | - |  | - | - | - |
| Germany | - | - | - |  | - | - | - | - |  | - | - | - |
| Ireland | - | - | - |  | - | - | - | - |  | - - | - | - |
| Netherlands | - | - | 107 |  | - | - | - | - |  | - | - | - |
| Norway | 66 | 30 | 302 |  | 10 | 45 | 5 | 718 |  | 867 | 290 | 12 |
| Russia | - | - |  |  | - | - | - | - |  | - | - | - |
| UK (England + Wales) | - | - |  |  | - | - | - | - |  | - | - | - |
| Estonia | - | - |  |  | - | - | - | - |  | - | - | - |
| Total | 66 | 30 | 409 |  | 10 | 45 | 5 | 718 |  | 867 | 290 | 12 |

Table 7.1.1.2. Western horse mackerel. Catches ( t ) in North Sea Subarea 4 and Skagerrak Division 3.a by country (Data submitted by Working Group members). Catches partly concern the North Sea horse mackerel.

| Country | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 8 | 34 | 7 | 55 | 20 | 13 | 13 | 9 | 10 |
| Denmark | 199 | 3,576 | 1,612 | 1,590 | 23,730 | 22,495 | 18,652 | 7,290 | 20,323 |
| Faroe Islands | 260 |  | - | - | - | - | - | - | - |
|  | 292 | 421 | 567 | 366 | 827 | 298 | 2312 | 1891 | 7841 |
| France | + | 139 | 30 | 52 | + | + | - | 3 | 153 |
| Germany, Fed.Rep. | 1,161 | 412 | - | - | - | - | - | - | - |
| Ireland | 101 | 355 | 559 | 2,0292 | 824 | 1602 | 6002 | 8503 | 1,0603 |
| Netherlands | 119 | 2,292 | 7 | 322 | 2 | 203 | 776 | 11,7283 | 34,4253 |
|  | - | - | - | 2 | 94 | - | - | - | - |
| Norway2 | - | - | - | - | - | - | 2 | - | - |
| Poland | 11 | 15 | 6 | 4 | - | 71 | 3 | 339 | 373 |
| Sweden | - | - | - | - | 3 | 998 | 531 | 487 | 5,749 |
| UK (Engl. + Wales) | - | - | - | - | 489 | - | - | - | - |
| UK (Scotland) |  |  |  |  |  |  |  |  |  |
| USSR |  |  |  |  |  |  |  |  |  |
| Total | 2,151 | 7,253 | 2,788 | 4,420 | 25,987 | 24,238 | 20,808 | 20,895 | 62,877 |


| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium | 10 | 13 | - | + | 74 | 57 | 51 | 28 | - |
| Denmark | 23,329 | 20,605 | 6,982 | 7,755 | 6,120 | 3,921 | 2,432 | 1,433 | 976 |
| Estonia | - | - | - | 293 | - | - | 17 | - | - |
| Faroe Islands | - | 942 | 340 | - | 360 | 275 | - | - | 296 |
| France | 248 | 220 | 174 | 162 | 302 | - | - | - | - |
| Germany, Fed.Rep. | 506 | $2,469^{4}$ | 5,995 | 2,801 | 1,570 | 1,014 | 1,600 | 7 | 37 |
| Ireland | - | 687 | 2,657 | 2,600 | 4,086 | 415 | 220 | 1,100 | 8,152 |
| Netherlands | 14,172 | 1,970 | 3,852 | 3,000 | 2,470 | 1,329 | 5,285 | 6,205 | 52 |
| Norway | 84,161 | 117,903 | 50,000 | 96,000 | 126,800 | 94,000 | 84,747 | 14,639 | 43,888 |
| Poland | - | - | - | - | - | - | - | - | - |
| Sweden | - | 102 | 953 | 800 | 697 | 2,087 | - | 95 | 1761 |
| UK (Engl. + Wales) | 10 | 10 | 132 | 4 | 115 | 389 | 478 | 40 | 10 |
| UK (N. Ireland) | - | - | 350 | - | - | - | - | - | - |
| UK (Scotland) | 2,093 | 458 | 7,309 | 996 | 1,059 | 7,582 | 3,650 | 2,442 | 10,511 |
| USSR / Russia (1992 -) | - | - | - | - | - | - | - | - | - |
| Unallocated+discards | $12,482^{3}$ | $-317^{3}$ | $-750^{3}$ | -2785 | $-3,270$ | 1,511 | -28 | 136 | $-31,615^{2}$ |
| Total | 112,047 | 145,062 | 77,904 | 114,133 | 140,383 | 112,580 | 98,452 | 26,125 | 34,068 |


| Country | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 19 | 21 | - | - | - | - | - | - | - |
| Denmark | 2,048 | 2,026 | 7 | 98 | 53 | 841 | 48 | 216 | 60 |
| Estonia | - | - | - | - | - | - | - | - | - |
| Faroe Islands | 28 | 908 | 24 | 0 | 671 | 5 | 76 | 35 | 0 |
| France | 379 | 60 | 49 | - | - | 255 | - | 1 | - |
| Germany | 4,620 | 4,072 | 0 | 0 | 4 | 534 | 0 | 44 | 1 |
| Ireland | - | 404 | 32 | 332 | 11 | 93 | 378 | - | - |
| Lithuania | - | - | - | - | - | - | - | - | - |
| Netherlands | 4,548 | 3,285 | 10 | 1 | 0 | 36 | 0 | 0 | 0 |
| Norway | 13,129 | 44,344 | 1,141 | 7,912 | 34,843 | 20,349 | 10,687 | 24,733 | 27,087 |
| Russia | - | - | 2 | - | - | - | - | - | - |
| Sweden | 1,761 | 1,957 | 1,009 | 68 | 561 | 1,002 | 567 | 216 | 0 |
| UK (Engl. + Wales) | 1 | 12 | - | - | - | - | 0 | - | - |
| UK (Scotland) | 3,041 | 1,658 | 3,054 | 3,161 | 252 | 0 | 0 | 22 | 61 |
| Unallocated+discards | 737 | -325 | 10 | 0 | 0 | -36 | 0 | 0 | 0 |
| Total | 30,311 | 58,422 | 5,338 | 11,572 | 36,395 | 23,079 | 11,756 | 25,267 | 27,210 |

${ }^{1}$ Includes Division 2.a. ${ }^{2}$ Estimated from biological sampling. ${ }^{3}$ Assumed to be misreported. ${ }^{4}$ Includes 13 t from the German Democratic Republic. ${ }^{5}$ Includes a negative unallocated catch of $-4,000 \mathrm{t} .{ }^{6}$ Negative values when there were overestimations of catch when comparing scientific with official data

| Country | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 74 | 2 | 207 | 61 | 19 | 9 | 0 | 23 |
| Faroe Islands | 3 | 55 | 0 | 8 | 0 | 0 | 0 | 53 |
| France | - | 1 | - | - | 268 | - | - | 17 |
| Germany, Fed.Rep. | 6 | 93 | 0 | 4 | 0 | 0 | 20 | 0 |
| Ireland | 651 | 298 | 342 | 14 | 755 | 25 | 7 | - |
| Netherlands | - | - | - | - | - | - | - | - |
| Lithuania | 22 | 0 | 7 | 339 | 81 | 92 | 0 | 310 |
| Norway | 4180 | 11631 | 57890 | 10556 | 13409 | 3183 | 6566 | 14051 |
| Sweden | 76 | 9 | 258 | 2 | 90 | 0 | 1 | 0 |
| UK (Engl. + Wales) | 31 | - | - | - | - | - | 16 | 203 |
| UK (Scotland) | 7 | 20 | 51 | 546 | 101 | 12 | 102 | 11 |
| Unallocated +discards | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 30 |
| Total | 5050 | 12110 | 58755 | 11531 | 14723 | 3320 | 6712 | 14699 |
| Country | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |  |
| Denmark | 37 | 7 | 21 | 289 | 183 | 22 | 11 |  |
| Faroe Islands | 0 | 0 | 67 | 0 | 6 | - | - |  |
| France | 12 | 4 | 1 | 2 | 98 | 0 | 2 |  |
| Germany, Fed.Rep. | 6 | 28 | 1 | 1 | 5 | 0.5 | 3 |  |
| Ireland | 8 | - | - | - | - | - | - |  |
| Netherlands | - | 0 | 14 | 7 | 72 | 1 | 27 |  |
| Lithuania | 12 | 130 | - | - | - | 0 | - |  |
| Norway | 8,887 | 8,765 | 9,880 | 8,601 | 8,154 | 10,376 | 3,651 |  |
| Sweden | 10 | 0 | 41 | 23 | 323 | 83 | 4 |  |
| UK (Engl. + Wales) | 134 | 13 | 4 | 0 | - | 0 | 0.5 |  |
| UK (Scotland) | 36 | 14 | - | - | 50 | - | 63 |  |
| Unallocated +discards | 32 | 97 | 87 | 162** | 339 | 1239 | 160 |  |
| Total | 9,175 | 9,057 | 10,117 | 9,085 | 9144 | 11,700 | 3,923 |  |

Table 7.1.1.3 Western horse mackerel. Catches ( t ) in Subarea 6 by country (Data submitted by Working Group members).

| Country | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 734 | 341 | 2,785 | 7 | - | - | - | 769 | 1,655 |
| Faroe Islands | - | - | 1,248 | - | - | 4,014 | 1,992 | 4,450 ${ }^{2}$ | 4,000 ${ }^{2}$ |
| France | 45 | 454 | 4 | 10 | 14 | 13 | 12 | 20 | 10 |
| Germany, Fed. Rep. | 5,550 | 10,212 | 2,113 | 4,146 | 130 | 191 | 354 | 174 | 615 |
| Ireland | - | - | - | 15,086 | 13,858 | 27,102 | 28,125 | 29,743 | 27,872 |
| Netherlands | 2,385 | 100 | 50 | 94 | 17,500 | 18,450 | 3,450 | 5,750 | 3,340 |
| Norway | - | 5 | - | - | - |  | 83 | 75 | 41 |
| Spain | - | - | - | - | - |  | -1 | _1 | -1 |
| UK (Engl. + Wales) | 9 | 5 | + | 38 | + | 996 | 198 | 404 | 475 |
| UK (N. Ireland) |  |  |  |  |  | - | - | - | - |
| UK (Scotland) | 1 | 17 | 83 | - | 214 | 1,427 | 138 | 1,027 | 7,834 |
| USSR. | - | - | - | - | - | - | - | - | - |
| Unallocated + disc |  |  |  |  |  | -19,168 | $-13,897$ | -7,255 | - |
| Total | 8,724 | 11,134 | 6,283 | 19,381 | 31,716 | 33,025 | 20,455 | 35,157 | 45,842 |
| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| Denmark | 973 | 615 | - | 42 | - | 294 | 106 | 114 | 780 |
| Faroe Islands | 3,059 | 628 | 255 | - | 820 | 80 | - | - | - |
| France | 2 | 17 | 4 | 3 | + | - | - | - | 53 |
| Germany, Fed. Rep. | 1,162 | 2,474 | 2,500 | 6,281 | 10,023 | 1,430 | 1,368 | 943 | 229 |
| Ireland | 19,493 | 15,911 | 24,766 | 32,994 | 44,802 | 65,564 | 120,124 | 87,872 | 22,474 |
| Netherlands | 1,907 | 660 | 3,369 | 2,150 | 590 | 341 | 2,326 | 572 | 1335 |
| Norway | - | - | - | - | - | - | - | - | - |
| Spain | -1 | -1 | 1 | 3 | - | - | - | - | - |
| UK (Engl. + Wales) | 44 | 145 | 1,229 | 577 | 144 | 109 | 208 | 612 | 56 |
| UK (N.Ireland) | - | - | 1,970 | 273 | - | - | - | - | 767 |
| UK (Scotland) | 1,737 | 267 | 1,640 | 86 | 4,523 | 1,760 | 789 | 2,669 | 14,452 |
| USSR/Russia (1992-) | - | 44 | - | - | - | - | - | - | - |
| Unallocated + disc. | 6,493 | 143 | -1,278 | -1,940 | $-6,960{ }^{3}$ | -51 | -41,326 | -11,523 | 837 |


| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total | 34,870 | 20,904 | 34,456 | 40,469 | 53,942 | 69,527 | 83,595 | 81,259 | 40,983 |
| Country | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| Denmark |  | 79 |  |  |  |  |  |  |  |
| Faroe Islands | - | - |  |  |  |  |  |  |  |
| France | 221 |  |  | 428 | 55 | 209 | 172 | 41 | 411 |
| Germany | 414 | 1031 | 209 | 265 | 149 | 1337 | 1413 | 1958 | 1025 |
| Ireland | 21951 | 31736 | 15843 | 20162 | 12341 | 20903 | 15702 | 12395 | 9780 |
| Lithuania |  |  |  |  |  |  |  |  | 2822 |
| Netherlands | 983 | 2646 | 686 | 600 | 450 | 847 | 3702 | 6039 | 1892 |
| Spain | - | - |  |  |  |  |  | 0 | 0 |
| UK (Engl.+Wales) | 227 | 344 | 41 | 91 |  | 46 | 5 | 52 |  |
| UK (N.Ireland) | 1132 | - | 79 | 272 | 654 | 530 | 249 | 210 | 82 |
| UK (Scotland) | 10147 | 4544 | 1839 | 3111 | 1192 | 453 | 377 | 62 | 43 |
| Unallocated+disc. | 98 | 1507 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total | 34815 | 41887 | 18697 | 24929 | 14840 | 24325 | 21619 | 20757 | 16055 |

${ }^{1}$ Included in Subarea 7. ${ }^{2}$ Includes Divisions 3.a, 4.a, b and 6.b. ${ }^{3}$ Includes a negative unallocated catch of $\mathbf{- 7 0 0 0}$ t.

| Country | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  | 58 | 1,131 | 433 | 856 | 3,045 |
| Faroe Islands |  | 573 |  | 66 |  |  |  |  |  |
| France |  | 73 |  |  | 246 |  |  | 195 | 65 |
| Germany | 1,835 | 5,097 | 635 | 773 | 6,508 | 671 | 8,616 | 4,194 | 1,980 |
| Ireland | 20,010 | 18,751 | 16,596 | 19,985 | 23,556 | 29,282 | 19,979 | 15,745 | 10,894 |
| Lithuania | 80 | 641 |  |  |  |  |  |  |  |
| Netherlands | 2,177 | 3,904 | 2,332 | 1,684 | 6,353 | 12,653 | 11,078 | 8,580 | 6,211 |
| Norway | 2 | 20 | 27 | 18 | 48 | 2 |  |  |  |
| Spain | 0 |  |  |  |  |  |  |  |  |
| UK (Engl. + Wales) | 332 |  |  | 463 |  |  | 451 | 18 | 58 |
| UK (N.Ireland) |  |  |  | 59 | 198 |  | 2,325 | 1,579 | 1,204 |


| Country | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | $\mathbf{2 0 1 5}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| UK (Scotland) | 38 | 588 | 243 | 89 | 2,528 | 1,231 | 385 | 1,277 | 696 |
| Unallocated+disc. | 0 | 0 | 0 | 0 | 230 | 2 | - | 123 |  |
| Total | 24,474 | 29,648 | 19,833 | 23,136 | 39,726 | 44,973 | 43,266 | 32,567 | 24,153 |


| Country | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  | 3,462 | 4,982 | 6,467 | 2,267 | 1,853 |
| Faroe Islands |  | 113 |  | 20 |  | - |
| France | 23 | 1,025 | 197 | 550 | 3 | 908 |
| Germany | 4,069 | 2,884 | 2,779 | 1,418 | 0 | - |
| Ireland | 15,381 | 15,123 | 17,959 | 21,109 | 9,187 | 8,530 |
| Lithuania | 2,510 |  |  |  |  | - |
| Netherlands | 9,246 | 5,497 | 11,921 | 14,421 | 5,202 | 1,309 |
| Norway |  |  |  |  |  |  |
| Spain |  |  |  |  |  |  |
| UK (Engl. + Wales) |  | 66 | 32 | 830 | 817 | 249 |
| UK (N.Ireland) | 0 |  | 1,026 | 1,907 | 1,229 | 417 |
| UK (Scotland) | 956 |  |  | 627 | 331** | 459 |
| Unallocated+disc. |  | 116 | 55 | 129 | 108 | 91 |
| Total | 32,186 | 28,286 | 38,950 | 47,480 | 19,146 | 13,818 |

Table 7.1.1.4. Western horse mackerel. Catches ( t ) in Subarea 7 by country (Data submitted by the Working Group members).

| Country | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium | - | 1 | 1 | - | - | + | + | 2 | - |
| Denmark | 5,045 | 3,099 | 877 | 993 | 732 | 1477 | 30408 | 27,368 | 33,202 |
| France | 1,983 | 2,800 | 2,314 | 1,834 | 2,387 | 1,881 | 3,801 | 2,197 | 1,523 |
| Germany, Fed.Rep. | 2,289 | 1,079 | 12 | 1,977 | 228 | - | 5 | 374 | 4,705 |
| Ireland | - | 16 | - | - | 65 | 100 | 703 | 15 | 481 |
| Netherlands | 23,002 | 25,000 | 27500 | 34,350 | 38,700 | 33,550 | 40,750 | 69,400 | 43,560 |


| Country | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Norway | 394 | - | - | - | - | - | - | - | - |
| Spain | 50 | 234 | 104 | 142 | 560 | 275 | 137 | 148 | 150 |
| UK (Engl. + Wales) | 12,933 | 2,520 | 2,670 | 1,230 | 279 | 1,630 | 1,824 | 1,228 | 3,759 |
| UK (Scotland) | 1 | - | - | - | 1 | 1 | + | 2 | 2,873 |
| USSR | - | - | - | - | - | 120 | - | - | - |
| Total | 45,697 | 34,749 | 33,478 | 40,526 | 42,952 | 39,034 | 77,628 | 100,734 | 90,253 |
| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| Faroe Islands | - | 28 | - | - | - | - | - | - | - |
| Belgium | - | + | - | - | - | 1 | - | - | 18 |
| Denmark | 34,474 | 30,594 | 28,888 | 18,984 | 16,978 | 41,605 | 28,300 | 43,330 | 60,412 |
| France | 4,576 | 2,538 | 1,230 | 1,198 | 1,001 | - | - | - | 30,571 |
| Germany, Fed.Rep. | 7,743 | 8,109 | 12,919 | 12,951 | 15,684 | 14,828 | 17,436 | 15,949 | 28,267 |
| Ireland | 12,645 | 17,887 | 19,074 | 15,568 | 16,363 | 15,281 | 58,011 | 38,455 | 43,624 |
| Netherlands | 43,582 | 111,900 | 104,107 | 109,197 | 157,110 | 92,903 | 116,126 | 114,692 | 131,701 |
| Norway | - | - | - | - | - | - | - | - | - |
| Spain | 14 | 16 | 113 | 106 | 54 | 29 | 25 | 33 | 6 |
| UK (Engl. + Wales) | 4,488 | 13,371 | 6,436 | 7,870 | 6,090 | 12,418 | 31,641 | 28,605 | 17,464 |
| UK (N.Ireland) | - | - | 2,026 | 1,690 | 587 | 119 | - | - | 1,093 |
| UK (Scotland) | + | 139 | 1,992 | 5,008 | 3,123 | 9,015 | 10,522 | 11,241 | 7,902 |
| Unallocated + discards | 28,368 | 7,614 | 24,541 | 15,563 | 4,010 | 14,057 | 68,644 | 26,795 | 58,718 |
| Total | 135,890 | 192,196 | 201,326 | 188,135 | 221,000 | 200,256 | 330,705 | 279,100 | 379,776 |
| Country | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| Faroe Islands | - | - |  | 550 | - | - | 3,750 | 3,660 |  |
| Belgium | - | - | - | - |  | - |  |  |  |
| Denmark | 25,492 | 19,166 | 13,794 | 20,574 | 10,094 | 10,499 | 11,619 | 9,939 | 6,838 |
| France | 22,095 | 25,007 | 20,401 | 9,401 | 5,220 | 5,010 | 5,726 | 7,108 | 6,680 |
| Germany | 24,012 | 13,392 | 9,045 | 7,583 | 10,212 | 13,319 | 16,259 | 9,582 | 6,511 |
| Ireland | 48,860 | 25,816 | 32,869 | 29,897 | 23,366 | 13,533 | 8,469 | 20,405 | 16,841 |


| Country | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Lithuania | - | - |  |  |  |  |  |  | 3,606 |
| Netherlands | 95,753 | 63,091 | 44,806 | 37,733 | 32,123 | 38,808 | 32,130 | 26,424 | 29,165 |
| Spain | - | 58 | 50 | 7 | 11 | 1 | 27 | 12 | 3 |
| UK (Engl. + Wales) | 11,925 | 7,249 | 4,391 | 5,913 | 4,393 | 3,411 | 4,097 | 2,670 | 2,754 |
| UK (N.Ireland) | 27 | - | 546 | 868 | 475 | 384 | 209 |  | 21 |
| UK (Scotland) | 5,095 | 4,994 | 5,142 | 1,757 | 1,461 | 268 | 1,146 | 59 | 365 |
| Unallocated+discards | 12,706 | 31,239 | -9,515 | 2,888 | 434 | 17,146 | 16,553 | 11,875 | 4,679 |
| Total | 245,965 | 190,012 | 121,530 | 117,170 | 87,788 | 102,379 | 99,985 | 91,733 | 77,463 |


| Country | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | 475 | 212 |  | - | - | - | 0 |  |  |
| Belgium |  |  |  | 19 | 2 |  | 14 |  |  |
| Denmark | 4856 | 1970 | 2710 | 5247 | 5831 | 2281 | 6373 | 5066 | 1474 |
| France | 2007 | 9703 |  | 260 | 7431 | 579 | 744 | 940 | 1552 |
| Germany | 3943 | 5693 | 14205 | 16847 | 14545 | 16391 | 15781 | 12948 | 7382 |
| Ireland | 8039 | 16282 | 23816 | 24491 | 14154 | 15893 | 15805 | 16922 | 10751 |
| Lithuania | 5387 | 4907 |  |  |  | - | 0 |  |  |
| Netherlands | 32654 | 28077 | 23263 | 65865 | 49207 | 53644 | 41562 | 15529 | 18100 |
| Norway | - | - | - | 40 |  | - | 0 |  |  |
| Spain | 11 | 11 | 6 | 3 |  | 10 | 0 |  |  |
| UK (Engl. + Wales) | 5119 | 3245 | 6257 | 12139 | 11688 | 12122 | 3388 | 4576 | 1798 |
| UK (Scotland) |  | 469 | 1119 | 1713 | 299 | 91 | 17 | 101 | 6 |
| Unallocated+discards | 6012 | -4624 | -10891 | 6511 | 1 | 3038 | 4399 | 974 | 1929 |
| Total | 68504 | 65946 | 60487 | 133136 | 103157 | 104049 | 88083 | 57055 | 42992 |


| Country | 2016 | 2017 | $\mathbf{2 0 1 8}$ | $\mathbf{2 0 1 9}$ | $\mathbf{2 0 2 0}$ | $\mathbf{2 0 2 1}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 314 | 1057 | 1,031 | 690 | 3,198 | 3,540 |
| France | 551 | 595 | 1,067 | 907 | 1,486 | 990 |
| Germany | 7313 | 4077 | 1,401 | 7,673 | 952 | 5,525 |


| Country | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ireland | 12193 | 7857 | 7,169 | 7,753 | 7,870 | 10,240 |
| Lithuania | 86 |  |  |  |  |  |
| Netherlands | 14415 | 8445 | 14,009 | 15,159 | 9,036 | 17,473 |
| Poland |  |  |  | 127 | 1,000 | 1,605 |
| Spain | 0 |  | 0 | 1 | 6 | 14 |
| UK (Engl. + Wales) | 820 | 478 | 2,410 | 2,862 | 679** | 2,401*** |
| UK (Scotland) |  |  |  |  | 3 | 92 |
| UK (Northern Ireland) |  |  | 52 | 0 | 2 | 933 |
| Unallocated+discards | 1692 | 830 | 548 | 918 | 311 | 677 |
| Total | 37384 | 23340 | 27,687 | 36,062 | 24,544 | 43,490 |

${ }^{2}$ French catches landed in the Netherlands **21t BMS landings included

Table 7.1.1.5. Western horse mackerel. Catches ( t ) in Subarea 8 by country (Data submitted by Working Group members).

| Country | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Denmark | - | - | - | - | - | 446 | 3,283 | 2,793 |  |  |
| France | 3,361 | 3,711 | 3.073 | 2,643 | 2,489 | 4,305 | 3,534 | 3,983 | 4,502 |  |
| Netherlands | - | - | - | - | -2 | -2 | -2 | -2 | - |  |
| Spain | 34,134 | 36,362 | 19,610 | 25,580 | 23,119 | 23,292 | 40,334 | 30,098 | 26,629 |  |
| UK (Engl.+Wales) | - | + | 1 | - | 1 | 143 | 392 | 339 | 253 |  |
| USSR | - | - |  |  |  | 20 | - | 656 | - | - |
| Total |  |  |  |  |  |  |  |  |  |  |



| Country | 2017 | $\mathbf{2 0 1 8}$ | 2019 | $\mathbf{2 0 2 0}$ | $\mathbf{2 0 2 1}$ |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Spain | 14,901 | 20,362 | 25,775 | 19,163 | 16,177 |  |
| UK (Engl. + Wales) | 2,907 | 1,921 | 1,755 | 1,104 | 713 |  |
| Unallocated+discards | 21,213 | 25,240 | 31,396 | 20,742 | 20,314 |  |
| Total | 2 |  |  |  |  |  |

${ }^{2}$ Included in Subarea 7. ${ }^{3}$ French catches landed in the Netherlands

Table 7.2.1.1. Western horse mackerel. The time series of Total Annual Egg Production (TAEP) estimates (10 ${ }^{12}$ eggs). (*) means preliminary.

| Year | TAEP |
| :--- | :--- |
| 1992 | 2094 |
| 1995 | 1344 |
| 1998 | 1242 |
| 2001 | 864 |
| 2004 | 884 |
| 2007 | 1486 |
| 2010 | 1033 |
| 2013 | 366 |
| 2016 | 311 |
| 2019 | 178 |
| $2022^{*}$ | 515 |

Table 7.2.2.1. Western horse mackerel. Time series of recruitment index estimated from the IBTS Surveys (2003-2021) in 2020-2022.

| Year | Index 2022 |  | Index 2021 | Index 2020 |
| :--- | :--- | :--- | :--- | :--- |
| 2003 | Mean | CV |  |  |
| 2004 | 2516442 | 0.30 | 732297 | 2434708 |
| 2005 | 2199332 | 0.31 | 2453310 | 2148828 |
| 2006 | 1501474 | 0.33 | 1499811 | 30882969 |
| 2007 | 7824230 | 0.29 | 3121579 | 7272792 |
| 2008 |  |  | 7481365 |  |


| Year | Index 2022 |  | Index 2021 | Index 2020 |
| :---: | :---: | :---: | :---: | :---: |
|  | Mean | CV |  |  |
| 2009 | 1127972 | 0.27 | 1148964 | 1135301 |
| 2010 | 872244 | 0.30 | 864772 | 860652 |
| 2011 | 175162 | 0.35 | 178188 | 180361 |
| 2012 | 4435133 | 0.31 | 4339882 | 4356450 |
| 2013 | 1099932 | 0.24 | 1111210 | 1092849 |
| 2014 | 2905589 | 0.24 | 2931963 | 2922237 |
| 2015 | 4123241 | 0.28 | 4060794 | 4030569 |
| 2016 | 5421010 | 0.29 | 5280009 | 5216531 |
| 2017 | 9395798 | 0.49 | 9460399 | 9450737 |
| 2018 | 5657414 | 0.29 | 5657414 | 4000271 |
| 2019 | 1637102 | 0.29 | 1637102 | 1636554 |
| 2020 | 878485 | 0.27 | 878484 |  |
| 2021 | 1015429 | 0.24 |  |  |

Table 7.2.2.2. Western horse mackerel. Time series of biomass from the PELACUS acoustic survey (in tonnes).

| Year | Biomass | CV |
| :---: | :---: | :---: |
| 1992 | 57188 | 0.32 |
| 1993 | 25028 | 0.32 |
| 1995 | 93825 | 0.32 |
| 1997 | 74364 | 0.32 |
| 1998 | 139395 | 0.32 |
| 1999 | 71744 | 0.32 |
| 2000 | 26192 | 0.32 |
| 2001 | 40864 | 0.32 |
| 2002 | 41788 | 0.32 |
| 2003 | 26647 | 0.32 |
| 2004 | 23992 | 0.32 |
| 2005 | 40082 | 0.32 |
| 2006 | 13934 | 0.32 |


| Year | Biomass | CV |
| :---: | :---: | :---: |
| 2007 | 28173 | 0.32 |
| 2008 | 33614 | 0.32 |
| 2009 | 24020 | 0.32 |
| 2010 | 53417 | 0.32 |
| 2011 | 7687 | 0.32 |
| 2012 | 15479 | 0.32 |
| 2013 | 5532 | 0.32 |
| 2014 | 30454 | 0.32 |
| 2015 | 67068 | 0.32 |
| 2016 | 32581 | 0.32 |
| 2017 | 13845 | 0.32 |
| 2018 | 9270 | 0.32 |
| 2019 | 13075 | 0.32 |
| 2020 | NA | NA |
| 2021 | 10233 | 0.32 |
| 2022 | 18584 | 0.32 |

Table 7.2.4.1. Western Horse Mackerel stock. Catch in numbers (thousands) at age by quarter and area in 2021 ( $15=15+$ group)

| Q1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 27.2.a | 27.6.a | 27.7.b | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.7.k. 1 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 176.35 | 752.15 | 120.21 | 63.17 | 2043.64 | 0.72 | 3156.24 |
| 2 | 0 | 0 | 133.80 | 95.92 | 886.13 | 495.17 | 166.46 | 1323.10 | 3.08 | 63.16 | 0 | 49.37 | 245.53 | 182.76 | 161.46 | 244.24 | 0.20 | 4050.36 |
| 3 | 0.04 | 79.67 | 1903.08 | 1228.80 | 11121.17 | 4629.76 | 3470.68 | 17349.72 | 39.43 | 826.03 | 0.02 | 24.51 | 19.61 | 161.41 | 65.21 | 167.85 | 0.10 | 41087.08 |
| 4 | 0.83 | 1866.14 | 5518.74 | 1954.61 | 14505.82 | 8440.11 | 7269.86 | 25341.21 | 62.72 | 1438.23 | 0.02 | 18.43 | 2.65 | 0 | 12.17 | 296.52 | 0.08 | 66728.15 |
| 5 | 0.98 | 2198.85 | 2768.86 | 483.61 | 3207.53 | 1381.36 | 1674.92 | 4888.14 | 14.71 | 714.82 | 0.01 | 44.08 | 6.45 | 0 | 108.37 | 629.84 | 0.18 | 18122.70 |
| 6 | 1.68 | 3765.30 | 5691.95 | 434.04 | 1711.22 | 597.99 | 676.09 | 2948.37 | 13.52 | 1840.62 | 0.01 | 89.32 | 13.09 | 0 | 266.84 | 1228.99 | 0.36 | 19279.40 |
| 7 | 11.83 | 26467.51 | 17066.48 | 1392.89 | 2000.11 | 1857.82 | 1084.79 | 3410.17 | 38.23 | 12477.64 | 0.01 | 60.23 | 9.77 | 0 | 370.69 | 636.97 | 0.25 | 66885.38 |
| 8 | 1.07 | 2389.65 | 1403.38 | 103.16 | 102.87 | 118.92 | 13.51 | 176.32 | 2.50 | 653.03 | 0 | 60.60 | 11.09 | 0 | 569.06 | 443.63 | 0.25 | 6049.04 |
| 9 | 1.43 | 3196.87 | 1538.38 | 89.18 | 73.97 | 127.01 | 6.37 | 262.43 | 2.86 | 844.09 | 0 | 54.47 | 10.49 | 0 | 629.95 | 279.81 | 0.22 | 7117.53 |
| 10 | 0.33 | 731.31 | 913.53 | 81.55 | 373.17 | 33.93 | 2.12 | 200.45 | 1.81 | 253.56 | 0 | 54.71 | 11.84 | 0 | 575.97 | 336.34 | 0.22 | 3570.83 |
| 11 | 0.30 | 660.54 | 574.12 | 23.90 | 19.82 | 14.39 | 0.66 | 42.37 | 0.77 | 112.98 | 0 | 47.10 | 9.25 | 0 | 527.26 | 259.22 | 0.19 | 2292.87 |
| 12 | 0.38 | 860.39 | 296.77 | 58.72 | 19.71 | 14.31 | 1.32 | 39.11 | 0.76 | 354.02 | 0 | 23.40 | 4.61 | 0 | 260.51 | 130.12 | 0.10 | 2064.24 |
| 13 | 1.08 | 2418.11 | 877.00 | 119.15 | 478.28 | 133.04 | 3.29 | 110.73 | 1.98 | 318.27 | 0 | 9.10 | 1.62 | 0 | 126.88 | 25.24 | 0.04 | 4623.81 |
| 14 | 0.18 | 395.74 | 53.53 | 1.90 | 1.57 | 1.14 | 0.01 | 3.12 | 0.06 | 1.25 | 0 | 6.66 | 0.96 | 0 | 83.60 | 27.88 | 0.03 | 577.62 |
| 15+ | 0.70 | 1568.27 | 232.92 | 541.08 | 183.30 | 157.48 | 11.87 | 285.16 | 1.59 | 217.52 | 0 | 18.83 | 2.84 | 0 | 237.99 | 77.26 | 0.08 | 3536.88 |


| Q2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 115.39 | 0.20 | 17.44 | 66.96 | 0.21 | 2.24 | 6.51 | 2.12 | 53.42 | 86.47 | 1362.73 | 38.75 | 332.22 | 50.88 | 0.11 | 2135.63 |
| 2 | 0 | 0.01 | 796.06 | 1.39 | 120.34 | 461.99 | 1.42 | 15.43 | 44.94 | 14.60 | 368.58 | 49.34 | 728.18 | 246.68 | 12.19 | 303.67 | 0.08 | 3164.89 |
| 3 | 0.01 | 0 | 206.90 | 0.36 | 31.28 | 120.07 | 0.37 | 4.01 | 11.68 | 3.79 | 95.79 | 80.20 | 273.22 | 327.36 | 306.39 | 3381.58 | 0.26 | 4843.28 |
| 4 | 0.24 | 0 | 20.97 | 0.04 | 3.17 | 11.43 | 0.04 | 0.41 | 1.18 | 0.38 | 10.45 | 66.82 | 70.57 | 15.81 | 1384.61 | 2320.77 | 0.23 | 3907.12 |
| 5 | 0.31 | 0 | 62.92 | 0.11 | 9.51 | 34.30 | 0.11 | 1.22 | 3.55 | 1.15 | 31.35 | 117.98 | 152.35 | 3.02 | 5285.07 | 1201.77 | 0.40 | 6905.14 |
| 6 | 0.50 | 0 | 62.92 | 0.11 | 9.51 | 34.30 | 0.11 | 1.22 | 3.55 | 1.15 | 31.35 | 127.66 | 173.69 | 1.82 | 6418.45 | 553.20 | 0.43 | 7419.97 |
| 7 | 3.49 | 0 | 125.84 | 0.22 | 19.02 | 68.59 | 0.22 | 2.44 | 7.10 | 2.31 | 62.70 | 98.57 | 140.05 | 0.24 | 4115.03 | 254.41 | 0.27 | 4900.52 |
| 8 | 0.54 | 0 | 41.95 | 0.07 | 6.34 | 22.86 | 0.07 | 0.81 | 2.37 | 0.77 | 20.90 | 56.33 | 159.58 | 0.15 | 1856.97 | 392.47 | 0.14 | 2562.33 |
| 9 | 0.50 | 0 | 83.89 | 0.15 | 12.68 | 45.73 | 0.15 | 1.63 | 4.74 | 1.54 | 41.80 | 43.45 | 182.72 | 0.22 | 1108.58 | 374.21 | 0.10 | 1902.08 |
| 10 | 0.21 | 0 | 20.97 | 0.04 | 3.17 | 11.43 | 0.04 | 0.41 | 1.18 | 0.38 | 10.45 | 40.53 | 186.71 | 0.19 | 1075.14 | 527.36 | 0.11 | 1878.31 |
| 11 | 0.12 | 0 | 41.95 | 0.07 | 6.34 | 22.86 | 0.07 | 0.81 | 2.37 | 0.77 | 20.90 | 38.14 | 156.55 | 0.24 | 743.91 | 495.42 | 0.08 | 1530.61 |
| 12 | 0.16 | 0 | 20.97 | 0.04 | 3.17 | 11.43 | 0.04 | 0.41 | 1.18 | 0.38 | 10.45 | 40.37 | 63.68 | 0.13 | 285.96 | 282.56 | 0.04 | 720.98 |
| 13 | 0.38 | 0 | 20.97 | 0.04 | 3.17 | 11.43 | 0.04 | 0.41 | 1.18 | 0.38 | 10.45 | 27.39 | 33.50 | 0.05 | 138.93 | 68.76 | 0.02 | 317.10 |
| 14 | 0.14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3.00 | 24.12 | 0.12 | 57.98 | 80.88 | 0.01 | 166.26 |
| 15+ | 0.46 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 37.17 | 93.32 | 1.32 | 232.62 | 431.50 | 0.05 | 796.44 |


| Q3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.3.a | 27.4.a | 27.6.a | 27.7.a | 27.7.b |  |  |  | \% |  | 27.j | 27.7.j. 2 |  | 27.8.a | 27.8.6 | 27.8.c | 27.8.9.6 | 27.8.e.w | 27.8.a | 27.8.a.2 | 27.8.e | Total |
| 0 | 0 | 0 | 0 | 0 |  | ${ }_{0}^{0} 8$ | 0.01 | 5.12 | 0 | 0.14 | ${ }_{0}^{0} 0$ | 0.00 | 0.02 | 0 | ${ }_{1}{ }_{148.35}$ | 79.42 502.62 |  | 149.96 |  | 0 | 0.01 |  | $\stackrel{234}{ }$ |
| 1 | 0 | 0 | 0 | 0 | 0 0.05 | 0.88 34.70 | 0.36 | 5.178 | 0.12 | 0.14 5.54 | 0.09 3.68 | 0.00 158.69 | 41.69 | 0 | 148.35 77.61 | ${ }^{5174.72}$ | 45.815 | 6467.41 3379.52 | 4.64 65.94 | 0.37 0.19 | 0.21 0.11 | 0.04 0.02 | 7175.72 3997.47 |
| 3 | 0 | 0 | 0 | 0 | 0.36 | 254.51 | 2.65 | 11.58 | 0.87 | 40.60 | 27.00 | 898.38 | 572.70 | 0.03 | 40.63 | 68.28 | 0 | ${ }_{1253.11}$ | 601.25 | 0.10 | 0.06 | 0.01 | 3772.13 |
| 4 | 0 | 0.01 | 1.54 | 0.01 | 0.58 | 412.80 | 4.30 | 18.94 | 1.41 | 65.86 | 43.79 | 1233.82 | 1152.02 | 0.06 | 44.82 | 70.92 | 0 | 1077.88 | 972.39 | 0.11 | 0.06 | 0.01 | 5101.32 |
| 5 | 0.11 | 0.62 | 163.11 | 0.65 | 0.49 | 348.35 | 3.63 | 15.86 | 1.19 | 55.58 | 36.95 | 891.00 | 1122.46 | 0.05 | 80.33 | 128.20 | 0.01 | 1384.43 | 2288.96 | 0.20 | 0.11 | 0.02 | 6522.28 |
| 6 | 0.05 | 0.29 | 76.29 | 0.31 | 1.16 | 830.86 | 8.65 | 37.04 | 2.84 | 132.56 | 88.14 | 2147.21 | 2656.01 | 0.11 | 121.08 | 201.64 | 0.03 | 1750.24 | 3778.11 | 0.30 | 0.17 | 0.03 | 11833.12 |
| 7 | 0.33 | 1.90 | 501.87 | 2.01 | 1.50 | 1072.86 | 11.17 | 47.71 | 3.67 | 171.16 | 113.81 | 1549.35 | 4652.98 | 0.15 | 47.18 | 110.32 | 0.05 | 765.52 | 1356.63 | 0.12 | 0.07 | 0.01 | 10410.34 |
| 8 | 0.83 | 4.78 | 1264.97 | 5.07 | 0.21 | 150.32 | 1.56 | 6.78 | 0.51 | 23.98 | 15.95 | 240.86 | 628.04 | 0.02 | 17.18 | 52.29 | 0.10 | 312.67 | 448.15 | 0.04 | 0.02 | 0 | 3174.33 |
| 9 | 0.32 | 1.85 | 488.88 | 1.96 | 0.08 | 57.14 | 0.59 | 3.22 | 0.20 | 9.12 | 6.06 | 56.77 | 272.91 | 0.01 | 22.34 | 64.42 | 0.10 | 289.49 | 702.94 | 0.06 | 0.03 | 0.01 | 1978.49 |
| 10 | 0.40 | 2.32 | 615.40 | 2.46 | 0.05 | 39.22 | 0.41 | 1.71 | 0.13 | 6.26 | 4.16 | 56.76 | 170.01 | 0.01 | 21.12 | 58.16 | 0.06 | 307.55 | 633.61 | 0.05 | 0.03 | 0.01 | 1919.89 |
| 11 | 0.14 | 0.80 | 211.62 | 0.85 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 | 17.69 | 36.36 | 0.14 | 298.00 | 502.49 | 0.04 | 0.02 | 0 | 1068.16 |
| 12 | 0.18 | 1.04 | 276.42 | 1.11 | 0.03 | 22.08 | 0.23 | 0.96 | 0.08 | 3.52 | 2.34 | 53.25 | 74.45 | 0 | 10.04 | 16.77 | 0.09 | 153.47 | 304.98 | 0.03 | 0.01 | 0 | 921.10 |
| 13 | 0.25 | 1.47 | 388.91 | 1.56 | 0.01 | 7.71 | 0.08 | 0.34 | 0.03 | 1.23 | 0.82 | 0.00 | 44.59 | 0 | 12.04 | 19.83 | 0.08 | 374.52 | 175.26 | 0.03 | 0.02 | 0 | 1028.79 |
| 14 | 0.32 | 1.85 | 490.47 | 1.96 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 | 12.33 | 18.00 | 0.04 | 410.97 | 154.31 | 0.03 | 0.02 | 0 | 1090.29 |
| 15+ | 0.94 | 5.45 | 1443.63 | 5.78 | 0.02 | 13.30 | 0.14 | 0.58 | 0.05 | 2.12 | 1.41 | 0.01 | 76.92 | 0 | 15.09 | 19.52 | 0 | 482.00 | 212.53 | 0.04 | 0.02 | 0 | 2279.56 |


| Q4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.3.a | 27.4.a | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1529.36 | 770.70 | 0.91 | 0.11 | 385.72 | 2686.80 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5293.83 | 1019.14 | 115.08 | 752.91 | 2587.13 | 9768.10 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4848.31 | 205.47 | 128.25 | 1528.48 | 2378.86 | 9089.36 |
| 3 | 0.21 | 0.21 | 0.09 | 154.76 | 0.01 | 6.35 | 0.02 | 1.53 | 99.06 | 2.04 | 20.14 | 1060.58 | 146.84 | 1880.53 | 971.11 | 81.18 | 3.02 | 390.59 | 370.27 | 5188.53 |
| 4 | 4.79 | 4.97 | 2.04 | 3601.82 | 0.04 | 66.25 | 0.06 | 4.95 | 432.64 | 6.63 | 65.32 | 2091.14 | 476.32 | 7292.19 | 684.38 | 78.66 | 1.53 | 159.01 | 352.97 | 15325.72 |
| 5 | 1.90 | 1.97 | 89.95 | 1339.08 | 0.02 | 23.65 | 0.03 | 2.49 | 303.72 | 3.33 | 32.86 | 1033.35 | 239.61 | 3610.44 | 738.85 | 92.46 | 1.59 | 169.87 | 374.19 | 8059.36 |
| 6 | 2.02 | 2.10 | 407.39 | 1114.13 | 0.03 | 13.78 | 0.05 | 4.07 | 323.65 | 5.45 | 53.73 | 2263.30 | 391.83 | 5528.57 | 943.76 | 132.27 | 0.94 | 278.33 | 400.97 | 11866.38 |
| 7 | 7.80 | 8.09 | 906.59 | 4961.01 | 0.06 | 42.88 | 0.09 | 8.15 | 502.41 | 10.91 | 107.52 | 4066.96 | 784.00 | 11653.37 | 593.19 | 117.05 | 0.17 | 197.97 | 191.20 | 24159.43 |
| 8 | 0.88 | 0.91 | 422.99 | 238.01 | 0.00 | 1.73 | 0.01 | 0.43 | 23.35 | 0.58 | 5.69 | 26.48 | 41.46 | 808.54 | 453.84 | 75.58 | 0.03 | 171.05 | 143.03 | 2414.59 |
| 9 | 0.72 | 0.75 | 308.92 | 235.90 | 0.00 | 0.65 | 0 | 0.22 | 11.96 | 0.30 | 2.91 | 13.57 | 21.24 | 414.53 | 622.91 | 76.68 | 0.12 | 251.06 | 209.56 | 2172.01 |
| 10 | 0.47 | 0.49 | 323.57 | 31.56 | 0.00 | 0.10 | 0 | 0.06 | 3.16 | 0.08 | 0.77 | 3.59 | 5.62 | 109.70 | 1064.44 | 110.50 | 0.20 | 357.48 | 453.97 | 2465.76 |
| 11 | 0.38 | 0.39 | 264.16 | 22.23 | 0.00 | 0.01 | 0 | 0.01 | 0.31 | 0.01 | 0.07 | 0.35 | 0.54 | 10.60 | 873.63 | 78.71 | 0.08 | 282.41 | 398.84 | 1932.73 |
| 12 | 0.75 | 0.78 | 513.55 | 54.35 | 0.00 | 0.20 | 0 | 0.12 | 6.51 | 0.16 | 1.58 | 7.38 | 11.55 | 225.63 | 422.02 | 27.59 | 0.10 | 105.73 | 234.75 | 1612.77 |
| 13 | 0.82 | 0.85 | 461.29 | 158.37 | 0.00 | 0.50 | 0 | 0.13 | 7.09 | 0.18 | 1.73 | 8.04 | 12.59 | 245.47 | 448.33 | 39.88 | 0.04 | 142.24 | 206.27 | 1733.82 |
| 14 | 0.48 | 0.50 | 351.51 | 11.94 | 0.00 | 0.01 | 0 | 0.01 | 0.36 | 0.01 | 0.09 | 0.41 | 0.64 | 12.53 | 505.66 | 54.10 | 0.02 | 140.06 | 240.94 | 1319.27 |
| 15+ | 1.26 | 1.30 | 885.97 | 59.48 | 0.00 | 0.16 | 0 | 0.10 | 5.23 | 0.13 | 1.27 | 5.93 | 9.29 | 181.38 | 585.91 | 81.93 | 0.04 | 151.62 | 263.23 | 2234.24 |

Table 7.2.4.1 cont. Western Horse Mackerel stock. Catch in numbers (thousands) at age by quarter and area in 2021 ( $\mathbf{1 5}=\mathbf{1 5 +}$ group)

| all Q |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.3.a | 27.4.a | 27.6.a | 27.7.a | 27.7.6 | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.7. 2 | 27.7.k. 1 | 27.7.7. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | 27.8.d. 2 | 27.8.e | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1534.21 | 850.13 | 0.91 | ${ }^{250.06}$ | 385.87 | 0.01 | 0.01 | 0 | 2921.20 |
| 1 | 0 | 0 | 0 | 0 | 0 | 116.28 | 0.20 | 17.45 | 72.09 | 0.21 | 2.39 | 6.61 | 2.12 | 53.44 | 0 | 0 | 5705.00 | 3636.64 | 319.84 | 7615.71 | 4686.29 | 1.20 | 0.21 | . 04 | 22235.72 |
| 2 | 0 | 0 | O | 0 | 0.06 | 966.67 | 1.39 | 216.62 | 1349.90 | 496.71 | 192.58 | 1371.87 | 176.36 | 473.58 | 0 | 0 | 5024.63 | 1353.90 | 610.43 | 5081.65 | 2992.71 | 0.47 | 0.11 | 0.02 | 20309.67 |
| 3 | 0.24 | 0.21 | 3.47 | 242.83 | 0.37 | 2398.53 | 0.38 | 1264.25 | 11351.89 | 4633.05 | 3603.28 | 18451.03 | 1088.45 | 3377.13 | 0.02 | 0.03 | 1116.46 | 442.28 | 491.80 | 2015.31 | 4520.95 | 0.46 | 0.06 | 0.01 | 55002.47 |
| 4 | 5.62 | 4.98 | 82.50 | 5663.86 | 0.62 | 6065.82 | 0.09 | 1967.03 | 14968.84 | 8448.18 | 7516.73 | 27480.82 | 1773.24 | 9896.42 | 0.02 | 0.06 | 814.46 | 222.81 | 17.34 | 2633.67 | 3942.66 | 0.41 | 0.06 | 0.01 | 91506.26 |
| 5 | 2.99 | 2.59 | 303.54 | 3664.04 | 0.51 | 3217.47 | 0.14 | 499.23 | 3561.40 | 1386.00 | 1798.13 | 5963.00 | 1146.47 | 5480.10 | 0.01 | 0.05 | 981.25 | 379.45 | 4.63 | 6947.74 | 4494.77 | 0.77 | 0.11 | 0.02 | 39834.41 |
| 6 | 3.75 | 2.39 | 552.06 | 5049.73 | 1.19 | 6616.41 | 0.16 | 456.27 | 2106.20 | 606.39 | 905.03 | 5304.62 | 2553.72 | 10057.83 | 0.01 | 0.11 | 1281.83 | 520.68 | 2.79 | 8713.86 | 5961.28 | 1.09 | 0.17 | 0.03 | 50697.59 |
| 7 | 19.95 | 9.99 | 1811.24 | 32432.52 | 1.57 | 18346.36 | 0.31 | 1431.23 | 2618.81 | 1872.62 | 1459.73 | 7600.88 | 2373.88 | 28849.56 | 0.01 | 0.15 | 799.16 | 377.19 | 0.47 | 5449.21 | 2439.21 | 0.63 | 0.07 | 0.01 | 107894.78 |
| 8 | 2.77 | 5.69 | 1745.98 | 2777.10 | 0.21 | 1600.11 | 0.08 | 111.50 | 155.86 | 120.08 | 50.68 | 221.32 | 285.59 | 2110.71 | 0 | 0.02 | 587.96 | 298.54 | 0.28 | 2909.75 | 1427.28 | 0.43 | 0.02 | 0 | 14411.98 |
| 9 | 2.47 | 2.60 | 850.69 | 3566.35 | 0.08 | 1682.37 | 0.15 | 102.68 | 134.87 | 127.65 | 25.68 | 286.97 | 82.41 | 1573.51 | 0 | 0.01 | 743.17 | 334.31 | 0.44 | 2279.09 | 1566.52 | 0.38 | 0.03 | 0.01 | 13362.44 |
| 10 | 1.20 | 2.81 | 962.83 | 824.70 | 0.06 | 975.20 | 0.04 | 85.19 | 389.48 | 34.18 | 12.92 | 209.48 | 64.57 | 543.82 | 0 | 0.01 | 1180.80 | 367.20 | 0.45 | 2316.14 | 1951.28 | 0.38 | 0.03 | 0.01 | 9922.76 |
| 11 | 0.81 | 1.19 | 491.26 | 722.13 | , | 616.56 | 0.07 | 30.24 | 42.99 | 14.47 | 2.75 | 45.12 | 2.08 | 144.52 | 0 | - | 976.55 | 280.87 | 0.46 | 1851.58 | 1655.98 | 0.32 | 0.02 | , | 6880.01 |
| 12 | 1.32 | 1.83 | 813.27 | 973.77 | 0.03 | 340.75 | 0.04 | 62.24 | 38.62 | 14.59 | 8.60 | 50.07 | 65.95 | 664.60 | 0 | 0 | 495.83 | 112.64 | 0.32 | 805.67 | 952.41 | 0.16 | 0.01 | 0 | 5402.73 |
| 13 | 2.16 | 2.32 | 893.84 | 2686.59 | 0.01 | 907.62 | 0.04 | 122.53 | 497.13 | 133.28 | 10.17 | 120.87 | 14.96 | 618.90 | 0 | 0 | 496.86 | 94.83 | 0.17 | 782.57 | 475.53 | 0.08 | 0.02 | 0 | 7860.50 |
| 14 | 0.98 | 2.35 | 860.15 | 454.85 | - | 53.58 | 0 | 1.90 | 1.93 | 1.15 | 0.21 | 3.53 | 0.70 | 13.78 | 0 | 0 | 527.64 | 97.18 | 0.18 | 692.60 | 504.02 | 0.07 | 0.02 | 0 | 3216.85 |
| 15+ | 2.90 | 6.76 | 2385.70 | 1773.09 | 0.02 | 247.55 | 0 | 541.31 | 189.11 | 157.66 | 18.13 | 292.59 | 10.88 | 475.90 | 0 | 0 | 657.00 | 197.60 | 1.37 | 1104.24 | 984.51 | 0.16 | 0.02 | 0 | 9046.51 |

Table 7.2.4.2. Western horse mackerel. Catch-at-age (thousands).

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0 | 3713 | 21072 | 134743 | 11515 | 13197 | 11741 | 8848 | 1651 | 414 | 1651 | 6582 | 18483 | 28679 | 19432 | 8210 |
| 1983 | 0 | 7903 | 2269 | 32900 | 53508 | 15345 | 44539 | 52673 | 17923 | 3291 | 5505 | 3386 | 17017 | 23902 | 38352 | 46482 |
| 1984 | 0 | 0 | 241360 | 4439 | 36294 | 149798 | 22350 | 38244 | 34020 | 14756 | 4101 | 0 | 639 | 1757 | 5080 | 50895 |
| 1985 | 0 | 1633 | 4901 | 602992 | 4463 | 41822 | 100376 | 12644 | 16172 | 6200 | 9224 | 339 | 850 | 3723 | 1250 | 34814 |
| 1986 | 0 | 0 | 0 | 1548 | 676208 | 8727 | 65147 | 109747 | 25712 | 21179 | 15271 | 3116 | 1031 | 855 | 292 | 51531 |
| 1987 | 0 | 99 | 493 | 0 | 2950 | 891660 | 2061 | 41564 | 90814 | 11740 | 9549 | 19363 | 8917 | 1398 | 200 | 32899 |
| 1988 | 876 | 27369 | 6112 | 2099 | 4402 | 18968 | 941725 | 12115 | 39913 | 67869 | 9739 | 16326 | 17304 | 5179 | 4892 | 32396 |
| 1989 | 0 | 0 | 0 | 20766 | 18282 | 5308 | 14500 | 1276730 | 12046 | 59357 | 83125 | 13905 | 24196 | 13731 | 8987 | 18132 |
| 1990 | 0 | 20406 | 45036 | 138929 | 61442 | 33298 | 10549 | 20607 | 1384850 | 37011 | 70512 | 101945 | 14987 | 34687 | 18077 | 56598 |
| 1991 | 20176 | 24021 | 56066 | 17977 | 159643 | 97147 | 49515 | 21713 | 17148 | 1028420 | 20309 | 12161 | 43665 | 8141 | 7053 | 25553 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 14888 | 229694 | 36332 | 80550 | 56280 | 255874 | 126816 | 48711 | 18992 | 23447 | 1099780 | 13409 | 23002 | 65250 | 11967 | 33246 |
| 1993 | 46 | 131108 | 109807 | 16738 | 62342 | 105760 | 325674 | 141148 | 68418 | 55289 | 30689 | 1075610 | 11373 | 24018 | 68137 | 32140 |
| 1994 | 3686 | 60759 | 911713 | 115729 | 53056 | 44520 | 38769 | 221863 | 106390 | 40988 | 43083 | 22380 | 918512 | 10143 | 14599 | 36635 |
| 1995 | 2702 | 233030 | 646753 | 526053 | 269658 | 74592 | 114649 | 36076 | 228687 | 113304 | 96624 | 59874 | 63187 | 951901 | 39278 | 148243 |
| 1996 | 10729 | 19774 | 659641 | 864188 | 189273 | 87562 | 52050 | 55914 | 53835 | 57361 | 56962 | 91690 | 67114 | 56012 | 349086 | 165611 |
| 1997 | 4860 | 110451 | 471611 | 732959 | 408648 | 256563 | 141168 | 143166 | 143769 | 123044 | 133166 | 96058 | 176730 | 98196 | 51674 | 283110 |
| 1998 | 744 | 91505 | 184443 | 488661 | 359590 | 217571 | 153136 | 119309 | 77494 | 67072 | 50108 | 58791 | 30535 | 65839 | 57583 | 141362 |
| 1999 | 14822 | 97561 | 83715 | 176919 | 265820 | 254516 | 212217 | 187196 | 147271 | 77622 | 35582 | 22909 | 34440 | 29743 | 41830 | 122176 |
| 2000 | 565 | 66210 | 130897 | 64801 | 119297 | 232346 | 202175 | 165745 | 109218 | 54365 | 14594 | 17509 | 18642 | 18585 | 10031 | 73174 |
| 2001 | 60561 | 93125 | 204360 | 166641 | 113659 | 120410 | 141419 | 259974 | 218002 | 110319 | 38576 | 22749 | 17102 | 14092 | 18857 | 64868 |
| 2002 | 14044 | 505717 | 122603 | 158114 | 123258 | 66640 | 68890 | 95052 | 132743 | 87285 | 46167 | 29692 | 25333 | 11305 | 12753 | 72682 |
| 2003 | 1913 | 323194 | 509889 | 141442 | 148989 | 89122 | 59047 | 48582 | 52305 | 102089 | 57089 | 31748 | 27158 | 8832 | 7683 | 40641 |
| 2004 | 22237 | 159011 | 116055 | 486195 | 81099 | 98855 | 69441 | 48969 | 32589 | 51953 | 54542 | 33298 | 12581 | 13407 | 4305 | 21278 |
| 2005 | 1305 | 74538 | 171420 | 310767 | 540649 | 69957 | 74746 | 61889 | 44443 | 22726 | 27019 | 42746 | 23677 | 6849 | 7491 | 18626 |
| 2006 | 1905 | 53322 | 58091 | 75505 | 91274 | 482229 | 57377 | 37222 | 41970 | 16865 | 11828 | 17073 | 32025 | 12877 | 7464 | 24645 |
| 2007 | 5121 | 32399 | 38598 | 40530 | 61938 | 112724 | 347284 | 48160 | 29112 | 21504 | 8728 | 7015 | 8462 | 14021 | 7618 | 18335 |
| 2008 | 30155 | 78121 | 24456 | 53525 | 57125 | 84358 | 54701 | 297879 | 49889 | 36692 | 25172 | 14466 | 12787 | 9269 | 13194 | 24124 |
| 2009 | 47421 | 86053 | 31431 | 56816 | 40104 | 36174 | 62700 | 57683 | 273217 | 68318 | 42063 | 30583 | 21230 | 8266 | 6811 | 39752 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | 4331 | 68198 | 122386 | 69381 | 29371 | 30496 | 51312 | 110033 | 73973 | 285281 | 70041 | 34486 | 24421 | 14887 | 14942 | 44201 |
| 2011 | 1136 | 17035 | 61864 | 106032 | 51259 | 35380 | 38626 | 59428 | 59031 | 61017 | 239472 | 88764 | 29187 | 17731 | 9783 | 35379 |
| 2012 | 5350 | 48100 | 42653 | 64221 | 171284 | 56012 | 37917 | 28132 | 25608 | 45490 | 41255 | 162118 | 50523 | 24043 | 11621 | 30567 |
| 2013 | 94165 | 138663 | 34651 | 34171 | 76847 | 248958 | 67370 | 25070 | 18447 | 20746 | 31217 | 20836 | 106242 | 21316 | 16279 | 24536 |
| 2014 | 19215 | 26080 | 83034 | 34591 | 28200 | 62102 | 152650 | 56679 | 21786 | 16441 | 23876 | 23654 | 24509 | 57284 | 25197 | 23878 |
| 2015 | 85629 | 108174 | 25416 | 51631 | 31604 | 24613 | 46201 | 118679 | 27331 | 12698 | 10883 | 12584 | 11794 | 7272 | 48586 | 15935 |
| 2016 | 133936 | 168323 | 97368 | 18662 | 31033 | 18762 | 14519 | 22754 | 80818 | 19004 | 10531 | 10298 | 14703 | 16212 | 18451 | 62769 |
| 2017 | 104771 | 135690 | 26426 | 132175 | 34464 | 49849 | 23046 | 14115 | 22170 | 52786 | 12603 | 6491 | 6110 | 6919 | 7284 | 33718 |
| 2018 | 25736 | 107004 | 42957 | 54376 | 257565 | 43887 | 39837 | 14438 | 8809 | 19014 | 44833 | 10875 | 8065 | 4589 | 3645 | 35529 |
| 2019 | 7643 | 53043 | 59271 | 50945 | 52717 | 280292 | 42996 | 38021 | 16292 | 12752 | 19572 | 33296 | 10418 | 4690 | 3940 | 30219 |
| 2020 | 22256 | 57801 | 40360 | 50895 | 17318 | 32781 | 162029 | 19134 | 13415 | 4799 | 4292 | 5888 | 14437 | 5012 | 2647 | 11550 |
| 2021 | 2921 | 22236 | 20310 | 55002 | 91506 | 39834 | 50698 | 107895 | 14412 | 13362 | 9923 | 6880 | 5403 | 7861 | 3217 | 9047 |

Table 7.2.4.3. Western horse mackerel. Marginal age-distribution (Timing = month of year, Fleet =1 [commercial], sex = mixed, and sample size = no. samples/100).

| year | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Timing | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 |
| Fleet | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| Sex | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| catch | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| year | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sample size | 6.8 | 6.8 | 6.8 | 6.8 | 6.8 | 6.8 | 6.8 | 6.8 | 6.8 | 6.8 | 4.5 | 7.5 | 6.1 | 4.8 | 6.3 | 7.5 | 6.2 | 5.1 | 2.8 | 3.2 | 3.6 |
| 0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.013 | 0.007 | 0.000 | 0.001 | 0.001 | 0.004 | 0.001 | 0.000 | 0.008 | 0.000 | 0.036 | 0.009 |
| 1 | 0.013 | 0.022 | 0.000 | 0.002 | 0.000 | 0.000 | 0.023 | 0.000 | 0.010 | 0.015 | 0.107 | 0.058 | 0.023 | 0.065 | 0.007 | 0.033 | 0.042 | 0.054 | 0.051 | 0.056 | 0.322 |
| 2 | 0.073 | 0.006 | 0.400 | 0.006 | 0.000 | 0.000 | 0.005 | 0.000 | 0.022 | 0.035 | 0.017 | 0.049 | 0.345 | 0.179 | 0.233 | 0.140 | 0.085 | 0.046 | 0.101 | 0.123 | 0.078 |
| 3 | 0.465 | 0.090 | 0.007 | 0.717 | 0.002 | 0.000 | 0.002 | 0.013 | 0.068 | 0.011 | 0.038 | 0.007 | 0.044 | 0.146 | 0.305 | 0.217 | 0.226 | 0.098 | 0.050 | 0.100 | 0.101 |
| 4 | 0.040 | 0.147 | 0.060 | 0.005 | 0.690 | 0.003 | 0.004 | 0.012 | 0.030 | 0.099 | 0.026 | 0.028 | 0.020 | 0.075 | 0.067 | 0.121 | 0.166 | 0.147 | 0.092 | 0.068 | 0.078 |
| 5 | 0.046 | 0.042 | 0.248 | 0.050 | 0.009 | 0.801 | 0.016 | 0.003 | 0.016 | 0.060 | 0.120 | 0.047 | 0.017 | 0.021 | 0.031 | 0.076 | 0.101 | 0.141 | 0.179 | 0.072 | 0.042 |
| 6 | 0.040 | 0.122 | 0.037 | 0.119 | 0.066 | 0.002 | 0.780 | 0.009 | 0.005 | 0.031 | 0.059 | 0.144 | 0.015 | 0.032 | 0.018 | 0.042 | 0.071 | 0.118 | 0.156 | 0.085 | 0.044 |
| 7 | 0.031 | 0.144 | 0.063 | 0.015 | 0.112 | 0.037 | 0.010 | 0.814 | 0.010 | 0.013 | 0.023 | 0.063 | 0.084 | 0.010 | 0.020 | 0.042 | 0.055 | 0.104 | 0.128 | 0.156 | 0.060 |
| 8 | 0.006 | 0.049 | 0.056 | 0.019 | 0.026 | 0.082 | 0.033 | 0.008 | 0.676 | 0.011 | 0.009 | 0.030 | 0.040 | 0.063 | 0.019 | 0.043 | 0.036 | 0.082 | 0.084 | 0.131 | 0.084 |
| 9 | 0.001 | 0.009 | 0.024 | 0.007 | 0.022 | 0.011 | 0.056 | 0.038 | 0.018 | 0.639 | 0.011 | 0.024 | 0.016 | 0.031 | 0.020 | 0.036 | 0.031 | 0.043 | 0.042 | 0.066 | 0.056 |
| 10 | 0.006 | 0.015 | 0.007 | 0.011 | 0.016 | 0.009 | 0.008 | 0.053 | 0.034 | 0.013 | 0.514 | 0.014 | 0.016 | 0.027 | 0.020 | 0.039 | 0.023 | 0.020 | 0.011 | 0.023 | 0.029 |
| 11 | 0.023 | 0.009 | 0.000 | 0.000 | 0.003 | 0.017 | 0.014 | 0.009 | 0.050 | 0.008 | 0.006 | 0.476 | 0.008 | 0.017 | 0.032 | 0.028 | 0.027 | 0.013 | 0.013 | 0.014 | 0.019 |
| 12 | 0.064 | 0.047 | 0.001 | 0.001 | 0.001 | 0.008 | 0.014 | 0.015 | 0.007 | 0.027 | 0.011 | 0.005 | 0.348 | 0.018 | 0.024 | 0.052 | 0.014 | 0.019 | 0.014 | 0.010 | 0.016 |
| 13 | 0.099 | 0.065 | 0.003 | 0.004 | 0.001 | 0.001 | 0.004 | 0.009 | 0.017 | 0.005 | 0.031 | 0.011 | 0.004 | 0.264 | 0.020 | 0.029 | 0.030 | 0.016 | 0.014 | 0.008 | 0.007 |
| 14 | 0.067 | 0.105 | 0.008 | 0.001 | 0.000 | 0.000 | 0.004 | 0.006 | 0.009 | 0.004 | 0.006 | 0.030 | 0.006 | 0.011 | 0.123 | 0.015 | 0.027 | 0.023 | 0.008 | 0.011 | 0.008 |
| 15 | 0.028 | 0.127 | 0.084 | 0.041 | 0.053 | 0.030 | 0.027 | 0.012 | 0.028 | 0.016 | 0.016 | 0.014 | 0.014 | 0.041 | 0.058 | 0.084 | 0.065 | 0.068 | 0.056 | 0.039 | 0.046 |


| year | 2003* | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Timing | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 |
| Fleet | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| Sex | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| catch | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sample size | 7.9 | 6.8 | 7.8 | 7.2 | 6.2 | 7.7 | 8.7 | 7.8 | 6.2 | 6.8 | 7.7 | 8.1 | 6.4 | 8.2 | 6.8 | 6.9 | 6.6 | 5.1 | 11.1 |
| 0 | 0.001 | 0.017 | 0.001 | 0.002 | 0.006 | 0.035 | 0.052 | 0.004 | 0.001 | 0.006 | 0.096 | 0.028 | 0.134 | 0.181 | 0.157 | 0.036 | 0.011 | 0.048 | 0.006 |
| 1 | 0.196 | 0.122 | 0.050 | 0.052 | 0.040 | 0.090 | 0.095 | 0.065 | 0.019 | 0.057 | 0.142 | 0.038 | 0.169 | 0.228 | 0.203 | 0.148 | 0.074 | 0.124 | 0.048 |
| 2 | 0.309 | 0.089 | 0.114 | 0.057 | 0.048 | 0.028 | 0.035 | 0.117 | 0.068 | 0.050 | 0.035 | 0.122 | 0.040 | 0.132 | 0.040 | 0.060 | 0.083 | 0.087 | 0.044 |
| 3 | 0.086 | 0.372 | 0.207 | 0.074 | 0.051 | 0.062 | 0.063 | 0.066 | 0.116 | 0.076 | 0.035 | 0.051 | 0.081 | 0.025 | 0.198 | 0.075 | 0.071 | 0.110 | 0.118 |
| 4 | 0.090 | 0.062 | 0.361 | 0.089 | 0.077 | 0.066 | 0.044 | 0.028 | 0.056 | 0.203 | 0.078 | 0.042 | 0.049 | 0.042 | 0.052 | 0.357 | 0.074 | 0.037 | 0.197 |
| 5 | 0.054 | 0.076 | 0.047 | 0.472 | 0.141 | 0.097 | 0.040 | 0.029 | 0.039 | 0.066 | 0.254 | 0.091 | 0.039 | 0.025 | 0.075 | 0.061 | 0.391 | 0.071 | 0.086 |
| 6 | 0.036 | 0.053 | 0.050 | 0.056 | 0.433 | 0.063 | 0.069 | 0.049 | 0.042 | 0.045 | 0.069 | 0.225 | 0.072 | 0.020 | 0.034 | 0.055 | 0.060 | 0.349 | 0.109 |
| 7 | 0.029 | 0.038 | 0.041 | 0.036 | 0.060 | 0.344 | 0.063 | 0.105 | 0.065 | 0.033 | 0.026 | 0.083 | 0.186 | 0.031 | 0.021 | 0.020 | 0.053 | 0.041 | 0.232 |
| 8 | 0.032 | 0.025 | 0.030 | 0.041 | 0.036 | 0.058 | 0.301 | 0.071 | 0.065 | 0.030 | 0.019 | 0.032 | 0.043 | 0.109 | 0.033 | 0.012 | 0.023 | 0.029 | 0.031 |
| 9 | 0.062 | 0.040 | 0.015 | 0.017 | 0.027 | 0.042 | 0.075 | 0.272 | 0.067 | 0.054 | 0.021 | 0.024 | 0.020 | 0.026 | 0.079 | 0.026 | 0.018 | 0.010 | 0.029 |
| 10 | 0.035 | 0.042 | 0.018 | 0.012 | 0.011 | 0.029 | 0.046 | 0.067 | 0.263 | 0.049 | 0.032 | 0.035 | 0.017 | 0.014 | 0.019 | 0.062 | 0.027 | 0.009 | 0.021 |
| 11 | 0.019 | 0.025 | 0.029 | 0.017 | 0.009 | 0.017 | 0.034 | 0.033 | 0.097 | 0.192 | 0.021 | 0.035 | 0.020 | 0.014 | 0.010 | 0.015 | 0.046 | 0.013 | 0.015 |


| year | $2003^{*}$ | 2004 | 2005 | 2006 | 2007 | $\mathbf{2 0 0 8}$ | $\mathbf{2 0 0 9}$ | $\mathbf{2 0 1 0}$ | $\mathbf{2 0 1 1}$ | $\mathbf{2 0 1 2}$ | $\mathbf{2 0 1 3}$ | $\mathbf{2 0 1 4}$ | $\mathbf{2 0 1 5}$ | $\mathbf{2 0 1 6}$ | $\mathbf{2 0 1 7}$ | $\mathbf{2 0 1 8}$ | $\mathbf{2 0 1 9}$ | $\mathbf{2 0 2 0}$ | $\mathbf{2 0 2 1}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 12 | 0.016 | 0.010 | 0.016 | 0.031 | 0.011 | 0.015 | 0.023 | 0.023 | 0.032 | 0.060 | 0.108 | 0.036 | 0.018 | 0.020 | 0.009 | 0.011 | 0.015 | 0.031 | 0.012 |
| 13 | 0.005 | 0.010 | 0.005 | 0.013 | 0.017 | 0.011 | 0.009 | 0.014 | 0.019 | 0.028 | 0.022 | 0.084 | 0.011 | 0.022 | 0.010 | 0.006 | 0.007 | 0.011 | 0.017 |
| 14 | 0.005 | 0.003 | 0.005 | 0.007 | 0.010 | 0.015 | 0.007 | 0.014 | 0.011 | 0.014 | 0.017 | 0.037 | 0.076 | 0.025 | 0.011 | 0.005 | 0.006 | 0.006 | 0.007 |
| 15 | 0.025 | 0.016 | 0.012 | 0.024 | 0.023 | 0.028 | 0.044 | 0.042 | 0.039 | 0.036 | 0.025 | 0.035 | 0.025 | 0.085 | 0.050 | 0.049 | 0.042 | 0.025 | 0.019 |

*From 2003 the marginal age composition is replaced by the age-length key in the assessment.
Table 7.2.4.4. Western horse mackerel. Conditional age-length key.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 2 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 2 | 11 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 3 | 18 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 13 | 15 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 1 | 24 | 63 | 32 | 7 | 2 | 2 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 8 | 72 | 88 | 22 | 8 | 2 | 1 | 4 | 5 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 2 | 41 | 111 | 57 | 11 | 14 | 18 | 12 | 1 | 0 | 0 | 0 | 1 | 0 |
| 2003 | 0 | 0 | 0 | 9 | 72 | 81 | 33 | 29 | 29 | 32 | 5 | 1 | 1 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 1 | 34 | 54 | 43 | 33 | 25 | 47 | 11 | 3 | 1 | 1 | 1 | 3 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 0 | 0 | 0 | 0 | 14 | 30 | 28 | 29 | 49 | 50 | 23 | 11 | 3 | 2 | 0 | 3 |
| 2003 | 0 | 0 | 0 | 0 | 1 | 8 | 22 | 23 | 33 | 52 | 19 | 5 | 7 | 2 | 2 | 5 |
| 2003 | 0 | 0 | 0 | 0 | 1 | 3 | 4 | 4 | 15 | 29 | 29 | 13 | 2 | 3 | 2 | 17 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 2 | 7 | 15 | 10 | 8 | 6 | 2 | 3 | 5 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 7 | 8 | 5 | 7 | 2 | 2 | 8 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 1 | 3 | 6 | 2 | 2 | 0 | 4 | 4 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 3 | 1 | 2 | 2 | 5 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 2 | 1 | 0 | 0 | 8 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 1 | 1 | 10 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 8 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 3 |
| 2004 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 17 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 52 | 126 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 51 | 186 | 14 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 29 | 164 | 44 | 27 | 6 | 3 | 2 | 2 | 2 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 4 | 95 | 71 | 64 | 21 | 5 | 2 | 13 | 3 | 4 | 1 | 0 | 0 | 1 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | 0 | 0 | 2 | 28 | 65 | 108 | 35 | 9 | 6 | 10 | 11 | 4 | 0 | 0 | 0 | 1 |
| 2004 | 0 | 0 | 1 | 2 | 36 | 73 | 50 | 9 | 9 | 21 | 5 | 7 | 0 | 1 | 0 | 2 |
| 2004 | 0 | 0 | 0 | 1 | 10 | 32 | 20 | 7 | 13 | 16 | 4 | 6 | 2 | 0 | 0 | 1 |
| 2004 | 0 | 0 | 0 | 0 | 2 | 4 | 11 | 5 | 8 | 8 | 12 | 3 | 4 | 0 | 1 | 2 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 0 | 3 | 4 | 3 | 3 | 2 | 0 | 0 | 3 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 3 | 1 | 1 | 3 | 1 | 1 | 1 | 6 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 3 | 0 | 2 | 0 | 1 | 0 | 3 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 1 | 2 | 1 | 0 | 7 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 1 | 2 | 1 | 0 | 2 | 3 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 0 | 2 | 1 | 1 | 5 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 3 | 0 | 0 | 3 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 2 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| 2005 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 1 | 42 | 54 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 75 | 151 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 61 | 230 | 4 | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 30 | 248 | 22 | 17 | 7 | 4 | 3 | 2 | 3 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 18 | 160 | 40 | 35 | 7 | 8 | 7 | 7 | 6 | 2 | 0 | 2 | 1 |
| 2005 | 0 | 0 | 0 | 3 | 37 | 45 | 51 | 18 | 8 | 12 | 9 | 6 | 2 | 1 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 3 | 21 | 39 | 26 | 8 | 19 | 20 | 10 | 3 | 0 | 0 | 3 |
| 2005 | 0 | 0 | 0 | 0 | 1 | 4 | 22 | 24 | 11 | 15 | 19 | 13 | 7 | 0 | 1 | 2 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 1 | 10 | 12 | 6 | 6 | 15 | 14 | 2 | 0 | 2 | 3 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 2 | 13 | 11 | 7 | 8 | 8 | 8 | 3 | 2 | 0 | 4 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 0 | 2 | 9 | 5 | 3 | 2 | 0 | 9 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 3 | 3 | 3 | 8 | 6 | 2 | 3 | 7 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 0 | 1 | 5 | 6 | 5 | 1 | 11 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 4 | 2 | 5 | 4 | 2 | 16 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 1 | 2 | 3 | 0 | 1 | 15 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 1 | 14 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 0 | 3 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 3 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2006 | 0 | 0 | 0 | 3 | 4 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 0 | 0 | 0 | 4 | 20 | 201 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 0 | 0 | 0 | 2 | 15 | 308 | 11 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 0 | 0 | 0 | 0 | 7 | 303 | 24 | 12 | 3 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 0 | 0 | 0 | 0 | 2 | 290 | 30 | 20 | 5 | 2 | 0 | 3 | 4 | 2 | 0 | 0 |
| 2006 | 0 | 0 | 0 | 0 | 1 | 129 | 67 | 34 | 31 | 5 | 1 | 6 | 8 | 7 | 0 | 0 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 54 | 46 | 36 | 24 | 6 | 7 | 6 | 9 | 6 | 5 | 1 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 14 | 22 | 21 | 27 | 8 | 6 | 6 | 8 | 5 | 3 | 2 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 6 | 9 | 10 | 9 | 6 | 5 | 2 | 4 | 10 | 2 | 7 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 2 | 4 | 9 | 6 | 4 | 2 | 2 | 8 | 3 | 4 | 7 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 4 | 0 | 3 | 5 | 3 | 3 | 6 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 2 | 2 | 3 | 4 | 3 | 3 | 6 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 2 | 5 | 1 | 2 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 3 | 2 | 5 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 4 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 3 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2007 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2007 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2007 | 0 | 0 | 0 | 1 | 12 | 2 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2007 | 0 | 0 | 0 | 0 | 27 | 9 | 234 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2007 | 0 | 0 | 0 | 0 | 7 | 7 | 334 | 9 | 2 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 2007 | 0 | 0 | 0 | 0 | 1 | 3 | 360 | 7 | 5 | 3 | 1 | 1 | 0 | 0 | 0 | 0 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 280 | 25 | 23 | 9 | 0 | 3 | 3 | 4 | 1 | 1 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 2 | 213 | 27 | 27 | 19 | 10 | 2 | 1 | 9 | 4 | 2 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 1 | 126 | 32 | 43 | 34 | 7 | 5 | 11 | 9 | 7 | 7 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 54 | 22 | 34 | 28 | 15 | 13 | 9 | 16 | 6 | 14 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 22 | 9 | 18 | 25 | 9 | 7 | 6 | 6 | 8 | 15 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 7 | 8 | 17 | 2 | 3 | 1 | 8 | 6 | 24 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 9 | 10 | 6 | 2 | 3 | 11 | 5 | 19 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 2 | 2 | 5 | 4 | 5 | 5 | 18 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 3 | 3 | 1 | 4 | 4 | 15 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 0 | 0 | 3 | 6 | 11 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 1 | 15 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 1 | 0 | 14 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2008 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2008 | 0 | 0 | 0 | 0 | 14 | 19 | 4 | 52 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2008 | 0 | 0 | 0 | 0 | 14 | 46 | 13 | 197 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2008 | 0 | 0 | 0 | 0 | 7 | 29 | 15 | 353 | 1 | 7 | 1 | 0 | 1 | 0 | 0 | 0 |
| 2008 | 0 | 0 | 0 | 0 | 5 | 18 | 9 | 391 | 9 | 8 | 2 | 2 | 0 | 1 | 1 | 0 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 6 | 5 | 358 | 27 | 18 | 7 | 3 | 2 | 1 | 4 | 3 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 3 | 2 | 276 | 39 | 32 | 12 | 2 | 7 | 3 | 8 | 7 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 188 | 39 | 35 | 27 | 6 | 5 | 7 | 4 | 8 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 79 | 25 | 29 | 28 | 7 | 2 | 7 | 13 | 16 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 42 | 12 | 24 | 25 | 9 | 7 | 6 | 10 | 18 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14 | 9 | 25 | 19 | 5 | 5 | 6 | 5 | 28 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 4 | 9 | 12 | 4 | 3 | 4 | 6 | 34 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 8 | 11 | 6 | 7 | 3 | 4 | 20 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 8 | 4 | 6 | 0 | 10 | 18 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 4 | 2 | 0 | 1 | 7 | 26 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 1 | 1 | 0 | 0 | 3 | 23 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 2 | 13 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 4 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2009 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 0 | 0 | 0 | 5 | 4 | 6 | 1 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 0 | 0 | 0 | 6 | 24 | 36 | 25 | 8 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 0 | 0 | 0 | 0 | 23 | 64 | 67 | 26 | 167 | 5 | 2 | 3 | 0 | 0 | 0 | 0 |
| 2009 | 0 | 0 | 0 | 0 | 5 | 41 | 70 | 36 | 262 | 10 | 4 | 1 | 0 | 1 | 1 | 0 |
| 2009 | 0 | 0 | 0 | 0 | 1 | 12 | 45 | 22 | 314 | 22 | 8 | 2 | 2 | 0 | 0 | 5 |
| 2009 | 0 | 0 | 0 | 0 | 0 | 2 | 28 | 14 | 301 | 32 | 17 | 6 | 2 | 4 | 1 | 2 |
| 2009 | 0 | 0 | 0 | 0 | 0 | 1 | 11 | 5 | 229 | 38 | 17 | 17 | 6 | 1 | 2 | 9 |
| 2009 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 154 | 25 | 21 | 15 | 6 | 4 | 7 | 19 |
| 2009 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 87 | 21 | 19 | 12 | 9 | 1 | 8 | 27 |
| 2009 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 44 | 10 | 12 | 10 | 2 | 6 | 4 | 32 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2009 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 17 | 4 | 10 | 15 | 3 | 4 | 3 | 26 |
| 2009 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 7 | 13 | 11 | 4 | 3 | 0 | 17 |
| 2009 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 7 | 8 | 3 | 3 | 1 | 18 |
| 2009 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 3 | 3 | 3 | 2 | 16 |
| 2009 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 1 | 0 | 20 |
| 2009 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 0 | 1 | 0 | 11 |
| 2009 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 6 |
| 2009 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2009 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2010 | 0 | 0 | 2 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 0 | 0 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 0 | 0 | 0 | 5 | 4 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 0 | 0 | 0 | 2 | 4 | 7 | 3 | 3 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 0 | 0 | 0 | 0 | 13 | 17 | 27 | 19 | 5 | 25 | 1 | 1 | 0 | 0 | 0 | 0 |
| 2010 | 0 | 0 | 0 | 0 | 4 | 12 | 17 | 26 | 12 | 69 | 3 | 2 | 1 | 1 | 0 | 1 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 2 | 13 | 31 | 11 | 103 | 3 | 0 | 4 | 0 | 0 | 1 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 1 | 10 | 13 | 11 | 145 | 4 | 5 | 1 | 1 | 1 | 1 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 12 | 6 | 149 | 9 | 6 | 3 | 1 | 1 | 5 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2010 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 133 | 6 | 12 | 5 | 2 | 1 | 8 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 86 | 10 | 9 | 4 | 4 | 3 | 15 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 3 | 57 | 8 | 10 | 3 | 2 | 1 | 6 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 30 | 9 | 7 | 6 | 3 | 2 | 11 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 18 | 10 | 5 | 7 | 1 | 2 | 16 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 14 | 8 | 7 | 8 | 3 | 3 | 15 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 12 | 2 | 7 | 4 | 3 | 3 | 13 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 3 | 6 | 1 | 4 | 0 | 17 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 17 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 9 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 1 |
| 2011 | 0 | 0 | 7 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 0 | 0 | 20 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 0 | 0 | 17 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 0 | 0 | 10 | 52 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 0 | 0 | 9 | 51 | 4 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 0 | 0 | 8 | 33 | 17 | 4 | 2 | 1 | 2 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 0 | 0 | 4 | 15 | 21 | 18 | 8 | 7 | 5 | 2 | 10 | 1 | 1 | 0 | 0 | 0 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2011 | 0 | 0 | 0 | 2 | 18 | 23 | 15 | 17 | 14 | 5 | 28 | 2 | 0 | 0 | 0 | 2 |
| 2011 | 0 | 0 | 0 | 0 | 2 | 10 | 18 | 28 | 17 | 7 | 81 | 1 | 0 | 1 | 0 | 1 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 3 | 6 | 27 | 19 | 7 | 120 | 3 | 2 | 1 | 0 | 2 |
| 2011 | 0 | 0 | 0 | 0 | 1 | 2 | 4 | 9 | 9 | 6 | 136 | 2 | 6 | 2 | 1 | 4 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 6 | 4 | 132 | 6 | 7 | 4 | 1 | 10 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 1 | 2 | 99 | 11 | 7 | 7 | 1 | 9 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 73 | 9 | 11 | 8 | 1 | 10 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 44 | 15 | 8 | 3 | 3 | 10 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 32 | 6 | 14 | 10 | 2 | 11 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 27 | 4 | 6 | 9 | 2 | 18 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 6 | 8 | 8 | 1 | 15 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 5 | 4 | 2 | 2 | 8 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 3 | 4 | 5 | 1 | 9 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 0 | 3 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2012 | 0 | 0 | 0 | 4 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 0 | 1 | 21 | 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2012 | 0 | 0 | 0 | 20 | 51 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 0 | 0 | 10 | 92 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 0 | 0 | 4 | 107 | 14 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 0 | 0 | 0 | 97 | 28 | 3 | 2 | 1 | 2 | 0 | 1 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 0 | 0 | 2 | 74 | 27 | 16 | 2 | 6 | 5 | 0 | 15 | 1 | 0 | 1 | 0 |
| 2012 | 0 | 0 | 0 | 0 | 26 | 34 | 20 | 9 | 16 | 16 | 5 | 44 | 0 | 1 | 0 | 1 |
| 2012 | 0 | 0 | 0 | 0 | 6 | 12 | 17 | 22 | 17 | 32 | 4 | 85 | 6 | 2 | 1 | 1 |
| 2012 | 0 | 0 | 0 | 0 | 0 | 4 | 8 | 13 | 26 | 26 | 8 | 113 | 2 | 4 | 0 | 4 |
| 2012 | 0 | 0 | 0 | 0 | 0 | 1 | 5 | 9 | 8 | 12 | 13 | 119 | 3 | 5 | 3 | 2 |
| 2012 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 7 | 12 | 1 | 118 | 7 | 5 | 2 | 4 |
| 2012 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 4 | 4 | 3 | 90 | 2 | 6 | 4 | 9 |
| 2012 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 5 | 71 | 6 | 6 | 4 | 8 |
| 2012 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 3 | 55 | 8 | 6 | 4 | 11 |
| 2012 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 25 | 3 | 5 | 5 | 16 |
| 2012 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 14 | 2 | 5 | 5 | 10 |
| 2012 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 1 | 2 | 4 | 3 |
| 2012 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 2 | 3 | 3 |
| 2012 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 5 |


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| 2012 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 |
| 2013 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 0 | 1 | 6 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 1 | 2 | 18 | 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 0 | 2 | 14 | 59 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 0 | 1 | 27 | 116 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 0 | 0 | 18 | 153 | 8 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 0 | 0 | 9 | 141 | 33 | 5 | 2 | 1 | 1 | 0 | 1 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 0 | 0 | 4 | 103 | 47 | 6 | 5 | 6 | 6 | 2 | 19 | 1 | 1 | 0 |
| 2013 | 0 | 0 | 0 | 0 | 2 | 44 | 38 | 14 | 6 | 19 | 16 | 4 | 56 | 4 | 2 | 0 |
| 2013 | 0 | 0 | 0 | 0 | 0 | 11 | 20 | 13 | 14 | 26 | 18 | 2 | 90 | 5 | 6 | 3 |
| 2013 | 0 | 0 | 0 | 0 | 0 | 3 | 10 | 13 | 10 | 15 | 13 | 7 | 119 | 4 | 2 | 3 |
| 2013 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 4 | 11 | 13 | 11 | 3 | 91 | 7 | 6 | 5 |
| 2013 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 4 | 0 | 0 | 9 | 3 | 68 | 5 | 7 | 3 |
| 2013 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 2 | 60 | 3 | 4 | 8 |
| 2013 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 2 | 0 | 49 | 6 | 3 | 9 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2013 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 29 | 4 | 9 | 7 |
| 2013 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 23 | 3 | 2 | 12 |
| 2013 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 13 | 3 | 8 | 8 |
| 2013 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 2 | 7 |
| 2013 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 4 |
| 2013 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 5 |
| 2013 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2013 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2014 | 0 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 0 | 0 | 5 | 6 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 0 | 0 | 8 | 22 | 4 | 9 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 0 | 0 | 6 | 17 | 10 | 16 | 27 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 0 | 0 | 4 | 6 | 8 | 34 | 54 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 0 | 0 | 0 | 0 | 8 | 24 | 83 | 21 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 0 | 0 | 0 | 0 | 2 | 17 | 76 | 35 | 2 | 1 | 2 | 1 | 0 | 3 | 0 | 0 |
| 2014 | 0 | 0 | 0 | 0 | 0 | 8 | 65 | 30 | 7 | 6 | 3 | 5 | 5 | 9 | 1 | 0 |
| 2014 | 0 | 0 | 0 | 0 | 1 | 4 | 38 | 23 | 3 | 5 | 8 | 6 | 10 | 27 | 6 | 3 |
| 2014 | 0 | 0 | 0 | 0 | 0 | 2 | 9 | 10 | 9 | 11 | 13 | 9 | 13 | 42 | 3 | 2 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2014 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 5 | 3 | 3 | 9 | 12 | 10 | 27 | 8 | 7 |
| 2014 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 6 | 2 | 3 | 6 | 8 | 31 | 4 | 5 |
| 2014 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 2 | 5 | 24 | 2 | 6 |
| 2014 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 4 | 16 | 8 | 5 |
| 2014 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 2 | 13 | 4 | 5 |
| 2014 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 5 | 3 |
| 2014 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 2 | 3 |
| 2014 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 3 |
| 2014 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2014 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 |
| 2014 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2015 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 0 | 2 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 0 | 0 | 8 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 0 | 0 | 22 | 5 | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 0 | 0 | 15 | 22 | 4 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2015 | 0 | 0 | 0 | 8 | 12 | 13 | 11 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 0 | 0 | 5 | 16 | 9 | 11 | 43 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 0 | 0 | 3 | 4 | 3 | 18 | 82 | 3 | 1 | 1 | 0 | 0 | 0 | 1 | 0 |
| 2015 | 0 | 0 | 0 | 0 | 1 | 5 | 15 | 85 | 8 | 2 | 2 | 1 | 1 | 1 | 5 | 1 |
| 2015 | 0 | 0 | 0 | 0 | 0 | 0 | 12 | 75 | 11 | 3 | 0 | 0 | 4 | 4 | 15 | 5 |
| 2015 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 36 | 10 | 6 | 1 | 5 | 9 | 5 | 34 | 5 |
| 2015 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 20 | 7 | 4 | 5 | 7 | 9 | 3 | 51 | 7 |
| 2015 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 9 | 0 | 10 | 6 | 5 | 10 | 4 | 43 | 12 |
| 2015 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 4 | 5 | 7 | 6 | 6 | 42 | 11 |
| 2015 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 1 | 2 | 1 | 32 | 9 |
| 2015 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 3 | 2 | 18 | 4 |
| 2015 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 1 | 5 | 5 |
| 2015 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 6 | 3 |
| 2015 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 1 |
| 2015 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 |
| 2016 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2016 | 0 | 0 | 25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 22 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 21 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 16 | 13 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 9 | 14 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 0 | 10 | 13 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 0 | 3 | 12 | 4 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 0 | 0 | 5 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 0 | 0 | 4 | 0 | 2 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 0 | 0 | 1 | 1 | 4 | 4 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 12 | 12 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 8 | 15 | 2 | 0 | 0 | 1 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 7 | 15 | 4 | 1 | 1 | 2 | 2 | 7 | 4 |
| 2016 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 7 | 7 | 2 | 0 | 2 | 5 | 3 | 5 | 7 |
| 2016 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 3 | 2 | 5 | 5 | 5 | 7 |
| 2016 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 1 | 2 | 4 | 7 | 6 |
| 2016 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 3 | 6 | 5 | 7 |
| 2016 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 5 | 13 | 7 |


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| 2016 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 9 | 3 |
| 2016 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 7 |
| 2016 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 1 | 6 | 5 |
| 2016 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 |
| 2016 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| 2017 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 10 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 10 | 91 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 10 | 46 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 10 | 60 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 4 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 29 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 22 | 34 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 23 | 74 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 19 | 79 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 7 | 40 | 70 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 1 | 22 | 98 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 0 | 8 | 97 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


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| 2017 | 0 | 0 | 4 | 104 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 0 | 0 | 112 | 23 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 0 | 1 | 105 | 53 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 0 | 0 | 69 | 112 | 44 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 0 | 1 | 47 | 88 | 128 | 39 | 5 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 0 | 0 | 27 | 50 | 145 | 83 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 0 | 0 | 6 | 29 | 117 | 136 | 50 | 4 | 7 | 1 | 0 | 0 | 0 | 0 | 2 |
| 2017 | 0 | 0 | 0 | 3 | 20 | 107 | 53 | 83 | 21 | 28 | 1 | 0 | 0 | 0 | 0 | 1 |
| 2017 | 0 | 0 | 0 | 0 | 6 | 73 | 24 | 27 | 99 | 74 | 11 | 0 | 0 | 0 | 1 | 2 |
| 2017 | 0 | 0 | 0 | 0 | 3 | 33 | 13 | 7 | 46 | 137 | 14 | 1 | 2 | 2 | 2 | 5 |
| 2017 | 0 | 0 | 0 | 0 | 2 | 7 | 3 | 11 | 40 | 97 | 80 | 7 | 2 | 3 | 8 | 6 |
| 2017 | 0 | 0 | 0 | 0 | 0 | 2 | 4 | 2 | 30 | 69 | 22 | 35 | 9 | 10 | 7 | 8 |
| 2017 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 10 | 47 | 16 | 20 | 31 | 16 | 15 | 6 |
| 2017 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 9 | 16 | 7 | 12 | 16 | 16 | 17 | 5 |
| 2017 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 14 | 6 | 10 | 6 | 9 | 27 | 4 |
| 2017 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 4 | 3 | 2 | 10 | 4 | 10 | 2 |
| 2017 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 3 | 2 | 0 | 1 | 2 | 1 |
| 2017 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 3 | 1 | 1 | 1 |


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| 2017 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2017 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
| 2017 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| 2018 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 9 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 13 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 14 | 118 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 3 | 160 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 2 | 68 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 18 | 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 18 | 76 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 11 | 83 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 54 | 42 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 56 | 31 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 66 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 55 | 61 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


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| 2018 | 0 | 42 | 102 | 41 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 21 | 184 | 100 | 49 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 10 | 112 | 104 | 167 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 0 | 70 | 119 | 431 | 11 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 0 | 15 | 113 | 584 | 52 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 0 | 0 | 52 | 531 | 79 | 27 | 3 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 0 | 0 | 6 | 409 | 146 | 49 | 10 | 3 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 0 | 0 | 3 | 175 | 203 | 140 | 39 | 13 | 6 | 0 | 1 | 0 | 0 | 0 | 1 |
| 2018 | 0 | 0 | 0 | 0 | 81 | 145 | 217 | 93 | 15 | 15 | 4 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 0 | 0 | 0 | 24 | 74 | 177 | 158 | 54 | 12 | 19 | 1 | 1 | 0 | 0 | 0 |
| 2018 | 0 | 0 | 0 | 0 | 3 | 34 | 130 | 59 | 138 | 61 | 55 | 8 | 0 | 0 | 0 | 2 |
| 2018 | 0 | 0 | 0 | 0 | 3 | 15 | 78 | 25 | 43 | 139 | 121 | 30 | 9 | 4 | 3 | 13 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 3 | 41 | 40 | 16 | 65 | 229 | 39 | 16 | 8 | 4 | 40 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 2 | 13 | 12 | 14 | 40 | 192 | 116 | 33 | 10 | 8 | 62 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 7 | 4 | 27 | 102 | 63 | 91 | 27 | 18 | 106 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 16 | 62 | 21 | 70 | 47 | 32 | 115 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 6 | 26 | 15 | 16 | 15 | 45 | 135 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 8 | 8 | 7 | 11 | 128 |


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| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 5 | 1 | 4 | 7 | 3 | 79 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 4 | 6 | 5 | 37 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 32 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 9 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10 |
| 2019 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 42 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


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| 2019 | 27 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 12 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 6 | 68 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 2 | 63 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 25 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 29 | 33 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 17 | 47 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 23 | 52 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 26 | 52 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 25 | 80 | 23 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 19 | 99 | 63 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 3 | 92 | 101 | 17 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 2 | 67 | 101 | 45 | 31 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 0 | 30 | 107 | 77 | 145 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 0 | 5 | 67 | 108 | 358 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 0 | 0 | 12 | 114 | 509 | 20 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2019 | 0 | 0 | 0 | 1 | 83 | 526 | 80 | 18 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 3 |


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| 2019 | 0 | 0 | 0 | 2 | 63 | 404 | 119 | 48 | 6 | 3 | 1 | 1 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 0 | 0 | 2 | 28 | 219 | 103 | 88 | 22 | 4 | 6 | 5 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 0 | 0 | 1 | 7 | 98 | 78 | 93 | 78 | 38 | 8 | 26 | 3 | 0 | 0 | 3 |
| 2019 | 0 | 0 | 0 | 0 | 2 | 40 | 42 | 110 | 33 | 75 | 49 | 61 | 7 | 0 | 0 | 3 |
| 2019 | 0 | 0 | 0 | 0 | 0 | 14 | 24 | 75 | 19 | 22 | 110 | 96 | 12 | 5 | 2 | 14 |
| 2019 | 0 | 0 | 0 | 0 | 0 | 2 | 8 | 53 | 17 | 11 | 54 | 136 | 29 | 3 | 2 | 38 |
| 2019 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 25 | 15 | 8 | 17 | 88 | 68 | 22 | 7 | 56 |
| 2019 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 10 | 9 | 8 | 15 | 45 | 35 | 37 | 21 | 71 |
| 2019 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 1 | 8 | 24 | 10 | 12 | 34 | 60 |
| 2019 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 1 | 2 | 13 | 8 | 3 | 11 | 71 |
| 2019 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 2 | 8 | 2 | 4 | 2 | 54 |
| 2019 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 2 | 0 | 34 |
| 2019 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 18 |
| 2019 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 8 |
| 2019 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 9 |
| 2019 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2019 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2020 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


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| 2020 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 50 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 34 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 38 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 56 | 29 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 24 | 107 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 4 | 203 | 40 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 4 | 136 | 75 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 0 | 97 | 111 | 10 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 0 | 21 | 109 | 16 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 0 | 12 | 89 | 66 | 23 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


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| 2020 | 0 | 0 | 0 | 58 | 76 | 35 | 83 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 0 | 1 | 24 | 69 | 60 | 185 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 0 | 0 | 1 | 40 | 101 | 333 | 25 | 3 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 0 | 0 | 3 | 6 | 121 | 321 | 31 | 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 0 | 0 | 0 | 5 | 58 | 322 | 68 | 24 | 2 | 4 | 0 | 4 | 0 | 0 | 0 |
| 2020 | 0 | 0 | 0 | 0 | 4 | 23 | 197 | 102 | 49 | 15 | 8 | 10 | 12 | 0 | 0 | 0 |
| 2020 | 1 | 0 | 0 | 0 | 0 | 4 | 74 | 62 | 113 | 18 | 10 | 19 | 41 | 5 | 0 | 6 |
| 2020 | 0 | 0 | 0 | 0 | 0 | 1 | 21 | 29 | 72 | 99 | 15 | 18 | 54 | 2 | 3 | 16 |
| 2020 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 21 | 39 | 35 | 77 | 24 | 56 | 8 | 4 | 28 |
| 2020 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 9 | 24 | 16 | 40 | 25 | 36 | 11 | 3 | 33 |
| 2020 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 15 | 9 | 19 | 8 | 27 | 24 | 15 | 4 | 39 |
| 2020 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 8 | 9 | 5 | 8 | 15 | 31 | 8 | 1 | 28 |
| 2020 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 2 | 6 | 3 | 6 | 6 | 13 | 10 | 16 |
| 2020 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 5 | 6 | 0 | 0 | 8 | 12 |
| 2020 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 4 | 4 | 0 | 0 | 10 |
| 2020 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 5 |
| 2020 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| 2021 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


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| 2021 | 21 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 23 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 28 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 32 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 31 | 66 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 28 | 122 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 4 | 189 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 0 | 152 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 4 | 96 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 7 | 98 | 31 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 7 | 56 | 30 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 1 | 31 | 60 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 1 | 24 | 81 | 33 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 0 | 15 | 125 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 0 | 5 | 115 | 71 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 0 | 5 | 66 | 120 | 18 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 0 | 0 | 41 | 111 | 87 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


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| 2021 | 0 | 0 | 12 | 123 | 160 | 55 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 0 | 0 | 2 | 85 | 264 | 69 | 25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 0 | 0 | 0 | 20 | 261 | 138 | 55 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 0 | 0 | 0 | 13 | 117 | 170 | 159 | 45 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 0 | 0 | 1 | 16 | 93 | 73 | 191 | 139 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 0 | 0 | 0 | 9 | 41 | 79 | 153 | 230 | 16 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 0 | 0 | 0 | 1 | 15 | 58 | 79 | 402 | 58 | 19 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 0 | 0 | 0 | 0 | 11 | 39 | 98 | 295 | 115 | 84 | 15 | 2 | 5 | 10 | 1 | 0 |
| 2021 | 0 | 0 | 0 | 1 | 6 | 17 | 74 | 265 | 81 | 152 | 31 | 2 | 14 | 15 | 0 | 3 |
| 2021 | 0 | 0 | 0 | 0 | 0 | 10 | 38 | 146 | 38 | 118 | 68 | 15 | 4 | 23 | 1 | 3 |
| 2021 | 0 | 0 | 0 | 0 | 0 | 3 | 19 | 61 | 19 | 44 | 69 | 38 | 12 | 23 | 5 | 17 |
| 2021 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 17 | 17 | 23 | 46 | 54 | 30 | 30 | 6 | 24 |
| 2021 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 14 | 17 | 19 | 16 | 10 | 59 | 54 | 15 | 27 |
| 2021 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 7 | 14 | 7 | 9 | 26 | 75 | 45 | 23 |
| 2021 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 5 | 6 | 4 | 3 | 22 | 78 | 13 |
| 2021 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 10 | 2 | 51 | 9 |
| 2021 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 12 | 1 |
| 2021 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 6 |


|  | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2021 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| 2021 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 |

## Table 7.2.4.5. Western horse mackerel. Catch-at-length distribution from the commercial fleet.

| year |  | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Timing |  | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 |
| Fleet |  | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| Sex |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| catch |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sample number |  | 42 | 50 | 40 | 47 | 53 | 57 | 37 | 46 | 87 | 68 | 49 | 48 | 66 | 63 | 82 | 101 | 108 | 104 | 96 | 51 | 111 |
| Length bins (cm) | 5 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 6 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 7 | 0.004 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 8 | 0.003 | 0.003 | 0.000 | 0.000 | 0.000 | 0.000 | 0.003 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 9 | 0.001 | 0.006 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.030 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 10 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.059 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 11 | 0.009 | 0.007 | 0.000 | 0.002 | 0.002 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.037 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 12 | 0.035 | 0.034 | 0.000 | 0.010 | 0.004 | 0.002 | 0.001 | 0.003 | 0.000 | 0.002 | 0.000 | 0.000 | 0.001 | 0.000 | 0.020 | 0.004 | 0.000 | 0.001 | 0.004 | 0.002 | 0.000 |
|  | 13 | 0.014 | 0.055 | 0.001 | 0.018 | 0.003 | 0.002 | 0.002 | 0.003 | 0.002 | 0.005 | 0.000 | 0.000 | 0.004 | 0.000 | 0.016 | 0.007 | 0.002 | 0.007 | 0.011 | 0.016 | 0.002 |
|  | 14 | 0.008 | 0.045 | 0.002 | 0.016 | 0.007 | 0.004 | 0.002 | 0.004 | 0.044 | 0.006 | 0.001 | 0.001 | 0.020 | 0.000 | 0.010 | 0.009 | 0.028 | 0.016 | 0.017 | 0.015 | 0.007 |
|  | 15 | 0.016 | 0.039 | 0.007 | 0.022 | 0.017 | 0.007 | 0.001 | 0.033 | 0.054 | 0.010 | 0.003 | 0.002 | 0.048 | 0.001 | 0.012 | 0.014 | 0.017 | 0.026 | 0.016 | 0.003 | 0.009 |
|  | 16 | 0.024 | 0.040 | 0.011 | 0.029 | 0.014 | 0.010 | 0.004 | 0.045 | 0.012 | 0.009 | 0.004 | 0.005 | 0.067 | 0.002 | 0.012 | 0.012 | 0.010 | 0.010 | 0.009 | 0.004 | 0.012 |


| year |  | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 17 | 0.042 | 0.049 | 0.011 | 0.020 | 0.006 | 0.014 | 0.008 | 0.021 | 0.008 | 0.009 | 0.010 | 0.009 | 0.052 | 0.002 | 0.008 | 0.018 | 0.010 | 0.003 | 0.008 | 0.011 | 0.010 |
|  | 18 | 0.044 | 0.054 | 0.016 | 0.025 | 0.007 | 0.013 | 0.012 | 0.020 | 0.014 | 0.009 | 0.017 | 0.009 | 0.043 | 0.003 | 0.011 | 0.019 | 0.022 | 0.008 | 0.005 | 0.016 | 0.010 |
|  | 19 | 0.044 | 0.037 | 0.021 | 0.035 | 0.012 | 0.012 | 0.012 | 0.008 | 0.024 | 0.010 | 0.017 | 0.022 | 0.026 | 0.006 | 0.024 | 0.028 | 0.027 | 0.013 | 0.011 | 0.019 | 0.015 |
|  | 20 | 0.052 | 0.030 | 0.031 | 0.042 | 0.018 | 0.012 | 0.024 | 0.009 | 0.036 | 0.026 | 0.016 | 0.034 | 0.022 | 0.015 | 0.024 | 0.047 | 0.029 | 0.029 | 0.018 | 0.019 | 0.013 |
|  | 21 | 0.061 | 0.033 | 0.027 | 0.091 | 0.054 | 0.023 | 0.036 | 0.014 | 0.019 | 0.057 | 0.030 | 0.046 | 0.022 | 0.025 | 0.021 | 0.055 | 0.043 | 0.051 | 0.030 | 0.046 | 0.027 |
|  | 22 | 0.072 | 0.031 | 0.027 | 0.109 | 0.120 | 0.039 | 0.076 | 0.044 | 0.024 | 0.062 | 0.041 | 0.035 | 0.022 | 0.028 | 0.019 | 0.041 | 0.060 | 0.069 | 0.038 | 0.034 | 0.029 |
|  | 23 | 0.098 | 0.034 | 0.032 | 0.117 | 0.120 | 0.086 | 0.123 | 0.065 | 0.032 | 0.044 | 0.048 | 0.039 | 0.026 | 0.024 | 0.026 | 0.023 | 0.072 | 0.121 | 0.038 | 0.030 | 0.039 |
|  | 24 | 0.112 | 0.054 | 0.026 | 0.092 | 0.113 | 0.161 | 0.102 | 0.067 | 0.031 | 0.034 | 0.059 | 0.049 | 0.026 | 0.026 | 0.031 | 0.016 | 0.065 | 0.135 | 0.053 | 0.047 | 0.048 |
|  | 25 | 0.087 | 0.077 | 0.029 | 0.088 | 0.084 | 0.139 | 0.109 | 0.081 | 0.037 | 0.033 | 0.051 | 0.072 | 0.045 | 0.030 | 0.032 | 0.022 | 0.058 | 0.109 | 0.097 | 0.021 | 0.059 |
|  | 26 | 0.069 | 0.063 | 0.040 | 0.069 | 0.071 | 0.086 | 0.114 | 0.101 | 0.049 | 0.041 | 0.041 | 0.076 | 0.075 | 0.036 | 0.031 | 0.026 | 0.039 | 0.077 | 0.126 | 0.041 | 0.065 |
|  | 27 | 0.059 | 0.044 | 0.071 | 0.063 | 0.058 | 0.068 | 0.099 | 0.110 | 0.084 | 0.067 | 0.050 | 0.066 | 0.087 | 0.060 | 0.038 | 0.033 | 0.042 | 0.048 | 0.132 | 0.103 | 0.075 |
|  | 28 | 0.043 | 0.032 | 0.094 | 0.042 | 0.048 | 0.049 | 0.069 | 0.097 | 0.105 | 0.092 | 0.055 | 0.052 | 0.076 | 0.102 | 0.060 | 0.037 | 0.050 | 0.033 | 0.103 | 0.171 | 0.102 |
|  | 29 | 0.027 | 0.026 | 0.106 | 0.031 | 0.038 | 0.034 | 0.048 | 0.072 | 0.098 | 0.119 | 0.083 | 0.064 | 0.058 | 0.118 | 0.075 | 0.060 | 0.056 | 0.032 | 0.067 | 0.117 | 0.113 |
|  | 30 | 0.021 | 0.025 | 0.107 | 0.019 | 0.028 | 0.024 | 0.030 | 0.053 | 0.066 | 0.106 | 0.117 | 0.087 | 0.050 | 0.112 | 0.093 | 0.083 | 0.069 | 0.032 | 0.050 | 0.091 | 0.116 |
|  | 31 | 0.014 | 0.021 | 0.111 | 0.014 | 0.024 | 0.017 | 0.020 | 0.041 | 0.043 | 0.078 | 0.101 | 0.094 | 0.054 | 0.109 | 0.095 | 0.092 | 0.074 | 0.039 | 0.042 | 0.052 | 0.087 |
|  | 32 | 0.012 | 0.023 | 0.098 | 0.008 | 0.019 | 0.022 | 0.016 | 0.033 | 0.035 | 0.062 | 0.072 | 0.073 | 0.046 | 0.096 | 0.063 | 0.098 | 0.066 | 0.039 | 0.034 | 0.033 | 0.055 |
|  | 33 | 0.009 | 0.025 | 0.047 | 0.009 | 0.021 | 0.028 | 0.013 | 0.023 | 0.033 | 0.041 | 0.052 | 0.055 | 0.035 | 0.077 | 0.063 | 0.088 | 0.057 | 0.032 | 0.032 | 0.029 | 0.030 |
|  | 34 | 0.008 | 0.029 | 0.027 | 0.010 | 0.024 | 0.031 | 0.014 | 0.016 | 0.032 | 0.026 | 0.043 | 0.036 | 0.025 | 0.047 | 0.029 | 0.069 | 0.045 | 0.028 | 0.025 | 0.028 | 0.022 |


| year |  | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 35 | 0.004 | 0.027 | 0.004 | 0.008 | 0.027 | 0.035 | 0.016 | 0.010 | 0.036 | 0.020 | 0.031 | 0.025 | 0.020 | 0.030 | 0.021 | 0.041 | 0.028 | 0.018 | 0.017 | 0.021 | 0.018 |
|  | 36 | 0.003 | 0.022 | 0.023 | 0.006 | 0.020 | 0.027 | 0.013 | 0.009 | 0.029 | 0.011 | 0.020 | 0.018 | 0.015 | 0.019 | 0.010 | 0.028 | 0.015 | 0.010 | 0.009 | 0.016 | 0.015 |
|  | 37 | 0.001 | 0.014 | 0.018 | 0.006 | 0.014 | 0.020 | 0.011 | 0.007 | 0.021 | 0.007 | 0.014 | 0.013 | 0.014 | 0.012 | 0.006 | 0.014 | 0.008 | 0.005 | 0.005 | 0.007 | 0.009 |
|  | 38 | 0.001 | 0.008 | 0.006 | 0.002 | 0.013 | 0.017 | 0.010 | 0.004 | 0.012 | 0.005 | 0.009 | 0.007 | 0.010 | 0.007 | 0.005 | 0.005 | 0.003 | 0.003 | 0.003 | 0.004 | 0.005 |
|  | 39 | 0.000 | 0.005 | 0.004 | 0.001 | 0.006 | 0.008 | 0.005 | 0.003 | 0.009 | 0.004 | 0.005 | 0.003 | 0.005 | 0.006 | 0.002 | 0.003 | 0.002 | 0.001 | 0.001 | 0.002 | 0.003 |
|  | 40 | 0.000 | 0.004 | 0.000 | 0.000 | 0.005 | 0.006 | 0.004 | 0.002 | 0.005 | 0.003 | 0.004 | 0.005 | 0.002 | 0.003 | 0.002 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.003 |
|  | 41 | 0.000 | 0.002 | 0.000 | 0.000 | 0.002 | 0.003 | 0.002 | 0.001 | 0.003 | 0.002 | 0.002 | 0.001 | 0.003 | 0.001 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.001 |
|  | 42 | 0.000 | 0.001 | 0.000 | 0.000 | 0.001 | 0.001 | 0.001 | 0.000 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 43 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.001 | 0.001 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 44 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 45 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 |
|  | 46 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 47 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 48 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 49 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 50 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 51 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Table 7.2.4.6. Western horse mackerel. Catch-at-length distribution from the PELACUS survey (fleet 5).

| year |  | 1995 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Timing |  | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 |
| Sex |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| catch |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sample number |  | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 |
| Length bins (cm) | 5 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 6 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 7 | 0.000 | 0.000 | 0.000 | 0.003 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.000 |
|  | 8 | 0.000 | 0.000 | 0.000 | 0.012 | 0.000 | 0.000 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.003 | 0.000 | 0.000 | 0.000 |
|  | 9 | 0.000 | 0.000 | 0.000 | 0.038 | 0.000 | 0.000 | 0.002 | 0.000 | 0.000 | 0.024 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.007 | 0.002 | 0.000 | 0.000 |
|  | 10 | 0.000 | 0.000 | 0.000 | 0.055 | 0.000 | 0.000 | 0.207 | 0.000 | 0.004 | 0.148 | 0.000 | 0.000 | 0.004 | 0.000 | 0.049 | 0.000 | 0.047 | 0.017 | 0.003 | 0.002 |
|  | 11 | 0.002 | 0.000 | 0.002 | 0.006 | 0.014 | 0.000 | 0.257 | 0.000 | 0.006 | 0.113 | 0.000 | 0.000 | 0.009 | 0.003 | 0.058 | 0.009 | 0.112 | 0.101 | 0.077 | 0.058 |
|  | 12 | 0.043 | 0.017 | 0.009 | 0.002 | 0.046 | 0.000 | 0.092 | 0.000 | 0.001 | 0.025 | 0.000 | 0.000 | 0.024 | 0.015 | 0.108 | 0.014 | 0.097 | 0.068 | 0.144 | 0.110 |
|  | 13 | 0.066 | 0.028 | 0.016 | 0.002 | 0.025 | 0.000 | 0.063 | 0.000 | 0.000 | 0.007 | 0.001 | 0.000 | 0.080 | 0.012 | 0.126 | 0.003 | 0.060 | 0.081 | 0.096 | 0.073 |
|  | 14 | 0.047 | 0.084 | 0.013 | 0.000 | 0.006 | 0.000 | 0.038 | 0.000 | 0.000 | 0.009 | 0.000 | 0.001 | 0.083 | 0.003 | 0.095 | 0.009 | 0.034 | 0.087 | 0.038 | 0.029 |
|  | 15 | 0.029 | 0.140 | 0.005 | 0.000 | 0.019 | 0.000 | 0.018 | 0.000 | 0.000 | 0.017 | 0.004 | 0.003 | 0.020 | 0.001 | 0.035 | 0.053 | 0.014 | 0.124 | 0.051 | 0.039 |
|  | 16 | 0.018 | 0.123 | 0.000 | 0.000 | 0.025 | 0.000 | 0.005 | 0.000 | 0.001 | 0.034 | 0.020 | 0.004 | 0.027 | 0.011 | 0.007 | 0.165 | 0.017 | 0.184 | 0.068 | 0.052 |
|  | 17 | 0.079 | 0.089 | 0.001 | 0.000 | 0.018 | 0.000 | 0.002 | 0.017 | 0.000 | 0.020 | 0.018 | 0.001 | 0.023 | 0.039 | 0.012 | 0.144 | 0.106 | 0.130 | 0.081 | 0.062 |


| year |  | 1995 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 18 | 0.148 | 0.045 | 0.005 | 0.000 | 0.003 | 0.000 | 0.004 | 0.024 | 0.000 | 0.012 | 0.019 | 0.003 | 0.021 | 0.066 | 0.020 | 0.059 | 0.120 | 0.039 | 0.091 | 0.069 |
|  | 19 | 0.163 | 0.073 | 0.005 | 0.000 | 0.001 | 0.000 | 0.002 | 0.019 | 0.001 | 0.001 | 0.017 | 0.012 | 0.020 | 0.081 | 0.022 | 0.059 | 0.076 | 0.029 | 0.072 | 0.055 |
|  | 20 | 0.083 | 0.008 | 0.005 | 0.000 | 0.007 | 0.000 | 0.005 | 0.016 | 0.018 | 0.002 | 0.009 | 0.057 | 0.024 | 0.195 | 0.036 | 0.057 | 0.043 | 0.036 | 0.039 | 0.030 |
|  | 21 | 0.032 | 0.031 | 0.007 | 0.002 | 0.012 | 0.000 | 0.013 | 0.018 | 0.126 | 0.002 | 0.047 | 0.117 | 0.013 | 0.235 | 0.053 | 0.059 | 0.034 | 0.032 | 0.050 | 0.039 |
|  | 22 | 0.012 | 0.017 | 0.003 | 0.007 | 0.007 | 0.002 | 0.010 | 0.030 | 0.123 | 0.008 | 0.087 | 0.171 | 0.011 | 0.089 | 0.059 | 0.052 | 0.031 | 0.028 | 0.032 | 0.026 |
|  | 23 | 0.014 | 0.026 | 0.007 | 0.035 | 0.023 | 0.004 | 0.004 | 0.056 | 0.129 | 0.026 | 0.073 | 0.142 | 0.022 | 0.039 | 0.083 | 0.073 | 0.035 | 0.024 | 0.019 | 0.027 |
|  | 24 | 0.028 | 0.032 | 0.011 | 0.066 | 0.064 | 0.025 | 0.008 | 0.073 | 0.078 | 0.035 | 0.072 | 0.070 | 0.026 | 0.009 | 0.100 | 0.061 | 0.031 | 0.012 | 0.027 | 0.058 |
|  | 25 | 0.042 | 0.053 | 0.003 | 0.076 | 0.125 | 0.109 | 0.047 | 0.098 | 0.083 | 0.063 | 0.071 | 0.064 | 0.024 | 0.034 | 0.068 | 0.053 | 0.021 | 0.001 | 0.024 | 0.056 |
|  | 26 | 0.042 | 0.040 | 0.008 | 0.039 | 0.123 | 0.244 | 0.083 | 0.179 | 0.136 | 0.087 | 0.090 | 0.086 | 0.038 | 0.028 | 0.026 | 0.045 | 0.028 | 0.000 | 0.020 | 0.033 |
|  | 27 | 0.025 | 0.042 | 0.029 | 0.029 | 0.109 | 0.293 | 0.074 | 0.134 | 0.141 | 0.091 | 0.136 | 0.083 | 0.048 | 0.027 | 0.011 | 0.039 | 0.027 | 0.000 | 0.013 | 0.026 |
|  | 28 | 0.023 | 0.030 | 0.099 | 0.044 | 0.084 | 0.141 | 0.037 | 0.098 | 0.058 | 0.088 | 0.103 | 0.076 | 0.077 | 0.016 | 0.007 | 0.017 | 0.022 | 0.001 | 0.013 | 0.026 |
|  | 29 | 0.031 | 0.044 | 0.212 | 0.146 | 0.094 | 0.089 | 0.015 | 0.097 | 0.037 | 0.069 | 0.077 | 0.051 | 0.127 | 0.027 | 0.007 | 0.009 | 0.013 | 0.001 | 0.009 | 0.025 |
|  | 30 | 0.029 | 0.047 | 0.275 | 0.179 | 0.100 | 0.062 | 0.008 | 0.061 | 0.029 | 0.059 | 0.056 | 0.039 | 0.134 | 0.021 | 0.003 | 0.002 | 0.007 | 0.001 | 0.012 | 0.032 |
|  | 31 | 0.017 | 0.016 | 0.166 | 0.120 | 0.067 | 0.021 | 0.001 | 0.041 | 0.022 | 0.033 | 0.042 | 0.014 | 0.080 | 0.013 | 0.006 | 0.000 | 0.002 | 0.000 | 0.012 | 0.032 |
|  | 32 | 0.009 | 0.017 | 0.078 | 0.062 | 0.016 | 0.008 | 0.001 | 0.028 | 0.005 | 0.017 | 0.040 | 0.004 | 0.047 | 0.016 | 0.005 | 0.003 | 0.003 | 0.000 | 0.005 | 0.014 |
|  | 33 | 0.005 | 0.000 | 0.024 | 0.029 | 0.010 | 0.002 | 0.000 | 0.006 | 0.003 | 0.009 | 0.014 | 0.002 | 0.014 | 0.008 | 0.003 | 0.002 | 0.004 | 0.000 | 0.001 | 0.004 |
|  | 34 | 0.004 | 0.000 | 0.009 | 0.021 | 0.003 | 0.000 | 0.000 | 0.002 | 0.000 | 0.002 | 0.003 | 0.000 | 0.006 | 0.009 | 0.001 | 0.001 | 0.002 | 0.003 | 0.001 | 0.002 |
|  | 35 | 0.004 | 0.000 | 0.004 | 0.012 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.001 | 0.000 | 0.000 | 0.001 | 0.002 | 0.001 | 0.004 | 0.001 | 0.000 | 0.000 | 0.000 |


| year |  | 1995 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 36 | 0.002 | 0.000 | 0.003 | 0.011 | 0.000 | 0.000 | 0.000 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.003 | 0.001 | 0.000 | 0.000 | 0.000 |
|  | 37 | 0.001 | 0.000 | 0.001 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.001 | 0.001 | 0.000 | 0.000 |
|  | 38 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 39 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 40 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 41 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 42 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 43 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 44 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 45 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.006 |
|  | 46 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.005 |
|  | 47 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.004 |
|  | 48 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.003 |
|  | 49 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 |
|  | 50 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 |
|  | 51 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Table 7.2.5.1. Western horse mackerel stock. Mean weight ( kg ) in catch-at-age by quarter and area in 2021 ( 15 = 15+ group)

| Q1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.6.a | 27.7.b | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.7.k. 1 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | NA |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.027 | 0.022 | 0.029 | 0.032 | 0.029 | 0.027 | 0.027 |
| 2 | 0 | 0 | 0.053 | 0.053 | 0.055 | 0.054 | 0.048 | 0.051 | 0.053 | 0.053 | 0.053 | 0.047 | 0.045 | 0.054 | 0.051 | 0.042 | 0.047 | 0.051 |
| 3 | 0.103 | 0.103 | 0.080 | 0.078 | 0.079 | 0.079 | 0.076 | 0.078 | 0.078 | 0.078 | 0.078 | 0.080 | 0.065 | 0.064 | 0.071 | 0.102 | 0.080 | 0.078 |
| 4 | 0.119 | 0.119 | 0.116 | 0.106 | 0.107 | 0.104 | 0.107 | 0.104 | 0.106 | 0.110 | 0.106 | 0.128 | 0.128 | 0 | 0.133 | 0.128 | 0.128 | 0.107 |
| 5 | 0.180 | 0.180 | 0.160 | 0.146 | 0.134 | 0.158 | 0.137 | 0.134 | 0.143 | 0.164 | 0.143 | 0.151 | 0.151 | 0 | 0.158 | 0.149 | 0.151 | 0.148 |
| 6 | 0.194 | 0.194 | 0.198 | 0.192 | 0.155 | 0.192 | 0.161 | 0.179 | 0.192 | 0.239 | 0.192 | 0.173 | 0.173 | 0 | 0.178 | 0.172 | 0.173 | 0.191 |
| 7 | 0.213 | 0.213 | 0.196 | 0.203 | 0.185 | 0.207 | 0.186 | 0.196 | 0.202 | 0.214 | 0.202 | 0.201 | 0.207 | 0 | 0.209 | 0.195 | 0.201 | 0.206 |
| 8 | 0.256 | 0.256 | 0.253 | 0.257 | 0.239 | 0.247 | 0.221 | 0.249 | 0.256 | 0.271 | 0.256 | 0.265 | 0.265 | 0 | 0.287 | 0.238 | 0.265 | 0.258 |
| 9 | 0.284 | 0.284 | 0.269 | 0.264 | 0.264 | 0.307 | 0.266 | 0.271 | 0.264 | 0.246 | 0.264 | 0.310 | 0.299 | 0 | 0.330 | 0.264 | 0.310 | 0.279 |
| 10 | 0.325 | 0.325 | 0.318 | 0.302 | 0.289 | 0.305 | 0.303 | 0.350 | 0.305 | 0.244 | 0.305 | 0.305 | 0.293 | 0 | 0.326 | 0.271 | 0.305 | 0.309 |
| 11 | 0.343 | 0.343 | 0.298 | 0.295 | 0.295 | 0.295 | 0.301 | 0.296 | 0.295 | 0.277 | 0.295 | 0.335 | 0.324 | 0 | 0.354 | 0.296 | 0.335 | 0.324 |
| 12 | 0.334 | 0.334 | 0.304 | 0.293 | 0.272 | 0.272 | 0.258 | 0.272 | 0.272 | 0.242 | 0.272 | 0.337 | 0.335 | 0 | 0.344 | 0.322 | 0.337 | 0.311 |
| 13 | 0.322 | 0.322 | 0.327 | 0.306 | 0.221 | 0.384 | 0.318 | 0.294 | 0.293 | 0.262 | 0.293 | 0.392 | 0.377 | 0 | 0.406 | 0.319 | 0.392 | 0.312 |
| 14 | 0.340 | 0.340 | 0.265 | 0.265 | 0.265 | 0.265 | 0.265 | 0.265 | 0.265 | 0.265 | 0.265 | 0.386 | 0.386 | 0 | 0.388 | 0.380 | 0.386 | 0.341 |
| 15+ | 0.346 | 0.346 | 0.399 | 0.358 | 0.311 | 0.478 | 0.411 | 0.505 | 0.393 | 0.335 | 0.393 | 0.448 | 0.447 | 0 | 0.438 | 0.478 | 0.448 | 0.377 |


| Q2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | NA |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 | 0.103 | 0.103 | 0.103 | 0.103 | 0.103 | 0.103 | 0.103 | 0.103 | 0.103 | 0.103 | 0.103 | 0.092 | 0.076 | 0.072 | 0.114 | 0.095 | 0.093 | 0.094 |
| 4 | 0.119 | 0.213 | 0.213 | 0.213 | 0.213 | 0.213 | 0.213 | 0.213 | 0.213 | 0.213 | 0.213 | 0.119 | 0.120 | 0.105 | 0.126 | 0.115 | 0.119 | 0.120 |
| 5 | 0.189 | 0.211 | 0.211 | 0.211 | 0.211 | 0.211 | 0.211 | 0.211 | 0.211 | 0.211 | 0.211 | 0.144 | 0.144 | 0.121 | 0.147 | 0.131 | 0.144 | 0.145 |
| 6 | 0.197 | 0.257 | 0.257 | 0.257 | 0.257 | 0.257 | 0.257 | 0.257 | 0.257 | 0.257 | 0.257 | 0.165 | 0.167 | 0.125 | 0.166 | 0.142 | 0.165 | 0.166 |
| 7 | 0.215 | 0.279 | 0.279 | 0.279 | 0.279 | 0.279 | 0.279 | 0.279 | 0.279 | 0.279 | 0.279 | 0.195 | 0.196 | 0.173 | 0.180 | 0.225 | 0.184 | 0.189 |
| 8 | 0.274 | 0.356 | 0.356 | 0.356 | 0.356 | 0.356 | 0.356 | 0.356 | 0.356 | 0.356 | 0.356 | 0.248 | 0.280 | 0.371 | 0.224 | 0.277 | 0.237 | 0.241 |
| 9 | 0.288 | 0.372 | 0.372 | 0.372 | 0.372 | 0.372 | 0.372 | 0.372 | 0.372 | 0.372 | 0.372 | 0.278 | 0.324 | 0.393 | 0.271 | 0.296 | 0.282 | 0.291 |
| 10 | 0.318 | 0.293 | 0.293 | 0.293 | 0.293 | 0.293 | 0.293 | 0.293 | 0.293 | 0.293 | 0.293 | 0.280 | 0.318 | 0.407 | 0.279 | 0.280 | 0.283 | 0.283 |
| 11 | 0.336 | 0.385 | 0.385 | 0.385 | 0.385 | 0.385 | 0.385 | 0.385 | 0.385 | 0.385 | 0.385 | 0.312 | 0.356 | 0.383 | 0.319 | 0.304 | 0.318 | 0.322 |
| 12 | 0.332 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.329 | 0.342 | 0.341 | 0.333 | 0.325 | 0.330 | 0.334 |
| 13 | 0.324 | 0.465 | 0.465 | 0.465 | 0.465 | 0.465 | 0.465 | 0.465 | 0.465 | 0.465 | 0.465 | 0.333 | 0.416 | 0.486 | 0.413 | 0.353 | 0.390 | 0.401 |
| 14 | 0.339 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.383 | 0.387 | 0.391 | 0.388 | 0.378 | 0.383 | 0.383 |
| 15+ | 0.353 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.429 | 0.471 | 0.465 | 0.466 | 0.473 | 0.469 | 0.469 |


| $\stackrel{\text { Q3 }}{\substack{\text { Ages }}}$ | 27.2.a | 27.3.a | 27.4.a | 27.6.a | 27.7.a | 27.7.b | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.7. 2 | 27.7.k. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | 27.8.d. 2 | 27.8.e | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| , | 0 | , | 0 | . | 0 | 0 | 0 | 0 | , | 0 | . | 0 | 0 | 0 | 0 | 0 | 0 | 0 | , | 0 | . | 0 | , |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 | 0 | 0 | 0 | 0 | 0.173 | 0.173 | 0.173 | 0.171 | 0.173 | 0.173 | 0.173 | 0.180 | 0.161 | 0.173 | 0.120 | 0.120 | 0.120 | 0.115 | 0.129 | 0.120 | 0.120 | 0.120 | 0.145 |
| 4 | 0.210 | 0.210 | 0.210 | 0.210 | 0.192 | 0.192 | 0.192 | 0.190 | 0.192 | 0.192 | 0.192 | 0.192 | 0.192 | 0.192 | 0.148 | 0.149 | 0.148 | 0.148 | 0.148 | 0.148 | 0.148 | 0.148 | 0.173 |
| 5 | 0.269 | 0.269 | 0.269 | 0.269 | 0.231 | 0.231 | 0.231 | 0.230 | 0.231 | 0.231 | 0.231 | 0.234 | 0.229 | 0.231 | 0.176 | 0.181 | 0.198 | 0.173 | 0.178 | 0.176 | 0.176 | 0.176 | 0.199 |
| 6 | 0.277 | 0.277 | 0.277 | 0.277 | 0.244 | 0.244 | 0.244 | 0.244 | 0.244 | 0.244 | 0.244 | 0.245 | 0.243 | 0.244 | 0.193 | 0.197 | 0.220 | 0.192 | 0.193 | 0.193 | 0.193 | 0.193 | 0.219 |
| 7 | 0.288 | 0.288 | 0.288 | 0.288 | 0.255 | 0.255 | 0.255 | 0.257 | 0.255 | 0.255 | 0.255 | 0.251 | 0.256 | 0.255 | 0.211 | 0.223 | 0.239 | 0.209 | 0.211 | 0.211 | 0.211 | 0.211 | 0.247 |
| 8 | 0.298 | 0.298 | 0.298 | 0.298 | 0.270 | 0.270 | 0.270 | 0.274 | 0.270 | 0.270 | 0.270 | 0.276 | 0.268 | 0.270 | 0.237 | 0.243 | 0.235 | 0.235 | 0.237 | 0.237 | 0.237 | 0.237 | 0.272 |
| 9 | 0.306 | 0.306 | 0.306 | 0.306 | 0.312 | 0.312 | 0.312 | 0.351 | 0.312 | 0.312 | 0.312 | 0.296 | 0.315 | 0.312 | 0.256 | 0.258 | 0.252 | 0.257 | 0.256 | 0.256 | 0.256 | 0.256 | 0.280 |
| 10 | 0.313 | 0.313 | 0.313 | 0.313 | 0.312 | 0.312 | 0.312 | 0.312 | 0.312 | 0.312 | 0.312 | 0.364 | 0.295 | 0.312 | 0.287 | 0.284 | 0.293 | 0.288 | 0.287 | 0.287 | 0.287 | 0.287 | 0.299 |
| 11 | ${ }_{0} 0.321$ | ${ }_{0}^{0.321}$ | 0.321 | 0.321 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | ${ }_{0}^{0.000}$ | 0.335 | ${ }_{0}^{0.330}$ | 0.338 | 0.348 | ${ }_{0}^{0.327}$ | 0.335 | ${ }_{0}^{0.335}$ | 0.335 | 0.332 |
| 12 | ${ }_{0} 0.328$ | 0.328 | 0.328 | 0.328 | 0.373 | 0.373 | ${ }^{0.373}$ | 0.373 | 0.373 | 0.373 | 0.373 | 0.418 | 0.342 | ${ }^{0.373}$ | 0.355 | 0.354 | 0.351 | 0.362 | 0.352 | 0.355 | 0.355 | 0.355 | 0.350 |
| 13 | ${ }_{0} 0.333$ | 0.333 | 0.333 | 0.333 | 0.368 | 0.368 | 0.368 | 0.368 | 0.368 | 0.368 | 0.368 | 0.368 | 0.368 | 0.368 | 0.388 | 0.386 | 0.382 | 0.392 | 0.380 | 0.388 | ${ }^{0.388}$ | 0.388 | 0.366 |
| 14 | 0.338 | 0.338 | 0.338 | 0.338 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.413 | 0.410 | 0.387 | 0.413 | 0.413 | 0.413 | 0.413 | 0.413 | 0.379 |
| 15+ | 0.359 | 0.359 | 0.359 | 0.359 | 0.328 | 0.328 | 0.328 | 0.328 | 0.328 | 0.328 | 0.328 | 0.328 | 0.328 | 0.328 | 0.485 | 0.486 | 0.423 | 0.481 | 0.493 | 0.485 | 0.485 | 0.485 | 0.398 |


| Q4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.3.a | 27.4.a | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 0.169 | 0.169 | 0.169 | 0.169 | 0.172 | 0.164 | 0.172 | 0.172 | 0.173 | 0.172 | 0.172 | 0.169 | 0.172 | 0.173 | 0.145 | 0.147 | 0.141 | 0.144 | 0.145 | 0.169 |
| 5 | 0.214 | 0.214 | 0.308 | 0.208 | 0.222 | 0.211 | 0.222 | 0.222 | 0.206 | 0.222 | 0.222 | 0.220 | 0.222 | 0.224 | 0.175 | 0.184 | 0.169 | 0.176 | 0.171 | 0.212 |
| 6 | 0.248 | 0.248 | 0.321 | 0.221 | 0.246 | 0.241 | 0.246 | 0.246 | 0.241 | 0.246 | 0.246 | 0.245 | 0.246 | 0.246 | 0.196 | 0.203 | 0.177 | 0.198 | 0.192 | 0.239 |
| 7 | 0.252 | 0.252 | 0.332 | 0.237 | 0.250 | 0.234 | 0.250 | 0.250 | 0.250 | 0.250 | 0.250 | 0.251 | 0.250 | 0.249 | 0.232 | 0.247 | 0.193 | 0.228 | 0.227 | 0.249 |
| 8 | 0.312 | 0.312 | 0.342 | 0.258 | 0.281 | 0.274 | 0.281 | 0.281 | 0.281 | 0.281 | 0.281 | 0.281 | 0.281 | 0.281 | 0.251 | 0.271 | 0.270 | 0.244 | 0.248 | 0.279 |
| 9 | 0.319 | 0.319 | 0.351 | 0.277 | 0.291 | 0.303 | 0.291 | 0.291 | 0.291 | 0.291 | 0.291 | 0.291 | 0.291 | 0.291 | 0.264 | 0.268 | 0.278 | 0.261 | 0.265 | 0.283 |
| 10 | 0.358 | 0.358 | 0.359 | 0.339 | 0.289 | 0.289 | 0.289 | 0.289 | 0.289 | 0.289 | 0.289 | 0.289 | 0.289 | 0.289 | 0.293 | 0.292 | 0.281 | 0.291 | 0.295 | 0.302 |
| 11 | 0.362 | 0.362 | 0.367 | 0.304 | 0.385 | 0.385 | 0.385 | 0.385 | 0.385 | 0.385 | 0.385 | 0.385 | 0.385 | 0.385 | 0.331 | 0.330 | 0.345 | 0.333 | 0.329 | 0.336 |
| 12 | 0.367 | 0.367 | 0.375 | 0.295 | 0.454 | 0.454 | 0.454 | 0.454 | 0.454 | 0.454 | 0.454 | 0.454 | 0.454 | 0.454 | 0.354 | 0.355 | 0.359 | 0.353 | 0.354 | 0.374 |
| 13 | 0.367 | 0.367 | 0.382 | 0.325 | 0.275 | 0.286 | 0.275 | 0.275 | 0.275 | 0.275 | 0.275 | 0.275 | 0.275 | 0.275 | 0.388 | 0.392 | 0.372 | 0.388 | 0.387 | 0.362 |
| 14 | 0.387 | 0.387 | 0.388 | 0.338 | 0.325 | 0.325 | 0.325 | 0.325 | 0.325 | 0.325 | 0.325 | 0.325 | 0.325 | 0.325 | 0.415 | 0.419 | 0.406 | 0.413 | 0.414 | 0.406 |
| 15+ | 0.408 | 0.408 | 0.410 | 0.372 | 0.507 | 0.507 | 0.507 | 0.507 | 0.507 | 0.507 | 0.507 | 0.507 | 0.507 | 0.507 | 0.476 | 0.483 | 0.479 | 0.481 | 0.470 | 0.450 |

Table 7.2.5.1 cont. Western horse mackerel stock. Mean weight ( kg ) in catch-at-age by quarter and area in 2021 ( $15=15+$ group)

| all Q |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\frac{\text { Ages }}{0}$ | $\frac{\text { 27.2.a }}{0}$ | $\frac{\text { 27.3.a }}{0}$ | $\frac{\text { 27.4.a }}{0}$ | $\frac{27.6 . a}{0}$ | $\frac{27.7 . a}{0}$ | 27.7.6 | $\frac{27.7 . \mathrm{c}}{0}$ | $\frac{27.7 . c .2}{0}$ | $\frac{27.7 . e}{0}$ | $\frac{27.7 . \mathrm{f}}{0}$ | $\frac{27.7 .8}{0}$ | $\frac{27.7 . \mathrm{h}}{0}$ | $\frac{27.7 . j}{0}$ | $\frac{27.7 .7 .2}{0}$ | $\frac{27.7 .1 .1}{0}$ | $\frac{27.7 .1 .2}{0}$ | $\frac{27.8 . a}{0.027}$ | $\frac{27.8 .6}{0.019}$ | $\frac{27.8 . \mathrm{c}}{0.049}$ | $\frac{27.7 . c . e}{0.022}$ | $\frac{27.8 . \mathrm{c} . \mathrm{w}}{0.047}$ | $\frac{27.8 . \mathrm{d}}{0.022}$ | $\frac{27.7 . \text { d. } 2}{0.021}$ | 27.8.e | ${ }_{\text {Total }} 0.027$ |
| 1 | 0 | 0 | 0 | 0 | 0.058 | 0.068 | 0.069 | 0.069 | 0.067 | 0.068 | 0.067 | 0.068 | 0.069 | 0.069 | 0.000 | 0.044 | 0.048 | 0.029 | 0.040 | 0.041 | 0.045 | 0.031 | 0.042 | 0.042 | 0.042 |
| 2 | 0 | 0 | 0 | 0 | 0.135 | 0.088 | 0.092 | 0.074 | 0.068 | ${ }_{0} 0.055$ | 0.055 | 0.052 | 0.127 | 0.096 | 0.053 | 0.146 | 0.075 | 0.056 | 0.057 | 0.078 | 0.071 | 0.061 | 0.079 | 0.079 | 0.072 |
| 3 | 0.146 | 0.153 | 0.140 | 0.136 | 0.172 | 0.092 | 0.106 | 0.079 | 0.080 | 0.079 | 0.078 | 0.083 | 0.174 | ${ }_{0} 0.138$ | ${ }^{0.078}$ | 0.173 | 0.113 | 0.091 | 0.069 | 0.112 | 0.101 | 0.096 | 0.120 | 0.120 | 0.090 |
| 4 | 0.162 | 0.169 | 0.157 | 0.152 | 0.191 | 0.122 | 0.188 | ${ }_{0}^{0.107}$ | 0.109 | 0.104 | 0.108 | 0.110 | 0.184 | 0.166 | 0.106 | 0.192 | 0.143 | ${ }_{0}^{0.138}$ | 0.108 | 0.136 | 0.127 | 0.129 | ${ }_{0}^{0.148}$ | 0.148 | 0.122 |
| 5 | 0.205 | 0.227 | 0.269 | 0.191 | 0.231 | 0.169 | 0.214 | 0.148 | 0.141 | 0.159 | 0.142 | 0.150 | 0.230 | 0.217 | 0.143 | 0.231 | 0.170 | 0.166 | 0.137 | 0.153 | 0.161 | 0.154 | 0.176 | 0.176 | 0.169 |
| 6 | 0.224 | 0.252 | 0.302 | 0.201 | 0.244 | 0.204 | 0.253 | 0.195 | 0.172 | 0.192 | 0.181 | 0.208 | 0.245 | 0.244 | 0.192 | 0.244 | 0.191 | 0.188 | 0.144 | 0.173 | 0.184 | 0.175 | 0.193 | 0.193 | 0.205 |
| 7 | 0.229 | 0.259 | 0.295 | 0.217 | 0.255 | 0.200 | 0.270 | 0.205 | 0.201 | 0.208 | 0.201 | 0.227 | 0.250 | 0.235 | 0.202 | 0.255 | 0.224 | 0.220 | 0.187 | 0.188 | 0.210 | 0.195 | 0.211 | 0.211 | 0.219 |
| 8 | 0.286 | 0.300 | 0.308 | 0.257 | 0.271 | 0.257 | 0.351 | 0.262 | 0.264 | 0.248 | 0.259 | 0.255 | 0.277 | 0.275 | 0.256 | 0.270 | 0.252 | 0.270 | 0.313 | 0.239 | 0.249 | 0.253 | 0.237 | 0.237 | 0.262 |
| 9 | 0.297 | 0.310 | 0.322 | 0.284 | 0.312 | 0.275 | 0.370 | 0.277 | 0.305 | 0.307 | 0.293 | 0.274 | 0.295 | 0.273 | 0.264 | 0.312 | 0.268 | 0.298 | 0.329 | 0.284 | 0.268 | 0.295 | 0.256 | 0.256 | 0.282 |
| 10 | 0.334 | 0.321 | 0.329 | 0.325 | 0.312 | 0.317 | 0.293 | 0.301 | 0.289 | 0.305 | 0.307 | 0.348 | 0.355 | 0.270 | 0.305 | 0.312 | 0.293 | 0.304 | 0.335 | 0.294 | 0.284 | 0.297 | 0.287 | 0.287 | 0.301 |
| 11 | 0.348 | 0.335 | 0.347 | 0.342 | 0.385 | 0.304 | 0.385 | 0.314 | 0.343 | 0.295 | 0.326 | 0.301 | 0.352 | 0.301 | 0.295 | 0.000 | 0.330 | 0.344 | 0.363 | 0.336 | 0.316 | 0.331 | 0.335 | 0.335 | 0.328 |
| 12 | 0.352 | 0.345 | 0.358 | 0.333 | 0.376 | 0.314 | 0.394 | 0.299 | 0.341 | 0.275 | 0.361 | 0.307 | 0.423 | 0.328 | ${ }^{0.272}$ | 0.373 | 0.351 | ${ }_{0} 0.347$ | 0.350 | 0.345 | 0.340 | 0.338 | 0.355 | 0.355 | 0.340 |
| 13 | 0.341 | 0.346 | 0.358 | 0.323 | 0.362 | 0.331 | 0.458 | 0.311 | 0.227 | 0.384 | 0.314 | 0.295 | 0.282 | 0.278 | 0.293 | 0.368 | 0.385 | 0.399 | 0.410 | 0.397 | 0.376 | 0.390 | 0.388 | 0.388 | 0.334 |
| 14 | 0.362 | 0.349 | 0.359 | 0.341 | 0.325 | 0.265 | 0.325 | 0.265 | 0.276 | 0.266 | 0.295 | 0.272 | 0.320 | 0.320 | 0.265 | 0.000 | 0.414 | 0.409 | 0.392 | 0.408 | 0.406 | 0.398 | 0.413 | 0.413 | 0.383 |
| 15+ | 0.377 | 0.368 | 0.378 | 0.348 | 0.335 | 0.395 | 0.507 | 0.358 | 0.316 | 0.478 | 0.406 | 0.504 | 0.490 | 0.399 | 0.393 | 0.328 | 0.473 | 0.477 | 0.466 | 0.469 | 0.477 | 0.463 | 0.485 | 0.485 | 0.408 |

Table 7.2.5.2. Western horse mackerel stock. Mean length (cm) in catch-at-age by quarter and area in 2021 ( $15=15+$ group)

| Q1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.6.a | 27.7.b | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.7.k. 1 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | , | 0 | 0 | 0 | 0 | 0 | 0 | 0 | NA |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14.3 | 13.1 | 14.6 | 15.0 | 14.6 | 14.3 | 14.3 |
| 2 | 0 | 0 | 19.1 | 19.1 | 19.5 | 19.3 | 18.2 | 18.8 | 19.1 | 19.1 | 19.1 | 17.4 | 17.0 | 18.2 | 17.8 | 16.7 | 17.4 | 18.7 |
| 3 | 23.9 | 23.9 | 22.0 | 21.8 | 22.0 | 22.0 | 21.5 | 21.8 | 21.8 | 21.9 | 21.8 | 20.9 | 19.4 | 19.4 | 20.0 | 22.7 | 20.9 | 21.8 |
| 4 | 25.0 | 25.0 | 25.0 | 24.4 | 24.6 | 24.2 | 24.0 | 24.3 | 24.4 | 24.6 | 24.4 | 24.7 | 24.7 | 0 | 25.1 | 24.7 | 24.7 | 24.4 |
| 5 | 28.5 | 28.5 | 27.7 | 27.0 | 26.6 | 27.6 | 26.5 | 26.4 | 26.9 | 27.9 | 26.9 | 26.2 | 26.2 | 0 | 26.6 | 26.1 | 26.2 | 27.1 |
| 6 | 29.3 | 29.3 | 29.8 | 29.4 | 28.0 | 29.5 | 28.1 | 29.0 | 29.4 | 30.9 | 29.4 | 27.5 | 27.5 | 0 | 27.8 | 27.4 | 27.5 | 29.2 |
| 7 | 30.3 | 30.3 | 29.6 | 30.0 | 29.4 | 30.0 | 29.2 | 29.7 | 29.9 | 30.4 | 29.9 | 28.9 | 29.1 | 0 | 29.3 | 28.6 | 28.9 | 30.0 |
| 8 | 32.1 | 32.1 | 32.4 | 32.4 | 31.8 | 32.4 | 31.0 | 32.1 | 32.4 | 32.6 | 32.4 | 31.7 | 31.7 | 0 | 32.6 | 30.6 | 31.7 | 32.1 |
| 9 | 33.3 | 33.3 | 33.0 | 32.9 | 32.9 | 34.0 | 33.1 | 32.9 | 32.9 | 32.5 | 32.9 | 33.5 | 33.1 | 0 | 34.3 | 31.7 | 33.5 | 33.1 |
| 10 | 34.8 | 34.8 | 34.9 | 33.6 | 32.8 | 33.9 | 34.4 | 34.2 | 33.9 | 32.1 | 33.9 | 33.4 | 32.9 | 0 | 34.1 | 32.1 | 33.4 | 33.9 |
| 11 | 35.4 | 35.4 | 34.1 | 34.0 | 34.0 | 34.0 | 34.4 | 34.0 | 34.0 | 33.4 | 34.0 | 34.5 | 34.1 | 0 | 35.2 | 33.2 | 34.5 | 34.6 |
| 12 | 35.2 | 35.2 | 34.2 | 34.4 | 33.2 | 33.2 | 32.6 | 33.2 | 33.2 | 32.1 | 33.2 | 34.7 | 34.6 | 0 | 35.0 | 34.2 | 34.7 | 34.3 |
| 13 | 34.7 | 34.7 | 35.1 | 34.5 | 31.5 | 35.5 | 35.1 | 33.9 | 33.8 | 32.8 | 33.8 | 36.5 | 36.0 | 0 | 37.0 | 34.0 | 36.5 | 34.3 |
| 14 | 35.4 | 35.4 | 32.6 | 32.6 | 32.6 | 32.6 | 32.6 | 32.6 | 32.6 | 32.6 | 32.6 | 36.4 | 36.4 | 0 | 36.5 | 36.2 | 36.4 | 35.3 |
| 15+ | 35.5 | 35.5 | 37.2 | 35.8 | 33.4 | 38.9 | 38.4 | 38.6 | 36.5 | 35.6 | 36.5 | 38.3 | 38.3 | 0 | 38.0 | 39.2 | 38.3 | 36.2 |


| $\begin{gathered} \text { Q2 } \\ \text { Ages } \end{gathered}$ | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | NA |
| 1 | 0 | 19 | 19 | 19 | 19 | 19 | 19 | 19 | 19 | 19 | 19 | 14 | 14 | 16 | 14 | 16 | 14 | 14 |
| 2 | 0 | 21 | 21 | 21 | 21 | 21 | 21 | 21 | 21 | 21 | 21 | 18 | 18 | 18 | 17 | 20 | 18 | 20 |
| 3 | 23.9 | 21.0 | 21.0 | 21.0 | 21.0 | 21.0 | 21.0 | 21.0 | 21.0 | 21.0 | 21.0 | 21.9 | 20.5 | 20.1 | 23.7 | 22.2 | 22.1 | 22.0 |
| 4 | 25.0 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 | 24.1 | 24.1 | 23.1 | 24.6 | 23.8 | 24.1 | 24.2 |
| 5 | 28.7 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 25.8 | 25.8 | 24.2 | 25.9 | 24.9 | 25.7 | 25.8 |
| 6 | 29.4 | 32.2 | 32.2 | 32.2 | 32.2 | 32.2 | 32.2 | 32.2 | 32.2 | 32.2 | 32.2 | 27.0 | 27.1 | 24.4 | 27.1 | 25.6 | 27.0 | 27.1 |
| 7 | 30.3 | 33.3 | 33.3 | 33.3 | 33.3 | 33.3 | 33.3 | 33.3 | 33.3 | 33.3 | 33.3 | 28.9 | 28.6 | 27.4 | 27.8 | 29.9 | 28.0 | 28.3 |
| 8 | 32.2 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 31.2 | 32.3 | 35.9 | 29.9 | 32.3 | 30.5 | 30.7 |
| 9 | 33.2 | 36.3 | 36.3 | 36.3 | 36.3 | 36.3 | 36.3 | 36.3 | 36.3 | 36.3 | 36.3 | 32.6 | 34.0 | 36.6 | 32.0 | 33.0 | 32.5 | 32.8 |
| 10 | 33.9 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 32.7 | 33.8 | 37.1 | 32.4 | 32.5 | 32.6 | 32.6 |
| 11 | 34.8 | 37.5 | 37.5 | 37.5 | 37.5 | 37.5 | 37.5 | 37.5 | 37.5 | 37.5 | 37.5 | 33.8 | 35.2 | 36.2 | 34.0 | 33.4 | 33.9 | 34.1 |
| 12 | 34.8 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 34.9 | 34.9 | 34.9 | 34.6 | 34.3 | 34.5 | 34.6 |
| 13 | 34.5 | 38.5 | 38.5 | 38.5 | 38.5 | 38.5 | 38.5 | 38.5 | 38.5 | 38.5 | 38.5 | 34.6 | 37.3 | 39.5 | 37.1 | 35.2 | 36.4 | 36.7 |
| 14 | 34.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 36.3 | 36.5 | 36.6 | 36.5 | 36.2 | 36.3 | 36.3 |
| 15+ | 35.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 37.8 | 39.0 | 38.9 | 38.8 | 39.1 | 38.9 | 38.9 |

Table 7.2.5.2 cont. Western horse mackerel stock. Mean length ( cm ) in catch-at-age by quarter and area in 2021 ( $15=15+$ group)

| Q3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.3.a | 27.4.a | 27.6.a | 27.7.a | 27.7.6 | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j ${ }^{\text {d }}$ | 27.7. ${ }^{\text {. } 2}$ | $\frac{27.7 . \mathrm{k} .2}{0}$ | ${ }_{\text {27.8.a }}^{13.16}$ | ${ }^{27.8 .6} 13.07$ | ${ }_{\text {27.8.c }}^{13.16}$ | $\frac{\text { 27.9.c.e }}{13.21}$ | ${ }_{\text {27.8.c.w }}^{13.16}$ | ${ }_{\text {27.8.d }}^{13.16}$ | $\frac{27.8 . \text { d. } 26}{13.16}$ | ${ }^{27.8 . e} 13.16$ | ${ }_{\text {Total }}^{13.1614922}$ |
| 1 | 0 | 0 | 0 | 0 | 17 | 17 | 17 | 17 | 17 | 17 | 17 | 17 | 17 | 17 | 17 | 16 | 18 | 17 | 17 | 17 | 17 | 17 | 13.1614922 17 |
| 2 | 0 | 0 | 0 | 0 | 25 | 25 | 25 | 24 | 25 | 25 | 25 | 24 | 27 | 25 | 21 | 20 | 19 | 21 | 24 | 21 | 21 | 21 | 21 |
| 3 | 0 | 0 | 0 | 0 | 26.9 | 26.9 | 26.9 | 26.7 | 26.9 | 26.9 | 26.9 | 27.0 | 26.7 | 26.9 | 24.1 | 24.1 | 24.1 | 23.8 | 24.8 | 24.1 | 24.1 | 24.1 | 25.4 |
| 4 | 30.5 | 30.5 | 30.5 | 30.5 | 28.2 | 28.2 | 28.2 | 28.0 | 28.2 | 28.2 | 28.2 | 28.0 | 28.3 | 28.2 | 26.0 | 26.0 | 26.0 | 26.0 | 26.0 | 26.0 | 26.0 | 26.0 | 27.3 |
| 5 | 31.1 | 31.1 | 31.1 | 31.1 | 30.0 | 30.0 | 30.0 | 30.0 | 30.0 | 30.0 | 30.0 | 30.1 | 30.0 | 30.0 | 27.6 | 27.9 | 28.8 | 27.5 | 27.7 | 27.6 | 27.6 | 27.6 | 28.6 |
| 6 | 31.4 | 31.4 | 31.4 | 31.4 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.3 | 30.7 | 30.5 | 28.6 | 28.8 | 29.9 | 28.5 | 28.6 | 28.6 | 28.6 | 28.6 | 29.5 |
| 7 | 32.0 | 32.0 | 32.0 | 32.0 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 | 31.0 | 31.1 | 31.1 | 29.5 | 30.0 | 30.8 | 29.3 | 29.5 | 29.5 | 29.5 | 29.5 | 30.8 |
| 8 | 32.4 | 32.4 | 32.4 | 32.4 | 32.2 | 32.2 | 32.2 | 32.2 | 32.2 | 32.2 | 32.2 | 32.1 | 32.2 | 32.2 | 30.7 | 31.0 | 30.6 | 30.6 | 30.7 | 30.7 | 30.7 | 30.7 | 31.9 |
| 9 | 32.8 | 32.8 | 32.8 | 32.8 | 33.4 | 33.4 | 33.4 | 34.1 | 33.4 | 33.4 | 33.4 | 31.5 | 33.8 | 33.4 | 31.6 | 31.6 | 31.4 | 31.6 | 31.5 | 31.6 | 31.6 | 31.6 | 32.2 |
| 10 | 33.1 | 33.1 | 33.1 | 33.1 | 33.0 | 33.0 | 33.0 | 33.0 | 33.0 | 33.0 | 33.0 | 34.5 | 32.5 | 33.0 | 32.8 | 32.7 | 33.0 | 32.9 | 32.8 | 32.8 | 32.8 | 32.8 | 32.9 |
| 11 | 33.5 | 33.5 | 33.5 | 33.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 34.6 | 34.5 | 34.8 | 35.1 | 34.4 | 34.6 | 34.6 | 34.6 | 34.4 |
| 12 | 33.8 | 33.8 | 33.8 | 33.8 | 35.7 | 35.7 | 35.7 | 35.7 | 35.7 | 35.7 | 35.7 | 36.5 | 35.1 | 35.7 | 35.4 | 35.3 | 35.2 | 35.6 | 35.3 | 35.4 | 35.4 | 35.4 | 34.9 |
| 13 | 34.0 | 34.0 | 34.0 | 34.0 | 35.8 | 35.8 | 35.8 | 35.8 | 35.8 | 35.8 | 35.8 | 35.8 | 35.8 | 35.8 | 36.5 | 36.4 | 36.3 | 36.6 | 36.2 | 36.5 | 36.5 | 36.5 | 35.5 |
| 14 | 34.2 | 34.2 | 34.2 | 34.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 37.3 | 37.2 | 36.5 | 37.3 | 37.3 | 37.3 | 37.3 | 37.3 | 35.9 |
| 15+ | 34.8 | 34.8 | 34.8 | 34.8 | 33.8 | 33.8 | 33.8 | 33.8 | 33.8 | 33.8 | 33.8 | 33.8 | 33.8 | 33.8 | 39.4 | 39.5 | 37.6 | 39.3 | 39.6 | 39.4 | 39.4 | 39.4 | 36.2 |


| Q4 Ages | 27.2.a | 27.3.a | 27.4.a | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 13.95 | 12.6 | 17.59 | 13.69 | 17.33 | 14.0492178 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 17 | 16 | 18 | 17 | 19 | 18 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | , | 0 | 0 | 20 | 21 | 19 | 21 | 20 | 20 |
| 3 | 26.8 | 26.8 | 26.8 | 26.8 | 26.9 | 25.7 | 26.9 | 26.9 | 26.4 | 26.9 | 26.9 | 27.0 | 26.9 | 26.8 | 23.8 | 24.4 | 24.0 | 23.4 | 24.0 | 25.8 |
| 4 | 27.7 | 27.7 | 27.7 | 27.7 | 27.5 | 27.4 | 27.5 | 27.5 | 27.0 | 27.5 | 27.5 | 27.2 | 27.5 | 27.6 | 25.8 | 25.9 | 25.6 | 25.8 | 25.8 | 27.4 |
| 5 | 29.8 | 29.8 | 32.3 | 29.6 | 30.0 | 29.8 | 30.0 | 30.0 | 28.3 | 30.0 | 30.0 | 30.1 | 30.0 | 30.1 | 27.6 | 28.1 | 27.2 | 27.6 | 27.4 | 29.6 |
| 6 | 30.9 | 30.9 | 32.8 | 30.2 | 31.0 | 31.1 | 31.0 | 31.0 | 30.3 | 31.0 | 31.0 | 31.0 | 31.0 | 31.1 | 28.7 | 29.0 | 27.7 | 28.8 | 28.5 | 30.7 |
| 7 | 31.3 | 31.3 | 33.3 | 30.9 | 31.3 | 30.8 | 31.3 | 31.3 | 31.2 | 31.3 | 31.3 | 31.5 | 31.3 | 31.3 | 30.4 | 31.0 | 28.5 | 30.2 | 30.2 | 31.3 |
| 8 | 33.0 | 33.0 | 33.7 | 31.8 | 32.4 | 32.5 | 32.4 | 32.4 | 32.4 | 32.4 | 32.4 | 32.4 | 32.4 | 32.4 | 31.3 | 32.0 | 32.1 | 31.0 | 31.2 | 32.2 |
| 9 | 33.4 | 33.4 | 34.1 | 32.6 | 32.8 | 33.5 | 32.8 | 32.8 | 32.8 | 32.8 | 32.8 | 32.8 | 32.8 | 32.8 | 31.9 | 32.0 | 32.5 | 31.8 | 31.9 | 32.4 |
| 10 | 34.4 | 34.4 | 34.4 | 34.8 | 32.8 | 32.8 | 32.8 | 32.8 | 32.8 | 32.8 | 32.8 | 32.8 | 32.8 | 32.8 | 33.1 | 33.0 | 32.6 | 33.0 | 33.2 | 33.3 |
| 11 | 34.6 | 34.6 | 34.7 | 33.6 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 36.5 | 34.5 | 34.5 | 35.0 | 34.6 | 34.4 | 34.5 |
| 12 | 34.8 | 34.8 | 35.0 | 33.1 | 37.8 | 37.8 | 37.8 | 37.8 | 37.8 | 37.8 | 37.8 | 37.8 | 37.8 | 37.8 | 35.3 | 35.4 | 35.5 | 35.3 | 35.3 | 35.5 |
| 13 | 35.0 | 35.0 | 35.3 | 34.3 | 32.7 | 33.1 | 32.7 | 32.7 | 32.7 | 32.7 | 32.7 | 32.7 | 32.7 | 32.7 | 36.5 | 36.6 | 36.0 | 36.5 | 36.5 | 35.4 |
| 14 | 35.5 | 35.5 | 35.5 | 34.7 | 34.5 | 34.5 | 34.5 | 34.5 | 34.5 | 34.5 | 34.5 | 34.5 | 34.5 | 34.5 | 37.4 | 37.5 | 37.1 | 37.3 | 37.3 | 36.8 |
| 15+ | 36.2 | 36.2 | 36.2 | 36.1 | 39.7 | 39.7 | 39.7 | 39.7 | 39.7 | 39.7 | 39.7 | 39.7 | 39.7 | 39.7 | 39.2 | 39.4 | 39.3 | 39.3 | 39.0 | 37.9 |

## Table 7.2.5.2 cont. Western horse mackerel stock. Mean length ( cm ) in catch-at-age by quarter and area in 2021 ( $15=15+$ group)

| $\begin{gathered} \text { all Q } \\ \text { Ages } \end{gathered}$ | 27.2.a | 27.3.a | 27.4.a | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.8 | 27.7.h | 27.7.j | 27.7.j. 2 | 27.7.k. 1 | 27.7.k.2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c. | 27.8.c.w | 27.8.d | 27.8.d.2 |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | , | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0 | 14.0 | 12.7 | 17.6 | 13.2 | 17.3 | 13.2 | 13.2 | 13.2 | 14.0 |
| 1 | 0 | 0 | 0 | 0 | 18.0 | 19.0 | 19.0 | 19.0 | 18.8 | 19.0 | 18.8 | 19.0 | 19.0 | 19.0 | 0 | 16.5 | 17.3 | 14.5 | 16.3 | 16.5 | 16.8 | 14.9 | 16.6 | 16.6 | 16.5 |
| 2 | 0 | 0 | 0 | 0 | 24.0 | 20.5 | 20.5 | 19.9 | 19.9 | 19.3 | 18.6 | 18.9 | 23.8 | 20.9 | 19.1 | 24.8 | 20.4 | 18.4 | 18.6 | 20.7 | 20.0 | 18.9 | 20.8 | 20.8 | 20.1 |
| 3 | 26.3 | 26.8 | 26.0 | 25.8 | 26.8 | 22.4 | 21.3 | 21.8 | 22.0 | 22.0 | 21.6 | 22.1 | 26.8 | 25.4 | 21.8 | 26.9 | 23.6 | 21.7 | 19.9 | 23.6 | 22.7 | 22.3 | 24.1 | 24.1 | 22.5 |
| 4 | 27.3 | 27.7 | 27.1 | 26.8 | 28.1 | 25.3 | 29.1 | 24.4 | 24.6 | 24.3 | 24.1 | 24.5 | 27.8 | 27.3 | 24.4 | 28.2 | 25.7 | 25.4 | 23.3 | 25.2 | 24.6 | 24.7 | 26.0 | 26.0 | 25.1 |
| 5 | 29.4 | 30.1 | 31.1 | 28.9 | 30.0 | 28.1 | 30.4 | 27.1 | 26.8 | 27.6 | 26.7 | 27.1 | 30.0 | 29.8 | 26.9 | 30.0 | 27.3 | 27.0 | 25.3 | 26.3 | 26.7 | 26.3 | 27.6 | 27.6 | 27.6 |
| 6 | 30.2 | 31.0 | 32.3 | 29.5 | 30.5 | 29.9 | 31.8 | 29.5 | 28.5 | 29.5 | 28.7 | 29.9 | 30.4 | 30.9 | 29.4 | 30.5 | 28.4 | 28.3 | 25.6 | 27.5 | 28.0 | 27.6 | 28.6 | 28.6 | 29.3 |
| 7 | 30.7 | 31.4 | 32.3 | 30.4 | 31.1 | 29.8 | 32.7 | 30.0 | 29.9 | 30.1 | 29.7 | 30.7 | 31.1 | 30.9 | 29.9 | 31.1 | 30.0 | 29.8 | 28.2 | 28.2 | 29.4 | 28.6 | 29.5 | 29.5 | 30.3 |
| 8 | 32.5 | 32.5 | 32.7 | 32.1 | 32.2 | 32.5 | 36.2 | 32.7 | 32.6 | 32.4 | 32.0 | 32.2 | 32.2 | 32.5 | 32.4 | 32.2 | 31.3 | 32.0 | 33.7 | 30.6 | 31.1 | 31.2 | 30.7 | 30.7 | 31.8 |
| 9 | 33.2 | 33.0 | 33.3 | 33.2 | 33.5 | 33.1 | 36.2 | 33.3 | 34.0 | 34.0 | 33.3 | 32.9 | 32.0 | 32.9 | 32.9 | 33.4 | 32.0 | 33.1 | 34.3 | 32.5 | 32.0 | 32.9 | 31.6 | 31.6 | 32.8 |
| 10 | 34.1 | 33.3 | 33.6 | 34.7 | 33.0 | 34.8 | 33.5 | 33.6 | 32.8 | 33.9 | 33.4 | 34.1 | 34.3 | 32.4 | 33.9 | 33.0 | 33.1 | 33.4 | 34.5 | 33.0 | 32.7 | 33.1 | 32.8 | 32.8 | 33.3 |
| 11 | 34.7 | 33.9 | 34.2 | 35.3 | 37.4 | 34.3 | 37.5 | 34.7 | 35.9 | 34.0 | 35.2 | 34.2 | 35.9 | 34.2 | 34.0 | 0.0 | 34.5 | 34.9 | 35.6 | 34.6 | 33.9 | 34.4 | 34.6 | 34.6 | 34.4 |
| 12 | 34.8 | 34.2 | 34.6 | 35.1 | 35.7 | 34.5 | 36.6 | 34.5 | 35.0 | 33.3 | 35.4 | 34.1 | 36.7 | 34.4 | 33.2 | 35.7 | 35.3 | 35.1 | 35.2 | 35.0 | 34.9 | 34.8 | 35.4 | 35.4 | 34.8 |
| 13 | 34.7 | 34.4 | 34.7 | 34.6 | 35.6 | 35.2 | 38.3 | 34.6 | 31.6 | 35.5 | 34.4 | 33.8 | 33.0 | 33.0 | 33.8 | 35.8 | 36.4 | 36.8 | 37.2 | 36.8 | 36.1 | 36.5 | 36.5 | 36.5 | 34.8 |
| 14 <br> $15+$ | 35.1 35.6 | 34.5 35.0 | 34.8 35.3 | 35.3 35.5 | 34.5 34.0 | 32.6 37.0 | 34.5 39.7 | 32.6 35.8 | 32.9 33.6 | 32.6 38.9 | ${ }_{3}^{33.5}$ | 32.8 38.6 | 34.3 39.3 | 34.3 36.9 | 32.6 36.5 | 0.0 33.8 | 37.3 39.1 | 37.2 39.2 | 36.7 38.9 | 37.1 38.9 | 37.1 39.2 | 36.8 38.8 |  | 37.3 39.4 | 36.2 36.9 |

Table 7.2.5.3. Western horse mackerel. Catch weights-at-age ( $\mathbf{k g}$ ), from Q1 and Q2 data (note that 2021 data is from Q1 and Q3).

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0.024 | 0.052 | 0.066 | 0.080 | 0.207 | 0.232 | 0.269 | 0.280 | 0.292 | 0.305 | 0.369 | 0.348 | 0.348 | 0.348 | 0.356 | 0.366 |
| 1983 | 0.024 | 0.052 | 0.066 | 0.080 | 0.171 | 0.227 | 0.257 | 0.276 | 0.270 | 0.243 | 0.390 | 0.348 | 0.348 | 0.348 | 0.356 | 0.366 |
| 1984 | 0.024 | 0.052 | 0.064 | 0.077 | 0.122 | 0.155 | 0.201 | 0.223 | 0.253 | 0.246 | 0.338 | 0.348 | 0.348 | 0.348 | 0.356 | 0.366 |
| 1985 | 0.024 | 0.052 | 0.066 | 0.081 | 0.148 | 0.140 | 0.193 | 0.236 | 0.242 | 0.289 | 0.247 | 0.241 | 0.251 | 0.314 | 0.346 | 0.321 |
| 1986 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.134 | 0.169 | 0.195 | 0.242 | 0.292 | 0.262 | 0.319 | 0.287 | 0.345 | 0.260 | 0.360 |
| 1987 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.126 | 0.150 | 0.171 | 0.218 | 0.254 | 0.281 | 0.336 | 0.244 | 0.328 | 0.245 | 0.373 |
| 1988 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.126 | 0.141 | 0.143 | 0.217 | 0.274 | 0.305 | 0.434 | 0.404 | 0.331 | 0.392 | 0.424 |
| 1989 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.103 | 0.131 | 0.159 | 0.127 | 0.210 | 0.252 | 0.381 | 0.400 | 0.421 | 0.448 | 0.516 |
| 1990 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.127 | 0.135 | 0.124 | 0.154 | 0.174 | 0.282 | 0.328 | 0.355 | 0.399 | 0.388 | 0.379 |
| 1991 | 0.024 | 0.052 | 0.066 | 0.080 | 0.121 | 0.137 | 0.143 | 0.144 | 0.150 | 0.182 | 0.189 | 0.303 | 0.323 | 0.354 | 0.365 | 0.330 |
| 1992 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.133 | 0.151 | 0.150 | 0.158 | 0.160 | 0.182 | 0.288 | 0.306 | 0.359 | 0.393 | 0.401 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.153 | 0.1661 | 0.173 | 0.172 | 0.170 | 0.206 | 0.238 | 0.308 | 0.327 | 0.376 | 0.421 |
| 1994 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.147 | 0.185 | 0.169 | 0.191 | 0.191 | 0.190 | 0.275 | 0.240 | 0.326 | 0.342 | 0.383 |
| 1995 | 0.024 | 0.052 | 0.059 | 0.066 | 0.119 | 0.096 | 0.152 | 0.166 | 0.178 | 0.187 | 0.197 | 0.222 | 0.215 | 0.246 | 0.237 | 0.298 |
| 1996 | 0.024 | 0.052 | 0.073 | 0.095 | 0.118 | 0.129 | 0.148 | 0.172 | 0.183 | 0.185 | 0.202 | 0.224 | 0.233 | 0.229 | 0.280 | 0.332 |
| 1997 | 0.024 | 0.052 | 0.066 | 0.080 | 0.112 | 0.124 | 0.162 | 0.169 | 0.184 | 0.188 | 0.208 | 0.241 | 0.229 | 0.268 | 0.286 | 0.266 |
| 1998 | 0.024 | 0.052 | 0.071 | 0.090 | 0.108 | 0.129 | 0.142 | 0.151 | 0.162 | 0.174 | 0.191 | 0.220 | 0.229 | 0.268 | 0.286 | 0.271 |
| 1999 | 0.024 | 0.052 | 0.081 | 0.110 | 0.120 | 0.130 | 0.160 | 0.170 | 0.180 | 0.190 | 0.210 | 0.241 | 0.233 | 0.268 | 0.286 | 0.274 |
| 2000 | 0.024 | 0.052 | 0.102 | 0.115 | 0.128 | 0.158 | 0.169 | 0.181 | 0.208 | 0.224 | 0.225 | 0.227 | 0.247 | 0.247 | 0.272 | 0.378 |
| 2001 | 0.020 | 0.048 | 0.077 | 0.109 | 0.133 | 0.160 | 0.169 | 0.176 | 0.187 | 0.205 | 0.220 | 0.241 | 0.265 | 0.244 | 0.266 | 0.308 |
| 2002 | 0.020 | 0.039 | 0.067 | 0.133 | 0.152 | 0.164 | 0.175 | 0.194 | 0.202 | 0.222 | 0.242 | 0.275 | 0.299 | 0.307 | 0.306 | 0.329 |
| 2003 | 0.022 | 0.060 | 0.089 | 0.114 | 0.142 | 0.160 | 0.175 | 0.178 | 0.194 | 0.205 | 0.226 | 0.249 | 0.267 | 0.286 | 0.278 | 0.317 |
| 2004 | 0.036 | 0.064 | 0.100 | 0.120 | 0.148 | 0.168 | 0.186 | 0.201 | 0.219 | 0.209 | 0.221 | 0.233 | 0.262 | 0.260 | 0.322 | 0.303 |
| 2005 | 0.023 | 0.053 | 0.071 | 0.114 | 0.136 | 0.158 | 0.184 | 0.196 | 0.197 | 0.202 | 0.222 | 0.230 | 0.247 | 0.281 | 0.268 | 0.344 |
| 2006 | 0.019 | 0.038 | 0.078 | 0.114 | 0.141 | 0.154 | 0.180 | 0.199 | 0.212 | 0.222 | 0.235 | 0.229 | 0.235 | 0.248 | 0.253 | 0.304 |
| 2007 | 0.024 | 0.048 | 0.067 | 0.092 | 0.130 | 0.150 | 0.163 | 0.186 | 0.210 | 0.233 | 0.248 | 0.256 | 0.264 | 0.286 | 0.310 | 0.347 |
| 2008 | 0.031 | 0.051 | 0.082 | 0.116 | 0.144 | 0.164 | 0.176 | 0.190 | 0.240 | 0.251 | 0.251 | 0.281 | 0.279 | 0.289 | 0.293 | 0.352 |
| 2009 | 0.025 | 0.047 | 0.070 | 0.107 | 0.156 | 0.177 | 0.187 | 0.203 | 0.225 | 0.252 | 0.270 | 0.292 | 0.306 | 0.322 | 0.316 | 0.370 |
| 2010 | 0.026 | 0.048 | 0.087 | 0.118 | 0.151 | 0.178 | 0.201 | 0.212 | 0.229 | 0.248 | 0.274 | 0.305 | 0.312 | 0.335 | 0.329 | 0.376 |


| Year | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2011 | 0.028 | 0.051 | 0.079 | 0.112 | 0.151 | 0.172 | 0.192 | 0.211 | 0.223 | 0.243 | 0.261 | 0.288 | 0.305 | 0.324 | 0.329 | 0.330 |
| 2012 | 0.044 | 0.060 | 0.087 | 0.118 | 0.151 | 0.175 | 0.198 | 0.213 | 0.232 | 0.256 | 0.266 | 0.286 | 0.312 | 0.307 | 0.347 | 0.357 |
| 2013 | 0.040 | 0.058 | 0.102 | 0.130 | 0.154 | 0.172 | 0.195 | 0.228 | 0.243 | 0.249 | 0.248 | 0.288 | 0.288 | 0.321 | 0.348 |  |
| 2014 | 0.032 | 0.053 | 0.094 | 0.127 | 0.143 | 0.180 | 0.201 | 0.224 | 0.247 | 0.259 | 0.273 | 0.278 | 0.289 | 0.311 | 0.304 | 0.353 |
| 2015 | 0.021 | 0.082 | 0.083 | 0.137 | 0.144 | 0.176 | 0.200 | 0.219 | 0.235 | 0.256 | 0.279 | 0.285 | 0.297 | 0.313 | 0.312 | 0.348 |
| 2016 | 0.016 | 0.055 | 0.096 | 0.133 | 0.164 | 0.192 | 0.200 | 0.225 | 0.249 | 0.254 | 0.306 | 0.295 | 0.310 | 0.335 | 0.337 | 0.339 |
| 2017 | 0.016 | 0.039 | 0.077 | 0.098 | 0.124 | 0.173 | 0.199 | 0.216 | 0.249 | 0.266 | 0.286 | 0.307 | 0.333 | 0.334 | 0.337 | 0.370 |
| 2018 | 0.013 | 0.028 | 0.074 | 0.092 | 0.113 | 0.161 | 0.207 | 0.236 | 0.231 | 0.270 | 0.282 | 0.295 | 0.336 | 0.339 | 0.327 | 0.358 |
| 2019 | 0.011 | 0.032 | 0.074 | 0.108 | 0.156 | 0.159 | 0.205 | 0.237 | 0.268 | 0.277 | 0.304 | 0.309 | 0.346 | 0.386 | 0.400 | 0.402 |
| 2020 | 0.026 | 0.028 | 0.051 | 0.083 | 0.121 | 0.170 | 0.181 | 0.235 | 0.259 | 0.288 | 0.297 | 0.315 | 0.318 | 0.373 | 0.371 | 0.386 |
| 2021 | 0.027 | 0.042 | 0.072 | 0.090 | 0.122 | 0.169 | 0.205 | 0.219 | 0.262 | 0.282 | 0.301 | 0.328 | 0.340 | 0.334 | 0.383 | 0.408 |

Table 7.2.6.1. Western horse mackerel. Maturity-at-age.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1983 | 0 | 0 | 0.3 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1984 | 0 | 0 | 0.1 | 0.6 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1985 | 0 | 0 | 0.1 | 0.4 | 0.8 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1986 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1987 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1988 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1989 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1990 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1991 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1992 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1993 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1994 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1995 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1996 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1997 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1998 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1999 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2000 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2001 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2002 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2003 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2004 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2005 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2006 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2007 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2008 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2009 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2011 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2012 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2013 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2014 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2015 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2016 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2017 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2018 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2019 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2020 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2021 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |

Table 7.2.8.1. Western horse mackerel. Potential fecundity ( $10^{6}$ eggs) per kg spawning female vs. weight in kg.

| 1987 <br> w | pfec. | 1992 |  | 1995 |  | 1998 |  | 2000 |  | 2001 |  | 2001 (cont) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | w | pfec. | w | pfec. | w | pfec. | w | pfec. | w | pfec. | w | pfec. |
| 0.168 | 1.524 | 0.105 | 1.317 | 0.13 | 1.307 | 0.172 | 1.318 | 0.258 | 0.841 | 0.086 | 0.688 | 0.165 | 1.382 |
| 0.179 | 0.916 | 0.109 | 2.056 | 0.157 | 1.246 | 0.104 | 0.867 | 0.268 | 0.747 | 0.08 | 0.812 | 0.166 | 1.579 |
| 0.192 | 2.083 | 0.11 | 1.869 | 0.168 | 1.699 | 0.112 | 1.312 | 0.304 | 1.188 | 0.081 | 0.535 | 0.167 | 1.479 |
| 0.233 | 1.644 | 0.112 | 1.772 | 0.179 | 1.135 | 0.206 | 0.382 | 0.311 | 1.411 | 0.095 | 0.88 | 0.113 | 0.527 |
| 0.213 | 1.066 | 0.115 | 1.188 | 0.189 | 1.529 | 0.207 | 0.78 | 0.337 | 0.613 | 0.11 | 1.164 | 0.14 | 0.876 |
| 0.217 | 2.392 | 0.119 | 1.317 | 0.168 | 1.1 | 0.109 | 1.133 | 0.339 | 1.571 | 0.113 | 1.106 | 0.122 | 0.589 |
| 0.277 | 1.617 | 0.12 | 1.413 | 0.209 | 1.497 | 0.132 | 1.02 | 0.341 | 1.522 | 0.095 | 0.823 | 0.12 | 0.68 |
| 0.279 | 1.018 | 0.123 | 1.293 | 0.215 | 1.524 | 0.2 | 1.088 | 0.355 | 1.056 | 0.11 | 0.883 | 0.121 | 0.578 |
| 0.274 | 1.62 | 0.123 | 1.991 | 0.218 | 1.616 | 0.152 | 1.417 | 0.357 | 0.604 | 0.108 | 0.823 | 0.139 | 0.723 |
| 0.3 | 1.513 | 0.131 | 1.617 | 0.226 | 1.883 | 0.149 | 1.004 | 0.367 | 1.15 | 0.097 | 0.741 | 0.144 | 1.213 |
| 0.32 | 1.647 | 0.135 | 0.793 | 0.22 | 1.324 |  |  | 0.393 | 1.279 | 0.101 | 0.853 | 0.144 | 1.265 |
| 0.273 | 1.956 | 0.131 | 1.039 | 0.236 | 1.221 |  |  | 0.393 | 0.668 | 0.106 | 1.133 | 0.171 | 0.956 |
| 0.212 | 2.83 | 0.136 | 1.06 | 0.261 | 1.21 |  |  | 0.413 | 0.694 | 0.107 | 0.935 | 0.121 | 0.607 |
| 0.268 | 1.687 | 0.138 | 1.489 | 0.245 | 1.445 |  |  | 0.421 | 1.339 | 0.107 | 0.494 | 0.122 | 0.689 |


| $\begin{gathered} 1987 \\ 0.32 \end{gathered}$ | 1.088 | 1992 |  | 1995 | 1998 | 2000 |  | 2001 |  | 2001 (cont) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0.147 | 1.214 | 0.306 | 1.693 | 0.423 | 0.798 | 0.11 | 0.85 | 0.139 | 0.915 |
| 0.318 | 1.208 | 0.151 | 1.158 | 0.314 | 1.312 | 0.445 | 1.03 | 0.111 | 0.67 | 0.153 | 0.943 |
| 0.343 | 1.933 | 0.16 | 1.349 | 0.46 | 1.575 | 0.446 | 1.208 | 0.103 | 0.632 | 0.154 | 0.709 |
| 0.378 | 1.429 | 0.165 | 1.359 | 0.449 | 1.43 | 0.152 | 0.643 | 0.111 | 0.547 | 0.156 | 0.773 |
| 0.404 | 1.849 | 0.165 | 0.945 |  |  | 0.165 | 0.579 | 0.118 | 0.88 | 0.162 | 1.158 |
| 0.428 | 2.236 | 0.167 | 1 |  |  | 0.175 | 0.596 | 0.107 | 0.944 | 0.174 | 1.389 |
| 0.398 | 1.538 | 0.168 | 1.545 |  |  | 0.179 | 0.997 | 0.104 | 0.724 | 0.175 | 1.426 |
| 0.431 | 1.223 | 0.18 | 1.299 |  |  | 0.19 | 0.744 | 0.111 | 0.86 | 0.179 | 1.248 |
| 0.432 | 1.465 | 0.174 | 1.487 |  |  | 0.197 | 0.613 | 0.11 | 0.728 | 0.179 | 1.236 |
| 0.421 | 1.843 | 0.178 | 1.594 |  |  | 0.203 | 0.702 | 0.111 | 0.544 | 0.18 | 2.353 |
| 0.481 | 1.757 | 0.185 | 1.475 |  |  | 0.219 | 0.472 | 0.129 | 0.935 | 0.184 | 2.255 |
| 0.494 | 1.611 | 0.195 | 1.41 |  |  | 0.223 | 0.806 | 0.114 | 0.901 | 0.139 | 0.931 |
| 0.54 | 1.754 | 0.203 | 1.937 |  |  | 0.227 | 0.606 | 0.114 | 0.557 | 0.161 | 1.037 |
| 0.564 | 2.255 | 0.205 | 1.534 |  |  | 0.289 | 1.273 | 0.151 | 1.377 | 0.162 | 0.893 |
| 0.585 | 1.221 | 0.213 | 1.577 |  |  | 0.294 | 1.395 | 0.153 | 1.596 | 0.169 | 0.691 |
|  |  | 0.222 | 0.958 |  |  | 0.3 | 1.305 | 0.154 | 1.699 | 0.18 | 1.609 |
|  |  | 0.275 | 2.444 |  |  |  |  | 0.103 | 0.679 | 0.185 | 1.776 |
|  |  |  |  |  |  |  |  | 0.12 | 1.14 | 0.211 | 2.102 |
|  |  |  |  |  |  |  |  | 0.12 | 0.631 | 0.224 | 1.466 |
|  |  |  |  |  |  |  |  | 0.121 | 0.834 | 0.162 | 0.849 |
|  |  |  |  |  |  |  |  | 0.144 | 0.626 | 0.17 | 0.668 |
|  |  |  |  |  |  |  |  | 0.116 | 0.668 | 0.187 | 1.453 |
|  |  |  |  |  |  |  |  | 0.118 | 1.194 | 0.198 | 1.371 |
|  |  |  |  |  |  |  |  | 0.112 | 0.779 | 0.219 | 1.847 |
|  |  |  |  |  |  |  |  | 0.126 | 0.782 | 0.22 | 1.578 |
|  |  |  |  |  |  |  |  | 0.139 | 1.244 | 0.201 | 0.878 |
|  |  |  |  |  |  |  |  | 0.119 | 1.212 | 0.206 | 1.196 |
|  |  |  |  |  |  |  |  | 0.109 | 0.755 | 0.223 | 1.115 |
|  |  |  |  |  |  |  |  | 0.122 | 0.841 | 0.225 | 1.43 |


| 1987 | 1992 | 1998 | 2000 | 2001 | 201 (cont) |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 8 | 0.131 | 0.929 | 0.233 | 1.724 |  |
|  | 0.135 | 0.862 | 0.241 | 1.131 |  |
|  | 0.142 | 1.834 | 0.219 | 0.96 |  |
|  | 0.146 | 1.689 | 0.237 | 1.33 |  |

Table 7.3.1.1. Western horse mackerel. Final assessment. Numbers-at-age (thousands).

| yea r | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 198 | 478369 | 122480 | 250831 | 576778 | 106579 | 137759 | 125194 |  |  | 43703 | 40717 | 47048 | 56449 | 70717 | 40383 | 25872 | 22981 | 20300 | 17888 | 15770 | 11748 |
| 2 | 00 | 0 | 0 | 0 | 0 | 0 | 0 | 752509 | 489208 | 2 | 5 | 2 | 3 | 1 | 9 | 4 | 1 | 1 | 2 | 8 | 20 |
| 198 | 150636 | 411383 | 105067 | 214228 | 490132 |  | 116212 | 105426 |  | 41139 | 36744 | 34230 | 39550 | 47452 | 59444 | 33946 | 21748 | 19317 | 17064 | 15036 | 11201 |
| 3 | 0 | 00 | 0 | 0 | 0 | 901783 | 0 | 0 | 633103 | 6 | 0 | 3 | 3 | 0 | 7 | 4 | 1 | 6 | 0 | 5 | 00 |
| 198 | 161894 | 129501 | 352454 |  | 181178 | 412063 |  |  |  | 52779 | 34286 | 30619 | 28522 | 32954 | 39537 | 49528 | 28283 | 18120 | 16095 | 14217 | 10585 |
| 4 | 0 | 0 | 00 | 894754 | 0 | 0 | 755036 | 970653 | 879452 | 8 | 7 | 0 | 2 | 0 | 0 | 8 | 6 | 1 | 0 | 3 | 20 |
| 198 | 212757 | 139193 | 110995 | 300430 |  | 152632 | 345841 |  |  | 73521 | 44111 | 28652 | 25585 | 23832 | 27535 | 33035 | 41383 | 23632 | 15140 | 13448 | 10032 |
| 5 | 0 | 0 | 0 | 00 | 757875 | 0 | 0 | 632296 | 811919 | 7 | 7 | 1 | 5 | 6 | 2 | 4 | 8 | 3 | 1 | 1 | 30 |
| 198 | 265939 | 182958 | 119388 |  | 255167 |  | 128654 | 290976 |  | 68213 | 61755 | 37048 | 24063 | 21486 | 20014 | 23123 | 27742 | 34753 | 19845 | 12714 | 95542 |
| 6 | 0 | 0 | 0 | 947683 | 00 | 640806 | 0 | 0 | 531476 | 9 | 8 | 5 | 0 | 9 | 6 | 8 | 6 | 4 | 9 | 4 | 2 |
| 198 | 522742 | 228643 | 156797 | 101742 |  | 214854 |  | 107670 | 243227 | 44400 | 56971 | 51570 | 30936 | 20092 | 17941 | 16711 | 19307 | 23164 | 29017 | 16570 | 90389 |
| 7 | 0 | 0 | 0 | 0 | 802401 | 00 | 537501 | 0 | 0 | 4 | 2 | 8 | 2 | 4 | 1 | 5 | 5 | 0 | 6 | 5 | 1 |
| 198 | 282829 | 449299 | 195724 | 133267 |  |  | 178975 |  |  | 20157 | 36784 | 47191 | 42714 | 25622 | 16640 | 14858 | 13840 | 15990 | 19184 | 24031 | 88581 |
| 8 | 0 | 0 | 0 | 0 | 857669 | 671669 | 00 | 446460 | 892982 | 80 | 6 | 3 | 4 | 4 | 7 | 8 | 4 | 2 | 1 | 9 | 9 |
| 198 | 317242 | 243054 | 384368 | 166110 | 112071 |  |  | 148048 |  | 73683 | 16626 | 30334 | 38913 | 35219 | 21126 | 13720 | 12251 | 11411 | 13183 | 15817 | 92848 |
| 9 | 0 | 0 | 0 | 0 | 0 | 715617 | 557387 | 00 | 368692 | 1 | 40 | 7 | 1 | 8 | 2 | 4 | 1 | 3 | 9 | 1 | 7 |
| 199 | 221323 | 272617 | 207900 | 326105 | 139615 |  |  |  | 122144 | 30392 | 60715 | 13697 | 24988 | 32054 | 29010 | 17401 | 11301 | 10091 |  | 10859 | 89505 |
| 0 | 0 | 0 | 0 | 0 | 0 | 934424 | 593360 | 460650 | 00 | 6 | 4 | 70 | 9 | 0 | 9 | 6 | 4 | 0 | 93993 | 2 | 3 |
| 199 | 391775 | 190071 | 232615 | 175388 | 271535 | 114958 |  |  |  | 98987 | 24617 | 49164 | 11090 | 20230 | 25949 | 23485 | 14087 |  |  |  | 81246 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 763457 | 482586 | 373756 | 50 | 0 | 3 | 20 | 7 | 5 | 5 | 1 | 91487 | 81689 | 76089 | 0 |
| 199 | 765957 | 336358 | 162000 | 195733 | 145419 | 222311 |  |  |  | 30056 | 79553 | 19778 | 39494 | 89081 | 16249 | 20842 | 18863 | 11314 |  |  | 71364 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 932988 | 616429 | 388597 | 5 | 40 | 0 | 2 | 8 | 6 | 5 | 1 | 4 | 73480 | 65610 | 7 |
| 199 | 696138 | 656750 | 285219 | 134713 | 159150 | 115988 | 174972 |  |  | 30160 | 23305 | 61657 | 15325 | 30599 | 69014 | 12588 | 16146 | 14612 |  |  | 60365 |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 728580 | 479399 | 6 | 9 | 60 | 4 | 3 | 3 | 5 | 3 | 7 | 87648 | 56922 | 2 |


| yea <br> r | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 199 | 638588 | 596165 | 554279 | 234609 | 107583 | 123921 |  | 132500 |  | 36011 | 22627 | 17474 | 46216 | 11485 | 22930 | 51715 |  | 12098 | 10949 |  | 49495 |
| 4 | 0 | 0 | 0 | 0 | 0 | 0 | 887447 | 0 | 548746 | 6 | 7 | 4 | 30 | 6 | 6 | 6 | 94329 | 6 | 3 | 65674 | 7 |
| 199 | 383672 | 546784 | 502804 | 455208 | 186872 |  |  |  |  | 41026 | 26888 | 16884 | 13035 | 34470 |  | 17100 | 38565 |  |  |  | 41805 |
| 5 | 0 | 0 | 0 | 0 | 0 | 834775 | 944280 | 669045 | 993332 | 3 | 6 | 8 | 4 | 40 | 85658 | 4 | 5 | 70342 | 90219 | 81648 | 3 |
| 199 | 215597 | 327698 | 456695 | 403769 | 349390 | 137985 |  |  |  | 69426 | 28619 | 18739 | 11762 |  | 24003 |  | 11906 | 26851 |  |  | 34790 |
| 6 | 0 | 0 | 0 | 0 | 0 | 0 | 600088 | 668167 | 469507 | 3 | 1 | 7 | 3 | 90786 | 90 | 59644 | 7 | 6 | 48975 | 62814 | 3 |
| 199 | 149721 | 184327 | 274770 | 370043 | 314529 | 263145 | 101527 |  |  | 33709 | 49763 | 20497 | 13416 |  |  | 17178 |  |  | 19214 |  | 29389 |
| 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 435502 | 481423 | 6 | 2 | 1 | 1 | 84191 | 64974 | 20 | 42682 | 85204 | 6 | 35046 | 4 |
| 199 | 257417 | 127654 | 152904 | 217182 | 276686 | 224255 | 181542 |  |  | 32085 | 22413 | 33049 | 13605 |  |  |  | 11395 |  |  | 12745 | 21818 |
| 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 686963 | 291691 | 0 | 1 | 6 | 2 | 89025 | 55858 | 43104 | 50 | 28313 | 56518 | 4 | 5 |
| 199 | 271147 | 220137 | 107141 | 124175 | 169819 | 209440 | 165981 | 132601 |  | 21086 | 23157 | 16164 | 23825 |  |  |  |  | 82115 |  |  | 24905 |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 498296 | 5 | 1 | 0 | 7 | 98062 | 64159 | 40253 | 31061 | 2 | 20402 | 40725 | 6 |
| 200 | 199939 | 231858 | 184700 |  |  | 128318 | 154694 | 120961 |  | 35935 | 15182 | 16659 | 11624 | 17130 |  |  |  |  | 59026 |  | 20829 |
| 0 | 0 | 0 | 0 | 869408 | 969665 | 0 | 0 | 0 | 959566 | 5 | 1 | 8 | 3 | 7 | 70499 | 46123 | 28936 | 22328 | 1 | 14665 | 5 |
| 200 | 118461 | 171252 | 195797 | 152133 |  |  |  | 117596 |  | 72376 | 27071 | 11430 | 12539 |  | 12891 |  |  |  |  | 44412 | 16775 |
| 1 | 00 | 0 | 0 | 0 | 695878 | 757304 | 985274 | 0 | 914706 | 5 | 8 | 6 | 6 | 87481 | 1 | 53049 | 34705 | 21773 | 16800 | 6 | 8 |
| 200 | 217936 | 101348 | 143970 | 159612 | 119702 |  |  |  |  | 66880 | 52839 | 19749 |  |  |  |  |  |  |  |  | 44603 |
| 2 | 0 | 00 | 0 | 0 | 0 | 531176 | 566053 | 727393 | 862551 | 3 | 7 | 9 | 83361 | 91432 | 63780 | 93979 | 38673 | 25299 | 15872 | 12247 | 8 |
| 200 | 106411 | 186545 | 853692 | 117895 | 126528 |  |  |  |  | 63928 | 49500 | 39082 | 14603 |  |  |  |  |  |  |  | 33871 |
| 3 | 0 | 0 | 0 | 0 | 0 | 922884 | 401720 | 423267 | 540682 | 8 | 5 | 7 | 2 | 61627 | 67587 | 47144 | 69464 | 28584 | 18699 | 11731 | 8 |
| 200 | 194900 |  | 157199 | 699736 |  |  |  |  |  | 40187 | 47451 | 36718 | 28981 | 10827 |  |  |  |  |  |  | 25976 |
| 4 | 0 | 910936 | 0 | 0 | 936037 | 977550 | 699678 | 301190 | 315498 | 0 | 7 | 4 | 3 | 0 | 45687 | 50103 | 34947 | 51492 | 21188 | 13861 | 7 |
| 200 | 148148 | 167032 |  | 130155 | 564898 |  |  |  |  | 24170 | 30754 | 36295 | 28078 | 22159 |  |  |  |  |  |  | 20917 |
| 5 | 0 | 0 | 770991 | 0 | 0 | 739481 | 760790 | 539744 | 231269 | 8 | 9 | 8 | 7 | 2 | 82778 | 34929 | 38303 | 26717 | 39364 | 16198 | 8 |
| 200 | 123143 | 126912 | 141144 |  | 104431 | 442621 |  |  |  | 17521 | 18290 | 23259 | 27442 | 21226 | 16750 |  |  |  |  |  | 17034 |
| 6 | 0 | 0 | 0 | 635985 | 0 | 0 | 569975 | 580750 | 409927 | 1 | 4 | 5 | 3 | 5 | 2 | 62569 | 26401 | 28951 | 20193 | 29752 | 1 |


| yea $r$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 200 | 195680 | 105562 | 107523 | 117133 |  |  | 346550 |  |  | 31624 | 13503 | 14089 | 17912 | 21131 | 16343 | 12896 |  |  |  |  | 15404 |
| 7 | 0 | 0 | 0 | 0 | 515397 | 829243 | 0 | 442570 | 448977 | 1 | 1 | 1 | 5 | 1 | 7 | 6 | 48173 | 20326 | 22289 | 15547 | 9 |
| 200 | 494533 | 167871 |  |  |  |  |  | 274343 |  | 35356 | 24883 | 10620 | 11079 | 14084 | 16614 | 12849 | 10139 |  |  |  | 13333 |
| 8 | 0 | 0 | 897001 | 898450 | 960040 | 415500 | 660907 | 0 | 349120 | 2 | 0 | 5 | 3 | 5 | 4 | 9 | 5 | 37874 | 15980 | 17524 | 3 |
| 200 | 127719 | 423942 | 142240 |  |  |  |  |  | 212255 | 26953 | 27269 | 19182 |  |  | 10853 | 12802 |  |  |  |  | 11623 |
| 9 | 0 | 0 | 0 | 744600 | 728405 | 762764 | 325533 | 513549 | 0 | 8 | 4 | 4 | 81855 | 85380 | 2 | 1 | 99011 | 78126 | 29182 | 12313 | 4 |
| 201 |  | 109316 | 357005 | 116404 |  |  |  |  |  | 15701 | 19911 | 20131 | 14156 |  |  |  |  |  |  |  |  |
| 0 | 938294 | 0 | 0 | 0 | 589646 | 560805 | 575922 | 242984 | 381021 | 70 | 4 | 0 | 3 | 60397 | 62992 | 80069 | 94444 | 73041 | 57633 | 21527 | 94826 |
| 201 |  |  |  | 288960 |  |  |  |  |  | 27269 | 11218 | 14214 | 14365 | 10099 |  |  |  |  |  |  |  |
| 1 | 344757 | 802108 | 916174 | 0 | 905176 | 443052 | 411492 | 416708 | 174525 | 2 | 40 | 5 | 4 | 8 | 43085 | 44933 | 57112 | 67364 | 52097 | 41107 | 82987 |
| 201 | 241707 |  |  |  | 223604 |  |  |  |  | 12380 | 19309 | 79368 | 10052 | 10156 |  |  |  |  |  |  |  |
| 2 | 0 | 294622 | 671386 | 739362 | 0 | 675707 | 322604 | 295259 | 296710 | 2 | 2 | 9 | 3 | 9 | 71400 | 30457 | 31762 | 40370 | 47615 | 36824 | 87712 |
| 201 | 105324 | 206608 |  |  |  | 167734 |  |  |  | 21187 |  | 13753 | 56507 |  |  |  |  |  |  |  |  |
| 3 | 0 | 0 | 246842 | 543011 | 574223 | 0 | 494832 | 232924 | 211602 | 2 | 88251 | 0 | 4 | 71553 | 72289 | 50814 | 21675 | 22603 | 28728 | 33884 | 88619 |
| 201 | 337547 |  | 172657 |  |  |  | 120958 |  |  | 14843 | 14834 |  |  | 39500 |  |  |  |  |  |  |  |
| 4 | 0 | 899701 | 0 | 198463 | 417618 | 425139 | 0 | 351342 | 164041 | 3 | 2 | 61733 | 96161 | 8 | 50012 | 50523 | 35512 | 15147 | 15796 | 20076 | 85606 |
| 201 | 239612 | 288478 |  | 139415 |  |  |  |  |  | 11656 | 10528 | 10513 |  |  | 27973 |  |  |  |  |  |  |
| 5 | 0 | 0 | 753259 | 0 | 153720 | 312140 | 310017 | 869306 | 250586 | 3 | 6 | 2 | 43733 | 68107 | 6 | 35415 | 35775 | 25145 | 10725 | 11184 | 74828 |
| 201 | 277767 | 205094 | 242971 |  | 110473 |  |  |  |  | 18558 |  |  |  |  |  | 20661 |  |  |  |  |  |
| 6 | 0 | 0 | 0 | 616682 | 0 | 118452 | 235918 | 231657 | 645708 | 9 | 86209 | 77817 | 77678 | 32307 | 50308 | 9 | 26157 | 26423 | 18572 | 7921 | 63524 |
| 201 | 363380 | 237716 | 172635 | 198638 |  |  |  |  |  | 47621 | 13667 |  |  |  |  |  | 15195 |  |  |  |  |
| 7 | 0 | 0 | 0 | 0 | 487532 | 848642 | 89202 | 175591 | 171363 | 6 | 8 | 63446 | 57250 | 57138 | 23762 | 37000 | 6 | 19237 | 19432 | 13658 | 52541 |
| 201 | 296823 | 311314 | 200926 | 142495 | 159543 |  |  |  |  | 13006 | 36104 | 10356 |  |  |  |  |  | 11507 |  |  |  |
| 8 | 0 | 0 | 0 | 0 | 0 | 382530 | 655178 | 68213 | 133605 | 9 | 2 | 5 | 48061 | 43362 | 43273 | 17995 | 28020 | 3 | 14568 | 14715 | 50129 |
| 201 | 135642 | 254120 | 262434 | 164828 | 113290 | 123488 |  |  |  |  |  | 26849 |  |  |  |  |  |  |  |  |  |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 | 290637 | 492372 | 50969 | 99551 | 96788 | 0 | 76992 | 35724 | 32228 | 32160 | 13373 | 20823 | 85515 | 10826 | 48187 |


| yea <br> r | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 202 | 108396 | 116033 | 213544 | 213720 | 129477 |  |  |  |  |  |  |  | 19520 |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 862859 | 920631 | 213963 | 360089 | 37155 | 72459 | 70397 | 8 | 55967 | 25966 | 23423 | 23373 | 9719 | 15133 | 62148 | 42886 |
| 202 | 816224 | 998917 | 912502 | 138485 | 134158 | 875566 | 581151 | 630308 | 143506 | 23906 | 25200 | 46793 | 43929 | 12192 | 38258 | 15285 | 13605 | 13479 | 5583 | 8454 | 62126 |
| 1 |  |  |  | 0 | 0 |  |  |  |  | 6 |  |  |  | 4 |  |  |  |  |  |  |  |

Table 7.3.1.2. Western horse mackerel. Final assessment. Fishing mortality-at-age.

| year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0.001 | 0.003 | 0.008 | 0.013 | 0.017 | 0.020 | 0.022 | 0.023 | 0.023 | 0.023 | 0.024 | 0.024 | 0.024 | 0.024 | 0.024 | 0.024 | 0.024 | 0.024 | 0.024 | 0.024 | 0.024 |
| 1983 | 0.001 | 0.005 | 0.011 | 0.018 | 0.023 | 0.028 | 0.030 | 0.031 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.033 | 0.033 | 0.033 | 0.033 | 0.033 | 0.033 |
| 1984 | 0.001 | 0.004 | 0.010 | 0.016 | 0.021 | 0.025 | 0.027 | 0.029 | 0.029 | 0.029 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 |
| 1985 | 0.001 | 0.003 | 0.008 | 0.013 | 0.018 | 0.021 | 0.023 | 0.024 | 0.024 | 0.024 | 0.024 | 0.025 | 0.025 | 0.025 | 0.025 | 0.025 | 0.025 | 0.025 | 0.025 | 0.025 | 0.025 |
| 1986 | 0.001 | 0.004 | 0.010 | 0.016 | 0.022 | 0.026 | 0.028 | 0.029 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 | 0.030 |
| 1987 | 0.001 | 0.005 | 0.013 | 0.021 | 0.028 | 0.033 | 0.036 | 0.037 | 0.038 | 0.038 | 0.038 | 0.038 | 0.038 | 0.038 | 0.039 | 0.039 | 0.039 | 0.039 | 0.039 | 0.039 | 0.039 |
| 1988 | 0.002 | 0.006 | 0.014 | 0.023 | 0.031 | 0.037 | 0.040 | 0.041 | 0.042 | 0.043 | 0.043 | 0.043 | 0.043 | 0.043 | 0.043 | 0.043 | 0.043 | 0.043 | 0.043 | 0.043 | 0.043 |
| 1989 | 0.002 | 0.006 | 0.014 | 0.024 | 0.032 | 0.037 | 0.041 | 0.042 | 0.043 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 |
| 1990 | 0.002 | 0.009 | 0.020 | 0.033 | 0.044 | 0.052 | 0.057 | 0.059 | 0.060 | 0.061 | 0.061 | 0.061 | 0.061 | 0.061 | 0.061 | 0.061 | 0.061 | 0.061 | 0.061 | 0.061 | 0.061 |
| 1991 | 0.003 | 0.010 | 0.023 | 0.037 | 0.050 | 0.059 | 0.064 | 0.067 | 0.068 | 0.069 | 0.069 | 0.069 | 0.069 | 0.069 | 0.069 | 0.069 | 0.069 | 0.069 | 0.069 | 0.069 | 0.069 |
| 1992 | 0.004 | 0.015 | 0.034 | 0.057 | 0.076 | 0.089 | 0.097 | 0.101 | 0.103 | 0.104 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 |
| 1993 | 0.005 | 0.020 | 0.045 | 0.075 | 0.100 | 0.118 | 0.128 | 0.133 | 0.136 | 0.137 | 0.138 | 0.138 | 0.138 | 0.139 | 0.139 | 0.139 | 0.139 | 0.139 | 0.139 | 0.139 | 0.139 |
| 1994 | 0.005 | 0.020 | 0.047 | 0.077 | 0.104 | 0.122 | 0.132 | 0.138 | 0.141 | 0.142 | 0.143 | 0.143 | 0.143 | 0.143 | 0.143 | 0.143 | 0.143 | 0.143 | 0.143 | 0.143 | 0.143 |


| year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 0.008 | 0.030 | 0.069 | 0.115 | 0.153 | 0.180 | 0.196 | 0.204 | 0.208 | 0.210 | 0.211 | 0.212 | 0.212 | 0.212 | 0.212 | 0.212 | 0.212 | 0.212 | 0.212 | 0.212 | 0.212 |
| 1996 | 0.007 | 0.026 | 0.060 | 0.100 | 0.133 | 0.157 | 0.171 | 0.178 | 0.181 | 0.183 | 0.184 | 0.184 | 0.184 | 0.185 | 0.185 | 0.185 | 0.185 | 0.185 | 0.185 | 0.185 | 0.185 |
| 1997 | 0.009 | 0.037 | 0.085 | 0.141 | 0.188 | 0.221 | 0.241 | 0.251 | 0.256 | 0.258 | 0.259 | 0.260 | 0.260 | 0.260 | 0.260 | 0.260 | 0.260 | 0.260 | 0.261 | 0.261 | 0.261 |
| 1998 | 0.006 | 0.025 | 0.058 | 0.096 | 0.128 | 0.151 | 0.164 | 0.171 | 0.174 | 0.176 | 0.177 | 0.177 | 0.177 | 0.178 | 0.178 | 0.178 | 0.178 | 0.178 | 0.178 | 0.178 | 0.178 |
| 1999 | 0.007 | 0.026 | 0.059 | 0.097 | 0.130 | 0.153 | 0.166 | 0.173 | 0.177 | 0.179 | 0.179 | 0.180 | 0.180 | 0.180 | 0.180 | 0.180 | 0.180 | 0.180 | 0.180 | 0.180 | 0.180 |
| 2000 | 0.005 | 0.019 | 0.044 | 0.073 | 0.097 | 0.114 | 0.124 | 0.129 | 0.132 | 0.133 | 0.134 | 0.134 | 0.134 | 0.134 | 0.134 | 0.134 | 0.134 | 0.134 | 0.134 | 0.134 | 0.134 |
| 2001 | 0.006 | 0.024 | 0.054 | 0.090 | 0.120 | 0.141 | 0.153 | 0.160 | 0.163 | 0.165 | 0.165 | 0.166 | 0.166 | 0.166 | 0.166 | 0.166 | 0.166 | 0.166 | 0.166 | 0.166 | 0.166 |
| 2002 | 0.006 | 0.022 | 0.050 | 0.082 | 0.110 | 0.129 | 0.141 | 0.147 | 0.150 | 0.151 | 0.152 | 0.152 | 0.152 | 0.152 | 0.152 | 0.152 | 0.152 | 0.152 | 0.152 | 0.152 | 0.152 |
| 2003 | 0.005 | 0.021 | 0.049 | 0.081 | 0.108 | 0.127 | 0.138 | 0.144 | 0.147 | 0.148 | 0.149 | 0.149 | 0.149 | 0.149 | 0.149 | 0.149 | 0.149 | 0.149 | 0.149 | 0.149 | 0.149 |
| 2004 | 0.004 | 0.017 | 0.039 | 0.064 | 0.086 | 0.101 | 0.110 | 0.114 | 0.116 | 0.117 | 0.118 | 0.118 | 0.118 | 0.118 | 0.119 | 0.119 | 0.119 | 0.119 | 0.119 | 0.119 | 0.119 |
| 2005 | 0.005 | 0.018 | 0.043 | 0.070 | 0.094 | 0.110 | 0.120 | 0.125 | 0.128 | 0.129 | 0.129 | 0.130 | 0.130 | 0.130 | 0.130 | 0.130 | 0.130 | 0.130 | 0.130 | 0.130 | 0.130 |
| 2006 | 0.004 | 0.016 | 0.036 | 0.060 | 0.081 | 0.095 | 0.103 | 0.107 | 0.109 | 0.110 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.112 | 0.112 | 0.112 |
| 2007 | 0.003 | 0.013 | 0.030 | 0.049 | 0.065 | 0.077 | 0.084 | 0.087 | 0.089 | 0.090 | 0.090 | 0.090 | 0.090 | 0.090 | 0.091 | 0.091 | 0.091 | 0.091 | 0.091 | 0.091 | 0.091 |
| 2008 | 0.004 | 0.016 | 0.036 | 0.060 | 0.080 | 0.094 | 0.102 | 0.107 | 0.109 | 0.110 | 0.110 | 0.110 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 |
| 2009 | 0.006 | 0.022 | 0.050 | 0.083 | 0.111 | 0.131 | 0.142 | 0.148 | 0.151 | 0.153 | 0.154 | 0.154 | 0.154 | 0.154 | 0.154 | 0.154 | 0.154 | 0.154 | 0.154 | 0.154 | 0.154 |
| 2010 | 0.007 | 0.027 | 0.061 | 0.102 | 0.136 | 0.160 | 0.174 | 0.181 | 0.185 | 0.186 | 0.187 | 0.187 | 0.188 | 0.188 | 0.188 | 0.188 | 0.188 | 0.188 | 0.188 | 0.188 | 0.188 |
| 2011 | 0.007 | 0.028 | 0.064 | 0.106 | 0.142 | 0.167 | 0.182 | 0.190 | 0.193 | 0.195 | 0.196 | 0.196 | 0.197 | 0.197 | 0.197 | 0.197 | 0.197 | 0.197 | 0.197 | 0.197 | 0.197 |
| 2012 | 0.007 | 0.027 | 0.062 | 0.103 | 0.137 | 0.162 | 0.176 | 0.183 | 0.187 | 0.188 | 0.189 | 0.190 | 0.190 | 0.190 | 0.190 | 0.190 | 0.190 | 0.190 | 0.190 | 0.190 | 0.190 |


| year | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5}$ | $\mathbf{1 6}$ | $\mathbf{1 7}$ | $\mathbf{1 8}$ | $\mathbf{1 9}$ | $\mathbf{2 0}$ |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2013 | 0.008 | 0.030 | 0.068 | 0.113 | 0.151 | 0.177 | 0.192 | 0.201 | 0.205 | 0.206 | 0.207 | 0.208 | 0.208 | 0.208 | 0.208 | 0.208 | 0.208 | 0.208 | 0.208 | 0.208 | 0.208 |  |
| 2014 | 0.007 | 0.028 | 0.064 | 0.105 | 0.141 | 0.166 | 0.180 | 0.188 | 0.192 | 0.193 | 0.194 | 0.195 | 0.195 | 0.195 | 0.195 | 0.195 | 0.195 | 0.195 | 0.195 | 0.195 | 0.195 |  |
| 2015 | 0.006 | 0.022 | 0.050 | 0.083 | 0.111 | 0.130 | 0.141 | 0.147 | 0.150 | 0.152 | 0.152 | 0.153 | 0.153 | 0.153 | 0.153 | 0.153 | 0.153 | 0.153 | 0.153 | 0.153 | 0.153 |  |
| 2016 | 0.006 | 0.022 | 0.051 | 0.085 | 0.114 | 0.134 | 0.145 | 0.151 | 0.154 | 0.156 | 0.157 | 0.157 | 0.157 | 0.157 | 0.157 | 0.157 | 0.157 | 0.157 | 0.157 | 0.157 | 0.157 |  |
| 2017 | 0.005 | 0.018 | 0.042 | 0.069 | 0.093 | 0.109 | 0.118 | 0.123 | 0.126 | 0.127 | 0.127 | 0.128 | 0.128 | 0.128 | 0.128 | 0.128 | 0.128 | 0.128 | 0.128 | 0.128 | 0.128 |  |
| 2018 | 0.005 | 0.021 | 0.048 | 0.079 | 0.106 | 0.125 | 0.136 | 0.141 | 0.144 | 0.146 | 0.146 | 0.147 | 0.147 | 0.147 | 0.147 | 0.147 | 0.147 | 0.147 | 0.147 | 0.147 | 0.147 |  |
| 2019 | 0.006 | 0.024 | 0.055 | 0.091 | 0.122 | 0.144 | 0.156 | 0.163 | 0.166 | 0.168 | 0.168 | 0.169 | 0.169 | 0.169 | 0.169 | 0.169 | 0.169 | 0.169 | 0.169 | 0.169 | 0.169 |  |
| 2020 | 0.003 | 0.014 | 0.031 | 0.052 | 0.069 | 0.081 | 0.088 | 0.092 | 0.094 | 0.095 | 0.095 | 0.095 | 0.095 | 0.095 | 0.095 | 0.095 | 0.095 | 0.095 | 0.095 | 0.095 | 0.095 |  |
| 2021 | 0.004 | 0.016 | 0.037 | 0.061 | 0.082 | 0.096 | 0.106 | 0.111 | 0.113 | 0.115 | 0.115 | 0.116 | 0.116 | 0.116 | 0.116 | 0.116 | 0.116 | 0.116 | 0.116 | 0.116 | $N A$ |  |

Table 7.3.1.3. Western horse mackerel. Final assessment. Stock summary table.

| Year | Recruit (thousands) | Total <br> Biomass | Spawning <br> Biomass | Catch | Yield/SSB | Fbar(1-3) | Fbar(4-8) | Fbar(1-10) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 47973500 | 3007480 | 2351030 | 61197 | 0.0260 | 0.0081 | 0.0218 | 0.0183 |
| 1983 | 1409080 | 3517800 | 2477960 | 90442 | 0.0365 | 0.0111 | 0.0299 | 0.0251 |
| 1984 | 1529300 | 4168190 | 2600980 | 96244 | 0.0370 | 0.0101 | 0.0271 | 0.0227 |
| 1985 | 2061350 | 4763890 | 3044610 | 96343 | 0.0316 | 0.0083 | 0.0223 | 0.0187 |
| 1986 | 2605390 | 5186500 | 4307560 | 137499 | 0.0319 | 0.0102 | 0.0274 | 0.0230 |
| 1987 | 5209590 | 5376550 | 5036810 | 187338 | 0.0372 | 0.0128 | 0.0345 | 0.0289 |
| 1988 | 2733740 | 5358590 | 5073630 | 210989 | 0.0416 | 0.0143 | 0.0384 | 0.0322 |
| 1989 | 3106180 | 5198870 | 4865000 | 209583 | 0.0431 | 0.0145 | 0.0392 | 0.0328 |
| 1990 | 2128810 | 4956330 | 4599300 | 275968 | 0.0600 | 0.0203 | 0.0546 | 0.0458 |
| 1991 | 3848570 | 4599130 | 4282640 | 287438 | 0.0671 | 0.0229 | 0.0617 | 0.0517 |
| 1992 | 7613940 | 4233590 | 3922920 | 393631 | 0.1003 | 0.0349 | 0.0941 | 0.0789 |
| 1993 | 6952140 | 3823250 | 3443150 | 453246 | 0.1316 | 0.0461 | 0.1242 | 0.1041 |
| 1994 | 6397130 | 3445010 | 2930260 | 412291 | 0.1407 | 0.0479 | 0.1290 | 0.1081 |
| 1995 | 3888900 | 3184580 | 2568610 | 538950 | 0.2098 | 0.0711 | 0.1915 | 0.1605 |
| 1996 | 2130800 | 2835840 | 2258940 | 422396 | 0.1870 | 0.0620 | 0.1671 | 0.1400 |
| 1997 | 1455050 | 2588010 | 2130370 | 534673 | 0.2510 | 0.0875 | 0.2357 | 0.1975 |
| 1998 | 2488060 | 2194950 | 1887400 | 325340 | 0.1724 | 0.0596 | 0.1605 | 0.1345 |
| 1999 | 2653150 | 1973540 | 1751820 | 298992 | 0.1707 | 0.0603 | 0.1624 | 0.1361 |
| 2000 | 1952630 | 1759820 | 1556520 | 202732 | 0.1302 | 0.0450 | 0.1211 | 0.1015 |
| 2001 | 11569200 | 1668810 | 1411810 | 229081 | 0.1623 | 0.0557 | 0.1499 | 0.1256 |
| 2002 | 1982640 | 1619510 | 1257880 | 196120 | 0.1559 | 0.0512 | 0.1380 | 0.1156 |
| 2003 | 988367 | 1638910 | 1167000 | 191856 | 0.1644 | 0.0505 | 0.1360 | 0.1140 |
| 2004 | 1801970 | 1650310 | 1171620 | 159742 | 0.1363 | 0.0402 | 0.1083 | 0.0908 |
| 2005 | 1366060 | 1652080 | 1362250 | 182001 | 0.1336 | 0.0442 | 0.1191 | 0.0998 |
| 2006 | 1137640 | 1584930 | 1418690 | 155827 | 0.1098 | 0.0380 | 0.1024 | 0.0859 |
| 2007 | 2071850 | 1505090 | 1359080 | 123356 | 0.0908 | 0.0310 | 0.0835 | 0.0700 |
| 2008 | 4713810 | 1447290 | 1290160 | 143349 | 0.1111 | 0.0381 | 0.1025 | 0.0859 |
| 2009 | 1196590 | 1381580 | 1183320 | 183782 | 0.1553 | 0.0534 | 0.1438 | 0.1205 |


| Year | Recruit <br> (thousands) | Total <br> Biomass | Spawning <br> Biomass | Catch | Yield/SSB | Fbar(1-3) | Fbar(4-8) | Fbar(1-10) |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2010 | 893847 | 1282810 | 1039260 | 203112 | 0.1954 | 0.0656 | 0.1765 | 0.1480 |
| 2011 | 338534 | 1155520 | 928577 | 193698 | 0.2086 | 0.0692 | 0.1865 | 0.1563 |
| 2012 | 2277900 | 1022860 | 880027 | 169859 | 0.1930 | 0.0674 | 0.1814 | 0.1521 |
| 2013 | 982006 | 905486 | 803142 | 165258 | 0.2058 | 0.0745 | 0.2006 | 0.1681 |
| 2014 | 3140900 | 797965 | 679877 | 136360 | 0.2006 | 0.0706 | 0.1901 | 0.1593 |
| 2015 | 2138820 | 737343 | 576525 | 98419 | 0.1707 | 0.0560 | 0.1507 | 0.1263 |
| 2016 | 2419130 | 735636 | 541909 | 98810 | 0.1823 | 0.0581 | 0.1564 | 0.1310 |
| 2017 | 2846550 | 753910 | 527801 | 82961 | 0.1572 | 0.0477 | 0.1284 | 0.1076 |
| 2018 | 2329200 | 803522 | 568172 | 101682 | 0.1790 | 0.0554 | 0.1491 | 0.1249 |
| 2019 | 1260210 | 839611 | 604308 | 124947 | 0.2068 | 0.0652 | 0.1755 | 0.1471 |
| 2020 | 1165290 | 842494 | 625449 | 76422 | 0.1222 | 0.0376 | 0.1014 | 0.0850 |
| 2021 | 816224 | 693991 | 81557 | 0.1175 | 0.0377 | 0.1015 | 0.0851 |  |

Table 7.4.1. Western Horse Mackerel. Short term prediction: INPUT DATA. *geometric mean of the recruitment time series from 1983 to 2021. ** from assessment output

| Age | N | Mat | M | PF | PM | Stock weight at age** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 816224 | 0.000 | 0.150 | 0 | 0 | 0.0091 |
| 1 | 998917 | 0.000 | 0.150 | 0 | 0 | 0.0251 |
| 2 | 912502 | 0.050 | 0.150 | 0 | 0 | 0.0493 |
| 3 | 1384850 | 0.250 | 0.150 | 0 | 0 | 0.0798 |
| 4 | 1341580 | 0.700 | 0.150 | 0 | 0 | 0.1166 |
| 5 | 875566 | 0.950 | 0.150 | 0 | 0 | 0.1519 |
| 6 | 581151 | 1.000 | 0.150 | 0 | 0 | 0.1824 |
| 7 | 630308 | 1.000 | 0.150 | 0 | 0 | 0.2087 |
| 8 | 143506 | 1.000 | 0.150 | 0 | 0 | 0.2312 |
| 9 | 239066 | 1.000 | 0.150 | 0 | 0 | 0.2503 |
| 10 | 25200 | 1.000 | 0.150 | 0 | 0 | 0.2664 |
| 11 | 46793 | 1.000 | 0.150 | 0 | 0 | 0.2799 |
| 12 | 43929 | 1.000 | 0.150 | 0 | 0 | 0.2911 |


| Age | $\mathbf{N}$ | Mat | $\mathbf{M}$ | PF | PM | Stock weight at age** |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 13 | 121924 | 1.000 | 0.150 | 0 | 0 | 0.3004 |
| 14 | 38258 | 1.000 | 0.150 | 0 | 0 | 0.3080 |
| 15 | 15285 | 1.000 | 0.150 | 0 | 0 | 0.3142 |
| 16 | 13605 | 1.000 | 0.150 | 0 | 0 | 0.3193 |
| 17 | 1.000 | 0.150 | 0 | 0 | 0.3234 |  |
| 18 | 1.000 | 0.150 | 0 | 0 | 0.3268 |  |
| 19 | 8454 | 1.000 | 0.150 | 0 | 0 | 0.3294 |
| 20 | 62126 |  | 0.150 | 0 | 0 | 0.3334 |

## Table 7.4.2. Western Horse Mackerel. Short term prediction; single area management option table. Assumption: Catch 2022: $71 \mathbf{1 3 8} \mathbf{t}$ (100\% of 2022 TOTAL TAC).

| Scenarios | $F_{\text {factor }}$ | $\mathrm{F}_{\text {bar }}$ | Catch_2022 | Catch_2023 | SSB_2023 | SSB_2024 | Change_SSB_2023-2024(\%) | Change_Catch_2022-2023(\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathrm{SSB}_{2024}=$ MSY $\mathrm{B}_{\text {trigger }}=\mathrm{B}_{\mathrm{pa}}=\mathrm{Bl}_{\mathrm{lim}}$ | The $B_{p a}, B_{l i m}$ and MSY $B_{\text {trigger }}$ options were left blank because $B_{\text {pa }}, B_{\text {lim }}$ and $M S Y$ B ${ }_{\text {trigger }}$ cannot be achieved in 2024, even with a zero catch in 2023. |  |  |  |  |  |  |  |
| $F=F_{M S Y}$ | 0.870 | 0.074 | 71138 | 73950 | 754163 | 737593 | -2.2 | 3.95 |
| $F=F_{\text {P05 }}=F_{p a}$ | 0.929 | 0.079 | 71138 | 78719 | 754163 | 733196 | -2.8 | 10.7 |
| $F=F_{\text {lim }}$ | 1.211 | 0.103 | 71138 | 101225 | 754163 | 712461 | -5.5 | 42.3 |
| $F=0$ | 0 | 0 | 71138 | 0 | 754163 | 805946 | 6.9 | -100 |
| $F=F_{2022}$ | 0.844 | 0.072 | 71138 | 71813 | 754163 | 739564 | -1.94 | 0.95 |
| PelAC proposed HCR | 0.341 | 0.015 | 71138 | 15513 | 754163 | 791583 | 4.96 | -78 |

### 7.16 Figures



Figure 7.1.1.1: Western horse mackerel. Catch by quarter and year for 2000-2021.


Figure 7.1.2.1. Western horse mackerel. Catch categories since 2000 (green bars indicate when countries have submitted catch data without specifying landings/discards).


Figure 7.1.3.1: Western horse mackerel. Catch by ICES Division and year for 1982-2021.


Figure 7.2.1.1: Western Horse mackerel egg production by half rectangle for period 3 (March 4th - April $\mathbf{8}^{\text {th }}, \mathbf{2 0 2 2}$ ). Circle areas and colour scale represent horse mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 7.2.1.2: Western Horse mackerel egg production by half rectangle for period 4 (April 9th - 29 ${ }^{\text {th }}, 2022$ ). Circle areas and colour scale represent horse mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 7.2.1.3: Western Horse mackerel egg production by half rectangle for period 5 (Apr 30th - May 31 ${ }^{\text {st }}$, 2022). Circle areas and colour scale represent horse mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 7.2.1.4: Western Horse mackerel egg production by half rectangle for period 6 (June 1st $\mathbf{- 3 0}{ }^{\text {th }}, \mathbf{2 0 2 2}$ ). Circle areas and colour scale represent horse mackerel stage l eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 7.2.1.5: Western mackerel egg production by half rectangle for period 7 (July 1st - July 31 ${ }^{\text {st, }}$ 2022). Circle areas and colour scale represent horse mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values


Figure 7.2.1.6: Provisional annual egg production curve for western horse mackerel for 2022, (black line). The curves for 2010, 2013, 2016 and 2019 are included for comparison


Figure 7.2.1.7. Provisional total annual egg production for western horse mackerel. Production figures back to 1992 are included for comparison.


Figure 7.2.2.1: Western horse mackerel. Trend of the fisheries independent indices of abundance used in the assessment of Western Horse mackerel. Top: Spawning index from egg survey; middle: recruitment index from IBTS survey; bottom: biomass estimates from PELACUS acoustic survey. Confidence intervals are shown as well.

2021 Western Stock: cat@ge by division


Figure 7.2.4.1: Western horse mackerel. Catch-at-age (millions) by ICES division in 2021.

Western Stock: cat@ge by Year


Figure 7.2.4.2: Western horse mackerel. Catch-at-age (millions) by Year.


Figure 7.2.4.3: Western horse mackerel. Catch-at-age - the area of bubbles is proportional to the catch number. Age 15 is a plus group.

Weight at age - 1st \& 2nd quarter

$$
\begin{aligned}
& -1-4-7-10-13 \\
& -2-5-8-11-14 \\
& -3-6-9-12-15
\end{aligned}
$$



Figure 7.2.5.1: Western horse mackerel. Weight at age in the catch (kg) by year.


Figure 7.2.5.2: Western horse mackerel. Weight at length in the stock (kg) as estimated by the stock assessment.


Figure 7.2.6.1: Western horse mackerel. Maturity at age as used in the assessment model.


Figure 7.2.10.1: Western horse mackerel. Length frequency distribution of the landing data as used in the assessment model.


Figure 7.2.10.2: Western horse mackerel. Stacked length frequency distribution of the landing data as used in the assessment model.


Figure 7.2.10.3: Western horse mackerel. Within-cohort consistency in the catch-at-age matrix, shown by plotting the log-catch of a cohort at a particular age against the log-catch of the same cohort at subsequent ages.


Figure 7.2.10.4: Western horse mackerel. Catch numbers at age composition by decade (year specifies start of decade i.e., $1980=1980-1999$, also note that 2020 only includes years 2020-2021).


Figure 7.2.10.5: Western horse mackerel. Data exploration. Correlation plots between indices of abundance (including 2021 data points). Size and shade of circle indicates magnitude of correlation and color indicated sign (blue positive, red negative).


Figure 7.2.11.1: Western horse mackerel. Model fitting. Fitting of the model to the fisheries-independent indices. From top to bottom: IBTS, egg survey, PELACUS. Dots represent observations (with confidence intervals) and blue line the model.


Figure 7.2.11.1 cont.: Western horse mackerel. Model fitting. Fitting of the model to the catch at age matrix from 1982 to 2002. Black joined dots represent observations and green line represents model.


Figure 7.2.11.1 cont.: Western horse mackerel. Model fitting. Fitting of the model to the length composition of the landing data from 2000 to 2021.


Figure 7.2.11.1 cont.: Western horse mackerel. Model fitting. Fitting of the model to the length composition of the acoustic survey.



Age (yr)

Figure 7.2.11.1 cont.: Western horse mackerel. Model fitting. Fitting of the model to the Age length comp of the catch.


Figure 7.2.11.2: Western horse mackerel. Model results. Spawning stock biomass ( 0.5 of the overall SSB only is shown; plot on the top) and recruitment estimates (plot on the bottom) from the assessment model from 1982 to $2022.95 \% \mathrm{Cl}$ are shown. Note this figure is a standard SS output. Whilst the $y$-axis denotes spawning biomass in mt , the axis values reflect the actual (data) values. Therefore, the axis values should be between 0 and 6 to correspond to the axis title.


Figure 7.2.11.2 cont.: Western horse mackerel. Model results. Fishing mortality estimates (Fbar ages 1-10) from the assessment model from 1982 to $\mathbf{2 0 2 1}$. $95 \% \mathrm{Cl}$ are shown.


Figure 7.2.11.3: Western horse mackerel. 5 years of retrospective analysis for SSB, F and Recruitment. Dash lines are the 2022 assessment confidence intervals.


Figure 7.2.11.4: Western horse mackerel. Historical model assessment results. Note: since the 2017 assessment, SSB is estimated on 1st of January. Prior to 2017 SSB has been estimated in May (spawning time).


Figure 7.10.1. Western horse mackerel. Top: comparison of (max) scientific advice, TAC (or sum of unilateral quota) and Total Catch. Bottom: percentage deviation from ICES advice, CoA is Catch over Advice, ToA is TAC over Advice.


Figure 7.13.1. Sensitivity of the model to the PELACUS data. Spawning biomass and fishing mortality (ages 1-10) as estimated in the model conducted in 2020 (in blue) and in a model with the same setup but excluding the PELACUS data for 2019 (in red).

## 8 Northeast Atlantic Mackerel

### 8.1 ICES Advice and International Management Applicable to 2021

From 2001 to 2007, the internationally agreed TACs covered most of the distribution area of the Northeast Atlantic mackerel. From 2008 to 2014, no agreement was reached among the Coastal States on the sharing of the mackerel quotas. In 2014, three of the Coastal States (European Union, Norway and the Faroe Islands) agreed on a Management Strategy for 2014 to 2018. In November 2018, the 2014 agreement was extended for a further two years until 2020. No agreement on the share of the stock has been reached for 2021 and 2022. Despite various agreements, the total declared quotas in each of the years 2015 to 2021 all exceeded the TAC advised by ICES. An overview of declared quotas and transfers for 2022, as available to WGWIDE, is given in the text table below. An estimate of the expected quota uptake in 2022 was carried out based on the fishery up to the end of August 2022 and knowledge of the fishery development from September to December 2021. Total removals of mackerel are expected to be approximately 1.1 million tonnes in 2022, exceeding the ICES advice for 2022 by approximately 336500 tonnes (30\%).

The quota figures and transfers in the text table below were based on national regulations, official press releases, and discard estimates.

Various international and national measures to protect mackerel are in operation throughout the mackerel catching countries. Refer to the stock annex for an overview.

| Estimation of 2022 catch <br> (t) | Unilateral <br> quotas | Reference | Expected <br> uptake | Justification |
| :--- | :--- | :--- | :--- | :--- |
| EU quota | 183359 | Record of fisheries con- <br> sultations between the <br> United Kingdom and the <br> European Union for 2022 | 183359 | Full uptake |
| UK quota | 210820 | Footnote ${ }^{1}$ | 210820 | Full uptake |
| Norwegian quota | 278222 | NEAFC HOD 04/2022 | 278222 | Full uptake |
| Inter-annual quota trans- <br> fer to 2022 | 23763 | NEAFC HOD 04/2022 | 23763 | Full uptake |
| Russian expected catches <br> 2022 | - | Last year's quota | 112319 | Russian expected catch ${ }^{3}$ |
| Icelandic quota | Footnote ${ }^{2}$ Footnote ${ }^{2}$ | Icelandic expected catch ${ }^{4}$ |  |  |
| Inter-annual quota trans- <br> fer to 2022 plus special <br> quota | 28026 | 20905 | Faroese Fisheries Minis- <br> try regulations No. <br> 182/2022 | 0 |


| Estimation of 2022 catch (t) | Unilateral quotas | Reference | Expected uptake | Justification |
| :---: | :---: | :---: | :---: | :---: |
| Greenland quota | 51670 | Ministry of Fisheries, Hunting and Agriculture, Greenland | 34000 | Greenland expected catch ${ }^{4}$ |
| Discards | 3129 | Previous years estimate | 3129 | Previous years estimate |
| Total expected catch (incl. discards) ${ }^{5,6}$ |  |  | 1131416 | tonnes |
| ${ }^{1}$ https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attachment_data/file/1059641/Outcomes_of_annual _negotiations for_UK fishing_opportunities_in_2021 and_2022.pdf |  |  |  |  |
| ${ }^{2}$ https://www.fiskistofa.is/veidar/aflaupplysingar/heildaraflamarksstada/?timabil=2122\&fyrirsp=4\&lang=is\&landhelgi=U |  |  |  |  |
| ${ }^{3}$ The Russian estimated uptake in 2022 was set as the Russian quota in 2021 scaled by the reduction of the ICES advised TAC from 2021 to 2022 (-6.7\%). |  |  |  |  |
| ${ }^{4}$ The estimated catch of mackerel in 2022 was obtained by summing recorded catches as of August 25 th 2022 and the catches from that date to end of the fishing season 2021, assuming a similar development in the fishery for the remainder of the 2022 season as in 2021. |  |  |  |  |
| ${ }^{5}$ Quotas refer to claims by each party for 2022 and include exchange to other parties. |  |  |  |  |

### 8.2 The Fishery

### 8.2.1 Fleet Composition in 2021

The total fleet can be considered to consist of the following components: freezer trawlers, purse seiners, pelagic trawlers, lines and jigging, and gillnets (see stock annex for detailed description of each component).

### 8.2.2 Fleet Behaviour in 2021

The northern summer fishery in Subarea 2 continued and increased significantly in 2021 with Norway reporting $80 \%$ of their catches in Division 2.a quarter 3. There was no fishery in Subarea 14 and a reduced fishery in Subarea 5. The Russian freezer trawler fleet operates over a wide area in northern international waters. This fleet targets herring and blue whiting in addition to mackerel. In 2021 the Russian vessels took the vast majority of their catch in Division 2.a.

Total catches from Icelandic vessels were similar to those in recent years and were in excess of 100 kt . The majority of the catch was taken in Division 2.a in quarter 3 of 2021, with very small catch also taken in Divisions 5.a and 5.b. In 2021 Greenland targeted mackerel in Division 2.a with no catch taken from Division 14.b. In 2019 Greenland fished in Division 14.b and in 2018 both Greenland and Iceland reported landings from this area. Catches from Greenland have increased in 2021 to 33 kt up from 27 kt in 2021 but lower than the peak of 63 kt in 2018. The Faroese fleet targeted mackerel during quarters 2,3 and 4 with $96 \%$ of the catches taken in Division 2.a. The remaining catch was taken in Division 5.b in quarter 2. No catch was reported from Divisions 4.a or 6.a as had been in previous years.

Fishing in the North Sea and west of the British Isles followed a traditional pattern, targeting mackerel on their spawning migration from the Norwegian deep in the northern North Sea, westwards around the north coast of Scotland and down the west coast of Scotland and Ireland. The majority of the Irish mackerel fishery took place in quarter 1 along the west coast of Scotland and Ireland, with the Scottish fleet operating in the same area at this time. The Scottish fishery in quarter 4 was more concentrated in the North Sea where $51 \%$ of the Scottish catch was taken.

In 2021 the Spanish fishery started at the beginning of March, as in previous years.

### 8.2.3 Recent Changes in Fishing Technology and Fishing Patterns

Northeast Atlantic mackerel, as a widely distributed species, is targeted by a number of different fishing métiers. Most of the fishing patterns of these métiers have remained unchanged during the most recent years (see stock annex), although the variation in timing of the spawning migration and geographical distribution can change from year to year and this affects the fishery in various areas. The most important changes in recent years are related to the geographical expansion of the northern summer fishery (Subareas 2,5 and 14) and changes in southern waters due to stricter TAC compliance by Spanish authorities. In 2020 and 2021 the northern summer fishery did not extend as far west as in previous years. In 2021 the summer fishery in Division 2.a increased substantially. The annual fishing pattern by statistical rectangle from 2003-2021 is shown in Section 1.8.

### 8.2.4 Recent regulations and management

Currently there is no agreement on a management strategy covering all parties fishing mackerel.
An overview of the technical measures, effort controls and management plans are given in the stock annex. Note that there may be additional existing international and national regulations that are not listed.

### 8.3 Quality and Adequacy of sampling Data from Commercial Fishery

The sampling of the commercial catch of Northeast Atlantic mackerel is summarised below:
$\left.\begin{array}{lllllll}\hline \text { Year } & \begin{array}{l}\text { WG Total Catch } \\ \text { (t) }\end{array} & \text { \% catch covered by sampling programme* } & \text { No. } & \text { No. } \\ \text { Samples }\end{array}\right)$

| Year | WG Total Catch <br> (t) | \% catch covered by sampling programme* | No. Samples | No. <br> Measured | No. <br> Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 617330 | 80 | 1212 | 148501 | 19779 |
| 2004 | 611461 | 79 | 1380 | 177812 | 24173 |
| 2005 | 543486 | 83 | 1229 | 164593 | 20217 |
| 2006 | 472652 | 85 | 1604 | 183767 | 23467 |
| 2007 | 579379 | 87 | 1267 | 139789 | 21791 |
| 2008 | 611063 | 88 | 1234 | 141425 | 24350 |
| 2009 | 734889 | 87 | 1231 | 139867 | 28722 |
| 2010 | 877272 | 91 | 1241 | 124695 | 29462 |
| 2011 | 948963 | 88 | 923 | 97818 | 22817 |
| 2012 | 899551 | 89 | 1216 | 135610 | 38365 |
| 2013 | 938299 | 89 | 1092 | 115870 | 25178 |
| 2014 | 1401788 | 90 | 1506 | 117250 | 43475 |
| 2015 | 1215827 | 88 | 2132 | 137871 | 24283 |
| 2016 | 1100135 | 89 | 2200 | 149216 | 21456 |
| 2017 | 1159641 | 87 | 2183 | 151548 | 24104 |
| 2018 | 1023144 | 83 | 1858 | 139590 | 20703 |
| 2019 | 839727 | 88 | 1835 | 141561 | 17646 |
| 2020 | 1039513 | 87 | 1430 | 142991 | 15685 |
| 2021 | 1081540 | 79 | 1783 | 76325 | 18736 |

Overall sampling effort in 2021 was lower than previous years with $79 \%$ of the catch sampled. It should be noted that this proportion is based on the total sampled catch and in 2021 there was no sampling reported from Russia. Nations with large, directed fisheries are capable of sampling $100 \%$ of their catch which may conceal deficiencies in sampling elsewhere.

The 2021 sampling levels by country are shown below.

| Country | Official <br> catch | \% WG catch covered by sampling <br> programme | No. Sam- <br> ples | No. Meas- <br> ured | No. Aged |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium | 110 | $0 \%$ | 5 | 111 | 111 |
| Denmark | 32813 | $93 \%$ | 20 | 1001 | 960 |
| Faroe Islands | 105096 | $98 \%$ |  |  |  |


| Country | Official catch | \% WG catch covered by sampling programme | No. Samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| France | 16686 | 0\% |  |  |  |
| Germany | 11996 | 92\% | 34 | 6389 | 310 |
| Greenland | 33360 | 100\% | 23 | 1768 | 138 |
| Iceland | 132109 | 96\% | 108 | 2823 | 4810 |
| Ireland | 60795 | 96\% | 51 | 8971 | 1899 |
| Lithuania | 6655 | 0\% |  |  |  |
| Netherlands | 24594 | 92\% | 34 | 2450 | 844 |
| Norway | 270653 | 96\% | 92 | 2423 | 2336 |
| Poland | 1779 | 0\% |  |  |  |
| Portugal | 4723 | 20\% | 77 | 2287 | 897 |
| Russia | 136176 | 0\% |  |  |  |
| Spain | 30085 | 99\% | 1134 | 32868 | 3207 |
| Sweden | 3514 | 0\% |  |  |  |
| UK (England \& Wales) | 22094 | 54\% | 154 | 9533 | 1747 |
| UK (Northern Ireland) | 16464 | 50\% | 1 | 164 | 50 |
| UK (Scotland) | 171840 | 92\% | 50 | 5537 | 1427 |

The majority of countries achieved a high level of sampling coverage. Belgian catches consist of by-catch in the demersal fisheries in the North Sea. France supplied a quantity of length-frequency data to the working group which can be utilised to characterise the selection of the fleet but requires an allocation of catch at age proportions from another sampled fleet in order to raise the data for use in the assessment. Russia, Sweden, Lithuania and Poland did not supply sampling information in 2021. Portugal sampled landings from 9.a only. England sampled landings from the handline fleet operating off the Cornish coast as well as from freezer trawlers. Cooperation between the Dutch and German sampling programmes is designed to provide complete coverage for the freezer trawlers operating under these national flags and also those of England and France. Catch sampling levels per ICES Division (for those with a WG catch of $>100 \mathrm{t}$ ) are shown below.

| Division | Official Catch (t) | WG Catch (t) | No. Samples | No. Measured | No Aged |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 0.03 | 0.03 | 0 | 0 | 0 |
| 2.a | 657905 | 657905 | 238 | 7828 | 8095 |
| $3 . \mathrm{a}$ | 439 | 439 | 0 | 0 | 0 |
| $3 . \mathrm{b}$ | 19 | 0 | 0 | 0 |  |


| Division | Official Catch (t) | WG Catch (t) | No. Samples | No. Measured | No Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $3 . \mathrm{c}$ | 24 | 24 | 0 | 0 | 0 |
| 3.d | 59 | 59 | 0 | 0 | 0 |
| 4.a | 216985 | 216985 | 96 | 12866 | 2244 |
| 4.b | 2837 | 2837 | 0 | 0 | 0 |
| 4.c | 1078 | 1078 | 1 | 92 | 25 |
| 5.a | 933 | 933 | 1 | 34 | 25 |
| 5.b | 4273 | 4273 | 2 | 102 | 75 |
| 6.a | 146583 | 146583 | 75 | 10372 | 2250 |
| 6.b | 52 | 52 | 0 | 0 | 0 |
| 7.a | 11 | 11 | 0 | 0 | 0 |
| 7.b | 2778 | 2778 | 3 | 366 | 102 |
| 7.c | 124 | 124 | 1 | 46 | 25 |
| 7.d | 6180 | 6180 | 1 | 59 | 25 |
| $7 . \mathrm{e}$ | 808 | 808 | 71 | 3747 | 1466 |
| 7.f | 258 | 258 | 75 | 5016 | 82 |
| 7.9 | 12 | 12 | 0 | 0 | 0 |
| 7.h | 38 | 38 | 1 | 111 | 25 |
| 7.j | 2740 | 2740 | 6 | 476 | 168 |
| 8.a | 1096 | 1096 | 1 | 54 | 1 |
| 8.b | 4338 | 4338 | 196 | 3950 | 66 |
| 8.c | 29120 | 29120 | 569 | 22553 | 2504 |
| 8.d | 40 | 40 | 0 | 0 | 0 |
| $8 . \mathrm{e}$ | 0.0096 | 0.0096 | 0 | 0 | 0 |
| $9 . \mathrm{a}$ | 2807 | 2807 | 446 | 8653 | 1558 |

In general, areas with insufficient sampling have relatively low levels of catch.

### 8.4 Catch Data

### 8.4.1 ICES Catch Estimates

## Missing 2021 data

In 2022, WGWIDE did not receive a data submission from Russia with the 2021 catch and sampling information. Preliminary catch data were available for 2021 from Russia and were reported by ICES division by year. From 2018-2020 the Russian catch accounted for an average of $13 \%$ of the total working group catch. The preliminary figure for 2021 also represents $13 \%$ of the total working group catch. Historically, preliminary catches are comparable to ICES final estimated catch.

The majority of the Russian catch was from ICES division 2.a and a three year average (20182020) was used to distribute the data by quarter. This resulted in the data assigned $7 \%$ in quarter $2,92 \%$ in quarter 3 and $1 \%$ in quarter 4. The remaining data was from Division $5 . \mathrm{b}$ and was all assumed to be taken in quarter 3 as has been the case in previous years by the Russian fleet.

Catch maps were produced by country to determine the fishing pattern particularly in Division 2.a quarter 3. The Russian fishery is concentrated mainly in Division 2.a with an average of $90 \%$ of the catch from Division 2.a. 1 and a consistent distribution of the fishery in the last three years. Iceland and the Faroes also fish in a similar area to the Russian fleet and use similar gear. Comparisons of the age and length data from Division 2.a quarter 3 were presented and a decision made to use samples from Iceland and the Faroes to allocate to the Russian catches.

## Total Catch 2021

The total ICES estimated catch for 2021 was 1081540 tonnes, an increase from 1039513 tonnes in 2020.

The combined 2021 TAC, arising from agreements and autonomous quotas, amounts to 1199 103 tonnes. The ICES catch estimate ( 1081540 tonnes) represents an undershoot of this but is still above the ICES advice of 852284 tonnes for 2021. The combined fishable TAC for 2022, as best ascertained by the Working Group (see Section 8.1), amounts to 1131416 tonnes.

Catches reported for 2021 and in previous Working Group reports are considered to be best estimates. In most cases, catch information comes from official logbook records. Other sources of information include catch processors. Some countries provide information on discards and slipped catch from observer programs and compliance reports. In several countries discarding is illegal. Spanish data is based on the official data supplied by the Fisheries General Secretary (SGP) but supplemented by scientific estimates which are recorded as unallocated catch in the ICES estimates. A detailed basis for the ICES catch estimates is presented in the stock annex.

The total catch as estimated by ICES is shown in Table 8.4.1.1. It is broken down by ICES area group and illustrates the development of the fishery since 1969.

## Discard Estimates

With a few exceptions, estimates of discards have been provided to the Working Group for the ICES Subareas and Divisions 6, 7/8.a,b,d,e and 3/4 (see Table 8.4.1.1) since 1978. Historical discard estimates were revised during the data compilation exercise undertaken for the 2014 benchmark assessment (ICES, 2014). The Working Group considers that the estimates for these areas are incomplete. In 2021, discard data for mackerel were provided by France, Ireland, Spain, Denmark, England, Scotland and Sweden. Total discards amounted to 3129 tonnes which is a decrease from 2020. The German, Dutch and Portuguese pelagic discard monitoring programmes
did not record any instances of discarding of mackerel. Estimates from the other countries supplying data include results from the sampling of demersal fleets.

### 8.4.2 Distribution of Catches

A significant change in the fishery took place between 2007 and 2009 with a greatly expanded northern fishery becoming established. This fishery has continued to the present but with a clear tendency for an eastern retraction, especially from the Greenlandic area and also western parts of the Icelandic area in the most recent three years. In 2021 there was only a small amount of catch from southern Iceland with the fishery moving further east. Of the total catch in 2021, Norway accounted for the greatest proportion (25\%) followed by Scotland (16\%), Russia (13\%), Iceland ( $12 \%$ ), Faroes ( $10 \%$ ) and Ireland ( $6 \%$ ). In the absence of sharing arrangements, the fishing parties declared unilateral quotas for 2021.

In 2021, catches in the northern areas (Subareas 1, 2, 5, 14) increased significantly and amounted to 663111 tonnes (see Table 8.4.2.1), an increase of 306126 tonnes on the 2020 catch. Norwegian catches were over 270 kt and Icelandic, Russian, Scottish and Faroese catches were all over 100 kt . Catches from Division 2.a accounted for $61 \%$ of the total catch in 2021. The wide geographical distribution of the fishery noted in previous years has continued.

The time series of catches by country from the North Sea, Skagerrak and Kattegat (Subarea 4, Division 3.a) is given in Table 8.4.2.2. Catches in 2021 decreased to 221340 tonnes from 457211 tonnes in 2020. The majority of the catch is from Subarea 4 with small catches were also reported in Divisions 3.a-d.

Catches in the western area (Subareas 6, 7 and Divisions 8.a, b, d and e) decreased in 2021 to 165 060 tonnes. This is a decrease of around 22000 t from 2020. The catches are detailed in Table 8.4.2.3.

Table 8.4.2.4 details the catches in the southern areas (Divisions 8.c and 9.a) which are taken almost exclusively by Spain and Portugal. The reported catch in 2021 of 31928 tonnes represents a decrease of almost 5600 tonnes from 2020. The catch is above the long-term average.

The distribution of catches by quarter (\%) is described in the text table below:

| Year | Q1 | Q2 | Q3 | Q4 |
| :---: | :---: | :---: | :---: | :---: |
| 1990 | 28 | 6 | 26 | 40 |
| 1991 | 38 | 5 | 25 | 32 |
| 1992 | 34 | 5 | 24 | 37 |
| 1993 | 29 | 7 | 25 | 39 |
| 1994 | 32 | 6 | 28 | 34 |
| 1995 | 37 | 8 | 27 | 28 |
| 1996 | 37 | 8 | 32 | 23 |
| 1997 | 34 | 11 | 33 | 22 |
| 1998 | 38 | 12 | 24 | 27 |
| 1999 | 36 | 9 | 28 | 27 |


| Year | Q1 | Q2 | Q3 | Q4 |
| :---: | :---: | :---: | :---: | :---: |
| 2000 | 41 | 4 | 21 | 33 |
| 2001 | 40 | 6 | 23 | 30 |
| 2002 | 37 | 5 | 29 | 28 |
| 2003 | 36 | 5 | 22 | 37 |
| 2004 | 37 | 6 | 28 | 29 |
| 2005 | 46 | 6 | 25 | 23 |
| 2006 | 41 | 5 | 18 | 36 |
| 2007 | 34 | 5 | 21 | 40 |
| 2008 | 34 | 4 | 35 | 27 |
| 2009 | 38 | 11 | 31 | 20 |
| 2010 | 26 | 5 | 54 | 15 |
| 2011 | 22 | 7 | 54 | 17 |
| 2012 | 22 | 6 | 48 | 24 |
| 2013 | 19 | 5 | 52 | 24 |
| 2014 | 20 | 4 | 46 | 30 |
| 2015 | 20 | 5 | 44 | 31 |
| 2016 | 23 | 4 | 44 | 29 |
| 2017 | 24 | 3 | 45 | 28 |
| 2018 | 20 | 3 | 40 | 37 |
| 2019 | 28 | 5 | 42 | 26 |
| 2020 | 31 | 4 | 34 | 31 |
| 2021 | 19 | 5 | 56 | 20 |

The quarterly distribution of catch from 2010-2020 is similar to recent years with the northern summer fishery in Q3 accounting for the greatest proportion of the total catch. The average proportion taken in quarter 3 from 2010-2020 is $46 \%$. In 2021 this proportion increased to $56 \%$ and is higher than the quarter 1 and quarter 4 catches which when combined account for $39 \%$ of the total. The proportion of the catch taken in quarter 2 has remained stable.
Catches per ICES statistical rectangle are shown in Figures 8.4.2.1 to 8.4.2.4. It should be noted that these figures are a combination of official catches and ICES estimates and may not indicate the true location of the catches or represent the location of the entire stock. These data are based on catches reported by all the major catching nations and represents almost the entire ICES estimated catch except Russia.

- First quarter 2021 (208 190 tonnes - 19\%)

The distribution of catches in the first quarter is shown in Figure 8.4.2.1. The proportion of the fishery taken in quarter 1 has decreased in 2021 with the Scottish and Irish pelagic fleets targeting mackerel in Divisions 6.a, 7.b and 7.j. Substantial catches are also taken by the Dutch owned freezer trawler fleet. The largest catches were taken in Division 6.a, as in recent years. The Spanish fisheries also take significant catches along the north coast of Spain during the first quarter.

- $\quad$ Second quarter 2021 (55 707 tonnes - 5\%)

The distribution of catches in the second quarter is shown in Figure 8.4.2.2. The quarter 2 fishery is traditionally the smallest and this was also the case in 2021. The most significant catches where those in Division 8.c and at the start of the summer fishery in northern waters by Icelandic, Norwegian and Russian fleets in Division 2.a.

- Third quarter 2021 (599 548 tonnes - 56\%)

Figure 8.4.2.3 shows the distribution of the quarter 3 catches. Large catches were taken throughout Division 2.a, with high concentrations in international waters. Fishing was carried out mainly by vessels from Russia, Norway, Iceland, the Faroes and Greenland. There were also catches from Division 4.a but very little from Division 5.a.

- Fourth quarter 2021 (218 186 tonnes - 20\%)

The fourth quarter distribution of catches is shown in Figure 8.4.2.4. The proportion of the catch taken in the fourth quarter has decreased from $31 \%$ in 2020 to $20 \%$ in 2021. The summer fishery in northern waters has largely finished with some catches reported from Division 2.a. The largest catches in quarter 4 are taken by Scotland around the Shetland Isles in Division 4.a.

### 8.4.3 Catch-at-Age

This catch in numbers relates to a total ICES estimated catch of 1081540 tonnes. These figures have been appended to the catch-at-age assessment table (see Table 8.7.1.2).

Age distributions of commercial catch were provided by Denmark, England, Germany, Faroes, Iceland, Ireland, the Netherlands, Norway, Portugal, Scotland, Northern Ireland and Spain. There remain gaps in the age sampling of catches, notably from France (length samples were provided), Russia, Sweden, Lithuania and Poland.
Catches for which there were no sampling data were converted into numbers-at-age using data from the most appropriate fleets. Accurate national fleet descriptions are required for the allocation of sample data to unsampled catches.

The catch numbers at age show a number of strong year classes in this fishery. Over $85 \%$ of the catch in numbers in 2021 consists of 2 to 10-year olds with the 2016-year class being the strongest. The 2016 year-class was strong in the fishery in previous years and accounted for $14 \%$ of the catch numbers at age in 2021. The 2019-year class, which are now 2 years old accounts for $13 \%$ of the catch numbers at age and were not evident in the fishery before. The 2015 year-class does not look as strong as the other year classes and represents $7 \%$ of the total. In 2021 there is a decrease in the proportion of fish in the plus group from $7 \%$ in 2020 to $5 \%$ in 2021. Year classes from 2009 and earlier are now in the plus group.

There is a small presence of juvenile (age 0) fish within the 2021 catch. As in previous years catches from Divisions 8.c and 9.a have contained a proportion of juveniles.

### 8.5 Biological Data

### 8.5.1 Length Composition of Catch

The mean length-at-age in the catch for 2021 is given in Table 8.5.1.1.
For the most common ages which are well sampled there is little difference to recent years. The length of juveniles is traditionally rather variable. The range of lengths recorded in 2021 for 0 group mackerel is $180 \mathrm{~mm}-222 \mathrm{~mm}$. The rapid growth of 0 -group fish combined with variations in sampling between northern and southern areas will contribute to the observed variability in the observed size of 0-group fish. Growth is also affected by fish density as indicated by a recent study which demonstrated a link between growth of juveniles and adults ( $0-4$ years) and the abundance of juveniles and adults (Jansen and Burns, 2015). A similar result was obtained for mature 3- to 8-year-old mackerel where a study over 1988-2014 showed declining growth rate since the mid-2000s to 2014, which was negatively related to both mackerel stock size and the stock size of Norwegian spring spawning herring (Ólafsdóttir et al., 2015).

### 8.5.2 Weights at Age in the Catch and Stock

The mean weight-at-age in the catch for 2021 are given in Table 8.7.1.3. There is a trend towards lighter weight-at-age for the most age classes (except 0 to 2 years old) starting around 2005, continuing until 2013 (Figure 8.5.2.1). This decrease in the catch mean weight-at-age seems to have stopped since 2013 and values for the last seven years do not show any particular trend for the older ages (age 6 and older) and are slightly increasing for younger ages (ages 1 to 5). These variations in weight-at-age are consistent with the changes noted in length in Section 8.5.1.

The Working Group used weight-at-age in the stock calculated as the average of the weight-atage in the three spawning components, weighted by the relative size of each component (as estimated by the 2022 egg survey for the southern and western components and the 2017 egg survey for the North Sea component). Mean weight-at-age in 2021 for the western component are estimated from Dutch, Irish and German commercial catch data, the biological sampling data taken during the egg surveys and during the Norwegian tagging survey. Only samples corresponding to mature fish, from areas and periods corresponding to spawning, as defined at the 2014 benchmark assessment (ICES, 2014) and laid out in the Stock Annex, were used to compute the mean weight-at-age in the western spawning component. For the North Sea spawning component, mean weight-at-age in 2021 were calculated from samples of the commercial catches collected from Divisions $4 . a$ and $4 . b$ in the second quarter of 2021 and samples collected during the 2021 egg survey conducted in the North Sea. Stock weights for the southern component, are based on samples from the Spanish catches and surveys in Divisions 8.c and 9a in the $2^{\text {nd }}$ quarter of the year. The mean weights in the three component and in the stock in 2021 are shown in the text table below.

As for the stock weights, the decreasing trend observed since 2005 for fish of age 3 and older seems to have stopped in 2013 and values in the last 8 years show an increasing trend (except for weights of ages 0 and 1 which have been stable, Figure 8.5.2.2).
\(\left.$$
\begin{array}{lllll}\hline & \text { North Sea Component } & \begin{array}{l}\text { Western } \\
\text { Component }\end{array}
$$ \& Southern Component \& NEA Mackerel <br>

Age \& \& \& 2021\end{array}\right]\)| Weighted mean* |
| :--- |
| 0 |
| 0.132 |
| 0.224 |

* Missing value of mean weight-at-age per component are replaced by component mean value in the calculation of the stock weights


### 8.5.3 Natural Mortality and Maturity Ogive

Natural mortality is assumed to be 0.15 for all age groups and constant over time.
The maturity ogive for 2021 was calculated as the average of the ogives of the three spawning components weighted by the relative size of each component calculated as described above for the stock weights. The ogives for the North Sea and Southern components are fixed over time. For the Western component the ogive is updated every year, using maturity data from commercial catch samples from Germany, Ireland, the Netherlands and the UK collected during the first and second quarters (ICES, 2014 and Stock Annex). The 2021 maturity ogives for the three components and for the mackerel stock are shown in the text table below.
\(\left.$$
\begin{array}{lllll}\hline \text { Age } & \begin{array}{l}\text { North Sea } \\
\text { Component }\end{array} & \begin{array}{l}\text { Western } \\
\text { Component }\end{array}
$$ \& \begin{array}{l}Southern <br>

Component\end{array} \& Mackerel\end{array}\right]\)| 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- |
| 1 | 0 | 0.15 | 0.02 |
| 2 | 0.37 | 0.59 | 0.54 |
| 4 | 1 | 1 | 1 |

A trend towards earlier maturation (increasing proportion mature at age 2) has been observed from around 2008 to 2015 . A change in the opposite direction has been observed since then and the proportion of fish mature at age since 2018 are now markedly lower than in the previous years and at levels comparable with the ones observed at the end of the 2000s (Figure 8.5.3.1).

### 8.6 Fishery Independent Data

### 8.6.1 International Mackerel Egg Survey

The ICES Triennial Mackerel and Horse Mackerel Egg Survey 2022 was carried out during January - July. The results have been used in the assessment for mackerel since 1977. Since 2004 and subsequent to demands for up-to-date data for the assessment, WGMEGS aims to provide a preliminary estimate of NEA mackerel biomass and western horse mackerel egg production in time for the assessment meetings within the same calendar year as the survey.

WGMEGS provided the preliminary results of the 2022 mackerel and horse mackerel egg survey for WGWIDE in August 2022. The final survey results will be available following the next WGMEGS meeting (April 2023). This is due to the extremely large numbers of plankton and fecundity samples to be analysed following the surveys as well as the tight deadline set by WGWIDE for delivering these estimates. A Working Document (WD07: O'Hea et al. 2022) with the preliminary results of the 2022 survey was presented to the WGWIDE 2022 meeting.

The 2022 survey was split into 6 separate sampling periods. Maximum deployment of effort in the Western area was during periods 3-6. Historically these periods would have coincided with
the expected peak spawning of both mackerel and horse mackerel. Recent years have seen mackerel peak spawning taking place during periods 3 and 5 . Due to the expansion of the spawning area which has been observed since 2007, survey effort allocation focuses on achieving full area coverage and delineation of the spawning boundaries.

Analyses of the plankton and fecundity samples were carried out according to the sampling protocols as described in the WGMEGS Survey Manual (ICES, 2019a) \& Fecundity manual (ICES, 2019b).

### 8.6.1.1 Data analysis for mackerel annual egg production

Stage 1 egg counts were converted to daily egg production using the development equations given in the survey manual. Procedures to estimate the total annual egg production are described in the WGMEGS Survey Manual (ICES, 2019a). Plots of the distribution of egg production for the western area are presented in Figures 8.6.1.1.1-8.6.1.1.5. The area coverage is described in detail in WD07 (O'Hea et al. 2019).
Figure 8.6.1.1.6 presents the egg production curve for the western area for the 2022 survey, along with those for the previous surveys for comparison. 2010 provided an unusually large spawning event early in the spawning season, 2013 yielded an even larger spawning event indicating that spawning was probably taking place well before the nominal start date of 10th February (day 42). In 2016 the first survey commenced on February 5th which is five days prior to the nominal start date. 2019 followed that of 2016 with no early peak spawning being recorded. The 2022 egg production curve is very similar to that of 2016, with peak spawning again occurring during Period 5. The expansion observed in western and northwestern areas during Periods 5 and 6 in 2016 was once again reported during 2022. During these periods it was not possible to fully delineate the northern and north-western boundaries, however the production in Periods 5 and 6 in the current survey year was lower in these northwestern areas. (Figures 8.6.1.1.4-5)=

Due to the cancelation of the Irish survey in June the area from 53 N to 61 N , and 3.5 W to 21 W could not be covered in period 6. In order to estimate the egg production in this uncovered area WGMEGS estimated the spawning area that was missed and also estimated mean daily egg production for the period. Positive stations (spawning area) were selected where stage 1 eggs were found in a rectangle on at least two occasions over last three MEGS surveys. MEGS estimated this amounted to 127 missed stations during the period and also estimated mean daily egg production for period 6 in 2022 at 19.58 stage 1 eggs $/ \mathrm{m}^{2} /$ day (WD07: $\mathrm{O}^{\prime}$ Hea et al. 2022).

The inclusion of the estimated egg abundance for the missing stations in Period 6 accounts for $10 \%$ of the annual egg production estimate in the western area for the 2022 survey.

The nominal end of spawning date of the 31st July is the same as used during previous survey years and the shape of the egg production curve for 2022 does not suggest that the end date needs to be altered. The provisional total annual egg production (TAEP) for the western area in 2022 was as $1.795 \times 10^{15}$. This is a $47 \%$ increase on the 2019 TAEP estimate which was $1.22 \times 10^{15}$.

Figure 8.6.1.1.7 shows the egg production curve for the southern area for the 2022 survey, along with those from previous surveys for comparison. The start date for spawning in the southern area was the $23^{\text {rd }}$ January. Portugal surveyed in Period 2 in division 9a. Sampling in the Cantabrian Sea where the majority of spawning occurs within the Southern area commenced on the $18^{\text {th }}$ March. The same end of spawning date of the $17^{\text {th }}$ July was used again this year and the spawning curve suggests that there is no reason for this to change. As in 2019 the survey periods were not
completely contiguous and this has been accounted for. The provisional total annual egg production (TAEP) for the southern area in 2022 was calculated as $3.21 \times 10^{14}$. This is a $25 \%$ decrease on the 2019 TAEP estimate of $4.23 \times 10^{14}$. A comparison of the total annual egg production (TAEP) for the western and southern area since 1998 is given below:

| Year | Western TAEP | Southern TAEP |
| :--- | :--- | :--- |
| 2022 | $1.795 \times 10^{15} \#$ | $3.21 \times 10^{14} \#$ |
| 2019 | $1.22 \times 10^{15}$ | $4.19 \times 10^{14}$ |
| 2016 | $2.20 \times 10^{15}$ | $2.25 \times 10^{14}$ |
| 2013 | $1.92 \times 10^{15}$ | $5.06 \times 10^{14}$ |
| 2010 | $1.36 \times 10^{15}$ | $4.59 \times 10^{14}$ |
| 2007 | $1.35 \times 10^{15}$ | $1.38 \times 10^{14}$ |
| 2004 | $1.54 \times 10^{15}$ | $3.18 \times 10^{14}$ |
| 2001 | $4.79 \times 10^{14}$ |  |
| 1998 |  |  |

The total annual egg production (TAEP) in 2022 for the western and southern components combined is $2.116 \times 10^{15}$. This is an increase in production of $29 \%$ compared to the 2019 estimate of $1.64 \times 10^{15}$ (Figure 8.6.1.1.8).

### 8.6.1.2 Mackerel fecundity and atresia estimation

Estimates of fecundity are given as preliminary realised fecundity which is the potential fecundity minus the atresia rate (for details see WD07: O'Hea et al. 2022). Atlantic mackerel samples were collected during survey periods $2-7$ over an area bounded by $59.36^{\circ} \mathrm{N} 14.20^{\circ} \mathrm{W}-36.54^{\circ} \mathrm{N}$ $2.32^{\circ} \mathrm{W}$. The analysis of potential fecundity is carried out by nine participating institutes. Preliminary fecundity results are based on 169 samples from periods 2 and 3 . The number of samples is higher than in 2019, when only 62 samples were available for the preliminary potential fecundity. The preliminary relative potential fecundity in 2022 is 1253 oocytes/gram female which is slightly higher than the preliminary estimate in 2019 of 1224 oocytes /gram female (Table 8.6.1.2.1). Due to time constraints no samples were analysed for atresia at the time of WGWIDE. For the preliminary estimation of the realised fecundity the mean atresia rate based on the previous seven surveys ( $6 \%$ ) was used. This resulted in a preliminary realised fecundity estimate for 2022 of 1178 oocytes/gram female fish.

### 8.6.1.3 Quality and reliability of the 2022 egg survey

The surveys in 2010 and 2013 were dominated by the issue of an early peak of western mackerel spawning and its close proximity to the nominal start date. In 2016 peak spawning reverted to May/June, a time that would traditionally be considered normal. In 2019, peak spawning in the western area was found to have occurred slightly earlier in Period 4 (Fig. 8.6.1.1.6). For 2022 the spawning pattern is remarkably similar to that reported for 2016.

The bulk of the spawning activity reported during historical surveys resulted from several egg production hotspots on and around the continental shelf edge and usually around the Celtic Sea and Porcupine Bank region. During 2016, high levels of spawning were recorded over a large area of the Northeast Atlantic with a large number of the stations being reported over deep water and well away from the continental shelf. In 2019 numbers of stage 1 eggs recorded on these northerly and western boundary stations were much reduced, although still present (Figures 8.6.1.1.4-5). This expansion was repeated in 2022 during Periods 5 and 6, however spawning densities recorded in these areas were significantly lower than reported in 2016 and 2019. Available surveys deployed during these periods were unable to fully delineate all boundaries. However, WGMEGS is satisfied that significant additional egg production is not being missed in these northern and western areas. Despite the inability to secure a northern spawning boundary for western area mackerel during periods 5 and 6, results from the recent exploratory MEGS surveys undertaken within these regions and reported to WGWIDE in 2021 (ICES, 2021a) provide reassurance that the fraction of spawning missed is a minor one and that the survey has indeed been successful in capturing the majority of spawning activity. An approach to estimate and account for the egg production missed as a result of the Irish survey cancellation in period 6 was developed and is detailed in WD07 (O'Hea et al. 2022).

### 8.6.1.4 Mackerel biomass estimates.

Based on the procedures of the WGMEGS Survey Manual (ICES, 2019a) \& Fecundity manual (ICES, 2019b) the preliminary spawning stock biomass (SSB) by components and components combined have been estimated as shown below using a preliminary fecundity estimate of 1178 oocytes/g female:

- $\quad 3.292 \mathrm{Mt}$ for western component (2019: 2.29 Mt ).
- $\quad 0.589 \mathrm{Mt}$ for southern component (2019: 0.80 Mt ).
- $\quad 3.881 \mathrm{Mt}$ for western and southern components combined (2019: 3.09Mt)


### 8.6.1.5 2022 North Sea mackerel egg survey

The North Sea Mackerel Egg Survey (NSMEGS) is designed to estimate the spawning stock biomass (SSB) of mackerel of the North Sea component of the Northeast Atlantic stock on a triennial basis. Prior to 2017 this survey was done utilizing the annual egg production method (AEPM). At the 2018 WGMEGS meeting, it was agreed to switch to the Daily Egg Production Method (DEPM) for the following survey NSMEGS (ICES, 2018b). The DEPM requires only one full sweep, in a short time period, over the entire mackerel spawning area, preferably during peak spawning time. A disadvantage of the DEPM is that it requires many more mackerel ovary samples to be collected in order to estimate batch fecundity and spawning fraction.

In 2022, the UK, Denmark and Norway conducted the North Sea survey between $5^{\text {th }}-24^{\text {th }}$ June. The spawning area (between $54^{\circ} \mathrm{N}$ and $62^{\circ} \mathrm{N}$ ) in the North Sea was surveyed with a single sweep. A total of 259 plankton stations and 38 pelagic trawl hauls were performed for the collection of mackerel adult and ichthyoplankton samples (O'Hea et al, WD08). The total area sampled in 2022 was slightly smaller than the area sampled during the previous survey in 2021.

The spatial daily egg production distribution is shown in Fig. 8.6.1.5.1. Procedures to estimate the Daily egg production are described in the WGMEGS Survey Manual (ICES, 2019a).

The DEP was calculated for the total investigated area. Provisional mackerel daily egg production for 2022 for the North Sea was estimated as $0.67 \times 10^{13} \mathrm{eggs}$. This is a $50 \%$ decrease on that reported for the 2021 survey.

## Adult parameters

Denmark sampled 1180 mackerel and collected ovary samples from 364 females. England sampled 225 mackerel and collecting ovary samples of 74 females. Norway collected 239 female mackerel (O'Hea et al, WD08). These samples were collected in June 2022 and at the time of writing, no analysis has yet been carried out. Analysis will take place before the end of 2022, with the results to be delivered prior to the WGMEGS meeting in April 2023.

### 8.6.1.6 2021 North Sea mackerel egg survey

In 2021 a North Sea Mackerel Egg Survey (NSMEGS, I1582) was carried out to estimate the spawning stock biomass (SSB) of mackerel of the North Sea component of the Northeast-Atlantic stock using DEPM methodology. The survey was designed to cover the whole spawning area in the North Sea $\left(53^{\circ} \mathrm{N}\right.$ to $\left.62^{\circ} \mathrm{N}\right)$.

The NSMEGS was carried out from $25^{\text {th }}$ May to $12^{\text {th }}$ June by The Netherlands, Denmark and Scotland (van Damme et al., WD03). The samples were collected and analysed according to the WGMEGS manuals (ICES 2019a, 2019b). A total of 294 plankton stations ,23 pelagic trawl hauls and 283 collected female samples were performed for the collection of mackerel adult and ichthyoplankton samples.

The spatial egg production distribution is shown in Fig. 8.6.1.6.1. The Daily egg production was calculated as $128 \times 10^{13}$ mackerel eggs for the total investigated area (Table 8.6.1.6.1).

The DEPM adult parameters were estimated with the data provided by the Netherlands (van Damme et al., WD03). Batch fecundity was estimated 18735 eggs/g. Corrected mean female weight was estimated as 331 g . Spawning fraction in the North Sea was calculated as $18 \%$ and sex ratio was 0.53. Adult parameters are presented in Table 8.6.1.6.2.

Using the DEP (stage Ia) for the entire sampled area and the estimated adult parameters for the North Sea component leads to an estimated SSB of $2380 \times 10^{3} \mathrm{t}$ in 2021.

### 8.6.2 Demersal trawl surveys in October - March (IBTS Q4 and Q1)

An index of survivors in the first autumn-winter (recruitment index) is normally derived from a geostatistical model fitted to catch data from bottom trawl surveys conducted during autumn and winter. A complete description of the data and model can be found in Jansen et al. (2015) and the NEA mackerel Stock Annex.

The data collection in 2022 Q1 was incomplete as two Scottish surveys (IBTS-NS Q1 and SWC Q1) were cancelled due to technical issues with R/V Scotia.

The area covered by IBTS-NS Q1 and SWC Q1 have historically been an important nursery area. A major fraction of the total estimated recruits are in this area in Q1 (Figure 8.6.2.1). The fraction varies from year to year and it was considered too uncertain to interpolate or assume that the same fraction was in the area as in Q1 2021. As a result, the recruitment index (survivors in the
first autumn-winter) has not been updated in 2022 to estimate the value for the 2021 year-class. The time series from ICES (ICES, 2021b) that is used for this years' assessment is shown in Figure 8.6.2.2.

### 8.6.3 International Ecosystem Summer Survey in Nordic Seas (IESSNS, A7806)

IESSNS is the only annual survey providing data used in the assessment and covers summer feeding distribution of mackerel age 3+ in Nordic Seas and was successfully conducted in 2022. Major survey results worth mentioning is that survey coverage expanded $32 \%$, compared to 2021 , as Greenlandic waters, north of $62^{\circ} \mathrm{N}$, were surveyed again and mackerel distribution south of Iceland demanded southward expansion of survey to latitude $61^{\circ} 15^{\prime}$ in the Iceland basin and on the Reykjanes ridge. Value of the mackerel index was impacted by two extremely large catches, 103 and 70 tonnes $\mathrm{km}^{-2}$, which contributed $33 \%$ of the biomass index value. Extreme catches also impacted index calculations in 2017, 2019 and 2020. Analytical work to develop index calculation method less sensitive to extreme catches will be undertaken at ICES Working Group on Improving use of Survey Data for Assessment and Advice (WGISDAA) annual meeting in October 2022. The western part of the northern Norwegian Sea (stratum 9) was oversampled as three surface trawl stations were added, at the dynamic stratum boundary, at only half the distance from next station, 35 nm instead of 70 nm . Mackerel was caught at two of these stations and the maximum catch per station was approximately one ton. All three stations were included in the index calculations and the dynamic stratum boundary extended 35 nm westward of these three stations. The zero-line for mackerel abundance was reached south and north of Iceland and in Greenlandic waters. It was not reached in the north-western and north-eastern part of the Norwegian Sea but given that the polar front with water too cold for mackerel is usually found close to the north-westernmost catches, we assume that the zero-line was practically reached there. Towards the Barents Sea the zero-line was not reached but this is considered of less quantitative importance based on low catch rates. The zero-line was not reached on the European shelf, where mackerel are present west of the British Isles and in the southern North Sea. The IESSNS cruise report is available as a working document to this report (WD01) and a detailed survey description is available in the mackerel Stock Annex.

The main results are that estimated total stock abundance and total biomass increased $43 \%$ compared to 2021. When the two extreme catches are excluded from index calculations in 2022, biomass and abundance is similar to 2021 values. Internal consistency increased compared to 2021, particularly for ages 5-8 years which had lower consistency than other ages in 2021. Abundance estimates by age are displayed in input data for the assessment (Table 8.7.1.9). Figures 8.6.3.1-2 display estimates of total stock abundance and stock biomass with confidence intervals, with and without two extreme catches in 2022, for the time series. Figures 8.6.3.3-4 show the internal consistency and catch curves for abundance at age from 2010 to 2022. Figures 8.6.3.6-7 display swept area trawl catch rate and mean mackerel density per rectangle for 2022, and mean mackerel density per rectangle for years 2010 and from 2012 to 2022.

### 8.6.4 Tag Recapture data

The following is a summary of the most important information on tag recapture data, more detailed info can be found in a working document attached to this report (Slotte and Hølleland, WD09). Information from steel tagging experiments conducted by Institute of Marine Research in Bergen (IMR) on mackerel at spawning grounds west of Ireland and British Isles in May-June and the respective recaptures at Norwegian factories with metal detectors (Tenningen et al. 2011)
was introduced to the mackerel assessment during ICES WKPELA 2014 (ICES, 2014). Data from release years 1980-2004, and recapture years 1986-2006 have been used in the update assessments following this benchmark. From 2011 onwards IMR changed tagging methodology to radio-frequency identification (RFID), more specifically passive integrated transponder tags (PIT-tags). This allowed for more automated data processing with recaptures from scanned landings at factories in Norway, Scotland and Iceland now being updated in real time to an IMR database over internet.

The data format is the same for both tag types; a table containing the numbers of tagged fish per year class in each release year, and the corresponding numbers scanned and recaptured of the same year classes in all years after release. The RFID data were considered to be a separate time series with a different scaling factor (survival) than the steel tags, and it has been used in update assessments following the ICES WKWIDE2017 benchmark (ICES, 2017). For steel tags data from ages 2-11 and all recapture years are used in the assessment. During the 2017 benchmark it was decided to use the same filtering for the RFID data from release year 2011 onwards. However, following decisions made during ICES IBPNEAMac 2019 (ICES 2019c) update assessments are now only using RFID data from release years 2013 onwards, ages 5-11 and recapture year 1 and 2 after release.

An overview of all RFID tagging data in terms of numbers tagged, biomass scanned, and numbers recaptured per year, and geographical distributions of data are shown in Figures 8.6.4.1-3. The exclusion of recapture years 3 and longer after release is due to potential tag loss over time, which seem evident in the RFID data (WD09). The exclusion of release years 2011-2012 is mainly based in lack of distributional coverage of scanned fishery, which changed significantly when more countries joined the program from 2014 onwards (Figure 8.6.4.2). The exclusion of ages 14 , was mainly because early in the time series these age groups were relatively few compared with the scanned fish year 1 and 2 after release, leading to some noise in the data. However, the age structure of tagged and scanned fish year 1-2 after release has developed over time series to be more overlapping, and high proportions of tagged mackerel are now at ages 2-4 (Figure 8.6.4.4).

Trends in year class abundance indices from RFID data based on recaptures year 1 and 2 after release now seem consistent and informative for assessment from ages 2-12 (Figure 8.6.4.5). Note that an alternative assessment at WGWIDE 2021 using these indices for the selected ages 5-11 instead of the regular data table resulted in negligible differences in SSB trend and same leave out RFID data effects; i.e. higher SSB in most recent years when excluding RFID data. Translating these abundance indices into different age-aggregated biomass indices also show comparable time trend with SSB from WGWIDE 2022 from release years 2013 onwards (Figure 8.6.4.5). Especially the marked decrease in SSB from 2017-2020 seem to follow the decline in the RFID biomass estimates, which may explain why leave out RFID runs from WGWIDE 2022 tends to lift the SSB upwards.

The signals of total mortality rate ( Z ) in fully mature fish ages 4-12 for year classes 2003-2014 tend to be higher in the RFID data than in the catch data with the data from final WGWIDE2022 assessment in between, whereas estimated $Z$ from the international trawl survey (IESSNS) is sticking out as the lowest of all sources (Figure 8.6.4.6).

The overall conclusion is that the RFID time series is slowly developing, but still is a very short time series. Nevertheless, the data seem quite informative for stock assessment, although showing higher total mortality rate signals than the other input data. Such conflicting trends suggest that year to year variations in assessment and leave out effects may frequently occur in coming years when time series are short. Finally, the new development of the time series suggests that the current filtering of RFID data for use in stock assessment should be revised in near future.

This especially counts for the inclusion of younger ages 2-4 that may be informative for incoming year classes to the stock.

### 8.6.5 Other surveys

### 8.6.5.1 International Ecosystem survey in the Norwegian Sea (IESNS, A3675)

After the mid-2000s an increasing amount of NEA mackerel has been observed in catches in the Norwegian Sea during the International Ecosystem survey in the Norwegian Sea in May (IESNS) targeting herring and blue whiting (Salthaug et al. 2019; 2020).

The spatial distribution pattern of mackerel in 2022 was quite similar to previous years, with mackerel present in the south-eastern Norwegian Sea. However, there were small catches of mackerel as far north as $68^{\circ} \mathrm{N}$ in 2021, but this year the catches only extended north to about $64^{\circ} \mathrm{N}$. This is the lowest northward extent of mackerel catches during IESNS after 2007 (first year with data from all participating vessels).

The IESNS survey provides valuable, although limited, quantitative information on mackerel. It is an acoustic survey and the trawl hauls are mainly targeting acoustic registrations of herring and blue whiting. Thus, the survey does not provide proper mackerel sampling in the vertical dimension and has too low trawl speed for representative sampling of all size groups of mackerel. Therefore, no further quantitative information can be drawn from these data.

### 8.6.5.2 Acoustic estimates of mackerel in the Iberian Peninsula and Bay of Biscay (PELACUS, A2548)

PELACUS survey data have not been processed on time for WGWIDE 2022 and therefore, no new information from the Bay of Biscay on mackerel distribution and abundance during spawning time is available.

### 8.7 Stock Assessment

### 8.7.1 Update assessment in 2022

The update assessment was carried out by fitting the state-space assessment model SAM (Nielsen and Berg, 2014) using the R library stockassessment (downloadable at install_github("fishfollower/SAM/stockassessment")) and adopting the configuration described in the Stock Annex.

The assessment model is fit to catch-at-age data for ages 0 to 12 (plus group) for the period 1980 to 2021 (with a strong down-weighting of the catches for the period 1980-1999) and three surveys: 1) SSB estimates from the triennial Mackerel Egg survey (every three years in the period 1992-2022),
2) a recruitment index from the western Europe bottom trawl IBTS Q1 and Q4 surveys (1998-2020, could not be updated for 2021) and
3) the abundance estimates for ages 3 to 11 from the IESSNS survey (2010, 20122022).

The model also incorporates tagging-recapture data from the Norwegian tagging program (for fish recaptured between 1980 and 2005 for the steel tags time series, and fish recaptured between 2014 and 2021 (age 5 and older at release) for the radio frequency tags time series).

Fishing mortality-at-age and recruitment are modelled as random walks, and there is a process error term on abundances at ages 1-11.

The differences in the new data used in this assessment compared to the last year's assessment were:

- Addition of the 2021 catch-at-age (0-12+), weights-at-age in the catch and in the stock and maturity ogive, proportions of natural and fishing mortality occurring before spawning.
- Addition of the 2022 abundance-at-age (3-11) index from IESSNS.
- Addition of the 2022 SSB index (preliminary value) from the 2022 mackerel egg survey
- The inclusion of the tag recaptures from 2021

Input parameters and configurations are summarized in Table 8.7.1.1. The input data are given in tables 8.7.1.2-9. Given the size of the data base, only the data from the last year of recaptures is given in this report (table 8.7.1.10).

### 8.7.2 Model diagnostics

## Parameter estimates

The estimated parameters and their uncertainty estimates are shown in table 8.7.2.1 and figure 8.7.2.1. The model estimates different observation standard deviations for young fish and for older fish. Reflecting the suspected high uncertainty in the catches of age 0 fish (mainly discards), the model gives a very poor fit to this data (large observation standard deviation). The standard deviation of the observation errors on catches of age 1 is lower, though still high, indicating a better fit. For the age 2 and older, the fit to the catch data is very good, with a very low observation standard deviation.

The observation standard deviations for the egg survey and the IESSNS surveys ages 4 to 11 are higher, indicating that the assessment gives a lower weight to the information coming from these surveys compared to that from the catches. The IESSNS age 3 is very poorly fit in the assessment (high observation standard deviation). Overdispersion of the tag recaptures has the same meaning as the observation standard deviations, but is not directly comparable.

The catchability of the egg survey is estimated to be 1.17 , greater than 1 , which implies that the assessment considers the egg survey index to be an overestimate of SSB. The catchabilities at age for the IESSNS increase from 0.79 for age 3 to 1.96 for age 9 . Since the IESSNS index is expressed as fish abundance, this also means that the assessment considers the IESSNS to provide overestimated abundance values for ages 4 to 11 . The post tagging mortality estimate is higher for the steel tags ( $\sim 40 \%$ ) than for the RFID tags ( $\sim 16 \%$ ).

The process error standard deviation (ages 1-11) is moderate as are the standard deviations of the F and recruitment random walks.

The catchability parameters for the egg survey, recruitment index and post tagging survival appear to be estimated more precisely than other parameters (table 8.7.2.1). The catchability for the IESSNS has a slightly higher standard deviation, except for age 3 which is significantly higher. Uncertainty on the observation standard deviations is larger for the egg survey, the IESSNS age 3 , for the recruitment index and for the catches at age 1 than for the other observations. The uncertainty on the observation variance estimates is not particularly high, especially for the data sources with the lowest observation variances, which are the most influential in the assessment (figure 8.7.2.2). Uncertainty on the overdispersion of the tag data is high. The standard deviation on the estimate of process error is low, and the standard deviations for the estimates of F random
walk variances of age 0 and 1 are both very high. The uncertainty on the random walk variance for recruitment is very large, indicating that the parameter was poorly estimated.

The estimated AR1 error correlation structure for the observations from the IESSNS survey age 3 to 11 has a high correlation between the errors of adjacent ages ( $\mathrm{r}=0.82$ ), decreasing exponentially with age difference (figure 8.7.2.3.). This high error correlation implies that the weight of this survey in the assessment is lower than for a model without correlation structure, which is also reflected in the high observation standard deviation for this survey.

There are some correlations between parameter estimates (figure 8.7.2.4):

- catchabilities are positively correlated (especially for the IESSNS age 4 to 11), and negatively correlated to the survival rate for the RFID tags. This simply represents the fact that all scaling parameters are linked, which is to be expected.
- the observation variance for the recruitment index is inversely correlated to the variance of the random walk of the recruitment. This implies that when the model relies less on the recruitment index, the estimated recruitment time series becomes smoother.
- the parameter related to the magnitude of the correlation in the AR1 matrix for the IESSNS is correlated to the observation variance for this survey, which reflects the fact that a strong correlation for the errors is linked to lower weight of the surveys.


## Residuals

The "one step ahead" (uncorrelated) residuals for the catches show some weak patterns in the residuals, with a prevalence of positive residuals for ages 3 to 10 between 2008 and 2014 and again in the last 2 years (figure 8.7.2.5). Empirical correlation plot in the residuals shows positive correlations between ages 3 to 12 (figure 8.7.2.6) which suggest that incorporating a correlation structure in the observation error for the catches might be appropriate. Residuals are of a similar size for all ages, indicating that the model configuration with respect to the decoupling of the observation variances for the catches is appropriate.

The residuals for the egg survey show a strong temporal pattern with large positive residuals for the period 2007-2010-2013, followed by large negative residuals for 2016, 2019 and 2022. This pattern reflects the fact that the model, based on all the information available, does not follow the trend present in the egg survey, with a steep decline between 2013 and 2016 (when the stock was at its highest) and the very low 2019 value. The relatively high observation variance for this survey indicates a poor fit with the egg survey due mainly to these observations which indicate a different trend to the other available observations.

Residuals for the IESSNS indices are relatively well balanced for most years. Despite the strong drop in the abundances at age in 2018 and 2021, the residuals for these years do not indicate any year effect (e.g. no large residuals of the same sign observed across ages). Correlations between age in the observation errors for this survey are explicitly modelled in the assessment, and as a result, empirical correlations in the one step ahead residuals between ages are low (figure 8.7.2.7).

Residuals for the recruitment index show no particular pattern, and appear to be relatively randomly distributed in the earlier years although positive residuals are consistently observed over the most recent 5 years, indicating that the model has difficulties agreeing with the sustained period of high values in the index.

Finally, inspection of the residuals for the tag recaptures (figure 8.7.2.8) did not show any specific pattern for the RFID data. For the steel tags, there is a tendency to have more positive residuals at the end of the period which could indicate that using a constant survival rate for this dataset may not be appropriate.

## Leave one out runs

In order to visualise impact of individual surveys on the estimated stock trajectories, the assessment was run leaving out successively each of the survey data sources (figure 8.7.2.9).

All leave one out runs showed parallel trajectories in SSB and Fbar, except for the run that excluded the RFID tag information, which shows a less steep decline in SSB since 2014. For recruitment, all runs also resulted in similar trajectories, except for the run without the recruitment index, in which recruitment decreased from high levels in the mid-2010s to historically low levels in the terminal assessment year.

Excluding the IESSNS survey resulted in estimates very close to the base case run, with slightly lower SSB and higher Fbar for the period covered by this survey. Removing the recruitment index had a similar effect on the estimated stock trajectories, but with a larger discrepancy in SSB for the most recent years. Without the recruitment index, the estimates of age 0 abundance are only informed by the catch data, which are considered uninformative by the model. As a result, the estimated recruitment for this specific run has a very different trajectory, which - despite the adjustments of year-class strength through the process error as fish become older - has an effect on SSB estimates. The exclusion of SSB estimates from the egg survey resulted in a larger estimated stock, exploited with a lower fishing mortality. The run leaving out the RFID also resulted in a higher SSB than in the assessment using all data for the years after 2017, and a slightly higher fishing mortality between 2011 and 2015, but lower after 2019. The magnitude of the effect of removing the RFID data is similar to that of removing other surveys.

As in previous years, the update assessment appears to trade-off the information coming from the IESSNS which leads to a more optimistic perception of the stock, and the information from the egg survey and the tags which suggests a more pessimistic perception of the stock.

## Additional sensitivity runs

A series of additional sensitivity runs were carried out to explore the potential influence of the additional uncertainty in some of the new data included in the 2022 update assessment namely the 2022 SSB index from the mackerel egg survey and the possible higher uncertainty in the 2021 catch-at-age data as a consequence of the missing Russian data.

- Sensitivity to assumption made for missing coverage in the 2022 egg survey index

The egg production assumed for the missing coverage in period 6 accounted for $9.4 \%$ of the 2022 SSB index from the mackerel egg survey. An alternative SAM assessment was run with a 2022 SSB index decreased by $9.4 \%$ ( 3.51 Mt ) compared to the value used in the update assessment $(3.88 \mathrm{Mt})$. The stock estimates for this run are almost identical to the update assessment (figure 8.7.2.10) which indicates that the current assessment is fairly robust to a $9.4 \%$ difference in the 2022 egg survey index of SSB

- Effect of increased uncertainty for 2021 catch-at-age

The current SAM assessment uses the same observation variance for all years, thereby considering that level of uncertainty in the catches does not vary over time. However, given the lack of sampling data from the Russian catch for 2021, and the uncertainty on the total catch value provided, the 2021 catches-at-age for the stock are potentially more uncertain than in normal years. In order to test the effect this potential larger uncertainty has on the assessment, the observation variance for the catch-at-age data for 2021 was increased by $50 \%$ compared to other years. This had no noticeable effect on the assessment (figure 8.7.2.11)

### 8.7.3 State of the Stock

The stock summary is presented in figure 8.7.3.1 and table 8.7.3.1. The stock numbers-at-age and fishing mortality-at-age are presented in tables 8.7.3.2-3. The spawning stock biomass is estimated to have increased almost continuously from just above 2 million tonnes in the late 1990s and early 2000s to 5.9 million tonnes in 2014 and 2015 and subsequently declined to reach a level just above 3.6 million tonnes in 2020 and increased slightly in 2021 to 3.9 million tonnes. The fishing mortality has declined from levels between $\mathrm{F}_{\mathrm{pa}}(0.36)$ and $\mathrm{F}_{\lim }(0.46)$ in the mid-2000s to levels at or below $\mathrm{F}_{\text {MSY }}(0.26)$ between 2010 to 2019 and increased sharply in the last two years to 0.31 in 2021. The recruitment time series from the assessment is not considered a reliable indicator of year-class strength (see section 8.7.5.1).

There are clear indications of changes in the selectivity of the fishery over the last 30 years (figure 8.7.3.2.). In the 1990s, the fishery seems to have had a steeper selection pattern (more rapid increase in fishing mortality with age). Between the end of the 1990s and the end of the 2000s, the selection on the ages 1 to 5 decreased, and selection of older fish (7 and older) increased. After 2008, the pattern started reversing towards a steeper selection pattern, until 2017. Since then, selection on age 2 to 5 decreased sharply, as the fishery targeted more the older part of the population (age 6 and older).

### 8.7.4 Quality of the assessment

## Parametric uncertainty

Large confidence intervals are associated with the SSB in the years before 1992 (figure 8.7.4.1 and figure 8.7.2.7). This results from the absence of information from the egg survey index, the downweighting of the information from the catches and the assessment being only driven by the tagging data and natural mortality in the early period. The confidence intervals become narrower from the early 1990s to the mid-2000s, corresponding to the period where information is available from the egg survey index, the tagging data and (partially) catches. The uncertainty increases slightly in the most recent years and the SSB estimate for 2021 is estimated with a precision of $+/-$ $24.8 \%$ (figure 8.7.4.1). There is generally also a corresponding large uncertainty on the fishing mortality, especially before 1995. The estimate of Fbart-8 in 2021 has a precision of $+/-27.4 \%$.

## Model instability

The retrospective analysis was carried out for 8 retro years, (or peels) by fitting the assessment using the 2022 data, removing successively 1 year of data (figure 8.7.4.2.). There was a systematic retrospective pattern found in Fbar for the older retrospective peels (current year -4 to current year -8) with a systematic downwards revision. The was also a pattern in the opposite direction for the SSB. However, this pattern is not apparent in the most recent peels (current -1 to -3 ), and the Mohn's rho value calculated over the last 5 years is of 0.18 for Fbar and -0.11 for SSB. Recruitment appears to be quite consistently estimated for the older peels (current -3 to -8 ), but the perception changed for the last 2 peels. This change is associated with an increase in the observation variance for the recruitment index, meaning that the recruitment estimates were more influenced by the recruitment index in the older peels, which was less the case in the last 2 peels.

## Model behaviour

The realisation of the process error in the model was also inspected. The process error expressed as annual deviations in abundances-at-age is shown in figure 8.7.4.3 which shows indications of some pattern across time and ages. There is a predominance of positive deviations in the recent years for ages 5 to 8 . While process error is assumed to be independent and identically
distributed, there is clear evidence of correlations in the realisation of the process error in the mackerel assessment, which appears to be correlated both across age-classes and years.

The temporal autocorrelation can also be visualised if the process error is expressed in term of biomass (process error expressed as deviations in abundances-at-age multiplied by weight at age and summed over all age classes, figure 8.7.4.4). Periods with positive values (when the model estimates larger global abundances-at-age than corresponding to the survival equation) have been alternating with periods with negative values (1991-1994, 2004-2005, and 2017-2019). For the years between 2007 and 2016, the biomass cumulated process error remains positive, and large (e.g. in 2013 - almost equivalent to the total catch weight). The reason for this aspect of model behaviour could not be identified.

### 8.7.5 Exploratory runs

### 8.7.5.1 Assessment starting at age 2

The age 0 estimates in the current assessment mainly rely on the recruitment index; the catch-atage 0 information is considered by the model as uninformative (large observation variance). Catch-at-age information becomes influential at age 2 (very low observation variance). The recruitment signal provided by abundances estimated at age 2 or 3 (when the fish enters the fishery), is different from the signal in the age 0 abundance (figure 8.7.5.1). Age 0 abundances are less variable than abundances at age 2 and 3 . For the period before 2012, there is a broad agreement in the perception of year class strength, although some year classes that do not appear particularly large at age 0 are perceived as very large at age 2 and 3 (e.g. 2002 year-class). For the more recent period (since the 2013 year-class), there is a greater discrepancy between recruitment at age 0 and that derived at older ages. While the age 0 abundances indicate very high recruitment for the year-classes 2012 to 2019, some of those year-classes appear as particularly poor based on age 2 and 3 abundances (2015, 2017 and 2018). As very little fishing occurs between ages 0 and 2 and 3 , exploitation is not likely to explain these changes in the perception of cohort strength. Such variations could be possibly due to variations in natural mortality (e.g. the strength of a cohort may not be fully determined at age 0 and processes occurring during the first years of life may still be determining year-class strength). However, some cohorts increase in size as they become older (e.g. 2002 and 2011), which clearly indicates that this is more likely a model artefact. The cohort strength at age 0 , based on the recruitment index, is progressively revised, due to the process error occurring on annual survival, so that cohort strength at age 2 corresponds to the information coming from the catches.

This discrepancy between the recruitment estimates at age 0 and the actual size of the cohort when entering the fishery implies that the age 0 recruitment does not give an accurate indication of year-class strength, and should not be used to make assumptions on stock development in the near future. The implications of starting the assessment at age 0 for the short term forecast done to compute the catch advice, however, are relatively limited, with the last estimated recruitment value (R2021 this year) contributing to around $6 \%$ of the catch in the advice year.

As very little fishing occurs on 0 and 1 year olds, and catch-at-age data is considered very noisy, and since there appears to be a disagreement between the recruitment index at age 0 and at older ages in the recent years, it does not seem appropriate to use age 0 or 1 as the youngest age in the assessment. An exploratory run was conducted starting the assessment at age 2 (and hence removing catch-at-age information for age 0 and 1 and the recruitment index, while retaining the remainder of the data and an unchanged model configuration).

Both the update and exploratory assessments give a very similar perception of the SSB and Fbar trajectories (figure 8.7.5.2), with only small differences in the last 2 years for both SSB and Fbar. The recruitment at age 2 (in blue on figure 8.7.5.2, note that the curve should be shifted
backwards by 2 years to compare year-class strength with the recruitment at age 0 , red curve) shows a much more variable year-class strength signal, with the same perception of year class strength as the age 0 recruitment for some years (broadly between year-classes 2000 and 2012), but a much lower estimated year-class strength since 2012.

In conclusion, both models are in broad agreement in terms of fit to the available data and stock trajectories such that the model starting at age 2 could be considered as potential alternative to the current model at the next benchmark workshop for this stock.

### 8.8 Short term forecast

The short-term forecast provides estimates of SSB and catch in 2023 and 2024 (given an assumed catch for the current (intermediate) year) and a range of management options for the catch in 2023.

All procedures used this year follow those used in the benchmark of 2014 as described in the stock annex.

### 8.8.1 Intermediate year catch estimation

Estimation of catch in the intermediate year (2022) is based on declared quotas, interannual transfers and information from the fisheries shown in the text table in Section 8.1.

### 8.8.2 Initial abundances at age

The recruitment estimate at age 0 from the assessment in the terminal assessment year (2021) was considered too uncertain to be used directly, because this year class has not yet fully recruited into the fishery. The last recruitment estimate is therefore normally replaced by predictions from the RCT3 software (Shepherd, 1997). The RCT3 software evaluates the historical performance of the IBTS recruitment index, by performing a linear regression between the index and the SAM estimates over the period 1998 to the year before the terminal year. The recruitment is then calculated as a weighted mean of the prediction from this linear regression based on the IBTS index value, and a time tapered geometric mean of the SAM estimates from 1990 to the year before the terminal year. The time tapered geometric mean gives the latest years more weight than a geometric mean. This is done because the recent productivity of the stock appears different than in the 1990's.

However, no IBTS index data point is available for 2021 and therefore the time tapered geometric mean was used without adjustments. This is as close to the standard procedure as practicable and leads to an expected recruitment of 5844 million.

### 8.8.3 Short term forecast

A deterministic short-term forecast was conducted using FLR (www.flr-project.org). Table 8.8.3.1 lists the input data to the forecast and tables 8.8.3.2 and 8.8.3.3 provide projections for various fishing mortality multipliers and catch constraints in 2023.

Assuming catches for 2022 of 1131 kt , F was estimated at 0.36 (above Fmsy) and SSB at 3.77 Mt (above $\mathrm{B}_{\mathrm{pa}}$ ) in spring 2022. If catches in 2023 equal the assumed catch for 2022, F is expected to increase to 0.40 (above $\mathrm{F}_{\mathrm{pa}}$ ) in 2023 with a corresponding decrease in SSB to 3.60 Mt in spring 2023. Assuming an F of 0.40 again in 2024, the SSB will further decrease to 3.33 Mt in spring 2024.

Following the MSY approach, exploitation in 2023 shall be at $\operatorname{FMSY}$ (0.26). This is equivalent to catches of 782 kt and a decrease in SSB to 3.68 Mt in spring 2023 ( $2 \%$ decrease). During the subsequent year, SSB will remain at a similar level ( 3.65 Mt ) in spring 2024.

### 8.9 Biological Reference Points

A management strategy evaluation Workshop on northeast Atlantic mackerel (MKMSEMAC) was conducted during 2020 (ICES, 2020) which resulted in the adoption of new reference points for NEA mackerel stock by ICES.

The table below summarises the currently used reference points.

| Framework | Reference point | Value | Technical basis | Source |
| :--- | :--- | :--- | :--- | :--- |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | 2.58 million <br> tonnes | $\mathrm{B}_{\text {pa }}$ | ICES (2020) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.26 | Stochastic simulations | ICES (2020) |
| Precautionary <br> approach | $\mathrm{B}_{\text {lim }}$ | 2.00 million <br> tonnes | $\mathrm{B}_{\text {loss }}$ (In 2003 from the 2019 WGWIDE assessment | ICES (2019c) |

### 8.10 Comparison with previous assessment and forecast

## Stock assessment output

The last available assessment used for providing advice was carried out in 2021 during the WGWIDE. The 2022 WGWIDE assessment gives a slightly different perception of the development of the stock, with a higher SSB estimated for the period 2014-2017 and a lower Fbar estimated over the period 2009-2018 (figure 8.10.1). The differences in the 2020 TSB, SSB and Fbar estimates between the 2021 and 2022 assessments are $-7.0 \%,-9.35 \%$ and $15.4 \%$ respectively.

|  | TSB 2020 | SSB 2020 | Fbar4-8 2020 |
| :--- | :--- | :--- | :--- |
| Assessment |  |  |  |
| 2021 WGWIDE Update | 5131499 tonnes | 3938555 tonnes | 0.249 |
| 2022 WGWIDE Update | 4772765 tonnes | 3570188 tonnes | 0.287 |
| Revision | $-7.0 \%$ | $-9.35 \%$ | $15.4 \%$ |

The addition of a new year of data modified only marginally the model parameters compared to last year (figure 8.10.2). The observation standard deviation has increased slightly for the IESSNS survey while it remained unchanged for the other data. Process variances all increased slightly,
except the process error on abundances at age 1-11 which remained unchanged. There was also a minor change in the catchabilities for the age 4 to 6 in the IESSNS survey.

The uncertainty on the estimates of the process variances have decreased slightly (especially for the recruitment random walk) but the uncertainty on other parameters is very similar to last year. The uncertainty on SSB and Fbart-s in $^{\text {in this year's assessment is lower for the recent period }}$ (for the estimates since 2010), but has increased slightly for the terminal year estimates (figure 8.7.4.1).

## Short term forecast

The estimation for the intermediate year (2021) catch used for the short-term forecast in the advice given last year was $10.87 \%$ higher than the actual catches as reported to WGWIDE 2022 (table below). The intermediate year assumption is made by summing the unilateral TAC declared, taking interannual transfers into account, and adding anticipated discards. During the WGWIDE, participants may provide information from their national administration and industry on the expected rate of use of the TAC (most often 100\%), which can lead to a modification of the expected national catches. In 2021, several countries were not able to catch their national TAC, due to restrictions on access to UK waters, and this could not be anticipated at the time of the working group. The undershoot of some of the national TACs lead to an assumed intermediate year catch that was too large. As the situation with regard to access to UK waters is still ongoing, an assumption regarding TAC undershoot is made in this year calculation of the intermediate year catch (see section 8.1).

Since the intermediate year catch was overestimated, the 2021 short-term forecast produced an underestimate of the 2021 SSB (by $9.78 \%$ ) and $13.6 \%$ overestimation of Fbar 2021.

|  | Catch (2021) | SSB (2021) | F $_{\text {bar4-8 (2021) }}$ |
| :--- | :--- | :--- | :--- |
| 2021 WGWIDE forecast | 1199103 t | 3510849 t | 0.35 |
| 2022 WGWIDE assessment | 1081541 t | 3891546 t | 0.31 |
| $\%$ difference | $10.87 \%$ | $-9.78 \%$ | $13.60 \%$ |

### 8.11 Management Considerations

Details and discussion on quality issues in this year's assessment is given in Section 8.7 above.
From 2001 to 2007, the internationally agreed TACs covered most of the distribution area of the Northeast Atlantic mackerel. From 2008 to 2014, no agreement was reached among the Coastal States on the sharing of the mackerel quotas. In 2014, three of the Coastal States (EU, NO and FO) agreed on a Management Strategy and sharing arrangement for 2014 to 2018. In November 2018, the agreement from 2014 was extended for two more years until 2020. There has been no new agreement on a new Management Strategy or share of the stock since then. Despite agreeing to abide by the ICES advice, the total declared quotas in each of the years 2015 to 2021 all exceed the advised catch by ICES (figure 8.11.1).

The mackerel in the Northeast Atlantic is traditionally characterised as three distinct 'spawning components': the southern component, the western component and the North Sea component. The basis for the components is derived from tagging experiments (ICES, 1974). However, the methods normally used to identify stocks or components (e.g., ectoparasite infections, blood phenotypes, otolith shapes and genetics) have not been able to demonstrate significant differences between animals from different components. A review of the mackerel in the North Sea, carried
out during WKWIDE 2017 (ICES, 2017) concluded that Northeast Atlantic mackerel should be considered as a single population (stock) with individuals that show stronger or weaker affinity for spawning in certain parts of the spawning area.

Since the mid-1970s, ICES has continuously recommended conservation measures for the North Sea component of the Northeast Atlantic mackerel stock (e.g., ICES, 1974; ICES, 1981). The measures advised by ICES to protect the North Sea spawning component (i.e., closed areas and minimum landing size) aimed to promote the conditions that make a recovery of this component possible.

The minimum landing size (MLS) for mackerel is currently set at 30 cm for the North Sea and 20 cm in the western area. The MLS of 30 cm in the North Sea was originally introduced by Norway in 1971 and was intended to protect the very strong 1969 year-class from exploitation in the industrial fishery (Pastoors, 2015). In the early 1990s, ICES recommended that, because of mixing of juvenile and adult mackerel on western waters fishing grounds, the adoption of a 30 cm minimum landing size for mackerel was not desirable as it could lead to increased discarding (ICES, 1990; 1991). A substantial part of the catch of (western) NEA mackerel is taken in ICES Division 4.a during the period October until mid-February to which the 30 cm MLS applies even though there is limited understanding on the effectiveness of minimum landing sizes in achieving certain conservation benefits (STECF, 2015).

### 8.12 Ecosystem considerations

An overview of the main ecosystem drivers possibly affecting the different life-stages of Northeast Atlantic mackerel and relevant observations are given in the Stock Annex. The discussion here is limited to recent features of relevance.

## Production (recruitment and growth)

Since 2012 the recruitment index (age 0) has been estimating substantially larger year-classes than what is later estimated at age 3 when they enter the fishery and the other surveys. It is not known if this mismatch is a sampling bias or altered mortality of the juveniles between age 0 and 3.

The rapid increase in stock size up until around 2015 was suggested to drive the recent expansion of the spawning northward into new areas (Jansen, 2016). There are several indications of a northward shift and/or expansion in spawning and nursery area towards northern and northeastern areas since 2016 (ICES, 2016; Nøttestad et al., 2018; Bjørdal, 2019; Bjørdal et al. in press). This northerly shift seems to have continued (Nøttestad et al., 2018). However, spawning in the Norwegian Sea was shown to be of little quantitative significance in 2021 (Burns and O'Hea, WD 15 to WGWIDE 2021 (ICES, 2021b)).

Growth (i.e. length- and weight-at-age) have declined substantially in recent times for all ages (e.g. 0-3 year-old in 1998-2012, Jansen and Burns, 2015; all ages in 2005-2015, Jansen and Burns, 2015; Ólafsdóttir et al., 2015). The variations in growth of mackerel in all ages are correlated with mackerel density, e.g. mean weight-at-length have been shown to be positively related to location, day-of-year, temperature and SSB. Furthermore, the density dependent regulation of growth from juvenile to adult mackerel, appears to reflect the spatial dynamics observed in the migration patterns during the feeding season. As such, growth rates of the juveniles were tightly correlated with the density of juveniles in the nursery areas (Jansen and Burns, 2015) and growth for adults (age 3-8) were correlated with the combined effects of mackerel and herring stock sizes (Ólafsdóttir et al., 2015). Conspecific density-dependence was most likely mediated via intensified competition associated with greater mackerel density, possibly also coinciding with decreased prey availability. Nevertheless, weight-at-age of mackerel both from the catches and the
surveys have increased during the last few years, particularly for the younger year classes from 1 to 6 years of age (ICES, 2019c; 2020; 2021b), coinciding with reduced abundance of mackerel in recent years.

## Drivers of the spatial distribution of mackerel

In the mid-2000s, the summer feeding distribution of Northeast Atlantic mackerel (Scomber scombrus) in Nordic Seas began expanding into new areas (Nøttestad et al., 2016). During the period 2007-2016 the mackerel distribution range increased three-fold and the centre-of-gravity shifted westward by 1650 km and northward by 400 km . Distribution range peaked in 2014 and was positively correlated to Spawning Stock Biomass (SSB) (ICES 2020). During this period mackerel stock expansion during the feeding season in summer increased from $1.3 \mathrm{mill} \mathrm{km}^{2}$ in 2007 to at least $2.9 \mathrm{mill} \mathrm{km}^{2}$ in 2014, mainly towards western and northern regions of the Nordic seas (Nøttestad et al., 2016). The distribution area was stable around 2.8-2.9 mill $\mathrm{km}^{2}$ during 20172019 (Nøttestad et al., 2017; 2019; ICES, 2018a). However, we witnessed a substantial shift in mackerel concentrations and distribution during summers of 2020-2021, when no mackerel were registered in Greenland waters, and a substantial decline was documented in Icelandic waters, whereas increased biomasses of mackerel were distributed in the central and northern part of the Norwegian Sea (Nøttestad et al., 2020b; WD09 in ICES 2021b). Overall, we have witnessed that mackerel had a much more eastern distribution in 2018-2022 compared to 2014-2017 (ICES, 2018a; Nøttestad et al., 2019; 2020b; 2021). Most of the surveyed mackerel still appears to be in the Norwegian Sea, but were more westerly distributed in 2022 than in the last 2 years. The survey coverage area was 2.9 million $\mathrm{km}^{2}$ in 2022, which is $32 \%$ larger coverage compared to 2021. Survey coverage was increased in the western areas (Iceland and Greenland waters) compared to in 2021. Furthermore, 0.28 million $\mathrm{km}^{2}$ was surveyed in the North Sea in July 2022.

Ólafsdóttir et al. (2018) modelled (GAM) IESSNS data (2007-2016) and found that mackerel was present in temperatures ranging from $5^{\circ} \mathrm{C}$ to $15^{\circ} \mathrm{C}$, but preferred areas between $9^{\circ} \mathrm{C}$ and $13^{\circ} \mathrm{C}$. The model showed that both mackerel occurrence and density were positively related to location, temperature, meso-zooplankton density and SSB. Thus, geographical expansion of mackerel during the summer feeding season in Nordic Seas was driven by increasing mackerel stock size and constrained by availability of preferred temperature and abundance of meso-zooplankton. However, these results are limited by time-series length (1997-2016; Olafsdottir et al., 2019). Notably, this seems to have changed during the most recent period from 2019 and onwards (e.g. high mackerel concentrations in 2020 at lower temperatures of $7-8{ }^{\circ} \mathrm{C}$; Nøttestad et al., 2019; 2020b; WD09 in ICES 2021b). It is not clear what causes this distributional shift, but the SST were $1-2^{\circ} \mathrm{C}$ lower in the western and south-western areas as compared to a 20-years mean (1999-2009), and substantially lower zooplankton concentrations in Icelandic and Greenland waters in 2019 and 2020 might partly explain such changes (ICES, 2018a; Nøttestad et al., 2019; 2020a). Marine climate with multi-decadal variability might also have affected the observed distributional changes but were not evaluated.

## Trophic interactions

There are strong indications for interspecific competition for food between mackerel, NSS-herring and blue whiting (Huse et al., 2012), where the competition between mackerel and herring being the best studied relationship. Both higher stomach fullness and prey shift for mackerel compared to herring during low stock size periods indicates that herring may suffer from this competition. Thus, an opportunistic (i.e. rapid shift in diet) and more generalist diet (i.e. wider range of prey) may be advantageous for mackerel in periods with low zooplankton abundances (Langøy et al. 2012; Debes et al. 2012; Óskarsson et al. 2015; Bachiller et al. 2016). Feeding activity seem to be highest in areas associated with colder water masses (Bachiller et al., 2016), and bioenergetics indicate that mackerel consumption may be as high as both herring and blue whiting in some years (122-135 mill t year ${ }^{-1}$, Bachiller et al. 2018). Distribution overlap between mackerel
and NSS herring during the summer feeding season is generally highest in the south-western part of the Norwegian Sea (Faroe and east Icelandic area) (Nøttestad et al., 2016; 2017; Ólafsdóttir et al., 2017). This spatiotemporal overlap between mackerel and herring have been present from 2016-2019 (ICES, 2018a, Nøttestad et al., 2016; 2017; 2019). In addition, increasing distribution overlaps in the north-western parts of the Norwegian Sea have also been observed since 2019 and onwards, which is in contrast to previous years (Nøttestad et al., 2019; 2020; WD09 in ICES 2021b). Overlapping distributions of mackerel and Norwegian spring-spawning herring (NSSH) were particularly present in the western and north-western part of the Norwegian Sea in 2022.
Recently, a number of predators have been highlighted as potential sources of mortality for mackerel. Although limited spatial overlap between marine mammals and mackerel during summers in the Nordic Seas (Nøttestad et al., 2019; Løviknes, 2019), orcas have been observed to actively search and hunt for mackerel schools (Nøttestad et al., 2014; Nøttestad et al., 2020a; 2021). Furthermore, the increases of 0 - and 1-groups mackerel found along major coastlines of Norway (2016-2018, Nøttestad et al., 2018; Bjørdal, 2019) have coincided with predation by increasing numbers of adult Atlantic bluefin tuna (Thynnus thunnus, Boge, 2019; Nøttestad et al., 2020b). Additionally, stomach samples from several species document that smaller sized mackerel is now eaten by different predators in northern waters (e.g. cod, saithe, marine mammals and seabirds; Bjørdal, 2019). Although, fewer 1-groups have been observed in coastal Norway waters in recent years (2019-2022, IESSNS; Nøttestad et al., 2019; 2020b; 2021; 2022) predation by the Atlantic bluefin tuna is still evident. The predation pressure and associated mortality from various predators on NEA mackerel (both juveniles and adults) are still unknown, but could have ecological impact in both time (i.e. population) and space (i.e. local and regional) (ICCAT, 2019; Nøttestad et al., 2020b).

### 8.13 References

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Table 8.4.1.1. NE Atlantic MackereI. ICES estimated catches by area (t). Discards not estimated prior to 1978 (data submitted by Working Group members).

| Year | Subarea 6 |  |  | Subarea 7 and Divisions 8.abde |  |  | Subareas 3 and 4 |  |  | Subareas 125 and 14 |  | Divisions 8.c and 9.a |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg Disc | Catch | Ldg Disc | Catch | Ldg | Disc | Catch |
| 1969 | 4800 |  | 4800 | 47404 |  | 47404 | 739175 |  | 739175 | 7 | 7 | 42526 | 42526 | 833912 |  | 833912 |
| 1970 | 3900 |  | 3900 | 72822 |  | 72822 | 322451 |  | 322451 | 163 | 163 | 70172 | 70172 | 469508 |  | 469508 |
| 1971 | 10200 |  | 10200 | 89745 |  | 89745 | 243673 |  | 243673 | 358 | 358 | 32942 | 32942 | 376918 |  | 376918 |
| 1972 | 13000 |  | 13000 | 130280 |  | 130280 | 188599 |  | 188599 | 88 | 88 | 29262 | 29262 | 361229 |  | 361229 |
| 1973 | 52200 |  | 52200 | 144807 |  | 144807 | 326519 |  | 326519 | 21600 | 21600 | 25967 | 25967 | 571093 |  | 571093 |
| 1974 | 64100 |  | 64100 | 207665 |  | 207665 | 298391 |  | 298391 | 6800 | 6800 | 30630 | 30630 | 607586 |  | 607586 |
| 1975 | 64800 |  | 64800 | 395995 |  | 395995 | 263062 |  | 263062 | 34700 | 34700 | 25457 | 25457 | 784014 |  | 784014 |
| 1976 | 67800 |  | 67800 | 420920 |  | 420920 | 305709 |  | 305709 | 10500 | 10500 | 23306 | 23306 | 828235 |  | 828235 |
| 1977 | 74800 |  | 74800 | 259100 |  | 259100 | 259531 |  | 259531 | 1400 | 1400 | 25416 | 25416 | 620247 |  | 620247 |
| 1978 | 151700 | 15100 | 166800 | 355500 | 35500 | 391000 | 148817 |  | 148817 | 4200 | 4200 | 25909 | 25909 | 686126 | 50600 | 736726 |
| 1979 | 203300 | 20300 | 223600 | 398000 | 39800 | 437800 | 152323 | 500 | 152823 | 7000 | 7000 | 21932 | 21932 | 782555 | 60600 | 843155 |
| 1980 | 218700 | 6000 | 224700 | 386100 | 15600 | 401700 | 87931 |  | 87931 | 8300 | 8300 | 12280 | 12280 | 713311 | 21600 | 734911 |
| 1981 | 335100 | 2500 | 337600 | 274300 | 39800 | 314100 | 64172 | 3216 | 67388 | 18700 | 18700 | 16688 | 16688 | 708960 | 45516 | 754476 |
| 1982 | 340400 | 4100 | 344500 | 257800 | 20800 | 278600 | 35033 | 450 | 35483 | 37600 | 37600 | 21076 | 21076 | 691909 | 25350 | 717259 |
| 1983 | 320500 | 2300 | 322800 | 235000 | 9000 | 244000 | 40889 | 96 | 40985 | 49000 | 49000 | 14853 | 14853 | 660242 | 11396 | 671638 |
| 1984 | 306100 | 1600 | 307700 | 161400 | 10500 | 171900 | 43696 | 202 | 43898 | 98222 | 98222 | 20208 | 20208 | 629626 | 12302 | 641928 |


| Year | Subarea 6 |  |  | Subarea 7 and Divisions 8.abde |  |  | Subareas 3 and 4 |  |  | Subareas 125 and 14 |  | Divisions 8.c and 9.a |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg Disc | Catch | Ldg Disc | Catch | Ldg | Disc | Catch |
| 1985 | 388140 | 2735 | 390875 | 75043 | 1800 | 76843 | 46790 | 3656 | 50446 | 78000 | 78000 | 18111 | 18111 | 606084 | 8191 | 614275 |
| 1986 | 104100 |  | 104100 | 128499 |  | 128499 | 236309 | 7431 | 243740 | 101000 | 101000 | 24789 | 24789 | 594697 | 7431 | 602128 |
| 1987 | 183700 |  | 183700 | 100300 |  | 100300 | 290829 | 10789 | 301618 | 47000 | 47000 | 22187 | 22187 | 644016 | 10789 | 654805 |
| 1988 | 115600 | 3100 | 118700 | 75600 | 2700 | 78300 | 308550 | 29766 | 338316 | 120404 | 120404 | 24772 | 24772 | 644926 | 35566 | 680492 |
| 1989 | 121300 | 2600 | 123900 | 72900 | 2300 | 75200 | 279410 | 2190 | 281600 | 90488 | 90488 | 18321 | 18321 | 582419 | 7090 | 589509 |
| 1990 | 114800 | 5800 | 120600 | 56300 | 5500 | 61800 | 300800 | 4300 | 305100 | 118700 | 118700 | 21311 | 21311 | 611911 | 15600 | 627511 |
| 1991 | 109500 | 10700 | 120200 | 50500 | 12800 | 63300 | 358700 | 7200 | 365900 | 97800 | 97800 | 20683 | 20683 | 637183 | 30700 | 667883 |
| 1992 | 141906 | 9620 | 151526 | 72153 | 12400 | 84553 | 364184 | 2980 | 367164 | 139062 | 139062 | 18046 | 18046 | 735351 | 25000 | 760351 |
| 1993 | 133497 | 2670 | 136167 | 99828 | 12790 | 112618 | 387838 | 2720 | 390558 | 165973 | 165973 | 19720 | 19720 | 806856 | 18180 | 825036 |
| 1994 | 134338 | 1390 | 135728 | 113088 | 2830 | 115918 | 471247 | 1150 | 472397 | 72309 | 72309 | 25043 | 25043 | 816025 | 5370 | 821395 |
| 1995 | 145626 | 74 | 145700 | 117883 | 6917 | 124800 | 321474 | 730 | 322204 | 135496 | 135496 | 27600 | 27600 | 748079 | 7721 | 755800 |
| 1996 | 129895 | 255 | 130150 | 73351 | 9773 | 83124 | 211451 | 1387 | 212838 | 103376 | 103376 | 34123 | 34123 | 552196 | 11415 | 563611 |
| 1997 | 65044 | 2240 | 67284 | 114719 | 13817 | 128536 | 226680 | 2807 | 229487 | 103598 | 103598 | 40708 | 40708 | 550749 | 18864 | 569613 |
| 1998 | 110141 | 71 | 110212 | 105181 | 3206 | 108387 | 264947 | 4735 | 269682 | 134219 | 134219 | 44164 | 44164 | 658652 | 8012 | 666664 |
| 1999 | 116362 |  | 116362 | 94290 |  | 94290 | 313014 |  | 313014 | 72848 | 72848 | 43796 | 43796 | 640311 |  | 640311 |
| 2000 | 187595 | 1 | 187595 | 115566 | 1918 | 117484 | 285567 | 165 | 304898 | 92557 | 92557 | 36074 | 36074 | 736524 | 2084 | 738608 |
| 2001 | 143142 | 83 | 143142 | 142890 | 1081 | 143971 | 327200 | 24 | 339971 | 67097 | 67097 | 43198 | 43198 | 736274 | 1188 | 737462 |


| Year | Subarea 6 |  |  | Subarea 7 and Divisions 8.abde |  |  | Subareas 3 and 4 |  |  | Subareas 125 and 14 |  |  | Divisions 8.c and 9.a |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch |
| 2002 | 136847 | 12931 | 149778 | 102484 | 2260 | 104744 | 375708 | 8583 | 394878 | 73929 |  | 73929 | 49576 |  | 49576 | 749131 | 23774 | 772905 |
| 2003 | 135690 | 1399 | 137089 | 90356 | 5712 | 96068 | 354109 | 11785 | 365894 | 53883 |  | 53883 | 25823 | 531 | 26354 | 659831 | 19427 | 679288 |
| 2004 | 134033 | 1705 | 134738 | 103703 | 5991 | 109694 | 306040 | 11329 | 317369 | 62913 | 9 | 62922 | 34840 | 928 | 35769 | 640529 | 19962 | 660491 |
| 2005 | 79960 | 8201 | 88162 | 90278 | 12158 | 102436 | 249741 | 4633 | 254374 | 54129 |  | 54129 | 49618 | 796 | 50414 | 523726 | 25788 | 549514 |
| 2006 | 88077 | 6081 | 94158 | 66209 | 8642 | 74851 | 200929 | 8263 | 209192 | 46716 |  | 46716 | 52751 | 3607 | 56358 | 454587 | 26594 | 481181 |
| 2007 | 110788 | 2450 | 113238 | 71235 | 7727 | 78962 | 253013 | 4195 | 257208 | 72891 |  | 72891 | 62834 | 1072 | 63906 | 570762 | 15444 | 586206 |
| 2008 | 76358 | 21889 | 98247 | 73954 | 5462 | 79416 | 227252 | 8862 | 236113 | 148669 | 112 | 148781 | 59859 | 750 | 60609 | 586090 | 37075 | 623165 |
| 2009 | 135468 | 3927 | 139395 | 88287 | 2921 | 91208 | 226928 | 8120 | 235049 | 163604 |  | 163604 | 107747 | 966 | 108713 | 722035 | 15934 | 737969 |
| 2010 | 106732 | 2904 | 109636 | 104128 | 4614 | 108741 | 246818 | 883 | 247700 | 355725 | 5 | 355729 | 50826 | 4640 | 55466 | 864229 | 13045 | 877272 |
| 2011 | 160756 | 1836 | 162592 | 51098 | 5317 | 56415 | 301746 | 1906 | 303652 | 398132 | 28 | 398160 | 26337 | 1807 | 28144 | 938070 | 10894 | 948963 |
| 2012 | 121115 | 952 | 122067 | 65728 | 9701 | 75429 | 218400 | 1089 | 219489 | 449325 | 1 | 449326 | 29809 | 3431 | 33240 | 884377 | 15174 | 899551 |
| 2013 | 132062 | 273 | 132335 | 49871 | 1652 | 51523 | 260921 | 337 | 261258 | 465846 | 15 | 465861 | 24867 | 2455 | 27322 | 933567 | 4732 | 938299 |
| 2014 | 180068 | 340 | 180408 | 93709 | 1402 | 95111 | 383887 | 334 | 384221 | 684082 | 91 | 684173 | 53591 | 4284 | 57875 | 1395337 | 6451 | 1401788 |
| 2015 | 134728 | 30 | 134757 | 98563 | 3155 | 101718 | 295877 | 34 | 295911 | 632493 | 78 | 632571 | 43735 | 7133 | 50869 | 1205396 | 10431 | 1215827 |
| 2016 | 206326 | 200 | 206526 | 37300 | 1927 | 39227 | 248041 | 570 | 248611 | 563440 | 54 | 563494 | 39056 | 3220 | 42276 | 1094163 | 5971 | 1100135 |
| 2017 | 225959 | 151 | 226110 | 21128 | 1992 | 23119 | 269404 | 400 | 269804 | 603806 | 62 | 603869 | 36512 | 227 | 36739 | 1156809 | 2832 | 1159641 |
| 2018 | 157239 | 90 | 157329 | 32037 | 1611 | 33649 | 341527 | 620 | 342147 | 455689 | 51 | 455740 | 33761 | 518 | 34279 | 1020254 | 2890 | 1023144 |
| 2019 | 122995 | 144 | 123139 | 32840 | 5902 | 38742 | 307235 | 812 | 308047 | 345019 | 18 | 345037 | 23832 | 931 | 24763 | 831920 | 7807 | 839727 |
| 2020 | 130577 | 341 | 130918 | 48806 | 8065 | 56871 | 456479 | 732 | 457211 | 356985 |  | 356985 | 37386 | 143 | 37529 | 1030233 | 9280 | 1039513 |
| 2021 | 146519 | 117 | 146635 | 15901 | 2524 | 18425 | 221019 | 423 | 221442 | 663111 |  | 663111 | 31862 | 65 | 31928 | 1078411 | 3129 | 1081540 |

Table 8.4.2.1. NE Atlantic Mackerel. ICES estimated catch ( t ) in Subareas $1,2,5$ and 14, 2000-2021 (Data submitted by Working Group members).

| Year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 1375 | 7 | 1 |  |  |  |  |  |  |  | 4845 |
| Estonia | 2673 | 219 |  |  |  |  |  |  |  |  |  |
| Faroe Islands | 5546 | 3272 | 4730 |  | 650 | 30 |  | 278 | 123 | 2992 | 66312 |
| France |  |  |  |  | 2 | 1 |  |  |  |  |  |
| Germany |  |  |  |  |  |  |  | 7 |  |  |  |
| Greenland |  |  |  |  |  |  |  |  |  |  |  |
| Iceland |  |  | 53 | 122 |  | 363 | 4222 | 36706 | 112286 | 116160 | 121008 |
| Ireland |  |  |  | 495 | 471 |  |  |  |  |  |  |
| Latvia |  |  |  |  |  |  |  |  |  |  |  |
| Lithuania | 2085 |  |  |  |  |  |  |  |  |  |  |
| Netherlands |  |  | 569 | 44 | 34 | 2393 |  | 10 | 72 |  | 90 |
| Norway | 31778 | 21971 | 22670 | 125481 | 10295 | 13244 | 8914 | 493 | 3474 | 3038 | 104858 |
| Poland |  |  |  |  |  |  |  |  |  |  |  |
| Sweden |  | 8 |  |  |  |  |  |  |  |  |  |
| United Kingdom |  | 54 | 665 | 692 | 2493 |  |  |  | 4 |  |  |
| Russia | 49101 | 41566 | 45811 | 40026 | 49489 | 40491 | 33580 | 35408 | 32728 | 414141 | 58613 |
| Misreported |  |  | -570 |  | -553 |  |  |  |  |  |  |
| Unallocated |  |  |  | -44 | 32 | -2393 |  | -10 | -18 |  |  |


| Year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Discards |  |  |  |  | 9 |  |  |  | 112 |  | 5 |
| Total | 92557 | 67097 | 73929 | 53883 | 62922 | 54129 | 46716 | 72891 | 148781 | 163604 | 355729 |
| Year | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| Denmark | 269 |  | 391 | 2345 | 4321 | 1 | 2 | 289 |  |  | 0.691 |
| Estonia |  |  | 13671 |  |  |  |  |  |  |  |  |
| Faroe Islands | 121499 | 107198 | 142976 | 103896 | 76889 | 61901 | 66194 | 52061 | 37418 | 33291 | 105096 |
| France | 2 |  | 197 | 8 | 36 |  |  | 733 |  | 8 | 0.2 |
| Germany |  | 107 | 74 |  | 2963 | 3499 | 4064 | 577 | 190 | 206 | 9 |
| Greenland | 621 | 74021 | 541481 | 875811 | 30351 | 36142 | 46388 | 62973 | 30241 | 26555 | 33360 |
| Iceland | 159263 | 149282 | 151103 | 172960 | 169333 | 170374 | 167366 | 168330 | 128008 | 151534 | 132109 |
| Ireland | 90 |  |  | 1725 | 6 | 2 |  |  |  |  |  |
| Latvia |  |  |  |  |  |  |  |  |  |  |  |
| Lithuania |  |  |  | 1082 |  | 1931 |  |  |  | 2 |  |
| Netherlands | 178 | 5 | 1 | 5887 | 6996 | 8599 | 7671 | 2697 | 13 | 0.73 |  |
| Norway | 43168 | 110741 | 33817 | 192322 | 204574 | 153228 | 167739 | 46853 | 22605 | 15937 | 256124 |
| Poland |  |  |  |  |  |  |  | 2 |  | 0.044 | 8.2 |
| Sweden |  | 4 | 825 | 3310 | 740 | 730 | 1720 | 910 |  | 220 | 228 |
| United Kingdom |  |  | 2 | 5534 | 7851 | 5240 | 4601 | 2009 |  | 426 |  |


| Year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Russia | 73601 | 74587 | 80812 | 116433 | 128433 | 121614 | 138061 | 118255 | 126543 | 128805 | 136176 |
| Misreported |  |  |  |  |  |  |  |  |  |  |  |
| Unallocated |  |  |  |  |  |  |  |  |  |  |  |
| Discards | 28 | 1 | 151 | 911 | 78 | 54 | 62 | 51 | 18 | 0.05 |  |
| Total | 398160 | 449326 | 465729 | 684173 | 632571 | 563315 | 603869 | 455740 | 345036 | 356985 | 663111 |

Table 8.4.2.2. NE Atlantic Mackerel. ICES estimated catch (t) in the North Sea, Skagerrak and Kattegat (Subarea 4 and Division 3.a), 2000-2021 (Data submitted by Working Group members).

| Year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 146 | 97 | 22 | 2 | 4 | 1 | 3 | 1 | 2 | 3 | 27 |
| Denmark | 27720 | 21680 | 343751 | 275081 | 25665 | 232121 | 242191 | 252171 | 26716 | 23491 | 36552 |
| Faroe Islands | 10614 | 18751 | 12548 | 11754 | 11705 | 9739 | 12008 | 11818 | 7627 | 6648 | 4639 |
| France | 1588 | 1981 | 2152 | 1467 | 1538 | 1004 | 285 | 7549 | 490 | 1493 | 686 |
| Germany Fed. Rep. | 78 | 4514 | 3902 | 4859 | 4515 | 4442 | 2389 | 5383 | 4668 | 5158 | 25621 |
| Iceland |  |  |  |  |  |  |  |  |  |  |  |
| Ireland | 9956 | 10284 | 20715 | 17145 | 18901 | 15605 | 4125 | 13337 | 11628 | 12901 | 14639 |
| Lithuania |  |  |  |  |  |  |  |  |  |  |  |
| Netherlands | 2262 | 2441 | 11044 | 6784 | 6366 | 3915 | 4093 | 5973 | 1980 | 2039 | 1300 |
| Norway | 142320 | 158401 | 161621 | 150858 | 147068 | 106434 | 113079 | 131191 | 114102 | 118070 | 129064 |
| Poland |  |  |  |  |  | 109 |  |  |  |  |  |


| Year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sweden | 49941 | 5090 | 52321 | 4450 | 4437 | 3204 | 3209 | 38581 | 36641 | 73031 | 34291 |
| United Kingdom | 58282 | 52988 | 61781 | 67083 | 62932 | 37118 | 28628 | 46264 | 37055 | 47863 | 52563 |
| Russia | 1672 | 1 |  |  |  | 4 |  |  |  |  | 696 |
| Misreported (Area 6.a) | 8591 | 39024 | 49918 | 62928 | 23692 | 37911 | 8719 |  | 17280 | 1959 |  |
| Unallocated | 34761 | 24873 | 22985 | -730 | -783 | 7043 | 171 | 2421 | 2039 | -629 | 660 |
| Discards | 1912 | 24 | 8583 | 11785 | 11329 | 4633 | 8263 | 4195 | 8862 | 8120 | 883 |
| Total | 304896 | 339970 | 394878 | 365894 | 317369 | 254374 | 209192 | 257208 | 236111 | 235049 | 247700 |
| Year | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| Belgium | 21 | 39 | 62 | 56 | 38 | 99 | 107 | 110 | 13 | 75 | 77 |
| Denmark | 32800 | 36492 | 31924 | 21340 | 35809 | 21696 | 27457 | 22207 | 25374 | 34375 | 28295 |
| Faroe Islands | 543 | 432 | 25 | 42919 | 25672 | 18193 | 12915 | 15475 | 17460 | 32860 |  |
| France | 1416 | 5736 | 1788 | 4912 | 7827 | 3448 | 5942 | 6714 | 5455 | 8959 | 5041 |
| Germany Fed. Rep. | 52911 | 4560 | 5755 | 4979 | 6056 | 10172 | 11185 | 12091 | 7778 | 15946 | 9939 |
| Iceland |  |  |  |  |  |  |  |  |  |  |  |
| Ireland | 15810 | 20422 | 13523 | 45167 | 34167 | 24437 | 35957 | 24567 | 1678 | 15395 | 11021 |
| Lithuania |  |  |  | 8340 |  | 596 |  |  |  | 813 | 6655 |
| Netherlands | 9881 | 6018 | 4863 | 24536 | 17547 | 11434 | 17401 | 13844 | 8957 | 18425 | 15983 |
| Norway | 162878 | 64181 | 130056 | 85409 | 36344 | 55089 | 51960 | 135715 | 135083 | 195515 | 14518 |


| Year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Poland |  |  |  |  | 24 |  | 0.721 | 4041 | 1394 | 16 | 559 |
| Sweden | 32481 | 4560 | 2081 | 1112 | 3190 | 2933 | 1981 | 3056 | 2152 | 3451 | 3277 |
| United Kingdom | 69858 | 75959 | 70840 | 145119 | 129203 | 99945 | 104499 | 103707 | 101890 | 130650 | 125553 |
| Russia |  |  | 4 |  |  |  |  |  | 0.12 |  |  |
| Misreported (Area 6.a) |  |  |  |  |  |  |  |  |  |  |  |
| Unallocated |  |  |  |  |  |  |  |  |  |  |  |
| Discards | 1906 | 1089 | 337 | 334 | 34 | 559 | 400 | 620 | 812 | 732 | 423 |
| Total | 303652 | 219489 | 261258 | 384221 | 295911 | 248611 | 269804 | 342147 | 308047 | 457211 | 221340 |

Table 8.4.2.3. NE Atlantic MackereI. ICES estimated catch ( t ) in the Western area (Subareas 6 and 7 and Divisions 8.a,b,d,e), 2000-2021 (Data submitted by Working Group members).

| Country | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium |  |  |  |  | 1 |  |  |  |  | 1 | 2 |
| Denmark | 82 | 835 |  | 113 |  |  |  | 6 | 10 |  | 48 |
| Estonia |  |  |  |  |  |  |  |  |  |  |  |
| Faroe Islands | 4863 | 2161 | 2490 | 2260 | 674 |  | 59 | 1333 | 3539 | 4421 | 36 |
| France | 17857 | 18975 | 19726 | 21213 | 18549 | 15182 | 14625 | 12434 | 14944 | 16464 | 10301 |
| Germany | 22901 | 20793 | 22630 | 19200 | 18730 | 14598 | 14219 | 12831 | 10834 | 17545 | 16493 |
| Guernsey |  |  |  |  |  |  | 10 |  |  |  |  |
| Ireland | 61277 | 60168 | 51457 | 49715 | 41730 | 30082 | 36539 | 35923 | 33132 | 48155 | 43355 |



| Faroe Islands | 8 |  |  | 3421 | 5851 | 13173 | 20559 | 13543 | 7787 | 2913 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| France | 11304 | 14448 | 12438 | 16627 | 17820 | 16634 | 16925 | 13974 | 12371 | 12816 | 11308 |
| Germany | 18792 | 14277 | 15102 | 23478 | 19238 | 9740 | 9608 | 7214 | 8936 | 8878 | 2049 |
| Greenland |  |  |  |  |  |  |  |  |  | 22 |  |
| Guernsey | 10 | 5 | 9 | 9 | 4 |  |  | 12 | 9 |  |  |
| Iceland |  |  |  |  |  |  |  |  | 69 |  |  |
| Ireland | 45696 | 42627 | 42988 | 56286 | 54571 | 52087 | 48957 | 42181 | 51635 | 58720 | 49731 |
| Isle of Man | 11 | 11 | 8 | 3 |  | 8 | 2 | 3 | 3 | 2 |  |
| Jersey | 7 | 8 | 8 | 7 | 3 | 3 | 0.003 | 3 | 2 | 5 |  |
| Lithuania | 23 |  |  | 176 | 554 | 13 |  |  |  |  |  |
| Netherlands | 18336 | 19794 | 16295 | 16242 | 15264 | 17896 | 18694 | 13851 | 13727 | 11895 | 8611 |
| Norway | 2019 | 1101 | 734 |  | 1313 | 1035 | 2657 | 4639 | 1420 | 221 | 11 |
| Poland |  |  |  |  |  |  |  | 14 | 2312 | 5286 | 1155 |
| Portugal |  |  |  |  |  |  |  |  | 46 | 35 | 32 |
| Russia |  |  |  |  |  | 30 |  |  | 1 | 10 |  |
| Spain | 1257 | 773 | 635 | 1796 | 951 | 1253 | 786 | 4471 | 1220 | 1784 | 704 |
| Sweden |  |  |  |  |  |  |  |  | 805 |  |  |
| United Kingdom | 111103 | 93775 | 92957 | 137195 | 110932 | 112268 | 116308 | 84309 | 50253 | 72637 | 84323 |


| Unallocated | 399 | 16 | -144 |  | 34 |  | 13 |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Discards | 7153 | 10654 | 2105 | 1742 | 3185 | 2126 | 2142 | 1701 | 6046 | 8405 | 2640 |
| Total | 219007 | 197496 | 183857 | 275519 | 236475 | 245754 | 249229 | 194180 | 161883 | 187788 | 165060 |

Table 8.4.2.4. NE Atlantic Mackerel. ICES estimated catch (t) in Divisions 8.c and 9.a, 2000-2021 (Data submitted by Working Group members). 9.b is included in 2020.

| Country | Div | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| France | 8.c | 177 | 151 | 43 | 55 | 168 | 383 | 392 | 44 | 283 |
| Portugal | 8.c |  |  |  |  |  |  | 1758 | 2302 | 4867.9437 |
| Portugal | 9.a | 2289 | 1509 | 2620 | 2605 | 2381 | 1753 | 2363 | 962 | 824 |
| Spain | 8.c |  |  | 43063 | 53401 | 50455 | 91043 | 38858 | 14709 | 17768 |
| Spain | 9.a |  |  | 7025 | 6773 | 6855 | 14569 | 7347 | 2759 | 845 |
| Discards | 8.c | 928 | 391 | 3606 | 156 | 73 | 725 | 4408 | 563 | 2187 |
| Discards | 9.a |  | 405 | 1 | 916 | 677 | 241 | 232 | 1245 | 1244 |
| Unallocated | 8.c | 28429 | 42851 |  |  |  |  |  | 4691 | 4144 |
| Unallocated | 9.a | 3946 | 5107 |  |  |  |  | 108 | 871 | 1076 |
| Total | 9.a | 6234 | 7021 | 9646 | 10293 | 9913 | 16562 | 10049 | 5836 | 3989 |
| Total |  | 35768 | 50414 | 56358 | 63906 | 60609 | 108713 | 55466 | 28146 | 33239 |
| Country | Div | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| France | 8.c | 220 | 171 | 21 | 106 | 83 | 50 | 43 | 96 | 93 |
| Portugal | 8.c | 5134 | 7334 | 6836 | 6069 | 3697 | 3709 | 3188 | 4189 | 3738 |
| Portugal | 9.a | 254 | 618 | 1456 | 619 | 634 | 855 | 706 | 575 | 953 |
| Spain | 8.c | 14617 | 33783 | 29726 | 26553 | 30893 | 27190 | 19148 | 31143 | 25272 |
| Spain | 9.a | 1162 | 2227 | 3853 | 2229 | 1206 | 1656 | 747 | 1379 | 1807 |


| Russia |  |  |  |  |  |  |  |  | 2 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Discards | 8.c | 1428 | 2821 | 4724 | 2469 | 84 | 324 | 760 | 28 | 18 |
| Discards | 9.9 | 1027 | 1463 | 2409 | 751 | 143 | 194 | 172 | 115 | 47 |
| Unallocated | 8.c | -573 | 8795 | 11 | 1357 |  | 300 |  |  |  |
| Unallocated | $9 . \mathrm{a}$ | 4053 | 662 | 1831 | 2123 |  |  |  |  |  |
| Total | $9 . \mathrm{a}$ | 6497 | 4308 | 9550 | 5722 | 1983 | 2736 | 1625 | 2070 | 2807 |
| Total |  | 27322 | 57874 | 50867 | 42276 | 36740 | 34279 | 24764 | 37529 | 31928 |

Table 8.5.1.1. NE Atlantic Mackerel. Mean length (mm) -at-age by area for 2021 (Q1-Q4).

| Age | 1 | 2.a | 3.a | 3.b | 3.c | 3.d | $4 . \mathrm{a}$ | 4.b | 4.c |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 292 | 279 | 293 | 292 | 292 | 289 | 289 | 278 | 286 |
| 2 | 322 | 312 | 322 | 319 | 316 | 318 | 317 | 323 | 317 |
| 3 | 338 | 335 | 340 | 343 | 336 | 326 | 341 | 333 | 333 |
| 4 | 348 | 345 | 350 | 350 | 347 | 352 | 350 | 351 | 354 |
| 5 | 356 | 355 | 358 | 359 | 355 | 360 | 359 | 358 | 363 |
| 6 | 364 | 366 | 372 | 372 | 372 | 361 | 373 | 368 | 368 |
| 7 | 370 | 372 | 378 | 376 | 377 | 359 | 376 | 372 | 377 |
| 8 | 376 | 377 | 381 | 381 | 380 | 378 | 380 | 380 | 383 |
| 9 | 381 | 381 | 385 | 383 | 384 | 387 | 382 | 384 | 390 |
| 10 | 386 | 383 | 389 | 388 | 388 | 389 | 385 | 387 | 383 |
| 11 | 390 | 388 | 391 | 386 | 389 | 377 | 387 | 389 | 384 |
| 12 | 394 | 393 | 399 | 400 | 398 | 386 | 399 | 398 | 392 |
| 13 | 398 | 395 | 400 | 399 | 396 | 398 | 397 | 397 | 397 |
| 14 | 402 | 396 | 406 | 406 | 399 | 404 | 392 | 396 | 389 |
| 15 | 399 | 398 | 412 |  |  |  | 402 | 407 | 398 |


| Age | 5.a | 5.b | 6.a | 6.b | 7.a | 7.c | 7.h |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  | 289 | 209 | 207 | 231 |  |  |  |
| 1 | 302 | 341 | 342 | 324 | 327 | 97 | 352 | 354 |
| 2 | 348 | 360 | 342 | 346 | 128 | 350 | 342 | 302 |
| 3 | 354 | 353 | 353 | 358 | 172 | 356 | 359 | 341 |
| 5 | 370 | 372 | 366 | 371 | 143 | 364 | 384 | 347 |
| 7 | 376 | 378 | 372 | 373 | 132 | 379 | 396 | 373 |
| 8 | 381 | 379 | 380 | 381 | 162 | 390 | 391 | 383 |
| 9 | 381 | 377 | 382 | 383 | 256 | 385 | 400 | 387 |


| 10 | 380 | 389 | 385 | 386 | 367 | 389 | 399 | 390 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 11 | 386 | 386 | 391 | 391 | 274 | 390 | 409 | 400 |
| 12 | 390 | 364 | 394 | 393 | 409 | 405 | 399 | 393 |
| 13 | 388 |  | 398 | 399 | 358 | 401 | 395 | 395 |
| 14 | 392 | 399 | 404 | 404 | 410 | 410 | 405 |  |
| 15 | 396 |  | 399 | 418 | 404 | 435 | 435 | 415 |

Table 8.5.1.1. NE Atlantic Mackerel. Mean length (mm) -at-age by area for 2021 (Q1-Q4) continued.

| Age | 7.j | 8.1 | 8.b | $8 . \mathrm{C}$ | 8.d | 9.1 | 9.a.N | All |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  | 181 | 180 | 217 |  | 206 | 222 | 178 |
| 1 | 178 | 247 | 216 | 225 | 278 | 258 | 256 | 195 |
| 2 | 295 | 295 | 294 | 315 | 326 | 317 | 279 | 306 |
| 3 | 338 | 316 | 323 | 341 | 324 | 329 | 310 | 328 |
| 4 | 349 | 347 | 347 | 354 | 346 | 340 | 338 | 343 |
| 5 | 356 | 363 | 360 | 364 | 366 | 368 | 357 | 354 |
| 6 | 367 | 376 | 372 | 373 | 375 | 385 | 379 | 368 |
| 7 | 374 | 382 | 379 | 377 | 379 | 390 | 383 | 373 |
| 8 | 382 | 391 | 389 | 383 | 385 | 395 | 389 | 379 |
| 9 | 379 | 389 | 387 | 386 | 388 | 413 | 394 | 382 |
| 10 | 387 | 393 | 392 | 391 | 393 | 411 | 395 | 384 |
| 11 | 388 | 396 | 393 | 399 | 400 | 470 | 401 | 388 |
| 12 | 392 | 415 | 416 | 403 | 406 | 410 | 414 | 395 |
| 13 | 400 | 406 | 406 | 408 | 406 | 440 | 409 | 396 |
| 14 | 395 | 405 | 415 | 422 | 405 |  | 423 | 396 |
| 15 | 415 |  |  |  |  |  |  | 399 |

Table 8.6.1.2.1. Fecundity and atresia for the assessment years, from 1998 to 2022 (Preliminary values). $n$ is the number of samples used, $n / g$ refers to the number of oocytes or atretic oocytes by gram of fish (*) means median not mean relative for potential fecundity.


Table 8.6.1.6.1. Daily egg production estimate (stage la) for mackerel in the North Sea using the DEPM in 2021.

| Year | DEP ${ }^{* 10^{13}}$ | CV DEP |
| :---: | :---: | :---: |
| 2021 | 1.28 | $16 \%$ |

Table 8.6.1.6.2. Estimated adult parameters and SSB for mackerel in the North Sea using the DEPM in 2021

| Year | $\mathbf{2 0 2 1}$ |
| :--- | :--- |
| Batch fecundity | 18735 |
| Relative batch fecundity (N/g) | 42.7 |
| CV Batch fecundity | 0.87 |
| Spawning fraction | 0.18 |
| Sex ratio | 0.53 |
| Female weight (g) | 331.4 |

Table 8.7.1.1. NE Atlantic mackerel. Input data and parameters and the model configurations for the assessment.

| Input data types and characteristics: |  |  |  |
| :--- | :--- | :--- | :--- |
| Name | Year range | Age range | Variable from year to year |
| Catch in tonnes | $1980-2021$ | Yes |  |
| Catch-at-age in numbers | $1980-2021$ | $0-12+$ | Yes |


| Weight-at-age in the commercial catch | 1980-2021 | 0-12+ | Yes |  |
| :---: | :---: | :---: | :---: | :---: |
| Weight-at-age of the spawning stock at spawning time. | 1980-2021 | 0-12+ | Yes |  |
| Proportion of natural mortality before spawning | 1980-2021 | 0-12+ | Yes |  |
| Proportion of fishing mortality before spawning | 1980-2021 | 0-12+ | Yes |  |
| Proportion mature-at-age | 1980-2021 | 0-12+ | Yes |  |
| Natural mortality | 1980-2021 | 0-12+ | No, fixed at | . 15 |
| Tuning data: |  |  |  |  |
| Type | Name | Year range |  | Age range |
| Survey (SSB) | ICES Triennial Mackerel and Horse Mackerel Egg Survey | $\begin{aligned} & \text { 1992, 1995, 1998, 2001, 2004, 2007, } \\ & \text { 2010, 2013,2016,2019,2022. } \end{aligned}$ |  | Not applicable (gives SSB) |
| Survey <br> (abundance index) | IBTS Recruitment index (log transformed) | 1998-2020 |  | Age 0 |
| Survey <br> (abundance index) | International Ecosystem Summer Survey in the Nordic Seas (IESSNS) | 2010, 2012-2022 |  | Ages 3-11 |
| Tagging/recapture | Norwegian tagging program | Steel tags : 1980 (release year)-2006 <br> (recapture years) Ages 5 and older <br> (age at release) <br> RFID tags : 2013 (release year) 2021  |  |  |
| SAM parameter configuration : |  |  |  |  |
| Setting | Value | Description |  |  |
| Coupling of fishing mortality states | 1/2/3/4/5/6/7/8/8/8/8/8/8 | Different $F$ states for ages 0 to 6 , one same F state for ages 7 and older |  |  |
| Correlated random walks for the fishing mortalities | 0 | F random walk of different ages are independent |  |  |
| Coupling of catchability parameters | $\begin{aligned} & \text { 0/0/0/0/0/0/0/0/0/0/0/0/0 } \\ & \text { 1/0/0/0/0/0/0/0/0/0/0/0/0 } \\ & \text { 2/0/0/0/0/0/0/0/0/0/0/0/0 } \\ & \text { 0/0/0/3/4/5/6/7/8/9/10/10/0 } \end{aligned}$ |  | No catchability para <br> One catchability para the egg <br> One catchability para the recruitment index <br> One catchability para group estimated for to11) | eter for the catches meter estimated for meter estimated for meter for each age e IESSNS (age 3 |
| Power law model | 0 | No power law model used for any of the surveys |  |  |


| Coupling of fishing <br> mortality random walk <br> variances | $1 / 2 / 3 / 3 / 3 / 3 / 3 / 3 / 3 / 3 / 3 / 3 / 3$ | Separate F random walk variances for age <br> 0, age 1 and a same variance for older ages |
| :--- | :--- | :--- |
| Coupling of log abundance <br> random walk variances | $1 / 2 / 2 / 2 / 2 / 2 / 2 / 2 / 2 / 2 / 2 / 2 / 2$ | Same variance used for the log abundance <br> random walk of all ages except for the <br> recruits (age 0) |
| Coupling of the <br> observation variances | Separate observation variances for age 0 <br> and 1 than for the older ages in the <br> catches |  |
| $0 / 0 / 0 / 0 / 0 / 0 / 0 / 0 / 0 / 0 / 0 / 0 / 0$ | One observation variance for the egg <br> survey |  |
| O/0/0/0/0/0/0/0/0/0/0/0/0 | One observation variance for the <br> recruitment index |  |
| Stock recruitment model | 0 | 2 observation variances for the IESSNS (age <br> 3 and ages 4 and older) |
| Correlation structure | "ID", "ID", "ID", "AR" | No stock-recruiment model |

Table 8.7.1.2. NE Atlantic Mackerel. CATCH IN NUMBER



## Table 8.7.1.3. NE Atlantic Mackerel. WEIGHTS AT AGE IN THE CATCH

```
Units : Kg
    year
```

age $1980 \quad 1981 \quad 1982198319841985 \quad 198619871988198919901991$
$\begin{array}{llllllllllllllllll}0.057 & 0.060 & 0.053 & 0.050 & 0.031 & 0.055 & 0.039 & 0.076 & 0.055 & 0.049 & 0.085 & 0.068\end{array}$
$\begin{array}{llllllllllllll}0.131 & 0.132 & 0.131 & 0.168 & 0.102 & 0.144 & 0.146 & 0.179 & 0.133 & 0.136 & 0.156 & 0.156\end{array}$
$\begin{array}{lllllllllllll}0.249 & 0.248 & 0.249 & 0.219 & 0.184 & 0.262 & 0.245 & 0.223 & 0.259 & 0.237 & 0.233 & 0.253\end{array}$
$\begin{array}{llllllllllll}0.285 & 0.287 & 0.285 & 0.276 & 0.295 & 0.357 & 0.335 & 0.318 & 0.323 & 0.320 & 0.336 & 0.327\end{array}$
$\begin{array}{llllllllllllllllll}0.345 & 0.344 & 0.345 & 0.310 & 0.326 & 0.418 & 0.423 & 0.399 & 0.388 & 0.377 & 0.379 & 0.394\end{array}$
0.3780 .3770 .3780 .3860 .3440 .4170 .4710 .4740 .4560 .4330 .4230 .423
$\begin{array}{lllllllllllll}0.454 & 0.454 & 0.454 & 0.425 & 0.431 & 0.436 & 0.444 & 0.512 & 0.524 & 0.456 & 0.467 & 0.469\end{array}$
$\begin{array}{lllllllllllll}0.498 & 0.499 & 0.496 & 0.435 & 0.542 & 0.521 & 0.457 & 0.493 & 0.555 & 0.543 & 0.528 & 0.506\end{array}$
$\begin{array}{lllllllllllll}8 & 0.520 & 0.513 & 0.513 & 0.498 & 0.480 & 0.555 & 0.543 & 0.498 & 0.555 & 0.592 & 0.552 & 0.554\end{array}$
$\begin{array}{llllllllllllll}9 & 0.542 & 0.543 & 0.541 & 0.545 & 0.569 & 0.564 & 0.591 & 0.580 & 0.562 & 0.578 & 0.606 & 0.609\end{array}$
$\begin{array}{lllllllllllllllllll}10 & 0.574 & 0.573 & 0.574 & 0.606 & 0.628 & 0.629 & 0.552 & 0.634 & 0.613 & 0.581 & 0.606 & 0.630\end{array}$
$\begin{array}{lllllllllllllll}0.590 & 0.576 & 0.574 & 0.608 & 0.636 & 0.679 & 0.694 & 0.635 & 0.624 & 0.648 & 0.591 & 0.649\end{array}$
$\begin{array}{llllllllllllll}12 & 0.580 & 0.584 & 0.582 & 0.614 & 0.663 & 0.710 & 0.688 & 0.718 & 0.697 & 0.739 & 0.713 & 0.708\end{array}$
year
age $1992199319941995199619971998 \quad 1999 \quad 2000 \quad 2001 \quad 2002 \quad 2003$
$\begin{array}{llllllllllllllllllll}0.051 & 0.061 & 0.046 & 0.072 & 0.058 & 0.076 & 0.065 & 0.062 & 0.063 & 0.069 & 0.052 & 0.081\end{array}$
$\begin{array}{lllllllllllll}0.167 & 0.134 & 0.136 & 0.143 & 0.143 & 0.143 & 0.157 & 0.176 & 0.135 & 0.172 & 0.160 & 0.170\end{array}$
$\begin{array}{lllllllllllll}0.239 & 0.240 & 0.255 & 0.234 & 0.226 & 0.230 & 0.227 & 0.235 & 0.227 & 0.224 & 0.256 & 0.267\end{array}$
$\begin{array}{lllllllllll}0.333 & 0.317 & 0.339 & 0.333 & 0.313 & 0.295 & 0.310 & 0.306 & 0.306 & 0.305 & 0.307 \\ 0.336\end{array}$
$\begin{array}{llllllllllllllllllll}0.397 & 0.376 & 0.390 & 0.390 & 0.377 & 0.359 & 0.354 & 0.361 & 0.363 & 0.376 & 0.368 & 0.385\end{array}$
$\begin{array}{llllllllllllllll}0.460 & 0.436 & 0.448 & 0.452 & 0.425 & 0.415 & 0.408 & 0.404 & 0.427 & 0.424 & 0.424 & 0.438\end{array}$
$\begin{array}{lllllllllll}0.495 & 0.483 & 0.512 & 0.501 & 0.484 & 0.453 & 0.452 & 0.452 & 0.463 & 0.474 & 0.461\end{array} 0.477$
$\begin{array}{lllllllllll}0.532 & 0.527 & 0.543 & 0.539 & 0.518 & 0.481 & 0.462 & 0.500 & 0.501 & 0.496 & 0.512\end{array} 0.522$
$\begin{array}{lllllllllllll}0.555 & 0.548 & 0.590 & 0.577 & 0.551 & 0.524 & 0.518 & 0.536 & 0.534 & 0.540 & 0.536 & 0.572\end{array}$
$\begin{array}{lllllllllllllll}9 & 0.597 & 0.583 & 0.583 & 0.594 & 0.576 & 0.553 & 0.550 & 0.569 & 0.567 & 0.577 & 0.580 & 0.612\end{array}$
$\begin{array}{lllllllllllllllllll}10 & 0.651 & 0.595 & 0.627 & 0.606 & 0.596 & 0.577 & 0.573 & 0.586 & 0.586 & 0.603 & 0.600 & 0.631\end{array}$
$\begin{array}{lllllllllllll}11 & 0.663 & 0.647 & 0.678 & 0.631 & 0.603 & 0.591 & 0.591 & 0.607 & 0.594 & 0.611 & 0.629 & 0.648\end{array}$
$\begin{array}{lllllllllllll}12 & 0.669 & 0.679 & 0.713 & 0.672 & 0.670 & 0.636 & 0.631 & 0.687 & 0.644 & 0.666 & 0.665 & 0.715\end{array}$
year
age 20042005 2006 2007 2008 2009 2010 $2011 \quad 2012 \quad 2013 \quad 20142015$
$\begin{array}{llllllllllllllllll}0 & 0.067 & 0.048 & 0.038 & 0.089 & 0.051 & 0.104 & 0.048 & 0.029 & 0.089 & 0.091 & 0.043 & 0.051\end{array}$
$\begin{array}{lllllllllllllll}1 & 0.156 & 0.151 & 0.071 & 0.120 & 0.105 & 0.153 & 0.118 & 0.113 & 0.123 & 0.173 & 0.126 & 0.154\end{array}$
$\begin{array}{lllllllllllllll}2 & 0.263 & 0.268 & 0.197 & 0.215 & 0.222 & 0.213 & 0.221 & 0.231 & 0.186 & 0.234 & 0.231 & 0.242\end{array}$
$\begin{array}{llllllllllllllll}3 & 0.323 & 0.306 & 0.307 & 0.292 & 0.292 & 0.283 & 0.291 & 0.282 & 0.284 & 0.277 & 0.282 & 0.294\end{array}$
$40.400 \quad 0.3660 .357 \quad 0.3720 .3700 .3310 .3310 .3340 .3400 .3360 .3240 .320$
$\begin{array}{lllllllllllllll}5 & 0.419 & 0.434 & 0.428 & 0.408 & 0.418 & 0.389 & 0.365 & 0.368 & 0.374 & 0.360 & 0.362 & 0.351\end{array}$
$\begin{array}{lllllllllllllllll}6 & 0.485 & 0.440 & 0.479 & 0.456 & 0.444 & 0.424 & 0.418 & 0.411 & 0.401 & 0.386 & 0.394 & 0.392\end{array}$
$\begin{array}{lllllllllllll}0.519 & 0.496 & 0.494 & 0.512 & 0.497 & 0.450 & 0.470 & 0.451 & 0.431 & 0.405 & 0.422 & 0.420\end{array}$
$\begin{array}{lllllllllllllll}0.554 & 0.539 & 0.543 & 0.534 & 0.551 & 0.497 & 0.487 & 0.494 & 0.469 & 0.431 & 0.443 & 0.443\end{array}$
$9 \quad 0.5730 .5560 .5840 .5730 .5710 .5380 .5150 .5400 .5030 .4540 .4670 .465$
$\begin{array}{llllllllllllll}10 & 0.595 & 0.583 & 0.625 & 0.571 & 0.620 & 0.586 & 0.573 & 0.580 & 0.537 & 0.472 & 0.482 & 0.489\end{array}$
$\begin{array}{lllllllllllll}11 & 0.630 & 0.632 & 0.636 & 0.585 & 0.595 & 0.599 & 0.603 & 0.611 & 0.537 & 0.493 & 0.523 & 0.522\end{array}$
$\begin{array}{llllllllllllll}12 & 0.684 & 0.655 & 0.689 & 0.666 & 0.662 & 0.630 & 0.630 & 0.664 & 0.585 & 0.554 & 0.589 & 0.561\end{array}$

[^8]Table 8.7.1.4. NE Atlantic Mackerel. WEIGHTS AT AGE IN THE STOCK

```
Units : Kg
    year
```



```
    0 0.063 0.063 0.063 0.063 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    1
    2
    3
    40.322 0.312 0.300 0.282 0.293 0.308 0.301 0.318 0.298}00.272 0.290 0.303
    5}00.3560.335 0.368 0.341 0.326 0.336 0.350 0.368 0.348 0.338 0.332 0.347
    6
```



```
    8}00.434 0.431 0.456 0.438 0.455 0.455 0.434 0.431 0.442 0.449 0.447 0.492
    9 0.438 0.454 0.455 0.475 0.489 0.447 0.428 0.483 0.466 0.432 0.494 0.500
    10}00.484 0.450 0.473 0.467 0.507 0.519 0.467 0.487 0.506 0.429 0.473 0.546
    11
    12 0.532 0.530 0.542 0.528 0.566 0.590 0.541 0.581 0.594 0.556 0.536 0.619
        year
age 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003
    0 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    1
    2}00.201 0.190 0.163 0.200 0.185 0.196 0.170 0.210 0.194 0.190 0.206 0.181
    3 0.260 0.266 0.240 0.278 0.250}00.257 0.251 0.260 0.253 0.246 0.245 0.251
    40.308 0.323 0.306 0.327 0.322 0.310 0.300 0.317 0.301 0.303 0.288 0.277
    5 0.360 0.359 0.368 0.385 0.372 0.356 0.348 0.356 0.357 0.342 0.333 0.341
    6
    7
    8 0.458 0.459 0.480 0.491 0.471 0.473 0.455 0.456 0.438 0.451 0.429 0.489
    9 0.487 0.480 0.496 0.511 0.513 0.505 0.475 0.489 0.464 0.484 0.458 0.490
    10}00.513 0.515 0.550 0.517 0.508 0.511 0.530 0.508 0.489 0.521 0.511 0.488
    11
    12 0.572 0.580 0.608 0.603 0.573 0.583 0.549 0.575 0.551 0.572 0.558 0.540
        year
age 2004 2005 2006 2007 2008 2009 2010 2011 2012 2013 2014 2015
    0 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    1
```



```
    3 0.258 0.221 0.236 0.206 0.207 0.214 0.199 0.223 0.216 0.179 0.199 0.247
    4 0.319 0.328 0.291 0.285 0.260 0.268 0.246 0.274 0.255 0.249 0.238 0.254
    5 0.356 0.378 0.333 0.329 0.346 0.295 0.296 0.332 0.288 0.280}0.3.291 0.288
    6
    7}00.449 0.464 0.413 0.448 0.393 0.386 0.389 0.389 0.360 0.341 0.341 0.350
    8}00.4820.481 0.437 0.452 0.448 0.437 0.407 0.430 0.390 0.375 0.387 0.381
    9 0.506 0.547 0.455 0.514 0.452 0.461 0.439 0.452 0.453 0.416 0.416 0.412
    10}00.519 0.538 0.469 0.538 0.478 0.517 0.489 0.495 0.498 0.441 0.466 0.447
    11 0.579 0.509 0.531 0.542 0.487 0.548 0.532 0.518}00.503 0.496 0.472 0.485
    12
        year
age 2016 2017 2018 2019 2020 2021
    0 0.000 0.000 0.000 0.000 0.000 0.000
    1}00.059 0.058 0.064 0.070 0.069 0.064
```



```
    3 0.238 0.237 0.266 0.250 0.252 0.261
    4 0.282 0.278 0.283 0.293 0.289 0.281
    5
    6
    7}00.3680.338 0.346 0.365 0.376 0.392
    8 0.385 0.377 0.364 0.371 0.394 0.416
    9}00.404 0.394 0.389 0.397 0.400 0.423
    10}00.424 0.426 0.419 0.428 0.423 0.446
    11
    12 0.473 0.499 0.491 0.481 0.488 0.496
```


## Table 8.7.1.5. NE Atlantic Mackerel. NATURAL MORTALITY

```
Units : NA
    year
age 1980 1981 1982 1983 1984 1985 1986 1987 1988 1989 1990 1991 1992 1993 1994
```



```
    1
    2
```



```
    4 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    5
    6
    0.15
```



```
    9 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
```




```
    12
        year
    age 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 2007 2008 2009
    0}00.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    1
    2
```



```
    4 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    5
    6
```




```
    9 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    10}00.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    11 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    12}00.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
        year
age 2010 2011 2012 2013 2014 2015 2016 2017 2018 2019 2020 2021
```



```
    1
    2
    3 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
```



```
    5
    0.15
    0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    8}00.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
```





```
    12}00.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
```


## Table 8.7.1.6. NE Atlantic Mackerel. PROPORTION MATURE

```
ear
age 1980
    0 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    1}00.093 0.097 0.097 0.098 0.102 0.102 0.102 0.102 0.102 0.102 0.102 0.102
    0.521 0.497 0.498 0.485 0.467 0. 516 0.522 0.352 0.360}0.0.372 0.392 0.435
    0.872 0.837 0.857 0.863 0.853 0.885 0.926 0.922 0.901 0.915 0.909 0.912
    4 0.949 0.934 0.930 0.940 0.938 0.940}00.983 0.994 0.989 0.994 0.996 0.991
    5 0.972 0.976 0.969 0.972 0.966 0.966 0.965 0.997 0.994 0.996 0.998 0.996
    6 0.984 0.984 0.987 0.999 1.000 1.000 1.000 1.000 1.000 1.000 1.000 0.996
    7 0.990 0.987 0.985 0.984 0.975 0.976 1.000 1.000 1.000 1.000 1.000 1.000
    8 1.000 0.999 0.999 0.999 0.999 0.999 0.991 0.992 0.991 0.993 0.995 1.000
    9 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    10 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    11 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    12 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
        year
age
        0.000 0.0, 199 (1994 2000 2001 2002 2003
        0.000 0.000 0.000 0.000 .000 0.000 0.000 0.000 0.000 0.000 0.000
    1 0.102 0.102 0.102 0.102 0.102 0.097 0.097 0.097 0.104 0.104 0.104 0.106
    2 0.520 0.534 0.621 0.599 0.586 0.621 0.688}0.50.669 0.692 0.675 0.710 0.690
    3}00.928 0.934 0.938 0.931 0.936 0.880 0.886 0.876 0.909 0.909 0.937 0.940
    4 0.996 0.996 0.994 0.993 1.000 0.993 0.994 0.989 0.989 0.987 0.992 0.988
    5 0.997 0.997 0.997 0.994 1.000 0.998 0.999 0.999 0.998 0.998 1.000 1.000
    6
    7 1.000 1.000 0.999 0.999 0.999 1.000 1.000 1.000 1.000 0.999 1.000 0.999
    8}1.0001.000 1.000 1.000 1.000 0.994 0.995 0.996 0.997 0.997 1.000 1.000
    9 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    10 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    11 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    12 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
        year
    age 2004 2005 2006 2007 2008 2009 2010 2011 2012 2013 2014 2015
    0 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    1 0.106 0.106 0.095 0.095 0.095 0.096 0.096 0.096 0.094 0.092 0.092 0.104
    2}00.761 0.616 0.589 0.546 0.524 0.541 0.667 0.655 0.604 0.683 0.675 0.763
    3 0.962 0.959 0.928 0.921 0.917 0.919 0.930}00.927 0.926 0.921 0.916 0.944
```

$4 \quad 0.9930 .9930 .9940 .9940 .9990 .9990 .9990 .9990 .9990 .998 \quad 0.9990 .998$ $5 \quad 0.9990 .9991 .0001 .000 \quad 0.9991 .0001 .0001 .000 \quad 0.9991 .0001 .000 \quad 0.999$ $6 \quad 1.0001 .0001 .0001 .0001 .0001 .00000 .9990 .9990 .999 \quad 0.99910 .9991 .000$ $7 \quad 0.9990 .9991 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .00010 .9991 .999$ 81.0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .000
91.0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .000
101.0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .0001 .000
$\begin{array}{lllllllllllll}11 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000\end{array}$
$\begin{array}{lllllllllllllllll}12 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000\end{array}$ year
age $2016 \quad 2017 \quad 2018 \quad 2019 \quad 2020 \quad 2021$
$0 \quad 0.000 \quad 0.000 \quad 0.000 \quad 0.000 \quad 0.000 \quad 0.000$
$\begin{array}{lllllll}1 & 0.127 & 0.125 & 0.105 & 0.105 & 0.105 & 0.119\end{array}$
$\begin{array}{lllllll}2 & 0.633 & 0.606 & 0.464 & 0.534 & 0.519 & 0.569\end{array}$
$3 \quad 0.9370 .9450 .902 \quad 0.910 \quad 0.9150 .943$
$400.997 \quad 0.998 \quad 0.998 \quad 0.999 \quad 0.998 \quad 0.999$
50.9991 .0001 .0001 .0001 .0001 .000
$6 \quad 1.0001 .0001 .0001 .0001 .000 \quad 0.999$
$7 \quad 0.999 \quad 0.999 \quad 0.9991 .000 \quad 1.000 \quad 0.999$
$8 \quad 1.0001 .0001 .0001 .0001 .0001 .000$
$9 \quad 1.0001 .0001 .0001 .0001 .0001 .000$
$101.0001 .0001 .000 \quad 1.000 \quad 1.000 \quad 1.000$
111.0001 .0001 .0001 .0001 .0001 .000
121.0001 .0001 .0001 .0001 .0001 .000

## Table 8.7.1.7. NE Atlantic Mackerel. FRACTION OF HARVEST BEFORE SPAWNING

## year

10.1660 .1660 .1660 .1660 .1660 .1660 .1660 .1660 .1660 .1660 .1390 .111
$20.2090 .2090 .2090 .209 \quad 0.2090 .2090 .209 \quad 0.209 \quad 0.2090 .209 \quad 0.240 \quad 0.272$
$30.209 \quad 0.209 \quad 0.209 \quad 0.209 \quad 0.209 \quad 0.209 \quad 0.209 \quad 0.209 \quad 0.209 \quad 0.209 \quad 0.240 \quad 0.272$
$4 \quad 0.2090 .209 \quad 0.209 \quad 0.209 \quad 0.2090 .209 \quad 0.209 \quad 0.209 \quad 0.209 \quad 0.209 \quad 0.240 \quad 0.272$
$\begin{array}{llllllllllllllllll}5 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.393 & 0.406\end{array}$
$\begin{array}{llllllllllllllllllll}6 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.393 & 0.406\end{array}$
$\begin{array}{llllllllllllllll}7 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.393 & 0.406\end{array}$
$\begin{array}{lllllllllllllll}8 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.393 & 0.406\end{array}$
$9 \quad 0.380 \quad 0.380 \quad 0.380 \quad 0.380 \quad 0.380 \quad 0.380 \quad 0.380 \quad 0.380 \quad 0.380 \quad 0.380 \quad 0.3930 .406$
$\begin{array}{lllllllllllllllllll}10 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.393 & 0.406\end{array}$
$\begin{array}{llllllllllllllllllllllllll}11 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.393 & 0.406\end{array}$
$\begin{array}{llllllllllllllllllll}12 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.380 & 0.393 & 0.406\end{array}$ year
$\begin{array}{lllllllllllllll}0.084 & 0.165 & 0.249 & 0.331 & 0.269 & 0.206 & 0.144 & 0.125 & 0.106 & 0.088 & 0.142 & 0.197\end{array}$
$\begin{array}{lllllllllllll}2 & 0.304 & 0.301 & 0.298 & 0.296 & 0.295 & 0.295 & 0.295 & 0.320 & 0.347 & 0.373 & 0.360 & 0.347\end{array}$
$\begin{array}{llllllllllll}2 & 0.304 & 0.301 & 0.298 & 0.296 & 0.295 & 0.295 & 0.295 & 0.320 & 0.347 & 0.373 & 0.360 \\ 0.347 \\ 3 & 0.304 & 0.301 & 0.298 & 0.296 & 0.295 & 0.295 & 0.295 & 0.320 & 0.347 & 0.373 & 0.360\end{array} 0.347$
$4 \quad 0.304 \quad 0.301 \quad 0.298 \quad 0.296 \quad 0.2950 .295 \quad 0.295 \quad 0.320 \quad 0.3470 .3730 .360 \quad 0.347$
$\begin{array}{lllllllllllllllll}5 & 0.419 & 0.444 & 0.469 & 0.494 & 0.494 & 0.494 & 0.495 & 0.461 & 0.426 & 0.392 & 0.408 & 0.425\end{array}$
$6 \quad 0.419 \quad 0.444 \quad 0.469 \quad 0.494 \quad 0.494 \quad 0.494 \quad 0.495 \quad 0.461 \quad 0.426 \quad 0.392 \quad 0.4080 .425$
$\begin{array}{llllllllllllll}7 & 0.419 & 0.444 & 0.469 & 0.494 & 0.494 & 0.494 & 0.495 & 0.461 & 0.426 & 0.392 & 0.408 & 0.425\end{array}$
$8 \quad 0.419 \quad 0.444 \quad 0.469 \quad 0.494 \quad 0.494 \quad 0.49410 .4950 .4610 .426 \quad 0.392 \quad 0.408 \quad 0.425$
$\begin{array}{llllllllllllll}9 & 0.419 & 0.444 & 0.469 & 0.494 & 0.494 & 0.494 & 0.495 & 0.461 & 0.426 & 0.392 & 0.408 & 0.425\end{array}$
$\begin{array}{lllllllllllll}10 & 0.419 & 0.444 & 0.469 & 0.494 & 0.494 & 0.494 & 0.495 & 0.461 & 0.426 & 0.392 & 0.408 & 0.425\end{array}$
$\begin{array}{lllllllllllllll}11 & 0.419 & 0.444 & 0.469 & 0.494 & 0.494 & 0.494 & 0.495 & 0.461 & 0.426 & 0.392 & 0.408 & 0.425\end{array}$
$\begin{array}{llllllllllllll}12 & 0.419 & 0.444 & 0.469 & 0.494 & 0.494 & 0.494 & 0.495 & 0.461 & 0.426 & 0.392 & 0.408 & 0.425\end{array}$
year
age $2004 \quad 2005 \quad 2006 \quad 2007$ 2008 $2009 \quad 2010 \quad 2011 \quad 2012 \quad 2013 \quad 2014 \quad 2015$
0.0000 .0000 .0000 .0000 .0000 .0000 .0000 .0000 .0000 .0000 .0000 .000

$2 \quad 0.334 \quad 0.317 \quad 0.300 \quad 0.284 \quad 0.2660 .249 \quad 0.232 \quad 0.176 \quad 0.119 \quad 0.064 \quad 0.1170 .171$
$\begin{array}{llllllllllllll}3 & 0.334 & 0.317 & 0.300 & 0.284 & 0.266 & 0.249 & 0.232 & 0.176 & 0.119 & 0.064 & 0.117 & 0.171\end{array}$
$40.3340 .3170 .300 \quad 0.284 \quad 0.2660 .249 \quad 0.232 \quad 0.176 \quad 0.119 \quad 0.064 \quad 0.1170 .171$
$50.4410 .4090 .376 \quad 0.3440 .310 \quad 0.275 \quad 0.2420 .2330 .2250 .2160 .2030 .189$
$\begin{array}{lllllllllllllll}6 & 0.441 & 0.409 & 0.376 & 0.344 & 0.310 & 0.275 & 0.242 & 0.233 & 0.225 & 0.216 & 0.203 & 0.189\end{array}$
$7 \begin{array}{lllllllllllll}7 & 0.441 & 0.409 & 0.376 & 0.344 & 0.310 & 0.275 & 0.242 & 0.233 & 0.225 & 0.216 & 0.203 & 0.189\end{array}$
$\begin{array}{lllllllllllll}8 & 0.441 & 0.409 & 0.376 & 0.344 & 0.310 & 0.275 & 0.242 & 0.233 & 0.225 & 0.216 & 0.203 & 0.189\end{array}$
$\begin{array}{llllllllllllllll}9 & 0.441 & 0.409 & 0.376 & 0.344 & 0.310 & 0.275 & 0.242 & 0.233 & 0.225 & 0.216 & 0.203 & 0.189\end{array}$
$\begin{array}{llllllllllllllll}10 & 0.441 & 0.409 & 0.376 & 0.344 & 0.310 & 0.275 & 0.242 & 0.233 & 0.225 & 0.216 & 0.203 & 0.189\end{array}$
$\begin{array}{llllllllllllll}11 & 0.441 & 0.409 & 0.376 & 0.344 & 0.310 & 0.275 & 0.242 & 0.233 & 0.225 & 0.216 & 0.203 & 0.189\end{array}$
$\begin{array}{lllllllllllll}12 & 0.441 & 0.409 & 0.376 & 0.344 & 0.310 & 0.275 & 0.242 & 0.233 & 0.225 & 0.216 & 0.203 & 0.189\end{array}$
year
age $201620172018 \quad 2019 \quad 20202021$
$0 \quad 0.000 \quad 0.000 \quad 0.000 \quad 0.000 \quad 0.000 \quad 0.000$
$1 \quad 0.1430 .232 \quad 0.393 \quad 0.581 \quad 0.533 \quad 0.187$
$\begin{array}{lllllll}2 & 0.224 & 0.153 & 0.179 & 0.183 & 0.184 & 0.090\end{array}$
$3 \quad 0.224 \quad 0.153 \quad 0.179 \quad 0.183 \quad 0.184 \quad 0.090$
$4 \quad 0.224 \quad 0.1530 .179 \quad 0.183 \quad 0.184 \quad 0.090$
$\begin{array}{llllllll}5 & 0.178 & 0.295 & 0.196 & 0.301 & 0.317 & 0.233\end{array}$
$6 \quad 0.178 \quad 0.295 \quad 0.196 \quad 0.301 \quad 0.317 \quad 0.233$
$\begin{array}{llllll}0.178 & 0.295 & 0.196 & 0.301 & 0.317 & 0.233\end{array}$
$8 \quad 0.178 \quad 0.2950 .196 \quad 0.301 \quad 0.317 \quad 0.233$
$9 \quad 0.1780 .2950 .196 \quad 0.301 \quad 0.3170 .233$
$\begin{array}{llllllll}10 & 0.178 & 0.295 & 0.196 & 0.301 & 0.317 & 0.233\end{array}$
$\begin{array}{lllllll}11 & 0.178 & 0.295 & 0.196 & 0.301 & 0.317 & 0.233\end{array}$
$\begin{array}{lllllll}12 & 0.178 & 0.295 & 0.196 & 0.301 & 0.317 & 0.233\end{array}$

Table 8.7.1.8. NE Atlantic Mackerel. FRACTION OF NATURAL MORTALITY BEFORE SPAWNING

## year

age $19801981 \quad 1982 \quad 1983 \quad 1984 \quad 1985 \quad 1986$
$\begin{array}{lllllllllllllllllllll}0 & 0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$1 \begin{array}{lllllllllllllllllllllll}1 & 0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$20.397 \quad 0.396 \quad 0.394 \quad 0.392 \quad 0.394 \quad 0.396 \quad 0.397 \quad 0.388 \quad 0.378 \quad 0.369 \quad 0.357 \quad 0.345$
$\begin{array}{llllllllllllllll}3 & 0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{llllllllllllllllllll}4 & 0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{llllllllllllllllllllllll}5 & 0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{llllllllllllllllllllllllll}6 & 0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{llllllllllllll}7 & 0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$8 \quad 0.397 \quad 0.396 \quad 0.394 \quad 0.392 \quad 0.394 \quad 0.396 \quad 0.397 \quad 0.388 \quad 0.378 \quad 0.369 \quad 0.357 \quad 0.345$
$\begin{array}{lllllllllllllllllllllll} & 0 & 0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$100.3970 .3960 .3940 .3920 .3940 .3960 .3970 .388 \quad 0.3780 .3690 .3570 .345$
$\begin{array}{llllllllllllllllllllllll}11 & 0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$
$\begin{array}{llllllllllllll}12 & 0.397 & 0.396 & 0.394 & 0.392 & 0.394 & 0.396 & 0.397 & 0.388 & 0.378 & 0.369 & 0.357 & 0.345\end{array}$ year

$0 \quad 0.333 \quad 0.341 \quad 0.349 \quad 0.357 \quad 0.339 \quad 0.322 \quad 0.3040 .3250 .3460 .3660 .3610 .355$
$1 \quad 0.333 \quad 0.341 \quad 0.349 \quad 0.357 \quad 0.339 \quad 0.322 \quad 0.3040 .325 \quad 0.346 \quad 0.366 \quad 0.361 \quad 0.355$
$\begin{array}{llllllllllllll}2 & 0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$\begin{array}{lllllllllllllllllll}3 & 0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$4 \quad 0.333 \quad 0.341 \quad 0.349 \quad 0.357 \quad 0.339 \quad 0.322 \quad 0.3040 .325 \quad 0.346 \quad 0.366 \quad 0.361 \quad 0.355$
$50.3330 .3410 .3490 .357 \quad 0.339 \quad 0.3220 .3040 .3250 .3460 .3660 .3610 .355$ $6 \quad 0.3330 .3410 .349 \quad 0.357 \quad 0.339 \quad 0.322 \quad 0.3040 .325 \quad 0.346 \quad 0.366 \quad 0.3610 .355$ $\begin{array}{llllllllllllllll}7 & 0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$\begin{array}{llllllllllllllll}8 & 0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$9 \quad 0.3330 .3410 .349 \quad 0.357 \quad 0.339 \quad 0.3220 .3040 .3250 .3460 .3660 .3610 .355$
$\begin{array}{lllllllllllllllllllllllll}10 & 0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$\begin{array}{lllllllllllllllllllll}11 & 0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$
$\begin{array}{llllllllllllllllllll}12 & 0.333 & 0.341 & 0.349 & 0.357 & 0.339 & 0.322 & 0.304 & 0.325 & 0.346 & 0.366 & 0.361 & 0.355\end{array}$ year
age $2004 \quad 2005 \quad 2006 \quad 2007 \quad 2008 \quad 2009 \quad 2010 \quad 2011 \quad 2012 \quad 2013 \quad 2014 \quad 2015$
$0 \quad 0.350 \quad 0.346 \quad 0.342 \quad 0.339 \quad 0.311 \quad 0.283 \quad 0.255 \quad 0.2520 .2490 .246 \quad 0.278 \quad 0.311$
$1 \quad 0.3500 .346 \quad 0.342 \quad 0.339 \quad 0.311 \quad 0.283 \quad 0.255 \quad 0.252 \quad 0.249 \quad 0.246 \quad 0.278 \quad 0.311$
$2 \begin{array}{lllllllllllll}2 & 0.350 & 0.346 & 0.342 & 0.339 & 0.311 & 0.283 & 0.255 & 0.252 & 0.249 & 0.246 & 0.278 & 0.311\end{array}$
$\begin{array}{llllllllllllll}3 & 0.350 & 0.346 & 0.342 & 0.339 & 0.311 & 0.283 & 0.255 & 0.252 & 0.249 & 0.246 & 0.278 & 0.311\end{array}$
$4 \quad 0.350 \quad 0.346 \quad 0.342 \quad 0.339 \quad 0.311 \quad 0.2830 .255 \quad 0.2520 .249 \quad 0.246 \quad 0.278 \quad 0.311$
$5 \quad 0.350 \quad 0.3460 .342 \quad 0.339 \quad 0.311 \quad 0.2830 .255 \quad 0.252 \quad 0.249 \quad 0.246 \quad 0.278 \quad 0.311$
$6 \quad 0.3500 .346 \quad 0.342 \quad 0.339 \quad 0.311 \quad 0.283 \quad 0.255 \quad 0.252 \quad 0.249 \quad 0.246 \quad 0.278 \quad 0.311$ $\begin{array}{lllllllllllll}0.350 & 0.346 & 0.342 & 0.339 & 0.311 & 0.283 & 0.255 & 0.252 & 0.249 & 0.246 & 0.278 & 0.311\end{array}$
$8 \quad 0.350 \quad 0.346 \quad 0.342 \quad 0.339 \quad 0.311 \quad 0.283 \quad 0.255 \quad 0.252 \quad 0.249 \quad 0.246 \quad 0.278 \quad 0.311$
$9 \quad 0.350 \quad 0.3460 .3420 .3390 .3110 .2830 .2550 .2520 .2490 .2460 .2780 .311$
$\begin{array}{lllllllllllllllllllll}10 & 0.350 & 0.346 & 0.342 & 0.339 & 0.311 & 0.283 & 0.255 & 0.252 & 0.249 & 0.246 & 0.278 & 0.311\end{array}$
$\begin{array}{llllllllllllllllll}11 & 0.350 & 0.346 & 0.342 & 0.339 & 0.311 & 0.283 & 0.255 & 0.252 & 0.249 & 0.246 & 0.278 & 0.311\end{array}$
$\begin{array}{lllllllllllllll}12 & 0.350 & 0.346 & 0.342 & 0.339 & 0.311 & 0.283 & 0.255 & 0.252 & 0.249 & 0.246 & 0.278 & 0.311\end{array}$ year
age $2016 \quad 2017 \quad 2018 \quad 2019 \quad 2020 \quad 2021$
$\begin{array}{lllllll}0 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329\end{array}$
$\begin{array}{llllllll}1 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329\end{array}$
$\begin{array}{lllllll}2 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329\end{array}$
$\begin{array}{llllllll}3 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329\end{array}$
$4 \quad 0.3430 .3270 .312 \quad 0.296 \quad 0.312 \quad 0.329$
$\begin{array}{llllllll}5 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329\end{array}$
$\begin{array}{lllllll}6 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329\end{array}$
$\begin{array}{lllllll}7 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329\end{array}$
$8 \quad 0.3430 .327 \quad 0.312 \quad 0.296 \quad 0.312 \quad 0.329$
$9 \quad 0.3430 .3270 .312 \quad 0.296 \quad 0.312 \quad 0.329$
$\begin{array}{llllllll}10 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329\end{array}$
$\begin{array}{llllllll}11 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329\end{array}$
$\begin{array}{lllllll}12 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329\end{array}$

## Table 8.7.1.9. NE Atlantic Mackerel. SURVEY INDICES

Some random text
103
SSB-egg-based-survey
19922022


Table 8.7.1.10. NE Atlantic Mackerel. RFID recapture data for the year 2021

| Releas e Yr | Recaptur e Yr | Year- <br> class | age at <br> release | Numbers scanned in recapture Yr | Numbers Released in Release Year | Numbers recaptured |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2019 | 2021 | 2008 | 11 | 28671260 | 2556 | 19 |
| 2019 | 2021 | 2009 | 10 | 32424008 | 2871 | 22 |
| 2019 | 2021 | 2010 | 9 | 73535492 | 4728 | 37 |
| 2019 | 2021 | 2011 | 8 | 91009636 | 9483 | 79 |
| 2019 | 2021 | 2012 | 7 | 54099631 | 6785 | 57 |
| 2019 | 2021 | 2013 | 6 | 43820959 | 8040 | 70 |
| 2019 | 2021 | 2014 | 5 | 82980291 | 5824 | 51 |
| 2020 | 2021 | 2009 | 11 | 32424008 | 2191 | 28 |
| 2020 | 2021 | 2010 | 10 | 73535492 | 5001 | 67 |
| 2020 | 2021 | 2011 | 9 | 91009636 | 5081 | 70 |
| 2020 | 2021 | 2012 | 8 | 54099631 | 5474 | 76 |
| 2020 | 2021 | 2013 | 7 | 43820959 | 2665 | 36 |
| 2020 | 2021 | 2014 | 6 | 82980291 | 4339 | 68 |
| 2020 | 2021 | 2015 | 5 | 35012724 | 1509 | 28 |

Table 8.7.2.1. NE Atlantic Mackerel. SAM parameter estimates for the 2022 update.

|  | esti- <br> mate | std.de <br> $\mathbf{v}$ | confidence interval lower <br> bound | confidence interval upper <br> bound |
| :--- | :--- | :--- | :--- | :--- |
| observation standard deviations |  |  |  |  |
| Catches age 0 | 0.89 | 0.18 | 0.63 | 1.28 |
| Catches age 1 | 0.35 | 0.25 | 0.21 | 0.58 |
| Catches age 2-12 | 0.11 | 0.15 | 0.08 | 0.15 |
| Egg survey | 0.32 | 0.25 | 0.20 | 0.53 |
| Recruitment index | 0.26 | 0.31 | 0.14 | 0.48 |
| IESSNS age 3 | 0.66 | 0.23 | 0.41 | 1.04 |
| IESSNS ages 4-11 | 0.45 | 0.14 | 0.34 | 0.60 |
| Recapture overdispersion <br> tags | 1.25 | 0.24 | 1.40 | 1.15 |


| random walk standard deviation |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| F age 0 | 0.28 | 0.43 | 0.12 | 0.68 |
| F age 1 | 0.18 | 0.46 | 0.07 | 0.45 |
| F age 2+ | 0.15 | 0.14 | 0.11 | 0.20 |
| N@age0 | 0.19 | 0.58 | 0.06 | 0.62 |
| process error standard deviation |  |  |  |  |
| N@age1-12+ | 0.21 | 0.08 | 0.18 | 0.24 |
| catchabilities |  |  |  |  |
| egg survey | 1.17 | 0.11 | 0.94 | 1.47 |
| recruitment index | $\begin{aligned} & 4.99 \mathrm{E}- \\ & 09 \end{aligned}$ | 0.11 | 3.97E-09 | $6.28 \mathrm{E}-09$ |
| IESSNS age 3 | 0.79 | 0.22 | 0.51 | 1.22 |
| IESSNS age 4 | 1.19 | 0.17 | 0.85 | 1.66 |
| IESSNS age 5 | 1.64 | 0.17 | 1.18 | 2.29 |
| IESSNS age 6 | 1.75 | 0.17 | 1.25 | 2.45 |
| IESSNS age 7 | 1.94 | 0.17 | 1.37 | 2.73 |
| IESSNS age 8 | 1.83 | 0.17 | 1.30 | 2.58 |
| IESSNS age 9 | 1.96 | 0.17 | 1.39 | 2.77 |
| IESSNS ages 10-11 | 1.88 | 0.17 | 1.34 | 2.63 |
| post tagging survival steal tags | 0.40 | 0.11 | 0.35 | 0.45 |
| post tagging survival RFID tags | 0.16 | 0.11 | 0.13 | 0.19 |

Table 8.7.3.1. NE Atlantic Mackerel. STOCK SUMMARY.

| Year | Recruitment (age 2) |  |  | SSB at spawning time |  |  | Total catch | F (ages 4-8) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Low | Value | High | Low | Value | High |  | Low | Value | High |
| 1980 | 741662 | 2256379 | 6864651 | 1931692 | 4095763 | 8684237 | 734950 | 0.13 | 0.21 | 0.33 |
| 1981 | 1483866 | 3954164 | 10536946 | 1908690 | 3610917 | 6831242 | 754045 | 0.13 | 0.21 | 0.33 |
| 1982 | 2251174 | 4258180 | 8054508 | 2034611 | 3475244 | 5935937 | 716987 | 0.13 | 0.21 | 0.32 |
| 1983 | 2476430 | 4189394 | 7087228 | 2412006 | 3685894 | 5632577 | 672283 | 0.14 | 0.21 | 0.31 |
| 1984 | 1048672 | 1945021 | 3607518 | 2800858 | 3989717 | 5683201 | 641928 | 0.14 | 0.21 | 0.31 |
| 1985 | 911867 | 1761799 | 3403932 | 2949017 | 4015844 | 5468602 | 614371 | 0.15 | 0.21 | 0.31 |
| 1986 | 2414771 | 4075432 | 6878145 | 2709488 | 3608231 | 4805089 | 602201 | 0.15 | 0.22 | 0.31 |
| 1987 | 1645418 | 2713304 | 4474253 | 2689623 | 3575982 | 4754438 | 654992 | 0.16 | 0.22 | 0.31 |
| 1988 | 1704717 | 2668074 | 4175837 | 2719049 | 3526114 | 4572730 | 680491 | 0.17 | 0.23 | 0.32 |
| 1989 | 2626296 | 3859714 | 5672396 | 2586241 | 3297725 | 4204941 | 585920 | 0.18 | 0.24 | 0.33 |
| 1990 | 1615372 | 2352987 | 3427412 | 2703312 | 3381620 | 4230127 | 626107 | 0.19 | 0.26 | 0.34 |
| 1991 | 1806495 | 2640685 | 3860081 | 2648612 | 3278987 | 4059393 | 675665 | 0.21 | 0.27 | 0.36 |
| 1992 | 1255408 | 1959426 | 3058250 | 2453803 | 3005914 | 3682253 | 760690 | 0.22 | 0.29 | 0.38 |
| 1993 | 1667703 | 2400478 | 3455227 | 2192074 | 2665757 | 3241797 | 824568 | 0.24 | 0.31 | 0.39 |
| 1994 | 1937535 | 2792997 | 4026163 | 1924591 | 2323071 | 2804054 | 819087 | 0.25 | 0.32 | 0.40 |
| 1995 | 1447844 | 2068160 | 2954245 | 1914732 | 2290278 | 2739483 | 756277 | 0.25 | 0.31 | 0.39 |
| 1996 | 1465444 | 2062647 | 2903224 | 1825764 | 2174598 | 2590081 | 563472 | 0.25 | 0.30 | 0.37 |


| Year | Recruitment (age 2) |  |  | SSB at spawning time |  |  | Total catch | F (ages 4-8) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Low | Value | High | Low | Value | High |  | Low | Value | High |
| 1997 | 1220370 | 1738105 | 2475486 | 1824264 | 2144607 | 2521204 | 573029 | 0.24 | 0.30 | 0.36 |
| 1998 | 1445736 | 2299083 | 3656118 | 1799608 | 2120261 | 2498049 | 666316 | 0.25 | 0.30 | 0.37 |
| 1999 | 1262250 | 1918742 | 2916672 | 1958221 | 2301530 | 2705027 | 640309 | 0.27 | 0.32 | 0.38 |
| 2000 | 1646261 | 2242303 | 3054147 | 1970480 | 2266622 | 2607272 | 738606 | 0.29 | 0.34 | 0.39 |
| 2001 | 1930231 | 2565296 | 3409305 | 1874159 | 2147511 | 2460732 | 737463 | 0.32 | 0.37 | 0.43 |
| 2002 | 863381 | 1152032 | 1537189 | 1778419 | 2054100 | 2372516 | 771422 | 0.33 | 0.39 | 0.46 |
| 2003 | 3687488 | 4794496 | 6233834 | 1718635 | 1993516 | 2312361 | 679287 | 0.34 | 0.41 | 0.48 |
| 2004 | 5011045 | 6668900 | 8875242 | 2204914 | 2608232 | 3085323 | 660491 | 0.32 | 0.38 | 0.45 |
| 2005 | 1745850 | 2356809 | 3181573 | 2013569 | 2404435 | 2871174 | 549514 | 0.26 | 0.31 | 0.36 |
| 2006 | 2532416 | 3427824 | 4639829 | 1888395 | 2242585 | 2663208 | 481181 | 0.24 | 0.28 | 0.33 |
| 2007 | 3624149 | 4994825 | 6883900 | 2021689 | 2380749 | 2803579 | 586206 | 0.27 | 0.31 | 0.37 |
| 2008 | 3651784 | 5064453 | 7023604 | 2354473 | 2800351 | 3330667 | 623165 | 0.25 | 0.30 | 0.36 |
| 2009 | 2640841 | 3641939 | 5022536 | 2915680 | 3480016 | 4153581 | 737969 | 0.23 | 0.27 | 0.33 |
| 2010 | 2971798 | 4083427 | 5610870 | 3290915 | 3901258 | 4624797 | 877272 | 0.22 | 0.26 | 0.32 |
| 2011 | 2520467 | 3464819 | 4762994 | 3791504 | 4504765 | 5352206 | 948963 | 0.22 | 0.26 | 0.31 |
| 2012 | 4214540 | 5786958 | 7946034 | 3548472 | 4225513 | 5031731 | 899551 | 0.20 | 0.24 | 0.29 |
| 2013 | 4871399 | 6709016 | 9239829 | 3935828 | 4705814 | 5626437 | 938299 | 0.19 | 0.24 | 0.29 |


| Year | Recruitment (age 2) |  |  | SSB at spawning time |  |  | Total catch | F (ages 4-8) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Low | Value | High | Low | Value | High |  | Low | Value | High |
| 2014 | 2729554 | 3753792 | 5162367 | 4924869 | 5898500 | 7064615 | 1401788 | 0.20 | 0.24 | 0.29 |
| 2015 | 2437757 | 3352679 | 4610983 | 4928398 | 5928946 | 7132621 | 1215827 | 0.18 | 0.22 | 0.27 |
| 2016 | 3652137 | 5035066 | 6941658 | 4638104 | 5598144 | 6756902 | 1100135 | 0.16 | 0.20 | 0.25 |
| 2017 | 1541675 | 2141687 | 2975221 | 4469022 | 5404599 | 6536037 | 1159641 | 0.17 | 0.21 | 0.25 |
| 2018 | 3434902 | 4853899 | 6859101 | 3859967 | 4667166 | 5643167 | 1023144 | 0.17 | 0.21 | 0.26 |
| 2019 | 2020927 | 2923019 | 4227783 | 3233573 | 3945436 | 4814014 | 839727 | 0.18 | 0.22 | 0.27 |
| 2020 | 1816780 | 2698510 | 4008168 | 2888782 | 3570188 | 4412324 | 1039513 | 0.23 | 0.29 | 0.36 |
| 2021 | 4313470 | 6801162 | 10723570 | 3044947 | 3891546 | 4973528 | 1081540 | 0.23 | 0.31 | 0.40 |
| 2022 | 2148239 | 4317473 | 8677141 |  | 3769326 ${ }^{+}$ |  |  |  |  |  |

$\dagger$ Estimated value from the forecast.

Table 8.7.3.2. NE Atlantic Mackerel. ESTIMATED POPULATION ABUNDANCE

|  | year |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 0 | 4977824 | 4592946 | 3796286 | 3630479 | 4120760 | 4018388 | 3996884 | 4085430 | 3714031 | 3542109 |
| 1 | 4770227 | 4667307 | 4444600 | 2747095 | 2554680 | 4153858 | 3337149 | 3297715 | 4020185 | 2989130 |
| 2 | 2256379 | 3954164 | 4258180 | 4189394 | 1945021 | 1761799 | 4075432 | 2713304 | 2668074 | 3859714 |
| 3 | 912180 | 1820022 | 3310754 | 4041761 | 4235132 | 1338773 | 1230503 | 4005081 | 2150334 | 2331117 |
| 4 | 1566134 | 705301 | 1375181 | 2796871 | 3678706 | 3995572 | 998216 | 837570 | 3714427 | 1671198 |
| 5 | 3406968 | 1168826 | 510471 | 949136 | 2140450 | 3017436 | 3135126 | 783356 | 528198 | 2969366 |
| 6 | 2631794 | 2415970 | 846616 | 379243 | 656491 | 1605024 | 2219979 | 2159711 | 600448 | 343216 |
| 7 | 822419 | 1788183 | 1638596 | 578072 | 268789 | 459973 | 1078736 | 1503046 | 1411112 | 463388 |
| 8 | 309077 | 572320 | 1245878 | 1138797 | 397837 | 195390 | 311847 | 770217 | 1046138 | 1067708 |
| 9 | 848340 | 215083 | 397825 | 867613 | 789693 | 278639 | 139301 | 210575 | 547400 | 732107 |
| 10 | 233382 | 590713 | 149594 | 276223 | 603333 | 546635 | 196960 | 96371 | 141292 | 376120 |
| 11 | 341202 | 162433 | 410639 | 104016 | 191710 | 418065 | 376311 | 135405 | 66188 | 92257 |
| 12 | 690723 | 718841 | 612445 | 708530 | 561998 | 521034 | 644266 | 694742 | 562107 | 422472 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 3297094 | 3318561 | 3327308 | 3098419 | 2964293 | 2859040 | 2952257 | 2923500 | 2987978 | 3289960 |
| 1 | 3097593 | 2570564 | 2856940 | 3087202 | 2565454 | 2505469 | 2258676 | 2639679 | 2400853 | 2618309 |
| 2 | 2352987 | 2640685 | 1959426 | 2400478 | 2792997 | 2068160 | 2062647 | 1738105 | 2299083 | 1918742 |
| 3 | 3890784 | 2115127 | 2534665 | 1621187 | 1967779 | 2382272 | 2153403 | 1924846 | 1227717 | 2350070 |
| 4 | 1827221 | 3033337 | 1514054 | 2017745 | 1086566 | 1416848 | 1800381 | 1772730 | 1630353 | 1239955 |
| 5 | 1073211 | 1243666 | 1917567 | 979620 | 1372311 | 673445 | 964445 | 1200871 | 1508741 | 1265166 |
| 6 | 1973779 | 775220 | 945978 | 1153956 | 583882 | 966540 | 489028 | 724376 | 855047 | 899888 |
| 7 | 214236 | 1223334 | 473593 | 569829 | 648711 | 343790 | 572199 | 320013 | 479077 | 614420 |
| 8 | 351975 | 137286 | 732262 | 309748 | 336390 | 280329 | 212476 | 344773 | 262374 | 310595 |
| 9 | 731514 | 249257 | 88556 | 410862 | 181936 | 175936 | 135539 | 150472 | 211003 | 180762 |
| 10 | 478615 | 497774 | 159923 | 52759 | 214721 | 108449 | 92183 | 85476 | 101841 | 131487 |
| 11 | 252493 | 302692 | 312443 | 97205 | 29473 | 129461 | 62308 | 49031 | 52952 | 63338 |
| 12 | 342343 | 388796 | 440913 | 466489 | 339941 | 219549 | 209528 | 169160 | 140089 | 124666 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 0 | 3144960 | 4433789 | 5064730 | 4013211 | 4940011 | 6087928 | 6119559 | 5148658 | 4756583 | 4639324 |
| 1 | 3035680 | 1857761 | 5101748 | 6175998 | 2832192 | 3897654 | 5945414 | 5618783 | 4333207 | 4134463 |
| 2 | 2242303 | 2565296 | 1152032 | 4794496 | 6668900 | 2356809 | 3427824 | 4994825 | 5064453 | 3641939 |
| 3 | 1790222 | 1727766 | 2474397 | 798852 | 3936615 | 5336987 | 1738772 | 2564331 | 4591346 | 5210179 |
| 4 | 1819767 | 1279107 | 1516076 | 1553732 | 750904 | 1902452 | 3240215 | 1511773 | 2043494 | 4067468 |
| 5 | 1010369 | 1237307 | 970982 | 902493 | 996705 | 540476 | 1055644 | 2123609 | 1264672 | 1661329 |
| 6 | 859823 | 660339 | 800265 | 568797 | 472792 | 482143 | 376210 | 757409 | 1140265 | 931368 |
| 7 | 619464 | 600842 | 404644 | 376524 | 265228 | 233351 | 284819 | 255528 | 426786 | 710562 |
| 8 | 375412 | 410741 | 343916 | 238702 | 183557 | 135455 | 133426 | 184948 | 179036 | 270518 |
| 9 | 191120 | 238796 | 226933 | 192419 | 115881 | 87835 | 74299 | 95034 | 102742 | 111945 |
| 10 | 113685 | 126738 | 126615 | 116014 | 91545 | 62816 | 53326 | 47505 | 59034 | 53904 |
| 11 | 70379 | 68495 | 62601 | 65601 | 46922 | 31568 | 32450 | 34341 | 22259 | 29446 |
| 12 | 122286 | 126402 | 111017 | 80331 | 56491 | 40612 | 39033 | 39894 | 31628 | 21099 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
| 0 | 5520852 | 6360676 | 5704912 | 5490274 | 5587694 | 5154715 | 5999571 | 6307925 | 6367101 | 7176591 |
| 1 | 4273885 | 5587697 | 6415242 | 4551906 | 4222134 | 5704906 | 3528918 | 5217025 | 4291009 | 4229297 |
| 2 | 4083427 | 3464819 | 5786958 | 6709016 | 3753792 | 3352679 | 5035066 | 2141687 | 4853899 | 2923019 |
| 3 | 3515562 | 3836598 | 2860930 | 5634314 | 7160193 | 3007330 | 2663718 | 4323592 | 1345459 | 3459984 |
| 4 | 4900988 | 3206377 | 3179936 | 2539120 | 5288131 | 4883570 | 2789692 | 2085552 | 2773684 | 904800 |
| 5 | 3065295 | 3582228 | 2550450 | 2653768 | 2532000 | 3851455 | 3615558 | 2164113 | 1356115 | 1440392 |
| 6 | 1322833 | 2226928 | 2584252 | 2393429 | 2527420 | 2092277 | 3059819 | 3088656 | 1481157 | 972355 |
| 7 | 584975 | 951404 | 1432488 | 1727766 | 2164657 | 1975678 | 1678923 | 2744688 | 2330127 | 1046798 |
| 8 | 391333 | 429065 | 634372 | 925526 | 1446550 | 1628395 | 1450473 | 1350611 | 1926895 | 1629024 |
| 9 | 175392 | 215371 | 282202 | 434894 | 648104 | 1042218 | 973051 | 1143531 | 1041363 | 1332009 |
| 10 | 76179 | 98128 | 130947 | 176676 | 286800 | 493250 | 571416 | 704991 | 709220 | 605128 |
| 11 | 26303 | 47692 | 55031 | 86098 | 98402 | 148441 | 246010 | 388856 | 490368 | 436225 |
| 12 | 32975 | 40716 | 51979 | 72920 | 71228 | 109057 | 146957 | 280765 | 354142 | 451171 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2020 | 2021 | 2022 |  |  |  |  |  |  |  |
| 0 | 6610775 | 6333542 | 6333542 |  |  |  |  |  |  |  |
| 1 | 6919855 | 5097986 | 5442078 |  |  |  |  |  |  |  |
| 2 | 2698510 | 6801162 | 4317473 |  |  |  |  |  |  |  |
| 3 | 2019484 | 2072539 | 5459511 |  |  |  |  |  |  |  |
| 4 | 2203869 | 1865960 | 1413880 |  |  |  |  |  |  |  |
| 5 | 733129 | 1694067 | 1344203 |  |  |  |  |  |  |  |
| 6 | 1008745 | 533527 | 989652 |  |  |  |  |  |  |  |
| 7 | 831732 | 991559 | 350945 |  |  |  |  |  |  |  |
| 8 | 850646 | 675470 | 542972 |  |  |  |  |  |  |  |
| 9 | 1081500 | 769063 | 399065 |  |  |  |  |  |  |  |
| 10 | 923342 | 698265 | 557747 |  |  |  |  |  |  |  |
| 11 | 437216 | 503532 | 472776 |  |  |  |  |  |  |  |
| 12 | 620191 | 465215 | 581817 |  |  |  |  |  |  |  |

Table 8.7.3.3. NE Atlantic Mackerel. ESTIMATED FISHING MORTALITY


```
    0}0.0090.009 0.009 0.009 0.009 0.009 0.009 0.009 0.009 0.009 0.008 0.008 0.008
    1 0.033 0.033 0.033 0.033 0.033 0.032 0.032 0.032 0.032 0.032 0.032 0.032 0.032
    0.058 0.058 0.057 0.057 0.058 0.057 0.057 0.057 0.057 0.058 0.058 0.059 0.059
    0.106 0.106 0.105 0.106 0.107 0.109 0.111 0.113 0.116 0.119 0.122 0.126 0.129
```




```
    0.238 0.238 0.240 0.241 0.245 0.249 0.255 0.261 0.268 0.281 0.293 0.304 0.314
    0.213 0.213 0.213 0.213 0.214 0.218 0.223 0.230}0.2390.239.253 0.275 0.304 0.335
    0.213 0.213 0.213 0.213}00.214 0.218 0.223 0.230 0.239 0.253 0.275 0.304 0.335
    0.213 0.213 0.213 0.213 0.214 0.218
```



```
    11 0.213 0.213 0.213 0.213 0.214 0.218 0.223 0.230 0.239 0.253 0.275 0.304 0.335
    12 0.213 0.213 0.213 0.213 0.214 0.218 0.223 0.230 0.239 0.253 0.275 0.304 0.335
        year
        1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005
        0.008 0.008 0.008 0.008 0.008 0.008 0.008 0.008 0.007 0.007 0.006 0.005 0.005
        0.032 0.032 0.032 0.032 0.032 0.031 0.031 0.031 0.029 0.029 0.025 0.020 0.018
        0.060 0.061 0.062 0.063 0.065 0.066 0.067 0.069 0.069 0.068 0.066 0.069 0.062
```




```
        0.251 0.254 0.259 0.268 0.281 0.300 0.317 0.341 0.327 0.334 0.328 0.314 0.275
        0.322 0.326 0.326 0.324 0.325 0.330 0.344 0.366 0.413 0.406 0.412 0.388 0.340
        0.367 0.385 0.376 0.349 0.328 0.329 0.343 0.355 0.411 0.475 0.529 0.480}0.3.360
    8 0.367 0.385 0.376 0.349 0.328}00.329 0.343 0.355 0.411 0.475 0.529 0.480 0.360
```



```
    10 0.367 0.385 0.376 0.349 0.328 0.329 0.343 0.355 0.411 0.475 0.529 0.480 0.360
    11 0.367 0.385 0.376 0.349 0.328 0.329 0.343 0.355 0.411 0.475 0.529 0.480 0.360
    12 0.367 0.385 0.376 0.349 0.328}0.3.329 0.343 0.355 0.411 0.475 0.529 0.480 0.360
        year
    age 2006 2007 2008 2009 2010 2011 2012 2013 2014 2015 2016 2017 2018
        0.005 0.005 0.005 0.004 0.004 0.004 0.003 0.003 0.003 0.002 0.001 0.001 0.001
        0.019 0.017 0.015 0.014 0.014 0.013 0.013 0.013 0.013 0.014 0.013 0.011 0.010
        0.054 0.044 0.038 0.036 0.037 0.037 0.038 0.039 0.041 0.042 0.045 0.047 0.050
        0.111 0.101 0.099 0.098 0.096 0.093 0.087 0.086 0.098 0.099 0.109 0.117 0.116
```



```
        0.246 0.254 0.246 0.235 0.234 0.222 0.215 0.212 0.234 0.210 0.203 0.217 0.230
        0.327 0.323 0.293 0.288}0.268 0.261 0.243 0.231 0.249 0.244 0.217 0.213 0.241
```



```
    8
```



```
    10 0.334 0.413 0.401 0.338 0.324 0.320 0.285 0.283 0.270 0.250 0.207 0.208 0.214
```




```
        year
age 2019 2020 2021 2022
        0.001 0.001 0.001 0.001
        0.010 0.009 0.008 0.008
        0.048 0.047 0.051 0.051
        0.114 0.109 0.116 0.116
        0.152 0.150 0.166 0.167
        0.247 0.262 0.264 0.264
        0.277 0.329 0.400 0.407
        0.213 0.348 0.355 0.342
    8 0.213 0.348 0.355 0.342
    9}00.213 0.348 0.355 0.342
    10 0.213 0.348 0.355 0.342
    11}00.213 0.348 0.355 0.342
    12 0.213 0.348 0.355 0.342
```

Table 8.8.3.1. NE Atlantic Mackerel. Short-term prediction: INPUT DATA

|  |  | $\Sigma$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2022 |  |  |  |  |  |  |  |  |
| 0 | 4497060 | 0.15 | 0.000 | 0.000 | 0.312 | 0.000 | 0.001 | 0.054 |
| 1 | 5025194 | 0.15 | 0.110 | 0.433 | 0.312 | 0.068 | 0.009 | 0.150 |
| 2 | 4317473 | 0.15 | 0.541 | 0.152 | 0.312 | 0.195 | 0.049 | 0.274 |
| 3 | 5459511 | 0.15 | 0.922 | 0.152 | 0.312 | 0.254 | 0.115 | 0.319 |
| 4 | 1413880 | 0.15 | 0.999 | 0.152 | 0.312 | 0.288 | 0.159 | 0.365 |
| 5 | 1344203 | 0.15 | 1.000 | 0.284 | 0.312 | 0.328 | 0.262 | 0.394 |
| 6 | 989652 | 0.15 | 1.000 | 0.284 | 0.312 | 0.354 | 0.335 | 0.429 |
| 7 | 350945 | 0.15 | 1.000 | 0.284 | 0.312 | 0.378 | 0.302 | 0.447 |
| 8 | 542972 | 0.15 | 1.000 | 0.284 | 0.312 | 0.394 | 0.302 | 0.463 |
| 9 | 399065 | 0.15 | 1.000 | 0.284 | 0.312 | 0.406 | 0.302 | 0.476 |
| 10 | 557747 | 0.15 | 1.000 | 0.284 | 0.312 | 0.432 | 0.302 | 0.498 |
| 11 | 472776 | 0.15 | 1.000 | 0.284 | 0.312 | 0.445 | 0.302 | 0.516 |
| 12+ | 581817 | 0.15 | 1.000 | 0.284 | 0.312 | 0.488 | 0.302 | 0.546 |
| 2023 |  |  |  |  |  |  |  |  |
| 0 | 4497060 | 0.15 | 0.000 | 0.000 | 0.312 | 0.000 | 0.001 | 0.054 |
| 1 | - | 0.15 | 0.110 | 0.433 | 0.312 | 0.068 | 0.009 | 0.150 |
| 2 | - | 0.15 | 0.541 | 0.152 | 0.312 | 0.195 | 0.049 | 0.274 |
| 3 | - | 0.15 | 0.922 | 0.152 | 0.312 | 0.254 | 0.115 | 0.319 |
| 4 | - | 0.15 | 0.999 | 0.152 | 0.312 | 0.288 | 0.159 | 0.365 |
| 5 | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.328 | 0.262 | 0.394 |
| 6 | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.354 | 0.335 | 0.429 |
| 7 | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.378 | 0.302 | 0.447 |
| 8 | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.394 | 0.302 | 0.463 |
| 9 | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.406 | 0.302 | 0.476 |
| 10 | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.432 | 0.302 | 0.498 |
| 11 | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.445 | 0.302 | 0.516 |
| 12+ | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.488 | 0.302 | 0.546 |
| 2024 |  |  |  |  |  |  |  |  |
| 0 | 4497060 | 0.15 | 0.000 | 0.000 | 0.312 | 0.000 | 0.001 | 0.054 |
| 1 | - | 0.15 | 0.110 | 0.433 | 0.312 | 0.068 | 0.009 | 0.150 |
| 2 | - | 0.15 | 0.541 | 0.152 | 0.312 | 0.195 | 0.049 | 0.274 |
| 3 | - | 0.15 | 0.922 | 0.152 | 0.312 | 0.254 | 0.115 | 0.319 |
| 4 | - | 0.15 | 0.999 | 0.152 | 0.312 | 0.288 | 0.159 | 0.365 |
| 5 | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.328 | 0.262 | 0.394 |
| 6 | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.354 | 0.335 | 0.429 |


|  |  | $\Sigma$ | $\frac{\lambda}{\sum_{3}^{2}}$ |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7 | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.378 | 0.302 | 0.447 |
| 8 | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.394 | 0.302 | 0.463 |
| 9 | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.406 | 0.302 | 0.476 |
| 10 | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.432 | 0.302 | 0.498 |
| 11 | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.445 | 0.302 | 0.516 |
| 12+ | - | 0.15 | 1.000 | 0.284 | 0.312 | 0.488 | 0.302 | 0.546 |

Table 8.8.3.2. NE Atlantic Mackerel. Short-term prediction: Multi-option table for 1131416 t catch in 2022 and a range of F-values in 2023.

| 2022 |  |  |  |
| :--- | :--- | :--- | :--- |
| TSB | SSB | F bar | Catch |
| 5011502 | 3769326 | 0.361 | 1131416 |


| 2023 | 2024 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TSB | SSB | Fbar | Catch | TSB | SSB | Implied change in the catch |
| 4703613 | 3834187 | 0 | 0 | 5260647 | 4442903 | -100\% |
| - | 3827916 | 0.01 | 33188 | 5232934 | 4408097 | -97\% |
| - | 3821661 | 0.02 | 66109 | 5205449 | 4373667 | -94\% |
| - | 3815422 | 0.03 | 98764 | 5178190 | 4339610 | -91\% |
| - | 3809199 | 0.04 | 131156 | 5151155 | 4305920 | -88\% |
| - | 3802991 | 0.05 | 163288 | 5124341 | 4272593 | -86\% |
| - | 3796799 | 0.06 | 195161 | 5097748 | 4239624 | -83\% |
| - | 3790623 | 0.07 | 226780 | 5071371 | 4207009 | -80\% |
| - | 3784463 | 0.08 | 258145 | 5045210 | 4174744 | -77\% |
| - | 3778318 | 0.09 | 289260 | 5019263 | 4142824 | -74\% |
| - | 3772188 | 0.1 | 320127 | 4993527 | 4111245 | -72\% |
| - | 3766074 | 0.11 | 350747 | 4968000 | 4080003 | -69\% |
| - | 3759975 | 0.12 | 381124 | 4942680 | 4049094 | -66\% |
| - | 3753892 | 0.13 | 411260 | 4917566 | 4018513 | -64\% |
| - | 3747824 | 0.14 | 441156 | 4892655 | 3988258 | -61\% |
| - | 3741771 | 0.15 | 470816 | 4867945 | 3958323 | -58\% |


| - | 3735734 | 0.16 | 500241 | 4843435 | 3928704 | -56\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| - | 3729712 | 0.17 | 529434 | 4819123 | 3899399 | -53\% |
| - | 3723704 | 0.18 | 558397 | 4795006 | 3870403 | -51\% |
| - | 3717712 | 0.19 | 587132 | 4771083 | 3841712 | -48\% |
| - | 3711735 | 0.2 | 615640 | 4747353 | 3813322 | -46\% |
| - | 3705773 | 0.21 | 643925 | 4723812 | 3785231 | -43\% |
| - | 3699826 | 0.22 | 671989 | 4700460 | 3757434 | -41\% |
| - | 3693894 | 0.23 | 699832 | 4677295 | 3729928 | -38\% |
| - | 3687977 | 0.24 | 727458 | 4654314 | 3702709 | -36\% |
| - | 3682074 | 0.25 | 754869 | 4631517 | 3675774 | -33\% |
| - | 3676187 | 0.26 | 782066 | 4608901 | 3649119 | -31\% |
| - | 3670314 | 0.27 | 809052 | 4586464 | 3622741 | -28\% |
| - | 3664456 | 0.28 | 835827 | 4564206 | 3596637 | -26\% |
| - | 3658612 | 0.29 | 862396 | 4542124 | 3570803 | -24\% |
| - | 3652783 | 0.3 | 888758 | 4520217 | 3545237 | -21\% |
| - | 3646969 | 0.31 | 914917 | 4498482 | 3519934 | -19\% |
| - | 3641169 | 0.32 | 940874 | 4476920 | 3494891 | -17\% |
| - | 3635384 | 0.33 | 966631 | 4455527 | 3470106 | -15\% |
| - | 3629613 | 0.34 | 992190 | 4434302 | 3445576 | -12\% |
| - | 3623856 | 0.35 | 1017552 | 4413244 | 3421297 | -10\% |
| - | 3618114 | 0.36 | 1042720 | 4392350 | 3397266 | -8\% |
| - | 3612386 | 0.37 | 1067695 | 4371621 | 3373481 | -6\% |
| - | 3606673 | 0.38 | 1092479 | 4351053 | 3349939 | -3\% |
| - | 3600973 | 0.39 | 1117074 | 4330646 | 3326635 | -1\% |
| - | 3595288 | 0.4 | 1141482 | 4310398 | 3303569 | 1\% |
| - | 3589617 | 0.41 | 1165704 | 4290308 | 3280736 | 3\% |
| - | 3583960 | 0.42 | 1189741 | 4270373 | 3258135 | 5\% |
| - | 3578317 | 0.43 | 1213597 | 4250594 | 3235762 | 7\% |
| - | 3572688 | 0.44 | 1237271 | 4230967 | 3213615 | 9\% |
| - | 3567073 | 0.45 | 1260767 | 4211492 | 3191690 | 11\% |


| - | 3561472 | 0.46 | 1284085 | 4192167 | 3169986 | 13\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| - | 3555885 | 0.47 | 1307228 | 4172992 | 3148499 | 16\% |
| - | 3550312 | 0.48 | 1330196 | 4153964 | 3127228 | 18\% |
| - | 3544752 | 0.49 | 1352992 | 4135082 | 3106169 | 20\% |
| - | 3539207 | 0.5 | 1375617 | 4116345 | 3085321 | 22\% |
| - | 3533675 | 0.51 | 1398072 | 4097752 | 3064680 | 24\% |
| - | 3528157 | 0.52 | 1420360 | 4079300 | 3044244 | 26\% |
| - | 3522652 | 0.53 | 1442481 | 4060990 | 3024010 | 27\% |
| - | 3517161 | 0.54 | 1464437 | 4042819 | 3003978 | 29\% |
| - | 3511684 | 0.55 | 1486230 | 4024787 | 2984143 | 31\% |
| - | 3506220 | 0.56 | 1507861 | 4006892 | 2964504 | 33\% |
| - | 3500769 | 0.57 | 1529332 | 3989132 | 2945058 | 35\% |
| - | 3495332 | 0.58 | 1550643 | 3971507 | 2925803 | 37\% |
| - | 3489909 | 0.59 | 1571797 | 3954015 | 2906737 | 39\% |
| - | 3484499 | 0.6 | 1592795 | 3936656 | 2887858 | 41\% |
| - | 3479102 | 0.61 | 1613639 | 3919427 | 2869163 | 43\% |
| - | 3473718 | 0.62 | 1634329 | 3902328 | 2850651 | 44\% |
| - | 3468348 | 0.63 | 1654867 | 3885357 | 2832319 | 46\% |
| - | 3462991 | 0.64 | 1675254 | 3868514 | 2814165 | 48\% |
| - | 3457647 | 0.65 | 1695492 | 3851797 | 2796187 | 50\% |
| - | 3452316 | 0.66 | 1715583 | 3835205 | 2778383 | 52\% |
| - | 3446998 | 0.67 | 1735527 | 3818737 | 2760751 | 53\% |
| - | 3441693 | 0.68 | 1755325 | 3802392 | 2743289 | 55\% |
| - | 3436402 | 0.69 | 1774980 | 3786168 | 2725995 | 57\% |
| - | 3431123 | 0.7 | 1794492 | 3770065 | 2708867 | 59\% |
| - | 3425857 | 0.71 | 1813863 | 3754081 | 2691903 | 60\% |
| - | 3420605 | 0.72 | 1833093 | 3738216 | 2675101 | 62\% |
| - | 3415365 | 0.73 | 1852185 | 3722468 | 2658460 | 64\% |
| - | 3410137 | 0.74 | 1871139 | 3706836 | 2641977 | 65\% |
| - | 3404923 | 0.75 | 1889957 | 3691320 | 2625652 | 67\% |


| - | 3399721 | 0.76 | 1908639 | 3675917 | 2609481 | 69\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| - | 3394533 | 0.77 | 1927188 | 3660628 | 2593463 | 70\% |
| - | 3389356 | 0.78 | 1945604 | 3645451 | 2577596 | 72\% |
| - | 3384193 | 0.79 | 1963888 | 3630385 | 2561880 | 74\% |
| - | 3379042 | 0.8 | 1982041 | 3615429 | 2546311 | 75\% |
| - | 3373903 | 0.81 | 2000065 | 3600582 | 2530888 | 77\% |
| - | 3368778 | 0.82 | 2017962 | 3585844 | 2515611 | 78\% |
| - | 3363664 | 0.83 | 2035731 | 3571212 | 2500476 | 80\% |
| - | 3358563 | 0.84 | 2053374 | 3556687 | 2485482 | 81\% |
| - | 3353475 | 0.85 | 2070892 | 3542267 | 2470628 | 83\% |
| - | 3348399 | 0.86 | 2088287 | 3527952 | 2455912 | 85\% |
| - | 3343335 | 0.87 | 2105559 | 3513740 | 2441333 | 86\% |
| - | 3338283 | 0.88 | 2122709 | 3499630 | 2426889 | 88\% |
| - | 3333244 | 0.89 | 2139739 | 3485622 | 2412578 | 89\% |
| - | 3328217 | 0.9 | 2156649 | 3471714 | 2398399 | 91\% |
| - | 3323202 | 0.91 | 2173442 | 3457907 | 2384351 | 92\% |
| - | 3318200 | 0.92 | 2190116 | 3444198 | 2370432 | 94\% |
| - | 3313209 | 0.93 | 2206675 | 3430587 | 2356641 | 95\% |
| - | 3308231 | 0.94 | 2223118 | 3417074 | 2342975 | 96\% |
| - | 3303264 | 0.95 | 2239446 | 3403657 | 2329434 | 98\% |
| - | 3298310 | 0.96 | 2255662 | 3390335 | 2316017 | 99\% |
| - | 3293368 | 0.97 | 2271765 | 3377108 | 2302722 | 101\% |
| - | 3288437 | 0.98 | 2287757 | 3363974 | 2289547 | 102\% |
| - | 3283519 | 0.99 | 2303638 | 3350934 | 2276491 | 104\% |
| - | 3278612 | 1 | 2319410 | 3337986 | 2263554 | 105\% |
| - | 3273718 | 1.01 | 2335073 | 3325129 | 2250733 | 106\% |
| - | 3268835 | 1.02 | 2350629 | 3312362 | 2238027 | 108\% |
| - | 3263964 | 1.03 | 2366078 | 3299685 | 2225435 | 109\% |
| - | 3259104 | 1.04 | 2381422 | 3287098 | 2212956 | 110\% |
| - | 3254257 | 1.05 | 2396660 | 3274598 | 2200589 | 112\% |


| - | 3249421 | 1.06 | 2411795 | 3262186 | 2188331 | $113 \%$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| - | 3244596 | 1.07 | 2426827 | 3249860 | 2176183 | $114 \%$ |
| - | 3239783 | 1.08 | 2441756 | 3237621 | 2164143 | $116 \%$ |
| - | 3234982 | 1.09 | 2456585 | 3225466 | 2152209 | $117 \%$ |

Table 8.8.3.3. NE Atlantic Mackerel. Short-term prediction: Management option table for 1131416 t catch in 2022 and a range of catch options in 2023.

| Rationale | $\begin{aligned} & \text { Catch } \\ & \text { (2023) } \end{aligned}$ | $\mathrm{F}_{\text {bar }}$ <br> (2023) | $\begin{aligned} & \text { SSB } \\ & \text { (2023) } \end{aligned}$ | $\begin{aligned} & \text { SSB } \\ & \text { (2024) } \end{aligned}$ | $\begin{aligned} & \text { \% SSB } \\ & \text { change * } \end{aligned}$ | \% catch change ** | \% advice change *** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Catch(2023) $=$ Zero | 0 | 0.000 | $\begin{aligned} & 38341 \\ & 87 \end{aligned}$ | $\begin{aligned} & 444290 \\ & 3 \end{aligned}$ | 2\% | -100\% | -100\% |
| $\begin{aligned} & \text { Catch }(2023)=2022 \text { catch } \\ & -20 \% \end{aligned}$ | 905133 | 0.306 | $\begin{aligned} & 36491 \\ & 47 \end{aligned}$ | $\begin{aligned} & 349483 \\ & 9 \end{aligned}$ | -3\% | -20\% | 14\% |
| Catch $(2023)=2022$ catch | 1131416 | 0.396 | $\begin{aligned} & 35976 \\ & 36 \end{aligned}$ | $\begin{aligned} & 333305 \\ & 1 \end{aligned}$ | -5\% | 0\% | 42\% |
| $\begin{aligned} & \text { Catch(2023) }=2022 \text { catch } \\ & +25 \% \end{aligned}$ | 1414270 | 0.517 | $\begin{aligned} & 35296 \\ & 67 \end{aligned}$ | $\begin{aligned} & 313112 \\ & 8 \end{aligned}$ | -6\% | 25\% | 78\% |
| Fbar(2023) $=0.26$ (Fmsy) | 782066 | 0.260 | $\begin{aligned} & 36761 \\ & 87 \end{aligned}$ | $\begin{aligned} & 364911 \\ & 9 \end{aligned}$ | -2\% | -31\% | -2\% |
| $\operatorname{Fbar}(2023)=0.36(\mathrm{Fpa})$ | 1042720 | 0.360 | $\begin{aligned} & 36181 \\ & 14 \end{aligned}$ | $\begin{aligned} & 339726 \\ & 6 \end{aligned}$ | -4\% | -8\% | 31\% |
| Fbar(2023) $=0.46($ Flim $)$ | 1284085 | 0.460 | $\begin{aligned} & 35614 \\ & 72 \end{aligned}$ | $\begin{aligned} & 316998 \\ & 6 \end{aligned}$ | -6\% | 13\% | 62\% |

* SSB 2023 relative to SSB 2022.
** Catch in 2023 relative to assumed catches in 2022 (1 131 416t). There is no internationally agreed TAC for 2022.
*** Catch in 2023 relative to the advice value for 2022 ( 794920 t ).


### 8.14 Figures



Figure 8.4.2.1. NE Atlantic Mackerel. Commercial catches in 2021, quarter 1.


Figure 8.4.2.2. NE Atlantic Mackerel. Commercial catches in 2021, quarter 2.


Figure 8.4.2.3. NE Atlantic Mackerel. Commercial catches in 2021, quarter 3.


Figure 8.4.2.4. NE Atlantic Mackerel. Commercial catches in 2021, quarter 4.


Figure 8.5.2.1. NE Atlantic mackerel. Weights-at-age in the catch.
mean weight in the stock


Figure 8.5.2.2. NE Atlantic mackerel. Weights-at-age in the stock.


Figure 8.5.3.1. NE Atlantic mackerel. Proportion of mature fish at age.


Figure 8.6.1.1.1. Mackerel egg production by half rectangle for period 3 (Mar $4^{\text {th }}-A p r 8^{\text {th }}$ ). Circle areas and colour scale represent mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 8.6.1.1.2. Mackerel egg production by half rectangle for period 4 (Apr $9^{\text {th }}-29^{\text {th }}$ ). Circle areas and colour scale represent mackerel stage leggs/m2/day by half rectangle. Crosses represent zero values.


Figure 8.6.1.1.3. Mackerel egg production by half rectangle for period 5 (Apr 30 ${ }^{\text {th }}-\mathrm{May} 31^{\text {st }}$ ). Circle areas and colour scale represent mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 8.6.1.1.4. Mackerel egg production by half rectangle for period 6 (June $1^{\text {st }}-30^{\text {th }}$ ). Circle areas and colour scale represent mackerel stage I eggs $/ \mathrm{m}^{2} /$ day by half rectangle. Crosses represent zero values.


Figure 8.6.1.5. Mackerel egg production by half rectangle for period 7 (July $1^{\text {st }}-31^{\text {st }}$ ). Circle areas and colour scale represent mackerel stage leggs/m2/day by half rectangle. Crosses represent zero values.

Western mackerel


Figure 8.6.1.1.6. Provisional annual egg production curve for mackerel in the western component in 2022, (black line). The curves for 2007, 201020132016 and 2019 are included for comparison.


Figure 8.6.1.1.7. Provisional annual egg production curve for mackerel in the southern component in 2022, (black line). The curves for 2007, 201020132016 and 2019 are included for comparison.


Figure 8.6.1.1.8. Combined mackerel TAEP estimates ( ${ }^{*} 10^{13}$ ) - 1992 - 2022.


Fig. 8.6.1.5.1. Stage 1a mackerel egg production (eggs/m2/day) by half rectangle for NSMEGS 2022. Circle areas and colour scale represent mackerel stage $1 \mathrm{eggs} / \mathrm{m} 2 /$ day by half rectangle. Crosses represent zero values.


Figure 8.6.1.6.1. Stage la mackerel egg production (eggs/m²/day) by half rectangle for NSMEGS 2021. Purple circles represent observed values, black circles represent interpolated values, and crosses represent observed zeros.


Figure 8.6.2.1. Spatial distribution of mackerel juveniles at age 0 in October to March. Average for cohorts from 19982020. Mackerel squared catch rates by trawl haul (circle areas represent catch rates in $\mathrm{kg} / \mathrm{km} 2$ ) overlaid on modelled squared catch rates per $10 \times 10 \mathrm{~km}$ rectangle. Each rectangle is coloured according to the expected squared catch rate in percent of the highest value for that year. See Jansen et al. (2015) for details. Red box indicates the approximate typical coverage of the IBTS-NS Q1 and SWC Q1 surveys.


Figure 8.6.2.2. Index of mackerel juveniles at age 0 in October to March proxied by annual integration of square root of expected catch in demersal trawl surveys (Blue lines). See Jansen et al. (2015) for details. * Rescaled


Figure 8.6.3.1. Estimated total stock numbers (TSN) of mackerel from IESSNS calculated using StoX for the years 2007 and from 2010 to 2022. Displayed is StoX baseline estimate (red dot) and a bootstrap estimate (black dot), calculated using 1000 replicates, with $90 \%$ confidence intervals (vertical line) based on the bootstrap. Analysis excludes the North Sea and survey coverage was incomplete in 2007 and 2011. For 2022, index value is also calculated excluding the two extreme catches (filled blue diamond) including $\mathbf{9 0 \%}$ confidence interval.


Figure 8.6.3.2. Estimated total stock biomass of mackerel from IESSNS calculated using StoX for the years 2007 and from 2010 to 2022. Displayed is StoX baseline estimate (red dot) and a bootstrap estimate (black dot), calculated using 1000 replicates, with $90 \%$ confidence intervals (vertical line) based on the bootstrap. Analysis excludes the North Sea and
survey coverage was incomplete in 2007 and 2011. For 2022, index value is also calculated excluding the two extreme catches (filled blue diamond) including 90\% confidence interval.


Figure 8.6.3.3. Internal consistency of the mackerel abundance index from the IESSNS survey including data from 2012 to 2022, excluding North Sea. Ages indicated by white numbers in grey diagonal cells. Statistically significant positive correlations ( $p<0.05$ ) are indicated by regression lines and red cells in upper left half. Correlation coefficients ( $r$ ) are given in the lower right half.


Figure 8.6.3.4. Mackerel catch curves from the estimate stock size at age from the IESSNS in 2010 and from 2012 to 2022, excluding the North Sea. Each cohort is marked by a uniquely coloured line that connects the estimates indicated by the respective ages.


Figure 8.6.3.5. Mackerel numbers by age from the IESSNS survey in 2022, excluding North Sea. Boxplot of abundance and relative standard error (CV) obtained by bootstrapping with 1000 replicates using StoX version 3.5.0.


Figure 8.6.3.6. Mackerel catch rates from predetermined surface trawl stations (circle size represents catch rate in $\mathrm{kg} / \mathrm{km} 2$ ) overlaid on mean catch rate per standardized rectangle ( $2^{\circ}$ lat. $\times 4^{\circ}$ lon.) from the 2022 IESSNS, including North Sea. Zero mackerel catches are displayed as grey crosses.


Figure 8.6.3.7. Mackerel annual distribution proxied by the absolute distribution of mean mackerel catch rates per standardized rectangles ( $2^{\circ}$ lat. x $4^{\circ}$ lon.), from predetermined surface trawl stations from IESSNS in 2010 to 2022, including North Sea. Colour scale goes from white $(=0)$ to red (= maximum value for the given year).


Figure 8.6.4.1. Number and distribution of RFID tagged mackerel from experiments west of Ireland and British Isles during 2011-2022. Note that data from releases 2011-2012 are not used in the stock assessment, based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019c), and data from experiments in 2021-2022 are not included as there are no full years with recaptures yet.


Figure 8.6.4.2. Biomass and distribution of catches scanned for RFID tagged mackerel during 2012-2021. Note that data from scanned catches in 2012-2013 are not used in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019c).


Figure 8.6.4.3. Distribution of recaptures of RFID tagged mackerel during 2012-2021. Note that data on recaptures in 2012-2013 are not used in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019c).

8.6.4.4. Overview of the relative year class distributions among RFID tagged mackerel per release year from experiments west of Ireland and British Isles in May-June compared with scanned and recaptured fish in year 1 and 2 after release of the same year classes. Note that data from releases in 2011-2012 are not used in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019c). Note also that it was decided to only use ages 5-11 in updated assessments, and limits for this age span is marked (vertical grey dotted lines) for each release year.


Figure 8.6.4.5. ( $A$ ) Trends in year class abundance ( $N=$ numbers released/numbers recaptured*numbers scanned) from RFID tag-recapture data based on aggregated data on recaptures and scanned numbers in year 1 and 2 after each release year. Data excluded in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019c), release years 2011-2012 and ages 2-4 and 12+, are marked with dotted lines in year class trends. (B) Trends in various age aggregated biomass indices from RFID tag-recapture data compared with the SSB ( $\pm 95$ confidence intervals) from the WGWIDE 2022 stock assessment. Data are based on a combination of estimated numbers by year class (A) scaled by survival parameter ( $\mathbf{0} \mathbf{1 5 9 9 \text { ) and weight at age in stock from WGWIDE 2022. Vertical dotted line marks the starting year }}$ where RFID tagging experiments are used in the stock assessment. Note that final year with RFID biomass estimates in 2020 is only based on recapture year 2021 and will likely change when adding recapture year 2022 in WGWIDE 2023.


Figure 8.6.4.6. Signals of total mortality rate (Z). (A) Trends in abundance of year classes 2003-2014 from unscaled input data (RFID, IESSNS and catches) and the WGWIDE2022 stock assessment. The estimated slope of decrease from the age 4 when it is fully recruited to the spawning stock until age 12 is interpreted as signal $Z$, grey dotted lines is $Z=0.4$. ( $B$ ) The estimated year class differences in $Z$ (with $95 \%$ confidence intervals), and corresponding differences between the various data sources.


Figure 8.7.2.1. NE Atlantic mackerel. Parameter estimates from the SAM model (and associated confidence intervals) for the WGWIDE 2022 update assessment. top left : estimated standard deviation for the observation errors, top centre : estimated overdispersion for the errors on the tag recaptures, top right : standard deviation for the processes, bottom : survey catchabilities and post-release survival of tagged fish.


Figure 8.7.2.2. NE Atlantic mackerel. Parameter uncertainty (standard deviation of estimate) versus parameter value for the observation variances.


Figure 8.7.2.3. NE Atlantic mackerel. Estimated AR1 error correlation structure for the observations from the IESSNS survey age 3 to 11 .


Figure 8.7.2.4. NE Atlantic mackerel. Correlation between parameter estimates from the SAM model for the WGWIDE 2022 update assessment


Figure 8.7.2.5. NE Atlantic mackerel. One Step Ahead Normalized residuals for the fit to the catch data (catch data prior to $\mathbf{2 0 0 0}$ in blue rectangle were not used to fit the model). Blue circles indicate positive residuals (observation larger than predicted) and filled red circles indicate negative residuals.


Figure 8.7.2.6. NE Atlantic mackerel. Empirical correlations between ages in the One Step Ahead residuals for the catch-at-age data.


Figure 8.7.2.7. NE Atlantic mackerel. Empirical correlations between ages in the One Step Ahead residuals for the IESSNS abundances-at-age.
-3.3-4


Figure 8.7.2.8. NE Atlantic mackerel. One step ahead residuals for the fit to the recaptures of tags in the final assessment. The x-axis represents the release year, and the $y$-axis is the number of years between tagging and recapture. Each panel correspond to a given age at release. Blue circles indicate positive residuals (observation larger than predicted) and filled red circles indicate negative residuals.


Figure 8.7.2.9. NE Atlantic mackerel. Leave one out assessment runs. SAM estimates of SSB , Fbar and recruitment, for assessments runs leaving out one of the observation data sets.


Figure 8.7.2.10. NE Atlantic mackerel. Sensitivity of the estimated stock trajectories to a $9.4 \%$ reduction of the 2022 SSB index from the egg survey (proportion of the total annual egg production corresponding to the interpolation done for the missing coverage in period 6 of the survey).


Figure 8.7.2.11. NE Atlantic mackerel. Sensitivity of the estimated stock trajectories to a $\mathbf{5 0 \%}$ increase in the observation error variance for the catches-at-age for the year 2021 (to account for a potential higher uncertainty due to the lack of data from Russia).


Figure 8.7.3.1. NE Atlantic mackerel. Perception of the NEA mackerel stock, showing the SSB, $\mathrm{F}_{\text {bar }} 4-8$ and recruitment (with $95 \%$ confidence intervals) from the SAM assessment.


Figure 8.7.3.2. NE Atlantic mackerel. Estimated selectivity for the period 1990 to 2021, calculated as the ratio of the estimated fishing mortality-at-age and the Fbar4-8 value in the corresponding year.


Figure 8.7.4.1. NE Atlantic mackerel. Uncertainty (standard deviation of the log values) of the estimates of SSB and Fbar from the SAM for the 2021 and 2022 WGWIDE assessments.


Figure 8.7.4.2. NE Atlantic mackerel. Analytical retrospective patterns (8 years back) of SSB, $\mathrm{F}_{\mathrm{bar}} 4-8$ and recruitment from the WGWIDE 2022 update assessment. the Mohn's rho values are calculated based on 5 retro years.


Figure 8.7.4.3. NE Atlantic mackerel. Process error expressed as annual deviations of abundances at age, for the 2022 WGWIDE assessment and from the 2021 WGWIDE assessment.


Figure 8.7.4.4. NE Atlantic mackerel. Model process error expressed in biomass cumulated across age-group for the 2022 WGWIDE assessment and for the 2021 WGWIDE assessment.


Figure 8.7.5.1. NE Atlantic mackerel. Model comparison of the cohort signal based on SAM estimates at age 0,2 and 3.


Figure 8.7.5.2. NE Atlantic mackerel. Model. comparison of the perception of the stocks from the WGWIDE 2022 assessment, and the assessment starting at age2.


Figure 8.10.1. NE Atlantic mackerel. Comparison of the stock trajectories from the WGWIDE 2021 (blue) and 2022 (red) update assessments.

scaling
parameters


Figure 8.10.2. NE Atlantic mackerel. Comparison of model parameters and their uncertainty for the 2022 WGWIDE and the 2021 WGWIDE assessment

Advice, TAC and catch


Figure 8.11.1. NE Atlantic mackerel. Top: comparison of the ICES advice, the agreed TAC, the sum of the unilateral quota and total catch. Bottom: calculated percentage of TAC over Advice, Sum of unilateral quota over Advice, Catch over Advice and Catch over Sum of unilateral quota.

## 9 Red gurnard in the Northeast Atlantic

### 9.1 General biology

The main biological features known for red gurnard (Aspitrigla (Chelidonichthys) cuculus) are described in the stock annex. This species is widely distributed in the North-east Atlantic from South Norway and North of the British Isles to Mauritania, on grounds between 20 and 250 m . This benthic species is abundant in the Channel (7de), the shelf West of Brittany (7h, 8a), and west of Scotland (6a), living on gravel or coarse sand. In the Channel, the size at first maturity is $\sim 25 \mathrm{~cm}$ at 3 years old (Dorel, 1986).

### 9.2 Stock identity and possible assessments areas

A compilation of datasets from bottom-trawl surveys undertaken within the project 'Atlas of the marine fishes of the northern European shelf' has produced a distribution map of red gurnard. Higher occurrences of red gurnard with patchy distribution have been observed along the Western approaches from the Shetlands Islands to the Celtic Seas and the Channel.

A continuous distribution of fish crossing the Channel and the area West of Brittany does not suggest a separation of the divisions 7 d from 7 e and 7 h . Therefore, a split of the population between these Ecoregions does not seem appropriate. Divergent trends in survey abundances have been observed within the assessment area, with a sustained spike in abundance in division 6a in the early 2010's which is not seen in surveys covering SA 7-8. Further investigations, such as morphometric studies, tagging and genetic population studies, would be needed to progress on stock boundaries, however SIMWG has advised that for now, there is insufficient evidence to carry out assessments on smaller spatial units.

### 9.3 Management regulations

Currently no technical measures are specifically applied to red gurnard or other gurnard species. The exploitation of red gurnard is submitted to the general regulation in the areas where they are caught. There is no minimum landing size set.

### 9.4 Fisheries data

Red gurnard is mainly landed as by-catch by demersal trawlers in mixed fisheries, predominantly in divisions 7d, 7e and 7h (tables 9.1-2). High discard rates and lack of resolution at a species level make interpretation of spatial trends in catches in other areas problematic.

### 9.4.1 Historical landings

Official landings of red gurnard reported to ICES are presented in tables 9.1 and 9.2. Before 1977, red gurnard was not specifically reported. Landings of gurnards are still not always reported at a species level, but rather as mixed gurnards (GUX). Use of this code is not consistent across countries reporting landings, and some report mixed landings of gurnards under the code GUU - which is unfortunately the FAO species code for Tub gurnard (Chelidonichthys lucerna).

A questionnaire was circulated to WGCATCH to gather information on how landings of gurnards are assigned to species. For those countries who responded, only Portugal has presented information on how the reporting of landings at a species level is achieved. Other countries accept the species code as declared at the point of landing, without further validation. This makes interpretation of the records of official landings, and trends within them difficult. Landings of gurnards (red, grey, tub and mixed) are shown in fig. 9.1.

International landings have fluctuated between 3452-5171 tonnes between 2006-2019. Landings in the most recent year (2021) were 2903 tonnes - the lowest on record. France is the main contributor of 'red gurnard' landings, with around $80 \%$ of landings from subarea $7 \mathrm{~d}-\mathrm{h}$ (Celtic Sea/English Channel). In the North Sea red gurnard landings are variable, but roughly evenly distributed between divisions $4 \mathrm{a}, \mathrm{b}$ and c . Landings from the west of Scotland and Ireland, and the Irish Sea (subarea $6 a-b, 7 a-c, 7 j$ ) and Bay of Biscay (division 8) have been consistently low.

### 9.4.2 Discards

Discard data for red gurnard has been provided for 2015-2021 through Intercatch (table 9.3). For those countries which provided data, discard rates are variable but high (table 9.4). Given the uncertainty associated with landings data, these figures should be treated with caution.

### 9.5 Survey data

Information on gurnard abundance are available in DATRAS for a number of surveys. Those covering the core area of the stock as determined by WKWEST (ICES, 2021) are the Scottish West Coast Groundfish Survey (SCOWCGFS and SC-IBTS), Irish Groundfish Survey (IGFS), English Channel Beam Trawl Survey (BTS), and French Q4 surveys EVHOE in the Celtic Sea and Bay of Biscay and CGFS in division 7d. Each of these surveys covers a specific area of red gurnard distribution and no individual survey covers the entire stock area. Lengths at age are available from CGFS-Q4 and (for some years) from IGFS.

## SCO-WCGFS and SC-IBTS.

Before 1996, red gurnard was scarce on the west of Scotland. The CPUE trended strongly upwards after 1997, reaching a peak in 2013, before declining to around the series average in recent years. The value for 2020 was sharply up on 2019, however it fell back in 2021(figs 9.2-3).

## CGFS (Q4).

Over the time-series 1988-2011, CPUE has fluctuated, peaked in 1994, reached a low in 2011, but is above long term mean since 2016. Values in 2021 were down (fig 9.4).

## EVHOE (Q4).

Over the period 1997-2020, the CPUE has fluctuated. It has been on an increasing trend since 2017, and 2020 is the second highest value in the series. Age reading of red gurnards caught during the EVHOE survey has been carried out in 2006 and routinely since 2008. They indicate that the individuals caught are mainly of age 1 and 2. Values in 2021 were down (fig. 9.4).

## IGFS.

The CPUE of red gurnard in the IGFS series has varied around the series mean without trend between 2002 and 2020. Values in 2021 were down (fig 9.5).

## BTS.

CPUE in this relatively short series has fluctuated without apparent trend since 2006. Values in 2021 were down (fig 9.5).

### 9.6 Biological sampling

Number-at-length information was provided by French and Portuguese landings and discards. There remains a lack of regular sampling for red gurnard in commercial landings and discarding to provide series of length or age compositions usable for a preliminary analytical assessment.

### 9.7 Biological parameters and other research

There is no update of growth parameters and available parameters from several authors are summarized in the Stock Annex. They vary widely. Available length-weight relationships are also shown in Stock Annex. Natural mortality has not been estimated in the areas studied at this Working Group. Accurate estimates of landings are still lacking for this species.

### 9.8 Assessment

Having explored the trends in available survey data, the delta-lognormal assessment method developed during WKWEST (ICES, 2021) was applied. This approach extracts the estimates of year effect from the log-normal part of the model (there is no temporal term in the binomial part), together with their associated standard error, and standardises the series relative to its mean value, to provide an index of biomass across the multiple surveys. Goodness of fit metrics of the model remain acceptable (figs 9.6-7) and the log-normal part of the model has an adjusted $\mathrm{r}^{2}$ value of 0.32 .
After a period of relative stability, the biomass indicator declined in 2019, before recovering strongly in 2020 (fig. 9.8). The indicator remains above the biomass limit reference level of 0.81 .

The influence of COVID-19 related disruption to surveys in the Channel during 2020 has not been investigated for this stock.

### 9.9 Data requirements

Gurnards are still not always reported by species, but rather as mixed gurnards. National approaches to validating the species composition of mixed gurnard landings are undocumented, other than for Portuguese landings. This makes interpretation of the records of official landings difficult. An international approach to the collection of data on species composition of gurnard landings is required to support the provision of advice for this stock.

### 9.10 References

Dorel, D. 1986. Poissons de l'Atlantique nord-est relations taille-poids. Institut Francais de Recherche pour l'Exploitation de la Mer. Nantes, France. 165 p.
ICES. 2021. Benchmark Workshop on selected stocks in the Western Waters in 2021 (WKWEST). ICES Scientific Reports. 3:31. 504 pp . https://doi.org/10.17895/ices.pub. 8137

Table 9.1. Red gurnard in subareas 3-8. Official landings by country in tonnes.

|  | $\underset{\sim}{E}$ | $\begin{aligned} & \text { 드̃ } \\ & \text { in } \end{aligned}$ |  | خ | $\begin{aligned} & \text { 訁 } \\ & \text { ù } \\ & \text { L } \\ & 0 \\ & 0 \end{aligned}$ |  | $\sum$ |  | T0 0 士 0 0 | 들 | - |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 313 | 0 | 4552 | 0 | 10 | 0 | 0 | 57 | 125 | 115 | 5172 |
| 2007 | 328 | 0 | 4494 | 1 | 4 | 0 | 0 | 66 | 127 | 156 | 5176 |
| 2008 | 352 | 0 | 4045 | 0 | 8 | 0 | 0 | 92 | 112 | 166 | 4775 |
| 2009 | 227 | 0 | 3310 | 0 | 6 | 0 | 1 | 160 | 150 | 263 | 4117 |
| 2010 | 237 | 0 | 3437 | 0 | 2 | 0 | 0 | 251 | 115 | 362 | 4404 |
| 2011 | 306 | 0 | 3176 | 1 | 2 | 0 | 1 | 295 | 134 | 257 | 4172 |
| 2012 | 306 | 0 | 2706 | 3 | 4 | 26 | 0 | 329 | 148 | 257 | 3779 |
| 2013 | 288 | 576 | 3154 | 3 | 9 | 16 | 2 | 267 | 113 | 329 | 4757 |
| 2014 | 263 | 399 | 3782 | 3 | 6 | 0 | 5 | 241 | 108 | 283 | 5090 |
| 2015 | 187 | 91 | 2919 | 2 | 3 | 0 | 0 | 210 | 122 | 341 | 3875 |
| 2016 | 238 | 87 | 2598 | 3 | 2 | 9 | 1 | 224 | 106 | 381 | 3646 |
| 2017 | 265 | 104 | 2396 | 0 | 1 | 9 | 4 | 226 | 113 | 335 | 3454 |
| 2018 | 314 | 89 | 2968 | 0 | 1 | 1 | 1 | 306 | 114 | 347 | 4141 |
| 2019 | 289 | 84 | 2448 | 1 | 5 | 3 | 0 | 247 | 117 | 478 | 3672 |
| 2020* | 211 | 105 | 2335 | 0 | 0 | 10 | 1 | 235 | 123 | 254 | 3273 |
| 2021* | 123 | 69 | 2240 |  | 0 | 8 | 0 | 160 | 117 | 370 | 2968 |
| 2021** | 90 | 31 | 2251 |  | 0 | 8 | 0 | 158 |  | 365 | 2903 |
| *preliminary data |  |  |  |  |  |  |  |  |  |  |  |

## Table 9.2. Red gurnard in subareas 3-8. Official landings by area in tonnes.

| Year | 4a | 4b | 4c | 5b | 6a | 6b | 7a | 7b | 7c | 7d | 7e | 7f | 7g | 7h | 7j | 7nk | 8a | 8b | 8c | 8d | 9a | 9nk | 10a | 12c | 10nk | 14a | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 13 | 83 | 64 | 0 | 32 | 1 | 11 | 9 | 12 | 1101 | 2803 | 229 | 16 | 446 | 5 | 0 | 153 | 60 | 1 | 5 | 9 | 115 | 0 | 0 | 1 | 0 | 5054 |
| 2007 | 12 | 120 | 55 | 2 | 21 | 0 | 7 | 7 | 15 | 1229 | 2674 | 246 | 15 | 437 | 4 | 0 | 139 | 59 | 3 | 2 | 125 | 0 | 0 | 0 | 2 | 0 | 5174 |
| 2008 | 34 | 64 | 54 | 0 | 28 | 3 | 5 | 7 | 16 | 1236 | 2451 | 249 | 9 | 408 | 5 | 0 | 66 | 24 | 3 | 1 | 109 | 0 | 3 | 0 | 0 | 0 | 4772 |
| 2009 | 58 | 59 | 92 | 0 | 94 | 2 | 4 | 8 | 6 | 1293 | 1557 | 112 | 22 | 510 | 7 | 0 | 98 | 40 | 1 | 3 | 148 | 0 | 1 | 0 | 0 | 0 | 4115 |
| 2010 | 79 | 63 | 86 | 0 | 101 | 46 | 13 | 8 | 10 | 1531 | 1608 | 132 | 23 | 433 | 9 | 0 | 100 | 33 | 0 | 2 | 114 | 0 | 0 | 0 | 1 | 0 | 4392 |
| 2011 | 66 | 29 | 51 | 0 | 69 | 54 | 13 | 5 | 6 | 1295 | 1753 | 124 | 20 | 372 | 9 | 0 | 112 | 46 | 1 | 3 | 133 | 0 | 1 | 0 | 0 | 1 | 4163 |
| 2012 | 83 | 71 | 78 | 0 | 51 | 7 | 8 | 2 | 5 | 1244 | 1441 | 145 | 53 | 294 | 2 | 0 | 83 | 50 | 8 | 1 | 136 | 4 | 1 | 0 | 0 | 1 | 3768 |
| 2013 | 88 | 109 | 60 | 0 | 47 | 0 | 10 | 2 | 6 | 1193 | 1692 | 170 | 58 | 477 | 2 | 0 | 79 | 72 | 532 | 1 | 155 | 0 | 2 | 0 | 0 | 0 | 4755 |
| 2014 | 102 | 52 | 68 | 0 | 47 | 3 | 7 | 1 | 2 | 1294 | 1642 | 115 | 19 | 1069 | 1 | 0 | 82 | 75 | 363 | 3 | 139 | 0 | 3 | 0 | 0 | 0 | 5087 |
| 2015 | 133 | 102 | 53 | 0 | 58 | 1 | 4 | 3 | 1 | 790 | 1553 | 87 | 6 | 703 | 1 | 0 | 95 | 70 | 81 | 2 | 128 | 0 | 2 | 0 | 0 | 0 | 3873 |
| 2016 | 112 | 83 | 117 | 0 | 76 | 1 | 11 | 3 | 1 | 906 | 1270 | 114 | 16 | 608 | 1 | 0 | 87 | 63 | 56 | 1 | 120 | 0 | 1 | 0 | 0 | 0 | 3645 |
| 2017 | 53 | 44 | 90 | 0 | 27 | 1 | 14 | 1 | 0 | 874 | 1424 | 83 | 38 | 473 | 3 | 0 | 78 | 48 | 59 | 1 | 142 | 0 | 1 | 0 | 0 | 0 | 3454 |
| 2018 | 109 | 40 | 113 | 0 | 43 | 0 | 7 | 0 | 0 | 903 | 1785 | 164 | 28 | 631 | 4 | 0 | 80 | 43 | 62 | 2 | 116 | 0 | 1 | 0 | 0 | 0 | 4131 |
| 2019 | 128 | 19 | 75 | 0 | 84 | 0 | 12 | 1 | 0 | 959 | 1516 | 75 | 24 | 477 | 5 | 5 | 73 | 38 | 65 | 0 | 109 | 0 | 0 | 2 | 0 | 0 | 3663 |
| 2020 | 58 | 13 | 65 | 2 | 53 | 4 | 10 | 1 | 4 | 680 | 1504 | 90 | 19 | 425 | 4 | 0 | 69 | 51 | 87 | 1 | 128 | 0 | 0 | 8 | 0 | 0 | 3273 |
| 2021* | 60 | 18 | 75 | 0 | 113 | 4 | 4 | 2 | 1 | 602 | 1390 | 46 | 15 | 471 | 4 | 0 | 62 | 40 | 62 | 0 | 119 | 0 | 1 | 8 | 0 | 0 | 3096 |

*preliminary data

Table 9.3. Red gurnard in subareas 3-8. Discards (t) by country, 2015-2021.

| Country | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| France | 1323 | 2249 | 2232 | 770 | 3132 | 292 | 623 |
| Ireland | 10 | 147 | 93 | 251 | 180 | 76 | 56 |
| Spain | 74 | 286 | 272 | 189 | 122 | 161 | 128 |
| UK (ENG) | 30 | 411 | 198 | 512 | 331 | 117 | 708 |
| UK (SCO) | 649 | 2056 | 3123 | 2795 | 1929 | 4270 | 757 |
| Total |  |  |  |  | 1515 |  |  |

Table 9.4. Red gurnard in subareas 3-8. Discarding of Red gurnard in the Northeast Atlantic, as a percentage of catch, by country, 2017-2021.

| Country | Discard rate (\%) |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
|  | 2017 | 2018 | 2019 | 2020 | 2021 |
| France | 48 | 21 | 56 | 11 | 22 |
| Ireland | 91 | 95 | 95 | 88 | 87 |
| Spain | 72 | 68 | 78 | 91 | 80 |
| UK (England) | 68 | 92 | 67 | 51 | 83 |
| UK (Scotland) |  |  | 45 |  |  |



Figure 9.1. Red gurnard in subareas 3-8. Official landings of red, grey, tub and mixed gurnards from SA3-8, 2006-2019.


Figure 9.2. Red Gurnard in subareas 3-8. Trends in mean abundance ( $\mathrm{kg} / \mathrm{hr}$ ) in the Q1 Scottish IBTS (1985-2010) and Q1 Scottish West Coast Groundfish Survey (2011-2021)


Figure 9.3. Red Gurnard in subareas 3-8. Trends in mean abundance ( $\mathrm{kg} / \mathrm{hr}$ ) in the Q4 Scottish IBTS (1990-2009) and Q4 Scottish West Coast Groundfish Survey (2011-2021)


Figure 9.4. Red Gurnard in subareas 3-8. Trends in mean abundance ( $\mathrm{kg} / \mathrm{hr}$ ) in the EVHOE (top) and French Channel Groundfish Survey (bottom)



Figure 9.5. Red Gurnard in subareas 3-8. Trends in mean abundance ( $\mathrm{kg} / \mathrm{hr}$ ) in the Irish Groundfish Survey (top) and English Channel Beam Trawl Survey (bottom)


Figure 9.6. Red Gurnard in subareas 3-8. Measures of goodness of fit of the lognormal part of the assessment model.


Figure 9.7. Red gurnard in subareas 3-8. Measures of goodness of fit of the binomial part of the assessment model.


Figure 9.8. Red gurnard in subareas 3-8. Results of the assessment model. Shading corresponds to 2 standard errors around the estimate. The dashed line represents MSY $B_{\text {trigger }}(0.81)$.

# 10 Striped red mullet in Subareas and Divisions 6, 7ac, e-k, 8, and 9a 

### 10.1 General biology

Striped red mullet (Mullus surmuletus) is a predominantly benthic species found along the coasts of Europe, southern Norway, and northern Scotland (northern Atlantic, Baltic Sea, North Sea, and the English Channel), up to the Northern part of West Africa, in the Mediterranean Basin, and in the Black Sea (Mahe et al., 2005). Young fish are distributed in lower salinity coastal areas, while adults have a more offshore distribution.

Adult red mullets feed on small crustaceans, annelid worms, and molluscs, using their chin barbels to detect prey and search the mud. As a consequence, striped red mullets are typically found on sandy, gravelly and shelly sediments where they can excavate sediment with their barbels and dislodge the small invertebrates. The main natural predators of striped red mullet are sea basses, pollacks, barracudas, monkfish, congers, and sharks (Caill-Milly et al., 2017).

Sexual maturity is reached at the beginning of the second year for males, followed by a marked decrease in growth rates, and at the end of the second or beginning of the third year for females which therefore continue their rapid growth a little longer (Déniel, 1991). In the English Channel, this species matures at approximately 16 cm (Mahe et al., 2005), while in the Bay of Biscay, the sizes of first sexual maturity are given by Dorel (1986) as males 16 cm , females 18 cm and a length at which $50 \%$ of the individuals are mature (the distinction between the two sexes is not mentioned) of 22 cm .

Spawning occurs in the spring and early summer -May to June according to Desbrosses (1933)with a spawning peak in June in the northern Bay of Biscay (N'Da and Déniel, 1993). Eggs and larvae average 2.8 mm and are pelagic (Sabatés et al., 2015). The hatching takes place after three days at $18^{\circ} \mathrm{C}$ and after eight days at a temperature of $9^{\circ} \mathrm{C}$ (Quéro and Vayne, 1997). After metamorphosis, juveniles become first demersal then benthic. At the age of one month, they measure about 5 cm and weigh 0.9 to 1.6 g . They show rapid growth during their first four months of life between July and October. Increases in length and mass are about 7 cm and 25 g on average during this period (N'Da and Déniel, 2005). The rate of growth declines sharply in October due to the cooling of water and the scarcity of trophic resources in the environment. These conditions contribute to the initiation of migration of red mullets to greater depths offshore. Until the age of two, there is no significant difference in size between males and females; they then measure $20-23 \mathrm{~cm}$. Sexual dimorphism is observed from the age of first maturity due to growth rates that will then differ between the two sexes. From age three, females exceed males in length by 4 cm on average and 7 cm beyond 5 years ( $\mathrm{N}^{\prime}$ Da et al., 2006).
The maximum reported age of the striped red mullet is 11 years (Quéro and Vayne, 1997; ICES, 2012), while the maximum length given is 44.5 cm in the Bay of Biscay (Dorel, 1986) and 40 cm elsewhere (Whitehead et al., 1984; Fischer et al., 1987). The maximum reported mass is 1 kg (Muus and Nielsen, 1999).

### 10.1 Management regulations

Prior to 2002, France enforced a minimum landing size of 16 cm . Since 2013 minimal size requirement has been established to 15 cm (France, 2013). There is no TAC for this stock.

### 10.2 Stock ID and possible management areas

In 2004 and 2005, a study using fish geometrical morphometry was carried out in the Eastern English Channel and the Bay of Biscay. It pointed out a morphological difference on striped red mullets between those from the Eastern English Channel and those from the Bay of Biscay (Mahe et al., 2014). Benzinou et al. (2013) conducted stock identification studies based on otolith and fish shape in European waters and showed that striped red mullet can be geographically divided into three zones:

- The Bay of Biscay (Northern Bay of Biscay - NBB, and Southern Bay of Biscay - SBB)
- A mixing zone composed of the Celtic Sea and the Western English Channel (CS + WEC)
- A northern zone composed of the Eastern English Channel and the North Sea (EEC + NS)

The distinction between the putative Biscay and Western Channel/Celtic Sea populations is supported by the distribution of landings at a statistical rectangle level (Figure 10.3.1). Examination of catch from surveys suggests striped red mullet in division 9 a are geographically distinct, with an area of higher abundance between Cabo Sao Vicente and the Tagus estuary, and an area where this species is mostly absent to the north. This assessment treats these putative components as one population. At present there are no management measures in place, however this structuring should be taken into account if measures are considered.

### 10.3 Fisheries data

Official landings have been recorded since 1975 and after early increases they have declined in recent years. Landings are mainly taken from Subarea 7 and 8 and France accounts for the majority of removals (Table 10.4.1-2). The striped red mullet is one species among set of benthic (demersal) species targeted by the French fleet, and is mainly caught by bottom trawlers with a mesh size of $70-99 \mathrm{~mm}$. In the Western English Channel striped red mullet is also caught by gillnets. Danish seine appeared in 2008 as a result of some trawlers converting to use seine gears.

The average characteristics of vessels in French fleets that caught red mullet from 2000 to 2015 are: 41.1 GRT, 191.1 kW engine power, 12.9 m length and 22 years of service. Net vessels are made up of the smallest units ( $85 \%$ are less than 12 m long), while $52 \%$ of bottom trawlers are less than 15 m ; the seiners are by far the largest and the oldest vessels (Caill-Milly et al., 2017).

The French activity on this species differs between the area composed by West Scotland/Celtic sea (including West Channel) and the area comprising the Bay of Biscay. In the first one, landings are mainly taken by bottom trawlers, followed by gillnet. In the second one, they are mainly done by bottom trawls, seine and nets. French activity in the Atlantic Iberian waters remains limited. The Spanish activity is located in the north ( $8 a, b$ ) and the south (8c) of the Bay of Biscay.

Discarding represented between $3 \%$ and $18 \%$ of the total catches in 2014-21 (Table 10.4.3). Since 2018, the discard rates are reported below $5 \%$. However, there are concerns about how these discards have been estimated due to the lack of discards data for some countries. From the data provided to Intercatch in 2020, discards are essentially composed of individuals measuring less than 18 cm (Figure 10.2).

### 10.4 Survey data, recruit series

Exchange data is available in DATRAS during 1997-2021 for the French EVHOE survey, covering the Bay of Biscay and Celtic Sea (fig. 10.5.1), during 2001-2016 for the northern Spanish
groundfish survey (SP-NSGFS), and from 2002 onwards for the Portuguese groundfish survey (PT-IBTS), covering the Portuguese coast. Relative total biomass in the EVHOE survey (fig. 10.5.2) are variable around the series mean between 1997 - 2011, before falling to a lower level thereafter. Similarly, catch rates in the PT-IBTS are at a low level in 2005, peak in 2010, before falling back to near the series mean in recent years. The mean stratified abundance from the Spanish NSGFS follows a similar trend: high variability around the mean before 2017, then low level since 2017. (fig. 10.5.3).

Biological sampling in the Bay of Biscay of sexual maturity and length measures were taken in 2009 by AZTI. French sampling started in 2004 in the Eastern Channel and in the south North Sea, and since 2008 in the Bay of Biscay. Since 2004, data (age, length, sexual maturity) are usually collected by France for the Eastern English Channel and the southern North Sea. France started to collect data for $8 \mathrm{a}, \mathrm{b}$ at the end of 2007. In 2007-2008, the striped red mullet otolith exchange had for goal to optimize age estimation between countries. In 2011, an Otolith Exchange Scheme was carried out, which was the second exercise for the Striped red mullet (Mullus surmuletus). Four readers of this exchange interpreted an images collection coming from the Bay of Biscay, the Spanish coasts and the Mediterranean coasts (Spain and Italy). A set of Mullus surmuletus otoliths ( $\mathrm{N}=75$ ) from the Bay of Biscay presented highest percentage of agreement ( $82 \%$ ). On 75 otoliths, 34 were read with $100 \%$ agreement ( $45 \%$ ) and thus a CV of $0 \%$. Modal age of these fishes was comprised between 0 and 3 years (Mahe et al., 2012).

## 10.5 $\quad 10.6 \quad$ Current research programs

Two research projects are currently investigating
(1) the evolution of striped red mullet abundance indices from fishery dependent data and
(2) the temporal evolution of the size and age at maturity for this species in the Bay of Biscay.

The first research project (ACOST) extend the analysis presented in
Caill-Milly et al. (2017) and Caill-Milly et al. (2019) and computes 4 abundances indices from 2005 to 2021 based on the landings per unit effort for 4 French fleets. The second project (MATO) updates the maturity data for the species in the Bay of Biscay thanks to a monthly longitudinal study over one reproduction cycle done in 2021. The final results will be published in 2022/2023 and the references will be added in the next report.

### 10.6 Analysis of stock trends/ assessment

Currently, an age structured analytical stock assessment has not been developed due to a short time-series of available data.

Data requirements - regular sampling of biological parameters of striped red mullet catches must be continued under DCF. Sampling in the Celtic Sea and in the Bay of Biscay started in 2008. In 2010 and 2011, sampling for age and maturity data was reduced compared to 2009, due to the end of the Nespman project. Since 2009, a concurrent sampling design carried out, should provide more data (length compositions) than in recent years.

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Table 10.4.1: Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8 and 9a. Official landings by country in tonnes.

| $\begin{aligned} & \text { § } \\ & \text { N/ } \end{aligned}$ |  | $\begin{aligned} & \text { 曾 } \\ & \text { OM. } \\ & \frac{1}{3} \end{aligned}$ |  |  | $\begin{aligned} & \text { Q } \\ & \text { D } \\ & \text { 亏ָu } \\ & \text { N } \end{aligned}$ |  | $\begin{aligned} & \overline{\mathrm{D}} \\ & \frac{10}{2} \\ & \text { ® } \end{aligned}$ |  |  |  |  |  | $\begin{aligned} & \frac{n}{0} \\ & 0.0 \end{aligned}$ | 듯 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 33 |  | 1947 | 8 |  | 16 |  | 1 |  | 115 |  | 10 | 387 | 170 | 2688 |  |
| 2007 | 43 |  | 1941 | 9 |  | 23 |  | 1 |  | 148 |  | 222 | 398 | 194 | 2978 |  |
| 2008 | 26 |  | 1394 | 9 |  | 22 |  | 0 |  | 165 |  | 169 | 394 | 165 | 2345 |  |
| 2009 | 20 |  | 1562 | 5 |  | 16 |  | 0 |  | 110 |  | 199 | 520 | 134 | 2567 |  |
| 2010 | 20 |  | 1743 | 5 |  | 8 |  | 0 |  | 128 |  | 276 | 479 | 133 | 2793 |  |
| 2011 | 21 |  | 1740 | 0 |  | 8 |  | 0 |  | 130 |  | 245 | 508 | 155 | 2806 |  |
| 2012 | 37 |  | 1342 | 0 |  | 7 |  | 1 |  | 125 |  | 217 | 332 | 122 | 2183 |  |
| 2013 | 28 |  | 932 | 5 |  | 4 |  | 0 |  | 50 |  | 187 | 246 | 71 | 1522 |  |
| 2014 | 12 |  | 926 | 5 |  | 2 |  | 0 |  | 2 |  | 221 | 265 | 53 | 1487 |  |
| 2015 | 23 |  | 1215 | 5 |  | 3 |  | 0 |  | 111 |  | 282 | 248 | 102 | 1989 |  |
| 2016 | 28 |  | 1179 | 0 |  | 4 |  | 0 |  | 69 |  | 204 | 194 | 83 | 1761 |  |
| 2017 | 36 |  | 997 | 0 |  | 10 |  | 0 |  | 13 |  | 154 | 327 | 64 | 1601 |  |
| 2018 | 37 |  | 896 | 0 |  | 0 |  | 0 |  | 95 |  | 122 | 321 | 67 | 1538 |  |
| 2019 | 30 |  | 1358 | 0 |  | 12 |  | 0 |  | 91 |  | 159 | 267 | 55 | 1973 |  |
| 2020 | 50 |  | 965 | 0 |  | 6 |  | 0 |  | 82 |  | 109 | 261 | 89 | 1562 |  |
| 2021 | 53 |  | 836 | 0 |  | 18 |  | 0 |  | 54 |  | 117 | 274 | 93 | 1445 |  |

Table 10.4.2: Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8 and 9a. Official landings by area in tonnes.

| Year | $\mathbf{6 a}$ | $\mathbf{6 b}$ | $\mathbf{7 a}$ | $\mathbf{7 b}$ | $\mathbf{7 c}$ | $\mathbf{7 e}$ | $\mathbf{7 f}$ | $\mathbf{7 g}$ | $\mathbf{7 h}$ | $\mathbf{7 j}$ | $\mathbf{7 k}$ | $\mathbf{8 a}$ | $\mathbf{8 b}$ | $\mathbf{8 c}$ | $\mathbf{8 d}$ | $\mathbf{8 e}$ | $\mathbf{9 a}$ | Total |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2006 | 0 | 0 | 1 | 1 | 0 | 869 | 50 | 24 | 103 | 11 | 0 | 1,023 | 468 | 71 | 28 | 0 | 39 | 2688 |
| 2007 | 1 | 0 | 1 | 1 | 1 | 1047 | 54 | 22 | 104 | 24 | 0 | 861 | 473 | 90 | 32 | 0 | 267 | 2978 |
| 2008 | 0 | 0 | 1 | 1 | 0 | 880 | 46 | 16 | 72 | 26 | 0 | 639 | 246 | 86 | 35 | 0 | 296 | 2345 |
| 2009 | 2 | 0 | 1 | 2 | 2 | 592 | 25 | 9 | 74 | 35 | 0 | 879 | 460 | 156 | 88 | 0 | 243 | 2567 |
| 2010 | 2 | 0 | 1 | 3 | 2 | 642 | 26 | 10 | 59 | 32 | 1 | 1,033 | 467 | 146 | 38 | 0 | 331 | 2793 |
| 2011 | 1 | 1 | 1 | 0 | 0 | 665 | 20 | 10 | 55 | 11 | 0 | 970 | 513 | 214 | 35 | 0 | 310 | 2806 |
| 2012 | 0 | 0 | 0 | 0 | 0 | 493 | 23 | 7 | 34 | 9 | 0 | 696 | 387 | 200 | 53 | 0 | 280 | 2183 |


| Year | $\mathbf{6 a}$ | $\mathbf{6 b}$ | $\mathbf{7 a}$ | $\mathbf{7 b}$ | $\mathbf{7 c}$ | $\mathbf{7 e}$ | $\mathbf{7 f}$ | $\mathbf{7 g}$ | $\mathbf{7 h}$ | $\mathbf{7 j}$ | $\mathbf{7 k}$ | $\mathbf{8 a}$ | $\mathbf{8 b}$ | $\mathbf{8 c}$ | $\mathbf{8 d}$ | $\mathbf{8 e}$ | $\mathbf{9 a}$ | Total |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2013 | 0 | 0 | 0 | 1 | 0 | 232 | 23 | 7 | 36 | 4 | 0 | 473 | 328 | 166 | 12 | 0 | 241 | 1522 |
| 2014 | 1 | 0 | 0 | 0 | 0 | 192 | 15 | 3 | 40 | 3 | 0 | 523 | 240 | 151 | 23 | 0 | 297 | 1487 |
| 2015 | 0 | 0 | 0 | 1 | 0 | 595 | 10 | 2 | 36 | 2 | 0 | 506 | 327 | 126 | 15 | 0 | 369 | 1989 |
| 2016 | 0 | 0 | 0 | 2 | 0 | 417 | 21 | 7 | 35 | 5 | 0 | 548 | 311 | 117 | 21 | 0 | 277 | 1761 |
| 2017 | 0 | 0 | 0 | 1 | 0 | 277 | 27 | 21 | 37 | 3 | 0 | 514 | 324 | 160 | 5 | 0 | 231 | 1601 |
| 2018 | 0 | 0 | 0 | 0 | 0 | 361 | 26 | 7 | 39 | 1 | 0 | 453 | 276 | 144 | 2 | 0 | 226 | 1538 |
| 2019 | 0 | 0 | 1 | 1 | 0 | 377 | 23 | 20 | 35 | 1 | 0 | 770 | 388 | 123 | 4 | 0 | 229 | 1973 |
| 2020 | 0 | 0 | 2 | 1 | 0 | 386 | 43 | 18 | 40 | 4 | 0 | 502 | 265 | 128 | 3 | 0 | 170 | 1562 |
| 2021 | 0 | 0 | 1 | 0 | 0 | 302 | 52 | 30 | 54 | 3 | 0 | 416 | 281 | 114 | 2 | 0 | 188 | 1445 |

Table 10.4.3: Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8 and 9a. Official discards by country in tonnes. Total is presented with the total discards rates in \%



Figure 10.3.1: Striped red mullet in Subareas and Divisions 6, 7a-c, e-f, 8 and 9a. Landings by statistical rectangle for BEL, FRA, IRE, PT, UK (E\&W), UK (SCO) from 2014 to 2020 (Fishery Dependent Information database 2021).


Figure 10.2: Striped red mullet in Subareas and Divisions 6, 7a-c, e-f, 8 and 9a. Length distribution from 2014 to 2021 from Intercatch (D: Discards, L: Landings)


Figure 10.5.1: EVHOE survey station map


Figure 10.5.2: Total biomass of striped red mullet in Subareas and Divisions 6, 7a-c, e-f, 8 and 9a., estimated from the EVHOE survey in tons, 1997-2021


Figure 10.5.3: Striped red mullet in Subareas and Divisions 6, 7a-c, e-f, 8 and 9a. Spain NSGFS mean stratified abundance in northern Spanish Shelf 1983-2020

## Annex 1: List of participants

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## Annex 2: Terms of Reference

## WGWIDE- Working Group on Widely Distributed Stocks

## Approved in November 2021

2021/2/FRSG20 The Working Group on Widely Distributed Stocks (WGWIDE), chaired by Andrew Campbell, Ireland, will meet 24-30 August 2022 in ICES HQ in Copenhagen to:
a) Address generic ToRs for Regional and Species Working Groups.

The assessments will be carried out on the basis of the stock annex. The assessments must be available for audit on the first day of the meeting.

Material and data relevant for the meeting must be available to the group no later than 14 days prior to the starting date.

WGWIDE will report by 2 September 2022 for the attention of ACOM.

Only experts appointed by national Delegates or appointed in consultation with the national Delegates of the expert's country can attend this Expert Group

## Generic ToRs for Regional and Species Working Groups

## Approved in November 2021

2021/2/FRSG01 The following ToRs apply to: AFWG, HAWG, NWWG, NIPAG, WGWIDE, WGBAST, WGBFAS, WGNSSK, WGCSE, WGDEEP, WGBIE, WGEEL, WGEF, WGHANSA and WGNAS.

The working group should focus on:
a) Consider and comment on Ecosystem and Fisheries overviews where available;
b) For the aim of providing input for the Fisheries Overviews, consider and comment on the following for the fisheries relevant to the working group:
i) descriptions of ecosystem impacts on fisheries
ii) descriptions of developments and recent changes to the fisheries
iii) mixed fisheries considerations, and
iv) emerging issues of relevance for management of the fisheries;
c) Conduct an assessment on the stock(s) to be addressed in 2022 using the method (assessment, forecast or trends indicators) as described in the stock annex; - complete and document an audit of the calculations and results; and produce a brief report of the work carried out regarding the stock, providing summaries of the following where relevant:
i) Input data and examination of data quality; in the event of missing or inconsistent survey or catch information refer to the ACOM document for dealing with COVID-

19 pandemic disruption and the linked template that formulates how deviations from the stock annex are to be reported.
ii) Where misreporting of catches is significant, provide qualitative and where possible quantitative information and describe the methods used to obtain the information;
iii) For relevant stocks (i.e., all stocks with catches in the NEAFC Regulatory Area), estimate the percentage of the total catch that has been taken in the NEAFC Regulatory Area in 2021.
iv) For category 3 and 4 stocks requiring new advice in 2022, implement the methods recommended by WKLIFE X (e.g. SPiCT, rfb, chr, rb rules) to replace the former 2 over 3 advice rule ( 2 over 5 for elasmobranchs). MSY reference points or proxies for the category 3 and 4 stocks
v) Evaluate spawning stock biomass, total stock biomass, fishing mortality, catches (projected landings and discards) using the method described in the stock annex;

1) for category 1 and 2 stocks, in addition to the other relevant model diagnostics, the recommendations and decision tree formulated by WKFORBIAS (see Annex 2 of https://www.ices.dk/sites/pub/Publication\ Reports/Ex-pert\ Group\ Report/Fisheries\ Resources\ Steering\ Group/2020/WKFORBIAS 2019.pdf) should be considered as guidance to determine whether an assessment remains sufficiently robust for providing advice.
2) If the assessment is deemed no longer suitable as basis for advice, consider whether it is possible and feasible to resolve the issue through an interbenchmark. If this is not possible, consider providing advice using an appropriate Category 2 to 5 approach.;
vi) The state of the stocks against relevant reference points;

Consistent with ACOM's 2020 decision, the basis for Fpa should be Fp. 05 .

1) 2. Where Fp. 05 for the current set of reference points is reported in the relevant benchmark report, replace the value and basis of Fpa with the information relevant for Fp. 05
1) 2. Where Fp. 05 for the current set of reference points is not reported in the relevant benchmark report, compute the Fp. 05 that is consistent with the current set of reference points and use as Fpa. A review/audit of the computations will be organized.
1) 3. Where Fp. 05 for the current set of reference points is not reported and cannot be computed, retain the existing basis for Fpa.
vii) Catch scenarios for the year(s) beyond the terminal year of the data for the stocks for which ICES has been requested to provide advice on fishing opportunities;
viii)Historical and analytical performance of the assessment and catch options with a succinct description of associated quality issues. For the analytical performance of category 1 and 2 age-structured assessments, report the mean Mohn's rho (assessment retrospective bias analysis) values for time series of recruitment, spawning stock biomass, and fishing mortality rate. The WG report should include a plot of this retrospective analysis. The values should be calculated in accordance with the "Guidance for completing ToR viii) of the Generic ToRs for Regional and Species

Working Groups - Retrospective bias in assessment" and reported using the ICES application for this purpose.
d) Produce a first draft of the advice on the stocks under considerations according to ACOM guidelines.
i. In the section 'Basis for the assessment' under input data match the survey names with the relevant "SurveyCode" listed ICES survey naming convention (restricted access) and add the "SurveyCode" to the advice sheet.
e) Review progress on benchmark issues and processes of relevance to the Expert Group.
i) update the benchmark issues lists for the individual stocks in SID;
ii) review progress on benchmark issues and identify potential benchmarks to be initiated in 2023 for conclusion in 2024;
iii) determine the prioritization score for benchmarks proposed for 2023-2024;
iv) as necessary, document generic issues to be addressed by the Benchmark Oversight Group (BOG)
f) Prepare the data calls for the next year's update assessment and for planned data evaluation workshops;
g) Identify research needs of relevance to the work of the Expert Group.
h) Review and update information regarding operational issues and research priorities on the Fisheries Resources Steering Group SharePoint site.
i) If not completed in 2020, complete the audit spread sheet 'Monitor and alert for changes in ecosystem/fisheries productivity' for the new assessments and data used for the stocks. Also note in the benchmark report how productivity, species interactions, habitat and distributional changes, including those related to climate-change, could be considered in the advice.

Information of the stocks to be considered by each Expert Group is available here.

## Annex 4: List of Stock Annexes

The table below provides an overview of the WGWIDE Stock Annexes. Stock Annexes for other stocks are available on the ICES website Library under the Publication Type "Stock Annexes". Use the search facility to find a particular Stock Annex, refining your search in the left-hand column to include the year, ecoregion, species, and acronym of the relevant ICES expert group.

| STOCK ID | STOCK NAME | LAST UPDATED | LINK |
| :---: | :---: | :---: | :---: |
| boc.27.6-8 | Boarfish (Capros aper) in Sub areas 6-8 (Celtic Seas, English Channel, and Bay of Biscay) | $\begin{aligned} & \text { September } \\ & 2020 \end{aligned}$ | boc.27.6-8 SA |
| gur.27.3-8 | Red gurnard (Chelidonichthys cuculus) in subareas 3-8 (Northeast Atlantic) | $\begin{aligned} & \text { September } \\ & 2021 \end{aligned}$ | gur.27.3-8 |
| her.27.1-24a514a | Herring (Clupea harengus) in subareas 1, 2, and 5, and in divisions 4.a and 14.a, Norwegian spring-spawning herring (the Northeast Atlantic and Arctic Ocean) | $\begin{aligned} & \text { September } \\ & 2021 \end{aligned}$ | her.27.1-24a514a_SA |
| hom.27.3a4bc7d | Horse mackerel (Trachurus trachurus) in divisions 3.a, 4.b-c, and 7.d (Skagerrak and Kattegat, southern and central North Sea, eastern English Channel) | $\begin{aligned} & \text { September } \\ & 2021 \end{aligned}$ | hom.27.3a4bc7d SA |
| hom.27.2a4a5b6a7a -ce-k8 | Horse mackerel (Trachurus trachurus) in Subarea 8 and divisions 2.a, 4.a, 5.b, 6.a, 7.a-c,e-k (the Northeast Atlantic) | $\begin{aligned} & \text { September } \\ & 2021 \end{aligned}$ | $\begin{aligned} & \text { hom.27.2a4a5b6a7a- } \\ & \text { ce-k8 SA } \end{aligned}$ |
| mac.27.nea | Mackerel (Scomber scombrus) in subareas 1-7 and 14 and divisions 8.a-e, 9.a (the Northeast Atlantic and adjacent waters) | September $2021$ | mac.27.nea SA |
| whb.27.1-91214 | Blue whiting (Micromesistius poutassou) in subareas 1-9, 12, and 14 (Northeast Atlantic and adjacent waters) | $\begin{aligned} & \text { September } \\ & 2021 \end{aligned}$ | whb.27.1-91214 SA |

## Annex 4: Audits

## 1 Audit of Norwegian spring spawning herring (her.27.124a514a)

Date: 05.09.2022
Auditor: Afra Egan, Anna Olafsdottir, Axelle

## General

The Norwegian springs-pawning herring is carried out using the XSAM model. This audit focuses on input data for the assessment and the WGWIDE report chapter.

For single stock summary sheet advice:

1) Assessment type: update/SALY
2) Assessment: analytical
3) Forecast: presented
4) Assessment model: XSAM with 3 survey fleets
5) Data issues: 2022 assessment input data are available on SharePoint in the folder "07.Software - 2022_her.27.1-24a514a_assessment".
Input data files were checked against the working group report tables
Data were the same in tables except for 2 instances:

Table 4.4.3.1 Catch numbers at age for 2020 differ from the input file - correction done. Table 4.4.4.1 Mean weights in the catch at age 1 does not match the input file (not used in the assessment)
The only available catch data from Russian Federation for 2021 was total catch by ICES division from ICES preliminary catch database, and no Russian catch samples were available. Historically, preliminary catches are comparable to ICES final estimated catch. There were adequate samples from other fishing nations operating in the same areas which were used to estimate catch at age and weight at age.
6) Consistency: This years' assessment is consistent with last years' assessment and the WG accepted the assessment.
7) Stock status: The fishing pressure on the stock is above $\mathrm{F}_{\mathrm{MSY}}$, and Fpa (but below $\mathrm{F}_{\mathrm{lim}}$ ). Spawning-stock size is above MSY $B_{\text {trigger, }} \mathrm{B}_{\mathrm{pa}}$, and $\mathrm{Blim}_{\text {lim }}$.
8) Management Plan: Agreed by the Coastal States in October 2018: the TAC shall be fixed to a fishing mortality of Fmgt $=0.14$, with a constraint of maximum $20 \%$ reduction and $25 \%$ increase relative to the TAC in the preceding year. If SSB is forecast to be lower than MSY $\mathrm{B}_{\text {trigger }}$ in the beginning of the quota year, F decreases linearly from $\mathrm{F}_{\text {mgt }}$ to $\mathrm{F}=0.05$ over the biomass range from $\mathrm{B}_{\text {trigger }}$ to $\mathrm{Blim}_{\text {lim }}$. The long-term management strategy has been evaluated by ICES and found to be consistent with the precautionary approach.

## General comments

The input data and assessment are documented as described in the stock annex and the report sections are well ordered. A table summarising the assessment settings in the stock annex would be useful and would make the audit easier.

The advice sheet was clearly and concisely written. Numbers and tables in the advice sheet were compared to the same information in the report and rounding differences highlighted and comments forwarded to the responsible person.

## Technical comments

To the best of our knowledge, the assessment has been performed correctly according to the stock annex.

Table and figure numbers and references to them in the text have been checked.

## Conclusions

The assessment has been performed correctly

## Checklist for audit process

## General aspects

- Has the EG answered those TORs relevant to providing advice? yes
- Is the assessment according to the stock annex description? I think so?
- If a management plan is used as the basis of the advice, has been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary? yes
- Have the data been used as specified in the stock annex? yes
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex?
- Is there any major reason to deviate from the standard procedure for this stock? no
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice? yes

Audit of Western Horse Mackerel data and assessment<br>Date: 07/09/2022<br>Auditor: Alessandro Orio, Sólvá Káradóttir Eliasen, Eleanor MacLeod, Richard Nash

## General

Western horse mackerel is assessed as a Category 1 stock. An SS3 model is run to determine the state of the stock in relation to reference points for western horse mackerel.

## For single stock summary sheet advice:

9) Assessment type: update
10) Assessment: analytical.
11) Forecast: presented
12) Assessment model: SS3 model with commercial catches (length and age data) and three survey indices: Triennial egg survey index (1992-2019); IBTS recruitment index; PELACUS acoustic biomass.
13) Data issues: No data issues.
14) Consistency: The view of the WG was that the assessment should be accepted. The Stock annex needs to be updated for the F and M before spawning used in the forecast (assumed at the beginning of the year in the current forecast) and for the new $F_{p a}$ value due the changed basis.
15) Stock status: Fishing pressure on the stock is above $\mathrm{F}_{\mathrm{MSy}}$ and between $\mathrm{F}_{\mathrm{pa}}$ and Flim; spawning-stock size is below MSY $\mathrm{B}_{\text {trigger, }} \mathrm{B}_{\mathrm{pa}}$ and $\mathrm{B}_{\text {lim }}$.
16) Management Plan: No management plan

## General comments

The assessment and forecast have been available for review. Input and output data were correct. A few inconsistencies were found in the advice sheet but these have been already corrected.

## Technical comments

Few inconsistencies are present in the stock annex. F and M before spawning in the forecast needs to be updated in the stock annex since in the forecast the spawning time is assumed to happen at the beginning of the year. The section on reference points needs to be updated with the new $F_{p a}$ due to the change of basis.
A thorough revision of the number of samples used for the different age and length frequency distributions in the assessment is suggested for the next benchmark iteration. There is a need to inspect the potential problems caused by the reweighting of both age length keys and age frequency distribution of the commercial catches using the same parameter. The fishing mortality estimated by the model is weighted by the population numbers but now the unweighted F can be obtained so it would be preferable to switch to that in the future to avoid extra calculations. Forecasts run directly in SS should be also considered during the next benchmark.
There are four tables in the tables section to which there, in the text section, are no references (Tables 7.2.4.3-7.2.4.6).

## Conclusions

The assessment has been performed correctly.

## Checklist for audit process

## General aspects

- Has the EG answered those TORs relevant to providing advice?


## Yes

- Is the assessment according to the stock annex description? Yes but it needs to be updated
- If a management plan is used as the basis of the advice, has been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary? Yes, no management plan
- Have the data been used as specified in the stock annex? Yes
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex?
Yes
- Is there any major reason to deviate from the standard procedure for this stock?

No

- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice?
Yes.


# 2 Audit of the Blue whiting assessment (whb.27.1-91214) 

Date: 01.09.2022
Auditor: Martin Pastoors, Leif Nottestad, Ed Farrell, Jessica Tingvall

## General

The blue whiting assessment is carried out using the SAM model and available on Stockassessment.org (WHB-2022). This audit focuses on input data, assessment, forecast and draft advice document.

## For single stock summary sheet advice:

17) Assessment type: update/SALY
18) Assessment: analytical
19) Forecast: presented; derived directly from the outputs of the SAM model. Appropriate settings according to Stock Annex.
20) Assessment model: SAM with 1 survey fleet
21) Data issues: Estimation of preliminary catch in 2022 difficult because of absence of Russian information. The tables in the report have been checked in relation to the input files used for the assessment and in relation with the tables in the advice summary.
22) Consistency: This years' assessment is consistent with last years' assessment, although there is very different outlook due to an incoming new year class. The WG accepted the assessment.
23) Stock status: The fishing pressure on the stock is above FMSY, FMGT and Fpa (but below Flim). Spawning-stock size is above MSY Btrigger, Bpa, and Blim.
24) Management Plan: Agreed by the Coastal States in October 2016 after evaluation of the management plan by ICES. The long-term management strategy was found to be consistent with the precautionary approach. However, the management plan was modified subsequent to the evaluation by ICES by including a clause to lift the limit on TAC change if the change was more than $40 \%$ (Clause 6). This modification has not been evaluated by ICES. Despite the agreement on the management plan by the Coastal States, the plan has not been effective due to a lack of agreement on the sharing of the TAC.

## General comments

The input data and assessment are documented as described in the stock annex and the report sections are well ordered.

## Technical comments

The code for the short term forecast is embedded in a large collection of code (_job_to_do_it_all.R) that is not running on stockassessment.org. If the stock is not being entered into TAF, it could be beneficial to at least include the forecast methodology directly on stockassessment.org. The text on the forecast in the stock annex needs updating as it is referring to code being available on stockassessment.org which is currently not the case.

## Conclusions

The assessment has been performed correctly

## Checklist for audit process

## General aspects

- Has the EG answered those TORs relevant to providing advice? Yes.
- Is the assessment according to the stock annex description? Yes (in some cases the SA will need minor updates)
- If a management plan is used as the basis of the advice, has been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary? The management plan has been agreed by the Coastal States in October 2016 after evaluation of the management plan by ICES. The long-term management strategy has been evaluated by ICES and was consistent with the precautionary approach. However, the management plan was modified subsequent to the evaluation by ICES by including a clause to lift the limit on TAC change if the change was more than $40 \%$ (Clause 6). This modification has not been evaluated by ICES. Despite the agreement on the management plan by the Coastal States, the plan has not been effective due to a lack of agreement on the sharing of the TAC.
- Have the data been used as specified in the stock annex? Yes.
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex? Yes.
- Is there any major reason to deviate from the standard procedure for this stock? No
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice? Yes


## Checking of Blue whiting report tables.

## Green $=$ checked and ok. NO highlight $=$ can't find the source data. Yellow = potential issue

1. Table 2.3.1.1. Blue whiting. ICES estimated catches (tonnes) by country for the period 1988-2021.
2. Table 2.3.1.2. Blue whiting. ICES estimated catches (tonnes) by country and ICES division for 2021.
3. Table 2.3.1.3. Blue whiting. ICES estimated catches (tonnes) by quarter and ICES division for 2021.
4. Table 2.3.1.4. Blue whiting. ICES estimated catches (tonnes) from the main fisheries 1988-2021 by area.
5. Table 2.3.1.5. Blue whiting. ICES estimates (tonnes) of catches, landings and discards by country for 2021.
6. Table 2.3.1.6. Blue whiting. ICES estimated catches (tonnes) inside and outside NEAFC regulatory area for 2021 by country.
7. Table 2.3.1.1.1. Blue whiting. ICES estimated catches (tonnes), the percentage of catch covered by the sampling programme, No. of age samples, No. of fish measured and No. of fish aged for 2000-2021.
8. Table 2.3.1.1.2. Blue whiting. ICES estimated catches (tonnes), the percentage of catch covered by the sampling programme (catch-at-age numbers), No. of length samples, No. of age samples, No. of fish measured,

## No. of fish aged, No. of fish aged by 1000 tonnes and No. of fish measured by 1000 tonnes by country for 2021.

9. Table 2.3.1.1.3. Blue whiting. ICES estimated catches (tonnes), No. of Age samples, No. of fish measured and No. of fish aged by country and quarter for 2021.
10. Table 2.3.1.1.3. (continued) Blue whiting. ICES estimated catches (tonnes), No. of Age samples, No. of fish measured and No. of fish aged by country and quarter for 2021.
11. Table 2.3.1.1.3. (continued) Blue whiting. ICES estimated catches (tonnes), No. of Age samples, No. of fish measured and No. of fish aged by country and quarter for 2021.
12. Table 2.3.1.1.4. Blue whiting. ICES estimated catches (tonnes), the percentage of catch covered by the sampling programme, No. of length samples, No. of age samples, No. of fish measured, No. of fish aged, No. of fish aged by 1000 tonnes and No. of fish measured by 1000 tonnes by ICES division for 2021.
13. Table 2.3.2.1. Blue whiting. ICES estimated preliminary landings (tonnes) in 2022 by quarter and ICES division. Data submitted to InterCatch.
14. Table 2.3.2.2. Blue whiting. ICES estimated preliminary catches (tonnes), the percentage of catch covered by the sampling programme, No. of samples, No. of fish measured, No. of fish aged, No. of fish aged by 1000 tonnes and No. of fish measured by 1000 tonnes by ICES division for 2022 preliminary data (quarters 1 and 2). Data submitted to InterCatch.
15. Table 2.3.2.3. Blue whiting. ICES estimates of catches (tonnes) in 2022, based on (initial) declared quotas and expected uptake estimated by WGWIDE.
16. Table 2.3.2.4. Blue whiting. Comparison of preliminary and final catches (tonnes).
17. Table 2.3.3.1. Blue whiting. Catch-at-age numbers (thousands) by year. Discards included since 2014. Values for 2022 are preliminary.
a. From 2011 onwards the data file (cn.dat) numbers are given to a few decimal places which is not show in the report tables.
b. There is a difference in how some of these decimals are rounded when at 0.5 in data. In some cases rounded up and in some rounded down.
i. 2016 age 3 in data is 2933271.5 and 2933271 in table
ii. 2018 age 9 in data is 90387.5 and 90387 in table
iii. 2021 age 7 in data is 1360104.5 and 1360104 in table
18. Table 2.3.4.1. Blue whiting. Individual mean weight ( kg ) at age in the catch. Preliminary values for 2022.
a. cw.dat. Checked. OK.
19. Table 2.3.5.1. Blue whiting. Natural mortality and proportion mature.
a. From mo.dat. OK
20. Table 2.3.7.1.1. Blue whiting. Time-series of StoX abundance estimates of blue whiting (millions) by age in the IBWSS. Total biomass in last column ( 1000 t ). Shaded values (ages 1-8; years 2004-2022) are used as input to the assessment
a. survey.dat. OK
21. Table 2.3.7.2.1. Blue whiting. Estimated abundance of 1 and 2 year old blue whiting from the International Ecosystem Survey in Nordic Seas (IESNS), 2003-2022.

Compare report table to table in BW_RecruitmentRank22.xls. There is one disagreement in the table in 2014. In the report age 1 and 2 in 2014 are 3893 and 2048 but in the xl table they are 3937 and 2030, respectively.
22. Table 2.3.7.2.2. Blue whiting. 1-group indices of blue whiting from the Norwegian winter survey (late January-early March) in the Barents Sea. (Blue whiting < 19 cm in total body length which most likely belong to 1-group.)
23. Table 2.3.7.2.3. Blue whiting. 1-group indices of blue whiting from the Icelandic bottom-trawl surveys, 1-group (< 22 cm in March).
24. Table 2.3.7.2.4. Blue whiting. 1-group indices of blue whiting from Faroese bottom-trawl surveys, 1-group (<= 23 cm in March).
25. Table 2.4.1.1. Blue whiting. Parameter estimates, from final assessment (2022) and retrospective analysis (2018-2021).
26. Table 2.4.1.2. Blue whiting. Mohn's rho by year and average over the last five years ( $\mathbf{n}=5$ ).
27. Table 2.4.1.3. Blue whiting. Estimated fishing mortalities. Catch data for 2022 are preliminary.
a. The 2023 data is in the model output but not in the report table. Is this correct??
28. Table 2.4.1.4. Blue whiting. Estimated stock numbers-at-age (thousands). Preliminary catch data for 2022 have been used.
a. The 2023 age 1 figure is missing from the report table but is in the model output table. Is this correct?
29. Table 2.4.1.5. Blue whiting. Estimated recruitment ( R ) in thousands, spawning-stock biomass (SSB) in tonnes, average fishing mortality for ages 3 to 7 (Fbar 3-7) and total-stock biomass (TBS) in tonnes. Preliminary catch data for 2022 are included.
a. Some of the 2023 values that are in the model output are not in the report table. Is this correct?
30. Table 2.4.6. Blue whiting. Model estimate of total catch weight (in tonnes) and Sum of Product of catch number and mean weight at age for ages 1-10+ (Observed catch). Preliminary catch data for 2022 are included.
31. Table 2.8.2.1.1. Blue whiting. Input to short-term projection (median values for exploitation pattern and stock numbers).
32. Table 2.8.2.1.2. Blue whiting. Deterministic forecast, intermediate year assumptions and recruitments.
33. Table 2.8.2.2.1. Blue whiting. Deterministic forecast (weights in tonnes).

Some checking of the assessment inputs and settings:

1. "\# preliminary year catches, the best guesses on total catch in the current (full) year (the catch of O-groups should be subtracted, but not done)" - JT: Has this been dealt with or not?
MV: The 0-group catch at age is very small in the preliminary Q1 and Q2 catches (they are mainly caught in the second half-year), however our best guess on the total catch weight is transformed to catch at age without taking account of the 0 -group. This will provide (an insignificant) bias, but we ignore that, for the preliminary data.
2. \#totalyield<- 1233169 \#\# best guess for 2021 - JT: Cannot find this number anywhere?

EB: this is the value the preliminary catches for 2021 used in last year's assessment, so I think this is not relevant anymore - we could have even deleted this line as we're not using it in this year's assessment.
totalyield<- 1107529 \#\# best guess for 2022 - ok!
3. Fpa<-0.32 in stock annex it says Fpa= 0.53. - JT: Not sure if this is a typo in the code or if 0.32 is correct, but in the stock annex it is Fpa $=0.53$ (refer to table in stock annex on page 23). The other BRFs are ok.
EB: thanks a lot for spotting, this is an old value! We'll change it to 0.32 in the stock annex.
4. JT: Configuration looks ok. I'm guessing that 1's in stock annex is equivalent to 0's in the script? See example below:
\# Coupling of fishing mortality STATES
\# Rows represent fleets.
\# Columns represent ages.
\# 12345678910 \# Age
1234567899 \# Catch - stock annex
\$keyLogFsta - in script
0123456788
MV: The configuration file was made for the ADMB version of SAM, but now where we use the TMB version it is fine to change the configuration to that (and maybe add a sentence that the configuration file is for use with the TMB version of SAM).

EB: I'll make this change in the stock annex.

# Audit of (Northeast Atlantic mackerel (mac.27.nea)) 

Date: $\quad 8^{\text {th }}$ September, 2022

Auditor: Eydna í Homrum, Sondre Hølleland, Esther Beukhof

- Audience to write for: ADG, ACOM, benchmark groups and EG next year.
- Aim is to audit (check if correct):
० the stock assessment- concentrate on the input data, settings and output data
from the assessment
o the correct use of the assessment output in the forecast, and check if forecast
settings are applied correctly
- Any deviations from the stock annex should be described sufficiently.
- By the conclusion of the working group, all update assessments should be audited suc-
cessfully.
- Store all audits on SharePoint for future reference.


## General

This audit focuses on the advice sheet and the WGWIDE report section on NEA Mackerel. The advice sheet is generally consistent with the report section. Some small inconsistencies in catch tables were identified between the advice sheet and the report. The assessment model performance was good, and a systematic downward revision in the retrospective pattern for F in recent years seems to be improved.

## For single stock summary sheet advice:

25) Assessment type: updated assessment (inter-benchmarked in 2019)
26) Assessment: analytical
27) Forecast: presented
28) Assessment model: A modified state-space Assessment Model (SAM) that is able to incorporate tag/recapture data - both historical steel tags (1980-2006) and recent RFID tags (2014-2021) together with three additional survey indices.
29) Data issues: For the IBTS age 0 index, no value for 2021 could be calculated due to technical issues with one survey vessel covering an historically important area. Therefore, the stock assessors had to deviate from the methodology in the stock annex for estimating recruitment for 2021 in the short-term forecast. The time-tapered geometric mean was estimated without the weighting procedure that uses the IBTS index and the SAM recruitment estimates combinedly. Instead, the time-tapered geometric mean was estimated using the SAM estimates only.
There was no submission of Russian data to WGWIDE this year, yet both preliminary catches for 2021 and final catches from 2018-2020 indicated a Russian proportion of the catches of $13 \%$. It is therefore considered appropriate to use the historic average (20182020) to assign catches to Russia in 2021 and to use samples from Iceland and the Faroes to allocate the Russian catches, as these countries fish largely within the same area and time of year.
30) Consistency: The retrospective bias (8 years considered), where the F has consistently been overestimated and SSB underestimated, is still present in older years but has become less apparent in recent years.
31) Stock status: $\operatorname{SSB}$ is above all reference points (MSY Btrigger, $B_{p a}$, and $B_{l i m}$ ) and $F$ is above $\mathrm{F}_{\text {msy }}$ but below $\mathrm{F}_{\mathrm{pa}}$ and $\mathrm{F}_{\text {lim. }}$.
32) Management Plan: There is no management strategy agreed for the stock, therefore ICES based its advice on the MSY approach. No agreement on the share of the stock has
been reached for 2022. Despite the acceptance of ICES advice, the total declared quotas in each of the years 2015 to 2021, all exceed the maximum catch advised by ICES.

## General comments

The report section reads well and most information is there. However, the report is not entirely updated to the fact that Russian catch data (catch at age and catch by rectangle) were not submitted to WGWIDE; smaller edits have been reported to the team responsible for the report chapter to be included in this year's report.
The advice sheet is well documented. WGWIDE decided to present the recruitment in the advice sheet as age 2 rather than as age 0 , as abundances of age 0 and age 1 do not reflect year class strength very well. Explanation for this is briefly stated in the figure captions of Figure 1 and 2 in the advice sheet, though not in the text of the sheet.

## Technical comments

The code and input data for the analysis (assessment, and short-term forecast) are all available on SharePoint. An auditor reran the assessment and short-term forecast, reproducing the reported results. Some adjustments were necessary to achieve this (e.g., adjusting paths, installing specific versions of $R$ packages etc.).

To the best of our knowledge, the assessment has been performed correctly according to the stock annex.

The report is rather long. Particularly the sections on surveys (used and unused) could be considerably shortened; at the time of reviewing the text, one survey-section (not used in the assessment) had not been updated.

Table and figure numbers and references to them in the text have been checked.

## Conclusions

The assessment has been performed correctly according to the stock annex.

## Checklist for audit process

## General aspects

- Has the EG answered those TORs relevant to providing advice?
- Is the assessment according to the stock annex description?
- If a management plan is used as the basis of the advice, has been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary?
- Have the data been used as specified in the stock annex?
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex?
- Is there any major reason to deviate from the standard procedure for this stock?
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice?


## Audit of (Stock name)

Date: 02/09/2022
Auditor: Are Salthaug

## General

Advice was provided in 2021 for both 2022 and 2023, thus this year's assessment is exploratory.

## For single stock summary sheet advice:

33) Assessment type: update/SALY
34) Assessment: trends - Category 3 with biennial advice
35) Forecast: not presented
36) Assessment model: Bayesian state space surplus production model fitted using catch data, 6 delta-lognormal estimated IBTS survey indices, and 1 acoustic survey estimate.
37) Data issues: No data issues
38) Consistency: This updated assessment is consistent with the assessment carried out in 2021
39) Stock status: Reference points are undefined.
40) Management Plan: A management strategy proposed by the Pelagic Advisory Council was evaluated and found to be precautionary (ICES, 2015). ICES provides advice for this stock following the precautionary approach, which in this case corresponds to the management strategy from the PelAC.

## General comments

The chapter is easy to follow and interpret.

## Technical comments

None

## Conclusions

The assessment has been performed correctly according to the procedure.

## Checklist for audit process

## General aspects

- Has the EG answered those TORs relevant to providing advice?
- Is the assessment according to the stock annex description?
- If a management plan is used as the basis of the advice, has been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary?
- Have the data been used as specified in the stock annex?
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex?
- Is there any major reason to deviate from the standard procedure for this stock?
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice?


# Audit of Red Gurnard stock assessment 

Date: 02.09.2022
Auditor: Patricia Gonçalves

## General

Assessment of this stock is not possible due to a lack of reliable catch data. Red gurnard is mainly landed as by-catch by demersal trawlers in mixed fisheries, predominantly in divisions $7 \mathrm{~d}, 7 \mathrm{e}$ and 7 h . High discard rates and lack of resolution at a species level make interpretation of spatial trends in catches in other areas problematic.

Landings by country and divisions are available from 2006 to 2021, discard data has been provided for 2015-2021 through Intercatch, 6 survey abundances index for the species area presented from around 1990 to 2021, with a combined biomass index built on these series.

## For single stock summary sheet advice:

1) Assessment type: delta-lognormal assessment (from WKWEST)
2) Assessment: trend analyses
3) Forecast: not presented
4) Assessment model: surveys indices combined using a delta-lognormal model in an index of biomass to evaluate stock trend
5) Data issues: general lack of catch data reported at species level
6) Consistency: undefined
7) Stock status: undefined.
8) Management Plan: there is no management plan.

## General comments

The section of red gurnard is very well structured and documented. The section includes a description regarding the lack of reporting data at species level and also the method used on the computation of a biomass index for this stock.

## Technical comments

## Conclusions

The combined biomass index has been correctly computed. There is no assessment for this stock.

## Checklist for audit process

## General aspects

- Has the EG answered those TORs relevant to providing advice?
- Is the assessment according to the stock annex description?
- If a management plan is used as the basis of the advice, has been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary?
- Have the data been used as specified in the stock annex?
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex?
- Is there any major reason to deviate from the standard procedure for this stock?
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice?


## Working Document to

## ICES Working Group on Widely Distributed Stocks (WGWIDE, No. 1)

ICES HQ, Copenhagen, Denmark, (hybrid meeting) 24. - 30. August 2022

## Cruise report from the International Ecosystem Summer Survey in the Nordic Seas (IESSNS) <br> $1^{\text {st }}$ July $-3^{\text {rd }}$ August 2022



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The International Ecosystem Summer Survey in the Nordic Seas (IESSNS) was performed within approximately 5 weeks from July $1^{\text {st }}$ to August $3^{\text {rd }}$ in 2022 using six vessels from Norway (2), Iceland (1), Faroe Islands (1), Greenland (1) and Denmark (1). The main objective is to provide annual age-segregated abundance index, with an uncertainty estimate, for northeast Atlantic mackerel (Scomber scombrus). The index is used as a tuning series in stock assessment according to conclusions from the 2017 and 2019 ICES mackerel benchmarks. A standardised pelagic swept area trawl method is used to obtain the abundance index and to study the spatial distribution of mackerel in relation to other abundant pelagic fish stocks and to environmental factors in the Nordic Seas, as has been done annually since 2010. Another aim is to
construct a new time series for blue whiting (Micromesistius poutassou) abundance index and for Norwegian spring-spawning herring (NSSH) (Clupea harengus) abundance index. This is obtained by utilizing standardized acoustic methods to estimate their abundance in combination with biological trawling on acoustic registrations. The time series for blue whiting and NSSH now consists of seven years (2016-2022).

The survey coverage area included in calculations of the mackerel index was 2.9 million $\mathrm{km}^{2}$ in 2022, which is $32 \%$ larger coverage compared to 2021 . Survey coverage was increased in the western areas (Iceland and Greenland waters) compared to in 2021. Furthermore, 0.28 million $\mathrm{km}^{2}$ was surveyed in the North Sea in July 2022, but those stations are excluded from the mackerel index calculations.

The total swept-area mackerel index in 2022 was 7.37 million tonnes in biomass and 17.51 billion in numbers, an increase by $43 \%$ for biomass and $43 \%$ for abundance compared to 2021 . In 2022, the most abundant year classes were 2020, 2019, 2010, 2011, respectively. The cohort internal consistency improved compared to last year, particularly for ages 5-8 years.

Most of the surveyed mackerel still appears to be in the Norwegian Sea. The mackerel were more westerly distributed than in the last 2 years.

The zero-line was reached south and north of Iceland and in the west in Greenland waters. It was not reached in the north-western and north-eastern part of the Norwegian Sea but given that the polar front with water too cold for mackerel is usually found close to the northwesternmost catches, we assume that the zero-line was practically reached here as well. Towards the Barents Sea the zero-line was not reached but considered of less quantitative importance based on low catch rates. The zero-line was not reached on the European shelf, where mackerel are present west of the British Isles and in the southern North Sea

Total number of NSSH recorded during IESSNS 2022 was 25.0 billion and the total biomass index was 7.14 million tonnes, or $22 \%$ (abundance) and 17\% (biomass) higher than in 2021. The 2016 year-class (6-year-olds completely dominated in the stock and contributed to $58 \%$ and $56 \%$ to the total biomass and total abundance, respectively, whereas the 2013 year-class ( 9 -year-olds) contributed $8 \%$ and $7 \%$ to the total biomass and total abundance, respectively. The 2016 year-class is fully recruited to the adult stock.

The zero-line of the distribution of the mature part of NSSH was considered to be reached in all directions. The group considered the acoustic biomass estimate of herring in 2022 to be of the similar quality as in the previous survey years. The herring was mainly observed in the upper surface layer as relatively small schools.

Total biomass of blue whiting registered during IESSNS 2022 was 2.2 million tons, which is to the same as in 2021. Estimated stock abundance (ages 1+) was 27.5 billion compared to 26.2 billion in 2021. Age 1 and 2 respectively, dominated the estimate in 2022 as they contributed to $44 \%$ and $33 \%$ (abundance) and $30 \%$ and $33 \%$ (biomass), respectively. The group considered the acoustic biomass estimate of blue whiting to be of good quality in the 2022 IESSNS as in the previous survey years.

As in previous years, there was overlap in the spatio-temporal distribution of mackerel and herring. This overlap occurred between mackerel and North Sea herring in the North Sea and partly in the southernmost part of the Norwegian Sea. There were also some overlapping distributions of mackerel and Norwegian spring-spawning herring (NSSH) particularly in the western, north-western part of the Norwegian Sea.

Other fish species also monitored are lumpfish (Cyclopterus lumpus) and Atlantic salmon (Salmo salar). Lumpfish was caught at $71 \%$ of surface trawl stations distributed across the surveyed area from southwestern part of Iceland, central part of North Sea to southwestern part of the Svalbard. Abundance was greater north of latitude $72^{\circ} \mathrm{N}$ compared to southern areas. A total of 60 North Atlantic salmon were caught in 38 stations both in coastal and offshore areas from $61^{\circ} \mathrm{N}$ to $76^{\circ} \mathrm{N}$ in the upper 30 m of the water column. The salmon ranged from 0.028 kg to 4.1 kg in weight, dominated by post-smolt and 1 sea-winter individuals. We caught from 1 to 6 salmon during individual surface trawl hauls. The length of the salmon ranged from 15 cm to 74 cm , with the highest fraction between 20 cm and 30 cm

Satellite measurements of sea surface temperature (SST) in the Northeast Atlantic in July 2022 show that parts of central Norwegian Sea and areas east and north of Iceland were slightly cooler than the long-term average for July 1990-2009. The northern regions of the Nordic Seas were slightly warmer than the average while the East Greenland Current was cooler that the long-term average. The SST in the Irminger Sea and Iceland Basin were slightly warmer than the average.

The zooplankton biomass varied between areas with a patchy distribution throughout the area. In the Norwegian Sea areas, the average zooplankton biomass was at similar level as last year, slightly lower in Icelandic waters, and higher in Greenlandic waters.

## 2 <br> Introduction

During approximately four weeks of survey in 2022 (1st of July to $3^{\text {rd }}$ of August), six vessels; the M/V "Eros" and M/V "Vendla" from Norway, "Jákup Sverri" operating from Faroe Islands, the R/V "Árni Friðriksson" from Iceland; R/V "Tarajoq" from Greenland and M/V "Ceton", operating in the North Sea by Danish scientists, participated in the International Ecosystem Summer Survey in the Nordic Seas (IESSNS).

The major aim of the coordinated IESSNS was to collect data on abundance, distribution, migration, and ecology of Northeast Atlantic (NEA) mackerel (Scomber scombrus) during its summer feeding migration phase in the Nordic Seas. The resulting abundance index will be used in the stock assessment of NEA mackerel at the annual meeting of ICES working group of widely distributed stocks (WGWIDE). The IESSNS mackerel index time series goes back to 2010. Since 2016, systematic acoustic abundance estimation of both Norwegian spring-spawning herring (Clupea harengus) and blue whiting (Micromesistius poutassou) have also been conducted. This is considered as potential input for stock assessment when the time series are sufficiently long. Furthermore, the IESSNS is a pelagic ecosystem survey collecting data on physical oceanography, plankton, and other fish species such as lumpfish and Atlantic salmon. Opportunistic whale observations are also recorded from Norway, Iceland, and Faroe Islands. The wide geographical coverage, standardization of methods, sampling on many trophic levels and international cooperation around this survey facilitates research on the pelagic ecosystem in the Nordic Seas, see e.g. Nøttestad et al. (2016), Jansen et al. (2016), Bachiller et al. (2018), Olafsdottir et al. (2019), Nikolioudakis et al. (2019).

The methods have evolved over time since the survey was initiated by Norway in the Norwegian Sea in the beginning of the 1990s. The main elements of standardization were conducted in 2010. Smaller improvements have been implemented since 2010. Faroe Islands and Iceland have participated in the joint mackerel-ecosystem survey since 2009. Greenland since 2013 and Denmark from 2018. Greenland did not participate in 2021 but was back in 2022 with their new research vessel R/V "Tarajoq".

The North Sea was included in the survey area for the fifth time in 2022, following the recommendations of WGWIDE. This was done by scientists from DTU Aqua, Denmark. The commercial fishing vessels "Ceton S205" was used. No problems applying the IESSNS methods were encountered. Area coverage, however, was restricted to the northern part of the North Sea at water depths deeper than 50 m (see Appendix 1 for comparison with the 2018-2021 results).

## 3

Material and methods

Coordination of the IESSNS 2022 was done during the WGIPS 2022 virtual meeting in January 2022, and by correspondence in spring and summer 2022. The participating vessels together with their effective survey periods are listed in Table 1.

Overall, the weather conditions were rougher than usual for the Norwegian vessels in the first part of the survey. However, in the second part, the weather conditions and progress were good. The Icelandic vessel, operating in Icelandic waters, experienced calm weather for duration of the survey with no survey delay,
and no CTD or WP2-net sampling was skipped due to high winds. The weather was worse than what is has been previous years for the Faroese vessel which operated in Faroese and Icelandic waters. This resulted in slow progression and the Icelandic vessel had to cover the northernmost transect line for R/V Jakup Sverri. The chartered vessel Ceton had good weather conditions throughout the survey.

During the IESSNS, the special designed pelagic trawl, Multpelt 832, has been applied by all participating vessels since 2012. This trawl is a product of cooperation between participating institutes in designing and constructing a standardized sampling trawl for the IESSNS. The work was led by trawl gear scientist John Willy Valdemarsen, Institute of Marine Research (IMR), Bergen, Norway (Valdemarsen et al. 2014). The design of the trawl was finalized during meetings of fishing gear experts and skippers at meetings in January and May 2011. Further discussions on modifications in standardization between the rigging and operation of Multpelt 832 was done during a trawl expert meeting in Copenhagen 17-18 August 2012, in parallel with the post-cruise meeting for the joint ecosystem survey, and then at the WKNAMMM workshop and tank experiments on a prototype (1:32) of the Multpelt 832 pelagic trawl, conducted as a sequence of trials in Hirtshals, Denmark from 26 to 28 February 2013 (ICES 2013a). The swept area methodology was also presented and discussed during the WGISDAA workshop in Dublin, Ireland in May 2013 (ICES 2013b). The standardization and quantification of catchability from the Multpelt 832 pelagic trawl was further discussed during the mackerel benchmark in Copenhagen in February 2014. Recommendations and requests coming out of the mackerel benchmark in February 2014, were considered and implemented during the IESSNS survey in July-August 2014 and in the surveys thereafter. Furthermore, recommendations and requests resulting from the mackerel benchmark in January-February 2017 (ICES 2017), were carefully considered and implemented during the IESSNS survey in July-August 2017. In 2018, the Faroese and Icelandic vessels employed new, redesigned cod-ends with the capacity to hold 50 tonnes. This was done to avoid the cod-end from bursting during hauling of large catches as occurred at three stations in the 2017 IESSNS.

Table 1. Survey effort by each of the five vessels during the IESSNS 2022. The number of predetermined ("fixed") trawl stations being part of the swept-area stations for mackerel in the IESSNS are shown after the total number of trawl stations.

| Vessel | Effective survey <br> period | Length of cruise <br> track (nmi) | Total trawl stations/ <br> Fixed stations | CTD stations | Plankton stations |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Árni Friðriksson | $4-21 / 7$ | 4082 | $48 / 46$ | 46 |  |
| Jákup Sverri | $1-17 / 7$ | 2768 | $33 / 27$ | 28 | 46 |
| Ceton | $3-12 / 7$ | 1905 | $38 / 34$ | 34 | 28 |
| Vendla | $5 / 7-3 / 8$ | 5369 | $74 / 60$ | 59 | - |
| Eros | $5 / 7-3 / 8$ | 5233 | $67 / 57$ | 56 | 59 |
| Tarajoq | $21 / 7-1 / 8$ | 1522 | $19 / 19$ | 19 | 56 |
| Total | $1 / 7-3 / 8$ | 20879 | $275 / 247$ | 242 | 19 |

### 3.1 Hydrography and Zooplankton

The hydrographical and plankton stations by all vessels combined are shown in Figure 1. Eros, Vendla, Árni Friðriksson and Jákup Sverri were all equipped with a SEABIRD CTD sensor and Árni Friðriksson and Jákup Sverri moreover also had a water rosette. Tarajoq used a SEABIRD SBE 19plus. Ceton used a Seabird SeaCat offline CTD. The CTD-sensors were used for recording temperature, salinity, and pressure (depth) from the surface down to 210 m , or to the bottom when at shallower depths.

Zooplankton was sampled with a WP2-net on 4 of 5 vessels, excluding Ceton which operates in the North Sea. Mesh sizes were $180 \mu \mathrm{~m}$ (Eros and Vendla) and $200 \mu \mathrm{~m}$ (Árni Friðriksson, Jákup Sverri and Tarajoq). The net was hauled vertically from a depth of 200 m (or bottom depth at shallower stations) to the surface at a speed of $0.5 \mathrm{~m} / \mathrm{s}$. All samples were split in two, one half preserved for species identification and enumeration, and the other half dried and weighed. The zooplankton was sorted into three size categories $(\mu \mathrm{m}),>2000,1000-2000,180 / 200-1000$, on the Norwegian and Faroese vessels; and two size fractions ( $\mu \mathrm{m}$ ), $>1000$ and 200-1000, on the Icelandic vessel. Detailed description of the zooplankton and CTD sampling is provided in the survey manual (ICES 2014a).
Two planned CTD and plankton stations were not taken due to bad weather. The number of stations taken by the different vessels is provided in Table 1.

### 3.2 Trawl sampling

All vessels used the standardized Multpelt 832 pelagic trawl (ICES 2013a; Valdemarsen et al. 2014; Nøttestad et al. 2016) for trawling, both for fixed surface stations and for trawling at greater depths to confirm acoustic registrations. Standardization of trawl deployment was emphasised during the survey as in previous years (ICES 2013a; ICES 2014b; ICES 2017). Sensors on the trawl doors, headrope and ground rope of the Multpelt 832 trawl recorded data, and allowed live monitoring, of effective trawl width (actually door spread) and trawl depth. The properties of the Multpelt 832 trawl and rigging on each vessel is reported in Table 2.

Trawl catch was sorted to the highest taxonomical level possible, usually to species for fish, and total weight per species recorded. The processing of trawl catch varied between nations. The Icelandic and Norwegian vessels sorted the whole catch to species but the Faroese vessel sub-sampled the catch before sorting if catches were more than 500 kg . Sub-sample size ranged from 90 kg (if it was clean catch of either herring or mackerel) to 200 kg (if it was a mixture of herring and mackerel). The biological sampling protocol for trawl catch varied between nations in number of specimens sampled per station (Table 3).

Results from the survey expansion southward into the North Sea are analyzed separately from the traditional survey grounds north of latitude $60^{\circ} \mathrm{N}$ as per stipulations from the 2017 mackerel benchmark meeting (ICES 2017). However, data collected with the IESSNS methodology from the Skagerrak and the northern and western part of the North Sea are now available for 2018, 2019, 2020, 2021 and 2022.

Table 2. Trawl settings and operation details during the international mackerel survey in the Nordic Seas from $1^{\text {st }}$ July to $3^{\text {rd }}$ August 2022. The column for influence indicates observed differences between vessels likely to influence performance. Influence is categorized as 0 (no influence) and + (some influence).

| Properties | Árni <br> Friðriksson | Vendla | Ceton | Jákup Sverri | Eros | Tarajoq | Influence |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Trawl producer | Hampiðjan new 2017 trawl | Egersund Trawl AS | Egersund Trawl AS | Vónin | Egersund Trawl AS | Hampiðjan | 0 |
| Warp in front of doors | Dynex-34 mm | Dynex -34 mm | Dynex | Dynex - 38 mm | Dynex-34 mm | Dynex-34 mm | + |
| Warp length during towing | 350 | 350 | 290-305 | 350 | 350-400 | 350 | 0 |
| Difference in warp <br> length port/starb. (m) | 16 | 2-10 | 10 | 0-7 | 5-10 | 10-20 | 0 |
| Weight at the lower wing ends (kg) | $2 \times 400 \mathrm{~kg}$ | $2 \times 400$ | $2 \times 400$ | $2 \times 400$ | $2 \times 400$ | $2 \times 500$ | 0 |
| Setback (m) | 14 | 6 | 6 | 6 | 6 | 6 | + |
| Type of trawl door | Jupiter | Seaflex $7.5 \mathrm{~m}^{2}$ adjustable hatches | Thybron type 15 | Vónin Twister | Seaflex $7.5 \mathrm{~m}^{2}$ adjustable hatches | T-20vf Flipper | 0 |
| Weight of trawl door $(\mathrm{kg})$ | 2200 | 1700 | 1970 | 1650 | 1700 | 2000 | + |
| Area trawl door ( $\mathrm{m}^{2}$ ) | 6 | 7.5 with $25 \%$ hatches (effective 6.5 ) | 7 | 4.5 | 7 with $50 \%$ hatches (effective 6.5) | 7 with $50 \%$ hatches (effective 6.5) | + |
| Towing speed (knots) mean (min-max) | 5.3 (4.6-5.7) | 4.6 (4.1-5.5) | 5.1 (4.5-5.6) | 4.4 (3.6-6) | 4.7 (4.1-5.725) | 4.9 (4.4-5.4) | + |
| Trawl height (m) mean (min-max) | 32 (26-41) | 28-37 | 30 (25-35) | 43 (35-50) | 25-32 | - | + |
| Door distance (m) mean (min-max) | 107 (95-115) | 121.8 (118-126) | 131.2 (126-137) | 115 (107-135) | 135 (113-140) | 105.4 (92-109) | + |
| Trawl width (m)* | 63.75 | 63.8 | 72.0 | 63.4 | 67.5 | 61.4 | + |
| Turn radius (degrees) | 5-10 | 5-12 | 5-10 | 5 BB turn | 5-8 SB turn | 6-8 SB turn | + |
| Fish lock front of codend | Yes | Yes | Yes | Yes | Yes | Yes | + |
| Trawl door depth (port, starboard, m) (minmax) | 3-21, 4-8 | 6-22, 8-23 | 6-15, 8-20 | 7-26, 7-20 | (6-20) | - | + |
| Headline depth (m) | 0 | 0 | 0 | 0 | 0 | 0 | + |
| Float arrangements on the headline | Kite +1 buoy on each wingtip | Kite with fender buoy +2 buoys on each wingtip | Kite with fender buoy + 2 buoys on each wingtip | Kite with +1 buoys on each wingtip | Kite +2 buoy on each wingtips | Kite +1 buoy on each wingtips | + |


| Weighing of catch | All weighted | All weighted | All weighted | Catch < 12 <br> tonnes weighed | All weighted | All weighted | + |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |

* calculated from door distance (Table 6)

Table 3. Protocol of biological sampling during the IESSNS 2022. Numbers denote the maximum number of individuals sampled for each species for the different determinations.

|  | Species | Faroes | Iceland | Norway | Denmark | Greenland |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Length measurements | Mackerel | 200/100* | 150 | 100 | $\geq 125$ | 100/50* |
|  | Herring | 200/100* | 200 | 100 | 75 | 100/50* |
|  | Blue whiting | 200/100* | 100 | 100 | 75 | 100/50* |
|  | Lumpfish | all | all | all | all | All |
|  | Salmon | - | all | all | - | All |
|  | Capelin |  | 100/50^^ | 25-30 |  | 25/25 |
|  | Other fish sp. | 20-50 | 50 | 25 | As appropriate | 25 |
| Weight, sex and | Mackerel | 15-25 | 50 | 25 | *** | 25 |
| maturity determination | Herring | 25-50 | 50 | 25 | 0 | 25 |
|  | Blue whiting | 15-50 | 50 | 25 | 0 |  |
|  | Lumpfish | 10 | $1^{\wedge}$ | 25 | 0 |  |
|  | Salmon | - | 0 | 25 | 0 | 0 |
|  | Capelin |  | 100/50^^ |  |  | 25 |
|  | Other fish sp. | 0 | 0 | 0 | 0 | 25 |
| Otoliths/scales collected | Mackerel | 15-25 | 25 | 25 | *** | 25 |
|  | Herring | 25-50 | 25 | 25 | 0 | 0 |
|  | Blue whiting | 15-50 | 50 | 25 | 0 | 0 |
|  | Lumpfish | 0 | $1^{\wedge}$ | 0 | 0 | 0 |
|  | Salmon | - | 0 | 0 | 0 | 50 |
|  | Capelin |  | 100/50^^ |  |  | 0 |
|  | Other fish sp. | 0 | 0 | 0 | 0 | 50 |
| Fat content | Mackerel | 0 | 10 | 0 | 0 | 0 |
|  | Herring | 0 | $10^{* *}$ | 0 | 0 | 0 |
|  | Blue whiting | 0 | 10 | 0 | 0 | 0 |
| Stomach sampling | Mackerel | 5 | 10 | 10 | 0 | 0 |
|  | Herring | 5 | $10^{* *}$ | 10 | 0 | 0 |
|  | Blue whiting | 5 | 10 | 10 | 0 | 0 |
|  | Other fish sp. | 0 | 0 | 10 | 0 | 0 |
| Tissue for genotyping | Mackerel | 0 | 0 | 0 | 0 | 0 |
|  | Herring | 0 | 0 | 25 | 0 | 0 |

*Length measurements / weighed individuals
**Sampled at every third station
*** Up to one fish per cm-group $<25 \mathrm{~cm}$, two fish $25-30 \mathrm{~cm}$ and three fish $>30 \mathrm{~cm}$ from each station was weighed and aged.
$\wedge$ All live lumpfish were tagged and released, only otoliths taken from fish which were dead when brought aboard.
${ }^{\wedge}$ Numbers changed from 100 to 50 during survey.
This year's survey was well synchronized in time and was conducted over a relatively short period (less than 5 weeks) given the large spatial coverage of around 2.9 million $\mathrm{km}^{2}$ (Figure 1). This was in line with recommendations put forward in 2016 that the survey period should be around four weeks with mid-point around 20th July. The main argument for this time period was to make the survey as synoptic as possible in space and time, and at the same time be able to finalize data and report for inclusion in the assessment for the same year.

## Underwater camera observations during trawling

M/V "Eros" and M/V "Vendla" employed an underwater video camera (GoPro HD Hero 4 and 5 Black Edition, www.gopro.com) to observe mackerel aggregation, swimming behaviour and possible escapement
from the cod end and through meshes. The camera was put in a waterproof box which tolerated pressure down to approximately 100 m depth. No light source was employed with cameras; hence, recordings were limited to day light hours. Some recordings were also taken during night-time when there was midnight sun and good underwater visibility. Video recordings were collected at 70 trawl stations. The camera was attached on the trawl in the transition between 200 mm and 400 mm meshes.

### 3.3 Marine mammals

Opportunistic observations of marine mammals were conducted by scientific personnel and crew members from the bridge between $5^{\text {th }}$ July and $2^{\text {nd }}$ August 2022 onboard M/V "Eros" and M/V "Vendla", and onboard R/V Árni Friðriksson from $4^{\text {th }}$ until $21^{\text {st }}$ July 2022. On board Jákup Sverri ( $1^{\text {st }}-17^{\text {th }}$ July) opportunistic observations were done from the bridge by crew members.

### 3.4 Lumpfish tagging

Lumpfish caught during the survey by vessels R/V "Árni Friðriksson", M/V "Eros", M/V "Vendla" and R/V Tarajoq were tagged with Peterson disc tags and released. When the catch was brought aboard, any lumpfish caught were transferred to a tank with flow-through sea water. After the catch of other species had been processed, all live lumpfish larger than $\sim 15 \mathrm{~cm}$ were tagged. The tags consisted of a plastic disc secured with a titanium pin which was inserted through the rear of the dorsal hump. Contact details of Biopol (www.biopol.is) were printed on the tag. The fish were returned to the tank until all fish were tagged. The fish were then released, and the time of release was noted which was used to determine the latitude and longitude of the release location.

### 3.5 Acoustics

## Multifrequency echosounder

The acoustic equipment onboard Vendla and Eros were calibrated $4^{\text {th }}$ July 2022 for 18, 38, 70, 120 and 200 kHz. Árni Friðriksson was calibrated $28^{\text {th }}$ of May 2022 for frequencies 18, 38, 70, 120 and 200 kHz . Jákup Sverri was calibrated on $24^{\text {th }}$ April 2022 for $18,38,120,200$ and 333 kHz . Tarajoq was calibrated on 20th May 2022 for 18, 38, 120, 200 and 333 kHz . Ceton did not conduct any acoustic data collection because no calibrated equipment was available, and acoustics are done in the same area and period of the year during the ICES coordinated North Sea herring acoustic survey (HERAS). All the other vessels used standard hydro-acoustic calibration procedure for each operating frequency (Foote 1987). CTD measurements were taken in order to get the correct sound velocity as input to the echosounder calibration settings.
Acoustic recordings were scrutinized to herring and blue whiting on daily basis using the post-processing software (LSSS, see Table 4 for details of the acoustic settings by vessel). Acoustic measurements were not conducted onboard Ceton in the North Sea. Species were identified and partitioned using catch information, characteristic of the recordings, and frequency between integration on 38 kHz and on other frequencies by a scientist experienced in viewing echograms.

To estimate the abundance from the allocated NASC-values the following target strengths (TS) relationships were used.

Blue whiting: TS = $20 \log (\mathrm{~L})-65.2 \mathrm{~dB}$ (rev. acc. ICES CM 2012/SSGESST:01)
Herring: TS = $20.0 \log (\mathrm{~L})-71.9 \mathrm{~dB}$

Table 4. Acoustic instruments and settings for the primary frequency ( 38 kHz ) during IESSNS 2022.

|  | R/V Árni <br> Friðriksson | M/V Vendla | Jákup Sverri | Eros | Tarajoq* |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Echo sounder | Simrad EK80 | Simrad EK60 | Simrad EK80 | Simrad EK80 | Simrad EK80 |
| Frequency (kHz) | $\begin{gathered} 18,38,70,120 \\ 200 \end{gathered}$ | $\begin{gathered} 18,38,70,120 \\ 200 \end{gathered}$ | $\begin{gathered} 18,38,70,120 \\ 200,333 \end{gathered}$ | $\begin{gathered} 18,38,70,120 \\ 200,333 \end{gathered}$ | $\begin{gathered} 18,38,70,120 \\ 200,333 \end{gathered}$ |
| Primary transducer | ES38-7 | ES38B | ES38-7 | ES38B | ES38-7 |
| Transducer installation | Drop keel | Drop keel | Drop keel | Drop keel | Drop keel |
| Transducer depth (m) | 9.6 | 8 | 6-9 | 6 | 7 |
| Upper integration limit (m) | 15 | 15 | 15 | 15 |  |
| Absorption coeff. (dB/km) | 10.5 | 9.9 | 9.5 | 9.3 |  |
| Pulse length (ms) | 1.024 | 1.024 | 1.024 | 1.024 | 1.024 |
| Band width (kHz) | 2.425 | 2.43 | 3.064 | 2.43 |  |
| Transmitter power (W) | 2000 | 2000 | 2000 | 2000 | 2000 |
| Angle sensitivity (dB) | 18 | 21.90 | 21.9 | 21.9 |  |
| 2-way beam angle ( dB ) | -20.30 | -20.70 | -20.6 | -20.7 |  |
| TS Transducer gain (dB) | 27.03 | 25.22 | 27.27 | 25.22 |  |
| SA correction (dB) | -0.04 | -0.73 | -0.01 | -0.72 |  |
| 3 dB beam width alongship: | 6.43 | 6.88 | 6.86 | 6.85 |  |
| 3 dB beam width athw. ship: | 6.43 | 6.76 | 6.89 | 6.79 |  |
| Maximum range (m) | 500 | 500 | 500 | 500 | 750 |
| Post processing software | LSSS v.2.12.0 | LSSS 2.12.0 | LSSS 2.12.0 | LSSS 2.12.0 | LSSS 2.12.0 |

M/V Ceton: No acoustic data collection because other survey in the same area in June/July (HERAS).
*Acoustic data collected but not post-processed at the time of report writing.

## Multibeam sonar

Both M/V Eros and M/V Vendla were equipped with the Simrad fisheries sonar SH90 (frequency range: $111.5-115.5 \mathrm{kHz}$ ), with a scientific output incorporated which allow the storing of the beam data for postprocessing. Acoustic multibeam sonar data was stored continuously onboard Eros and Vendla for the entire survey.

## Cruise tracks

The six participating vessels followed predetermined survey lines with predetermined surface trawl stations (Figure 1). Calculations of the mackerel index are based on swept area approach with the survey area split into 10 strata, of which 6 are permanent (1,2,3,7,10 and 13) and four dynamic (4,5,6 and 9) (Figure 2). Distance between predetermined surface trawl stations is constant within stratum but variable
between strata and ranged from 35-90 nmi. The survey design using different strata is done to allow the calculation of abundance indices with uncertainty estimates, both overall and from each stratum in the software program StoX (see Salthaug et al. 2017). Temporal survey progression by vessel along the cruise tracks in July-August 2022 is shown in Figure 3. The cruising speed was between $10-11$ knots if the weather permitted, otherwise the cruising speed was adapted to the weather situation.


Figure 1. Fixed predetermined trawl stations (shown for CTD and WP2) included in the IESSNS from July 1st to August $3^{\text {rd }}$ 2022. At each station a 30 min surface trawl haul, a CTD station ( $0-500 \mathrm{~m}$ ) and WP2 plankton net samples ( $0-200 \mathrm{~m}$ depth) was performed.


Figure 2. Permanent and dynamic strata used in StoX for IESSNS 2022. The survey area is split into 10 strata, of which 6 are permanent $(1,2,3,7,10$ and 13$)$ and four dynamic ( $4,5,6$ and 9 ). The former stratum 8 (along the Norwegian coast) was merged into adjacent strata 1 and 7 . The former stratum 11 (southern Greenland) has not been surveyed the last few years. The former stratum 12 (offshore south of Iceland) is not used any longer, since the southern boundaries of strata 5 and 6 have been converted to dynamic boundaries. For original strata boundaries see WGIPS manual (ICES 2014a).


Figure 3. Temporal survey progression by vessel along the cruise tracks during IESSNS 2022: Blue represents effective survey start (1'st of July) progressing to red representing a five-week span (survey ended $3^{\text {rd }}$ of August). As Ceton and Tarajoq did not submit acoustics, they have been represented by station positions.

### 3.6 StoX

The recorded acoustic and biological data were analysed using the StoX software package which has been used for some years now for WGIPS coordinated surveys. A description of StoX can be found in Johnsen et al. (2019) and here: www.imr.no/forskning/prosjekter/stox. Mackerel swept-area abundance index, excluding the North Sea, was calculated using StoX version 3.5.0. The herring and blue whiting acoustic abundance indices were calculated using StoX version 3.4.0.

### 3.7 Swept area index and biomass estimation

This year the input data for the swept area calculations were taken from the ICES database in contrast to previous years where the input data were extracted from the PGNAPES database.
The swept area age segregated index is calculated separately for each stratum (see stratum definition in Figure 2). Individual stratum estimates are added together to get the total estimate for the whole survey area which is approximately defined by the area between $60^{\circ} \mathrm{N}$ and $77^{\circ} \mathrm{N}$ and $40^{\circ} \mathrm{W}$ and $20^{\circ} \mathrm{E}$ in 2022 . The density of mackerel on a trawl station is calculated by dividing the total number caught by the assumed area swept by the trawl. The area swept is calculated by multiplying the towed distance by the horizontal opening of the trawl. The horizontal opening of the trawl is vessel specific, and the average value across all hauls is calculated based on door spread (Table 5 and Table 6). An estimate of total number of mackerel in a
stratum is obtained by taking the average density based on the trawl stations in the stratum and multiplying this with the area of the stratum.

Table 5. Descriptive statistics for trawl door spread, vertical trawl opening and tow speed for each vessel during IESSNS 2022 at predetermined surface trawl stations. Number of trawl stations used in calculations is also reported. Horizontal trawl opening was calculated using average vessel values for trawl door spread and tow speed (details in Table 6).

|  | Jákup Sverri | RV Árni Friðriksson | Eros | Vendla | Ceton | Tarajoq |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Trawl doors horizontal spread (m) |  |  |  |  |  |  |
| Number of stations | 27 | 44 | 57 | 60 | 34 | 19 |
| Mean | 115 | 107 | 122 | 112 | 131.2 | 105.4 |
| max | 125 | 115 | 136 | 120 | 136.7 | 109.4 |
| min | 107 | 95 | 115 | 100 | 126.4 | 92.4 |
| st. dev. | 4.1 | 3.9 | 4.8 | 4.0 | 2.7 |  |
| Vertical trawl opening (m) |  |  |  |  |  |  |
| Number of stations | 27 | 45 | 59 | 60 | 34 | - |
| Mean | 43 | 31.7 | 35 | 32.5 | 29.5 | - |
| max | 47 | 25.8 | 33 | 37.0 | 35.5 | - |
| min | 35 | 41.3 | 25 | 18.8 | 24.9 | - |
| st. dev. | 3.8 | 3.0 | 2.9 | 4.33 | 2.2 | - |
| Horizontal trawl opening (m) |  |  |  |  |  |  |
| Mean | 63.4 | 63.75 | 67.5 | 63.8 | 72.0 | 61.4 |
| Speed (over ground, nmi) |  |  |  |  |  |  |
| Number of stations | 27 | 45 | 57 | 60 | 34 | 19 |
| Mean | 4.4 | 5.3 | 4.5 | 4.7 | 5.1 | 4.9 |
| max | 6 | 5.7 | 5.3 | 5.6 | 5.6 | 5.4 |
| min | 3.4 | 4.6 | 3.0 | 4.1 | 4.5 | 4.4 |
| st. dev. | 0.5 | 0.2 | 0.5 | 0.3 | 0.2 | 0.2 |

Horizontal trawl opening was calculated using average vessel values for trawl door spread and tow speed (Table 6). The estimates in the formulae were based on flume tank simulations in 2013 (Hirtshals, Denmark) where formulas were developed from the horizontal trawl opening as a function of door spread, for two towing speeds, 4.5 and 5 knots:

Towing speed 4.5 knots: Horizontal opening $(m)=0.441$ * Door spread $(m)+13.094$
Towing speed 5.0 knots: Horizontal opening $(m)=0.3959$ * Door spread (m) 20.094

Table 6. Horizontal trawl opening as a function of trawl door spread and towing speed. Relationship based on simulations of horizontal opening of the Multpelt 832 trawl towed at 4.5 and 5 knots, representing the speed range in the 2014 survey, for various door spread. See text for details. In 2017, the towing speed range was extended from 5.0 to 5.2, in 2020 the door spread was extended to 122 m and in 2022 the towing speed range was extended down to 4.3 knots and up to 5.5 knots. See also Appendix 4.

|  |  |  |  | Towing speed (knots) |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Door spread (m) | 4.3 | 4.4 | 4.5 | 4.6 | 4.7 | 4.8 | 4.9 | 5 | 5.1 | 5.2 | 5.3 | 5.4 | 5.5 |
| 100 | 56.6 | 57 | 57.2 | 57.7 | 58.2 | 58.7 | 59.2 | 59.7 | 60.2 | 60.7 | 61.1 | 61.6 | 62.1 |
| 101 | 56.9 | 57.3 | 57.6 | 58.1 | 58.6 | 59.1 | 59.6 | 60.1 | 60.6 | 61.1 | 61.5 | 62 | 62.4 |
| 102 | 57.3 | 57.7 | 58.1 | 58.6 | 59 | 59.5 | 60 | 60.5 | 60.9 | 61.4 | 61.9 | 62.4 | 62.8 |
| 103 | 57.7 | 58.1 | 58.5 | 59 | 59.5 | 59.9 | 60.4 | 60.9 | 61.3 | 61.8 | 62.3 | 62.7 | 63.2 |
| 104 | 58.2 | 58.6 | 59 | 59.4 | 59.9 | 60.3 | 60.8 | 61.3 | 61.7 | 62.2 | 62.6 | 63.1 | 63.5 |
| 105 | 58.6 | 59 | 59.4 | 59.9 | 60.3 | 60.8 | 61.2 | 61.7 | 62.1 | 62.6 | 63 | 63.5 | 63.9 |
| 106 | 59 | 59.4 | 59.8 | 60.3 | 60.7 | 61.2 | 61.6 | 62.1 | 62.5 | 62.9 | 63.4 | 63.8 | 64.3 |
| 107 | 59.5 | 59.9 | 60.3 | 60.7 | 61.2 | 61.6 | 62 | 62.5 | 62.9 | 63.3 | 63.8 | 64.2 | 64.6 |
| 108 | 59.9 | 60.3 | 60.7 | 61.1 | 61.6 | 62 | 62.4 | 62.9 | 63.3 | 63.7 | 64.1 | 64.6 | 65 |
| 109 | 60.4 | 60.8 | 61.2 | 61.6 | 62 | 62.4 | 62.8 | 63.2 | 63.7 | 64.1 | 64.5 | 64.9 | 65.3 |
| 110 | 60.8 | 61.2 | 61.6 | 62 | 62.4 | 62.8 | 63.2 | 63.6 | 64.1 | 64.5 | 64.9 | 65.3 | 65.6 |
| 111 | 61.3 | 61.6 | 62 | 62.4 | 62.8 | 63.2 | 63.6 | 64 | 64.4 | 64.8 | 65.2 | 65.6 | 66 |
| 112 | 61.7 | 62.1 | 62.5 | 62.9 | 63.3 | 63.7 | 64 | 64.4 | 64.8 | 65.2 | 65.6 | 66 | 66.3 |
| 113 | 62.2 | 62.5 | 62.9 | 63.3 | 63.7 | 64.1 | 64.4 | 64.8 | 65.2 | 65.6 | 65.9 | 66.3 | 66.6 |
| 114 | 62.6 | 63 | 63.4 | 63.7 | 64.1 | 64.5 | 64.9 | 65.2 | 65.6 | 66 | 66.3 | 66.6 | 67 |
| 115 | 63.1 | 63.5 | 63.8 | 64.2 | 64.5 | 64.9 | 65.3 | 65.6 | 66 | 66.3 | 66.7 | 67 | 67.3 |
| 116 | 63.6 | 63.9 | 64.3 | 64.6 | 65 | 65.3 | 65.7 | 66 | 66.4 | 66.7 | 67 | 67.3 | 67.6 |
| 117 | 64 | 64.4 | 64.7 | 65 | 65.4 | 65.7 | 66.1 | 66.4 | 66.8 | 67.1 | 67.4 | 67.7 | 68 |
| 118 | 64.5 | 64.8 | 65.1 | 65.5 | 65.8 | 66.1 | 66.5 | 66.8 | 67.2 | 67.5 | 67.8 | 68 | 68.3 |
| 119 | 64.9 | 65.3 | 65.6 | 65.9 | 66.2 | 66.6 | 66.9 | 67.2 | 67.6 | 67.9 | 68.1 | 68.4 | 68.6 |
| 120 | 65.4 | 65.7 | 66 | 66.3 | 66.6 | 67 | 67.3 | 67.6 | 67.9 | 68.2 | 68.5 | 68.7 | 68.9 |
| 121 | 65.8 | 66.1 | 66.5 | 66.8 | 67.1 | 67.4 | 67.7 | 68 | 68.3 | 68.6 | 68.8 | 69 | 69.3 |
| 122 | 66.2 | 66.5 | 66.9 | 67.2 | 67.5 | 67.8 | 68.1 | 68.4 | 68.7 | 69 | 69.1 | 69.4 | 69.6 |

### 4.1 Hydrography

Satellite measurements (NOAA OISST) of sea surface temperature (SST) in the central areas in the Northeast Atlantic in July 2022 were slightly cooler than the long-term average for July 1990-2009 based on SST anomaly plots (Figure 4). The northern regions of the Nordic Seas were slightly warmer than the average while the East Greenland Current was cooler that the long-term average. The SST in the Irminger Sea and Iceland Basin were slightly warmer than the average.

It should be mentioned that the NOAA SST are sensitive to the weather conditions (i.e. wind and cloudiness) prior to and during the observations and do therefore not necessarily reflect the oceanographic condition of the water masses in the areas, as seen when comparing detailed in situ features of SSTs between years (Figures 4-5). However, since the anomaly is based on the average for the whole month of July, it should give representative results of the surface temperature.

In situ measurements from the survey showed that the upper layer ( 10 m depth) in 2022 generally was slightly cooler than 2021, except for the northern areas with slightly warmer surface layer (Figure 5, upper left panel). However, in the deeper layers ( 50 m and deeper; Figure 5 , upper right panel and bottom panels), the hydrographical features in the area were similar to previous years. The increased presence of the East Icelandic Current visible in the surface might be due to the relatively cold July month in 2022 with less summer stratification in the that area. At all depths there is a clear signal from the cold East Icelandic Current which carries cold and fresh water into the central and south-eastern part of the Norwegian Sea. Along the Norwegian Shelf and in the southernmost areas, the water masses are dominated by warmer waters of Atlantic origin.


Figure 4. Annual sea surface temperature anomaly ( -4 to $+4^{\circ} \mathrm{C}$ ) in Northeast Atlantic for the month of July from 2010 to 2022 showing warm and cold conditions in comparison to the average for July 1990-2009. Based on monthly averages of daily Optimum Interpolation Sea Surface Temperature (Ver. 2.1 NOAA OISST, AVHRR-only, Banzon et al. 2016, https://www.ncdc.noaa.gov/oisst).


Figure 5. Temperature $\left({ }^{\circ} \mathrm{C}\right)$ at 10, 50, 100 and 400 m depth in Nordic Seas and the North Sea in July-August 2022. 500 m and 2000 m depth contours are shown in light grey.

### 4.2 Zooplankton

The zooplankton biomass varied between areas with a patchy distribution throughout the area (Figure 6a). In the Norwegian Sea areas, the average zooplankton biomass was at the same level as last year.

The time-series of average zooplankton biomass averaged by three subareas: Greenland region, Iceland region and the Norwegian Sea region is shown in Figure 6b (see definitions in legend). In the Greenland area an increase was observed in 2022 compared to the low 2020 value (not surveyed in 2021). In the Icelandic region the level was the same as in 2021. The Greenland and Iceland time-series co-vary (20142020, $2022 \mathrm{r}=0.89$ ). The biomass index in the Norwegian Sea varied less compared to the other two indices, and showed a slight decrease in 2022 from a relatively stable level since 2013 (Figure 6b). The lower variability might in part be explained by the more homogeneous oceanographic conditions in the area defined as Norwegian Sea.

These plankton indices should be treated with some caution as it is only a snapshot of the standing stock biomass, not of the actual production in the area, which complicates spatio-temporal comparisons.


Figure 6a. Zooplankton biomass ( g dw/m2, 0-200 m) in Nordic Seas in July-August 2022.500 m and 2000 m depth contours are shown in light grey.


Figure 6b. Zooplankton biomass indices ( $\mathrm{g} \mathrm{dw} / \mathrm{m} 2,0-200 \mathrm{~m}$ ). Time-series (2010-2022) of mean zooplankton biomass for three subareas within the survey range: Norwegian Sea (between $14^{\circ} \mathrm{W}-17^{\circ} \mathrm{E}$ \& north of $61^{\circ} \mathrm{N}$ ), Icelandic waters $\left(14^{\circ} \mathrm{W}-30^{\circ} \mathrm{W}\right)$ and Greenlandic waters (2014-2022, west of $30^{\circ} \mathrm{W}$ ).

### 4.3 Mackerel

The total swept-area mackerel index in 2022 was 7.37 million tonnes in biomass and 17.51 billion in numbers, an increase of $43 \%$ for biomass and $43 \%$ for abundance compared to 2021 . The survey coverage area (excl. the North Sea, 0.28 million $\mathrm{km}^{2}$ ) was 2.9 million $\mathrm{km}^{2}$ in 2022 , which is $32 \%$ larger compared to 2021. The mackerel catch rates varied from zero to 103 tonnes $/ \mathrm{km}^{2}$ (mean $=2.3$ tonnes $/ \mathrm{km}^{2}$, with two very large values ( 70 and 103, see CPUE by station in Figure 7 together with the mean catch rates per $2^{\circ}$ lat. x $4^{\circ}$ lon. rectangles). These two hauls contributed with $33 \%$ of the total biomass index (Appendix 3). This is also explains the very high uncertainty of the estimate. It is worth noting that western part of the northern Norwegian Sea (stratum 9) was oversampled as three surface trawl stations were added, at the dynamic stratum boundary, at only half the distance from next station, 35 nm instead of 70 nm . Mackerel was caught at all these station and max catch per station was about one ton. All three stations were included in the index calculations and the dynamic stratum boundary extended 35 nm westward of these three stations.

Most of the surveyed mackerel still appears to be in the Norwegian Sea. The mackerel were more westerly distributed than in the last 2 years.

The zero-line was reached south and north of Iceland and in the west in Greenland waters. It was not reached in the northwestern and northeastern part of the Norwegian Sea but given that the polar front with water too cold for mackerel is usually found close the northwestern most catches, we assume that the zeroline was practically reached here as well. Towards the Barent Sea the zero-line was not reached but considered of less quantitative importance based on low catch rates. The zero-line was not reached on the European shelf, where mackerel are present west of the British Isles and in the southern North Sea (Campbell, 2021).


Figure 7. Mackerel catch rates by Multpelt 832 pelagic trawl haul at predetermined surface trawl stations (circle areas represent catch rates in $\mathrm{kg} / \mathrm{km}^{2}$ ) overlaid on mean catch rates per standardized rectangles ( $2^{\circ}$ lat. $x 4^{\circ}$ lon.) in Nordic Seas in July-August 2022.


Figure 8. Annual distribution of mackerel proxied by the absolute distribution of mean mackerel catch rates per standardized rectangles ( $2^{\circ}$ lat. x $4^{\circ}$ lon.), from Multpelt 832 pelagic trawl hauls at predetermined surface trawl stations in Nordic Seas in June-August 2010-2022. Colour scale goes from white (=0) to red (= maximum value for the highest year).


Figure 9. Annual distribution of mackerel proxied by the relative distribution of mean mackerel catch rates per standardized rectangles ( $2^{\circ}$ lat. x $4^{\circ}$ lon.), from Multpelt 832 pelagic trawl hauls at predetermined surface trawl stations stations in Nordic Seas in June-August 2010-2022. Colour scale goes from white (= 0) to red (= maximum value for the given year).


Figure 10. Average weight of mackerel at predetermined surface trawl stations during IESSNS 2022.

The mackerel weight varied between 48 to 872 g with an average of 388 g . The length of mackerel caught in the pelagic trawl hauls onboard the five vessels varied from 18 to 46 cm , with an average of 33 cm . Individuals in the length range $30-31 \mathrm{~cm}$ and $36-40 \mathrm{~cm}$ dominated in numbers and biomass. Mackerel length distribution followed the same overall pattern as previous years both in the Norwegian Sea, with increasing size towards the distribution boundaries in the north and the north-west, and in the western area with increasing size westward (Figure 10). The spatial distribution and overlap between the major pelagic fish species (mackerel, herring, blue whiting) in 2022 according to surface trawl catches is shown in Figure 11.


Figure 11. Distribution and spatial overlap between mackerel, herring, and blue whiting, at all surface trawl stations during IESSNS 2022. Vessel tracks are shown as continuous lines and predetermined surface trawl stations with no catch of the three species is displayed as + .

## Swept area analyses from standardized pelagic trawling with Multpelt 832

The swept area estimates of mackerel biomass from the 2022 IESSNS were based on abundance of mackerel per stratum (see strata definition in Figure 2) and calculated in StoX version 3.5.0. Mackerel abundance index in 2022 was slightly lower than the time series mean of 18.9 billion (Table 7a; Figure 12) and the biomass index was slightly higher than the mean of 7.28 million tons (Table 7c). Mackerel estimates of abundance, biomass and mean weight by age and length are displayed in Table 7d. There is no pattern in changing size-at-age between years (Table 7 b ). In 2022, the most abundant year-classes were respectively 2020 (age 2), 2019 (age 3), 2012 (age 10), and 2011 (age 11) (Figure 13). Mackerel of age 1, 2 and to some extent also age 3 are not completely recruited to the survey (Figure 15), information on recruitment is therefore uncertain. Variance in age index estimation is provided in Figure 14.

The overall internal consistency was slightly improved compared to last year (Figure 16). There is a good to strong internal consistency for the younger ages (1-5 years) and older ages (9-14 years) with r between 0.70 and 0.91 . However, the internal consistency is more variable between age 5 to 9 , with $r=0.43$ between 5 and 6 years ( $\mathrm{r}=0.43$ ) and $\mathrm{r}=0.22$ between 7 and 8 years. The reason for the relatively low consistency for these year groups are not clear.

Mackerel index calculations from the catch in the North Sea (Figure 2) were excluded from the index calculations presented in the current chapter to facilitate comparison to previous years and because the 2017 mackerel benchmark stipulated that trawl stations south of latitude $60{ }^{\circ} \mathrm{N}$ be excluded from index
calculations (ICES 2017). Results from the mackerel index calculations for the North Sea are presented in Appendix 1.

The indices used for NEA mackerel stock assessment in WGIWIDE are the number-at-age indices for age 3 to 11 year (Table 7a).


Figure 12. Estimated total stock biomass (upper panel) and total stock numbers (lower panel) of mackerel from StoX for the years 2007 and from 2010 to 2022. The red dots are baseline estimates, the black dots are mean of 1000 bootstrap replicates while the error bars represent $90 \%$ confidence intervals based on the bootstrap.


Figure 13. Mackerel age distribution in numbers (\%) and in biomass (\%) from IESSNS 2022.


Figure 14. Number by age for mackerel in 2022. Boxplot of abundance and relative standard error (CV) obtained by bootstrapping with 1000 replicates using the StoX software.

Table 7. a-d) StoX baseline (point estimate) time series of the IESSNS showing (a) age-disaggregated abundance indices of mackerel (billions), (b) mean weight (grams) per age, (c) estimated biomass at age (million tonnes) in 2007 and from 2010 to 2022, and (d) estimates of abundance, biomass and mean weight by age and length.

| a) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year\Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14(+) | Tot N |
| 2007 | 1.33 | 1.86 | 0.90 | 0.24 | 1.00 | 0.16 | 0.06 | 0.04 | 0.03 | 0.01 | 0.01 | 0.00 | 0.01 | 0.00 | 5.65 |
| 2010 | 0.03 | 2.80 | 1.52 | 4.02 | 3.06 | 1.35 | 0.53 | 0.39 | 0.20 | 0.05 | 0.03 | 0.02 | 0.01 | 0.01 | 13.99 |
| 2011 | 0.21 | 0.26 | 0.87 | 1.11 | 1.64 | 1.22 | 0.57 | 0.28 | 0.12 | 0.07 | 0.06 | 0.02 | 0.01 | 0.00 | 6.42 |
| 2012 | 0.50 | 4.99 | 1.22 | 2.11 | 1.82 | 2.42 | 1.64 | 0.65 | 0.34 | 0.12 | 0.07 | 0.02 | 0.01 | 0.01 | 15.91 |
| 2013 | 0.06 | 7.78 | 8.99 | 2.14 | 2.91 | 2.87 | 2.68 | 1.27 | 0.45 | 0.19 | 0.16 | 0.04 | 0.01 | 0.02 | 29.57 |
| 2014 | 0.01 | 0.58 | 7.80 | 5.14 | 2.61 | 2.62 | 2.67 | 1.69 | 0.74 | 0.36 | 0.09 | 0.05 | 0.02 | 0.00 | 24.37 |
| 2015 | 1.20 | 0.83 | 2.41 | 5.77 | 4.56 | 1.94 | 1.83 | 1.04 | 0.62 | 0.32 | 0.08 | 0.07 | 0.04 | 0.02 | 20.72 |
| 2016 | <0.01 | 4.98 | 1.37 | 2.64 | 5.24 | 4.37 | 1.89 | 1.66 | 1.11 | 0.75 | 0.45 | 0.20 | 0.07 | 0.07 | 24.81 |
| 2017 | 0.86 | 0.12 | 3.56 | 1.95 | 3.32 | 4.68 | 4.65 | 1.75 | 1.94 | 0.63 | 0.51 | 0.12 | 0.08 | 0.04 | 24.22 |
| 2018 | 2.18 | 2.50 | 0.50 | 2.38 | 1.20 | 1.41 | 2.33 | 1.79 | 1.05 | 0.50 | 0.56 | 0.29 | 0.14 | 0.09 | 16.92 |
| 2019 | 0.08 | 1.35 | 3.81 | 1.21 | 2.92 | 2.86 | 1.95 | 3.91 | 3.82 | 1.50 | 1.25 | 0.58 | 0.59 | 0.57 | 26.4 |
| 2020 | 0.04 | 1.10 | 1.43 | 3.36 | 2.13 | 2.53 | 2.53 | 2.03 | 2.90 | 3.84 | 1.50 | 1.18 | 0.92 | 0.98 | 26.47 |
| 2021 | 0.09 | 2.13 | 0.71 | 1.22 | 1.53 | 0.37 | 1.29 | 0.81 | 1.05 | 0.97 | 0.93 | 0.46 | 0.34 | 0.33 | 12.22 |
| 2022 | 0.02 | 3.91 | 2.36 | 0.94 | 1.31 | 1.04 | 0.60 | 0.96 | 1.00 | 1.86 | 1.61 | 0.90 | 0.56 | 0.45 | 17.51 |

b)

| Year $\backslash$ Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2007 | 133 | 233 | 323 | 390 | 472 | 532 | 536 | 585 | 591 | 640 | 727 | 656 |
| 2010 | 133 | 212 | 290 | 353 | 388 | 438 | 512 | 527 | 548 | 580 | 645 | 683 |
| 2011 | 133 | 278 | 318 | 371 | 412 | 440 | 502 | 537 | 564 | 541 | 570 | 632 |
| 2012 | 112 | 188 | 286 | 347 | 397 | 414 | 437 | 458 | 488 | 523 | 514 | 615 |
| 209 | 509 |  |  |  |  |  |  |  |  |  |  |  |


| 2013 | 96 | 184 | 259 | 326 | 374 | 399 | 428 | 445 | 486 | 523 | 499 | 547 | 677 |
| ---: | ---: | ---: | ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2014 | 228 | 275 | 288 | 335 | 402 | 433 | 459 | 477 | 488 | 533 | 603 | 544 | 537 |
| 2015 | 128 | 290 | 333 | 342 | 386 | 449 | 463 | 479 | 488 | 505 | 559 | 568 | 583 |
| 2016 | 95 | 231 | 324 | 360 | 371 | 394 | 440 | 458 | 479 | 488 | 494 | 523 | 511 |
| 2017 | 86 | 292 | 330 | 373 | 431 | 437 | 462 | 487 | 536 | 534 | 542 | 574 | 589 |
| 2018 | 67 | 229 | 330 | 390 | 420 | 449 | 458 | 477 | 486 | 515 | 534 | 543 | 575 |
| 2019 | 153 | 212 | 325 | 352 | 428 | 440 | 472 | 477 | 490 | 511 | 524 | 564 | 545 |
| 2020 | 99 | 213 | 315 | 369 | 394 | 468 | 483 | 507 | 520 | 529 | 539 | 567 | 575 |
| 2021 | 140 | 253 | 357 | 377 | 409 | 451 | 467 | 487 | 497 | 505 | 516 | 523 | 544 |
| 2022 | 125 | 263 | 330 | 408 | 438 | 431 | 462 | 508 | 525 | 519 | 531 | 531 | 549 |

c)

| Year $\backslash$ Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | $14(+)$ | Tot B |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2007 | 0.18 | 0.43 | 0.29 | 0.09 | 0.47 | 0.09 | 0.03 | 0.02 | 0.02 | 0.01 | 0.01 | 0.00 | 0.01 | 0.00 | 1.64 |
| 2010 | 0.00 | 0.59 | 0.44 | 1.42 | 1.19 | 0.59 | 0.27 | 0.20 | 0.11 | 0.03 | 0.02 | 0.01 | 0.01 | 0.00 | 4.89 |
| 2011 | 0.03 | 0.07 | 0.28 | 0.41 | 0.67 | 0.54 | 0.29 | 0.15 | 0.07 | 0.04 | 0.03 | 0.01 | 0.01 | 0.00 | 2.69 |
| 2012 | 0.06 | 0.94 | 0.35 | 0.73 | 0.72 | 1.00 | 0.72 | 0.30 | 0.17 | 0.06 | 0.03 | 0.01 | 0.00 | 0.00 | 5.09 |
| 2013 | 0.01 | 1.43 | 2.32 | 0.70 | 1.09 | 1.15 | 1.15 | 0.56 | 0.22 | 0.10 | 0.08 | 0.02 | 0.01 | 0.01 | 8.85 |
| 2014 | 0.00 | 0.16 | 2.24 | 1.72 | 1.05 | 1.14 | 1.23 | 0.80 | 0.36 | 0.19 | 0.05 | 0.03 | 0.01 | 0.00 | 8.98 |
| 2015 | 0.15 | 0.24 | 0.80 | 1.97 | 1.76 | 0.87 | 0.85 | 0.50 | 0.30 | 0.16 | 0.04 | 0.04 | 0.02 | 0.01 | 7.72 |
| 2016 | $<0.01$ | 1.15 | 0.45 | 0.95 | 1.95 | 1.72 | 0.83 | 0.76 | 0.53 | 0.37 | 0.22 | 0.10 | 0.04 | 0.04 | 9.11 |
| 2017 | 0.07 | 0.03 | 1.18 | 0.73 | 1.43 | 2.04 | 2.15 | 0.86 | 1.04 | 0.33 | 0.28 | 0.07 | 0.05 | 0.03 | 10.29 |
| 2018 | 0.15 | 0.57 | 0.16 | 0.93 | 0.50 | 0.63 | 1.07 | 0.85 | 0.51 | 0.26 | 0.30 | 0.16 | 0.08 | 0.05 | 6.22 |
| 2019 | 0.01 | 0.29 | 1.24 | 0.43 | 1.25 | 1.26 | 0.92 | 1.86 | 1.87 | 0.77 | 0.65 | 0.33 | 0.32 | 0.32 | 11.52 |
| 2020 | $<0.01$ | 0.23 | 0.45 | 1.24 | 0.84 | 1.18 | 1.22 | 1.03 | 1.51 | 2.03 | 0.81 | 0.67 | 0.53 | 0.58 | 12.33 |
| 2021 | 0.01 | 0.54 | 0.25 | 0.46 | 0.62 | 0.17 | 0.60 | 0.39 | 0.52 | 0.49 | 0.48 | 0.24 | 0.18 | 0.19 | 5.15 |
| 2022 | 0.00 | 1.03 | 0.78 | 0.39 | 0.57 | 0.45 | 0.28 | 0.49 | 0.52 | 0.97 | 0.85 | 0.48 | 0.31 | 0.26 | 7.37 |


| d) <br> Length (cm) | Age in years (year class) |  |  |  |  |  |  |  |  |  |  |  |  |  | Number$\left(10^{\wedge} 6\right)$ | Biomass$\left(10^{\wedge} 6 \mathrm{~kg}\right)$ | Mean weight (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13+ | NA |  |  |  |
|  | 2021 | 2020 | 2019 | 2018 | 2017 | 2016 | 2015 | 2014 | 2013 | 2012 | 2011 | 2010 |  |  |  |  |  |
| 18-19 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0 | 46.7 |
| 19-20 | 8 |  |  |  |  |  |  |  |  |  |  |  |  |  | 8 | 0 | 58.1 |
| 20-21 | 3 |  |  |  |  |  |  |  |  |  |  |  |  |  | 3 | 0 | 66.4 |
| 21-22 | 3 |  |  |  |  |  |  |  |  |  |  |  |  |  | 3 | 0 | 74.5 |
| 22-23 |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 0 | 0 | 88.0 |
| 23-24 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 0 |  |
| 24-25 |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 0 | 0 | 126.0 |
| 25-26 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 0 |  |
| 26-27 | 0 |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 0 | 166.0 |
| 27-28 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 0 |  |
| 28-29 | 8 | 64 |  |  |  |  |  |  |  |  |  |  |  |  | 72 | 15 | 214.4 |
| 29-30 |  | 805 | 30 | 3 |  |  |  |  |  |  |  |  |  |  | 838 | 200 | 239.1 |
| 30-31 |  | 1809 | 9 |  | 4 | 3 |  |  |  |  |  |  |  |  | 1825 | 471 | 258.1 |
| 31-32 |  | 993 | 353 | 2 | 34 | 5 |  |  |  |  |  |  |  |  | 1386 | 390 | 281.7 |
| 32-33 |  | 178 | 637 | 25 | 5 | 5 |  |  |  |  |  |  |  |  | 851 | 265 | 311.5 |
| 33-34 |  | 34 | 711 | 96 | 43 | 10 | 3 |  | 0 | 0 |  |  |  |  | 896 | 301 | 336.3 |
| 34-35 | 0 | 16 | 384 | 95 | 133 | 52 | 0 |  |  |  |  | 0 |  |  | 681 | 248 | 363.6 |
| 35-36 |  | 3 | 204 | 70 | 104 | 279 | 125 | 13 | 7 | 3 | 2 |  |  |  | 808 | 313 | 387.6 |
| 36-37 |  |  | 26 | 477 | 219 | 236 | 77 | 38 | 1 | 17 | 26 | 0 | 4 |  | 1120 | 471 | 420.5 |
| 37-38 |  | 4 | 1 | 168 | 439 | 269 | 153 | 127 | 84 | 403 | 97 | 43 | 11 |  | 1799 | 835 | 464.1 |
| 38-39 |  |  | 1 | 7 | 171 | 161 | 158 | 461 | 195 | 435 | 527 | 295 | 226 |  | 2639 | 1321 | 500.5 |
| 39-40 |  | 4 | 0 | 1 | 157 | 17 | 41 | 198 | 511 | 465 | 497 | 301 | 188 |  | 2382 | 1256 | 527.5 |
| 40-41 |  |  |  |  | 0 | 3 | 28 | 111 | 174 | 493 | 341 | 159 | 297 |  | 1606 | 910 | 566.5 |
| 41-42 |  |  |  | 0 |  | 4 | 12 | 4 | 19 | 40 | 98 | 82 | 203 |  | 464 | 280 | 606.3 |
| 42-43 |  |  |  |  |  |  |  | 2 | 5 | 6 | 17 | 8 | 56 |  | 94 | 61 | 642.4 |
| 43-44 |  |  |  |  |  |  |  | 3 |  |  | 1 | 9 | 21 |  | 33 | 22 | 687.6 |
| 44-45 |  |  |  |  |  |  |  |  |  |  |  |  | 3 |  | 3 | 2 | 704.0 |
| 45-46 |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 1 | 1 | 803.8 |
| 46-47 |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 0 | 0 | 872.0 |
| TSN(mill) | 23.4 | 3909.5 | 2355.9 | 944.4 | 1307.8 | 1043.4 | 598.2 | 956.1 | 995.9 | 1862.0 | 1605.7 | 897.6 | 1011.3 | 2.2 | 17513.5 | 7365 |  |
| TSB(1000 t) | 2.9 | 1028.7 | 777.1 | 385.4 | 572.3 | 449.4 | 276.5 | 485.8 | 522.7 | 967.2 | 851.5 | 476.6 | 567.8 | 1.4 | 7365.3 |  |  |
| Mean length(cm) | 22.7 | 30.2 | 32.7 | 35.5 | 36.4 | 36.2 | 37.1 | 38.2 | 38.8 | 38.6 | 38.9 | 39.0 |  |  |  |  |  |
| Mean weight(g) | 125 | 263 | 330 | 408 | 438 | 431 | 462 | 508 | 525 | 519 | 531 | 531 |  |  |  |  |  |

Table 8. Bootstrap estimates from StoX (based on 1000 replicates) of mackerel in 2022. Numbers by age and total number (TSN) are in millions and total biomass (TSB) in million tons.

|  | Age | 5th percentile | Median | $\begin{array}{r} \text { 95th } \\ \text { percentile } \end{array}$ | Mean | SD | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 3.9 | 20.3 | 41.5 | 21.3 | 12.0 | 0.56 |
|  | 2 | 1945.0 | 3822.1 | 6590.4 | 3974.3 | 1416.0 | 0.36 |
|  | 3 | 1019.0 | 2341.4 | 4200.5 | 2384.2 | 1002.9 | 0.42 |
|  | 4 | 382.1 | 950.4 | 1858.6 | 988.8 | 483.8 | 0.49 |
|  | 5 | 575.8 | 1311.0 | 2357.8 | 1380.1 | 551.4 | 0.40 |
|  | 6 | 617.4 | 1006.7 | 1609.3 | 1043.2 | 306.7 | 0.29 |
|  | 7 | 434.8 | 602.8 | 845.6 | 618.9 | 136.3 | 0.22 |
|  | 8 | 704.6 | 972.9 | 1250.1 | 980.5 | 166.5 | 0.17 |
|  | 9 | 696.4 | 977.0 | 1367.3 | 991.6 | 207.9 | 0.21 |
|  | 10 | 874.3 | 1801.7 | 3269.0 | 1872.5 | 763.0 | 0.41 |
|  | 11 | 1068.4 | 1534.8 | 2206.6 | 1567.8 | 353.6 | 0.23 |
|  | 12 | 487.9 | 808.9 | 1340.7 | 849.8 | 277.5 | 0.33 |
|  | 13 | 283.9 | 522.3 | 983.6 | 556.4 | 236.2 | 0.42 |
|  | 14 | 162.4 | 241.0 | 343.9 | 245.3 | 55.7 | 0.23 |
|  | 15 | 88.7 | 141.7 | 201.7 | 142.8 | 34.9 | 0.24 |
|  | 16 | 33.6 | 78.2 | 112.2 | 74.5 | 25.6 | 0.34 |
|  | 17 | 6.5 | 14.1 | 25.4 | 14.8 | 5.6 | 0.38 |
|  | 18 | 1.1 | 6.0 | 12.7 | 6.6 | 3.6 | 0.55 |
|  | 19 | 0.0 | 2.5 | 7.6 | 2.7 | 2.7 | 1.03 |
| TSN |  | 11388 | 17196 | 26156 | 17719 | 4558 | 0.26 |
| TSB |  | 4.82 | 7.23 | 10.89 | 7.44 | 1.87 | 0.25 |



Figure 15. Catch curves for the years 2010; 2012-2022. Each cohort of mackerel is marked by a uniquely coloured line that connects the estimates indicated by the respective ages.


Figure 16 Internal consistency of the of mackerel density index from 2012 to 2022. Ages indicated by white numbers in grey diagonal cells. Statistically significant positive correlations ( $p<0.05$ ) are indicated by regression lines and red cells in upper left half. Correlation coefficients ( $r$ ) are given in the lower right half.

The swept area method assumes that potential distribution of mackerel outside the survey area - both vertically and horizontally - is a constant percentage of the total biomass. In some years, this assumption may be violated, e.g. mackerel may be distributed below the footrope of the trawl or if the proportion of mackerel outside the survey coverage varies among years. In order to improve the precision of the swept area estimate it would be beneficial to extend the survey coverage further south, such that it covers the southwestern waters south of $60^{\circ} \mathrm{N}$, e.g. UK waters.

The standard swept area method using the average horizontal trawl opening by each participating vessel (ranging 56.6.5-75.4 m; Table 5), assuming that a constant fraction of the mackerel inside the horizontal trawl opening are caught. Further, that if mackerel is distributed below the depth of the trawl (footrope), this fraction is assumed constant from year to year.

As in previous years, there was overlap in the spatio-temporal distribution of mackerel and herring (Figure 11). This overlap occurred mostly between mackerel and Norwegian spring-spawning herring (NSSH) in the western, north-western and north-eastern part of the Norwegian Sea.

### 4.4 Norwegian spring-spawning herring

Norwegian spring-spawning herring (NSSH) was recorded in the southwestern (east and north of Iceland), central and northern part of the Norwegian Sea basin (Figure 17a). The acoustic registrations in the eastern parts of the Norwegian Sea were low in July 2022. A relatively large part of the adult NSSH stock was distributed north of $68^{\circ} \mathrm{N}$ (Figure 17a). Herring registrations south of $62^{\circ} \mathrm{N}$ in the eastern part were allocated to a different stock, North Sea herring, while the herring to the south and west in Icelandic waters (west of $14^{\circ} \mathrm{W}$ south of Iceland) were allocated to Icelandic summer-spawners - these were removed from the biomass estimation of NSSH, except some putative North Sea herring in the southeastern area north of Shetland (Figure 17b).

The total number of NSSH recorded during IESSNS 2022 was 25.0 billion and the total biomass index was 7.14 million tonnes, or $22 \%$ (abundance) and $17 \%$ (biomass) higher than 2021 (Table 10 and 11).

The 2016 year-class (6 year-olds) completely dominated in the stock and contributed $58 \%$ and $56 \%$ to the total biomass and total abundance, respectively, whereas the 2013 year-class ( 9 year-olds) contributed $8 \%$ and $7 \%$ to the total biomass and total abundance, respectively (Figure 18 and Table 9). The 2016 year-class is fully recruited to the adult stock.

Bootstrap estimates of numbers by age are shown in Figure 18. The uncertainty (CV) around the age disaggregated abundance indices from the 2022 survey was very low, except for the highly dominating 6 year-olds (2016 year class) (Figure 18).

The internal consistency among year classes was generally very high for age classes 4 years and older, with the lowest correlation, for the youngest year classes, as expected since they are not fully recruited into the survey (Figure 19).

The 0-boundary of the distribution of the adult part of NSSH was considered to be reached in all directions. The herring was mainly observed in the upper surface layer as relatively small schools. This shallow distribution of herring might have led to an unknown portion of herring being in the "blind zone" above the transducer depth of the vessels (i.e. shallower than 10-15 m, Table 4), and therefore not being registered by the vessels. The group considered the acoustic biomass estimate of herring in 2022 to be of the similar quality as in the previous survey years.


Figure 17a. The sa/Nautical Area Scattering Coefficient (NASC) values of herring along the cruise tracks in 2022 presented as contour lines. Values north of $62^{\circ} \mathrm{N}$, and east of $14^{\circ} \mathrm{W}$, are considered to be Norwegian spring-spawning herring. South and west of this area the herring observed are other stocks, i.e. Icelandic summer spawners, Faroese autumn spawners and North Sea herring in the southeast.


Figure 17b. The $\mathrm{s}_{\mathrm{A}} /$ Nautical Area Scattering Coefficient (NASC) values of Norwegian spring-spawning herring along the cruise tracks in 2022, presented as bar plot.


Figure 18. Abundance by age for Norwegian spring-spawning herring during IESSNS 2022. Boxplot of abundance and relative standard error (CV) obtained by bootstrapping with 500 replicates using the StoX software.

Table 9. Estimates of abundance, mean weight and mean length of Norwegian spring-spawning herring based on calculation in StoX (bootstrap) for IESSNS 2022.


Table 10. IESSNS bootstrap time series (mean of 1000 replicates) from 2016 to 2022. StoX abundance estimates of Norwegian spring-spawning herring (millions).

| Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ | TSB(1000 t) |  |
| 2016 | 38 | 119 | 747 | 577 | 1622 | 1636 | 1967 | 1588 | 1274 | 2001 | 2164 | 6245 | 6676 |  |
| 2017 | 1232 | 240 | 1318 | 4653 | 1003 | 1184 | 795 | 1716 | 1004 | 1115 | 1657 | 4040 | 5821 |  |
| 2018 | 0 | 587 | 656 | 864 | 3054 | 924 | 1172 | 746 | 971 | 1078 | 663 | 2704 | 4379 |  |
| 2019 | 0 | 143 | 1910 | 616 | 1101 | 3487 | 814 | 751 | 510 | 780 | 470 | 4660 | 4794 |  |
| 2020 | 0 | 15 | 117 | 8280 | 1710 | 2367 | 4087 | 696 | 520 | 305 | 594 | 1827 | 5991 |  |
| 2021 | 1 | 4 | 184 | 398 | 12117 | 1045 | 1398 | 2226 | 502 | 361 | 393 | 1641 | 6103 |  |
| 2022 | 0 | 681 | 1008 | 1251 | 1301 | 14135 | 914 | 1211 | 1734 | 477 | 433 | 1325 | 7143 |  |

Table 11. IESSNS baseline time series from 2016 to 2022. StoX abundance estimates of Norwegian springspawning herring (millions).

| Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ | TSB(1000 t) |  |
| 2016 | 41 | 146 | 752 | 604 | 1637 | 1559 | 2010 | 1614 | 1190 | 2023 | 2151 | 6467 | 6753 |  |
| 2017 | 1216 | 248 | 1285 | 4586 | 1056 | 1188 | 816 | 1794 | 1022 | 1131 | 1653 | 4119 | 5885 |  |
| 2018 | 0 | 577 | 722 | 879 | 3078 | 931 | 1264 | 734 | 948 | 1070 | 694 | 2792 | 4465 |  |
| 2019 | 0 | 153 | 1870 | 590 | 1067 | 3475 | 859 | 702 | 520 | 700 | 463 | 4808 | 4780 |  |
| 2020 | 0 | 7 | 111 | 8082 | 1697 | 2335 | 4102 | 714 | 491 | 294 | 590 | 1833 | 5930 |  |
| 2021 | 1 | 3 | 196 | 388 | 11988 | 1109 | 1342 | 2292 | 491 | 365 | 386 | 1649 | 6085 |  |
| 2022 | 0 | 724 | 984 | 1225 | 1339 | 14071 | 960 | 1172 | 1762 | 434 | 432 | 1329 | 7135 |  |



Lower right panels show the Coefficient of Correlation ( $r$ )
Figure 19. Internal consistency for Norwegian spring-spawning herring within the IESSNS 2022. The upper left part of the plots shows the relationship between log index-at-age within a cohort. Linear regression line shows the best fit to the log-transformed indices. The lower-right part of the plots shows the correlation coefficient ( r ) for the two ages plotted in that panel. The background colour of each panel is determined by the $r$ value, where red equates to $r=1$ and white to $r<0$.

### 4.5 Blue whiting

Blue whiting was distributed in parts of the survey area dominated by warm Atlantic waters and had a continuous distribution from the southern boundary of the survey area $\left(60^{\circ} \mathrm{N}\right)$ to Spitsbergen $\left(72{ }^{\circ} \mathrm{N}\right)$. High blue whiting density (sA-values) was observed in the southern part of the Norwegian Sea, along the Norwegian continental slope, around the Faroe Islands, and southeast of Iceland. Concentrations of older fish (age2+) were low, and they were mainly observed on the continental slopes, both in the eastern and the
southern part of the Norwegian Sea (Figure 20). The distribution in 2022 is comparable to the last two years with juvenile blue whiting recorded south and southwest of Iceland. As in previous years no blue whiting was registered in the cold East Icelandic Current, between Iceland and Jan Mayen.

The total biomass of blue whiting registered during IESSNS 2022 was 2.2 million tons (Table 12), which is about the same level as in 2021. Estimated stock abundance (ages 1+) was 27.5 billion compared to 26.2 billion in 2021. Age 1 and 2 respectively, dominated the estimate in 2022 as they contributed to $44 \%$ and $33 \%$ (abundance) and $30 \%$ and $33 \%$ (biomass), respectively.

Bootstrap estimates of numbers by age, with uncertainty estimates, for blue whiting during IESSNS 2022 are shown in Figure 21. The baseline point estimates from 2016-2022 are shown in Table 13. The internal consistency among year classes is shown in Figure 22 and indicates very good internal consistency for ages 3-5, and moderate to good fit for other ages.
The group considered the acoustic biomass estimate of blue whiting to be of good quality in the 2022 IESSNS as in the previous survey years.


Figure 20a. The $\mathrm{s}_{\mathrm{A}} /$ Nautical Area Scattering Coefficient (NASC) values of blue whiting along the cruise tracks in IESSNS 2022.


Figure 20b. The sA/Nautical Area Scattering Coefficient (NASC) values of blue whiting along the cruise tracks in IESSNS 2022. Presented as bar plot.

Table 12. Estimates of abundance, mean weight and mean length of blue whiting based on calculation in StoX (bootstrap) for IESSNS 2022.

| $\begin{aligned} & \text { Length } \\ & (\mathrm{cm}) \\ & \hline \end{aligned}$ | Age in years (year class) |  |  |  |  |  |  |  |  |  |  | Number$\left(10^{\wedge} 6\right)$ | Biomass <br> $\left(10^{\wedge} 6 \mathrm{~kg}\right)$ | Mean weight (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |  |  |  |  |
|  | 2022 | 2021 | 2020 | 2019 | 2018 | 2017 | 2016 | 2015 | 2014 | 2013 | 2012 |  |  |  |
| 10-11 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 11-12 | 135.2 |  |  |  |  |  |  |  |  |  |  | 135.2 | 1.1 | 8.2 |
| 12-13 | 414.1 |  |  |  |  |  |  |  |  |  |  | 414.1 | 4.7 | 11.3 |
| 13-14 | 236.6 |  |  |  |  |  |  |  |  |  |  | 236.6 | 3.5 | 14.9 |
| 14-15 | 169.0 |  |  |  |  |  |  |  |  |  |  | 169.0 | 2.9 | 17.1 |
| 15-16 |  |  |  |  |  |  |  |  |  |  |  |  | 0.2 | 22.0 |
| 16-17 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 17-18 |  |  |  |  |  |  |  |  |  |  |  |  | 0.4 | 30.0 |
| 18-19 |  | 152.9 |  |  |  |  |  |  |  |  |  | 152.9 | 6.2 | 37.2 |
| 19-20 |  | 1567.2 |  |  |  |  |  |  |  |  |  | 1567.2 | 68.3 | 44.1 |
| 20-21 |  | 4498.5 |  |  |  |  |  |  |  |  |  | 4498.5 | 225.8 | 50.8 |
| 21-22 |  | 4136.4 | 277.3 | 44.9 |  |  |  |  |  |  |  | 4458.5 | 251.9 | 57.1 |
| 22-23 |  | 1687.7 | 902.5 |  |  |  |  |  |  |  |  | 2590.2 | 166.9 | 64.0 |
| 23-24 |  | 484.9 | 2723.7 | 21.6 |  |  |  |  |  |  |  | 3230.2 | 244.4 | 76.6 |
| 24-25 |  | 84.2 | 2921.4 | 101.8 |  |  |  |  |  |  |  | 3107.4 | 263.9 | 85.7 |
| 25-26 |  | 5.9 | 1837.0 | 336.5 |  |  |  |  |  |  |  | 2179.4 | 207.8 | 95.5 |
| 26-27 |  | 4.0 | 729.4 | 396.6 | 19.4 | 6.8 |  |  |  |  |  | 1156.3 | 121.6 | 106.5 |
| 27-28 |  |  | 243.2 | 564.3 | 144.2 | 6.5 |  |  |  |  |  | 958.2 | 107.7 | 115.1 |
| 28-29 |  | 1.1 | 99.4 | 437.5 | 151.5 | 11.7 |  | 46.8 | 26.3 |  |  | 774.4 | 95.5 | 127.3 |
| 29-30 |  |  |  | 81.2 | 240.6 | 34.8 | 67.3 | 65.6 | 101.5 | 54.1 | 54.1 | 699.3 | 90.1 | 133.3 |
| 30-31 |  |  | 14.4 | 190.4 | 8.9 | 19.7 | 125.3 | 43.1 | 249.8 |  |  | 651.7 | 96.1 | 154.1 |
| 31-32 |  |  |  |  | 174.0 | 26.1 | 178.4 | 36.0 | 64.3 | 74.0 |  | 552.8 | 89.0 | 167.6 |
| 32-33 |  |  |  |  | 97.6 | 43.9 | 53.9 | 26.7 | 145.2 |  |  | 367.3 | 66.5 | 187.2 |
| 33-34 |  |  |  |  | 47.2 | 65.8 | 66.9 | 35.7 | 72.8 |  | 6.4 | 294.8 | 58.3 | 200.8 |
| 34-35 |  |  |  |  |  | 64.9 | 7.0 | 49.6 | 18.4 |  |  | 139.8 | 29.7 | 221.0 |
| 35-36 |  |  |  |  |  | 24.4 | 10.9 |  | 11.9 |  |  | 47.2 | 11.8 | 244.2 |
| 36-37 |  |  |  |  |  | 7.8 |  |  |  | 19.5 | 6.4 | 33.7 | 8.7 | 267.6 |
| 37-38 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 38-39 |  |  |  |  |  |  |  |  |  |  |  |  | 0.5 | 285.0 |
| 39-40 |  |  |  |  |  |  |  |  | 0.7 |  |  | 0.7 | 0.2 | 282.6 |
| TSN(mill) | 955 | 12623 | 9748 | 2175 | 883 | 313 | 510 | 303 | 691 | 148 | 67 | 28503.1 |  |  |
| cv (TSN) | 1.04 | 0.18 | 0.17 | 0.27 | 0.35 | 0.36 | 0.37 | 0.34 | 0.34 | 0.50 | 0.79 | 0.11 |  |  |
| TSB(1000 t) | 12.2 | 683.9 | 826.3 | 240.1 | 127.5 | 58.4 | 81.9 | 48.5 | 111.4 | 22.9 | 9.0 | 2223.7 |  |  |
| cv (TSB) | 1.04 | 0.18 | 0.17 | 0.27 | 0.36 | 0.38 | 0.37 | 0.35 | 0.34 | 0.46 | 0.71 | 0.12 |  |  |
| Mean length(cm) | 12.5 | 21.3 | 24.0 | 26.8 | 29.6 | 32.0 | 31.0 | 31.1 | 31.0 | 31.6 | 32.3 |  |  |  |
| Mean weight(g) | 13 | 60 | 87 | 114 | 152 | 190 | 167 | 173 | 167 | 168 | 180 |  |  |  |



Figure 21. Number by age with uncertainty for blue whiting during IESSNS 2022. Boxplot of abundance and relative standard error (CV) obtained by bootstrapping with 1000 replicates using the StoX software.

Table 13. IESSNS baseline time series from 2016 to 2022. StoX abundance estimates of blue whiting (millions).

| Age |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ | TSB(1000t) |
| 2016 | 3869 | 5609 | 11367 | 4373 | 2554 | 1132 | 323 | 178 | 177 | 8 | 233 | 2283 |
| 2017 | 23137 | 2558 | 5764 | 10303 | 2301 | 573 | 250 | 18 | 25 | 0 | 25 | 2704 |
| 2018 | 0 | 915 | 1165 | 3252 | 6350 | 3151 | 900 | 385 | 100 | 52 | 41 | 2039 |
| 2019 | 2153 | 640 | 1933 | 2179 | 4348 | 5434 | 1151 | 209 | 229 | 5 | 8 | 2028 |
| 2020 | 4066 | 5804 | 2996 | 1629 | 1205 | 1718 | 1990 | 939 | 201 | 21 | 30 | 1806 |
| 2021 | 4023 | 18056 | 2300 | 1664 | 841 | 982 | 1543 | 609 | 60 | 91 | 74 | 2238 |
| 2022 | 978 | 12454 | 9773 | 2279 | 904 | 314 | 520 | 303 | 678 | 177 | 71 | 2241 |



Figure 22. Internal consistency for blue whiting within the IESSNS. The upper left part of the plots shows the relationship between log index-at-age within a cohort. Linear regression line shows the best fit to the log-transformed indices. The lower-right part of the plots shows the correlation coefficient (r) for the two ages plotted in that panel. The background colour of each panel is determined by the r value, where red equates to $\mathrm{r}=1$ and white to $\mathrm{r}<0$.

### 4.6 Other species

## Lumpfish (Cyclopterus lumpus)

Lumpfish was caught in $71 \%$ of trawl stations across the five vessels (Figure 23) and where lumpfish was caught, $69 \%$ of the catches were $\leq 10 \mathrm{~kg}$. Lumpfish was distributed across the entire survey area, from east of Greenland to the Barents Sea in the northeast part of the covered area.

Abundance was greatest north of $71^{\circ} \mathrm{N}$, with lower densities in the central Norwegian Sea and mostly absent directly south of Iceland, and south and southwest of the North Sea. The zero line was not hit to the northeast, northwest and southwest of the survey so it is likely that the distribution of lumpfish extends beyond the survey coverage. The length of lumpfish caught varied from 5 to 51 cm with a bimodal distribution with the left peak ( $5-20 \mathrm{~cm}$ ) likely corresponding to 1 -group lumpfish and the right peak consisting of a mixture of age groups (Figure 24). For fish $\geq 20 \mathrm{~cm}$ in which sex was determined, the males exhibited a unimodal distribution with a peak around $25-27 \mathrm{~cm}$. The females also exhibited a bimodal distribution but with a peak around $24-30 \mathrm{~cm}$ and another around $35-45 \mathrm{~cm}$. Generally, the mean length and mean weight of the lumpfish was highest in Faroese waters, and around Iceland and along the shelf edges of Norway and lowest in the central and northern Norwegian Sea.
A total of 294 fish ( 67 by R/V "Árni Friðriksson", 83 by M/V "Eros", 96 by M/V Vendla and 48 by Tarajoq) between 5 and 52 cm were tagged during the survey (Figure 25).


Figure 23. Lumpfish catches at surface trawl stations during IESSNS 2022.


Figure 24. Length distribution of a) all lumpfish caught during the survey and b) length distribution of fish in which sex was determined.


Figure 25. Number tagged, and release location, of lumpfish. Insert shows the length distribution of the tagged fish.

## Salmon (Salmo salar)

A total of 60 North Atlantic salmon were caught in 38 stations both in coastal and offshore areas from $61^{\circ} \mathrm{N}$ to $76^{\circ} \mathrm{N}$ in the upper 30 m of the water column during IESSNS 2022 (Figure 26). The salmon ranged from 0.028 kg to 4.1 kg in weight, dominated by post-smolt and 1 sea-winter individuals. We caught from 1 to 6 salmon during individual surface trawl hauls. The length of the salmon ranged from 15 cm to 74 cm , with the highest fraction between 20 cm and 30 cm .


Figure 26. Catches of salmon at surface trawl stations during IESSNS 2022.

## Capelin (Mallotus villosus)

Capelin was caught in the surface trawl on 22 stations primarily along the cold fronts: Between East Greenland and Iceland, west and North of Jan Mayen and at the entrance to the Barents Sea (Figure 27). This is 10 stations more than in 2021 partly because of the lack of Greenland coverage in 2021 and partly because of more stations with capelin around Iceland this year (11 in 2022, 6 in 2021).


Figure 27. Presence of capelin in surface trawl stations during IESSNS 2022.

### 4.7 Marine Mammals

Opportunistic whale observations were done by $M / V$ "Eros" and $M / V$ "Vendla" from Norway in addition to R/V "Árni Friðriksson" from Iceland and R/V "Jákup Sverri" from Faroe Islands in from 1st July to 3rd August 2022 (Figure 28). Overall, 711 marine mammals of 11 different species were observed, which was a decrease from an overall 1029 marine mammals and eight species observed in 2021.

The species that were observed included fin whales (Balaenoptera physalus), minke whales (Balaenoptera acutorostrata), humpback whales (Megaptera novaeangliae), Northern bottlenose whales (Hyperoodon ampullatus), pilot whales (Globicephala sp.), killer whales (Orcinus orca), sperm whales (Physeter macrocephalus), sei whales (Baleanoptera borealis), white sided dolphins (Lagenorhynchus acutus) white beaked dolphins (Lagenorhynchus albirostris), harbour porpoise (Phocoena phocoena). A basking shark (Cetorhinus maximus) was also observed during the survey. The dominant number of marine mammal observations were found around Iceland, Faroe Islands and along the continental shelf between the north-eastern part of the Norwegian Sea and in a line between Finnmark to southwest of Svalbard. We observed very few marine mammals in the central part of the Norwegian Sea in July 2022. Fin whales ( $n=48$, group size $=1-12$ (average group size $=2.5)$ ) and humpback whales $(\mathrm{n}=44$, group size $=1-30$ (average group size $=3.9)$ ) dominated among the large whale species, and they were present west and northwest of Iceland and from Norwegian coast outside Finnmark stretching north/northwest via Bear Island to southwest of Svalbard. Very few sperm whales $(\mathrm{n}=8$, group size $=1$ (average group size $=1.0)$ ) where observed. Killer whales ( $\mathrm{n}=$ 121 , group size $=1-30$ (average groups size $=10.1$ )) dominated in the southern, northern and north-eastern part of the Norwegian Sea, partly overlapping and presumably feeding on NEA mackerel in the upper water masses. Pilot whales $(\mathrm{n}=30$, group size $=5-15$ (average groups size $=10$ )) where mostly observed in Faroese waters during IESSNS 2022. A sei whale and one northern bottlenose whale were observed in Icelandic waters, whereas a basking shark was observed in Faroese waters. White beaked dolphins ( $\mathrm{n}=229$,
group size $=1-22($ average groups size $=8.5))$ were present in the northern part of the Norwegian Sea. Two pods of white sided dolphins (group size $=15$ ) were observed in the southern part of the Norwegian Sea. Minke whales $(\mathrm{n}=53$, group size $=1-10$ (average group size $=1.7)$ ) were distributed over large areas from western coast of Norway to western part of Iceland, and from $60^{\circ} \mathrm{N}$ to $75^{\circ} \mathrm{N}$, including overlapping and likely feeding on NSS herring in the upper 40 m of the water column. There is available a new publication summarizing the main results on marine mammals from the IESSNS surveys from 2013 to 2018, with major focus on hot spot areas of fin whales and humpback whales from 2013 to 2018 (Løviknes et al. 2021)


Figure 28. Overview of all marine mammals sighted during IESSNS 2022.

| The group suggested the following recommendation from WGIPS | To whom |
| :--- | :---: |
| The occasional large catches of mackerel have a relatively large impact on the overall <br> results and possibly bias the stock indices. WGIPS recommends that the ability of the <br> present and alternative methods (such as more advanced statistical models) to <br> represent this overdispersion is evaluated, preferably at the WGISDAA meeting 25.- <br> 27.October, 2022. | National <br> istitutes and <br> WGISDAA |
| The surveys conducted by Denmark in 2018-2022 have clearly demonstrated that the <br> IESSNS methodology works also for the northern North Sea (i.e. north and west from | WGWIDE, RCG |
| Doggerbank) and the Skagerrak area deeper than 50 m. The survey provides essential <br> fishery-independent information on the stock during its feeding migration in summer <br> and WGIPS recommends that the Danish survey should continue as a regular annual <br> survey. | NANSEA |
| It is recommended that WGIPS contacts the country representatives for the IESSNS <br> survey to update the respective sections (e.g. trawl performance, trawl station data <br> collection) in the survey manual prior to the WGIPS meeting 23.-27.January 2023. | WGIPS |


| Action points | Responsible |
| :---: | :---: |
| Criteria and guidelines should be established for discarding substandard trawl stations using live monitoring of headline, footrope and trawl door vertical depth, and horizontal distance between trawl doors. For predetermined surface trawl station, discarded hauls should be repeated until performance is satisfactory. <br> Explicit guideline for incomplete trawl hauls is to repeat the station or exclude it from future analysis. It is not acceptable to visually estimate mackerel catch, it must be hauled onboard and weighed. If predetermined trawl hauls are not satisfactory according to criteria the station will be excluded from mackerel index calculations, i.e. treated as if it does not exist, but not as a zero mackerel catch station. | All |
| All survey participants are encouraged to continue the international tagging of lumpfish. | All |
| We encourage registrations of opportunistic marine mammal observations. | All |
| We should consider calculating the zooplankton index from annually gridded field polygons to extract area-mean time-series. WGINOR is currently working on Norwegian Sea polygons, and further work on this issue will start when their work is finalized. | All |
| In 2022 the IESSNS survey in the North Sea has been conducted for five consecutive years (2018-2022). It is recommended that a comprehensive report is written about the major results from the NEA mackerel time series from the IESSNS surveys in the North Sea, where the internal consistency between years in the survey for selected age groups is also evaluated. A major aim will be to at some stage evaluate and consider the possibility to include and implement the IESSNS survey in the North Sea as an abundance index used in ICES for NEA mackerel. | $\begin{aligned} & \text { DTU-Aqua } \\ & \text { (KW) } \end{aligned}$ |

## 7 Survey participants

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## 10 <br> Appendices

## Appendix 1

Denmark joined the IESSNS in 2018 for the first time extending the original survey area into the North Sea. The commercial fishing vessels "Ceton S205" was used. No problems applying the IESSNS methods were encountered. Area coverage, however, was restricted to the northern part of the North Sea at water depths larger 50 m . No plankton samples were taken, and no acoustic data were recorded because this is covered by the HERAS survey in June/July in this area.

In 2022, 34 stations were taken (PT and CTD. The locations of stations differed slightly from the previous year focussing on the area north and west of Doggerbank and extended into the eastern Skagerrak. However, due to shortage of available survey time only 34 out of the planned 38 stations were covered.

Average mackerel catch in 2022 amounted $1689 \mathrm{~kg} / \mathrm{km}^{2}$, which was considerably lower than in the previous year (2021: $2429 \mathrm{~kg} / \mathrm{km}^{2}$ ) but higher or similar than in the period 2018-2020 (2020: $1318 \mathrm{~kg} / \mathrm{km}^{2}$, 2019: 1009 $\mathrm{kg} / \mathrm{km}^{2}, 2018: 1743 \mathrm{~kg} / \mathrm{km}^{2}$ ). The length and age composition indicate a relative low amount of small $(<25 \mathrm{~cm})$ individuals whereas the abundance of older ( $\geq$ age 2 ) mackerel was on a similar level than in the previous year (Fig. A.1.).

StoX (version 3.5.0) estimate of mackerel biomass in the North Sea for 2022 is 471948 tonnes (Table A1-1) which is the second highest biomass values in the time series. The biomass and abundance estimates are based on a preliminary defined polygon for the surveyed area covered in all years since 2018 in which the northern border was set to $60^{\circ} \mathrm{N}$ (border to stratum 1; Fig. 2), and the eastern, southern, and western limits were either the coastline or extrapolated using half the longitudinal or latitudinal distance between the adjacent stations. The area of this polygon is $278525 \mathrm{~km}^{2}$.

For 11 out of 35 individuals in the size range of 18 to 20 cm the first wintering was not visible applying the standard age reading procedure. These fish should be attributed to the 2021-year class rather than be treated as 0-group fish considering the spawning period of mackerel in the North Sea. However, the aspect of the non-visible first age ring, which might be related to the presently prevailing warm winter conditions in the North Sea, warrants further investigations.

Based on the experiences made in the previous years, new limits for the stratum in the North were defined which shall be used for the station allocation for future surveys (Fig. A2). The northern limit for the North Sea and the Skagerrak were defined as $60^{\circ} \mathrm{N}$ and $59^{\circ} \mathrm{N}$, respectively. The western geographical limit in the North Sea was set to $1^{\circ} 30^{\prime} \mathrm{W}$ in the north and $2^{\circ} 30^{\prime} \mathrm{W}$ further south following the UK coastline where the Inner Moray Firth and the Firth of Forth were excluded because mackerel were not recorded there and a high abundance of 0-group gadoids, sandeel and other species makes a quantitative analysis of the catches very time consuming. The easter limit in the Skagerrak was set to $11^{\circ} \mathrm{E}$, and the southern limit in the North Sea was approximated by the 50 m isobath, which is about the shallowest depth limit for a safe setting of the Multpelt 832 trawl.

Table A1-1. StoX (version 3.5.0) baseline estimates of age segregated and length segregated mackerel indices for the North Sea in 2022.

| Length$(\mathrm{cm})$ | Age in years / Year class |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Number$\left(10^{\wedge} 6\right)$ | Biomass <br> (ton) | Mean weight <br> (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |  |  |  |
|  | 2022 | 2021 | 2020 | 2019 | 2018 | 2017 | 2016 | 2015 | 2014 | 2013 | 2012 | 2011 | 2010 | 2009 | 2008 |  |  |  |
| 17-18 |  | 0.1 |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.1 | 4 | 40 |
| 18-19 | 15.5 | 15.3 |  |  |  |  |  |  |  |  |  |  |  |  |  | 30.8 | 1488 | 48 |
| 19-20 | 36.3 | 87.1 |  |  |  |  |  |  |  |  |  |  |  |  |  | 123.4 | 6753 | 55 |
| 20-21 | 1.8 | 120.4 |  |  |  |  |  |  |  |  |  |  |  |  |  | 122.1 | 8024 | 66 |
| 21-22 |  | 42.0 |  |  |  |  |  |  |  |  |  |  |  |  |  | 42.0 | 3162 | 75 |
| 22-23 |  | 12.6 |  |  |  |  |  |  |  |  |  |  |  |  |  | 12.6 | 1153 | 92 |
| 23-24 |  | 11.3 |  |  |  |  |  |  |  |  |  |  |  |  |  | 11.3 | 1237 | 109 |
| 24-25 |  | 26.7 |  |  |  |  |  |  |  |  |  |  |  |  |  | 26.7 | 3318 | 124 |
| 25-26 |  | 12.6 |  |  |  |  |  |  |  |  |  |  |  |  |  | 12.6 | 1747 | 139 |
| 26-27 |  | 7.4 |  |  |  |  |  |  |  |  |  |  |  |  |  | 7.4 | 1161 | 157 |
| 27-28 |  | 15.3 |  |  |  |  |  |  |  | 0.8 |  |  |  |  |  | 16.1 | 3013 | 187 |
| 28-29 |  | 147.9 | 23.2 |  |  |  |  |  |  |  |  |  |  |  |  | 171.1 | 36138 | 211 |
| 29-30 |  | 496.5 | 23.2 |  |  |  |  |  |  |  |  |  |  |  |  | 519.7 | 126715 | 244 |
| 30-31 |  | 204.9 | 160.3 |  |  |  |  |  |  |  |  |  |  |  |  | 365.2 | 97338 | 266 |
| 31-32 |  | 26.2 | 134.1 | 13.3 |  |  |  |  |  |  |  |  |  |  |  | 173.6 | 49252 | 284 |
| 32-33 |  |  | 103.7 | 13.1 | 0.6 |  |  |  |  |  |  |  |  |  |  | 117.4 | 36622 | 312 |
| 33-34 |  |  | 35.2 | 30.1 | 5.4 | 0.6 |  |  |  |  |  |  |  |  |  | 71.3 | 23661 | 332 |
| 34-35 |  |  | 3.6 | 29.6 | 18.9 | 2.3 |  |  |  |  |  |  |  |  |  | 54.3 | 19943 | 367 |
| 35-36 |  |  |  | 5.7 | 13.5 | 7.6 | 6.6 | 4.4 |  |  |  |  |  |  |  | 37.8 | 14858 | 393 |
| 36-37 |  |  |  | 0.7 | 8.9 | 11.3 | 7.1 | 0.2 | 0.5 | 0.8 |  |  |  |  |  | 29.5 | 12106 | 410 |
| 37-38 |  |  |  |  | 1.5 | 6.3 | 9.4 | 3.9 | 0.1 |  |  |  |  |  |  | 21.1 | 9138 | 433 |
| 38-39 |  |  |  |  |  | 1.2 | 0.7 | 4.1 | 2.4 | 0.5 |  |  |  |  |  | 8.9 | 4416 | 498 |
| 39-40 |  |  |  |  | 1.1 | 4.2 | 2.5 | 0.7 | 0.9 | 0.5 |  |  |  |  |  | 9.8 | 4963 | 504 |
| 40-41 |  |  |  |  |  | 1.1 | 0.8 | 1.3 | 0.5 | 0.3 | 0.7 |  |  |  |  | 4.6 | 2537 | 549 |
| 41-42 |  |  |  |  |  |  |  | 1.1 |  | 0.1 |  |  |  |  |  | 1.3 | 699 | 542 |
| 42-43 |  |  |  |  |  |  |  | 0.4 |  |  |  |  | 0.4 |  | 0.1 | 1.0 | 648 | 675 |
| 43-44 |  |  |  |  |  |  |  |  |  | 1.8 |  |  |  |  |  | 1.8 | 1250 | 682 |
| 44-45 |  |  |  |  |  |  | 1.3 |  |  |  |  |  |  |  |  | 1.3 | 1281 | 950 |
| TSN (mill) | 53.6 | 1226.4 | 483.3 | 92.4 | 49.8 | 34.4 | 28.5 | 16.3 | 4.3 | 4.8 | 0.7 | 0.0 | 0.4 | 0.0 | 0.1 | 1,995 | 472626 |  |
| TSB (ton) | 2913 | 242385 | 136351 | 30981 | 19206 | 14533 | 13103 | 7731 | 2195 | 2535 | 345 | 0 | 259 | 0 | 90 | 472,626 |  |  |
| Mean length (cm) | 18.7 | 26.8 | 30.8 | 33.0 | 34.7 | 36.3 | 36.9 | 37.4 | 38.2 | 38.1 | 40.0 |  | 42.0 |  | 42.0 |  |  |  |
| Mean weight (g) | 54 | 198 | 282 | 335 | 385 | 422 | 460 | 474 | 511 | 525 | 525 |  | 638 |  | 746 |  |  |  |



Fig. A1-1. Comparison of length and age distribution of mackerel in the North Sea 2018 to 2022.


Fig. A1-2. Limits of the North Sea stratum for future surveys and sampling positions achieved in the period 2018-2022.

## Appendix 2:

The mackerel index is calculated on all valid surface stations. That means, that invalid and potential extra surface stations and deeper stations need to be excluded. Below is the exclusion list used when calculating the mackerel abundance index for IESSNS 2022 (Table A2-1). Map of included and excluded trawl stations displayed in Figure A2-1.

Table A2-1: Trawl station exclusion list and average horizontal trawl opening per vessel for IESSNS 2022 for calculating the mackerel abundance index.

| Vessel | Country | Horizontal trawl <br> opening $(\mathrm{m})$ | Exclusion list <br> Cruise | Stations |
| :--- | :--- | :--- | :--- | :--- |
| Vendla | Norway | 67.5 | 2022816 | $60,75,80,82,85,88,90,91$, |
| Eros | Norway | 63.5 | 2022817 | $95,104,109,113,120,124$ |
|  |  |  | $28,30,44,46,51,55,59,63$, |  |
| R/V Árni Friðriksson | Iceland | 63.75 | A8-2022 | $72,73,91$ |
| R/V Jákup Sverre | Faro Islands | 63.4 | 2230 | $5,23,24,35,46,61^{*}$ |
| R/V Tarajoq | Greenland | 61.4 | TA-2022-04 | none |
| Ceton | Denmark | 72.0 | IESSNS2022 | none |

* Observe that in PGNAPES and the national database station numbers are 4-digit numbers preceded by 2230 (e.g. '22300005')


Figure A2-1. IESSNS 2022. Surface trawl stations included (filled dark blue rectangle) and excluded (filled light blue rectangle) in calculations of mackerel age segregated index used in the assessment. Strata boundary also displayed (grey solid lines).

## Appendix 3: Impact of large hauls on abundance and biomass estimates

In 2022 there were two large mackerel hauls. In order to investigate the effect of these on the StoX estimates, an additional run of Sto $X$ was made without these hauls (Figure A3-1).
If the two stations with the highest catches (slightly above 20 tons on each) are removed, the baseline estimate of total abundance is reduced by $34 \%$ and the baseline estimate of total biomass is reduced by $33 \%$ (from 7.37 to 4.91 million tons). Moreover, the relative standard error of total abundance from 1000 bootstrap replicates is $26 \%$ when all stations are used, while becomes reduced to $12 \%$ when the two highest stations are removed. The relative standard error of total biomass from 1000 bootstrap replicates is $25 \%$ when all stations are used, while becomes reduced to $11 \%$ when the two highest stations are removed.


Figure A3-1. StoX runs with (black/red 2022 dot) and without (blue 2022 dot) large hauls. Biomass (left panel) and abundance (right panel).

## Appendix 4:

Horizontal trawl opening of the Multpelt 832 trawl is a function of trawl door spread and tow speed (Table 6 in the 2022 report). The estimates in table 6 are originally based on flume tank simulations in 2013 (Hirtshals, Denmark) where two formulas were empirically derived for two towing speeds, 4.5 and 5 knots:

Towing speed 4.5 knots: $\quad$ Horizontal opening $(m)=0.441 *$ Door spread $(m)+13.094$
Towing speed 5.0 knots: Horizontal opening $(m)=0.3959$ * Door spread (m) +20.094
In 2017, the towing speed range was increased to 5.2 knots, i.e. an extrapolation of the trawl opening as a function of door spread and speed was performed. In 2022 the towing speed range was further extended down to 4.3 knots and up to 5.5 knots, using a kriging gridding method, see figure A4-1.


Figure A3-1. Table 6 in the report shown as a plot.

Working document 02, WGWIDE 2022

## PFA self-sampling report for WGWIDE 2022

Martin Pastoors, 24/08/2022 12:41:15 (v2)
PFA report 2022_07

## Executive summary

The Pelagic Freezer-trawler Association (PFA) is an association that has nine member companies that together operate 15 (in 2021) freezer trawlers in six European countries (www.pelagicfish.eu). In 2015, the PFA has initiated a self-sampling program that expands the ongoing monitoring programs on board of pelagic freezer-trawlers aimed at assessing the quality of fish. The expansion in the selfsampling program consists of recording of haul information, recording the species compositions by haul and regularly taking length measurements from the catch. The self-sampling is carried out by the vessel quality managers on board of the vessels, who have a long experience in assessing the quality of fish, and by the skippers/officers with respect to the haul information. The scientific coordination of the self-sampling program is carried out by Martin Pastoors (PFA chief science officer) with support of Floor Quirijns (contractor). The self-sampling program has been incrementally implemented in the fishery and by 2018 all vessels in the PFA fleet participated in the self-sampling.

This report for WGWIDE presents an overview of the results of the Pelagic Freezer-Trawler Association (PFA) self-sampling program for the fisheries for widely-distributed pelagic stocks: Northeast Atlantic mackerel, Blue whiting, Horse mackerel and Atlanto-scandian herring (herring caught north of 62 degrees). The selection of hauls to be included in the analyses was based on first summing all catches by vessel, trip, species and week. For each vessel-trip-species-week combination, the proportion of the species in the catch were calculated. The following filter criteria have applied to the weekly data:

- for horse mackerel: latitude >45, proportion in the catch > 10\%, weekly catch > 10 tonnes
- for mackerel : latitude $>45$, proportion in the catch $>10 \%$, weekly catch $>10$ tonnes
- for blue whiting : latitude >50, proportion in the catch > $10 \%$, weekly catch $>10$ tonnes
- for herring : division = 27.2.a, proportion in the catch > 10\%, weekly catch > 10 tonnes

Trips from 2016 up to 11/08/2022 have been processed for this overview. Pelagic fisheries within the Pelagic Freezer-trawler Association are carried out by vessels from different countries. Overall, around $48 \%$ of the catch volume of trips in this overview were taken by Dutch trawlers, 22\% German trawlers, $14 \%$ UK trawlers and $16 \%$ other countries. Blue whiting constitutes the majority of the catch in those trips (54\%), followed by mackerel (23\%) and horse mackerel (12\%). Atlanto-scandian herring only constitutes around $3 \%$ of the volume in the PFA widely distributed fishery. Note that the North Sea herring fishery is not included in this overview.

The Mackerel fishery takes place from October through to March of the subsequent year. Bycatches of mackerel may also occur during other fisheries,e.g. for horse mackerel or herring. Overall, the selfsampling activities for the mackerel fisheries during the years 2016-2022 (up to 11/08/2022) covered 465 fishing trips with 6352 hauls, a total catch of 386474 tonnes and 103745 individual length measurements. The main fishing areas are ICES division 27.4.a and division 27.6.a. Compared to the previous years, mackerel in the catch in 2021 have been relatively large with a median length of 36.4 cm compared to 33.6-36.2 in the preceding years. Also, the median weight has been somewhat higher with median weight of 435 gram compared to 385-422 gram in the preceding years.

The Western horse mackerel fishery takes place from October through to March of the subsequent year. Overall, the self-sampling activities for the Western horse mackerel fisheries during the years 2016-2022 (up to 11/08/2022) covered 250 fishing trips with 3316 hauls, a total catch of 128553 tonnes and 130146 individual length measurements. The main fishing areas are ICES division 27.6.a, division 27.7.b and division 27.7.j. Western horse mackerel have a wide range in the length distributions in the catch. Median lengths in divisions 27.6.a, 27.7.b and 27.7.j have fluctuated between 25.2 and 31.9 cm (with one low median length of 22.7 cm in 27.6.a in 2018). In ICES division 27.7.h, median lengths in the catch have been smaller and fluctuated between 20.7 and 24.5 cm .

The North Sea horse mackerel fishery takes place from October through to January of the subsequent year. Overall, the self-sampling activities for the North Sea horse mackerel fisheries during the years 2016-2022 (up to 11/08/2022) covered 109 fishing trips with 900 hauls, a total catch of 46322 tonnes and 38983 individual length measurements. The main fishing areas is ICES division 27.7.d with some minor catches in 27.4.c. Catches in division 27.4.a have been counted as Western Horse mackerel. North Sea horse mackerel have a narrow range in the length distributions in the catch. Median lengths in division 27.7.d have fluctuated between 20.7 and 24.3 cm .

The blue whiting fishery takes place from February through to May although some minor fisheries for blue whiting may remain over the other months. Overall, the self-sampling activities for the horse mackerel fisheries during the years 2016-2022 (up to 11/08/2022) covered 320 fishing trips with 8234 hauls, a total catch of 810714 tonnes and 466229 individual length measurements. The main fishing areas are ICES division 27.6.a, division 27.7.c and division 27.7.k. Compared to the previous years, blue whiting in the catches during 2020-2022 have been relatively large with a median length of 27.8 cm compared to 24.1-24.5 in the preceding years.

The fishery for Atlanto-scandian herring (ASH) is a relatively smaller fishery for PFA and takes place mostly in October. Overall, the self-sampling activities for the horse mackerel fisheries during the years 2016-2022 (up to 11/08/2022) covered 32 fishing trips with 297 hauls, a total catch of 17705 tonnes and 5147 individual length measurements. Only the herring fishery in ICES division 27.2.a is considered for ASH. Note that there are herring catches in other divisions within the selected trips. These are trips where North Sea herring has been fished with some bycatches of mackerel for
example. Atlanto-scandian herring have a relatively narrow range in the length distributions in the catch. Median lengths have been between 30 and 35 cm .

In this 2022 self-sampling report, a standardized CPUE calculation has been included for the first time for most of the stocks. The standardized CPUE is based on a GLM model with a negative binomial distribution. The response variable is the catch by week and vessel, with an offset of the log effort (number of fishing days per week) and explanatory variables year, GT category, month, division and depth category. An assumed technical efficiency increase of $2.5 \%$ per year has been included in the fitting of the model (Rousseau et al 2019)

## 1 Introduction

The Pelagic Freezer-trawler Association (PFA) is an association that has nine member companies that together operate 18 freezer trawlers (in 2022) in six European countries (www.pelagicfish.eu). In 2015, the PFA has initiated a self-sampling program that expands the ongoing monitoring programs on board of pelagic freezer-trawlers by the specialized crew of the vessels. The primary objective of that monitoring program is to assess the quality of fish. The expansion in the self-sampling program consists of recording of haul information, recording the species compositions per haul and regularly taking random length-samples from the catch. The self-sampling is carried out by the vessel quality managers on board of the vessels, who have a long experience in assessing the quality of fish, and by the skippers/officers with respect to the haul information. The scientific coordination of the selfsampling program is carried out by Martin Pastoors (PFA chief science officer) with support of Floor Quirijns (contractor).

## 2 Overview of self-sampling methodology

The PFA self-sampling program has been implemented incrementally on many vessels that belong to the members of the PFA. The self-sampling program is designed in such a way that it follows as closely as possible the working practices on board of the different vessels and that it delivers relevant information for documenting the performance of the fishery and to assist stock assessments of the stocks involved. The following main elements can be distinguished in the self-sampling protocol:

- haul information (date, time, position, weather conditions, environmental conditions, gear attributed, estimated catch, optionally: species composition)
- batch information (total catch per batch=production unit, including variables like species, average size, average weight, fat content, gonads $\mathrm{y} / \mathrm{n}$ and stomach fill)
- linking batch and haul information (essentially a key of how much of a batch is caught in which of the hauls)
- length information (length frequency measurements, either by batch or by haul)

The self-sampling information is collected using standardized Excel worksheets. Each participating vessel will send in the information collected during a trip by the end of the trip. The data will be checked and added to the database by Floor Quirijns and/or Martin Pastoors, who will also generate standardized trip reports (using RMarkdown) which will be sent back to the vessel within one or two days. The compiled data for all vessels is being used for specific purposes, e.g. reporting to expert groups, addressing specific fishery or biological questions and supporting detailed biological studies. The PFA publishes an annual report on the self-sampling program.

A major feature of the PFA self-sampling program is that it is tuned to the capacity of the vessel-crew to collect certain kinds of data. Depending on the number of crew and the space available on the vessel, certain types of measurements can or cannot be carried out. That is why the program is essentially tuned to each vessel separately. And that is also the reason that the totals presented in this report can be somewhat different dependent on which variable is used. For example the estimate of total catch is different from the sum of the catch per species because not all vessels have supplied data on the species composition of the catch.

In order to supply relevant information to WGWIDE, the PFA self-sampling data has been filtered using the following approach. First, all catches per vessel, trip and species have been summed by week. For each vessel-trip-species-week combination, the proportion of the species in the catch were calculated. Then the following filter criteria have applied to the weekly data:

- for horse mackerel: latitude $>45$, proportion in the catch $>10 \%$, catch $>10$ tonnes
- for mackerel : latitude > 45, proportion in the catch > 10\%, catch > 10 tonnes
- for blue whiting : latitude >50, proportion in the catch > 10\%, catch > 10 tonnes
- for herring : division $=$ 27.2.a, proportion in the catch $>10 \%$, catch $>10$ tonnes For this report, data have been processed for 2016-2022 (up to 11/08/2022).


## 3 Results

### 3.1 General

An overview of all the self-sampled trips for mac, hom, whb, her_ash in 27.2.a, 27.4.a, 27.6.a, 27.7.b, 27.7.j, 27.7.h, 27.4.c, 27.7.d, 27.7.c, 27.7.k, 27.5.b. The percentage non-target species is defined as the catch of non-pelagic species relative to the catch of pelagic species.

| year | nvessels | ntrips | ndays | nhauls | catch | catch/day | nontarget | nlength | nbio |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  |  |
| 2016 | 9 | 45 | 591 | 1,307 | 113,900 | 193 | $0.50 \%$ | 65,212 | 0 |  |
| 2017 | 12 | 62 | 840 | 1,781 | 177,887 | 212 | $0.26 \%$ | 91,357 | 0 |  |
| 2018 | 16 | 86 | 1,219 | 2,677 | 253,237 | 208 | $0.22 \%$ | 170,306 | 641 |  |
| 2019 | 16 | 97 | 1,226 | 2,658 | 224,886 | 183 | $0.29 \%$ | 124,288 | 1,055 |  |
| 2020 | 17 | 112 | 1,424 | 3,038 | 305,282 | 214 | $0.36 \%$ | 163,955 | 2,379 |  |
| 2021 | 19 | 119 | 1,398 | 2,874 | 282,097 | 202 | $0.52 \%$ | 138,481 | 1,411 |  |
| $2022^{*}$ | 18 | 62 | 733 | 1,694 | 144,718 | 197 | $0.84 \%$ | 65,457 | 4,004 |  |
| $($ all) |  | 583 | 7,431 | 16,029 | $1,502,007$ |  |  | 819,056 | 9,490 |  |

Table 3.1.1: PFA fisheries for widely distributed species Self-sampling Summary of number of vessels, trips, days, hauls, catch (tonnes), catch per day and number of fish measured. * denotes incomplete year

## Catch and number of self-sampled hauls by year and division



Table 3.1.2: PFA fisheries for widely distributed species Self-sampling Summary of catch (top) and number of hauls (bottom) per year and division. * denotes incomplete year

## Catch and number of self-sampled hauls by year and month

| month | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022* | all | perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Jan | 12,789 | 28,644 | 25,647 | 35,499 | 37,485 | 51,537 | 41,028 | 232,629 | 15.5\% |
| Feb | 10,196 | 19,369 | 32,600 | 32,829 | 28,300 | 31,967 | 28,025 | 183,285 | 12.2\% |
| Mar | 16,154 | 29,388 | 32,673 | 27,992 | 47,769 | 36,936 | 40,093 | 231,004 | 15.4\% |
| Apr | 14,420 | 28,510 | 58,665 | 28,857 | 66,042 | 29,472 | 25,878 | 251,844 | 16.8\% |
| May | 7,763 | 12,367 | 30,227 | 21,332 | 29,189 | 14,466 | 8,521 | 123,866 | 8.2\% |
| Jun | 1,649 | 0 | 6,866 | 1,498 | 4,219 | 2,467 | 0 | 16,699 | 1.1\% |
| Jul | 1,977 | 665 | 791 | 6,185 | 1,566 | 12,330 | 1,174 | 24,688 | 1.6\% |
| Aug | 886 | 6,545 | 4,551 | 3,844 | 4,234 | 4,779 | 0 | 24,839 | 1.7\% |
| Sep | 1,990 | 9,898 | 8,334 | 7,775 | 12,586 | 9,134 | 0 | 49,717 | 3.3\% |
| Oct | 18,517 | 17,478 | 22,975 | 25,417 | 27,648 | 39,924 | 0 | 151,960 | 10.1\% |
| Nov | 18,307 | 21,875 | 20,385 | 22,205 | 27,061 | 30,033 | 0 | 139,865 | 9.3\% |
| Dec | 9,251 | 3,148 | 9,522 | 11,453 | 19,184 | 19,052 | 0 | 71,610 | 4.8\% |
| (all) | 113,900 | 177,887 | 253,237 | 224,886 | 305,282 | 282,097 | 144,718 | 1,502,007 | 100.0\% |

month 201620172018 2019 2020 2021 202 * perc

| Jan | 174 | 311 | 309 | 452 | 355 | 568 | 482 | 2,651 | $16.5 \%$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Feb | 142 | 206 | 325 | 362 | 287 | 344 | 301 | 1,967 | $12.3 \%$ |
| Mar | 160 | 226 | 297 | 314 | 410 | 333 | 389 | 2,129 | $13.3 \%$ |
| Apr | 114 | 201 | 494 | 289 | 574 | 240 | 359 | 2,271 | $14.2 \%$ |
| May | 105 | 145 | 372 | 250 | 312 | 167 | 144 | 1,495 | $9.3 \%$ |
| Jun | 14 | 0 | 77 | 23 | 97 | 42 | 0 | 253 | $1.6 \%$ |
| Jul | 25 | 12 | 10 | 75 | 26 | 113 | 19 | 280 | $1.7 \%$ |
| Aug | 5 | 58 | 39 | 41 | 53 | 33 | 0 | 229 | $1.4 \%$ |
| Sep | 38 | 130 | 145 | 149 | 154 | 187 | 0 | 803 | $5.0 \%$ |
| Oct | 204 | 198 | 232 | 299 | 295 | 398 | 0 | 1,626 | $10.1 \%$ |
| Nov | 223 | 269 | 291 | 315 | 331 | 305 | 0 | 1,734 | $10.8 \%$ |
| Dec | 103 | 25 | 86 | 89 | 144 | 156 | 0 | 603 | $3.8 \%$ |
| (all) | 1,307 | 1,781 | 2,677 | 2,658 | 3,038 | 2,886 | 1,694 | 16,041 | $100.0 \%$ |

Table 3.1.3: PFA fisheries for widely distributed species Self-sampling summary of catch (top) and number of hauls (bottom) per year and month.

## Catch and number of self-sampled hauls by year and country (flag)

| flag | 2016 | 2017 |  | 2018 | 2019 | 2020 | 2021 | 2022* | * all | perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DEU | 27,803 | 27,500 |  | 55,468 | 40,385 | 69,108 | 54,075 | 5 26,246 | 6300,585 | 20.0\% |
| FR | 0 |  | 0 | 11,936 | 19,356 | 14,506 | 12,257 | 7 9,128 | 8 67,184 | 4.5\% |
| LIT | 0 |  | 0 | 0 | 1,414 | 13,744 | 23,150 | 0 6,467 | 7 44,775 | 3.0\% |
| NL | 68,790 | 114,844 |  | 139,403 | 106,898 | 117,284 | 124,171 | 1 69,345 | $5 \quad 740,736$ | $49.3 \%$ |
| POL | 0 |  | 0 | 15,966 | 28,022 | 54,615 | 29,675 | 5 13,599 | 9141,877 | 9.4\% |
| UK | 17,306 | 35,543 |  | 30,464 | 28,811 | 36,026 | 35,341 | 1 19,932 | 2 203,423 | 13.5\% |
| NA | 0 |  | 0 | 0 | 0 | 0 | 3,428 |  | $0 \quad 3,428$ | 0.2\% |
| (a11) | 113,900 | 177,887 |  | 253,237 | 224,886 | 305,282 | 282,097 | 7144,718 | 8 1,502,007 | 100.0\% |
| flag | 2016 | 2017 | 2018 | 82019 | 2020 | 2021 | 2022* | all | perc |  |
| DEU | 340 | 276 | 637 | 7456 | 623 | 463 | 269 | 3,064 1 | 19.1\% |  |
| FR | 0 | 0 | 236 | 6 357 | 243 | 205 | 165 | 1,206 | 7.5\% |  |
| LIT | 0 | 0 | 0 | 034 | 142 | 165 | 36 | 377 | 2.4\% |  |
| NL | 807 | 1,177 1, | 1,403 | 3 1,314 | 1,374 | 1,385 | 886 | 8,346 5 | 52.1\% |  |
| POL | 0 | 0 | 111 | 183 | 322 | 187 | 113 | 916 | 5.7\% |  |
| UK | 160 | 328 | 290 | - 314 | 334 | 394 | 225 | 2,045 12 | 12.8\% |  |
| NA | 0 | 0 | 0 | 00 | 0 | 75 | 0 | 75 | 0.5\% |  |
| (all) | 1,307 | 1,781 2, | 2,677 | 7 2,658 | 3,038 | 2,874 | 1,694 1 | 16,029 10 | 00.0\% |  |

Table 3.1.4: PFA fisheries for widely distributed species Self-sampling summary of catch (top) and number of hauls (bottom) per year and month.

## Catch by species and year

| species <br> perc | english_name | scientific_name | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | all |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| whb | blue whiting | Micromesistius poutassou | 48,666 | 79,108 | 154,733 | 113,262 | 174,647 | 149,325 | 90,974 | 810,715 |
| 54.0\% |  |  |  |  |  |  |  |  |  |  |
| mac | mackerel | Scomber scombrus | 33,544 | 63,026 | 55,756 | 54,005 | 84,290 | 69,094 | 26,569 | 386,283 |
| 25.7\% |  |  |  |  |  |  |  |  |  |  |
| hom | horse mackerel | Trachurus trachurus | 21,808 | 20,853 | 28,497 | 31,565 | 25,061 | 33,995 | 13,096 | 174,876 |
| 11.6\% |  |  |  |  |  |  |  |  |  |  |
| her | herring | Clupea harengus | 4,509 | 6,870 | 7,851 | 17,286 | 9,154 | 19,912 | 3,123 | 68,704 |
| 4.6\% |  |  |  |  |  |  |  |  |  |  |
| arg | argentines | Argentina spp | 1,560 | 2,596 | 4,097 | 4,566 | 7,036 | 5,457 | 9,595 | 34,906 |
| 2.3\% |  |  |  |  |  |  |  |  |  |  |
| her_ash | NA | NA | 2,109 | 4,913 | 1,367 | 3,373 | 3,563 | 2,379 | 0 | 17,706 |
| 1.2\% |  |  |  |  |  |  |  |  |  |  |
| boc | boarfish | Capros aper | 226 | 245 | 153 | 288 | 603 | 844 | 680 | 3,039 |
| 0.2\% |  |  |  |  |  |  |  |  |  |  |
| pil | pilchard | Sardina pilchardus | 719 | 61 | 371 | 155 | 32 | 325 | 140 | 1,805 |
| 0.1\% |  |  |  |  |  |  |  |  |  |  |
| hke | hake | Merluccius merluccius | 266 | 107 | 270 | 197 | 181 | 239 | 333 | 1,593 |
| 0.1\% |  |  |  |  |  |  |  |  |  |  |
| spr | sprat | Sprattus sprattus | 382 | 0 | 0 | 0 | 415 | 138 | 0 | 934 |
| 0.1\% |  |  |  |  |  |  |  |  |  |  |
| sqr | squid | Loligo vulgaris | 0 | 0 | 8 | 8 | 26 | 133 | 55 | 229 |
| 0.0\% |  |  |  |  |  |  |  |  |  |  |
| had | haddock | Melanogrammus aeglefinus | 11 | 5 | 15 | 46 | 42 | 66 | 37 | 222 |
| 0.0\% |  |  |  |  |  |  |  |  |  |  |
| brb | black seabream | Spondyliosoma cantharus | 29 | 2 | 22 | 3 | 83 | 5 | 3 | 148 |
| 0.0\% |  |  |  |  |  |  |  |  |  |  |
| bor | boarfish | Caproidae | 0 | 0 | 0 | 0 | 0 | 59 | 73 | 132 |
| 0.0\% |  |  |  |  |  |  |  |  |  |  |
| whg | whiting | Merlangius merlangus | 13 | 0 | 24 | 31 | 31 | 30 | 2 | 130 |
| 0.0\% |  |  |  |  |  |  |  |  |  |  |
| oth | NA | NA | 57 | 101 | 74 | 102 | 119 | 95 | 37 | 585 |
| 0.0\% |  |  |  |  |  |  |  |  |  |  |
| (all) | (all) | (all) | 113,900 | 177,887 | 253,237 | 224,886 | 305,282 | 282,097 | 144,718 | 1,502,007 |
| 100.0\% |  |  |  |  |  |  |  |  |  |  |

Table 3.1.5: PFA fisheries for widely distributed species Self-sampling Summary of total catch (tonnes) by species. OTH refers to all other species that are not the main target species

## Haul positions

An overview of all self-sampled hauls in the PFA fisheries for widely distributed species.


Figure 3.1.1: PFA fisheries for widely distributed species Self-sampling haul positions. $N$ indicates the number of hauls.

## Catches for the main target species

Summed catches (tonnes) of the main target species aggregated in rectangles.


Figure 3.1.2: PFA fisheries for widely distributed species Self-sampling catch per species and per rectangle. $N$ indicates the number of hauls. Catch refers to the total catch per year.

## Catch rates (catch/day) for the main target species



Figure 3.1.3: Average catch per day, per species and per rectangle. $N$ indicates the number of hauls; avg refers to the average catch per day.

## Average surface temperature by quarter and by rectangle.



Figure 3.1.4: PFA fisheries for widely distributed species Average surface temperature (C) by year and quarter. $N$ indicates the number of hauls. Avg refers to the average temperature.

## Average fishing depth.



Figure 3.1.5: PFA fisheries for widely distributed species Average fishing depth ( $m$ ) by year and quarter. N indicates the number of hauls. Avg refers to the average fishing depth.

## Average wind force.



Figure 3.1.6: PFA fisheries for widely distributed species Average windforce (Bft) by year and quarter. $N$ indicates the number of hauls. Avg refers to the average windforce.

### 3.2 Northeast Atlantic mackerel (MAC, Scomber scombrus)

Northeast Atlantic mackerel self-sampling summary.

| species | year | nvessels | ntrips | ndays | nhauls | catch | catch/day | nlength | nbio |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| mac | 2016 | 9 | 30 | 213 | 395 | 32,894 | 154 | 6,964 | 0 |
| mac | 2017 | 11 | 48 | 386 | 690 | 62,715 | 162 | 11,614 | 0 |
| mac | 2018 | 16 | 56 | 501 | 841 | 55,186 | 110 | 13,700 | 32 |
| mac | 2019 | 15 | 72 | 615 | 1,105 | 53,525 | 87 | 17,894 | 476 |
| mac | 2020 | 17 | 84 | 712 | 1,258 | 83,876 | 118 | 31,381 | 646 |
| mac | 2021 | 18 | 78 | 606 | 1,054 | 68,466 | 113 | 11,294 | 684 |
| mac | 2022 | 14 | 40 | 296 | 538 | 26,515 | 90 | 6,591 | 3,733 |
| $($ all | (all) |  | 408 | 3,329 | 5,881 | 383,176 |  | 99,438 | 5,571 |

Table 3.2.1: Northeast Atlantic mackerel. Self-sampling summary with the number of days, hauls, trips, vessels, catch (tonnes), catch rate (ton/day), number of fish measured, number of biological observations.

Northeast Atlantic mackerel. Catch by division

| species | division | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | $2022^{*}$ | all | perc |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| mac | $27.2 . a$ | 7,381 | 12,967 | 4,803 | 204 | 706 | 9 | 0 | 26,069 | $6.8 \%$ |
| mac | $27.4 . a$ | 15,291 | 17,325 | 28,511 | 24,293 | 50,545 | 44,514 | 11,715 | 192,194 | $50.2 \%$ |
| mac | $27.6 . a$ | 8,678 | 28,288 | 18,071 | 21,298 | 15,847 | 21,989 | 9,854 | 124,025 | $32.4 \%$ |
| mac | $27.7 . \mathrm{b}$ | 186 | 3,640 | 1,111 | 5,386 | 6,044 | 1,094 | 4,539 | 21,999 | $5.7 \%$ |
| mac | $27.7 . j$ | 1,359 | 496 | 2,689 | 2,345 | 10,734 | 861 | 406 | 18,889 | $4.9 \%$ |
| (all) | (all) | 32,894 | 62,715 | 55,186 | 53,525 | 83,876 | 68,466 | 26,515 | 383,176 | $100.0 \%$ |

Table 3.2.2: Northeast Atlantic mackerel. Self-sampling summary with the catch (tonnes) by year and division

Northeast Atlantic mackerel. Catch by month

| species | month | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | $2022 *$ | all | perc |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| mac | Jan | 7,848 | 18,550 | 11,546 | 18,715 | 20,750 | 14,806 | 12,735 | 104,950 | $27.4 \%$ |
| mac | Feb | 1,189 | 8,199 | 7,297 | 11,862 | 19,376 | 5,678 | 6,942 | 60,544 | $15.8 \%$ |
| mac | Mar | 139 | 4,469 | 1,292 | 4,374 | 5,114 | 2,840 | 6,613 | 24,841 | $6.5 \%$ |
| mac | Apr | 701 | 955 | 1,226 | 1,326 | 604 | 366 | 98 | 5,276 | $1.4 \%$ |
| mac | May | 30 | 288 | 192 | 489 | 1,239 | 97 | 71 | 2,406 | $0.6 \%$ |
| mac | Jun | 124 | 0 | 60 | 96 | 173 | 35 | 0 | 489 | $0.1 \%$ |
| mac | Jul | 192 | 89 | 0 | 262 | 83 | 907 | 55 | 1,588 | $0.4 \%$ |
| mac | Aug | 120 | 237 | 59 | 431 | 296 | 360 | 0 | 1,503 | $0.4 \%$ |
| mac | Sep | 943 | 9,096 | 4,779 | 3,039 | 6,284 | 2,624 | 0 | 26,765 | $7.0 \%$ |
| mac | Oct | 13,857 | 7,866 | 19,437 | 11,457 | 20,161 | 30,743 | 0 | 103,521 | $27.0 \%$ |
| mac | Nov | 7,625 | 11,595 | 8,934 | 1,473 | 9,461 | 10,009 | 0 | 49,097 | $12.8 \%$ |
| mac | Dec | 128 | 1,370 | 363 | 0 | 334 | 0 | 0 | 2,195 | $0.6 \%$ |
| (all) | (all) | 32,894 | 62,715 | 55,186 | 53,525 | 83,876 | 68,466 | 26,515 | 383,176 | $100.0 \%$ |

Table 3.2.3: Northeast Atlantic mackerel. Self-sampling summary with the catch (tonnes) by year and month

## Northeast Atlantic mackerel. Catch by country

| species | flag | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | $2022 *$ | all | perc |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| mac | DEU | 6,127 | 6,934 | 9,760 | 8,735 | 22,795 | 10,305 | 7,859 | 72,515 | $18.9 \%$ |
| mac | FR | 0 | 0 | 8,096 | 8,962 | 6,375 | 7,086 | 2,997 | 33,516 | $8.7 \%$ |
| mac | LIT | 0 | 0 | 0 | 0 | 827 | 6,876 | 0 | 7,704 | $2.0 \%$ |
| mac | NL | 16,107 | 29,171 | 12,670 | 14,885 | 27,424 | 20,674 | 5,035 | 125,966 | $32.9 \%$ |
| mac | POL | 0 | 0 | 4,051 | 3,601 | 5,502 | 1,771 | 0 | 14,926 | $3.9 \%$ |
| mac | UK | 10,660 | 26,610 | 20,608 | 17,341 | 20,952 | 19,704 | 10,815 | 126,691 | $33.0 \%$ |
| mac | NA | 0 | 0 | 0 | 0 | 0 | 2,049 | 0 | 2,049 | $0.5 \%$ |
| (all) | (all) | 32,894 | 62,715 | 55,186 | 53,525 | 83,876 | 68,466 | 26,707 | 383,368 | $100.0 \%$ |

Table 3.2.4: Northeast Atlantic mackerel. Self-sampling summary with the catch (tonnes) by year and country

## Northeast Atlantic mackerel. Catch by rectangle



Figure 3.2.1: Northeast Atlantic mackerel. Catch per per rectangle. $N$ indicates the number of hauls; Catch refers to the total catch per year.

Northeast Atlantic mackerel. Catchrate (ton/day) by rectangle


Figure 3.2.2: Northeast Atlantic mackerel. Catchrate (ton/day) per rectangle. $N$ indicates the number of hauls; Avg refers to the average catchrate per rect.

Northeast Atlantic mackerel. Spatio-temporal evolution of catch by month and rectangle


Figure 3.2.3: Northeast Atlantic mackerel. Spatio-temporal evolution of the catches per rectangle and month. $N$ indicates the number of hauls; C refers to the total catch by year and month.

Northeast Atlantic mackerel. Catch proportion at depth


Figure 3.2.4: Northeast Atlantic mackerel. Catch proportion at depth. N indicates the number of hauls.

Northeast Atlantic mackerel. Length distributions of the catch


Figure 3.2.5: Northeast Atlantic mackerel. Length distributions by year (top) and by year and division (bottom). Nobs refers to the number of observations; median denotes the median length.

Northeast Atlantic mackerel. Length distributions as proportions by (large) rectangle


Figure 3.2.6: Northeast Atlantic mackerel. Length distributions as proportions by large rectangle. Ind. refers to the number of length measurements

Northeast Atlantic mackerel. Average length, weight and fat content by year and month


Figure 3.2.7: Northeast Atlantic mackerel. Average length, average weight, and average fat content. Nobs indicates the number of measurements, median indicates the median values

## Northeast Atlantic mackerel (MAC). Standardized CPUE

Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with $\log ($ days $)$ as offset. It is assumed that a $2.5 \%$ annual efficiency increase takes place (Rousseau et al 2019).


Figure 3.2.8: Northeast Atlantic mackerel. Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with log(days) as offset

### 3.3 Western horse mackerel (HOM, Trachurus trachurus) <br> Western horse mackerel self-sampling summary.

| species | year | nvessels | ntrips | ndays | nhauls | catch | catch/day | nlength | nbio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: | ---: | ---: |
| hom | 2016 | 7 | 21 | 171 | 314 | 13,382 | 78 | 11,154 | 0 |
| hom | 2017 | 10 | 25 | 161 | 304 | 11,578 | 72 | 8,176 | 0 |
| hom | 2018 | 13 | 35 | 244 | 431 | 21,412 | 88 | 21,756 | 0 |
| hom | 2019 | 15 | 47 | 363 | 668 | 24,022 | 66 | 14,172 | 25 |
| hom | 2020 | 14 | 40 | 268 | 508 | 16,334 | 61 | 13,531 | 203 |
| hom | 2021 | 17 | 53 | 366 | 643 | 26,576 | 73 | 24,753 | 59 |
| hom | 2022 | 14 | 28 | 166 | 330 | 12,183 | 73 | 8,976 | 269 |
| (all) | (all) |  | 249 | 1,739 | 3,198 | 125,486 |  | 102,518 | 556 |

Table 3.3.1: Western horse mackerel. Self-sampling summary with the number of days, hauls, trips, vessels, catch (tonnes), catch rate (ton/day), number of fish measured, number of biological observations.

Western horse mackerel. Catch by division

| species | division | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | $2022 *$ | all | perc |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| hom | 27.4.a | 7 | 6 | 0 | 11 | 13 | 1,007 | 9 | 1,054 | $0.8 \%$ |
| hom | $27.6 . a$ | 4,751 | 5,343 | 12,067 | 13,849 | 5,901 | 1,564 | 552 | 44,027 | $35.1 \%$ |
| hom | $27.7 . \mathrm{b}$ | 4,313 | 4,741 | 2,250 | 4,176 | 5,226 | 4,743 | 335 | 25,784 | $20.5 \%$ |
| hom | $27.7 . \mathrm{h}$ | 1,297 | 1,329 | 6,282 | 984 | 55 | 8,551 | 197 | 18,695 | $14.9 \%$ |
| hom | $27.7 . j$ | 3,015 | 159 | 813 | 5,002 | 5,138 | 10,712 | 11,089 | 35,927 | $28.6 \%$ |
| (all) | (all) | 13,382 | 11,578 | 21,412 | 24,022 | 16,334 | 26,576 | 12,183 | 125,486 | $100.0 \%$ |

Table 3.3.2: Western horse mackerel. Self-sampling summary with the catch (tonnes) by year and division

Western horse mackerel. Catch by month

| species month | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | $2022 *$ | all | perc |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| hom | Jan | 3,350 | 6,666 | 10,627 | 9,610 | 7,017 | 4,894 | 10,232 | 52,397 | $41.8 \%$ |
| hom | Feb | 5,361 | 3,052 | 5,392 | 3,257 | 4,774 | 6,634 | 1,264 | 29,734 | $23.7 \%$ |
| hom | Mar | 60 | 212 | 3,027 | 1,284 | 1,237 | 245 | 413 | 6,478 | $5.2 \%$ |
| hom | Apr | 174 | 0 | 31 | 45 | 0 | 6 | 0 | 257 | $0.2 \%$ |
| hom | May | 176 | 156 | 7 | 42 | 529 | 2 | 0 | 911 | $0.7 \%$ |
| hom | Jun | 2 | 0 | 227 | 1,357 | 642 | 0 | 0 | 2,228 | $1.8 \%$ |
| hom | Jul | 1,728 | 112 | 15 | 5,342 | 420 | 5,809 | 274 | 13,699 | $10.9 \%$ |
| hom | Aug | 0 | 0 | 0 | 8 | 0 | 1,005 | 0 | 1,013 | $0.8 \%$ |
| hom | Sep | 0 | 0 | 429 | 335 | 0 | 4,300 | 0 | 5,065 | $4.0 \%$ |
| hom | Oct | 27 | 15 | 126 | 259 | 1 | 831 | 0 | 1,259 | $1.0 \%$ |
| hom | Nov | 1,608 | 1,262 | 1,410 | 2,483 | 1,713 | 2,629 | 0 | 11,105 | $8.8 \%$ |
| hom | Dec | 896 | 103 | 120 | 0 | 0 | 221 | 0 | 1,340 | $1.1 \%$ |
| (all) | (all) | 13,382 | 11,578 | 21,412 | 24,022 | 16,334 | 26,576 | 12,183 | 125,486 | $100.0 \%$ |

Table 3.3.3: Western horse mackerel. Self-sampling summary with the catch (tonnes) by year and month

## Western horse mackerel. Catch by country

| species | flag | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | $2022 *$ | all | perc |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| hom | DEU | 3,710 | 1,803 | 4,069 | 2,602 | 977 | 4,155 | 725 | 18,042 | $14.4 \%$ |
| hom | FR | 0 | 0 | 622 | 864 | 1,370 | 788 | 1,400 | 5,043 | $4.0 \%$ |
| hom | NL | 9,211 | 9,239 | 14,617 | 18,011 | 11,535 | 18,234 | 9,605 | 90,452 | $72.1 \%$ |
| hom | POL | 0 | 0 | 0 | 4 | 1,005 | 1,210 | 0 | 2,219 | $1.8 \%$ |
| hom | UK | 461 | 535 | 2,104 | 2,541 | 1,447 | 2,014 | 452 | 9,555 | $7.6 \%$ |
| hom | NA | 0 | 0 | 0 | 0 | 0 | 175 | 0 | 175 | $0.1 \%$ |
| (all) | (all) | 13,382 | 11,578 | 21,412 | 24,022 | 16,334 | 26,576 | 12,183 | 125,486 | $100.0 \%$ |

Table 3.3.4: Western horse mackerel. Self-sampling summary with the catch (tonnes) by year and country

## Western horse mackerel. Catch by rectangle



Figure 3.3.1: Western horse mackerel. Catch per per rectangle. N indicates the number of hauls; Catch refers to the total catch per year.

## Western horse mackerel. Catchrate (ton/day) by rectangle



Figure 3.3.2: Western horse mackerel. Catchrate (ton/day) per rectangle. $N$ indicates the number of hauls; Avg refers to the average catchrate per rect.

## Western horse mackerel. Spatio-temporal evolution of catch by month and rectangle



Figure 3.3.3: Western horse mackerel. Spatio-temporal evolution of the catches per rectangle and month. $N$ indicates the number of hauls; C refers to the total catch by year and month.

## Western horse mackerel. Catch proportion at depth



Figure 3.3.4: Western horse mackerel. Catch proportion at depth. $N$ indicates the number of hauls.

Western horse mackerel. Length distributions of the catch


Figure 3.3.5: Western horse mackerel. Length distributions by year (top) and by year and division (bottom). Nobs refers to the number of observations; median denotes the median length.

Western horse mackerel. Length distributions as proportions by (large) rectangle


Figure 3.3.6: Western horse mackerel. Length distributions as proportions by large rectangle. Ind. refers to the number of length measurements

Western horse mackerel. Average length, weight and fat content by year and month


Figure 3.3.7: Western horse mackerel. Average length, average weight, and average fat content. Nobs indicates the number of measurements, median indicates the median values

## Western horse mackerel (HOM). Standardized CPUE

Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with $\log$ (days) as offset. It is assumed that a $2.5 \%$ annual efficiency increase takes place (Rousseau et al 2019).


Figure 3.3.8: Western horse mackerel. Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with log(days) as offset

### 3.4 North Sea horse mackerel (HOM, Trachurus trachurus)

North Sea horse mackerel self-sampling summary.

| species | year | nvessels | ntrips | ndays | nhauls | catch | catch/day | nlength | nbio |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| hom | 2016 | 5 | 16 | 77 | 130 | 6,359 | 83 | 6,313 | 0 |
| hom | 2017 | 6 | 14 | 81 | 156 | 8,568 | 106 | 1,013 | 0 |
| hom | 2018 | 5 | 13 | 80 | 146 | 7,079 | 88 | 4,349 | 0 |
| hom | 2019 | 8 | 14 | 78 | 143 | 7,417 | 95 | 9,448 | 0 |
| hom | 2020 | 7 | 21 | 94 | 150 | 8,726 | 93 | 10,685 | 829 |
| hom | 2021 | 8 | 22 | 94 | 153 | 7,259 | 77 | 6,320 | 0 |
| hom | 2022 | 5 | 9 | 17 | 22 | 914 | 54 | 855 | 0 |
| (all) | (all) |  | 109 | 521 | 900 | 46,322 |  | 38,983 | 829 |

Table 3.4.1: North Sea horse mackerel. Self-sampling summary with the number of days, hauls, trips, vessels, catch (tonnes), catch rate (ton/day), number of fish measured, number of biological observations.

North Sea horse mackerel. Catch by division
species division 2016 2017 2018 2019 2020 2021 2022 all perc

| hom | $27.4 . c$ | 0 | 1,371 | 853 | 369 | 898 | 1,149 | 0 | 4,640 | $10.0 \%$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| hom | $27.7 . \mathrm{d}$ | 6,358 | 7,198 | 6,226 | 7,048 | 7,829 | 6,111 | 914 | 41,682 | $90.0 \%$ |
| (all) | (all) | 6,359 | 8,568 | 7,079 | 7,417 | 8,726 | 7,259 | 914 | 46,322 | $100.0 \%$ |

Table 3.4.2: North Sea horse mackerel. Self-sampling summary with the catch (tonnes) by year and division

North Sea horse mackerel. Catch by month

| species | month | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | $2022 *$ | all | perc |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| hom | Jan | 0 | 2,362 | 892 | 1,382 | 2 | 1,013 | 538 | 6,189 | $13.4 \%$ |
| hom | Feb | 879 | 0 | 310 | 0 | 0 | 97 | 376 | 1,662 | $3.6 \%$ |
| hom | Mar | 38 | 0 | 0 | 0 | 0 | 0 | 0 | 38 | $0.1 \%$ |
| hom | Jun | 0 | 0 | 0 | 0 | 6 | 25 | 0 | 31 | $0.1 \%$ |
| hom | Jul | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $0.0 \%$ |
| hom | Aug | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | $0.0 \%$ |
| hom | Sep | 447 | 135 | 1,471 | 2,009 | 3,860 | 422 | 0 | 8,344 | $18.0 \%$ |
| hom | oct | 1,802 | 4,490 | 1,391 | 1,967 | 1,834 | 2,349 | 0 | 13,833 | $29.9 \%$ |
| hom | Nov | 2,873 | 1,581 | 2,018 | 1,110 | 1,463 | 1,218 | 0 | 10,263 | $22.2 \%$ |
| hom | Dec | 312 | 0 | 998 | 949 | 1,561 | 2,134 | 0 | 5,954 | $12.9 \%$ |
| (all) | (all) | 6,359 | 8,568 | 7,079 | 7,417 | 8,726 | 7,259 | 914 | 46,322 | $100.0 \%$ |

Table 3.4.3: North Sea horse mackerel. Self-sampling summary with the catch (tonnes) by year and month

## North Sea horse mackerel. Catch by country

| species | flag | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | $2022 *$ | all | perc |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| hom | DEU | 593 | 0 | 1,378 | 958 | 0 | 0 | 0 | 2,930 | $6.3 \%$ |
| hom | FR | 0 | 0 | 422 | 400 | 238 | 202 | 0 | 1,261 | $2.7 \%$ |
| hom | LIT | 0 | 0 | 0 | 1,373 | 0 | 0 | 0 | 1,373 | $3.0 \%$ |
| hom | NL | 2,383 | 4,887 | 1,578 | 1,682 | 4,167 | 2,356 | 436 | 17,487 | $37.8 \%$ |
| hom | UK | 3,383 | 3,682 | 3,701 | 3,004 | 4,322 | 3,674 | 478 | 22,243 | $48.0 \%$ |
| hom | NA | 0 | 0 | 0 | 0 | 0 | 1,028 | 0 | 1,028 | $2.2 \%$ |
| (all) | (all) | 6,359 | 8,568 | 7,079 | 7,417 | 8,726 | 7,259 | 914 | 46,322 | $100.0 \%$ |

Table 3.4.4: North Sea horse mackerel. Self-sampling summary with the catch (tonnes) by year and country

## North Sea horse mackerel. Catch by rectangle



Figure 3.4.1: North Sea horse mackerel. Catch per per rectangle. $N$ indicates the number of hauls; Catch refers to the total catch per year.

North Sea horse mackerel. Catchrate (ton/day) by rectangle


Figure 3.4.2: North Sea horse mackerel. Catchrate (ton/day) per rectangle. N indicates the number of hauls; Avg refers to the average catchrate per rect.

North Sea horse mackerel. Spatio-temporal evolution of catch by month and rectangle


Figure 3.4.3: North Sea horse mackerel. Spatio-temporal evolution of the catches per rectangle and month. $N$ indicates the number of hauls; C refers to the total catch by year and month.

North Sea horse mackerel. Catch proportion at depth


Figure 3.4.4: North Sea horse mackerel. Catch proportion at depth. $N$ indicates the number of hauls.

North Sea horse mackerel. Length distributions of the catch


Figure 3.4.5: North Sea horse mackerel. Length distributions by year (top) and by year and division (bottom). Nobs refers to the number of observations; median denotes the median length.

North Sea horse mackerel. Length distributions as proportions by (large) rectangle


Figure 3.4.6: North Sea horse mackerel. Length distributions as proportions by large rectangle. Ind. refers to the number of length measurements

North Sea horse mackerel. Average length, weight and fat content by year and month


Figure 3.4.7: North Sea horse mackerel. Average length, average weight, and average fat content. Nobs indicates the number of measurements, median indicates the median values

## North Sea horse mackerel (HOM). Standardized CPUE

Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with $\log$ (days) as offset. It is assumed that a $2.5 \%$ annual efficiency increase takes place (Rousseau et al 2019).


Figure 3.4.8: North Sea horse mackerel. Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with log(days) as offset

### 3.5 Blue whiting (WHB, Micromesistius pouttasseu)

Blue whiting self-sampling summary.

| species | year | nvessels | ntrips | ndays | nhauls | catch | catch/day | nlength | nbio |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| whb | 2016 | 8 | 22 | 198 | 462 | 40,535 | 205 | 27,315 | 0 |
| whb | 2017 | 8 | 32 | 343 | 753 | 78,325 | 228 | 63,682 | 0 |
| whb | 2018 | 12 | 42 | 550 | 1,375 | 149,723 | 272 | 112,492 | 0 |
| whb | 2019 | 14 | 46 | 457 | 1,089 | 109,234 | 239 | 50,057 | 0 |
| whb | 2020 | 13 | 57 | 670 | 1,581 | 168,786 | 252 | 83,177 | 178 |
| whb | 2021 | 14 | 52 | 532 | 1,185 | 138,946 | 261 | 58,391 | 0 |
| whb | 2022 | 15 | 33 | 406 | 962 | 87,325 | 215 | 34,068 | 0 |
| (all) | (all) |  | 284 | 3,156 | 7,407 | 772,874 |  | 429,182 | 178 |

Table 3.5.1: Blue whiting. Self-sampling summary with the number of days, hauls, trips, vessels, catch (tonnes), catch rate (ton/day), number of fish measured, number of biological observations.

Blue whiting. Catch by division


Table 3.5.2: Blue whiting. Self-sampling summary with the catch (tonnes) by year and division
Blue whiting. Catch by month

| whb | Jan | 85 | 185 | 957 | 4,287 | 9,527 | 29,603 | 14,391 | 59,034 | 7.6\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| whb | Feb | 1,683 | 8,027 | 19,108 | 17,504 | 4,051 | 18,915 | 16,468 | 85,755 | 11.1\% |
| whb | Mar | 15,317 | 24,683 | 26,954 | 21,389 | 41,128 | 30,134 | 32,907 | 192,513 | 24.9\% |
| whb | Apr | 13,328 | 27,316 | 55,518 | 26,391 | 61,978 | 25,146 | 19,539 | 229,216 | 29.7\% |
| whb | May | 5,001 | 9,390 | 24,093 | 15,465 | 22,506 | 8,571 | 4,020 | 89,045 | 11.5\% |
| whb | Jun | 697 | 0 | 5,004 | 0 | 697 | 0 | 0 | 6,398 | $0.8 \%$ |
| whb | Jul | 10 | 0 | 0 | 7 | 13 | 0 | 0 | 30 | 0.0\% |
| whb | Aug | 0 | 1,265 | 4,219 | 337 | 2,043 | 0 | 0 | 7,864 | 1.0\% |
| whb | Sep | 50 | 538 | 414 | 246 | 1,327 | 2 | 0 | 2,576 | 0.3\% |
| whb | Oct | 266 | 39 | 92 | 407 | 2,401 | 4 | 0 | 3,209 | 0.4\% |
| whb | Nov | 1,665 | 5,623 | 6,413 | 13,841 | 7,283 | 11,275 | 0 | 46,099 | 6.0\% |
| whb | Dec | 2,432 | 1,260 | 6,952 | 9,361 | 15,834 | 15,296 | 0 | 51,135 | 6.6\% |
| (all) | (all) | 40,535 | 78,325 | 149,723 | 109,234 | 168,786 | 138,946 | 87,325 | 772,874 | 100.0\% |

Table 3.5.3: Blue whiting. Self-sampling summary with the catch (tonnes) by year and month Blue whiting. Catch by country

| species | flag | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | $2022 *$ | all |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| whb | DEU | 13,545 | 15,914 | 35,831 | 23,479 | 39,647 | 33,190 | 16,635 | 178,240 | $23.1 \%$ |
| whb | FR | 0 | 0 | 1,625 | 4,892 | 5,069 | 2,786 | 4,188 | 18,561 | $2.4 \%$ |
| whb | LIT | 0 | 0 | 0 | 0 | 10,146 | 15,807 | 6,467 | 32,421 | $4.2 \%$ |
| whb | NL | 26,940 | 59,027 | 98,499 | 53,538 | 60,454 | 52,365 | 41,147 | 391,969 | $50.7 \%$ |
| whb | POL | 0 | 0 | 11,764 | 23,192 | 45,791 | 26,288 | 11,237 | 118,273 | $15.3 \%$ |
| whb | UK | 50 | 3,385 | 2,004 | 4,133 | 7,678 | 8,510 | 7,650 | 33,410 | $4.3 \%$ |
| (all) | (all) | 40,535 | 78,325 | 149,723 | 109,234 | 168,786 | 138,946 | 87,325 | 772,874 | $100.0 \%$ |

Table 3.5.4: Blue whiting. Self-sampling summary with the catch (tonnes) by year and country

Blue whiting. Catch by rectangle


Figure 3.5.1: Blue whiting. Catch per per rectangle. $N$ indicates the number of hauls; Catch refers to the total catch per year.

Blue whiting. Catchrate (ton/day) by rectangle


Figure 3.5.2: Blue whiting. Catchrate (ton/day) per rectangle. $N$ indicates the number of hauls; Avg refers to the average catchrate per rect.

Blue whiting. Spatio-temporal evolution of catch by month and rectangle


Figure 3.5.3: Blue whiting. Spatio-temporal evolution of the catches per rectangle and month. N indicates the number of hauls; C refers to the total catch by year and month.

Blue whiting. Catch proportion at depth


Figure 3.5.4: Blue whiting. Catch proportion at depth. $N$ indicates the number of hauls.

Blue whiting. Length distributions of the catch


Figure 3.5.5: Blue whiting. Length distributions by year (top) and by year and division (bottom). Nobs refers to the number of observations; median denotes the median length.

Blue whiting. Length distributions as proportions by (large) rectangle


Figure 3.5.6: Blue whiting. Length distributions as proportions by large rectangle. Ind. refers to the number of length measurements

Blue whiting. Average length, weight and fat content by year and month


Figure 3.5.7: Blue whiting. Average length, average weight, and average fat content. Nobs indicates the number of measurements, median indicates the median values

## Blue whiting (WHB). Standardized CPUE

Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with $\log$ (days) as offset. It is assumed that a $2.5 \%$ annual efficiency increase takes place (Rousseau et al 2019).


Figure 3.5.8: Blue whiting. Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with log(days) as offset

### 3.6 Atlanto-scandian herring (HER_ASH, Clupea harengus)

Atlanto-scandian herring self-sampling summary.


Table 3.6.1: Atlanto-scandian herring. Self-sampling summary with the number of days, hauls, trips, vessels, catch (tonnes), catch rate (ton/day), number of fish measured, number of biological observations.

## Atlanto-scandian herring. Catch by division

| species | division | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | all | perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| her_ash | 27.2.a | 2,109 | 4,913 | 1,367 | 3,373 | 3,563 | 2,379 | 17,706 | 100.0\% |
| (all) | (all) | 2,109 | 4,913 | 1,367 | 3,373 | 3,563 | 2,379 | 17,706 | 100.0\% |

Table 3.6.2: Atlanto-scandian herring. Self-sampling summary with the catch (tonnes) by year and division

Atlanto-scandian herring. Catch by month

| species | month | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | all | perc |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| her_ash | May | 0 | 0 | 0 | 0 | 26 | 0 | 26 | $0.1 \%$ |
| her_ash | Aug | 0 | 118 | 52 | 0 | 61 | 0 | 232 | $1.3 \%$ |
| her_ash | Sep | 54 | 7 | 405 | 362 | 53 | 0 | 881 | $5.0 \%$ |
| her_ash | Oct | 2,055 | 4,788 | 910 | 2,184 | 2,480 | 1,659 | 14,076 | $79.5 \%$ |
| her_ash | Nov | 0 | 0 | 0 | 828 | 942 | 721 | 2,491 | $14.1 \%$ |
| (all) | (all) | 2,109 | 4,913 | 1,367 | 3,373 | 3,563 | 2,379 | 17,706 | $100.0 \%$ |

Table 3.6.3: Atlanto-scandian herring. Self-sampling summary with the catch (tonnes) by year and month

Atlanto-scandian herring. Catch by country

| species | flag | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | all | perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| her_ash | DEU | 1,237 | 707 | 0 | 719 | 1,036 | 721 | 4,419 | 25.0\% |
| her_ash | LIT | 0 | 0 | 0 | 0 | 1,098 | 0 | 1,098 | 6.2\% |
| her_ash | NL | 775 | 4,185 | 1,367 | 2,654 | 524 | 1,659 | 11,164 | 63.1\% |
| her_ash | POL | 0 | 0 | 0 | 0 | 859 | 0 | 859 | 4.9\% |
| her_ash | UK | 97 | 21 | 0 | 0 | 48 | 0 | 166 | 0.9\% |
| (all) | (all) | 2,109 | 4,913 | 1,367 | 3,373 | 3,563 | 2,379 | 17,706 | 100.0\% |

Table 3.6.4: Atlanto-scandian herring. Self-sampling summary with the catch (tonnes) by year and country

## Atlanto-scandian herring. Catch by rectangle



Figure 3.6.1: Atlanto-scandian herring. Catch per per rectangle. $N$ indicates the number of hauls; Catch refers to the total catch per year.

Atlanto-scandian herring. Catchrate (ton/day) by rectangle


Figure 3.6.2: Atlanto-scandian herring. Catchrate (ton/day) per rectangle. N indicates the number of hauls; Avg refers to the average catchrate per rect.

## Atlanto-scandian herring. Spatio-temporal evolution of catch by month and rectangle



Figure 3.6.3: Atlanto-scandian herring. Spatio-temporal evolution of the catches per rectangle and month. $N$ indicates the number of hauls; C refers to the total catch by year and month.

Atlanto-scandian herring. Catch proportion at depth


Figure 3.6.4: Atlanto-scandian herring. Catch proportion at depth. $N$ indicates the number of hauls.

Atlanto-scandian herring. Length distributions of the catch


Figure 3.6.5: Atlanto-scandian herring. Length distributions by year (top) and by year and division (bottom). Nobs refers to the number of observations; median denotes the median length.

Atlanto-scandian herring. Length distributions as proportions by (large) rectangle


Figure 3.6.6: Atlanto-scandian herring. Length distributions as proportions by large rectangle. Ind. refers to the number of length measurements

Atlanto-scandian herring. Average length, weight and fat content by year and month


Figure 3.6.7: Atlanto-scandian herring. Average length, average weight, and average fat content. Nobs indicates the number of measurements, median indicates the median values

## 4 Discussion and conclusions

The PFA self-sampling program has been carried out for the seventh year in a row (2015-2021). Here, results have been presented for the years 2016-2022 in terms of meta-information on the sampling (number of vessels, trips, days and length measurements per area and/or season), in terms of the spatio-temporal distribution of catches and the length and weight compositions by area and/or season.

The definition of what constitutes the 'widely distributed fishery' has been approached by selecting all combination of vessel-trip-weeks where hauls were taken in a certain area and where the catch composition consisted of a minimum percentage of certain species (blue whiting, mackerel, horse mackerel, Atlanto-scandian herring) and a minimum weekly catch of 10 tons. Although for herring we aimed to select only trips for Atlanto-scandian herring (in division 27.2.a) some trips with North Sea herring have been included because they were combined with some fishing for mackerel.

Trips from 2016 up to 11/08/2022 have been processed for this overview. Pelagic fisheries within the Pelagic Freezer-trawler Association are carried out by vessels from different countries. Overall, around $48 \%$ of the catch volume of trips in this overview were taken by Dutch trawlers, $22 \%$ German trawlers, $14 \%$ UK trawlers and $16 \%$ other countries. Blue whiting constitutes the majority of the catch in those trips (54\%), followed by mackerel (23\%) and horse mackerel (12\%). Atlanto-scandian herring only constitutes around $3 \%$ of the volume in the PFA widely distributed fishery. Note that the North Sea herring fishery is not included in this overview.

The Mackerel fishery takes place from October through to March of the subsequent year. Bycatches of mackerel may also occur during other fisheries,e.g. for horse mackerel or herring. Overall, the selfsampling activities for the mackerel fisheries during the years 2016-2022 (up to 11/08/2022) covered 465 fishing trips with 6352 hauls, a total catch of 386474 tonnes and 103745 individual length measurements. The main fishing areas are ICES division 27.4.a and division 27.6.a. Compared to the previous years, mackerel in the catch in 2021 have been relatively large with a median length of 36.4 cm compared to 33.6-36.2 in the preceding years. Also, the median weight has been somewhat higher with median weight of 435 gram compared to 385-422 gram in the preceding years.

The Western horse mackerel fishery takes place from October through to March of the subsequent year. Overall, the self-sampling activities for the Western horse mackerel fisheries during the years 2016-2022 (up to 11/08/2022) covered 250 fishing trips with 3316 hauls, a total catch of 128553 tonnes and 130146 individual length measurements. The main fishing areas are ICES division 27.6.a, division 27.7.b and division 27.7.j. Western horse mackerel have a wide range in the length distributions in the catch. Median lengths in divisions 27.6.a, 27.7.b and 27.7.j have fluctuated between 25.2 and 31.9 cm (with one low median length of 22.7 cm in 27.6.a in 2018). In ICES division 27.7.h, median lengths in the catch have been smaller and fluctuated between 20.7 and 24.5 cm .

The North Sea horse mackerel fishery takes place from October through to January of the subsequent year. Overall, the self-sampling activities for the North Sea horse mackerel fisheries during the years 2016-2022 (up to 11/08/2022) covered 109 fishing trips with 900 hauls, a total catch of 46322 tonnes and 38983 individual length measurements. The main fishing areas is ICES division 27.7.d with some minor catches in 27.4.c. Catches in division 27.4.a have been counted as Western Horse mackerel. North Sea horse mackerel have a narrow range in the length distributions in the catch. Median lengths in division 27.7.d have fluctuated between 20.7 and 24.3 cm .

The blue whiting fishery takes place from February through to May although some minor fisheries for blue whiting may remain over the other months. Overall, the self-sampling activities for the horse mackerel fisheries during the years 2016-2022 (up to 11/08/2022) covered 320 fishing trips with 8234 hauls, a total catch of 810714 tonnes and 466229 individual length measurements. The main fishing areas are ICES division 27.6.a, division 27.7.c and division 27.7.k. Compared to the previous years, blue whiting in the catches during 2020-2022 have been relatively large with a median length of 27.8 cm compared to 24.1-24.5 in the preceding years.

The fishery for Atlanto-scandian herring (ASH) is a relatively smaller fishery for PFA and takes place mostly in October. Overall, the self-sampling activities for the horse mackerel fisheries during the years 2016-2022 (up to 11/08/2022) covered 32 fishing trips with 297 hauls, a total catch of 17705 tonnes and 5147 individual length measurements. Only the herring fishery in ICES division 27.2.a is considered for ASH. Note that there are herring catches in other divisions within the selected trips. These are trips where North Sea herring has been fished with some bycatches of mackerel for example. Atlanto-scandian herring have a relatively narrow range in the length distributions in the catch. Median lengths have been between 30 and 35 cm .

In this 2022 self-sampling report, a standardized CPUE calculation has been included for the first time for most of the stocks. The standardized CPUE is based on a GLM model with a negative binomial distribution. The response variable is the catch by week and vessel, with an offset of the log effort (number of fishing days per week) and explanatory variables year, GT category, month, division and depth category. An assumed technical efficiency increase of $2.5 \%$ per year has been included in the fitting of the model (Rousseau et al 2019).

## 5 Acknowledgements

The skippers, officers and the quality managers of many of the PFA vessels are putting in a lot of effort to make the PFA the self-sampling work. Without their efforts, there would be no self-sampling.

## 6 References and publications

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## 7 More information

Please contact Martin Pastoors (mpastoors@pelagicfish.eu) if you would have any questions on the PFA self-sampling program or the specific results presented here.

## 8 Northeast Atlantic mackerel: detailed tables

Northeast Atlantic mackerel Sampling overview


Northeast Atlantic mackerel Length frequencies 2021


| mac | 2021 | 1 | 27 | 27.7.b | TL | 32 | 1 | 158 | 0.0003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 33 | 1 | 9116 | 0.0202 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 34 | 3 | 27349 | 0.0605 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 35 | 1 | 5243 | 0.0116 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 36 | 7 | 44463 | 0.0983 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 37 | 11 | 69334 | 0.1534 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 38 | 18 | 125023 | 0.2765 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 39 | 15 | 112427 | 0.2487 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 40 | 7 | 44475 | 0.0984 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 41 | 1 | 5243 | 0.0116 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 42 | 1 | 9116 | 0.0202 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 28 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 29 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 30 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 31 |  | 128253 | 0.0968 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 32 | 12 | 256507 | 0.1935 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 33 | 9 | 192380 | 0.1452 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 34 | 8 | 171004 | 0.1290 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 35 | 8 | 106877 | 0.0806 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 36 | 11 | 149629 | 0.1129 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 37 | 9 | 85502 | 0.0645 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 38 | 7 | 149629 | 0.1129 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 39 | 3 | 21375 | 0.0161 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 32 | 1 | 19069 | 0.0216 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 33 | 3 | 54106 | 0.0613 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 34 | 7 | 77064 | 0.0873 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 35 | 9 | 115011 | 0.1303 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 36 | 14 | 180151 | 0.2040 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 37 | 11 | 161067 | 0.1824 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 38 | 8 | 99500 | 0.1127 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 39 | 6 | 94816 | 0.1074 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 40 | 4 | 66247 | 0.0750 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 42 | 1 | 15967 | 0.0181 |
| mac | 2021 | 2 | 27 | 27.7.b | TL | 38 | 1 | 6127 | 0.5000 |
| mac | 2021 | 2 | 27 | 27.7.b | TL | 40 | 1 | 6127 | 0.5000 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 24 | 1 | 9442 | 0.0054 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 25 | 5 | 43353 | 0.0249 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 26 | 12 | 97836 | 0.0561 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 27 | 2 | 10111 | 0.0058 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 28 | 2 | 9810 | 0.0056 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 29 | 10 | 98209 | 0.0563 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 30 | 23 | 131613 | 0.0755 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 31 | 16 | 127653 | 0.0732 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 32 | 28 | 166532 | 0.0955 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 33 | 23 | 149258 | 0.0856 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 34 | 26 | 179596 | 0.1030 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 35 | 23 | 173859 | 0.0998 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 36 | 25 | 179053 | 0.1027 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 37 | 24 | 155003 | 0.0889 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 38 | 19 | 123471 | 0.0708 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 39 | 11 | 74916 | 0.0430 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 40 | 2 | 13203 | 0.0076 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 14 | 4 | 14157 | 0.0952 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 15 | 17 | 60167 | 0.4048 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 16 | 14 | 49549 | 0.3333 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 17 | 5 | 17696 | 0.1190 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 18 | 2 | 7078 | 0.0476 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 28 | 9 | 112126 | 0.1232 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 29 | 13 | 203625 | 0.2238 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 30 | 11 | 141073 | 0.1550 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 31 | 7 | 76669 | 0.0843 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 32 | 4 | 25451 | 0.0280 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 33 | 6 | 23804 | 0.0262 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 34 | 4 | 35446 | 0.0390 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 35 | 13 | 115133 | 0.1265 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 36 | 8 | 64557 | 0.0710 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 37 | 2 | 23926 | 0.0263 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 38 | 7 | 66987 | 0.0736 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 39 | 2 | 17577 | 0.0193 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 40 | 1 | 3499 | 0.0038 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 27 | 1 | 2840 | 0.0020 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 28 | 16 | 151373 | 0.1086 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 29 | 11 | 47096 | 0.0338 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 30 | 9 | 53242 | 0.0382 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 31 | 12 | 62300 | 0.0447 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 32 | 6 | 27933 | 0.0200 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 33 | 18 | 116680 | 0.0837 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 34 | 14 | 154071 | 0.1105 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 35 | 29 | 191132 | 0.1371 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 36 | 32 | 259037 | 0.1858 |


| mac | 2021 | 3 | 27 | 27.7.j | TL | 37 | 32 | 150305 | 0.1078 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 38 | 15 | 80969 | 0.0581 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 39 | 7 | 70033 | 0.0502 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 40 | 5 | 16515 | 0.0118 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 41 | 1 | 10581 | 0.0076 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 35 | 4 | 1061 | 0.0337 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 36 | 4 | 1141 | 0.0362 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 37 | 14 | 18943 | 0.6011 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 38 | 19 | 5141 | 0.1631 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 39 | 10 | 2534 | 0.0804 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 40 | 8 | 2202 | 0.0699 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 41 | 2 | 490 | 0.0155 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 23 | 1 | 11982 | 0.0002 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 24 | 9 | 108983 | 0.0015 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 25 | 21 | 231669 | 0.0031 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 26 | 76 | 793340 | 0.0107 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 27 | 138 | 1561556 | 0.0210 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 28 | 138 | 1922592 | 0.0258 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 29 | 190 | 2253457 | 0.0303 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 30 | 253 | 2750277 | 0.0370 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 31 | 318 | 3004111 | 0.0404 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 32 | 392 | 4157073 | 0.0559 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 33 | 482 | 5170802 | 0.0695 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 34 | 651 | 6524440 | 0.0877 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 35 | 902 | 8714365 | 0.1171 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 36 | 1037 | 9457713 | 0.1271 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 37 | 1253 | 11146124 | 0.1498 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 38 | 1084 | 8877236 | 0.1193 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 39 | 603 | 4858841 | 0.0653 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 40 | 263 | 2138178 | 0.0287 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 41 | 69 | 522747 | 0.0070 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 42 | 17 | 176805 | 0.0024 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 43 | 3 | 24148 | 0.0003 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 44 | 2 | 5576 | 0.0001 |
| mac | 2021 | 4 | 27 | 27.7.j | TL | 30 | 1 | 1413 | 1.0000 |

Northeast Atlantic mackerel Length frequencies 2022


| mac | 2022 | 1 | 27 | 27.6.a | TL | 38 | 274 | 2565070 | 0.1704 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 39 | 214 | 2171710 | 0.1442 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 40 | 117 | 1193579 | 0.0793 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 41 | 29 | 282289 | 0.0187 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 42 | 7 | 89517 | 0.0059 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 20 | 3 | 34527 | 0.0034 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 21 | 5 | 57545 | 0.0056 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 22 | 7 | 80563 | 0.0079 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 23 | 7 | 80563 | 0.0079 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 24 | 4 | 46036 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 25 | 5 | 57545 | 0.0056 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 26 | 4 | 46036 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 27 | 4 | 35245 | 0.0035 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 28 | 3 | 31860 | 0.0031 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 29 | 7 | 84735 | 0.0083 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 30 | 25 | 216377 | 0.0212 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 31 | 24 | 196195 | 0.0192 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 32 | 31 | 296552 | 0.0291 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 33 | 45 | 315556 | 0.0310 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 34 | 72 | 507485 | 0.0498 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 35 | 162 | 1101891 | 0.1081 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 36 | 203 | 1318078 | 0.1293 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 37 | 201 | 1323574 | 0.1298 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 38 | 270 | 1866023 | 0.1831 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 39 | 211 | 1395078 | 0.1369 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 40 | 110 | 751821 | 0.0738 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 41 | 34 | 283483 | 0.0278 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 42 | 6 | 46318 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 43 | 3 | 9300 | 0.0009 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 44 | 1 | 10778 | 0.0011 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 25 | 2 | 6604 | 0.0129 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 26 | 1 | 3302 | 0.0064 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 27 | 3 | 14565 | 0.0284 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 28 | 6 | 20668 | 0.0403 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 29 | 10 | 46168 | 0.0899 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 30 | 20 | 74332 | 0.1448 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 31 | 16 | 73536 | 0.1432 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 32 | 13 | 59549 | 0.1160 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 33 | 15 | 55851 | 0.1088 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 34 | 15 | 44410 | 0.0865 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 35 | 16 | 30610 | 0.0596 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 36 | 17 | 21822 | 0.0425 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 37 | 11 | 26273 | 0.0512 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 38 | 8 | 18422 | 0.0359 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 39 | 6 | 12616 | 0.0246 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 40 | 3 | 4754 | 0.0093 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 34 | 4 | 20848 | 0.1319 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 35 | 5 | 31786 | 0.2011 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 36 | 5 | 28235 | 0.1786 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 37 | 6 | 27219 | 0.1722 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 38 | 6 | 28043 | 0.1774 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 39 | 3 | 14125 | 0.0894 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 40 | 1 | 2102 | 0.0133 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 41 | 1 | 2102 | 0.0133 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 42 | 1 | 3602 | 0.0228 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 26 | 1 | 3261 | 0.0010 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 29 | 9 | 102965 | 0.0330 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 30 | 68 | 847404 | 0.2719 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 31 | 65 | 831571 | 0.2668 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 32 | 53 | 568987 | 0.1825 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 33 | 36 | 409378 | 0.1313 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 34 | 11 | 109308 | 0.0351 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 35 | 4 | 54797 | 0.0176 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 36 | 4 | 67573 | 0.0217 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 37 | 5 | 74347 | 0.0239 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 38 | 3 | 47380 | 0.0152 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 32 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 34 | 5 | 4369 | 0.4168 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 35 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 36 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 37 | 2 | 1747 | 0.1667 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 38 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 40 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 31 | 1 | 4751 | 0.0457 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 32 | 2 | 17221 | 0.1656 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 33 | 1 | 12469 | 0.1199 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 34 | 1 | 9293 | 0.0894 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 35 | 3 | 14255 | 0.1371 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 36 | 2 | 9503 | 0.0914 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 37 | , | 13494 | 0.1298 |


| mac | 2022 | 3 | 27 | $27.7 . j$ | TL | 38 | 2 | 18246 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

## 9 Western horse mackerel: detailed tables

Western horse mackerel Sampling overview


Western horse mackerel Length frequencies 2021


| mac | 2021 | 1 | 27 | 27.7.b | TL | 32 | 1 | 158 | 0.0003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 33 | 1 | 9116 | 0.0202 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 34 | 3 | 27349 | 0.0605 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 35 | 1 | 5243 | 0.0116 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 36 | 7 | 44463 | 0.0983 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 37 | 11 | 69334 | 0.1534 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 38 | 18 | 125023 | 0.2765 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 39 | 15 | 112427 | 0.2487 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 40 | 7 | 44475 | 0.0984 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 41 | 1 | 5243 | 0.0116 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 42 | 1 | 9116 | 0.0202 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 28 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 29 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 30 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 31 | 6 | 128253 | 0.0968 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 32 | 12 | 256507 | 0.1935 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 33 | 9 | 192380 | 0.1452 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 34 | 8 | 171004 | 0.1290 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 35 | 8 | 106877 | 0.0806 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 36 | 11 | 149629 | 0.1129 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 37 | 9 | 85502 | 0.0645 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 38 | 7 | 149629 | 0.1129 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 39 | 3 | 21375 | 0.0161 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 32 | 1 | 19069 | 0.0216 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 33 | 3 | 54106 | 0.0613 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 34 | 7 | 77064 | 0.0873 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 35 | 9 | 115011 | 0.1303 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 36 | 14 | 180151 | 0.2040 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 37 | 11 | 161067 | 0.1824 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 38 | 8 | 99500 | 0.1127 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 39 | 6 | 94816 | 0.1074 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 40 | 4 | 66247 | 0.0750 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 42 | 1 | 15967 | 0.0181 |
| mac | 2021 | 2 | 27 | 27.7.b | TL | 38 | 1 | 6127 | 0.5000 |
| mac | 2021 | 2 | 27 | 27.7.b | TL | 40 | 1 | 6127 | 0.5000 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 24 | 1 | 9442 | 0.0054 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 25 | 5 | 43353 | 0.0249 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 26 | 12 | 97836 | 0.0561 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 27 | 2 | 10111 | 0.0058 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 28 | 2 | 9810 | 0.0056 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 29 | 10 | 98209 | 0.0563 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 30 | 23 | 131613 | 0.0755 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 31 | 16 | 127653 | 0.0732 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 32 | 28 | 166532 | 0.0955 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 33 | 23 | 149258 | 0.0856 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 34 | 26 | 179596 | 0.1030 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 35 | 23 | 173859 | 0.0998 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 36 | 25 | 179053 | 0.1027 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 37 | 24 | 155003 | 0.0889 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 38 | 19 | 123471 | 0.0708 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 39 | 11 | 74916 | 0.0430 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 40 | 2 | 13203 | 0.0076 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 14 | 4 | 14157 | 0.0952 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 15 | 17 | 60167 | 0.4048 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 16 | 14 | 49549 | 0.3333 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 17 | 5 | 17696 | 0.1190 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 18 | 2 | 7078 | 0.0476 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 28 | 9 | 112126 | 0.1232 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 29 | 13 | 203625 | 0.2238 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 30 | 11 | 141073 | 0.1550 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 31 | 7 | 76669 | 0.0843 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 32 | 4 | 25451 | 0.0280 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 33 | 6 | 23804 | 0.0262 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 34 | 4 | 35446 | 0.0390 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 35 | 13 | 115133 | 0.1265 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 36 | 8 | 64557 | 0.0710 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 37 | 2 | 23926 | 0.0263 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 38 | 7 | 66987 | 0.0736 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 39 | 2 | 17577 | 0.0193 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 40 | 1 | 3499 | 0.0038 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 27 | 1 | 2840 | 0.0020 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 28 | 16 | 151373 | 0.1086 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 29 | 11 | 47096 | 0.0338 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 30 | 9 | 53242 | 0.0382 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 31 | 12 | 62300 | 0.0447 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 32 | 6 | 27933 | 0.0200 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 33 | 18 | 116680 | 0.0837 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 34 | 14 | 154071 | 0.1105 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 35 | 29 | 191132 | 0.1371 |
| mac | 2021 | 3 |  | 27.7.j | TL | 36 | 32 | 259037 | 0.1858 |


| mac | 2021 | 3 | 27 | 27.7.j | TL | 37 | 32 | 150305 | 0.1078 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 38 | 15 | 80969 | 0.0581 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 39 | 7 | 70033 | 0.0502 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 40 | 5 | 16515 | 0.0118 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 41 | 1 | 10581 | 0.0076 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 35 | 4 | 1061 | 0.0337 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 36 | 4 | 1141 | 0.0362 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 37 | 14 | 18943 | 0.6011 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 38 | 19 | 5141 | 0.1631 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 39 | 10 | 2534 | 0.0804 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 40 | 8 | 2202 | 0.0699 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 41 | 2 | 490 | 0.0155 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 23 | 1 | 11982 | 0.0002 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 24 | 9 | 108983 | 0.0015 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 25 | 21 | 231669 | 0.0031 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 26 | 76 | 793340 | 0.0107 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 27 | 138 | 1561556 | 0.0210 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 28 | 138 | 1922592 | 0.0258 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 29 | 190 | 2253457 | 0.0303 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 30 | 253 | 2750277 | 0.0370 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 31 | 318 | 3004111 | 0.0404 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 32 | 392 | 4157073 | 0.0559 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 33 | 482 | 5170802 | 0.0695 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 34 | 651 | 6524440 | 0.0877 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 35 | 902 | 8714365 | 0.1171 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 36 | 1037 | 9457713 | 0.1271 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 37 | 1253 | 11146124 | 0.1498 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 38 | 1084 | 8877236 | 0.1193 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 39 | 603 | 4858841 | 0.0653 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 40 | 263 | 2138178 | 0.0287 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 41 | 69 | 522747 | 0.0070 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 42 | 17 | 176805 | 0.0024 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 43 | 3 | 24148 | 0.0003 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 44 | 2 | 5576 | 0.0001 |
| mac | 2021 | 4 | 27 | 27.7.j | TL | 30 | 1 | 1413 | 1.0000 |

Western horse mackerel Length frequencies 2022

| mac | 2022 | 1 | 27 | 27.4.a | TL | 26 | 9 | 33098 | 0.0014 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 27 | 52 | 340439 | 0.0144 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 28 | 44 | 336042 | 0.0142 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 29 | 54 | 396568 | 0.0168 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 30 | 59 | 388841 | 0.0165 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 31 | 62 | 440724 | 0.0187 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 32 | 68 | 505194 | 0.0214 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 33 | 84 | 482556 | 0.0205 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 34 | 159 | 1033678 | 0.0438 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 35 | 241 | 1684979 | 0.0714 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 36 | 392 | 2663419 | 0.1129 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 37 | 576 | 4055936 | 0.1719 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 38 | 626 | 4511764 | 0.1913 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 39 | 481 | 3759345 | 0.1594 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 40 | 256 | 1875541 | 0.0795 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 41 | 92 | 770072 | 0.0326 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 42 | 21 | 243746 | 0.0103 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 43 | 2 | 7439 | 0.0003 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 44 | 3 | 58790 | 0.0025 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 17 | 1 | 4150 | 0.0003 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 18 | 1 | 4150 | 0.0003 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 19 | 1 | 4150 | 0.0003 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 23 | 4 | 15424 | 0.0010 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 24 | 5 | 30313 | 0.0020 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 25 | 8 | 34891 | 0.0023 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 26 | 12 | 85246 | 0.0057 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 27 | 14 | 137644 | 0.0091 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 28 | 24 | 273784 | 0.0182 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 29 | 22 | 150308 | 0.0100 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 30 | 45 | 398239 | 0.0265 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 31 | 64 | 554722 | 0.0368 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 32 | 84 | 782862 | 0.0520 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 33 | 120 | 1156090 | 0.0768 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 34 | 95 | 882994 | 0.0586 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 35 | 115 | 1091725 | 0.0725 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 36 | 105 | 1096835 | 0.0728 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 37 | 209 | 2050525 | 0.1362 |


| mac | 2022 | 1 | 27 | 27.6.a | TL | 38 | 274 | 2565070 | 0.1704 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 39 | 214 | 2171710 | 0.1442 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 40 | 117 | 1193579 | 0.0793 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 41 | 29 | 282289 | 0.0187 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 42 | 7 | 89517 | 0.0059 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 20 | 3 | 34527 | 0.0034 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 21 | 5 | 57545 | 0.0056 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 22 | 7 | 80563 | 0.0079 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 23 | 7 | 80563 | 0.0079 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 24 | 4 | 46036 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 25 | 5 | 57545 | 0.0056 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 26 | 4 | 46036 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 27 | 4 | 35245 | 0.0035 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 28 | 3 | 31860 | 0.0031 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 29 | 7 | 84735 | 0.0083 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 30 | 25 | 216377 | 0.0212 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 31 | 24 | 196195 | 0.0192 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 32 | 31 | 296552 | 0.0291 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 33 | 45 | 315556 | 0.0310 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 34 | 72 | 507485 | 0.0498 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 35 | 162 | 1101891 | 0.1081 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 36 | 203 | 1318078 | 0.1293 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 37 | 201 | 1323574 | 0.1298 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 38 | 270 | 1866023 | 0.1831 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 39 | 211 | 1395078 | 0.1369 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 40 | 110 | 751821 | 0.0738 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 41 | 34 | 283483 | 0.0278 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 42 | 6 | 46318 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 43 | 3 | 9300 | 0.0009 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 44 | 1 | 10778 | 0.0011 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 25 | 2 | 6604 | 0.0129 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 26 | 1 | 3302 | 0.0064 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 27 | 3 | 14565 | 0.0284 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 28 | 6 | 20668 | 0.0403 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 29 | 10 | 46168 | 0.0899 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 30 | 20 | 74332 | 0.1448 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 31 | 16 | 73536 | 0.1432 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 32 | 13 | 59549 | 0.1160 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 33 | 15 | 55851 | 0.1088 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 34 | 15 | 44410 | 0.0865 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 35 | 16 | 30610 | 0.0596 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 36 | 17 | 21822 | 0.0425 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 37 | 11 | 26273 | 0.0512 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 38 | 8 | 18422 | 0.0359 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 39 | 6 | 12616 | 0.0246 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 40 | 3 | 4754 | 0.0093 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 34 | 4 | 20848 | 0.1319 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 35 | 5 | 31786 | 0.2011 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 36 | 5 | 28235 | 0.1786 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 37 | 6 | 27219 | 0.1722 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 38 | 6 | 28043 | 0.1774 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 39 | 3 | 14125 | 0.0894 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 40 | 1 | 2102 | 0.0133 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 41 | 1 | 2102 | 0.0133 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 42 | 1 | 3602 | 0.0228 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 26 | 1 | 3261 | 0.0010 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 29 | 9 | 102965 | 0.0330 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 30 | 68 | 847404 | 0.2719 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 31 | 65 | 831571 | 0.2668 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 32 | 53 | 568987 | 0.1825 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 33 | 36 | 409378 | 0.1313 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 34 | 11 | 109308 | 0.0351 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 35 | 4 | 54797 | 0.0176 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 36 | 4 | 67573 | 0.0217 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 37 | 5 | 74347 | 0.0239 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 38 | 3 | 47380 | 0.0152 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 32 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 34 | 5 | 4369 | 0.4168 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 35 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 36 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 37 | 2 | 1747 | 0.1667 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 38 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 40 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 31 | 1 | 4751 | 0.0457 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 32 | 2 | 17221 | 0.1656 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 33 | 1 | 12469 | 0.1199 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 34 | 1 | 9293 | 0.0894 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 35 | 3 | 14255 | 0.1371 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 36 | 2 | 9503 | 0.0914 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 37 | 1 | 13494 | 0.1298 |


| mac | 2022 | 3 | 27 | $27.7 . j$ | $T L$ | 38 | 2 | 18246 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

## 10 North Sea horse mackerel: detailed tables

North Sea horse mackerel Sampling overview


North Sea horse mackerel Length frequencies 2021


| mac | 2021 | 1 | 27 | 27.7.b | TL | 32 | 1 | 158 | 0.0003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 33 | 1 | 9116 | 0.0202 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 34 | 3 | 27349 | 0.0605 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 35 | 1 | 5243 | 0.0116 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 36 | 7 | 44463 | 0.0983 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 37 | 11 | 69334 | 0.1534 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 38 | 18 | 125023 | 0.2765 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 39 | 15 | 112427 | 0.2487 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 40 | 7 | 44475 | 0.0984 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 41 | 1 | 5243 | 0.0116 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 42 | 1 | 9116 | 0.0202 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 28 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 29 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 30 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 31 | 6 | 128253 | 0.0968 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 32 | 12 | 256507 | 0.1935 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 33 | 9 | 192380 | 0.1452 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 34 | 8 | 171004 | 0.1290 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 35 | 8 | 106877 | 0.0806 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 36 | 11 | 149629 | 0.1129 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 37 | 9 | 85502 | 0.0645 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 38 | 7 | 149629 | 0.1129 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 39 | 3 | 21375 | 0.0161 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 32 | 1 | 19069 | 0.0216 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 33 | 3 | 54106 | 0.0613 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 34 | 7 | 77064 | 0.0873 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 35 | 9 | 115011 | 0.1303 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 36 | 14 | 180151 | 0.2040 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 37 | 11 | 161067 | 0.1824 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 38 | 8 | 99500 | 0.1127 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 39 | 6 | 94816 | 0.1074 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 40 | 4 | 66247 | 0.0750 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 42 | 1 | 15967 | 0.0181 |
| mac | 2021 | 2 | 27 | 27.7.b | TL | 38 | 1 | 6127 | 0.5000 |
| mac | 2021 | 2 | 27 | 27.7.b | TL | 40 | 1 | 6127 | 0.5000 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 24 | 1 | 9442 | 0.0054 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 25 | 5 | 43353 | 0.0249 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 26 | 12 | 97836 | 0.0561 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 27 | 2 | 10111 | 0.0058 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 28 | 2 | 9810 | 0.0056 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 29 | 10 | 98209 | 0.0563 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 30 | 23 | 131613 | 0.0755 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 31 | 16 | 127653 | 0.0732 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 32 | 28 | 166532 | 0.0955 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 33 | 23 | 149258 | 0.0856 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 34 | 26 | 179596 | 0.1030 |
| mac | 2021 | 3 | 27 | 27.4.a | tL | 35 | 23 | 173859 | 0.0998 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 36 | 25 | 179053 | 0.1027 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 37 | 24 | 155003 | 0.0889 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 38 | 19 | 123471 | 0.0708 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 39 | 11 | 74916 | 0.0430 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 40 | 2 | 13203 | 0.0076 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 14 | 4 | 14157 | 0.0952 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 15 | 17 | 60167 | 0.4048 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 16 | 14 | 49549 | 0.3333 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 17 | 5 | 17696 | 0.1190 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 18 | 2 | 7078 | 0.0476 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 28 | 9 | 112126 | 0.1232 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 29 | 13 | 203625 | 0.2238 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 30 | 11 | 141073 | 0.1550 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 31 | 7 | 76669 | 0.0843 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 32 | 4 | 25451 | 0.0280 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 33 | 6 | 23804 | 0.0262 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 34 | 4 | 35446 | 0.0390 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 35 | 13 | 115133 | 0.1265 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 36 | 8 | 64557 | 0.0710 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 37 | 2 | 23926 | 0.0263 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 38 | 7 | 66987 | 0.0736 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 39 | 2 | 17577 | 0.0193 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 40 | 1 | 3499 | 0.0038 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 27 | 1 | 2840 | 0.0020 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 28 | 16 | 151373 | 0.1086 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 29 | 11 | 47096 | 0.0338 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 30 | 9 | 53242 | 0.0382 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 31 | 12 | 62300 | 0.0447 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 32 | 6 | 27933 | 0.0200 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 33 | 18 | 116680 | 0.0837 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 34 | 14 | 154071 | 0.1105 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 35 | 29 | 191132 | 0.1371 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 36 | 32 | 259037 | 0.1858 |


| mac | 2021 | 3 | 27 | 27.7.j | TL | 37 | 32 | 150305 | 0.1078 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 38 | 15 | 80969 | 0.0581 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 39 | 7 | 70033 | 0.0502 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 40 | 5 | 16515 | 0.0118 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 41 | 1 | 10581 | 0.0076 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 35 | 4 | 1061 | 0.0337 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 36 | 4 | 1141 | 0.0362 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 37 | 14 | 18943 | 0.6011 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 38 | 19 | 5141 | 0.1631 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 39 | 10 | 2534 | 0.0804 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 40 | 8 | 2202 | 0.0699 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 41 | 2 | 490 | 0.0155 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 23 | 1 | 11982 | 0.0002 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 24 | 9 | 108983 | 0.0015 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 25 | 21 | 231669 | 0.0031 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 26 | 76 | 793340 | 0.0107 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 27 | 138 | 1561556 | 0.0210 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 28 | 138 | 1922592 | 0.0258 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 29 | 190 | 2253457 | 0.0303 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 30 | 253 | 2750277 | 0.0370 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 31 | 318 | 3004111 | 0.0404 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 32 | 392 | 4157073 | 0.0559 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 33 | 482 | 5170802 | 0.0695 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 34 | 651 | 6524440 | 0.0877 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 35 | 902 | 8714365 | 0.1171 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 36 | 1037 | 9457713 | 0.1271 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 37 | 1253 | 11146124 | 0.1498 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 38 | 1084 | 8877236 | 0.1193 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 39 | 603 | 4858841 | 0.0653 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 40 | 263 | 2138178 | 0.0287 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 41 | 69 | 522747 | 0.0070 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 42 | 17 | 176805 | 0.0024 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 43 | 3 | 24148 | 0.0003 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 44 | 2 | 5576 | 0.0001 |
| mac | 2021 | 4 | 27 | 27.7.j | TL | 30 | 1 | 1413 | 1.0000 |

North Sea horse mackerel Length frequencies 2022

| mac | 2022 | 1 | 27 | 27.4.a | TL | 26 | 9 | 33098 | 0.0014 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 27 | 52 | 340439 | 0.0144 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 28 | 44 | 336042 | 0.0142 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 29 | 54 | 396568 | 0.0168 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 30 | 59 | 388841 | 0.0165 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 31 | 62 | 440724 | 0.0187 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 32 | 68 | 505194 | 0.0214 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 33 | 84 | 482556 | 0.0205 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 34 | 159 | 1033678 | 0.0438 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 35 | 241 | 1684979 | 0.0714 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 36 | 392 | 2663419 | 0.1129 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 37 | 576 | 4055936 | 0.1719 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 38 | 626 | 4511764 | 0.1913 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 39 | 481 | 3759345 | 0.1594 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 40 | 256 | 1875541 | 0.0795 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 41 | 92 | 770072 | 0.0326 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 42 | 21 | 243746 | 0.0103 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 43 | 2 | 7439 | 0.0003 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 44 | 3 | 58790 | 0.0025 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 17 | 1 | 4150 | 0.0003 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 18 | 1 | 4150 | 0.0003 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 19 | 1 | 4150 | 0.0003 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 23 | 4 | 15424 | 0.0010 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 24 | 5 | 30313 | 0.0020 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 25 | 8 | 34891 | 0.0023 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 26 | 12 | 85246 | 0.0057 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 27 | 14 | 137644 | 0.0091 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 28 | 24 | 273784 | 0.0182 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 29 | 22 | 150308 | 0.0100 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 30 | 45 | 398239 | 0.0265 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 31 | 64 | 554722 | 0.0368 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 32 | 84 | 782862 | 0.0520 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 33 | 120 | 1156090 | 0.0768 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 34 | 95 | 882994 | 0.0586 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 35 | 115 | 1091725 | 0.0725 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 36 | 105 | 1096835 | 0.0728 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 37 | 209 | 2050525 | 0.1362 |


| mac | 2022 | 1 | 27 | 27.6.a | TL | 38 | 274 | 2565070 | 0.1704 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 39 | 214 | 2171710 | 0.1442 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 40 | 117 | 1193579 | 0.0793 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 41 | 29 | 282289 | 0.0187 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 42 | 7 | 89517 | 0.0059 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 20 | 3 | 34527 | 0.0034 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 21 | 5 | 57545 | 0.0056 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 22 | 7 | 80563 | 0.0079 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 23 | 7 | 80563 | 0.0079 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 24 | 4 | 46036 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 25 | 5 | 57545 | 0.0056 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 26 | 4 | 46036 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 27 | 4 | 35245 | 0.0035 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 28 | 3 | 31860 | 0.0031 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 29 | 7 | 84735 | 0.0083 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 30 | 25 | 216377 | 0.0212 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 31 | 24 | 196195 | 0.0192 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 32 | 31 | 296552 | 0.0291 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 33 | 45 | 315556 | 0.0310 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 34 | 72 | 507485 | 0.0498 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 35 | 162 | 1101891 | 0.1081 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 36 | 203 | 1318078 | 0.1293 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 37 | 201 | 1323574 | 0.1298 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 38 | 270 | 1866023 | 0.1831 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 39 | 211 | 1395078 | 0.1369 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 40 | 110 | 751821 | 0.0738 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 41 | 34 | 283483 | 0.0278 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 42 | 6 | 46318 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 43 | 3 | 9300 | 0.0009 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 44 | 1 | 10778 | 0.0011 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 25 | 2 | 6604 | 0.0129 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 26 | 1 | 3302 | 0.0064 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 27 | 3 | 14565 | 0.0284 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 28 | 6 | 20668 | 0.0403 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 29 | 10 | 46168 | 0.0899 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 30 | 20 | 74332 | 0.1448 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 31 | 16 | 73536 | 0.1432 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 32 | 13 | 59549 | 0.1160 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 33 | 15 | 55851 | 0.1088 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 34 | 15 | 44410 | 0.0865 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 35 | 16 | 30610 | 0.0596 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 36 | 17 | 21822 | 0.0425 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 37 | 11 | 26273 | 0.0512 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 38 | 8 | 18422 | 0.0359 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 39 | 6 | 12616 | 0.0246 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 40 | 3 | 4754 | 0.0093 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 34 | 4 | 20848 | 0.1319 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 35 | 5 | 31786 | 0.2011 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 36 | 5 | 28235 | 0.1786 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 37 | 6 | 27219 | 0.1722 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 38 | 6 | 28043 | 0.1774 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 39 | 3 | 14125 | 0.0894 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 40 | 1 | 2102 | 0.0133 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 41 | 1 | 2102 | 0.0133 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 42 | 1 | 3602 | 0.0228 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 26 | 1 | 3261 | 0.0010 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 29 | 9 | 102965 | 0.0330 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 30 | 68 | 847404 | 0.2719 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 31 | 65 | 831571 | 0.2668 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 32 | 53 | 568987 | 0.1825 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 33 | 36 | 409378 | 0.1313 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 34 | 11 | 109308 | 0.0351 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 35 | 4 | 54797 | 0.0176 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 36 | 4 | 67573 | 0.0217 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 37 | 5 | 74347 | 0.0239 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 38 | 3 | 47380 | 0.0152 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 32 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 34 | 5 | 4369 | 0.4168 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 35 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 36 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 37 | 2 | 1747 | 0.1667 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 38 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 40 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 31 | 1 | 4751 | 0.0457 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 32 | 2 | 17221 | 0.1656 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 33 | 1 | 12469 | 0.1199 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 34 | 1 | 9293 | 0.0894 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 35 | 3 | 14255 | 0.1371 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 36 | 2 | 9503 | 0.0914 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 37 | 1 | 13494 | 0.1298 |


| $\operatorname{mac}$ | 2022 | 3 | $2727.7 . j$ | TL | 38 | 218246 | 0.1755 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| mac | 2022 | 3 | $2727.7 . j$ | TL | 39 | 14751 | 0.0457 |

## 11 Blue whiting: detailed tables

Blue whiting Sampling overview


Blue whiting Length frequencies 2021


| mac | 2021 | 1 | 27 | 27.7.b | TL | 32 | 1 | 158 | 0.0003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 33 | 1 | 9116 | 0.0202 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 34 | 3 | 27349 | 0.0605 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 35 | 1 | 5243 | 0.0116 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 36 | 7 | 44463 | 0.0983 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 37 | 11 | 69334 | 0.1534 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 38 | 18 | 125023 | 0.2765 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 39 | 15 | 112427 | 0.2487 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 40 | 7 | 44475 | 0.0984 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 41 | 1 | 5243 | 0.0116 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 42 | 1 | 9116 | 0.0202 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 28 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 29 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 30 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 31 | 6 | 128253 | 0.0968 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 32 | 12 | 256507 | 0.1935 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 33 | 9 | 192380 | 0.1452 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 34 | 8 | 171004 | 0.1290 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 35 | 8 | 106877 | 0.0806 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 36 | 11 | 149629 | 0.1129 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 37 | 9 | 85502 | 0.0645 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 38 | 7 | 149629 | 0.1129 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 39 | 3 | 21375 | 0.0161 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 32 | 1 | 19069 | 0.0216 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 33 | 3 | 54106 | 0.0613 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 34 | 7 | 77064 | 0.0873 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 35 | 9 | 115011 | 0.1303 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 36 | 14 | 180151 | 0.2040 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 37 | 11 | 161067 | 0.1824 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 38 | 8 | 99500 | 0.1127 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 39 | 6 | 94816 | 0.1074 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 40 | 4 | 66247 | 0.0750 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 42 | 1 | 15967 | 0.0181 |
| mac | 2021 | 2 | 27 | 27.7.b | TL | 38 | 1 | 6127 | 0.5000 |
| mac | 2021 | 2 | 27 | 27.7.b | TL | 40 | 1 | 6127 | 0.5000 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 24 | 1 | 9442 | 0.0054 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 25 | 5 | 43353 | 0.0249 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 26 | 12 | 97836 | 0.0561 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 27 | 2 | 10111 | 0.0058 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 28 | 2 | 9810 | 0.0056 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 29 | 10 | 98209 | 0.0563 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 30 | 23 | 131613 | 0.0755 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 31 | 16 | 127653 | 0.0732 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 32 | 28 | 166532 | 0.0955 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 33 | 23 | 149258 | 0.0856 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 34 | 26 | 179596 | 0.1030 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 35 | 23 | 173859 | 0.0998 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 36 | 25 | 179053 | 0.1027 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 37 | 24 | 155003 | 0.0889 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 38 | 19 | 123471 | 0.0708 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 39 | 11 | 74916 | 0.0430 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 40 | 2 | 13203 | 0.0076 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 14 | 4 | 14157 | 0.0952 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 15 | 17 | 60167 | 0.4048 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 16 | 14 | 49549 | 0.3333 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 17 | 5 | 17696 | 0.1190 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 18 | 2 | 7078 | 0.0476 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 28 | 9 | 112126 | 0.1232 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 29 | 13 | 203625 | 0.2238 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 30 | 11 | 141073 | 0.1550 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 31 | 7 | 76669 | 0.0843 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 32 | 4 | 25451 | 0.0280 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 33 | 6 | 23804 | 0.0262 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 34 | 4 | 35446 | 0.0390 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 35 | 13 | 115133 | 0.1265 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 36 | 8 | 64557 | 0.0710 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 37 | 2 | 23926 | 0.0263 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 38 | 7 | 66987 | 0.0736 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 39 | 2 | 17577 | 0.0193 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 40 | 1 | 3499 | 0.0038 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 27 | 1 | 2840 | 0.0020 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 28 | 16 | 151373 | 0.1086 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 29 | 11 | 47096 | 0.0338 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 30 | 9 | 53242 | 0.0382 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 31 | 12 | 62300 | 0.0447 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 32 | 6 | 27933 | 0.0200 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 33 | 18 | 116680 | 0.0837 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 34 | 14 | 154071 | 0.1105 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 35 | 29 | 191132 | 0.1371 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 36 | 32 | 259037 | 0.1858 |


| mac | 2021 | 3 | 27 | 27.7.j | TL | 37 | 32 | 150305 | 0.1078 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 38 | 15 | 80969 | 0.0581 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 39 | 7 | 70033 | 0.0502 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 40 | 5 | 16515 | 0.0118 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 41 | 1 | 10581 | 0.0076 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 35 | 4 | 1061 | 0.0337 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 36 | 4 | 1141 | 0.0362 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 37 | 14 | 18943 | 0.6011 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 38 | 19 | 5141 | 0.1631 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 39 | 10 | 2534 | 0.0804 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 40 | 8 | 2202 | 0.0699 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 41 | 2 | 490 | 0.0155 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 23 | 1 | 11982 | 0.0002 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 24 | 9 | 108983 | 0.0015 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 25 | 21 | 231669 | 0.0031 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 26 | 76 | 793340 | 0.0107 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 27 | 138 | 1561556 | 0.0210 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 28 | 138 | 1922592 | 0.0258 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 29 | 190 | 2253457 | 0.0303 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 30 | 253 | 2750277 | 0.0370 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 31 | 318 | 3004111 | 0.0404 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 32 | 392 | 4157073 | 0.0559 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 33 | 482 | 5170802 | 0.0695 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 34 | 651 | 6524440 | 0.0877 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 35 | 902 | 8714365 | 0.1171 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 36 | 1037 | 9457713 | 0.1271 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 37 | 1253 | 11146124 | 0.1498 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 38 | 1084 | 8877236 | 0.1193 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 39 | 603 | 4858841 | 0.0653 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 40 | 263 | 2138178 | 0.0287 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 41 | 69 | 522747 | 0.0070 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 42 | 17 | 176805 | 0.0024 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 43 | 3 | 24148 | 0.0003 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 44 | 2 | 5576 | 0.0001 |
| mac | 2021 | 4 | 27 | 27.7.j | TL | 30 | 1 | 1413 | 1.0000 |

Blue whiting Length frequencies 2022


| mac | 2022 | 1 | 27 | 27.6.a | TL | 38 | 274 | 2565070 | 0.1704 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 39 | 214 | 2171710 | 0.1442 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 40 | 117 | 1193579 | 0.0793 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 41 | 29 | 282289 | 0.0187 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 42 | 7 | 89517 | 0.0059 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 20 | 3 | 34527 | 0.0034 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 21 | 5 | 57545 | 0.0056 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 22 | 7 | 80563 | 0.0079 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 23 | 7 | 80563 | 0.0079 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 24 | 4 | 46036 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 25 | 5 | 57545 | 0.0056 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 26 | 4 | 46036 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 27 | 4 | 35245 | 0.0035 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 28 | 3 | 31860 | 0.0031 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 29 | 7 | 84735 | 0.0083 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 30 | 25 | 216377 | 0.0212 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 31 | 24 | 196195 | 0.0192 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 32 | 31 | 296552 | 0.0291 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 33 | 45 | 315556 | 0.0310 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 34 | 72 | 507485 | 0.0498 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 35 | 162 | 1101891 | 0.1081 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 36 | 203 | 1318078 | 0.1293 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 37 | 201 | 1323574 | 0.1298 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 38 | 270 | 1866023 | 0.1831 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 39 | 211 | 1395078 | 0.1369 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 40 | 110 | 751821 | 0.0738 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 41 | 34 | 283483 | 0.0278 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 42 | 6 | 46318 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 43 | 3 | 9300 | 0.0009 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 44 | 1 | 10778 | 0.0011 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 25 | 2 | 6604 | 0.0129 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 26 | 1 | 3302 | 0.0064 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 27 | 3 | 14565 | 0.0284 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 28 | 6 | 20668 | 0.0403 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 29 | 10 | 46168 | 0.0899 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 30 | 20 | 74332 | 0.1448 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 31 | 16 | 73536 | 0.1432 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 32 | 13 | 59549 | 0.1160 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 33 | 15 | 55851 | 0.1088 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 34 | 15 | 44410 | 0.0865 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 35 | 16 | 30610 | 0.0596 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 36 | 17 | 21822 | 0.0425 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 37 | 11 | 26273 | 0.0512 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 38 | 8 | 18422 | 0.0359 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 39 | 6 | 12616 | 0.0246 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 40 | 3 | 4754 | 0.0093 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 34 | 4 | 20848 | 0.1319 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 35 | 5 | 31786 | 0.2011 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 36 | 5 | 28235 | 0.1786 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 37 | 6 | 27219 | 0.1722 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 38 | 6 | 28043 | 0.1774 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 39 | 3 | 14125 | 0.0894 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 40 | 1 | 2102 | 0.0133 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 41 | 1 | 2102 | 0.0133 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 42 | 1 | 3602 | 0.0228 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 26 | 1 | 3261 | 0.0010 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 29 | 9 | 102965 | 0.0330 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 30 | 68 | 847404 | 0.2719 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 31 | 65 | 831571 | 0.2668 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 32 | 53 | 568987 | 0.1825 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 33 | 36 | 409378 | 0.1313 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 34 | 11 | 109308 | 0.0351 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 35 | 4 | 54797 | 0.0176 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 36 | 4 | 67573 | 0.0217 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 37 | 5 | 74347 | 0.0239 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 38 | 3 | 47380 | 0.0152 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 32 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 34 | 5 | 4369 | 0.4168 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 35 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 36 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 37 | 2 | 1747 | 0.1667 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 38 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 40 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 31 | 1 | 4751 | 0.0457 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 32 | 2 | 17221 | 0.1656 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 33 | 1 | 12469 | 0.1199 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 34 | 1 | 9293 | 0.0894 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 35 | 3 | 14255 | 0.1371 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 36 | 2 | 9503 | 0.0914 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 37 | 1 | 13494 | 0.1298 |


| mac | 2022 | 3 | 27 | $27.7 . j$ | $T L$ | 38 | 2 | 18246 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

## 12 Atlanto-scandian herring: detailed tables

Atlanto-scandian herring Sampling overview


Atlanto-scandian herring Length frequencies 2021


| mac | 2021 | 1 | 27 | 27.7.b | TL | 32 | 1 | 158 | 0.0003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 33 | 1 | 9116 | 0.0202 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 34 | 3 | 27349 | 0.0605 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 35 | 1 | 5243 | 0.0116 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 36 | 7 | 44463 | 0.0983 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 37 | 11 | 69334 | 0.1534 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 38 | 18 | 125023 | 0.2765 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 39 | 15 | 112427 | 0.2487 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 40 | 7 | 44475 | 0.0984 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 41 | 1 | 5243 | 0.0116 |
| mac | 2021 | 1 | 27 | 27.7.b | TL | 42 | 1 | 9116 | 0.0202 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 28 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 29 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 30 | 1 | 21375 | 0.0161 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 31 | 6 | 128253 | 0.0968 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 32 | 12 | 256507 | 0.1935 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 33 | 9 | 192380 | 0.1452 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 34 | 8 | 171004 | 0.1290 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 35 | 8 | 106877 | 0.0806 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 36 | 11 | 149629 | 0.1129 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 37 | 9 | 85502 | 0.0645 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 38 | 7 | 149629 | 0.1129 |
| mac | 2021 | 1 | 27 | 27.7.j | TL | 39 | 3 | 21375 | 0.0161 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 32 | 1 | 19069 | 0.0216 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 33 | 3 | 54106 | 0.0613 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 34 | 7 | 77064 | 0.0873 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 35 | 9 | 115011 | 0.1303 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 36 | 14 | 180151 | 0.2040 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 37 | 11 | 161067 | 0.1824 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 38 | 8 | 99500 | 0.1127 |
| mac | 2021 | 2 | 27 | 27.6.a | tL | 39 | 6 | 94816 | 0.1074 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 40 | 4 | 66247 | 0.0750 |
| mac | 2021 | 2 | 27 | 27.6.a | TL | 42 | 1 | 15967 | 0.0181 |
| mac | 2021 | 2 | 27 | 27.7.b | TL | 38 | 1 | 6127 | 0.5000 |
| mac | 2021 | 2 | 27 | 27.7.b | TL | 40 | 1 | 6127 | 0.5000 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 24 | 1 | 9442 | 0.0054 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 25 | 5 | 43353 | 0.0249 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 26 | 12 | 97836 | 0.0561 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 27 | 2 | 10111 | 0.0058 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 28 | 2 | 9810 | 0.0056 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 29 | 10 | 98209 | 0.0563 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 30 | 23 | 131613 | 0.0755 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 31 | 16 | 127653 | 0.0732 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 32 | 28 | 166532 | 0.0955 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 33 | 23 | 149258 | 0.0856 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 34 | 26 | 179596 | 0.1030 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 35 | 23 | 173859 | 0.0998 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 36 | 25 | 179053 | 0.1027 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 37 | 24 | 155003 | 0.0889 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 38 | 19 | 123471 | 0.0708 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 39 | 11 | 74916 | 0.0430 |
| mac | 2021 | 3 | 27 | 27.4.a | TL | 40 | 2 | 13203 | 0.0076 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 14 | 4 | 14157 | 0.0952 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 15 | 17 | 60167 | 0.4048 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 16 | 14 | 49549 | 0.3333 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 17 | 5 | 17696 | 0.1190 |
| mac | 2021 | 3 | 27 | 27.6.a | TL | 18 | 2 | 7078 | 0.0476 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 28 | 9 | 112126 | 0.1232 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 29 | 13 | 203625 | 0.2238 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 30 | 11 | 141073 | 0.1550 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 31 | 7 | 76669 | 0.0843 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 32 | 4 | 25451 | 0.0280 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 33 | 6 | 23804 | 0.0262 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 34 | 4 | 35446 | 0.0390 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 35 | 13 | 115133 | 0.1265 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 36 | 8 | 64557 | 0.0710 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 37 | 2 | 23926 | 0.0263 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 38 | 7 | 66987 | 0.0736 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 39 | 2 | 17577 | 0.0193 |
| mac | 2021 | 3 | 27 | 27.7.b | TL | 40 | 1 | 3499 | 0.0038 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 27 | 1 | 2840 | 0.0020 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 28 | 16 | 151373 | 0.1086 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 29 | 11 | 47096 | 0.0338 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 30 | 9 | 53242 | 0.0382 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 31 | 12 | 62300 | 0.0447 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 32 | 6 | 27933 | 0.0200 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 33 | 18 | 116680 | 0.0837 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 34 | 14 | 154071 | 0.1105 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 35 | 29 | 191132 | 0.1371 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 36 | 32 | 259037 | 0.1858 |


| mac | 2021 | 3 | 27 | 27.7.j | TL | 37 | 32 | 150305 | 0.1078 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 38 | 15 | 80969 | 0.0581 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 39 | 7 | 70033 | 0.0502 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 40 | 5 | 16515 | 0.0118 |
| mac | 2021 | 3 | 27 | 27.7.j | TL | 41 | 1 | 10581 | 0.0076 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 35 | 4 | 1061 | 0.0337 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 36 | 4 | 1141 | 0.0362 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 37 | 14 | 18943 | 0.6011 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 38 | 19 | 5141 | 0.1631 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 39 | 10 | 2534 | 0.0804 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 40 | 8 | 2202 | 0.0699 |
| mac | 2021 | 4 | 27 | 27.2.a | TL | 41 | 2 | 490 | 0.0155 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 23 | 1 | 11982 | 0.0002 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 24 | 9 | 108983 | 0.0015 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 25 | 21 | 231669 | 0.0031 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 26 | 76 | 793340 | 0.0107 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 27 | 138 | 1561556 | 0.0210 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 28 | 138 | 1922592 | 0.0258 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 29 | 190 | 2253457 | 0.0303 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 30 | 253 | 2750277 | 0.0370 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 31 | 318 | 3004111 | 0.0404 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 32 | 392 | 4157073 | 0.0559 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 33 | 482 | 5170802 | 0.0695 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 34 | 651 | 6524440 | 0.0877 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 35 | 902 | 8714365 | 0.1171 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 36 | 1037 | 9457713 | 0.1271 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 37 | 1253 | 11146124 | 0.1498 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 38 | 1084 | 8877236 | 0.1193 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 39 | 603 | 4858841 | 0.0653 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 40 | 263 | 2138178 | 0.0287 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 41 | 69 | 522747 | 0.0070 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 42 | 17 | 176805 | 0.0024 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 43 | 3 | 24148 | 0.0003 |
| mac | 2021 | 4 | 27 | 27.4.a | TL | 44 | 2 | 5576 | 0.0001 |
| mac | 2021 | 4 | 27 | 27.7.j | TL | 30 | 1 | 1413 | 1.0000 |

Atlanto-scandian herring Length frequencies 2022

| mac | 2022 | 1 | 27 | 27.4.a | TL | 26 | 9 | 33098 | 0.0014 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 27 | 52 | 340439 | 0.0144 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 28 | 44 | 336042 | 0.0142 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 29 | 54 | 396568 | 0.0168 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 30 | 59 | 388841 | 0.0165 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 31 | 62 | 440724 | 0.0187 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 32 | 68 | 505194 | 0.0214 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 33 | 84 | 482556 | 0.0205 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 34 | 159 | 1033678 | 0.0438 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 35 | 241 | 1684979 | 0.0714 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 36 | 392 | 2663419 | 0.1129 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 37 | 576 | 4055936 | 0.1719 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 38 | 626 | 4511764 | 0.1913 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 39 | 481 | 3759345 | 0.1594 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 40 | 256 | 1875541 | 0.0795 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 41 | 92 | 770072 | 0.0326 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 42 | 21 | 243746 | 0.0103 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 43 | 2 | 7439 | 0.0003 |
| mac | 2022 | 1 | 27 | 27.4.a | TL | 44 | 3 | 58790 | 0.0025 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 17 | 1 | 4150 | 0.0003 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 18 | 1 | 4150 | 0.0003 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 19 | 1 | 4150 | 0.0003 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 23 | 4 | 15424 | 0.0010 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 24 | 5 | 30313 | 0.0020 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 25 | 8 | 34891 | 0.0023 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 26 | 12 | 85246 | 0.0057 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 27 | 14 | 137644 | 0.0091 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 28 | 24 | 273784 | 0.0182 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 29 | 22 | 150308 | 0.0100 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 30 | 45 | 398239 | 0.0265 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 31 | 64 | 554722 | 0.0368 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 32 | 84 | 782862 | 0.0520 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 33 | 120 | 1156090 | 0.0768 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 34 | 95 | 882994 | 0.0586 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 35 | 115 | 1091725 | 0.0725 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 36 | 105 | 1096835 | 0.0728 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 37 | 209 | 2050525 | 0.1362 |


| mac | 2022 | 1 | 27 | 27.6.a | TL | 38 | 274 | 2565070 | 0.1704 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 39 | 214 | 2171710 | 0.1442 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 40 | 117 | 1193579 | 0.0793 |
| mac | 2022 | 1 | 27 | 27.6.a | tL | 41 | 29 | 282289 | 0.0187 |
| mac | 2022 | 1 | 27 | 27.6.a | TL | 42 | 7 | 89517 | 0.0059 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 20 | 3 | 34527 | 0.0034 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 21 | 5 | 57545 | 0.0056 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 22 | 7 | 80563 | 0.0079 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 23 | 7 | 80563 | 0.0079 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 24 | 4 | 46036 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 25 | 5 | 57545 | 0.0056 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 26 | 4 | 46036 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 27 | 4 | 35245 | 0.0035 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 28 | 3 | 31860 | 0.0031 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 29 | 7 | 84735 | 0.0083 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 30 | 25 | 216377 | 0.0212 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 31 | 24 | 196195 | 0.0192 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 32 | 31 | 296552 | 0.0291 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 33 | 45 | 315556 | 0.0310 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 34 | 72 | 507485 | 0.0498 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 35 | 162 | 1101891 | 0.1081 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 36 | 203 | 1318078 | 0.1293 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 37 | 201 | 1323574 | 0.1298 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 38 | 270 | 1866023 | 0.1831 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 39 | 211 | 1395078 | 0.1369 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 40 | 110 | 751821 | 0.0738 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 41 | 34 | 283483 | 0.0278 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 42 | 6 | 46318 | 0.0045 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 43 | 3 | 9300 | 0.0009 |
| mac | 2022 | 1 | 27 | 27.7.b | TL | 44 | 1 | 10778 | 0.0011 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 25 | 2 | 6604 | 0.0129 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 26 | 1 | 3302 | 0.0064 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 27 | 3 | 14565 | 0.0284 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 28 | 6 | 20668 | 0.0403 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 29 | 10 | 46168 | 0.0899 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 30 | 20 | 74332 | 0.1448 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 31 | 16 | 73536 | 0.1432 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 32 | 13 | 59549 | 0.1160 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 33 | 15 | 55851 | 0.1088 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 34 | 15 | 44410 | 0.0865 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 35 | 16 | 30610 | 0.0596 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 36 | 17 | 21822 | 0.0425 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 37 | 11 | 26273 | 0.0512 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 38 | 8 | 18422 | 0.0359 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 39 | 6 | 12616 | 0.0246 |
| mac | 2022 | 1 | 27 | 27.7.j | TL | 40 | 3 | 4754 | 0.0093 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 34 | 4 | 20848 | 0.1319 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 35 | 5 | 31786 | 0.2011 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 36 | 5 | 28235 | 0.1786 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 37 | 6 | 27219 | 0.1722 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 38 | 6 | 28043 | 0.1774 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 39 | 3 | 14125 | 0.0894 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 40 | 1 | 2102 | 0.0133 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 41 | 1 | 2102 | 0.0133 |
| mac | 2022 | 2 | 27 | 27.6.a | TL | 42 | 1 | 3602 | 0.0228 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 26 | 1 | 3261 | 0.0010 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 29 | 9 | 102965 | 0.0330 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 30 | 68 | 847404 | 0.2719 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 31 | 65 | 831571 | 0.2668 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 32 | 53 | 568987 | 0.1825 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 33 | 36 | 409378 | 0.1313 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 34 | 11 | 109308 | 0.0351 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 35 | 4 | 54797 | 0.0176 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 36 | 4 | 67573 | 0.0217 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 37 | 5 | 74347 | 0.0239 |
| mac | 2022 | 3 | 27 | 27.4.a | TL | 38 | 3 | 47380 | 0.0152 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 32 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 34 | 5 | 4369 | 0.4168 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 35 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 36 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 37 | 2 | 1747 | 0.1667 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 38 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.b | TL | 40 | 1 | 873 | 0.0833 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 31 | 1 | 4751 | 0.0457 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 32 | 2 | 17221 | 0.1656 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 33 | 1 | 12469 | 0.1199 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 34 | 1 | 9293 | 0.0894 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 35 | 3 | 14255 | 0.1371 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 36 | 2 | 9503 | 0.0914 |
| mac | 2022 | 3 | 27 | 27.7.j | TL | 37 | 1 | 13494 | 0.1298 |


| mac | 2022 | 3 | 27 | $27.7 . j$ | $T L$ | 38 | 2 | 18246 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

# North Sea mackerel daily egg production and spawning stock biomass estimation in 2021 

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## Introduction

The North Sea Mackerel Egg Survey (NSMEGS) is designed to estimate the spawning stock biomass (SSB) of mackerel of the North Sea spawning component of the Northeast-Atlantic stock on a triennial basis. Up to and including 2017 this was undertaken utilizing the annual egg production method (AEPM). This method estimates and combines total annual egg production (TAEP), realized fecundity per gram female, and sex (male to female) ratio to calculate SSB.

Spatial and temporal coverage in the North Sea was reduced with the withdrawal of Norway from the NSMEGS in 2014, with the Netherlands left as the sole survey participant in 2015 and 2017. In 2020 Denmark was recruited as a new participant for the NSMEGS, but due to the Covid-19 pandemic and the implementation of associated measures it was not possible to complete the survey in 2020. After consultation with WGMEGS chairs and the mackerel assessor it was agreed to postpone the survey to 2021.

An issue for the NSMEGS is that since 1982 it has been impossible to collect and sample pre-spawning mackerel, which are necessary in order to estimate the potential fecundity. For SSB estimation using the AEPM, the realized fecundity value used was from the 1982 estimate (Iversen and Adoff, 1983). Also, the planned coverage for 2020 (which was postponed to 2021) of the mackerel spawning in the North Sea, both temporally and spatially, was far from ideal for the Annual Egg Production Method (AEPM; ICES 2018). Consequently, WGMEGS discussed utilizing the Daily Egg Production Method (DEPM) for the NSMEGS. The DEPM requires only one full sweep, in a short time period, of the entire mackerel spawning area, and preferably during peak spawning time, in order to estimate the Daily Egg Production (DEP). A disadvantage of the DEPM is that it requires many more mackerel ovary samples to be collected to estimate batch fecundity and spawning fraction. Considering the pros and cons of the AEPM and DEPM for the NSMEGS, at the 2018 meeting WGMEGS decided to switch to the DEPM for the NSMEGS in 2020 ( which was then postponed to 2021; ICES 2018).

## Survey

In 2021 Netherlands and Denmark conducted the NSMEGS. Whilst completing an exploratory egg survey along the Norwegian Sea, similar to those in 2017 and 2018 to the west of Faroes, Scotland was also able to contribute several additional survey transects within the Northern North Sea that were then incorporated into the 2021 NSMEGS dataset.

During 2021, Covid 19 measures continued to pose significant challenges that impeded the execution of the survey plan. The Dutch vessel was not permitted to enter foreign harbours during survey breaks, instead being required to undertake the long steam back to a Dutch harbour. As a consequence the Netherlands was unable to sample the most northerly transect. However, Scotland was able to complete this transect during their exploratory survey.

The samples were collected and analysed according to the WGMEGS manuals (ICES 2019a, 2019b). The Netherlands and Scotland sampled eggs with a Gulf VII plankton sampler while Denmark used a Nackthai
sampler. The Netherlands and Denmark utilised a $500 \mu \mathrm{~m}$ plankton net whereas Scotland used a $250 \mu \mathrm{~m}$ plankton net. At each station a double oblique haul was performed from the surface to 5 m above the bottom, a maximum depth of 200 m , or 20 m below the thermocline in case of stratification of the water column. Temperature and salinity were measured during the haul with a CTD mounted on top of the plankton sampler. Electronic flowmeters were mounted on the plankton sampler to monitor flow.

The NSMEGS was carried out from $25^{\text {th }}$ May to $12^{\text {th }}$ J une (Table 1). During this period the spawning area between $53^{\circ} \mathrm{N}$ and $62^{\circ} \mathrm{N}$ was surveyed once, receiving a single coverage (Fig. 1). The survey is designed to cover the entire spawning area with samples collected every half ICES statistical rectangle (ICES, 2014). In total 294 plankton stations were sampled. In 26 of the half rectangles more than one plankton sample was collected (Fig. 1a). These rectangles were used to estimate the CV and variance of the DEP. On each transect at least one pelagic trawl haul was performed for the collection of mackerel adult samples (Fig. 1b).

Following the WGMEGS manual temperature at 5 m depth was used to estimate egg development (ICES 2019a). For the DEPM only the mackerel eggs in development stage 1 A are used to estimate daily egg production.

## Results

## Mackerel daily egg production

During the survey the weather was fine. Denmark and Scotland managed to sample all their planned plankton stations. The Netherlands missed 4 plankton stations due to technical issues and limited sampling time.

The spatial egg distribution is shown in Fig. 2. The standard MEGS interpolation rules (ICES, 2019a) were applied where needed (see interpolated stations in Fig. 2). The interpolated egg production accounted for $7.3 \%$ of the DEP. The egg distribution is comparable to previous surveys in the same area and period, with the highest numbers of eggs found in the south western area. Previous surveys did not sample above $59^{\circ} \mathrm{N}$ and no comparison with previous years is available for this area.

The DEP was calculated for the total investigated area (Table 2). For comparison with the previous survey, a DEP was also calculated for the area between 53.5 and $59^{\circ} \mathrm{N}$ and $0.5^{\circ} \mathrm{W}$ and $5.5^{\circ} \mathrm{E}$, which was the area sampled in 2017 in the same period of the year (extended period 2 of 2017; see Fig. 2 for sampled area in 2017). DEP of 2021 was $10 \%$ higher compared to 2017 (Table 3), however the sampled area in 2021 was also larger (9\%) due to coastal stations not sampled or interpolated in 2017.

## Adult parameters

Denmark sampled 817 mackerel and collected ovary samples of 119 females. Of these 34 were suitable for estimating batch fecundity, and 112 for POF analyses for spawning fraction estimation. The Netherlands sampled 524 mackerel during the survey and collected ovary samples of 164 females. Of these 164 ovaries 73 qualified for batch fecundity estimation, and 108 for POF analyses.

Denmark did not deliver the results of the batch fecundity and POF analyses. In agreement with the chairs of WGMEGS, the DEPM adult parameters were therefore estimated with the data provided by the Netherlands. Adult parameters are presented in Table 4.

Of the samples analysed for batch fecundity 54 could be used for batch fecundity estimation. In these samples the batch was clearly separated from the standing stock of vitellogenic oocytes. In the remaining 19 samples the new batch of oocytes was not separated from the standing stock. Batch fecundity was 18735 eggs (Table 4). This is higher compared to the estimate of 12391 in the Atlantic in 2019 (ICES, 2021). Corrected female weight was lower compared to the Atlantic in 2019, 331 and 346 grammes respectively. Spawning fraction in the North Sea was $18 \%$, while this was $23 \%$ in the Atlantic in 2019. Sex ratio was 0.53 and this was similar compared to the Atlantic.

## SSB

Using the stage 1 A (stage duration of 1 A is 1 day) egg data and the estimated adult parameters, the DEP for the entire sampled area in 2021 amounts to an SSB of $2380 * 10^{3}$ tonnes (Table 4). This estimate is
an order of magnitude higher compared to the estimates of previous surveys in the North Sea using the AEPM. The SSB estimated in 2017 using the AEPM was 287 * $10^{3}$ tonnes.

The total area sampled in 2021 was much larger compared to the area sampled in 2017 (Fig. 2). In 2017 sampling was only conducted south of $59^{\circ} \mathrm{N}$. In 2021 sampling was carried out as far as $62^{\circ} \mathrm{N}$ with substantial numbers of eggs being found in this northern area (Fig. 2). In the area above $59^{\circ} \mathrm{N}$ there maybe overlap with the western component.

For comparison between 2021 and 2017 a DEPM estimation of SSB was done using the egg production in the area between 53.5 and $59^{\circ} \mathrm{N}$. No adult parameters were available for 2017, so these were assumed to be same as in 2021 . The SSB in the area between 53.5 and $59^{\circ} \mathrm{N}$ is substantially lower compared to the entire sampled area in 2021 and would be $915 * 10^{3}$ tonnes (Table 5). In 2017 the SSB would be 821 * $10^{3}$ tonnes. For 2017 this is 3 times higher compared to the AEPM estimate of 287 * $10^{3}$ tonnes. Kraus et al. (2012) and Köster et al. (2020) compared the AEPM and DEPM methods for a time-series of cod in the Baltic. They found the trend and SSB in most years were similar using both methods and similar to the ICES estimate of SSB. However, in years with high SSB the two methods diverged (Kraus et al. 2012, Köster et al. 2020).

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Figure 1. Number of samples for NSMEGS 2021; plankton samples per half ICES rectangle (left) and pelagic trawl hauls for mackerel adult samples (right; all hauls included).


Figure 2. Stage 1A mackerel egg production (eggs/m²/day) by half rectangle for NSMEGS 2021. Purple circles represent observed values, black circles represent interpolated values, and crosses represent
observed zeros. Dashed line shows sampled area in extended period 2 in 2017 which was used for comparison calculation between the years.

Table 1. NSMEGS surveys cruise dates in 2021 (For Scotland only stations used in the NSMEGS DEP calculation are shown.)

| Country | NL | DK | SCO |
| :--- | :---: | :---: | :---: |
| Period | 1 | 1 | 1 |
| Dates | $25.05-12.06$ | $31.05-9.06$ | $8.06-11.06$ |
| Plankton stations sampled | 174 | 91 | 29 |
| Pelagic trawl hauls | 12 | 10 | 1 |

Table 2. Daily egg production estimate (stage 1A) in the North Sea.

| Year | DEP * 10 |  |
| :---: | :---: | :---: |
|  | CV | DEP |
| 2021 | 1.28 | $16 \%$ |

Table 3. Comparison of Daily Egg production (stage 1) between 2021 and 2017, in the area between 53.5 and $59^{\circ} \mathrm{N}$.

| Year | $\mathbf{2 0 2 1}$ | 2017 Extended period 2 |
| :---: | :---: | :---: |
| DEP * $\mathbf{1 0}^{\mathbf{1 2}}$ | 4.94 | 4.43 |
| Area sampled <br> $\mathbf{( * ~}^{\mathbf{1 0}} \mathbf{1 1}^{\mathbf{2}} \mathbf{)}$ | 2.25 | 2.01 |

Table 4. Adult parameters and SSB.

| Year | $\mathbf{2 0 2 1}$ |
| ---: | :---: |
| Batch fecundity | 18735 |
| Relative batch <br> fecundity (N/ g) | 42.7 |
| CV Batch fecundity | 0.87 |
| Spawning fraction | 0.18 |
| Sex ratio | 0.53 |
| Female weight (g) | 331.4 |
| SSB (* 103 |  |

Table 5. Comparisons EPM calculation of stage 1 eggs between 2021 and 2017 (extended period 2). (For 2017 the same batch fecundity, S and R are used as for 2021, as these data were not available for 2017.)

|  |  | SSB (*10 $\mathbf{3}^{\mathbf{3}}$ tonnes |  |  |
| :--- | :---: | :---: | :---: | :---: |
| Year | DEP $* \mathbf{1 0}^{\mathbf{1 2}}$ | AEPM | DEPM (below <br> $\mathbf{5 9} \mathbf{} \mathbf{N}$ ) | DEPM (total <br> area) |
| $\mathbf{2 0 2 1}$ | 4.94 | - | 915 | 2380 |
| $\mathbf{2 0 1 7}$ Extended period 2 | 4.43 | 287 | 821 | - |

## DISTRIBUTION AND ABUNDANCE OF NORWEGIAN SPRINGSPAWNING HERRING DURING THE SPAWNING SEASON IN 2022

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Slotte (IMR)
Cruise leader(s): Are Salthaug and Erling Kåre Stenevik (IMR)

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## Summary (English):

During the period 14-27th of February 2022 the spawning grounds of Norwegian spring-spawning herring from Møre ( $62^{\circ} 15^{\prime} \mathrm{N}$ ) to Troms ( $71^{\circ} \mathrm{N}$ ) were covered acoustically by the commercial vessels MS Eros and MS Vendla. The estimated biomass was about 18 \% lower, and the estimated total number was about $29 \%$ lower this year compared to the last year's survey. The uncertainty of the estimates in 2022 was approximately equal to last year. The surveyed population of NSS herring was dominated by the 2016 year class; $52 \%$ in numbers and $46 \%$ in biomass. The 2016 year class was reduced by $37 \%$ in numbers from last year's survey. Most of the spawning stock was found outside Lofoten and Vesterålen this year, further north and more concentrated than usual. The observed maturity indicates a bit later spawning compared to last year and like last year a more northern spawning than normal. As usual, the herring in the southern part of the spawning area were older than those found in the northern part. The estimates of relative abundance from the survey in 2022 are recommended to be used in this year's ICES stock assessment of Norwegian spring-spawning herring.

## Summary (Norwegian):

I perioden 14. - 27. Februar 2022 ble gytefeltene til norsk vårgytende sild fra Møre ( $62^{\circ} 15^{\prime} \mathrm{N}$ ) til Troms $\left(71^{\circ} \mathrm{N}\right)$ dekket akustisk med de kommersielle fartøyene MS Eros og MS Vendla. Den estimerte biomassen var omtrent 18 \% lavere, og det estimerte antallet omtrent 29 \% lavere sammenlignet med fjorårets tokt. Usikkerheten i årets estimat er på samme nivå som i fjor. Gytebestanden var dominert av 2016-årsklassen med 52 \% i antall og 46 \% i vekt. Sammenlignet med toktet i fjor var antallet av 2016-årsklassen redusert med $37 \%$. Mesteparten av gytebestanden befant seg vest av Lofoten og Vesterålen i år. Sammenlignet med tidligere år stod silda lenger nord og var mer konsentrert. Sammenlignet med toktet i fjor var silda kommet noe senere i modningsprosessen i år. I likhet med tidligere år så var det mer eldre sild i den sørlige delen av gyteområdet og silda i nord var yngre. Det anbefales å bruke estimatene av relativ mengde fra toktet i 2022 i ICES sin bestandsvurdering av norsk vårgytende sild.

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## 1 - Introduction

Acoustic surveys on Norwegian spring-spawning herring during the spawning season has been carried out regularly since 1988, with some breaks (in 1992-1993, 1997, 2001-2004 and 2009-2014). In 2015 the survey was initiated again partly based on the feedback from fishermen and fishermen's organizations that IMR should conduct more surveys on this commercially important stock. Since then this survey, hereafter termed the NSSH spawning survey, has continued using hired commercial fishing vessels. In the ICES benchmark assessment of NSS herring in 2016 it was decided to use the data from this time series as input to the stock assessment, together with the ecosystem survey in the Norwegian Sea in May and catch data. Thus, the results from the NSSH spawning survey, have significant influence on the ICES catch advice.

The objective of the NSSH spawning survey 2022 was to continue the time series of abundance estimates, both mean estimates and uncertainty, for use in the ICES WGWIDE stock assessment. Moreover, other biological information about the surveyed spawning stock of Norwegian spring-spawning herring is also presented: spatial distribution of biomass and acoustic densities, total biomass and stock numbers with sample uncertainty, spatial patterns in age and maturity and variations in temperature.

## 2 - Material and methods

## 2.1 - Survey design

During the period $14-27^{\text {th }}$ of February 2022 (same period as in 2017-2021) the spawning grounds from Møre ( $62^{\circ} 15^{\circ} \mathrm{N}$ ) to Troms $\left(71^{\circ} \mathrm{N}\right)$ were covered acoustically by the commercial fishing vessels MS Eros and MS Vendla. The survey was planned based on information from the previous spawning cruises and the distribution of the herring fishery during the autumn 2021 up to the survey start February $14^{\text {th }} 2022$ (Figure 1). The fishery prior to the survey in 2022 indicated that the herring wintering in the Norwegian Sea were entering the coast in the Træna deep south of Røst as observed in previous years. However, unlike previous years the fishery did not move south of Røst before the survey started. Like in the last winter season an extensive fishery in October-February 2021/2022 occurred along the continental slope north of Andenes in addition to the fishery in the Kvænangen fjord area that also have been taking place the five previous years. Biological samples from catches from the northern fishery indicate that the 2016 year class dominated in this area. The survey coverage was therefore planned to also take account of a potentially large flux of herring entering the spawning area from the north. As seen from Figure 1, the fishery during the survey in 2022 mainly took place between Røst and Træna (66.3-67.4 ${ }^{\circ} \mathrm{N}$ ) which is farther north than usual at this time.

The survey design followed a standard stratified design (Jolly and Hampton 1990), where the survey area was stratified before the survey start according to the assumed density structures of herring during the spawning migration (based on previous surveys and fisheries). All strata except the northernmost one were covered with a zigzag design since this is the most efficient use of survey effort (Harbitz 2019). The survey planner function in the Rstox_1.11 package in $r$ was used to generate the transects, and this function generates survey tracks with uniform coverage of strata and a random starting position in the start of each stratum. Each straight line in the zigzag track within a stratum was considered as a transect and a primary sampling unit (Simmonds and MacLennan 2005). Transit tracks between strata, i.e. from the end of the zigzag in one stratum to the start of the zigzag in the next stratum, were not used as primary sampling units. At the start of the survey in 2022 the fishing fleet was located west of Røst and it was estimated that the fleet had moved south to the Træna area around $66.5^{\circ} \mathrm{N}$ when the survey entered this area. Hence, the survey coverage (see Aglen 1989) was planned to be relatively low south of $65^{\circ} \mathrm{N}$ since it was assumed that the fishing fleet followed the front of the herring migrating south and that the abundance of herring south of the fleet therefore was insignificant.

## 2.2 - Biological sampling

Trawl sampling was planned to be carried out on a regular basis during the survey to confirm the acoustic observations and to be able to give estimates of abundance for different size and age groups. Both vessels used a Multpelt 832 scientific sampling trawl with small meshed ( 20 mm ) inner net in the codend and a slit (so called "splitt") close to the codend to avoid too large catches. The following variables of individual herring were analysed from each station with herring catch: total weight in grams and total length in cm (rounded down to the nearest 0.5 cm ) of up to 100 individuals per sample. In addition, age from scales, sex, maturity stage, stomach fullness and gonad weight in grams were measured in up to 50 individuals per sample. Some genetic samples and otoliths were also collected to be used in later research projects.

## 2.3 - Additional data collection

CTD casts (using Seabird 911 systems) were taken by both vessels, spread out haphazardly in the survey area. ADCP data was recorded on Eros as described in Annex 2 in Salthaug et al. (2020). These data will later be used to analyse swimming speed and direction of herring below the vessel.

## 2.4 - Acoustic data processing

Echosounder data from the 38 kHz transducers was, as usual, the basis for measurement of fish density. The software

LSSS version 2.12 .0 was use for post-processing. Echogram scrutinization was carried out by the cruise leader and the chief instrument officer. Data was partitioned into the following categories: "herring", "other" and "air bubbles" (upper 20 meters from the transducer near field).

## 2.5 - Abundance estimation methods

The acoustic density values were stored by species category in nautical area scattering coefficient (NASC) [m $\mathrm{m}^{2} \mathrm{n} . \mathrm{mi}^{-2}$ ] units (MacLennan et al. 2002) in a database with a horizontal resolution of 0.1 nmi and a vertical resolution of 10 m , referenced to the sea surface. To estimate the mean and variance of NASC, we use the methods established by Jolly and Hampton (1990) and implemented in the software Stox version 3.3 (Johnsen et al. 2019). The primary sampling unit is the sum of all elementary NASC samples of herring along the transect multiplied with the resolution distance. The transect ( $t$ ) has NASC value ( $s$ ) and distance length $L$. The average NASC (S) in a stratum $(i)$ is then:

$$
\begin{equation*}
\hat{S}_{i}=\frac{1}{n_{i}} \cdot \sum_{i=1}^{n_{i}} w_{i t} s_{i t} \tag{1}
\end{equation*}
$$

where $w_{\text {it }}=L_{i t} / L_{i t} \quad\left(t=1,2, . . n_{i}\right)$ are the lengths of the $n_{i}$ sample transects, and

$$
\begin{equation*}
L_{i}=\frac{1}{n_{i}} \sum_{i=1}^{n_{i}} L_{i t} \tag{2}
\end{equation*}
$$

The final mean NASC is given by weighting by stratum area, A ;


Variance by stratum is estimated as:

$$
\begin{equation*}
\left.\left.\hat{V}\left|\hat{S}_{i}\right|=\frac{n}{n_{i}-1} \sum_{t=1}^{n} w_{i t}^{2} \right\rvert\, s_{i t}-\bar{s}\right\}^{2} \text { with } \bar{s}_{i}=\frac{1}{n_{i}} \cdot \sum_{i=1}^{n_{i}} s_{i} \tag{4}
\end{equation*}
$$

Where $w_{i t}=L_{i t} / L_{i t} \quad\left(t=1,2, . . n_{i}\right)$ are the lengths of the $n_{i}$ sample transects.

The global variance is estimated as
$\hat{V}(\hat{S})=\frac{\sum_{i} A_{i=1}^{2} \hat{V}(\hat{S})}{\left(\sum_{i} A\right)^{2}}$

The global relative standard error of NASC
$R S E=100 \sqrt{\frac{\hat{V}(\hat{S})}{N}} / \hat{S}$
where $N$ is number of strata.

In order to verify acoustic observations and to analyse year class structure over the surveyed area, trawling was carried out regularly along the transects. All trawl stations with herring were used to derive a common length distribution for all transect within the respective strata. All stations had equal weight.

Relative standard error by number of individuals by age group was estimated by combining Monto Carlo selection from estimated NASC distributions by stratum with bootstrapping techniques of the assigned trawl stations.

The acoustic estimates presented in this report use the 38 kHz NASC, and the mean was calculated for data scrutinized as herring and collected along the transects (acoustic recordings taken during trawling, and for experimental activity are excluded). The number of herring ( $N$ ) in each length group ( $I$ ) within each stratum ( $i$ ) is then computed as:
$N_{I}=\frac{f_{i} \cdot \hat{S}_{i} \cdot A_{i}}{\langle\sigma\rangle}$
Where
$f_{l}=\frac{n_{l} L_{i}^{2}}{\sum_{l=1}^{m} n_{l} L_{l}}$
is the "acoustic contribution" from the length group $L_{l}$ to the total energy and $\left\langle\mathrm{s}_{i}\right\rangle$ is the mean nautical area scattering coefficient $\left[\mathrm{m}^{2} / \mathrm{nmi}^{2}\right]$ (NASC) of the stratum. A is the area of the stratum [ $\mathrm{nmi}{ }^{2}$ ] and $\sigma$ is the mean backscattering cross section at length $L_{I}$. The conversion from number of fish by length group ( $I$ ) to number by age is done by estimating an age ratio from the individuals of length group (I) with age measurements. Similar, the mean weight by length and age grouped is estimated.

The mean target strength (TS) is used for the conversion where $\sigma=4 \pi 10^{(T S / 10)}$ is used for estimating the mean backscattering cross section. Traditionally, TS $=20 \operatorname{logL}-71.9$ (Foote 1987) has been used for mean target strength of herring during the spawning surveys, however, several papers question this mean target strength. Ona (2003) describes how the target strength of herring may change with changes with depth, due to swimbladder compression. He measured the mean target strength of herring to be TS $=20 \log L-2.3 \log (1+z / 10)-65.4$ where $z$ is depth in meters. Given that previous surveys were estimated using Foote (1987), the estimation this year was also done with this TS, for direct comparison and possible inclusion in the stock assessment by ICES WGWIDE 2021 as another year in the time series.

## 3 -Results and discussion

## 3.1 - Survey coverage

The cruise tracks of the NSSH spawning survey in 2022, together with pelagic trawl stations and CTD stations are shown in Figure 2. As mentioned above, the coverage south of $65^{\circ} \mathrm{N}$ was fairly low since we expected low abundance in this area, which turned out to be the case (see below). Thus, most of the available survey effort was used to carry out dense coverage of the strata north of $65^{\circ} \mathrm{N}$. The survey coverage (see Aglen 1989) of the first three strata was 5,7 and 9 respectively (starting from south) and 11 in the four next strata with zigzag transects. The northernmost stratum with parallel transects had a survey coverage of 9 . Pelagic trawl hauls were carried out regularly (Fig. 2 ) in the areas where herring like marks were observed on the echo sounder, to confirm the acoustic observations based on species composition in the catch and to obtain biological samples like size, maturity stage and age of herring. A total of 34 CTD casts were carried out in the surveyed area (Fig. 2). Nautical area scattering coefficients (NASC) allocated to herring from acoustic transects by each nautical mile are shown in Figure 3. Significant herring marks on the echosounders started to occur slightly north of $66^{\circ} \mathrm{N}$, which is unusually far north in mid-February, and herring was observed in the entire area north of this. South of Lofoten the herring was mainly distributed around the shelf edge of the Røst bank, but outside Lofoten and Vesterålen herring was also observed on the banks nearer land. North of Vesterålen the herring was distributed along the shelf edge as usual, and the zero-line was established in the north around $70.9^{\circ} \mathrm{N}$. Capelin marks started to appear around $69.7^{\circ} \mathrm{N}$ (confirmed by trawl samples) and was observed regularly north of this, in particular around the shelf edge area in the northernmost part. The herring schools appeared to be deeper and clearly separated from the more shallow capelin schools, an observation that the trawl sampling also supported. No more capelin results are presented in this report as the focus is on herring.

## 3.2 - Estimates of abundance

The abundance estimates from this survey are viewed as relative, i.e. as indices of abundance, since there are highly uncertain scaling parameters like acoustic target strength and compensation for herring migrating in the opposite direction of the survey. The abundance estimates are shown in Table 1 and 2. The 2016 year class (age 6) dominated both in numbers (52 \%) and biomass (46 \%), followed by the 2013 year class (age 9) which contributed $12 \%$ in numbers and $15 \%$ in biomass. Compared with the point estimates from last year (see Salthaug et al. 2021) the 2016 year class was reduced by $37 \%$ in numbers and the 2013 year class by 22 \%. The point estimate of total stock biomass (TSB) in the survey area was 3.302 million tons which is $18 \%$ lower than last year's estimate (mean of 1000 bootstrap replicates). The time series of total stock biomass from the survey is shown in Figure 4. The point estimate of total stock number (TSN) in the survey area was 12.2 billion which is $29 \%$ lower than last year's estimate. The time series of total stock number from the survey is shown in Figure 5. This year's estimates of TSB and TSN are slightly below the respective means of the time series. The relative standard error (CV) of the TSB and TSN estimates in 2021 are both $17 \%$ (Tab. 1 and 2). These estimates of sample uncertainty are quite similar to those from the two previous surveys. The CV per age (Tab. 1 and 2) shows the normal pattern with high uncertainty for the very young and old year classes and moderate (20-30 \%) for the most abundant ages in the survey. Figure 6 shows estimates of number per year class in the eight most recent surveys. The estimated numbers from the survey in 2022 seems to decline as excepted for the year classes that are fully recruited to the survey. In addition, like in the most recent surveys the 2016 and 2013 year classes are estimated to be the most abundant which shows that this survey is internally consistent. Mean weight and length from the 2021 spawning survey are shown in Table 3. The Stox project used to calculate abundance and related parameters is openly available and can be found here:
http://metadata.nmdc.no/metadata-api/landingpage/2870f9f21da64f3a01641dfe12512b33

## 3.3 - Spatial distribution of the stock

The relative distribution of the estimated biomass per stratum is shown in Figure 7. This year most of the biomass
( $84 \%$ ) was found in the two strata west of Lofoten and Vesterålen, while only a small fraction was found in the strata to the north and south of these. The spawning stock was much more concentrated and further north than usual this year. Age compositions per stratum are shown in Figure 8. The southernmost stratum where herring was recorded was dominated by herring older than eight years, which is consistent with earlier observations; the largest and oldest fish are in the front of the spawning migration. The 2016 year class dominated in the rest of the strata, and the proportion of younger herring was as usual highest in the north.

The pattern with large and old fish in the southern part of the spawning area and younger and older herring in the north has been thoroughly discussed in Slotte and Dommasnes, 1997, 1998, 1999, 2000; Slotte, 1998b; Slotte, 1999a, Slotte 2001, Slotte et al. 2000, Slotte \& Tangen 2005, 2006). The main hypothesis is that this could be due to the high energetic costs of migration, which is relatively higher in small compared to larger fish (Slotte, 1999b). Large fish and fish in better condition will have a higher migration potential and more energy to invest in gonad production and thus the optimal spawning grounds will be found farther south (Slotte and Fiksen, 2000), due to the higher temperatures of the hatched larvae drifting northwards and potentially better timing to the spring bloom (Vikebø et al. 2012).

Figure 9 shows the proportion of different maturation stages in each stratum. Most of the herring was classified as maturing or ripe, and the proportions of maturing herring were larger than last year which indicates later spawning this year. The old herring in the southernmost stratum was dominated by maturing individuals indicating that these fish would swim further south before spawning. The fishery also indicated that this was the case since catches moved further south after the survey covered the area (see Fig. 1). A small fraction of the herring outside Lofoten and Vesterålen were spawning and this, together with the large proportion of ripe individuals, indicate that much of the 2016 year class spawned in this area. Like last year this shows a very northern spawning this year, which also was confirmed through the fishery that was very low at the historically important spawning grounds off Møre.

## 3.4 - Geographical variation in temperatures experienced by the herring

Temperatures experienced by herring from close to the surface and down to 250 m are shown in Figure 10 for the areas south and north of $67^{\circ} \mathrm{N}$, for the years after 2016 when the survey has been carried out in the same period (latter half of February). The temperatures in 2022 varied from $7.7^{\circ} \mathrm{C}$ at 250 m depth south of $67 \infty \mathrm{~N}$ to $5.4^{\circ} \mathrm{C}$ at 5 m depth north of $67^{\circ} \mathrm{N}$. The temperatures near the surface were quite low this year, and also varied more with depth compared to earlier years. At typical spawning depths of herring at 100-200 m depth, the temperature conditions were quite similar to those observed during the most recent NSSH spawning surveys.

## 3.5 - Quality of the survey

In 2022 both vessels were equipped with multifrequency equipment on a drop keel. The weather conditions were exceptionally good this year so that acoustic data with good quality was recorded and trawling on registrations could be carried out all of the time. No correction for air bubble attenuation (as described in Annex 3 in Slotte et al. 2019) had to be carried out this year due to the nice weather. As opposed to last year the zero line was clearly established in the north, and we are not aware of any observations that indicates presence of mature NSS herring outside the survey area during the survey this year. To conclude, the acoustic and biological data recorded in 2022 on the NSSH spawning survey were of satisfactory quality and the estimates from the survey are recommended to be used in the stock assessment of Norwegian spring-spawning herring in 2022.

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## 5 - Tables

Table 1. Abundance estimates (mill ion individuals) of Norwegian spring-spawning herring during the spawning survey 14.-27.
February 2022 , based on 1000 bootstrap replicates.

| Age | 5th percentile | Median | 95th percentile | Mean | SD | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 1 | 23 | 62 | 27 | 19 | 0.72 |
| 3 | 13 | 71 | 134 | 72 | 36 | 0.50 |
| 4 | 51 | 154 | 310 | 162 | 78 | 0.48 |
| 5 | 406 | 738 | 1148 | 760 | 234 | 0.31 |
| 6 | 4473 | 6314 | 8475 | 6393 | 1256 | 0.20 |
| 7 | 205 | 308 | 458 | 317 | 76 | 0.24 |
| 8 | 377 | 557 | 788 | 563 | 126 | 0.22 |
| 9 | 1066 | 1500 | 2063 | 1515 | 298 | 0.20 |
| 10 | 174 | 294 | 458 | 301 | 89 | 0.30 |
| 11 | 303 | 477 | 707 | 486 | 122 | 0.25 |
| 12 | 175 | 297 | 439 | 301 | 79 | 0.26 |
| 13 | 137 | 247 | 393 | 255 | 80 | 0.31 |
| 14 | 206 | 380 | 584 | 385 | 119 | 0.31 |
| 15 | 37 | 71 | 122 | 73 | 26 | 0.36 |
| 16 | 227 | 384 | 602 | 395 | 117 | 0.30 |
| 17 | 18 | 52 | 109 | 57 | 29 | 0.50 |
| 18 | 36 | 86 | 157 | 89 | 37 | 0.41 |
| 20 | 0 | 13 | 42 | 15 | 15 | 1.04 |
| TSN | 8910 | 12126 | 15591 | 12183 | 2051 | 0.17 |

Table 2. Abundance estimates (thousand tons ) of Norwegian spring-spawning herring during the spawning survey 14.-27. February 2022 , based on 1000 bootstrap replicates.

| Age | 5th percentile | Median | 95th percentile | Mean | SD | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0 | 1 | 4 | 2 | 1 | 0.77 |
| 3 | 1 | 8 | 18 | 9 | 5 | 0.59 |
| 4 | 7 | 23 | 44 | 24 | 11 | 0.48 |
| 5 | 76 | 131 | 204 | 136 | 41 | 0.30 |
| 6 | 1083 | 1511 | 2035 | 1533 | 303 | 0.20 |
| 7 | 57 | 87 | 128 | 89 | 22 | 0.24 |
| 8 | 115 | 169 | 239 | 171 | 38 | 0.22 |
| 9 | 336 | 478 | 660 | 481 | 96 | 0.20 |
| 10 | 58 | 102 | 160 | 104 | 31 | 0.30 |
| 11 | 104 | 165 | 245 | 168 | 42 | 0.25 |
| 12 | 64 | 108 | 158 | 109 | 29 | 0.26 |
| 13 | 51 | 92 | 147 | 95 | 30 | 0.32 |
| 14 | 75 | 138 | 213 | 140 | 43 | 0.31 |
| 15 | 14 | 27 | 46 | 28 | 10 | 0.36 |


| Age | 5th percentile | Median | 95th percentile | Mean | SD | CV |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 16 | 87 | 148 | 232 | 151 | 45 | 0.30 |  |
| 17 | 6 | 19 | 41 | 21 | 11 | 0.51 |  |
| 18 | 13 | 0 | 33 | 60 | 34 | 14 | 0.41 |
| 20 | 2424 | 5 | 16 | 6 | 6 | 1.03 |  |
| TSB | 3291 | 4246 | 3302 | 557 | 0.17 |  |  |

Table 3. Estimated length and weight of individuals by age group of Norwegian spring-spawning herring during the spawning survey 14.-27. February 2022 , based on 1000 bootstrap replicates.

| Age | mean weight (g) | CV weight | mean length (cm) | CV length |
| :---: | :---: | :---: | :---: | :---: |
| 2 | 56.7 | 0.063 | 21.2 | 0.017 |
| 3 | 105.7 | 0.230 | 24.9 | 0.053 |
| 4 | 137.6 | 0.066 | 27.4 | 0.017 |
| 5 | 171.3 | 0.026 | 29.1 | 0.006 |
| 6 | 230.0 | 0.012 | 31.3 | 0.003 |
| 7 | 277.0 | 0.021 | 33.0 | 0.005 |
| 8 | 301.1 | 0.018 | 34.1 | 0.005 |
| 9 | 315.2 | 0.010 | 34.3 | 0.003 |
| 10 | 343.4 | 0.018 | 35.6 | 0.007 |
| 11 | 342.3 | 0.019 | 35.6 | 0.006 |
| 12 | 362.1 | 0.017 | 36.6 | 0.003 |
| 13 | 371.7 | 0.021 | 36.9 | 0.004 |
| 14 | 362.5 | 0.017 | 36.5 | 0.005 |
| 15 | 373.7 | 0.023 | 37.1 | 0.006 |
| 16 | 380.9 | 0.014 | 37.2 | 0.003 |
| 17 | 362.5 | 0.037 | 37.3 | 0.008 |
| 18 | 379.1 | 0.024 | 37.1 | 0.011 |
| 20 | 387.6 | 0.032 | $37.0$ | 0.000 |

## 6 - Figures



Figure 1. Distribution of commercial catches of Norwegian spring-spawning herring from October 2021 until February 2022, based on electronic logbooks. Each point represent one catch, only catches larger than 10 tons are shown.


Figure. 2. Cruise tracks (mostly acoustic transects), pelagic trawl stations (triangles), and CTD stations (Z) covered by Eros and Vendla on the Norwegian spring-spawning herring spawning survey 14.-27. February 2022.


Figure 3. Acoustic densities (NASC) of herring recorded during the Norwegian spring-spawning herring spawning survey 14.-27. February 2022. Points represent NASC values per nautical mile. Depth contours are shown for $50 \mathrm{~m}, 100 \mathrm{~m}, 150 \mathrm{~m}, 200 \mathrm{~m}, 500 \mathrm{~m}$, 1000 m, 1500 m and 2000 m.

## SPAWNING SURVEY,TSB



Figure 4. Estimates of total biomass from the Norwegian spring-spawning herring spawning surveys during1988-2022. The estimates are mean of 1000 bootstrap replicates and the error bars represent $90 \%$ confidence intervals.

## SPAWNING SURVEY,TSN



Figure 5. Estimates of total number from the Norwegian spring-spawning herring spawning surveys during1988-2022. The estimates are mean of 1000 bootstrap replicates and the error bars represent $90 \%$ confidence intervals.


Figure 6. Abundance by year class estimated during the Norwegian spring-spawning herring spawning surveys 2015-2022 (mean of 1000 bootstrap replicates). Legend: Separate colour for each survey year.


Figure 7. Relative distribution by stratum of the biomass of herring from the Norwegian spring-spawning herring spawning survey $14 .-$ 27. February 2022.


Figure 8. Age distribution per stratum from the Norwegian spring-spawning herring spawning survey 14.-27. February 2022 . The area of the bubbles is scaled with the total number estimated in each stratum.


Figure 9. Proportions of different maturity stages from the Norwegian spring-spawning herring spawning survey 14.-27. February 2022.


Figure 10. Mean temperatures at 5, 20, 50, 100, 150, 250 m in the area covered during the Norwegian spring-spawning herring spawning surveys in 2017-2022.

Working Document to

Working Group on International Pelagic Surveys (WGIPS)
23 - 27 January 2023
and
Working Group on Widely Distributed Stocks (WGWIDE)
24 - 30 August 2022

# INTERNATIONAL ECOSYSTEM SURVEY IN NORDIC SEA (IESNS) <br> in April - May 2022 

Post-cruise meeting on Teams, 14-16 June 2022

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## Introduction

In April-May 2022, four research vessels and one hired commercial vessel participated in the International ecosystem survey in the Nordic Seas (IESNS); R/V Dana, Denmark (joint survey by Denmark, Germany, Ireland, The Netherlands and Sweden), R/V Jákup Sverri, Faroe Islands, R/V Árni Friðriksson, Iceland, R/V G.O. Sars, Norway and M/S Resolute, United Kingdom (UK). It should be noted that this was the first year that UK participated in the survey, and the plan is to continue the participation in the coming years. The Barents Sea is usually surveyed by a Russian research vessel, but that was not possible in 2022. The aim of the survey was to cover the whole distribution area of the Norwegian Spring-spawning herring with the objective of estimating the total abundance of the herring stock, in addition to collect data on plankton and hydrographical conditions in the area. The survey was initiated by the Faroes, Iceland, Norway and Russia in 1995. Since 1997 also the EU participated (except 2002 and 2003) and from 2004 onwards it was more integrated into an ecosystem survey.

This report represents analyses of data from this International survey in 2022 that are stored in the PGNAPES database and the ICES acoustic database and supported by national survey reports from some survey participants (Dana: Cruise Report R/V Dana Cruise 03/2022. International Ecosystem survey in the Nordic Seas (IESNS) in 2022, Árni Friðriksson: Report on Survey A5-2022, Bjarnason, 2022, Jákup Sverri: Preliminary Report Cruise no. 2216).

## Material and methods

Coordination of the survey was done during the WGIPS meeting in January 2022 and by correspondence. Planning of the acoustic transects and hydrographic stations and plankton stations were carried out by using the survey planner function in the r package Rstox version 1.11 (see https://www.hi.no/en/hi/forskning/projects/stox). The survey planner function generates the survey plan (transect lines) in a cartesian coordinate system and transforms the positions to the geographical coordinate system (longitude, latitude) using the azimuthal equal distance projection, which ensures that distances, and also equal coverage, if the method used is designed with this prerequisite, are preserved in the transformation. Figure 1 shows the planned acoustic transects and hydrographic and plankton stations in each stratum. Only parallel transects were used this year, however, because the transects follow great circles they appear bended in a Mercator projection. The participating vessels together with their effective survey periods are listed in the table below:

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| Vessel | Institute | Survey period |
| :--- | :--- | :--- |
| Dana | DTU Aqua - National Institute of Natural Resources, <br> Denmark | $22 / 04-20 / 05$ |
| G.O. Sars | Institute of Marine Research, Bergen, Norway | $26 / 04-30 / 05$ |
| Jákup Sverri | Faroe Marine Research Institute, Faroe Islands | $28 / 04-08 / 05$ |
| Árni Friðriksson | Marine and Freshwater Research Institute, Iceland | $04 / 05-23 / 05$ |
| Resolute | CEFAS, United Kingdom | $24 / 04-06 / 05$ |

Note that Resolute covered the UK EEZ in the southernmost part of the IESNS survey area, but this area was also covered by G.O. Sars and Dana. The reason for this double coverage was to ensure consistency with previous year's surveys (the UK coverage went well and these data were used in the abundance estimation). Figure 2 shows the cruise tracks, Figure 3 the hydrographic and WPII plankton stations and, Figure 4 Macroplankton trawl and Multinet stations and Figure 5 the pelagic trawl stations. Survey effort by each vessel is detailed in Table 1. Daily contacts were maintained between the vessels during the course of the survey, primarily through electronic mail. The temporal progression of the survey is shown in Figure 6. UK also covered an area south of the IESNS survey area and this is described in Annex A.

In general, the weather conditions did not affect the survey even if there were some days that were not favourable and trawling, WP2 and Multinet sampling at some stations were prevented. The survey was based on scientific echosounders using 38 kHz frequency. Transducers were calibrated with the standard sphere calibration (Foote et al., 1987) prior to the survey. Salient acoustic settings are summarized in the text table below.

Acoustic instruments and settings for the primary frequency (boldface).

|  | Dana | G. O. Sars | Arni <br> Frioriksson | Jákup Sverri | Resolute |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Echo sounder | Simrad EK60 | Simrad EK80 | Simrad EK80 | Simrad EK80 | Simrad EK80 |
| Frequency (kHz) | 38 | $38,18,70$, | $38,18,70$, | $18,38,70$, | 38,200 |
| Primary |  | $120,200,333$ <br> transducer | ES38BP | ES 38-7 | ES38-7 |

All participants except UK used the same post-processing software (LSSS). The UK data were, however, scrutinized using Echoview. Scrutinization was carried out according to an agreement at a PGNAPES scrutinizing workshop in Bergen in February 2009 (ICES 2009), and "Notes from acoustic Scrutinizing workshop in relation to the IESNS", Reykjavík 3.-5. March 2015 (Annex 4 in ICES 2015). Generally, acoustic recordings were scrutinized on daily basis and species identified and partitioned using catch information, characteristic of the recordings, and frequency between integration on 38 kHz and on other frequencies by a scientist experienced in viewing echograms. Immediately after the 2022 survey an online
meeting was held to standardise the scrutiny and to agree on particularly difficult scrutiny situations encountered. All vessels used a large or medium-sized pelagic trawl as the main tool for biological sampling. The salient properties of the trawls, plankton nets and hydrographic equipment are as follows:

|  | Dana | G.O. Sars | Arni <br> Friðriksson | Jákup Sverri | Resolute |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Trawl dimensions |  |  |  |  |  |
| Circumference (m) |  | 496 | 832 | 832 | 972 |
| Vertical opening (m) | 20-30 | 25-30 | 20-35 | 44-55 | 30-50 |
| Mesh size in codend (mm) | 20/40 | 24 | 20 | 45 | 100 |
| Typical towing speed (kn) | 3.5-4.5 | 3.0-4.5 | 3.1-5.0 | 3.7 (3-4.5) | 3.5-5 |
| Plankton sampling |  |  |  |  |  |
| Sampling net | WP2 | WP2 | WP2 | WP2 | WP2 |
| Standard sampling depth (m) | 200 | 200 | 200 | 200 | 200 |
| Hydrographic sampling |  |  |  |  |  |
| CTD unit | SBE911 | SBE911 | SBE911 | SBE911 | SAIV SD208 |
| Standard sampling depth (m) | 1000 | 1000 | 1000 | 1000 | 250 |

Catches from trawl hauls were sorted and weighed; fish were identified to species level, when possible, and other taxa to higher taxonomic levels. A subsample of herring, blue whiting and mackerel were sexed, aged, and measured for length and weight, and their maturity status was estimated using established methods. An additional sample of fish was measured for length. For the Norwegian, Icelandic and Faroese vessel, a smaller subsample of stomachs was sampled for further analyses on land. As part of a coming age reading and stock identity workshop, genetic samples were collected of herring. Salient biological sampling protocols for trawl catches are listed in the table below.

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|  | Species | Dana | G.O. Sars |  | Jákup | Resolute |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Friðriksson | Sverri |  |
| Length measurements | Herring | 200-300 | 100 | 300 | 100-300 | 100 |
|  | Blue whiting | 200-300 | 100 | 50 | 100-200 | 100 |
|  | Mackerel | 100-200 | 100 | 50 | 100-200 | 100 |
|  | Other fish sp. | 50 | 30 | 30 | 100-150 | 30 |
| Weighed, sexed and maturity determination |  |  |  |  |  |  |
|  | Herring | 50 | 25-100 | 100 | 50* | 50 |
|  | Blue whiting | 50 | 25-100 | 50 | $50^{*}$ | 50 |
|  | Mackerel | 50 | 25-100 | 50 | 50 | 50 |
|  | Other fish sp. | 0 | 0 | 0 | $0^{*}$ | 0 |
| Otoliths/scales collected | Herring | 50 | 25-30 | 100 | 50 | 50 |
|  | Blue whiting | 50 | 25-30 | 50 | 25-50 | 50 |
|  | Mackerel | 0 | 25-30 | 50 | 50 | 50 |
|  | Other fish sp. | 0 | 0 | 0 | 0 | 0 |
| Stomach sampling | Herring | 0 | 10 | 10 | 5 | 0 |
|  | Blue whiting | 0 | 10 | 10 | 5 | 0 |
|  | Mackerel | 0 | 10 | 10 | 5 | 0 |
|  | Other fish sp. | 0 | 0 | 0 | 0 | 0 |
| Genetic samples | Herring | 50 |  |  | 25 | 50 |

* Number of weighed individuals significantly higher.

Acoustic data were analysed using the StoX software package (version 3.4.0) which has been used for some years now for WGIPS coordinated surveys. A description of StoX can be found in Johnsen et al. (2019) and here: https://www.hi.no/en/hi/forskning/projects/stox. Estimation of abundance from acoustic surveys with StoX is carried out according to the stratified transect design model developed by Jolly and Hampton (1990). This method requires pre-defined strata, and the survey area was therefore split into 5 strata with pre-defined acoustic transects (this year only 4 strata, as the Barents Sea was not surveyed). Within each stratum, parallel transects with equal distances were used. The distance between transects was based on available survey time, and the starting point of the first transect in each stratum was randomized. This approach allows for robust statistical analyses of uncertainty of the acoustic estimates. The strata and transects used in StoX are shown in Figure 2. Generally, and in accordance with most WGIPS coordinated surveys, all trawl stations within a given stratum with catches of the target species (either blue whiting or herring) were assigned to all transects within the stratum, and the length distributions were weighted equally within the stratum.

The following target strength (TS)-to-fish length (L) relationships were used:
Blue whiting: TS $=20 \log (\mathrm{~L})-65.2 \mathrm{~dB}$ (ICES 2012)
Herring: $\quad \mathrm{TS}=20.0 \log (\mathrm{~L})-71.9 \mathrm{~dB}$ (Foote et al. 1987)
The target strength for herring is the traditionally one used while this target strength for blue whiting was first applied in 2012 (ICES 2012).

The hydrographical and plankton stations by survey are shown in Figure 3. Most vessels collected hydrographical data using a SBE 911 CTD. Maximum sampling depth was 1000 m . Zooplankton was sampled by WPII nets on all vessels, according to the standard procedure for the surveys. Mesh sizes were 180 or $200 \mu \mathrm{~m}$. The net was hauled vertically from 200 m to the surface or from the bottom whenever bottom depth was less than 200 m . All samples were split in two and one half was preserved in formalin while the other half was dried and weighed. The samples for dry weight were size fractionated before drying by sieving the samples through $2000 \mu \mathrm{~m}$ and $1000 \mu \mathrm{~m}$ sieves, giving the size fractions $180 / 200-1000 \mu \mathrm{~m}, 1000-2000 \mu \mathrm{~m}$, and $>2000 \mu \mathrm{~m}$. Data are presented as mg total dry weight per $\mathrm{m}^{2}$. For the zooplankton distribution map, all stations are presented. Interpolation was carried out using Bratseth's Successive Correction Method (Bratsheth, 1986). This method was designed specifically for marine data, and it uses bottom depth to calculate the similarity among the interpolation points. More specifically, it uses objective analysis with a Gaussian correlation function where the effective distance between the observations and the nodes of the interpolation grids is defined based on the difference in bottom depths, as follows:
$r^{2}=r_{x}^{2}+r_{y}^{2}+\left(\lambda \frac{H_{\mathrm{a}}-H_{\mathrm{o}}}{H_{\mathrm{a}}+H_{\mathrm{o}}}\right)^{2}$
where rx and ry is the geographic distance in the zonal and meridional directions, and Ha and Ho are the bottom depths at the analysis and observation points, respectively (Skagseth and Mork, 2012). The analysis was done using an R script based on a MATLAB routine developed by Kjell Arne Mork (Mork et al. 2014). For the time series, stations in the Norwegian Sea delimited to east of $14^{\circ} \mathrm{W}$ and west of $20^{\circ} \mathrm{E}$ have been included. Estimates of the statistical distribution of the zooplankton biomass indices is done by simple bootstrapping by re-sampling with replacement.

## Results and Discussion

## Hydrography

The temperature distributions in the ocean, averaged over selected depth intervals; 0$50 \mathrm{~m}, 50-200 \mathrm{~m}$, and $200-500 \mathrm{~m}$, are shown in Figures $7 \mathrm{a}-\mathrm{c}$. The temperatures in the surface layer $(0-50 \mathrm{~m})$ ranged from below $0^{\circ} \mathrm{C}$ in the Greenland Sea to $9-10^{\circ} \mathrm{C}$ in the southern part of the Norwegian Sea (Figure 7a). The Arctic front was encountered south of $65^{\circ} \mathrm{N}$ east of Iceland extending eastwards towards about $2^{\circ} \mathrm{W}$ where it turned north-eastwards to $65^{\circ} \mathrm{N}$ and then almost straight northwards. This front was sharper below 50 m than above. Further to west at about $8^{\circ} \mathrm{W}$ another front runs northward to Jan Mayen, the Jan Mayen Front, that was most distinct in the upper 200 m . The warmer North Atlantic water formed a broad tongue that stretched far northwards along the Norwegian coast with temperatures about $6{ }^{\circ} \mathrm{C}$ to the Bear Island at $74.5^{\circ} \mathrm{N}$ in the surface layer.

Relative to the long-term mean, from 1995 to 2021, the temperatures at $0-50 \mathrm{~m}$ were below the mean in most of the Norwegian Sea (Figure 7a). Below 50 m depth, the
patterns were more fragmented, but the Norwegian Sea was still in general colder than the long-term mean (Figures 7b-c). Largest negative temperature anomalies were between Iceland and Faroe Islands due to a more southern located IcelandFaroe front compared to the long-term mean. This was found for all depths, and the temperatures in this region were in some locations $3{ }^{\circ} \mathrm{C}$ lower than the mean (Figures $7 \mathrm{a}-\mathrm{c}$ ). Also, in the centre of the Norwegian Basin, the temperatures were $1{ }^{\circ} \mathrm{C}$ lower than the mean, probably because of a more eastern located Arctic front. Warmest regions, relative to the long-term mean, were in the eastern Greenland Sea, with temperatures $2{ }^{\circ} \mathrm{C}$ higher than the mean, and in some areas below 50 m depth in southern and southwestern parts of the Norwegian Sea.

Two main features of the circulation in the Norwegian Sea, where the herring stock is grazing, are the Norwegian Atlantic Current (NWAC) and the East Icelandic Current (EIC). The NWAC with its offshoots forms the northern limb of the North Atlantic current system and carries relatively warm and salty water from the North Atlantic into the Nordic Seas. The EIC, on the other hand, carries Arctic waters. To a large extent this water derives from the East Greenland Current, but to a varying extent, some of its waters may also have been formed in the Iceland and Greenland Seas. The EIC flows into the southwestern Norwegian Sea where its waters subduct under the Atlantic waters to form an intermediate Arctic layer. While such a layer has long been known in the area north of the Faroes and in the Faroe-Shetland Channel, it is in the last four decades a similar layer has been observed all over the Norwegian Sea. Also, in periods this layer has been less well-defined.

This circulation pattern creates a water mass structure with warm Atlantic Water in the eastern part of the area and more Arctic conditions in the western part. The NWAC is rather narrow in the southern Norwegian Sea, but when meeting the Vøring Plateau off Mid Norway it is deflected westward. The western branch of the NWAC reaches the area of Jan Mayen at about $71^{\circ} \mathrm{N}$. Further northward in the Lofoten Basin the lateral extent of the Atlantic water gradually narrows again, apparently under topographic influence of the mid-ocean ridge. It has been shown that atmospheric forcing largely controls the distribution of the water masses in the Nordic Seas. Hence, the lateral extent of the NWAC, and consequently the position of the Arctic Front, that separates the warm North Atlantic waters from the cold Arctic waters, is correlated with the large-scale distribution of the atmospheric sea level pressure. The local air-sea heat flux in addition influence the upper layer and it is found that it can explain about half of the year-to-year variability of the ocean heat content in the Norwegian Sea.

## Zooplankton

The zooplankton biomass ( mg dry weight $\mathrm{m}-2$ ) in the upper 200 m is shown in Figure 8. Sampling stations were evenly spread over the area, covering Atlantic water, Arctic water, and the Arctic frontal zone. The highest zooplankton biomasses
were found in the eastern and southeastern parts. Within the eastern area, several locations had high biomass and a large patch was found at ca. $3^{\circ} \mathrm{W}$ and $64.5^{\circ} \mathrm{N}$. Lower biomasses were found in central and western parts of the Norwegian Sea.

Figure 9 shows the zooplankton indices for the sampling area (delimited to east of $14^{\circ} \mathrm{W}$ and west of $20^{\circ} \mathrm{E}$ ). To examine regional biomass differences, the area was divided into 4 sub-areas 1) East of Iceland, 2) the Jan Mayen Arctic front, 3) the Lofoten Basin (covering the northern Norwegian Sea, and 4) the Norwegian Sea Basin (covering the southern Norwegian Sea). The zooplankton biomass index for 2022 was respectively: $4563,6627,9237$ and 9962 mg dry weight $\mathrm{m}^{-2}$, and while the subareas east of Iceland and Jan Mayen arctic front showed a decrease compared to last year, the Lofoten- and Norwegian Basin increased. The zooplankton biomass indices for the Norwegian Sea in May have been estimated since 1995. All subareas had a high biomass period until mid-2000, and a lower period thereafter. The decrease was most pronounced in the Iceland Sea, where the reduction was $59 \%$. In the Lofoten- and Norwegian Basins there has been an increasing trend during the low-biomass period.

The reasons for the changes in zooplankton biomass are not obvious. It is worth noting that the period with lower zooplankton biomass coincides with higher-thanaverage heat content in the Norwegian Sea (ICES, 2020) and reduced inflow of Arctic water into the southwestern Norwegian Sea (Kristiansen et al., 2019). Timing effects, such as match/mismatch with the phytoplankton bloom, can also affect the zooplankton abundance. The high biomass of pelagic fish feeding on zooplankton has been suggested to be one of the main causes for the reduction in zooplankton biomass. However, carnivorous zooplankton and not pelagic fish may be the main predators of zooplankton in the Norwegian Sea (Skjoldal et al., 2004), and we do not have good data on the development of the carnivorous zooplankton stocks.

## Norwegian spring-spawning herring

Survey coverage in the Norwegian Sea was considered adequate in 2022. The zeroline was believed to be reached for adult NSS herring in most of the areas. It is recommended that the results from IESNS 2022 can be used for assessment purpose. The herring was primarily distributed in the central and southwestern area (Figure 10). In the westernmost area old herring dominated, but in general, the 2016-yearclass was the most abundant year class throughout the survey area. It is a commonly observed pattern that the older fish are distributed in the southwest while the younger fish are found closer to the nursery areas in the Barents Sea (Figure 11).

Six-year-old herring (2016-year class) dominated both in terms of number (49\%) and biomass ( $48 \%$ ) on basis of the StoX bootstrap estimates for the Norwegian Sea (Table 2). The abundance of the 2016 year-class decreased by $19 \%$ compared to last year's estimate which could be expected since this year-class was fully recruited to
the survey last year (Figure 12). The second largest year-class in the survey was the 2013 year-class ( $10 \%$ in numbers), and older age groups (10-18 years old) contributed with less than $10 \%$ to the abundance estimate. Uncertainty estimates for number at age based on bootstrapping within StoX are shown in Figure 13 and Table 2. The relative standard error (CV) is $21 \%$ both for the total biomass and for the total numbers estimate, and the relative standard error for the dominating age groups is around 20-30 \% (Figure 13).

The total estimate of herring in the Norwegian Sea from the 2022 survey was 19.8 billion in number and the biomass was 4.4 million tonnes. The biomass estimate is $13 \%$ lower than the 2021 survey estimate and also the estimated number is about $13 \%$ lower than in 2021. The biomass estimate decreased significantly from 2009 to 2012 and has since then been rather stable at 4.2 to 5.9 million tonnes with similar confidence interval (Figure 14), with the lowest abundance occurring in 2017. The 2016 year class now appears to be fully recruited, distributed widely in the feeding area and more dominant than the older year classes.

There was no coverage of juvenile herring in strata 5 (the Barents Sea) in May 2022.
In the last 6 years, there have been concerns regarding age reading of herring, because the age distributions from the different participants have showed differences - particularly older specimens appear to have uncertain ages. A scale and otolith exchange has been ongoing for some time, where scales and otoliths for the same fish have been sampled. As a follow-up on that work, a new exchange and following workshop are currently being planned for April 2023. The survey group emphasizes the necessity of having this workshop before next year's survey takes place.

With respect to age-reading concerns in the recent years, the comparison between the nations in this year's survey for the most part appeared to be in good agreement (Figure 15).

Recently, concerns have been raised by the survey groups for the International ecosystem surveys in the Nordic Seas (IESNS and IESSNS) on mixing issues between Norwegian spring-spawning herring and other herring stocks (e.g. Icelandic summer-spawning, Faroese autumn-spawning, Norwegian summer-spawning and North Sea type autumn-spawning herring) occurring in some of the fringe regions in the Norwegian Sea. Until now, fixed cut lines have been used by the survey group to exclude herring of presumed other types than NSS herring, however this simple procedure is thought to introduce some contamination of the stock indices of the target NSS herring. WGIPS noted in their 2019 report that the separation of different herring stock components is an issue in several of the surveys coordinated in WGIPS and the needs for development of standardized stock splitting methods was also noted in the WKSIDAC (ICES 2017).

## Blue whiting

Boostrap estimates of abundance, biomass, mean length and mean weight of blue whiting during IESNS 2022 are shown in Table 3. The estimated biomass was 1.5 million tons $(\mathrm{CV}=0.13)$ which is a $76 \%$ increase from last year's estimate, and one of the two highest estimates after 2007 (together with the 2016 estimate). The estimated total abundance was 17.2 billion ( $\mathrm{CV}=0.13$ ) which is a $112 \%$ increase from last year's estimate. The stock is totally dominated by 1 and 2 year old (2021 and 2022 year classes) and the estimates of total abundance, abundance of age 1 and abundance of age 2 are all the highest observed after 2007. Uncertainty estimates for numbers at age based on bootstrapping with StoX are shown in Figure 18 and Table 3.

The spatial distribution of blue whiting in 2022 is shown in Figure 16. As usual, most of the fish was registered in the eastern part of the Norwegian Sea. However, higher concentrations than in later years were observed in more central areas, in particular around the zero meridian in the southern part. This corresponds well with the high abundance estimate. The largest fish was found in the northwestern part of the of the survey area this year (Figure 17). Comparison of the size and age distributions of blue whiting by stratum and country are shown in Figure 19 and 20, and they seem to be in fairly good agreement.

## Mackerel

Trawl catches of mackerel are shown in Figure 21. Mackerel was present in the southern and eastern part of the Norwegian Sea in the beginning of May. This year the catches did not extend as far north as compared with recent years, only north to circa $64^{\circ} \mathrm{N}$. This is the lowest northward extent of mackerel catches during IESNS after 2007 (first year with data from all participating vessels). No further quantitative information can be drawn from these data as this survey is not designed to monitor mackerel.

General recommendations and comments
Recommendation
Addressed to

1. Continue the methodological research in distinguishing WGIPS between herring and blue whiting in the interpretation of echograms.
2. It is recommended that the the planned age reading

WG workshop in April 2023 also includes a session n how to deal with stock components of herring in the IESNSsurvey.

## Next year's post-cruise meeting

We will aim for next meeting in 13-15 June 2023. The final decision will be made at the next WGIPS meeting.

## Concluding remarks

- The sea temperature in 2022 was generally below the long-term mean (1995-2021) in the Norwegian Sea, but the pattern was more fragmented below 50 m depth. The Arctic front in the southern Norwegian Sea was more southerly and easterly located in 2022 compared to the long-term mean.
- The 2022 indices of meso-zooplankton biomass in the Norwegian Sea and adjoining waters were fairly similar to last year's estimates.
- The total biomass estimate of NSSH in herring in the Norwegian Sea was 4.4 million tonnes, which is a $13 \%$ decrease from the 2021 survey estimate. The estimate of total number of NSSH was 19.8 billion, which is $13 \%$ lower than in the 2021 survey. The survey followed the pre-planned protocol and the survey group recommends using the abundance estimates in the analytical assessment.
- The 2016 year class of NSSH dominated in the survey indices both in numbers (49\%) and biomass (48\%). The abundance of the 2016 year-class decreased by $19 \%$ compared to last year's estimate
- The biomass of blue whiting measured in the 2022 survey increased by $76 \%$ from last year's survey and $112 \%$ in terms of numbers. The stock is dominated by the 2020 and 2021 year classes) and the estimates of total abundance, abundance of age 1 and abundance of age 2 are all the highest observed after 2007.


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## Tables

Table 1. Survey effort by vessel for the International ecosystem survey in the Nordic Seas in May June 2022.

| Vessel | Effective <br> survey <br> period | Effective <br> acoustic <br> cruise <br> track <br> (nm) | Trawl <br> stations | Ctd <br> stations | Aged <br> fish <br> (HER) | Length <br> fish <br> (HER) | Plankton <br> stations |
| :--- | :--- | :--- | :---: | :---: | :---: | :---: | :---: |
| Dana | $26 / 4-16 / 5$ | 2495 | 20 | 36 | 253 | 873 | 35 |
| Jákup Sverri | $28 / 4-8 / 5$ | 1464 | 19 | 23 | 325 | 1093 | 23 |
| Árni Fridriksson | $8 / 5-23 / 5$ | 3013 | 14 | 40 | 863 | 2747 | 34 |
| G.O. Sars | $26 / 4-30 / 5$ | 5103 | 37 | 60 | 375 | 1107 | 59 |
| Resolute | $24 / 4-06 / 5$ | 1158 | 11 | 22 | 290 | 537 | 22 |
| Total |  | $\mathbf{1 3 2 3 3}$ | $\mathbf{1 0 1}$ | $\mathbf{1 8 1}$ | $\mathbf{2 1 0 6}$ | $\mathbf{6 3 5 7}$ | $\mathbf{1 7 3}$ |

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Table 2. IESNS 2022 in the Norwegian Sea. Estimates of abundance, mean weight and mean length of Norwegian spring-spawning herring. The estimates are mean of 1000 bootstrap replicates in Stox.

| Length (cm) | Age in years (year class) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Number <br> (10^6) | $\begin{aligned} & \text { Biomass } \\ & \left(10^{\wedge} 6 \mathrm{~kg}\right) \end{aligned}$ | Mean weight (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 3 2019 |  |  |  |  |  |  | $\begin{array}{r} 10 \\ 2012 \end{array}$ | $\begin{array}{r} 11 \\ 2011 \end{array}$ | $\begin{array}{r} 12 \\ 2010 \end{array}$ | $\begin{array}{r} 13 \\ 2009 \end{array}$ | $\begin{array}{r} 14 \\ 2008 \end{array}$ |  | $\begin{array}{r} 16 \\ 2006 \end{array}$ | $\begin{array}{r} 17 \\ 2005 \end{array}$ | $\begin{array}{r} 18 \\ 2004 \end{array}$ | Unknown |  |  |  |
| 17-18 | 18.6 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 18.6 | 0.7 | 38.0 |
| 18-19 | 37.3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 37.3 | 1.6 | 42.5 |
| 19-20 | 27.4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 27.4 | 1.5 | 56.0 |
| 20-21 | 113.7 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 4.2 | 117.8 | 7.2 | 59.5 |
| 21-22 | 107.8 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 107.8 | 7.8 | 72.6 |
| 22-23 | 116.3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.6 | 116.9 | 9.7 | 82.9 |
| 23-24 | 71.4 | 22.9 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 94.3 | 8.9 | 93.3 |
| 24-25 |  | 46.8 | 142.5 |  | 5.9 |  |  |  |  |  |  |  |  |  |  |  |  | 1.7 | 197.0 | 21.3 | 108.6 |
| 25-26 |  | 61.2 | 229.3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 290.6 | 33.6 | 116.3 |
| 26-27 |  | 27.1 | 252.4 | 49.3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 328.9 | 44.7 | 134.8 |
| 27-28 |  | 72.1 | 134.8 | 5.8 | 6.8 |  |  |  |  |  |  |  |  |  |  |  |  |  | 219.5 | 33.0 | 152.2 |
| 28-29 |  | 46.7 | 94.5 | 168.3 | 57.5 | 37.7 |  |  |  | 12.8 |  |  |  |  |  |  |  |  | 417.3 | 70.3 | 168.7 |
| 29-30 | 14.7 | 46.9 | 174.7 | 336.4 | 304.1 | 81.4 | 116.3 |  | 58.3 |  |  |  |  |  |  |  |  |  | 1132.8 | 210.4 | 185.2 |
| 30-31 |  | 28.4 | 149.5 | 297.3 | 1411.4 | 239.3 | 378.3 | 187.0 | 29.2 | 26.0 |  |  |  |  |  |  |  |  | 2746.4 | 549.4 | 199.1 |
| 31-32 |  | 30.8 | 24.6 | 212.9 | 3210.3 | 353.7 | 374.9 | 411.2 | 79.3 |  |  | 88.9 |  |  |  |  |  |  | 4786.7 | 1034.1 | 215.0 |
| 32-33 |  |  | 4.7 | 203.8 | 2986.8 | 144.5 | 138.6 | 383.8 | 113.2 | 29.2 | 68.7 | 21.1 |  |  |  |  |  |  | 4094.4 | 956.8 | 232.9 |
| 33-34 |  |  |  | 12.0 | 1427.9 | 98.0 | 163.1 | 243.8 | 121.0 | 6.9 | 110.7 |  |  | 6.5 |  |  |  |  | 2189.9 | 554.7 | 254.2 |
| 34-35 |  |  |  |  | 190.5 | 157.7 | 213.7 | 491.8 | 10.9 | 4.8 |  |  |  |  |  |  |  |  | 1069.5 | 299.5 | 280.0 |
| 35-36 |  |  |  |  | 29.5 | 38.3 | 197.5 | 235.6 | 56.4 | 77.0 | 39.2 | 31.1 | 10.4 |  | 7.2 | 15.6 |  |  | 737.8 | 219.3 | 296.8 |
| 36-37 |  |  |  |  | 2.7 |  | 57.8 | 99.3 | 70.3 | 80.7 | 60.1 | 32.4 | 29.5 | 35.6 | 6.1 | 14.1 |  |  | 488.7 | 154.9 | 316.9 |
| 37-38 |  |  |  |  |  |  |  | 11.1 | 38.1 | 60.1 | 32.6 | 97.2 | 72.0 | 56.7 | 33.9 | 10.9 |  |  | 412.5 | 139.7 | 338.7 |
| 38-39 |  |  |  |  |  |  |  |  |  | 24.2 | 13.6 | 22.7 | 3.4 | 28.6 | 26.1 | 17.6 |  |  | 136.2 | 49.7 | 363.3 |
| 39-40 |  |  |  |  |  |  |  |  |  | 17.1 |  |  |  | 5.4 | 7.0 | 6.0 | 5.6 | 0.2 | 41.5 | 15.1 | 366.1 |
| 40-41 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 5.0 |  |  | 2.5 | 7.5 | 3.1 | 408.0 |
| TSN(mill) | 507.2 | 383.0 | 1207.1 | 1285.8 | 9633.2 | 1150.5 | 1640.3 | 2063.6 | 576.6 | 338.9 | 324.9 | 293.4 | 115.3 | 132.9 | 85.4 | 64.2 | 5.6 |  | 19817.1 |  |  |
| cv (TSN) | 0.59 | 0.49 | 0.45 | 0.34 | 0.23 | 0.36 | 0.37 | 0.34 | 0.40 | 0.31 | 0.42 | 0.40 | 0.39 | 0.35 | 0.44 | 0.45 | 1.12 |  | 0.21 |  |  |
| TSB(1000 t) | 37.7 | 58.0 | 182.1 | 252.4 | 2132.2 | 266.1 | 400.6 | 531.5 | 152.2 | 102.0 | 89.7 | 86.2 | 37.1 | 45.1 | 29.8 | 20.5 | 2.0 |  | 4427.0 |  |  |
| cv (TSB) | 0.55 | 0.48 | 0.41 | 0.35 | 0.23 | 0.34 | 0.35 | 0.32 | 0.38 | 0.31 | 0.39 | 0.35 | 0.39 | 0.36 | 0.46 | 0.46 | 1.12 |  | 0.21 |  |  |
| Mean length(cm) | 21.2 | 27.6 | 27.9 | 30.0 | 31.5 | 32.2 | 33.0 | 33.6 | 34.0 | 35.8 | 35.2 | 35.9 | 36.6 | 36.9 | 37.3 | 36.7 | 39.0 |  |  |  |  |
| Mean weight(g) | 76.0 | 165.2 | 169.1 | 199.6 | 223.0 | 246.3 | 262.7 | 273.6 | 285.3 | 314.2 | 299.6 | 320.7 | 321.4 | 341.9 | 346.6 | 319.4 | 365.4 |  |  |  |  |

Table 3. IESNS 2022 in the Norwegian Sea. Estimates of abundance, mean weight and mean length of blue whiting. The estimates are mean of 1000 bootstrap replicates in Stox.

| Length (cm) | Age in years (year class) |  |  |  |  | $\begin{array}{r} 6 \\ 2016 \\ \hline \end{array}$ | $\begin{array}{r} 7 \\ 2015 \end{array}$ | $\begin{array}{r} 8 \\ 2014 \end{array}$ | Unknown | Number <br> (10^6) | Biomass$\left(10^{\wedge} 6 \mathrm{~kg}\right)$ | Mean weight (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 |  |  |  |  |  |  |  |
|  | 2021 | 2020 | 2019 | 2018 | 2017 |  |  |  |  |  |  |  |
| 14-15 | 7.6 |  |  |  |  |  |  |  | 2.6 | 10.2 | 0.1 | 16.0 |
| 15-16 | 232.7 |  |  |  |  |  |  |  |  | 232.7 | 4.9 | 20.8 |
| 16-17 | 1304.5 | 29.8 |  |  |  |  |  |  |  | 1334.3 | 32.5 | 24.4 |
| 17-18 | 4114.3 | 122.2 |  |  |  |  |  |  |  | 4236.5 | 125.6 | 29.7 |
| 18-19 | 5637.5 | 135.3 |  |  |  |  |  |  |  | 5772.8 | 199.4 | 34.6 |
| 19-20 | 4229.8 | 161.9 | 6.7 |  |  |  |  |  |  | 4398.5 | 173.8 | 39.9 |
| 20-21 | 1206.1 | 387.6 | 66.5 |  |  |  |  |  |  | 1660.2 | 78.4 | 47.5 |
| 21-22 | 271.7 | 1526.6 | 123.7 |  |  |  |  |  |  | 1922.0 | 109.8 | 57.4 |
| 22-23 | 135.6 | 2649.2 | 58.5 |  |  |  |  |  |  | 2843.2 | 183.6 | 65.5 |
| 23-24 | 1.9 | 2821.4 | 207.0 |  |  |  |  |  |  | 3030.3 | 221.0 | 74.5 |
| 24-25 | 27.0 | 2116.0 | 308.7 |  |  |  |  |  |  | 2451.8 | 199.0 | 83.2 |
| 25-26 |  | 495.9 | 277.6 | 12.9 |  |  |  |  |  | 786.4 | 72.5 | 93.1 |
| 26-27 |  | 117.2 | 145.7 | 27.8 |  |  |  |  |  | 290.7 | 30.4 | 105.0 |
| 27-28 |  | 11.7 | 34.6 | 25.9 | 31.6 | 7.1 | 9.4 |  |  | 120.2 | 14.2 | 118.4 |
| 28-29 |  |  | 50.1 | 13.5 |  |  | 4.9 |  |  | 68.5 | 9.0 | 128.6 |
| 29-30 |  |  |  |  | 2.3 | 9.2 | 16.7 | 12.9 | 0.0 | 41.2 | 5.9 | 141.6 |
| 30-31 |  |  |  | 17.6 | 20.8 |  | 10.0 | 17.7 |  | 66.1 | 10.5 | 159.2 |
| 31-32 |  |  |  |  | 26.5 | 20.2 | 5.7 |  |  | 52.3 | 9.7 | 182.3 |
| 32-33 |  |  |  |  |  |  | 46.2 | 16.4 | 0.2 | 62.8 | 12.6 | 199.5 |
| 33-34 |  |  |  |  |  |  | 9.5 | 8.0 | 0.1 | 17.7 | 4.2 | 239.4 |
| 34-35 |  |  |  |  | 7.9 |  |  |  | 3.4 | 11.3 | 3.0 | 271.5 |
| 35-36 |  |  |  |  |  |  |  |  |  |  |  |  |
| 36-37 |  |  |  |  | 2.2 |  |  |  |  | 2.2 | 0.7 | 330.0 |
| TSN(mill) | 17169 | 10575 | 1279 | 98 | 91 | 36 | 102 | 55 |  | 29411.9 |  |  |
| cv (TSN) | 0.16 | 0.15 | 0.20 | 0.39 | 0.36 | 0.51 | 0.54 | 0.54 |  | 0.13 |  |  |
| TSB(1000 t) | 603.3 | 729.5 | 105.7 | 11.9 | 15.2 | 5.9 | 17.7 | 10.5 |  | 1500.6 |  |  |
| cv (TSB) | 0.15 | 0.16 | 0.19 | 0.40 | 0.38 | 0.53 | 0.55 | 0.53 |  | 0.13 |  |  |
| Mean length(cm) | 18.2 | 22.7 | 24.1 | 27.2 | 29.7 | 29.9 | 30.6 | 30.5 |  |  |  |  |
| Mean weight(g) | 36 | 72 | 85 | 121 | 167 | 159 | 168 | 183 |  |  |  |  |

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Figures


Figure 1. The pre-planned strata and transects for the IESNS survey in 2022 (red: EU, dark blue: Norway, yellow: Faroes Islands, violet: UK, green: Iceland). Hydrographic stations and plankton stations are shown as blue circles with diamonds. All the transects have numbered waypoints for each 30 nautical mile and at the ends.

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Figure 2. Cruise tracks and strata (with numbers) for the IESNS survey in May 2022.


Figure 3. IESNS survey in May 2022: location of hydrographic and WPII plankton stations.

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Figure 4. IESNS survey in May 2022: location of Macroplankton/Krill trawl and Multinet stations.


Figure 5. IESNS survey in May 2022: location of pelagic trawl stations.

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Figure 6. Temporal progression IESNS in April-May 2022.

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Figure 7a. Temperature (left) and temperature anomaly (right) averaged over $0-50 \mathrm{~m}$ depth in May 2021. Anomaly is relative to the 1995-2019 mean.


Figure 7b. Same as above but averaged over 50-200 m depth.


Figure 7c. Same as above but averaged over 200-500 m depth.


Figure 8. Distribution of zooplankton biomass ( mg dry weight $\mathrm{m}^{-2}$ ) in the upper 200 m in May 2022.


Figure 9. Indices of zooplankton biomass ( mg dry weight $\mathrm{m}^{-2}$ ) sampled by WP2 in May in the Norwegian Sea and adjacent waters from 1995-2022.

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(a)

(b)


Figure 10. Distribution of Norwegian spring-spawning herring as measured during the IESNS survey in May 2021 in terms of NASC values $\left(\mathrm{m}^{2} / \mathrm{nm}^{2}\right)$ averaged for every 1 nautical mile.

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Figure 11. Mean length of Norwegian spring-spawning herring in all hauls in May 2022.
 Cohort year class

Figure 12. Tracking of the Total Stock Number at age (TSN, in billions) of Norwegian spring-spawning herring for each cohort since 2004 from age 2 to age 6. From 2008, stock is estimated using the StoX software. Prior to 2008, stock was estimated using BEAM.

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Figure 13. Norwegian spring-spawning herring in the Norwegian Sea: R boxplot of abundance and relative standard error (CV) obtained by bootstrapping with 1000 replicates using the StoX software.

## IESNS,TSB



Figure 14. Biomass estimates of Norwegian-spring spawning herring in the IESNS survey (Barents Sea, east of $20^{\circ} \mathrm{E}$, is excluded) from 1996 to 2022 as estimated using BEAM (1996-2007; calculated on basis of rectangles) and as estimated with the software StoX (2008-2021; bootstrap means with $90 \%$ confidence interval; calculated on basis of standard stratified transect design).

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Figure 15. Comparison of the age distributions of NSS-herring by stratum and country in IESNS 2022. The strata are shown in Figure 3.
(a)

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(b)


Figure 16. Distribution of blue whiting as measured during the IESNS survey in May 2022 in terms of NASC values $\left(\mathrm{m}^{2} / \mathrm{nm}^{2}\right)$ (a) averaged for every 1 nautical mile.

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Figure 17. Mean length of blue whiting in all hauls in IESNS 2022. The strata are shown.


Figure 18. Blue whiting in the Norwegian Sea: R boxplot of abundance and relative standard error (CV) obtained by bootstrapping with 1000 replicates using the StoX software.


Figure 19. Comparison of the length distributions of blue whiting by stratum and country in IESNS 2022. The strata are shown in Figure 3.


Figure 20. Comparison of the age distributions of blue whiting by stratum and country in IESNS 2022. The strata are shown in Figure 3.

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Figure 21. Pelagic trawl catches of mackerel in IESNS 2022.

## ANNEX A

## UK contribution to IESNS 2022

## Background

In 2022 the UK participated to the IESNS survey by running a full survey on a chartered vessel that covered the UK EEZ within the IESNS survey area and an additional area south to $62^{\circ} \mathrm{N}$, which is currently considered as the southern boundary of the Norwegian Springspawning herring stock. The main objective of the survey was to determine the distribution abundance and age structure of herring and blue whiting in the area south to the IESNS traditional coverage and detect and quantify potential mixing between different herring stocks (e.g. NSSH, NSAS, WoS).

## Materials and methods

The survey was conducted onboard the commercial pelagic trawler F/V Resolute from $24 / 04 / 2022$ to $06 / 05 / 2022$. All the details about characteristics of the vessel, sampling, acoustic settings used, and data processing are listed in the previous section of this report. The acoustic transects and location of the hydrographic and plankton stations are shown in fig. A1. The survey area was split into 2 strata: a northern stratum that included the area north of $62^{\circ}$ N which overlapped with the same area covered by the RV Dana and a southern stratum that covered the rest of the survey area (Fig. A2-a). For blue whiting, the southern stratum was further split into 2 additional strata to account for the habitat preferences of the species (Fig.A2-b).

## Results and discussion

In total 9 acoustic transects were completed covering a total of 1158 nmi of acoustic sampling unit. A total of 11 pelagic trawls were carried out to provide groundtruth information about the species and size composition and to collect biological information (Fig. A3). In addition, CTD and plankton sampling were performed on 22 fixed stations.
Herring was patchily distributed over the whole survey area with higher densities located primarily around the Shetlands and at the southernmost transect of the survey located west of Orkney (Fig. A4). Herring size ranged from 21 to 33.5 cm with larger sizes found in the northern part of the survey area (Fig. A5). The total biomass estimate was $450,258 \mathrm{t}$ (northern stratum: 43.550, southern stratum: 406,708 ) and a total number of 2.89 billion. Three-yearsold and four-years-old herring were the most abundant age classes in terms of numbers accounting for $23 \%$ and $21 \%$ respectively of the total estimate (Fig. A6). The relative standard error (CV) is $40 \%$ for both the total biomass and for the total numbers estimate.
Blue whiting was mainly distributed over the slope area in the north and western part of the survey areas (Fig. A7). Blue whiting aggregations primarily consisted of continuous and dense layers distributed between $200-400 \mathrm{~m}$ depth in the water column. Blue whiting size ranged from 16 to 33.5 cm with an overall average of 22.5 cm (Fig. A8). The total biomass
estimate was 449,656 t (northern stratum: $261,872 \mathrm{t}$, southern stratum: $187,784 \mathrm{t}$ ) and a total number of 6.4 billion. Two-years-old was the most abundant age class in terms of numbers accounting for $89 \%$ of the total estimate (Fig A9). The relative standard error (CV) is $24 \%$ for both the total biomass and for the total numbers estimate.
Mackerel was caught in almost all the trawls carried out. The size ranged from 18 to 41 cm with an overall average size of 33 cm (Fig. A10). No further quantitative information can be drawn from these data as this survey was not designed to monitor mackerel.

## Future work

Genetic analysis is planned to be performed on herring fin clips samples collected during the survey ( 290 samples collected across 7 locations) to characterise the different stocks present in the survey area and the potential level of mixing with the Norwegian spring spawning herring.


Figure A1 - Acoustic transects and location of hydrographic and plankton stations.


Figure A2 - Strata used for biomass estimation for herring (a) and blue whiting (b).


Figure A3-Location and catch composition of the pelagic trawl stations.

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Figure A4 - Distribution of herring in terms of NASC values $\left(\mathrm{m}^{2} / \mathrm{nm}^{2}\right)$ averaged for every 1 nautical mile.


Figure A5 - Distribution of the mean length of herring measured in the pelagic trawl catches.

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Figure A6 - Boxplot of herring abundance at age and relative standard error (CV) obtained by bootstrapping using the StoX software.

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Figure A7 - Distribution of blue whiting in terms of NASC values $\left(\mathrm{m}^{2} / \mathrm{nm}^{2}\right)$ averaged for every 1 nautical mile.


Figure A8 - Distribution of the mean length of blue whiting measured in the pelagic trawl catches.


Figure A9 - Boxplot of blue whiting abundance at age and relative standard error (CV) obtained by bootstrapping using the StoX software.

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Figure A10 - Distribution of the mean length of mackerel measured in the pelagic trawl catches.

# North Sea mackerel total egg production for 2022 using the daily egg production method 

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## Introduction

The North Sea Mackerel Egg Survey (NSMEGS) is designed to estimate the spawning stock biomass (SSB) of mackerel of the North Sea spawning component of the Northeast-Atlantic stock on a triennial basis. Up to and including 2017 this was undertaken utilizing the annual egg production method (AEPM) and generally undertaken in the year following the survey covering the western components. This method estimates and combines total annual egg production (TAEP), realized fecundity per gram female, and sex (male to female) ratio to calculate SSB.

Spatial and temporal coverage in the North Sea was reduced with the withdrawal of Norway from the NSMEGS in 2014, with the Netherlands left as the sole survey participant in 2015 and 2017. In 2020 Denmark was recruited as a new participant for the NSMEGS, and in 2021 the UK (England) announced that they were willing to participate.

An issue for the NSMEGS is that since 1982 it has been impossible to collect and sample pre-spawning mackerel, which are necessary in order to estimate the potential fecundity. For SSB estimation using the AEPM, the realized fecundity value used was from the 1982 estimate (Iversen and Adoff, 1983). For a number of years it was recognised that an AEPM survey wasn't producing the best results for the North Sea. Therefore, at the WGMEGS meeting in 2018 a decision was made to use the Daily Egg Production Method (DEPM) for future North Sea surveys (ICES 2018). The DEPM requires only one full sweep, in a short time period, over the entire mackerel spawning area, preferably during peak spawning time. A disadvantage of the DEPM is that it requires many more mackerel ovary samples to be collected to estimate batch fecundity and spawning fraction.

## Survey

In 2022 the UK and Denmark conducted the North Sea survey. Whilst planning the survey it became apparent that the vessel time available from the two countries would not be sufficient to cover the area. As a result, Norway agreed to survey the four northernmost transects in the North Sea at the start of their period 6 survey.

The samples were collected and analysed according to the WGMEGS manuals (ICES 2019a, 2019b). UK and Norway sampled eggs with a Gulf VII plankton sampler while Denmark used a Nackthai sampler. The UK and Denmark utilised a $500 \mu \mathrm{~m}$ plankton net which is standard protocol for the North Sea due to issues with clogging, while Norway used a $250 \mu \mathrm{~m}$ mesh. At each station a double oblique haul was performed from the surface to 5 m above the bottom, a maximum depth of 200 m , or 20 m below the thermocline in case of stratification of the water column. Temperature and salinity were measured during the haul with a CTD mounted on top of the plankton sampler. Either electronic or mechanical flowmeters were mounted on the plankton sampler to monitor flow.

The NSMEGS was carried out from $5^{\text {th }}-24^{\text {th }}$ June (Table 1). During this period the spawning area between $54^{\circ} \mathrm{N}$ and $62^{\circ} \mathrm{N}$ was surveyed once, receiving a single coverage (Fig. 1). The survey is designed to cover the entire spawning area with samples collected every half ICES statistical rectangle (ICES, 2014). In total 259 plankton stations were sampled, with 19 stations interpolated. On each of the Danish transects at least one pelagic trawl haul was performed for the collection of mackerel adult samples. Due to problems with their fishing gear CEFAS carried out a number of rod and line fishing events.

Following the WGMEGS manual temperature at 5 m depth was used to estimate egg development (ICES 2019a). For the DEPM only the mackerel eggs in development stage 1 A are used to estimate daily egg production.

## Results

## Mackerel daily egg production

The spatial egg distribution is shown in Fig. 1. Standard MEGS interpolation rules (ICES, 2019a) were applied where needed. Egg distributions are comparable to 2021, however egg numbers seem to be more evenly distributed throughout the survey area this year.

The total area sampled in 2022 was slightly smaller than the area sampled in 2021, the first full transect was started at $54^{\circ} 15^{\prime} \mathrm{N}$ compared to $53^{\circ} 15^{\prime} \mathrm{N}$ in 2021. The two southern transects were sampled but there were issues with many of the stations re the accuracy of the flow data. This resulted in three valid stations south of $54^{\circ} \mathrm{N}$ with a further three being interpolated. The invalid stations do give an indication of the presence and absence (qualitative data) of mackerel stage 1 A and above over this area.

The DEP was calculated for the total investigated area (Table 2). Total egg production for 2022 was $0.6699 * 10^{13}$ eggs. This is a $50 \%$ decrease on egg numbers reported in 2021 (Table 3).

## Adult parameters

Denmark conducted 33 hauls, from which they sampled 1180 mackerel and collected ovary samples from 364 females. England conducted 20 rod and line fishing events of which 9 were positive, biologically sampling 225 mackerel and collecting ovary samples of 74 females. Norway collected 239 female mackerel samples from 5 fishing hauls, (Table 1). As these samples were collected in June no analysis has been carried out on them. Batch fecundity and POF counting will take place before the end of the year, with the results to be delivered prior to the WGMEGS meeting in April 2023.

SSB
As there are no data available from the adult parameters, WGMEGS is just reporting egg production for 2022.

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Figure 1. Heat map of Stage 1A mackerel egg production (eggs. $\mathrm{m}^{-2}$. day ${ }^{-1}$ ) by half rectangle for the North Sea, 2022. Grey circles represent observed values, crosses represent observed zeros.

Table 1. NSMEGS surveys cruise dates in 2022 (For Norway only stations used in the NSMEGS DEP calculation are shown). UK=UK England, DK=Denmark, NO=Norway.

| Country | UK | DK | NO |
| :--- | :---: | :---: | :---: |
| Period | 1 | 1 | 1 |
| Dates (2022) | $5.06-24.06$ | $08.06-17.06$ | $7.06-19.06$ |
| Plankton stations sampled | 135 | 79 | 45 |
| Pelagic trawl hauls |  | 33 | 5 |
| Positive rod and line <br> events | 9 |  |  |

Table 2. Total egg production using the Daily egg production estimate (stage 1A abundance) in the North Sea for 2022.

| Year | DEP $\mathbf{1 1 0}^{\mathbf{1 3}}$ | CV DEP |
| :---: | :---: | :---: |
| 2022 | 0.67 |  |

Table 3. Comparison of total stage 1A egg production for 2022 and 2021 in the North Sea estimated by the Daily Egg production method.

| Year | $\mathbf{2 0 2 2}$ | $\mathbf{2 0 2 1}$ |
| :---: | :---: | :---: |
| DEP $\boldsymbol{* 1 0}^{\mathbf{1 3}}$ | 0.67 | 1.28 |

# 2022 Mackerel and Horse Mackerel Egg Survey <br> Preliminary Results 

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## 1 Introduction

The mackerel and horse mackerel egg survey is an ICES-coordinated international study in the north east Atlantic conducted during the first half of 2022. This study is a combined plankton and fishery investigation formed by a series of individual surveys which have taken place triennially since the late 1970s and is coordinated by the ICES Working Group on Mackerel and Horse Mackerel Egg Surveys (WGMEGS). Historically a North sea mackerel egg survey is carried out in the year after the western and southern surveys. in 2022, due to the presence of new participants, the all surveys were carried out in the same year

The main objective of this series of individual cruises from January until July is to produce both an index and a direct estimate of the biomass of the north east Atlantic mackerel stock and an index for the southern and western horse mackerel stocks. The results have been used in the assessment for mackerel since 1977 and from 1992 for horse mackerel. The mackerel and horse mackerel egg survey is still a principal source of data providing fisheries independent information for these stocks.

The general method is to quantify the freshly spawned eggs in the water column on the spawning grounds. To be able to establish a relationship between eggs and biomass of the spawning stock, the fecundity of the females must also be determined. This is undertaken by sampling ovaries before and during spawning. In cases where the annual egg production method is applied the potential fecundity is counted from whole mount volumetric subsamples using a dissecting microscope while atresia is counted histologically from slides. Realised fecundity is estimated as potential fecundity minus atresia. The realised fecundity is used in combination with the calculated number of freshly spawned eggs in the water to estimate the spawning stock biomass.

To provide reliable estimates of spawned eggs and fecundity an extensive coverage of the spawning area is required both in time and space. The spawning of the southern horse mackerel stock and mackerel starts in late December off the Portuguese coast. Spawning proceeds further north along the continental shelf edge as water temperature increases during late winter and spring. In the past peak spawning of mackerel has normally occurred in April-May in the area of the Sole Banks with an extension to the Porcupine Bank. Whilst
the distribution and timing of peak western horse mackerel spawning has remained fairly stable during recent surveys the same cannot be said for NEA mackerel. The 2010 and 2013 MEGS surveys saw peak mackerel spawning in February - March with 2013 also demonstrating a shift in the geographical centre of spawning further south within the southern Biscay region. Since then however mackerel spawning is now observed over a large region of the Northeast Atlantic both on and off the continental shelf, ranging as far west as Hatton Bank, as far north as Iceland and the Faroe Islands and in recent years around the Shetland Islands and the Norwegian coast in the Northeast.

This survey report presents the preliminary results of the 2022 mackerel and horse mackerel egg survey provided for WGWIDE in August 2022. The survey report and the analysis will be finalised during the next WGMEGS meeting in April 2023. Although every effort was made to ensure that WGWIDE were provided with the most recent and accurate data-set, WGMEGS cannot guarantee that there will not be changes prior to the analysis being finalised. This is due to the extremely large numbers of plankton and fecundity samples to be analysed following the surveys as well as the tight deadline set by WGWIDE for delivering these estimates. This has resulted in a very limited time within which to process the 2022 MEGS data.

## Survey effort

As a consequence of the long spawning period and the large survey area involved, the mackerel and horse mackerel egg surveys have always relied on broad international participation. In 2022 a total of 18 individual cruises were carried out, 16 in the Atlantic and 2 in the North sea, for a total of 321 at-sea survey days. Individual contributions were; Spain (IEO: 42 days at sea, AZTI: 30 days), Scotland ( 53 days), the Netherlands ( 39 days), Ireland ( 28 days), Portugal ( 34 days), Germany ( 23 days), Norway ( 15 days), Faroe Islands (14 days), England ( 23 days) and Denmark (14 days). Denmark joined the group in 2020 and participated in the 2021 North Sea survey along with the Netherlands. England rejoined the group in 2021 and in 2022 conducted the North Sea survey in participation with Denmark.

## Survey design

The aim of the triennial egg survey is to determine the annual egg production (AEP). This is calculated using the mean daily egg production rates per pre-defined sampling period for the complete spawning area of the Northeast Atlantic Mackerel and Horse Mackerel Stocks. To achieve this, one plankton haul per each half rectangle (separated by approximately $15-20 \mathrm{NM}$, depending on latitude) is conducted on alternating transects covering the complete spawning area. The 2022 egg survey was designed in order to maximise both the spatial and temporal coverage in each of the sampling periods. Given the very large area to be surveyed this design minimises the chances of under/overestimation of the egg production (ICES 2008).

The 2022 survey plan was split into 6 sampling periods (Table 1). Portugal were assigned to start the survey in the southern area during Period 2. No sampling was scheduled to take place in ICES division 9a after Period 2. Sampling of the western area commenced in Period 3, and included coverage of the west of Scotland, west of Ireland, Biscay and the Cantabrian Sea. Surveying in the Cantabrian Sea ended at the end of Period 5. In Periods 6 and 7 the surveys were designed to identify a southern boundary of spawning and to survey all areas north of this boundary.

Maximum deployment of effort in the western area was during Periods three, four, five and six. Historically these periods would have coincided with the expected peak spawning of both mackerel and horse mackerel. Recent years have seen mackerel peak spawning taking place during Periods 3 and 5 .

Due to the expansion of the spawning area which has been observed since 2007 the emphasis was even more focused on full area coverage and delineation of the spawning boundaries. Cruise leaders had been asked to cover their entire assigned area using alternate transects and then use any remaining time to fill in the missed transects.

Table 1. Participating countries, vessels, areas covered, dates and sampling periods of the 2022 surveys.

| Country | Vessel | Area | Dates | Period |
| :--- | :--- | :--- | :--- | :--- |
| Portugal | Vizconde de Eza | Portugal | Jan 23 | Feb 26 |

## Processing of samples

The analysis of the plankton and fecundity samples were carried out according to the sampling protocols as described in the WGMEGS Manuals for Survey (ICES, 2019a) and Fecundity (ICES, 2019b).

A total of 1780 plankton samples were collected and sorted. Mackerel and horse mackerel eggs were identified and the egg development stages determined. Depending on the vessel facilities and the experience of the participants this was done either during the cruise or back in the national institutes.

Double micropipette samples and slices from ovaries of mackerel were taken during each survey. Additional samples were collected during periods 3 and 4 by participants in an effort to carry out DEPM analysis, along with AEPM analysis. Fecundity sampling for horse mackerel only took place during the expected peak spawning Periods, 6 and 7.

In order to increase the number of samples available for fecundity analysis additional mackerel gonads were collected from some Dutch pelagic vessels, and also on the Dutch and Irish Blue whiting surveys in Periods 2, 3 and 4.

After each survey the ovary screening and fecundity samples were shared between the participating research institutes for histological and whole mount analysis to determine the realised fecundity (potential fecundity minus atresia). Screening samples, and fecundity samples, have to be analysed in the laboratory upon return from sea. These procedures are not straightforward and require time. The last histology samples were collected in July and because of the narrow time frame only a selection of the fecundity samples have been analysed up to this date. Samples were therefore only analysed from sampling Periods 2 and 3 for the preliminary estimate.

Horse mackerel is considered to be an indeterminate spawner and therefore since 2007 IPMA has adopted the DEPM methodology for the southern horse mackerel stock (div. 9a). The egg survey design in the western area is directed at the AEP method for mackerel which produces an estimate of SSB. Fecundity samples for horse mackerel were taken during the survey in the western areas in order to develop a modified DEPM approach for estimating the biomass of the horse mackerel stocks. Additional samples were collected during the Irish WESPAS survey in the Celtic Sea and west of Ireland in Periods 6 and 7.

Even though the partial processing of the screening samples has identified ovaries to be analysed for DEPM, none of these samples have been analysed yet.

## Survey coverage and mackerel egg production by period

Period 2 - Portugal started the 2022 survey series on January $23^{\text {rd }}$. This is a DEPM survey mainly targeting the southern horse mackerel stock and is designed for this purpose, but it provides mackerel egg samples as well. The survey is usually undertaken between Cadiz and Galicia and is confined to ICES division 9a.

Period 3 - Period 3 marks the commencement of the western area surveys as well as a continuation of sampling in the southern area. Sampling was undertaken by Ireland (West of Scotland, west of Ireland, Celtic Sea), Germany (Celtic Sea) and AZTI (northern Biscay). Further south the Bay of Biscay, Cantabrian Sea and Galicia were covered by Spain (IEO).

No eggs were found by Ireland in northern waters so after a number of days the vessel turned south and sampled in the Celtic sea. Due to issues with Covid cases among the crew the German survey was delayed starting, however it subsequently linked with the Irish vessel. Both IEO and AZTI suffered difficulties with their vessels, and lost a number of sampling days, however full coverage was achieved (Fig. 1.1).

Egg numbers were quite low to the west of Ireland, however further south large numbers of eggs were found close to the 200 m contour line. In Biscay and the Cantabrian Sea IEO and AZTI recorded a number of stations with large egg numbers. 298 stations were sampled and there were only 13 interpolations. There were 52 replicate samples with the majority being completed in the Cantabrian Sea.

Period 4 - This period was covered by three surveys. Scotland sampled the area from the northwest of Ireland to the Shetland islands. Germany surveyed west of Ireland, Celtic sea and northern Biscay while IEO completed the survey coverage in southern Biscay and the Cantabrian Sea (Fig. 1.2).

Due to difficulties in acquiring diplomatic clearance the Scottish survey was unable to sample in Irish waters. As a result Germany extended their survey area to ensure continuity of sampling coverage.

Once again moderate levels of eggs were recorded throughout the area, with the highest concentrations still being found close to the 200 m contour line. Large egg numbers were recorded to the west of Scotland, however numbers were lower than those reported for 2019 within this area and time period. 327 stations
were sampled and there were 46 interpolations. 52 replicate samples were taken and once again most of these were collected from the Cantabrian Sea.

Period 5 - In Period 5, the entire spawning area from the Cantabrian Sea to the West of Scotland, and up to Faroese waters at around $61^{\circ} \mathrm{N}$ was surveyed by AZTI, the Netherlands, Scotland, and Faroes.

Spawning in the Cantabrian Sea was tailing off with only low egg numbers being found. Throughout Biscay and into the southern Celtic Sea numbers were generally low to moderate (Fig. 1.3). This pattern continued west of Ireland, to around $54^{\circ} \mathrm{N}$, with spawning remaining on and around the Shelf edge. North of this however, and similar to that noted in 2016 and 2019, spawning activity fanned out both westwards and northwards. Due to the large area Scotland had to survey their vessel was forced to restrict exploration of the western boundary around the SW of Rockall Bank. Egg numbers in 2022 within this area were lower than reported in 2019 so while the western boundary wasn't delineated, MEGS is happy that major egg production isn't being missed. North of this the Faroese survey completed stations North of Hatton Bank and up towards the Icelandic coast. Some egg production was found to the north of Rockall, however the largest number of eggs were encountered west of the Shetlands. In total 444 stations were sampled and there were 214 interpolations. No replicate samples were taken.

Period 6 - During period 6 northern Biscay, from $46^{\circ} N$ and also the Celtic Sea were covered by the Netherlands while Ireland was to cover west of Ireland and also west of Scotland. Norway surveyed the area north of $59^{\circ} \mathrm{N}$ from the south of Iceland to the Norwegian coast, as well as carrying out four transects in the northern North Sea to assist England and Denmark provide full coverage for the DEPM survey.

Ireland was due to charter a research vessel from Northern Ireland to conduct the survey. One week before the survey was due to depart this vessel had to go to dry dock for emergency repairs. After much searching a smaller Welsh RV was contracted. Once at sea however it quickly became clear that the replacement vessel was not going to be suitable for the survey. Only two successful stations were carried out before a decision was eventually made to abandon the survey. Norway and Netherlands both completed their survey sampling successfully.

Low levels of spawning were observed in Biscay and to the south to the West of Ireland and Porcupine bank (Fig. 1.4). Similarly in the northern area spawning was persistent at low levels, apart again from the area west of the Shetland. Due to an unavoidable reduction in the number of survey days available Norway was unable to secure either the western or northern boundary in the northern area, however Netherlands secured the western boundary in their area. 184 stations were sampled with 36 interpolations. No replicate stations were completed.

Period 7 - This period was covered entirely by Scotland sampling on alternate transects in the area from $47^{\circ} 15 \mathrm{~N}$ in the south to north of the Hebrides and $59^{\circ} \mathrm{N}$ (Fig. 1.5). Due to the lack of eggs encountered the Scottish survey adhered very closely to the 200 m contour and 144 stations were sampled with 24 interpolations. 2 replicate station was completed. Only very low levels of spawning were observed and these were confined to the continental shelf and shelf edge with all spawning boundaries being delineated successfully.


Figure 1.1: Mackerel egg production by half rectangle for period 3 ( $\mathrm{Mar} 4^{\text {th }}-\mathrm{Apr} 8^{\text {th }}$ ). Circle areas and colour scale represent mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 1.2: Mackerel egg production by half rectangle for period 4 (Apr $9^{\text {th }}-29^{\text {th }}$ ). Circle areas and colour scale represent mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 1.3: Mackerel egg production by half rectangle for period 5 (Apr $30^{\text {th }}-$ May $31^{\text {st }}$ ). Circle areas and colour scale represent mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 1.4: Mackerel egg production by half rectangle for period 6 (June $1^{\text {st }}-30^{\text {th }}$ ). Circle areas and colour scale represent mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 1.5: Mackerel egg production by half rectangle for period 7 (July $1^{\text {st }}-31^{\text {st }}$ ). Circle areas and colour scale represent mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.

## 2 Results - MACKEREL

## Stage 1 Egg production in the Western Areas

The cancelling of the Irish survey in period 6 was addressed by MEGS. The group estimated the spawning area that was missed and also estimated mean daily egg production for the period. The survey area from 53 N to 61 N , and 3.5 W to 21 W was looked at for the 2013, 2016 and 2019 surveys. Positive stations were selected where stage 1 eggs were found in a rectangle on at least two occasions over these three surveys (Fig. 2.1, blue rectangles). MEGS estimated this amounted to 127 missed stations during the period and also estimated mean daily egg production for period 6 in 2022 at 19.58 stage $1 \mathrm{eggs} / \mathrm{m}^{2} /$ day. Figure 2.2 shows the spawning curve for 2022 , with and without the correction for the Irish survey.
2010 provided an unusually large spawning event early in the spawning season, 2013 yielded an even larger spawning event indicating that spawning was probably taking place well before the nominal start date of $10^{\text {th }}$ February (Fig. 2.3). In 2016 the first survey commenced on February $5^{\text {th }}$ which is five days prior to the nominal start date. That year however mackerel migration was later and slower than that recorded in the previous two surveys (Fig. 2.3 \& Table 2).
In 2016 concern was expressed that survey coverage may have underestimated the total egg production estimate. The expansion observed in western and northwestern areas during Periods 5 and 6 in 2016 was once again reported during 2022, however this year production in Periods 5 and 6 was lower in these northwestern areas. The 2022 spawning curve is very similar to that of 2016, with peak spawning again occurring during Period 5. Annual egg production since 1992 is shown in Figure 2.4. Mackerel egg production by period since 2004 is shown in Figure 2.5.

In 2017 and 2018 MEGS organised exploratory egg surveys in this region. These surveys provide significant evidence that while some spawning has been missed the loss of egg abundance is not sufficiently large to significantly impact the SSB estimate.
Overall, the inclusion of the estimated egg abundance for the missing stations in Period 6 has a impact of $10 \%$ on the annual egg production 2022.


Figure 2.1: Area, blue colour, from period 6 where it is estimated eggs would have been found


Figure 2.2: 2022 spawning curve showing uncorrected (left) and corrected (right) egg estimates for Period 6 (black line). The left hand plot shows the data from the Netherlands and Norwegian surveys. The right hand plot includes the addition of the estimated egg abundance calculated for the missing Irish Period 6 survey.

The nominal end of spawning date of the $31^{\text {st }}$ July is the same as was used during previous survey years and the shape of the egg production curve for 2022 does not suggest that the chosen end date needs to be altered. The provisional total annual egg production (TAEP) for the western area in 2022 was calculated as 1.795 * $\mathbf{1 0}^{\mathbf{1 5}}$ (Table 2). This is a $47 \%$ increase on the 2019 TAEP estimate which was $1.22 * 10^{15}$.


Figure 2.3: Provisional annual egg production curve for mackerel in the western spawning component in 2022, (black line). The curves for 2010, 2013, 2016 and 2019 are included for comparison.


Figure 2.4: Provisional annual egg production for 2022 for the western spawning component.

Bars from 1992 are included for comparison.


Figure 2.5: Egg production by period for the western spawning component since 2004

Table 2. Western estimate of mackerel total stage I egg production by period using the histogram method for 2022.

| Dates | Period | Days | Annual stage I egg production * $10{ }^{15}$ |
| :---: | :---: | :---: | :---: |
| Feb $5^{\text {th }}-$ Mar $3^{\text {rd }}$ | Pre 3 | 31 | 0.09 |
| Mar $4^{\text {th }}-$ April $8^{\text {th }}$ | 3 | 36 | 0.325 |
| Apr $9^{\text {th }}-$ April $26^{\text {th }}$ | 4 | 18 | 0.120 |
| April $27^{\text {th }}-$ Apr $29^{\text {th }}$ | 4-5 | 3 | 0.043 |
| Apr 30 ${ }^{\text {th }}-$ May $31^{\text {st }}$ | 5 | 32 | 0.853 |
| Jun $1^{\text {st }}-5^{\text {th }}$ | 5-6 | 5 | 0.067 |
| Jun $6^{\text {th }}-$ June $22^{\text {nd }}$ | 6 | 17 | 0.21 |
| June $23{ }^{\text {rd }}$ - July $4^{\text {th }}$ | 6-7 | 12 | 0.081 |
| July $5^{\text {th }}$ - July $25^{\text {th }}$ | 7 | 21 | 0.007 |
| July $26^{\text {th }}-31^{\text {st }}$ | Post 7 | 6 | 0.0003 |
| Total |  |  | 1.795 |

## Stage 1 Egg production in the Southern Areas

The start date for spawning in the southern area was the $23^{\text {rd }}$ January (Table 3). Portugal surveyed in Period 2 in division 9a. Sampling in the Cantabrian Sea where the majority of spawning occurs within the Southern area commenced on the $18^{\text {th }}$ March. The same end of spawning date of the $17^{\text {th }}$ July was used again this year and the spawning curve suggests that there is no reason for this to change (Fig. 2.4). As in 2019 the survey periods were not completely contiguous and this has been accounted for (Table 3). The mackerel egg production by period since 2004 is shown in Figure 2.6. The provisional total annual egg production (TAEP) for the southern area in 2022 was calculated as $\mathbf{3 . 2 1}{ }^{*} \mathbf{1 0}^{\mathbf{1 4}}$ (Table 3). This is a $\mathbf{2 5 \%}$ decrease on the 2019 TAEP estimate which was $4.23 * 10^{14}$ (Fig. 2.5).


Figure 2.4: Provisional annual egg production curve for mackerel in the southern spawning component for 2022, black line). The curves for 2010, 2013, 2016 and 2019 are included for comparison.


Figure 2.5: Provisional annual egg production for the southern spawning component for 2022.
Bars from 1992 are included for comparison.


Figure 2.6: Egg production by period for the southern spawning component since 2004

Table 3. Southern estimate of mackerel total stage I egg production by period using the histogram method for 2022.

| Dates | Period | Days | Annual stage I egg production $\times 10{ }^{14}$ |
| :---: | :---: | :---: | :---: |
| Feb $1^{\text {st }}-$ Mar $17^{\text {th }}$ <br> March $18^{\text {th }}-$ April $2^{\text {nd }}$ <br> April $3^{\text {rd }}$ <br> April $4^{\text {th }}-25^{\text {th }}$ <br> Apr $26^{\text {th }}-$ May $1^{\text {st }}$ <br> May $2^{\text {nd }}-4^{\text {th }}$ <br> May $5^{\text {th }}-J u l y ~ 17^{\text {th }}$ | $\begin{aligned} & 2-3 \\ & 3 \\ & 3-4 \\ & 4 \\ & 4-5 \\ & 5 \\ & \text { Post } 5 \end{aligned}$ | $\begin{aligned} & 45 \\ & 16 \\ & 1 \\ & 22 \\ & 6 \\ & 71 \\ & 3 \end{aligned}$ | 1.52 1.27 0.052 0.323 0.026 0.003 0.014 |
| Total | 3.212 |  |  |

## Total egg production

Total annual eggs production (TAEP) for both the western and southern components combined in 2022 is $\mathbf{2 . 1 1 6}$ * $\mathbf{1 0}^{\mathbf{1 5} .}$ (Fig. 2.3). This is an increase in production of $\mathbf{2 9 \%}$ compared to $2019,1.64 * 10^{15}$ (Fig. 2.3).


Figure 2.3: Combined mackerel TAEP estimates $\left({ }^{*} 10^{13}\right)$ - 1992 - 2022.

## Fecundity - Preliminary estimates

## Adult Parameters

Fecundity Sample distribution

Atlantic mackerel samples were collected during periods 2-7 spread over an area with a bounding box of $59.36 \mathrm{~N} 14.20 \mathrm{~W}-36.54 \mathrm{~N} 2.32 \mathrm{~W}$. Nine institutes participated. The histological screening of samples was performed by five institutes while fecundity was analysed by six of them.

As usual for the preliminary report, only samples from Periods 2 and 3 were selected. This is because there is not enough time to analyse samples from the other periods. For the final report samples from the other periods will be included also. Experience from earlier surveys is that the preliminary estimate and the final estimate is close.

## Screening

Potential fecundity counts were based on whole mount samples taken from maturing females which had not
started spawning. To select these samples, a histological screening procedure was used followed by a screening procedure on the selected whole mount samples.

A total of 918 samples were screened, of which 793 were from periods 2 and 3 (
Table ). Of those, 482 samples showed spawning markers, i.e. migratory nucleus stage (MIG), hydrated oocytes, eggs, and post ovulatory follicles (POFs). A total of 175 samples from periods 2-3 showed presence of atresia without considering those that were classified as "spent" or having "massive atresia".

From previous survey reports we know that POF scoring has varied considerably between periods. WKFATHOM2 (2018) discussed this issue and came up with more detailed criteria for POF staging. Looking at screening results from 2022, POFs were identified less frequently than in 2019 for periods 2 and 3, i.e. 58 \% vs 74\% (Table 4).

Table 4. POF scoring using histology by periods 2-3.

| Period | Screened | Spawning <br> Markers | POFs | Fecundity <br> Histology | Fecundity <br> Whole <br> mount | Atresia <br> Presence |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 32 | 24 | 21 | 2 | 2 | 3 |
| 3 | 675 | 541 | 494 | 38 | 33 | 156 |

Results from previous surveys showed that POF scoring could vary considerably between periods. At WKFATHOM2 (ICES 2018) this issue was discussed and more detailed criteria for POF staging were elaborated. Looking at screening results from 2022, POFs were identified less frequently than in 2019 for periods 2 and 3, i.e. $58 \%$ vs $74 \%$ (Table 5).

Table 5. POF scoring using histology (Periods 2-3).

| Period | No POF | POF | \%POF | \%POF <br> 2019 |
| :--- | :--- | :--- | :--- | :--- |
| 2 | 66 | 55 | 52 | 66 |
| 3 | 260 | 404 | 60 | 74 |
| $2-3$ | 326 | 459 | 58 | 74 |

A total of 159 samples from periods 2-3 showed presence of atresia without considering those that were classified as "spent" or having "massive atresia" (Table ).

Looking at the oocyte stage most of the samples in periods 2-3 were at MIG or hydrated oocyte stage ( $\mathrm{n}=545$ ) and that less than half $(\mathrm{n}=217)$ were in vitellogenic oocyte stage.

## Potential fecundity

For the 2022 preliminary estimate of potential fecundity, 169 samples were available, which represents $21 \%$ of all samples screened for periods 2 and 3 . This number is much higher than in 2019 , when 34 samples were available for the preliminary report.

The potential fecundity estimate is based on samples from pre-spawning fish. The pre-spawning status is confirmed using a detailed histology screening procedure that detects the most advanced oocyte stage (stage 1-5) as well as spawning markers (POF's, post ovulatory follicles and eggs). This year the fecundity estimate is based on samples that may also include the MIG oocyte stage. This is different from previous surveys (in recent time) where the most advanced oocyte stage included was stage 3 (advanced vitellogenesis). However, the MIG oocyte stage is not a true spawning marker, but a marker that shows that spawning likely will take place within a few days. For previous surveys samples with MIG's were excluded for precautious reasons.

Since the 2013 MEGS survey, the median has been used for relative fecundity estimation rather than the mean which was used previously. The reason for the change is related to the fact that that unlike the mean, the median is not influenced by extreme values. A posterior analysis showed that the median for relative potential fecundity was close to the arithmetic mean in most years. The largest difference was in 2013, but even then, the median was within the confidence interval of the potential fecundity arithmetic mean. WGMEGS 2018 (ICES 2018) discussed whether to use the trimmed mean instead of the median for the potential fecundity estimate. A trimmed mean is preferred for calculation of confidence intervals. However, until the time-series data is reanalyzed in the near future, it was decided that the relative fecundity estimate should still be based on the median rather than the mean.

The distribution of relative potential fecundity values (Figure 2.4) was close to a normal distribution and ranged from 623 to 1972 ( $\mathrm{n} / \mathrm{g}$ ). The distribution was almost similar both for samples with the MIG oocyte stage (stage 4) and stage 3 (Figure 2.4). The median value for stage 3 samples was 1247 (mean 1282, SD 290) while for the MIG stage the median was 1256 (mean 1300, SD 267). This shows that including samples with MIG's in the fecundity estimate have not significantly changed the median or mean value, and that our previous cautious procedure excluding MIG's is probably unnecessary.


Figure 2.4. Relative fecundity preliminary estimation in 2022. The panels show the distribution (in \%) of relative fecundity using samples in which the most advanced oocyte stage present was 3 (advanced vitellogenesis, top panel), samples where the most advanced oocyte stage was MIG (stage 4, middle panel) and the combined histogram (bottom panel).

The preliminary relative potential fecundity in 2022 was slightly higher than in 2019 (1253 and 1191, respectively)

Table 6 Estimate of relative fecundity ( $\mathrm{n} / \mathrm{g}$ fish) and statistics.

| Year | N | Median | Mean | sd | Max | Min | $95 \%$ CI |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2022 | 169 | 1253 | 1288 | 283 | 1972 | 623 | $1252-1324$ |
| 2019 | 34 | 1215 | 1263 | 285 | 2029 | 564 | $1163-1362$ |

## Biological data of fish samples to fecundity

The distribution of fish length, weight, Fulton's condition factor ( $100 \times$ weight/length ${ }^{3}$ ), and gonad-somatic index (GSI; $100 \times$ Ovary weight/Fish weight) is shown in Figure 2.5.

Similar to the previous surveys only fish with condition factor between 0.5 and 1.2, and GSI between 1 and 25 were included (ICES 2014) in the fecundity and atresia estimates. For this preliminary estimation, no females needed to be excluded from the analysis based on these biological parameters.


Figure 2.5. Fish length and weight, Fulton's Condition index and GSI of individuals analysed for fecundity.

## Atresia

Atresia is the loss of oocytes by reabsorption before spawning and must be subtracted from the potential fecundity (whole mount fecundity counting) to estimate the realised fecundity. In this preliminary report, intensity of atresia can not be presented due to the time consumed for the histology screening.

The prevalence of atresia estimated by histological screening may however be a good indicator of the level of atresia. Prevalence of atresia is defined as the percentage of spawning fish which have early stage atresia (early
alpha-atresia). Among the 559 samples considered the prevalence of atresia estimated was 0.28 , (fish from period 2-3, excluding spent fish and fish with massive atresia).

## Realised fecundity

Realised fecundity is defined as the potential fecundity minus the loss by atresia. The loss by atresia is a function of both intensity of atresia and prevalence of atresia. The intensity of atresia for 2022 is still unavailable, therefore the loss was calculated from the average loss from the surveys since 2001 (Table ). The relative loss by atresia from this period (2001-2019) ranged from 6-9\% (average 6\%).

Based on this, the preliminary realised fecundity-estimate for 2022 was 1178 oocytes/gram female. The estimate is well within the observed range of realized fecundity (1009-1209, average 1087 egg per gram female) from all previous surveys back to 2001 (Table 7). For the three most recent surveys, realized fecundity varied between 1087 and 1209 eggs per gram female (average 1148).

Table 7. Summary table of mackerel fecundity and atresia by survey year.

|  | Survey year |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  | 2022 |
|  |  |  |  |  |  |  |  | Prel. |
| Fecundity samples ( n ) | 187 | 205 | 176 | 74 | 132 | 97 | 62 | 169 |
| Prevalence of atresia ( n ) | 290 | 348 | 416 | 511 | 732 | 713 | 895 | 559 |
| Intensity of atresia ( n ) | 290 | 348 | 416 | 511 | 56 | 66 | 64 |  |
| Relative potential fecundity ( $\mathrm{n} / \mathrm{g}$ ) | 1097 | 1127 | 1098 | 1140 | 1257* | 1159* | 1191* | 1253* |
| Prevalence of atresia | 0.2 | 0.28 | 0.38 | 0.33 | 0.22 | 0.3 | 0.28 | 0.28 |
| Geometric mean intensity of atresia ( $\mathrm{n} / \mathrm{g}$ ) | 40 | 33 | 30 | 26 | 27 | 30 | 20 |  |
| Potential fecundity lost per day ( $\mathrm{n} / \mathrm{g}$ ) | 1.07 | 1.25 | 1.48 | 1.16 | 0.8 | 1.2 | 0.73 |  |
| Potential fecundity lost ( $\mathrm{n} / \mathrm{g}$ ) | 64 | 75 | 89 | 70 | 48 | 72 | 44 | 75 |
| Relative potential fecundity lost (\%) | 6 | 7 | 9 | 6 | 4 | 6 | 4 | 6 |
| Realised fecundity ( $\mathrm{n} / \mathrm{g}$ )* | 1033 | 1052 | 1009 | 1070 | 1209 | 1087 | 1147 | 1178 |

*Median not mean relative potential fecundity.

## Biomass estimation

Total spawning stock biomass (SSB) was estimated using a preliminary fecundity estimate of 1178 oocytes/g female, a sex ratio of 1:1 and a raising factor of 1.08 (ICES, 1987) to convert pre-spawning to spawning fish. This gave an estimate of spawning stock biomass of:

- 3.292 million tonnes for western component (2019: 2.29).
- 0.589 million tonnes for southern component (2019: 0.80).
- 3.881 million tonnes for western and southern components combined (2019: 3.09)


## 3 Results - HORSE MACKEREL

## Horse mackerel egg production by period

Period 3 - In period 3 horse mackerel spawning started in the Cantabrian Sea and southern Biscay, but numbers of eggs found were very low. Higher spawning took place in the Celtic Sea but numbers were still low (Fig. 3.1).

Period 4 - Horse mackerel spawning continued in the Cantabrian Sea, extending into southern Biscay. Eggs were again found in the Celtic Sea but numbers were lower than in period 3 (Fig. 3.2).

Period 5 - Horse mackerel spawning continues in the Cantabrian Sea, Celtic Sea and northern Bay of Biscay, but still in low numbers. Some eggs were also found south and west of Ireland (Fig. 3.3).

Period 6 -Spawning continued in northern Biscay, the Celtic Sea and to the southwest of Ireland. For the first time in a number of years large numbers of eggs were reported in a number of stations close to the 200 m contour. Peak spawning took place in this period (Fig. 3.4).

Period 7 - Eggs were found from northern Biscay to west of Scotland, being concentrated off the southwest of Ireland. In general egg numbers were low but occasional stations with moderate to high counts were observed (Fig. 3.5).


Figure 3.1: Horse mackerel egg production by half rectangle for period 3 (March $4^{\text {th }}-$ April $8^{\text {th }}$ ). Circle areas and colour scale represent horse mackerel stage I eggs $/ \mathrm{m}^{2} /$ day by half rectangle. Crosses represent zero values.


Figure 3.2: Horse mackerel egg production by half rectangle for period 4 (April $9^{\text {th }}-29^{\text {th }}$ ). Circle areas and colour scale represent horse mackerel stage I eggs $/ \mathrm{m}^{2} /$ day by half rectangle. Crosses represent zero values.


Figure 3.3: Horse mackerel egg production by half rectangle for period 5 (Apr 30 ${ }^{\text {th }}-$ May $31^{\text {st }}$ ). Circle areas and colour scale represent horse mackerel stage I eggs $/ \mathrm{m}^{2} /$ day by half rectangle. Crosses represent zero values.


Figure 3.4: Horse mackerel egg production by half rectangle for period 6 (June $1^{\text {st }}-30^{\text {th }}$ ). Circle areas and colour scale represent horse mackerel stage I eggs/m²/day by half rectangle. Crosses represent zero values.


Figure 3.5: Horse mackerel egg production by half rectangle for period 7 (July $1^{\text {st }}-$ July $31^{\text {st }}$ ). Circle areas and colour scale represent horse mackerel stage I eggs $/ \mathrm{m}^{2} /$ day by half rectangle. Crosses represent zero values.

## TAEP results - Western Horse Mackerel

Period number and duration are the same as those used to estimate the western mackerel stock, as are the dates defining the start and end of spawning (Table 6). The shape of the egg production curve does not suggest that those dates should be altered for 2022 (Fig. 3.6). An exercise, similar to the one carried out for mackerel in period 6 , was not carried out for horse mackerel as MEGS feel that the Netherlands period 6 survey delineated the northern boundary of horse mackerel spawning during this period. The total annual egg
 which was the lowest estimate of annual egg production ever recorded for this species (Fig. 3.7). Horse mackerel egg production by period since 2007 is shown in Figure 3.8.


Figure 3.6: Provisional annual egg production curve for western horse mackerel for 2022, (black line). The curves for 2010, 2013, 2016 and 2019 are included for comparison.

Western horse mackerel Annual egg production


Figure 3.7: Provisional total annual egg production for western horse mackerel. Production figures back to 1992 are included for comparison.

Western horse mackerel Egg production by period


Figure 3.8: Egg production by period for the western horse mackerel spawning component since 2007

Table 6: Western estimate of horse mackerel total stage I egg production by period using the histogram method for 2022.

| Dates | Period | Days | Annual stage I egg production * $10{ }^{15}$ |
| :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { Feb } 1^{\text {st }}-\text { Mar }^{\text {rd }} \\ & \text { Mar } 4^{\text {th }}-\text { April } 8^{\text {th }} \\ & \text { Apr } 9^{\text {th }}-26^{\text {th }} \\ & \text { Apr } 27^{\text {th }}-29^{\text {th }} \\ & \text { Apr } 30^{\text {th }}-\text { May } 31^{\text {st }} \\ & \text { Jun } 1^{\text {st }}-5^{\text {th }} \\ & \text { Jun } 6^{\text {th }}-22^{\text {nd }} \\ & \text { June } 5-{\text { July } 4^{\text {th }}}^{\text {Jun }} \end{aligned}$ |  Pre 3 <br> 3  <br> $4-5$  <br> 5  <br> $5-6$  <br> 6  <br> $6-7$  | $\begin{aligned} & 31 \\ & 36 \\ & 18 \\ & 3 \\ & 32 \\ & 3 \\ & 5 \\ & 17 \\ & 12 \end{aligned}$ | 0.016 0.055 0.016 0.003 0.038 0.043 0.223 0.091 |
| July $5^{\text {th }}-25$ th | 7 | 21 | 0.028 |
| July $26^{\text {th }}-31^{\text {st }}$ | Post 7 | 6 | 0.001 |
| Total |  |  | 0.514 |

## Fecundity investigations

This year for horse mackerel only DEPM ovary samples were collected during Periods 6 and 7, during peak of spawning. In addition to those samples collected during the MEGS surveys additional samples were collected from the Irish WESPAS surveys in periods 6 and 7 . Since horse mackerel fecundity is at this moment not used for estimating the spawning stock biomass the focus of the fecundity analysis has been on mackerel. Therefore, at this time no horse mackerel fecundity results are ready to be presented. All samples will be analysed and results presented at the 2023 WGMEGS meeting.

## DEPM results -Western Horse Mackerel

The horse-mackerel egg data of the DEPM survey are still under revision. Samples will be analyzed before and results will be presented to the 2023 WGMEGS meeting.

## 4 Discussion

Since 2004 and subsequent to demands for up-to-date data for the assessment, WGMEGS has endeavored to provide an estimate of NEA mackerel biomass and western horse mackerel egg production within the same calendar year as the survey and in time for the assessment meetings taking place. This report represents the preliminary results of the 2022 egg survey. WGMEGS cannot guarantee that there will be no changes prior to the presentation of the final survey results at WGMEGS in April 2023. However, despite the tight deadline nearly all plankton samples were analyzed for mackerel (southern and western area) and horse mackerel (western area only) stage 1 eggs. Portugal still has to supply data for their Period 2 survey in division 9 a. Historically not many mackerel are caught during this survey therefore only negligible changes in the total egg production values are to be expected

As with 2019 no fecundity samples from Period 1 were available, instead samples from Periods 2 and 3 were included in the potential fecundity estimate. For the final fecundity estimate the later periods will also be included, as was done for previous surveys. No estimate of loss by atresia is yet available for 2022. The realised fecundity estimate is therefore based on the average atretic loss found in the period from 2001-2019. Since the atretic loss has always been a small number compared to the potential fecundity, using this average value will likely not give a large error. The prevalence of atresia for $2022(28 \%)$ is comparable to previous survey estimates, it is thus highly likely that the atretic loss will also be at the same level. Atretic loss will however be analysed and included in the final fecundity estimate at the WGMEGS meeting in 2023.

Previous surveys in 2010 and 2013 were dominated by the issue of the early peak of western mackerel spawning and its close proximity to the nominal start date. In 2016 peak spawning reverted to May / June, a time that would traditionally be considered normal. In 2019, peak spawning in the western area was found to have occurred slightly earlier in Period 4. For 2022 the spawning pattern is remarkably similar to that reported for 2016.

During 2016, high levels of spawning were recorded over a large area of the Northeast Atlantic with a large number of the stations being reported over deepwater and well away from the continental shelf. In 2019 numbers of stage 1 eggs recorded on these northerly and western boundary stations were much reduced, although still present. The expansion was repeated in 2022 during Periods 5 and 6 , however spawning densities recorded in these areas were significantly lower than reported in 2016 and 2019. Available surveys deployed during these periods were unable to fully delineate all boundaries however WGMEGS are satisfied that significant additional egg production is not being missed in these northern and western areas.

For the first time in a number of surveys western horse mackerel has shown an increase in egg production.

The MEGS group is confident that this survey accurately reflects the spawning patterns as exhibited by both species and as is presented in this working document. Despite the inability to secure a northern spawning boundary for western mackerel during periods 5 and 6 , results from the recent exploratory MEGS surveys undertaken within these regions and reported to WGWIDE in 2021 (ICES,
2021) provide reassurance that the fraction of spawning missed is a minor one and that the survey has indeed been successful in capturing the majority of spawning activity. The potential issue arising from the missing Irish survey has also been satisfactorily addressed.

## 5 References

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## Blue whiting

An updated alternative assessment including more surveys*

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## Introduction

During WGWIDE 2020 we saw how vulnerable a stock assessment is when we only have one survey input to base the assessment on, and that survey is cancelled. In 2020 it was due to the covid-19 pandemic, but in the future there might be other unforeseen events that may cause the survey being cancelled or something may go wrong in the data collection so that we do not have reliable data for a specific year. To avoid this issue of potentially having no fishery independent data and make the assessment more robust against problems with the IBWSS, we will in this report consider including the IESNS and IESSNS survey data for blue whiting in the assessment.

## Data description

For the IESNS survey we have data from 2008 to 2022 and for the IESSNS from 2016 to 2022 . We use ages from 1-4+ and 1-6+ from the two surveys. This age selection was made based on the consistency plots in Figure 4. From the original assessment, we also have catch data (ages 1-10+, 1981-2022) and the IBWSS (ages 1-8, 2004-2022), where 2010 and 2020 is missing. The model has been configured based on data available in 2020, but we will include everything that is available at the time of the WGWIDE 2022 meeting in 24.-30. August 2022. An overview of the data selected for the alternative assessment is found in Figure 5 and each time series is plotted in Figure 6 for each age group and Figure 7 for each year class.

## Model description

Today's assessment is using the R package stockassessment and the SAM model. Including additional survey data as input in this framework is a relatively simple task. The effort is mostly needed for deciding how to set up the configuration of the model. The procedure of how we have selected the model configuration is that we have included the two additional survey data sources and start out with a default SAM configuration. Then we start at the top of the configuration and make incremental changes and compare different settings until we get the best model fit in terms of AIC. Then we move on to the next configuration setting. We only consider configurations that are somewhat sensible. For instance, we do not consider putting the same catchability on 1 year old and 8 -year-old fish, with some other catchability for those in-between. We only consider cases where neighbouring age groups share the same parameters. The final configuration file is included in the appendix. For details on diagnostic, see appendix.

## Model output

Once we have fitted the model, we can look at model output. In Figure 1 we have plotted SSB, Fbar and recruitment for the period 1980-2022 according to the fitted model. The black line with grey confidence

[^10]interval is the official WGWIDE2022 assessment model for comparison.
In terms of SSB, the two models follow each other closely, with a slightly lower SSB for the alternative model in recent years. The main difference is clearly that we get smaller confidence intervals, i.e. higher accuracy, by adding more data to the model. For Fbar the picture is more or less the same. The two models are close to each other, only the alternative model point estimate is higher than WGWIDE for the last 3-4 years.In recruitment we see a bigger discrepancy. The alternative model gives a higher recruitment in 2016 and also for the last two years, 2021 and 2022. This is most likely due to high values for these years in the two additional survey indices. The confidence intervals are narrower for the alternative model compared to WGWIDE2022. Hence, the alternative assessment is consistent with the WGWIDE2022 assessment, but it has higher accuracy.

## Leave-out analysis

A standard diagnostic is to leave out one survey at the time and see what effect this has on the output. This is achieved by taking out one data source at the time and refitting the model. This can give us an idea of how that particular data source affects the total. The leaveout plots are presented in Figure 2.

For the SSB the differences are small and the four curves are close to each other. If we take out IESSNS the SSB is slightly lower and if we take out IESNS it increases in the recent years. Taking out IBWSS increases the uncertainty the most, which is natural as it is the largest survey in terms of observations. We also see a similar pattern for Fbar. For the recruitment, taking out IESSNS will give the lowest recruitment, while if we take out IBWSS we get the highest for 2021. Going back in time, the leaveout scenarios give more or less the same result.

Another interesting scenario we can run is: What if we take out all the surveys and run the SAM model with only catch data. The results of such a model run is presented in Figure 3 compared to the WGWIDE2022 assessment. In short, it gives a lower point estimate for SSB and Recruitment and higher Fbar. It also widens the confidence intervals when taking out all surveys.

## Conclusion

This exploratory model run shows that it is possible to include IESNS and IESSNS into the SAM model for Blue Whiting. It reduces the uncertainty and may provide more information about the younger fish. It will certainly reduce the risk for not having any survey to base the assessment on, by having two-three surveys instead of just one. The data is already being collected, and ready to use.

## Appendix

## Diagnostics

## Jit run

A jitter run means that we re-estimate the model using randomly selected initial values and report the maximum difference in each parameter and model output. Ideally there should not be any major changes due to the initial values. The results from the jitter run indicates that there is little effect on the different model parameters due to varying the initial values.

| \#\# | $\max (\mid$ delta\| $)$ |
| :--- | :---: |
| \#\# logFpar | $1.708855 \mathrm{e}-12$ |
| \#\# logSdLogFsta | $1.119327 \mathrm{e}-12$ |
| \#\# logSdLogN | $3.281819 \mathrm{e}-13$ |
| \#\# logSdLogObs | $3.246112 \mathrm{e}-12$ |
| \#\# logSdLogTotalObs | $5.225820 \mathrm{e}-12$ |
| \#\# transfIRARdist | $1.073452 \mathrm{e}-11$ |
| \#\# itrans_rho | $2.763567 \mathrm{e}-12$ |





Figure 1: Model output in terms of SSB, Fbar and recruitment with 95 percent confidence intervals.




Figure 2: Leaveout plots for alternative assessment.




Figure 3: Comparison of assessment with catch only vs WGWIDE2021 assessment.

| \#\# logFScaleMSY | $6.995491 \mathrm{e}-01$ |
| :--- | ---: |
| \#\# implicitFunctionDelta | $5.901470 \mathrm{e}-01$ |
| \#\# logScaleFmsy | $5.912467 \mathrm{e}-01$ |
| \#\# logScaleFmax | $5.875044 \mathrm{e}-01$ |
| \#\# logScaleF01 | $6.409693 \mathrm{e}-01$ |
| \#\# logScaleFcrash | $6.993916 \mathrm{e}-01$ |
| \#\# logScaleFext | $5.391787 \mathrm{e}-01$ |
| \#\# logScaleFlim | $6.193291 \mathrm{e}-01$ |
| \#\# logF | $1.629119 \mathrm{e}-10$ |
| \#\# logN | $2.133138 \mathrm{e}-10$ |
| \#\# missing | $2.507221 \mathrm{e}-10$ |
| \#\# ssb | $5.337661 \mathrm{e}-04$ |
| \#\# fbar | $4.384876 \mathrm{e}-11$ |
| \#\# rec | $9.316012 \mathrm{e}-03$ |
| \#\# catch | $8.541672 \mathrm{e}-05$ |
| \#\# logLik | $2.537490 \mathrm{e}-10$ |

## Simulation study

Another test is to do a simulation study, where we simulate the processes going into the model and compare this to the model output based on the observations. Ideally, the simulations should stay within the $95 \%$ confidence intervals with a probability of 0.95 . Here we use 50 simulations. It seems that most of the simulations fall within the confidence intervals, with some exceptions. This is expected.

## Retrospective plots

Peeling off one year at the time and fitting the model based on those data. In the retrospective plots (Figure 13) we can see how well the last year's assessment fits with what the model predicts with one more year of data. Mohn's $\rho$ for the retrospective analysis of SSB, Fbar and recruitment is respectively, 0.0069, -0.0094 and -0.0736.

## Figures



Figure 4: Internal consistency/correlation plots for IBWSS, IESNS and IESSNS. We use $\log (x+1)$ to avoid issues when $x$ is 0 . For IBWSS ages 1-8 are used, while in the alternative model $1-4+$ and $1-6+$ is used for IESNS and IESSNS, respectively.


Figure 5: Dataplot showing for which ages and years we use observations from the different data sources. For all except IBWSS the oldest age group is a plus group.


Figure 6: Time series for all data sources on log scale - one line per age group.


Figure 7: Time series of the different data sources on log scale - one line per year class.

## Config

Here we print out the configuration file for the alternative assessment.

```
print(conf)
## $minAge
## [1] 1
##
## $maxAge
## [1] 10
##
## $maxAgePlusGroup
## [1] 1 0 1 1
##
## $keyLogFsta
## V1 V2 V3 V4 V5 V6 V7 V8 V9 V10
## [1,] 0
## [2,] -1 -1 -1 -1 -1 -1 -1 -1 -1 
## [3,] -1 -1 1
## [4,] -1 -1 1
##
## $corFlag
## [1] 2
##
## $keyLogFpar
## V1 V2 V3 V4 V5 V6 V7 V8 V9 V10
## [1,] -1 -1 -1 -1 -1 -1 -1 -1 -1 
```





Figure 8: QQ-normality plots for model residuals by data source.


Figure 9: QQ-normality plots for model residuals by data source.


Figure 10: Boxplots of residuals by age for each fleet.


## IESNS



## IBWSS



Figure 11: Correlation plot (model estimated).


## IESNS



## IBWSS



IESSNS


Figure 12: Empirical correlation plot.




Figure 13: Retrospective plots for SSB, Fbar and Recruitment.

```
## [2,] 0
## [3,] 5 5 6 7 7 7 7 -1 -1 -1 -1 -1 
## [4,] 8
##
## $keyQpow
## V1 V2 V3 V4 V5 V6 V7 V8 V9 V10
## [1,] -1 -1 1-1 -1 -1 -1 -1 -1 -1 
## [2,] -1 -1 1
## [3,] -1 -1 -1 -1 -1 -1 -1 -1 -1 -1 -1
## [4,] -1 -1 -1 -1 -1 -1 -1 -1 -1 -1 -1 
##
## $keyVarF
## V1 V2 V3 V4 V5 V6 V7 V8 V9 V10
## [1,] 0
## [2,] -1 1 -1 1
## [3,] -1 -1 1
## [4,] -1 -1 1
##
## $keyVarLogN
## [1] 0
##
## $keyVarObs
## V1 V2 V3 V4 V5 V6 V7 V8 V9 V10
## [1,] 0
## [2,] 4, 5
## [3,] 9 9 9 10 10 -1 -1 -1 -1 -1 -1
## [4,] 111 11 111 11 111 11 (1)
##
## $obsCorStruct
## [1] AR AR AR AR
## Levels: ID AR US
##
## $keyCorObs
## V1 V2 V3 V4 V5 V6 V7 V8 V9
## [1,] 0
## [2,] 2
## [3,] 4 4 4 5 5 -1 -1 -1 -1 -1 -1
## [4,] 6 6 6 6 6 6 6 < -1 -1 -1 -1
##
## $stockRecruitmentModelCode
## [1] 0
##
## $noScaledYears
## [1] 0
##
## $keyScaledYears
## numeric(0)
##
## $keyParScaledYA
## <0 x O matrix>
##
## $fbarRange
## [1] 3 7
##
```

```
## $keyBiomassTreat
## [1] -1 -1 -1 -1
##
## $obsLikelihoodFlag
## [1] LN ALN LN LN
## Levels: LN ALN
##
## $fixVarToWeight
## [1] 0
##
## $fracMixF
## [1] 0
##
## $fracMixN
## [1] 0
##
## $fracMixObs
## [1] 0 0 0 0
##
## $constRecBreaks
## numeric(0)
##
## $predVarObsLink
## V1 V2 V3 V4 V5 V6 V7 V8 V9 V10
## [1,] -1 -1 -1 -1 -1 -1 -1 -1 -1 -1
## [2,] -1 -1 -1 -1 -1 -1 -1 -1 NA NA
## [3,] -1 -1 -1 -1 NA NA NA NA NA NA
## [4,] -1 -1 -1 -1 -1 -1 NA NA NA NA
##
## $hockeyStickCurve
## [1] 20
##
## $stockWeightModel
## [1] 0
##
## $keyStockWeightMean
## [1] NA NA NA NA NA NA NA NA NA NA
##
## $keyStockWeightObsVar
## [1] NA NA NA NA NA NA NA NA NA NA
##
## $catchWeightModel
## [1] 0
##
## $keyCatchWeightMean
## [1] NA NA NA NA NA NA NA NA NA NA
##
## $keyCatchWeightObsVar
## [1] NA NA NA NA NA NA NA NA NA NA
##
## $matureModel
## [1] 0
##
## $keyMatureMean
```

```
## [1] NA NA NA NA NA NA NA NA NA NA
##
## $mortalityModel
## [1] 0
##
## $keyMortalityMean
## [1] NA NA NA NA NA NA NA NA NA NA
##
## $keyMortalityObsVar
## [1] NA NA NA NA NA NA NA NA NA NA
##
## $keyXtraSd
## [,1] [,2] [,3] [,4]
```


# The 2022 updated RFID tag-recapture data on NEA mackerel Trends in abundance with different filtering 

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## Summary

A full overview and update of the RFID tagging experiments of mackerel 2011-2022, as well as the recaptures and scanned fish 2012-2021 is given. Since the benchmarking process during ICES IBPNEAMac 2019 and decisions therein, the data included in the SAM stock assessment has been filtered to only include mackerel tagged at ages 5-11, release years 2013 and later and recaptures limited to year 1 and 2 after release. The RFID data set used as input to the SAM stock assessment is a complex one with numbers released per age in a release year, and the numbers scanned and recaptured of these year classes annually in all the years after release; i.e not typical abundance indices per age per year as normally included in age-based stock assessments. Hence, the overview does not only focus on the input data themselves and quality assurance of these, but the actual trends they show for both the different year classes and biomass. Special effort in put on demonstrating trends in actual data included in assessment compared with other ways of filtering the data, such as including more age groups and more years with recaptures after release then the current assessment. Finally, the year class trends, mortality trends in the RFID data are compared with the other age-based input data from commercial catches and the international trawl survey in the Norwegian Sea (IESSNS).

## Background

The Institute of Marine Research in Bergen (IMR) has conducted tagging experiments on mackerel on annual basis since 1968, both in the North Sea and to the west of Ireland during the spawning season May-June. Information from steel-tagged mackerel tagged west of Ireland and British Isles was introduced in the mackerel assessment during ICES WKPELA 2014 (ICES, 2014), and data from release years 1980-2004, and recapture years 1986-2006 has been used in the update assessments after this. The steel tag experiments continued to 2009, with recaptures to 2010, but this part of the data was at the time considered less representative and was excluded.

What is used in the SAM stock assessment is a table of data showing numbers of steel tagged fish per year class in each release year, and the corresponding numbers scanned and recaptured of the same year classes in all years after release. The steel tag data and the corresponding trends in the data in terms of index of total biomass and year class abundance by year is described in (Tenningen et al., 2011).

The steel tag methodology involved a whole lot of manual processes, demanding a lot of effort and reducing the possibility to scan larger proportions of the landings. The tags were recovered at metal detector/deflector gate systems installed at plants processing mackerel for human consumption. External personnel were hired to monitor the systems during processing. Among the typical 50 fish deflected, they had to find the tagged fish with a hand-hold detector and send the fish to IMR for further analysis. It was decided in the end to go for a change in methodology to radio-frequency identification (RFID), which would allow for more automatic processes and increased proportion of scanned landings.

## RFID tag recapture methodology and data quality assurance

The RFID tagging project on NEA mackerel was initiated in 2011 by IMR, and the data were used in update assessments after the ICES WKWIDE2017 benchmark meeting (ICES, 2017b). The data format was the same as for steel tags, but the time series were treated with a different scaling parameter in the assessment.

RFID is a technology that uses radio waves to transfer data from an electronic tag, called an RFID tag, through a reader for the purpose of identifying and tracking the object. The tags used for mackerel are passive, commonly called PIT-tags, specifically developed for tagging fish and animals. They are made of biocompatible glass (specific type used for mackerel is ISO FDX-B $134,3 \mathrm{kHz}, 3.85 \times 23 \mathrm{~mm}$ glass tags) which are equipped with a one-time programmable microchip with a unique ID. Information to the reader is released as it passes an electric field in the antenna system, and information is automatically updated in an IMR database over internet. When tagging and releasing the fish, information is also synced to the IMR database regularly over internet.

There is a web-based software solution (SmartSeaFish) and database that is used to track the different scanning systems at the factories, import data on catch information, and biological sampling data of released fish and screened catches. Based on this information the software is used to allocate the biological data to releases and catches, and to further estimate numbers released every year, and the concurrent numbers screened and recaptured over the next years (by year class).

The development of the tagging data time series is dependent on the work from each country's research institutes, fisheries authorities or the industry it selves to provide additional data about catches screened through the RFID systems, such as total catch weight, position of catch (ICES rectangle), mean weight in catch, etc. Regular biological sampling of the catches landed at these factories is also needed. Altogether, these data are essential for the estimation of numbers screened per year class. Responsible scientists in Norway, Iceland, Faroes and Scotland have been following up the factories, and delivering the catch data and biological data. Currently the responsibilities are as below:

Iceland: Anna Olavsdottir (HAFRO) responsible scientist

- uploading catch data and biological data to SmartSeaFish database
- allocating recaptures and biological samples to the different landings
- testing the 3 Icelandic factories for efficiency, 10 test tags in 10 different landings every year.
- initiates servicing of RFID-antenna systems if needed
- 

Scotland: Steve Mackingson (Scottish Pelagic Fishermen's Association) responsible scientist

- uploading catch data to SmartSeaFish database (we still use Norwegian biological data from same period/ICES area)
- allocating recaptures to the different landings
- testing the 5 Scottish factories for efficiency, 10 test tags in 10 different landings every year/season.
- initiates servicing of RFID-antenna systems if needed
- 

Norway: Aril Slotte (IMR) responsible scientist for the Norwegian RFID tagging program for mackerel and herring, main responsible for final estimations needed to procuce the data table delivered to ICES WGWIDE

- uploading catch data and biological data to SmartSeaFish database
- allocating recaptures and biological samples to the different landings (including biological data to Scottish landings)
- Norway now has 15 factories with RFID antenna systems for scanning mackerel and herring. All factories are serviced 1 time per year and when there are apparent issues to be solved

A new monitoring system has been developed (Figure 1), which is now placed at all 15 Norwegian factories and the 3 Icelandic factories. This monitoring system is continuously overviewing that RFID antennas and readers are functioning. Voltage variations are measured and every 15 min the reading capabilities are tested automatically with a status tag, and these tests are also stored in the SmartFish database for further analyses of efficiency. This monitoring system has replaced the manual testing with 10 test tags in 10 different landings every year/season. The plan is that same systems are placed out at the Scottish factories, or any new factories installing the system.

Based on results from the online monitoring system in addition to the manual test off recapture efficiencies or the online monitoring, responsible scientists decides if data from a factory have to be excluded from final estimation and data input to ICES WGWIDE assessment. Factories that do not function properly are put in an 'out of order' list (Figure 2), where catch data and recapture data from these 'out of order' periods are excluded during estimation.

To conclude regarding quality assurance, we have made progress and the current monitoring of efficiencies at factories that has been raised as a main issue is now at an acceptable level. Still, there is need for more quality control of both all raw tag-recapture data, biological data and allocations of these to landings, as well as the final estimations of data included in the ICES WGWIDE stock assessment. In the future we potentially need to develop annual sworkshops prior to the assessment, where more scientists go through the new data being updated from new tagging experiments, as well as recaptures from all previous experiments, undertake quality assurance of the data and other analyses of the trends in the data outside of the assessment model. The idea is that this should work similarly as post-cruise meetings where all involved scientists take part in final report.

## Status of updated RFID tag recapture data

The RFID tagging technology is clearly a more cost-effective than the old steel tag technology. We are now scanning about 10 times more biomass than during the period with steel tags. An overview of the RFID tagging data in terms of numbers tagged, biomass scanned, and numbers recaptured is given in Tables 1-3, and geographical distributions of data in Figures 3-6.

During the period 2011 - 2022 as many as 556953 mackerel have been tagged with RFID (Table 1). This includes an experiment off the Norwegian Coast on young mackerel in September 2011 as well as five experiments carried out in August in Iceland 2015-2019, none of which are included as input data in the assessment. Data from the releases at the spawning grounds in May-June of Ireland and the Hebrides are the only data included in the assessment.

By 26. August 2022 as many as 10124 RFID-tagged mackerel have been recaptured from all experiments. Looking only recaptures 2012-2021 full years and the experiments of Ireland and British Isles used in update asssessments, 8488 mackerel has been recaptured at landings scanned at 25 European factories processing mackerel for human consumption (Tables 2-3). The project started with RFID antenna reader systems connected to conveyor belt systems at 8 Norwegian factories in 2012. Now there are 5 operational systems at 4 factories in UK (Denholm has 2 RFID systems) and 3 in Iceland. Norway has installed RFID systems at 8 more factories in 2017-2018, most of which with the purpose of scanning Norwegian spring spawning herring catches (IMR started tagging herring in 2016), but some also processing mackerel. Recently one factory, Pelagia Austevoll is terminated, so currently 15 factories are scanning for RFID tags in Norway. More systems are also bought by Ireland (3), which up to now has been non-operational.

During ICES WGWIDE 2018 (ICES, 2018d) meeting bias issues were described for RFID tag data, in addition to potential weighting issues of the tag data inside the model. After the intermediate benchmark meeting ICES IBPNEAMac 2019 (ICES, 2019a), these issues were overcome by using a subset of data for release years (exclude 2011-2012), recapture years (only use recaptures from year 1 and 2 after release) and age groups (exclude youngest fish ages 2-4, use ages 5-11). This is now the subset of data to be used in update assessments.

The exclusion of release years 2011-2012, and recapture years 2012-2013 is mainly based in lack of distributional coverage of scanned fishery, which changed significantly when more countries joined the program and scanned landings from 2014 onwards (Figures 4-5).

The exclusion of recaptures in year 3 or longer after the release year was because data indicated tag loss over time, and that the large majority was recaptured prior to year 3 after release. In year recaptures are not used. However, following recaptures from in year (years out=0) and further through year 1-3+ after tagging, it is apparent that tagged fish are quite quickly distributed in the fishery, and the distributional patterns of recaptures are maintained over time (Figure 6). Hence, potentially more recapture years could be included it one overcame how to adjust for potential tag loss.

The exclusion of ages 1-4, was mainly based in noisy data from these age groups, and the fact that in the early tagging years fish in these age groups were relatively few compared with the scanned fish year 1 and 2 after release. The few fish from these ages were not considered representative for the behaviour of the year classes. However, over time this picture has changed considerable. The age structure of tagged and scanned fish year 1-2 after release are now overlapping, and high proportions of tagged mackerel are now at ages 2-4 (Figure 7). This means that given current filtering we will exclude large proportions of the RFID tag recapture data in coming years, so this is a decision that will have to be revised. Hence, in the following focus is on the actual trends and consistency in the RFID tag data, having in mind that the current filtering may have to be revised in near future.

## Status of RFID tag recapture data trends and consistency for use in stock assessment

Estimates of year class abundance for unfiltered RFID tag-recapture data show trends over time that seems informative for stock assessment (Figure 8), and this is also supported by the tests of consistency in the data (Figure 9), implying a potential for including younger age groups in future assessments.

However, the information coming from the RFID tag data is easier to interpret when comparing age aggregated biomass indices estimated from the RFID data (based on year 1-2 with scanning and recaptures) with SSB from the stock assessment, as shown in Figure 10. The decision to exclude release years 2011-2012 is supported by this plot, showing noisy estimates above the confidence intervals of the assessment. However, by including only release years 2013 onwards as in current assessments, the biomass trend in the RFID tag data is more in line with the SSB of the assessment, especially the decrease in SSB from 2017-2020 is also very evident regardless of ages aggregated from RFID data. This again signifies that over time, and in a future benchmark process, information of tag recaptures from younger age groups may be included again if trends are informative for the assessment.

In recent years we have seen a trend that the information from RFID tag recapture data about abundance in a release year increase when adding one more year with recaptures and scanned data. Figures 11-12 illustrates this issue for single year classes as well as various age aggregated abundance estimates. This supports the decision to stick to only using recapture and scanned data for year 1 and 2 after release. Moreover, it also implies the last year included in the stock assessment always based on s will be revised in next update assessment, with a recent clear tendency that adding the second year with data lifts the perception of abundance in a release year.

One more way of looking at the information from RFID tag recapture data relative to the other sources of input data and the stock assessment itself, is to compare signals of total mortality rate ( $Z$ ) by estimating slope of decrease in abundance of year classes 2003-2014 of fully mature fish aged 4-12
(Figure 13). Here it is apparent that mortality signals from RFID data seem informative following a steady decrease as the catch data, whereas IESSNS data sticks out as a bit noisier trend. When looking at the estimated Z for each data source, it is evident that the RFID data show signals of higher mortality rate than the catch data, whereas Z estimates for the IESSNS data are even lower. Z estimates from the WGWIDE2022 assessment are also above the catch data, but below the tag estimates, signifying that the model put some weight on the tag data. Note that RFID data shows more uncertain estimates of $Z$ for recent year classes with very few years, fewer than the other sources, which means the estimates may change over time.

The overall conclusion is still that the RFID data seems quite informative, and that the current filtering and exclusion of data for use in stock assessment should be revised in near future. Only looking at the relative year class structure in the tag estimates 2019-2020 compared with the structure seen in catch data, IESSNS, and WGWIDE2022 assessment (Figure 14), we see very similar structure. In addition, here it is evident that the RFID-estimates also how large new year classes such as 2016, which is not used yet in assessment because of the exclusion of young fish. Also noticeable is that yearclasses such as 20122013 are relatively smaller in the RFID-estimates than in the IESSNS and catch data, which likely may be due to age reading issues. The tag estimates are based on fewer readers.

Finally, on a totally different issue. Do mackerel growing up in the North Sea belong to a specific component? Figure 15 demonstrates that recaptures from very young fish tagged in the North Sea at the western Norwegian coast (Bømlo Island) over the year adapted the same migration pattern as the fish tagged at older ages along Ireland-Hebrides. This supports the hypothesis that mackerel growing up in the North Sea do not belong to a North Sea component, but to a large dynamic mackerel population changing migration pattern and spawning areas as the stock fluctuates in abundance and age structure.

Link to official publication of all raw data needed to produce input data set to the assessment is: Aril Slotte (IMR), Anna Ólafsdóttir (MFRI), Sigurður Pór Jónsson (MFRI), Jan Arge Jacobsen (FAMRI) and Steve Mackinson (SPFA) (2021) PIT-tag time series for studying migrations and use in stock assessment of North East Atlantic mackerel (Scomber Scombrus) http://metadata.nmdc.no/metadataapi/landingpage/f9e8b1cff4261cf6575e70e56c4c3b3e This is the correct citation when using the data. The data are available through this link as various APIs that are updated daily. There is also an R-package https://github.com/IMRpelagic/taggart can be used to download data from the APIs.

## Tables

Table 1. Overview of numbers released in the different RFID tagging experiments, and numbers recaptured per year. Recaptures from experiments and recapture years used in 2022 stock assessment, based on decisions in the ICES IBPNEAMac 2019 (ICES 2019) are outlined and marked grey. However, note that these numbers also include recaptures from some factories excluded in the final estimation of tag table used in the stock assessment 2022 (see Tables 2-3), due to low efficiency or misfunctions. Note that recaptures in 2022 are preliminary by 26. August.

| Survey | N -Released | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | All years |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Iceland 2015 | 806 | 0 | 0 | 0 | 6 | 2 | 3 | 0 | 0 | 0 | 0 | 0 | 11 |
| Iceland 2016 | 4884 | 0 | 0 | 0 | 0 | 59 | 48 | 28 | 19 | 13 | 10 | 0 | 177 |
| Iceland 2017 | 3890 | 0 | 0 | 0 | 0 | 0 | 0 | 28 | 27 | 9 | 13 | 4 | 5 |
| Iceland 2018 | 1872 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 16 | 13 | 8 | 3 | 86 |
| Iceland 2019 | 3614 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 25 | 12 | 3 | 45 |
| Norway2011 | 31253 | 9 | 31 | 24 | 32 | 26 | 16 | 20 | 7 | 13 | 6 | 2 | 185 |
| Ireland-Hebrides 2011 | 18645 | 27 | 24 | 29 | 24 | 17 | 5 | 9 | 7 | 3 | 2 | 0 | 186 |
| Ireland-Hebrides 2012 | 32135 | 31 | 57 | 60 | 64 | 34 | 21 | 12 | 5 | 6 | 5 | 4 | 299 |
| Ireland-Hebrides 2013 | 22792 | 0 | 26 | 89 | 104 | 61 | 30 | 21 | 10 | 8 | 5 | 1 | 355 |
| Ireland-Hebrides 2014 | 55184 | 0 | 0 | 112 | 311 | 277 | 139 | 91 | 44 | 45 | 29 | 3 | 1051 |
| Ireland-Hebrides 2015 | 43905 | 0 | 0 | 0 | 115 | 217 | 177 | 93 | 49 | 41 | 20 | 9 | 721 |
| Ireland-Hebrides 2016 | 43956 | 0 | 0 | 0 | 0 | 124 | 324 | 183 | 121 | 92 | 48 | 12 | 904 |
| Ireland-Hebrides 2017 | 56073 | 0 | 0 | 0 | 0 | 0 | 134 | 344 | 174 | 146 | 80 | 21 | 899 |
| Ireland-Hebrides 2018 | 38136 | 0 | 0 | 0 | 0 | 0 | 0 | 204 | 248 | 229 | 132 | 37 | 850 |
| Ireland-Hebrides 2019 | 51179 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 290 | 541 | 435 | 123 | 1389 |
| Ireland-Hebrides 2020 | 48968 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 517 | 811 | 207 | 1535 |
| Ireland-Hebrides 2021 | 49173 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 755 | 269 | 1024 |
| Ireland-Hebrides 2022 | 50488 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 400 | 400 |
| All surveys | 556953 | 67 | 138 | 314 | 656 | 817 | 925 | 1037 | 1004 | 1705 | 2362 | 1099 | 10124 |
| All Ireland-Hebrides | 510634 | 58 | 107 | 290 | 618 | 730 | 830 | 957 | 948 | 1628 | 2322 | 1086 | 9574 |

Table 2. Overview of numbers of tonnes scanned for RFID tags per factory per year. Data from years used in 2022 stock assessment (2014 and onwards), based on decisions in the ICES IBPNEAMac 2019 (ICES 2019), are outlined and marked grey. Based on an evaluation of efficiency of the scanners, data from some factories are excluded as they were not functioning or having poor data quality, and these are not marked grey.

| Factory | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | All years |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FO01 Vardin Pelagic | 0 | 0 | 10460 | 11565 | 7895 | 4844 | 0 | 0 |  |  | 34763 |
| GB01 Denholm Coldstore | 0 | 0 | 0 | 4377 | 4710 | 5365 | 7806 | 5191 | 8809 | 8957 | 45215 |
| GB01 Denholm Factory | 0 | 0 | 14939 | 17509 | 18840 | 17913 | 13609 | 12018 | 13951 | 6284 | 115064 |
| GB02 Lunar Freezing Peterhead | 0 | 0 | 22586 | 17830 | 16473 | 9745 | 9857 | 14300 | 24382 | 24751 | 139924 |
| GB03 Lunar Freezing Fraserburgh | 0 | 0 | 0 | 8797 | 14282 | 12684 | 9452 | 5729 |  |  | 50943 |
| GB04 Pelagia Shetland | 0 | 0 | 21436 | 41117 | 40200 | 26935 | 25350 | 15128 | 22573 | 18312 | 211051 |
| GB05 Northbay Pelagic | 0 | 0 | 0 | 0 | 0 | 0 | 15353 | 12667 | 15478 | 19377 | 62875 |
| IC01 Vopnafjord | 0 | 0 | 18577 | 18772 | 21716 | 22935 | 18869 | 18547 | 21191 | 15729 | 156336 |
| IC02 Neskaupstad | 0 | 0 | 0 | 6288 | 21887 | 19558 | 16757 | 26633 | 28180 | 32216 | 151519 |
| IC03 Höfn | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10592 | 13488 | 10087 | 34167 |
| NO01 Pelagia Egersund Seafood | 20930 | 21442 | 36724 | 14375 | 15905 | 0 | 48373 | 25404 | 51013 | 37196 | 271361 |
| NO02 Skude Fryseri | 7546 | 8250 | 16719 | 14172 | 8671 | 16760 | 3108 | 1285 | 17661 | 18611 | 112783 |
| NO03 Pelagia Austevoll | 6405 | 6134 | 10314 | 4203 | 2216 | 0 | 7293 | 3533 | 8351 |  | 48449 |
| NO04 Pelagia Florø | 9986 | 12838 | 17379 | 12592 | 7749 | 0 | 0 | 0 |  |  | 60544 |
| NO05 Pelagia Måløy | 13344 | 14632 | 13942 | 21051 | 15762 | 22405 | 13341 | 8591 | 21287 | 22724 | 167079 |
| NO06 Pelagia Selje | 17731 | 26878 | 39525 | 41209 | 29897 | 35416 | 28972 | 32047 | 31678 | 34835 | 318189 |
| N007 Pelagia Liavågen | 9442 | 10968 | 22395 | 18144 | 13911 | 19989 | 12398 | 11888 | 17487 | 21515 | 158138 |
| N008 Brødrene Sperre | 14425 | 15048 | 20182 | 34307 | 36736 | 18814 | 34280 | 8515 | 32333 | 28283 | 242924 |
| NO09 Lofoten Viking | 0 | 0 | 0 | 0 | 0 | 0 | 3380 | 2457 | 3823 | 17924 | 27584 |
| NO10 Pelagia Træna | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 10509 | 10509 |
| NO11 Nergård Sild | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2524 | 2527 |
| NO12 Pelagia Lødingen | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 950 | 4883 | 5833 |
| NO13 Pelagia Tromsø | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 180 | 180 |
| NO14 Nils Sperre | 0 | 0 | 0 | 0 | 0 | 0 | 28304 | 26272 | 30265 | 33901 | 118742 |
| NO15 Grøntvedt Pelagic | 0 | 0 | 0 | 0 | 0 | 0 | 6411 | 0 | 0 | 6778 | 13190 |
| NO16 Vikomar | 0 | 0 | 0 | 0 | 0 | 0 | 12512 | 6480 | 15679 | 16915 | 51585 |
| All factories | 99808 | 116190 | 265178 | 286310 | 276850 | 233363 | 315426 | 247277 | 378582 | 392491 | 2611475 |

Table 3. Overview of numbers of RFID tagged mackerel recaptured per factory per year. Only recaptures from Ireland surveys (Table 1) that are used as basis stock assessment are shown. Recaptures from years used in 2022 stock assessment from 2014 and onwards, based on decisions in the ICES IBPNEAMac 2019 (ICES 2019), are outlined and marked grey. Based on an evaluation of efficiency of the scanners, data from some factories are excluded as they were not functioning or having poor data quality, and these are not marked grey. See Table 2 for biomass scanned.

| Factory | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | All years |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FO01 Vardin Pelagic | 0 | 0 | 13 | 35 | 20 | 11 | 0 | 0 | 0 | 0 | 79 |
| GB01 Denholm Coldstore | 0 | 0 | 0 | 10 | 10 | 24 | 36 | 19 | 46 | 61 | 206 |
| GB01 Denholm Factory | 0 | 0 | 25 | 62 | 77 | 113 | 54 | 53 | 92 | 64 | 540 |
| GB02 Lunar Freezing Peterhead | 0 | 0 | 32 | 49 | 60 | 38 | 41 | 54 | 123 | 137 | 534 |
| GB03 Lunar Freezing Fraserburgh | 0 | 0 | 0 | 9 | 14 | 7 | 25 | 34 | 0 | 0 | 89 |
| GB04 Pelagia Shetland | 0 | 0 | 21 | 124 | 148 | 137 | 98 | 82 | 134 | 134 | 878 |
| GB05 Northbay Pelagic | 0 | 0 | 0 | 0 | 0 | 0 | 57 | 59 | 81 | 136 | 333 |
| IC01 Vopnafjord | 0 | 0 | 22 | 55 | 65 | 59 | 62 | 54 | 146 | 180 | 643 |
| ICO2 Neskaupstad | 0 | 0 | 0 | 19 | 65 | 54 | 35 | 114 | 127 | 284 | 698 |
| IC03 Höfn | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 44 | 65 | 117 | 226 |
| N001 Pelagia Egersund Seafood | 10 | 22 | 18 | 7 | 1 | 0 | 137 | 80 | 184 | 184 | 643 |
| NO02 Skude Fryseri | 5 | 6 | 21 | 17 | 25 | 51 | 13 | 3 | 34 | 88 | 263 |
| NO03 Pelagia Austevoll | 1 | 1 | 7 | 4 | 0 | 0 | 28 | 17 | 48 | 0 | 106 |
| NO04 Pelagia Florø | 5 | 12 | 27 | 21 | 16 | 0 | 0 | 0 | 0 | 0 | 81 |
| NO05 Pelagia Måløy | 5 | 13 | 18 | 43 | 37 | 77 | 36 | 28 | 97 | 121 | 475 |
| NO06 Pelagia Selje | 15 | 27 | 37 | 76 | 59 | 85 | 87 | 153 | 172 | 257 | 968 |
| N007 Pelagia Liavågen | 10 | 11 | 29 | 31 | 26 | 97 | 48 | 51 | 111 | 138 | 552 |
| NO08 Brødrene Sperre | 7 | 15 | 20 | 56 | 107 | 77 | 52 | 12 | 0 | 99 | 445 |
| NO09 Lofoten Viking | 0 | 0 | 0 | 0 | 0 | 0 | 10 | 3 | 5 | 66 | 84 |
| NO10 Pelagia Træna | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 67 | 67 |
| NO11 Nergård Sild Senjahopen | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10 | 10 |
| NO12 Pelagia Lødingen | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 16 | 17 |
| NO14 Nils Sperre | 0 | 0 | 0 | 0 | 0 | 0 | 109 | 68 | 73 | 80 | 330 |
| NO15 Grøntvedt Pelagic | 0 | 0 | 0 | 0 | 0 | 0 | 11 | 0 | 0 | 18 | 29 |
| NO16 Vikomar | 0 | 0 | 0 | 0 | 0 | 0 | 18 | 20 | 89 | 65 | 192 |
| All factories | 58 | 107 | 290 | 618 | 730 | 830 | 957 | 948 | 1628 | 2322 | 8488 |

Figures


Figure 1. Example of how the new monitoring systems looks like. It follows the traffic light systems, where red implies that we currently may have issues with either voltage variations or reduced efficiency of RFID tags.


Figure 2. Example of how it looks like in the SmartSeaFish web-based software where factories having issues with recapture efficiency are put in an 'Out of order' list. Catch data and recapture data from these factories and periods are excluded in final estimation of data table being included in the ICES WGWIDE stock assessment.


Figure 3. Distribution of RFID tagged mackerel from experiments west of Ireland-Hebrides during 2011-2022. Number of released fish is summed per ICES rectangle. See Table 1 for details on numbers released. Note that data from releases 2011-2012 are not used in the stock assessment, based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019), and data from experiments in 2021-2022 are not included as there are no full years with recaptures yet.


Figure 4. Distribution (summed per ICES rectangle) of catches scanned for RFID tagged mackerel during 2012-2021. Note that data on scanned catches in 2012-2013 are not used in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019). Detailed data on scanned biomass per factory and year are given in Table


Figure 5. Distribution (summed per ICES rectangle) of recaptures of RFID tagged mackerel during 2012-2021. Note that data on recaptures in 2012-2013 are not used in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019). Detailed data on recaptures per factory and year are given in Table 3.


Figure 6. Distribution (summed per ICES rectangle) of recaptures of RFID tagged mackerel related to release years 2011-2015 and years after release ( $0=$ same year as tagging, $1=$ year after tagging etc.). Note that data on recaptures from 2011-2012 release years and from year 0 and 3+ after tagging are not used in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019). Note also tha $t$ in 2011 scanning had not started (Figure 4), so no in year recaptures.


Figure 6 continued for release years 2016-2021. Preliminary recaptures in 2022 are not included as allocations to catches are not completed.


Figure 7. Overview of the relative year class distribution among RFID tagged mackerel per release year from experiments west of Ireland-Hebrides in May-June, compared with the number scanned and recaptured in year 1 and 2 after release of the same year classes. See Figure 3 for distribution of the tagged fish and the respective distribution of recaptures in year 1 and 2 after release in Figures 4-5. Note that data from releases in 2011-2012 are not used in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019). Note also that it was decided to only use ages 5-11 in updated assessments, and limits for this age span is marked (vertical grey dotted lines) for each release year. Details on actual numbers released and recaptured are given in Table 1 and 3, also for other tagging experiments not included in the stock assessment.


Figure 8. Trends in year class abundance ( $\mathrm{N}=$ numbers released/numbers recaptured*numbers scanned year 1 and 2 after release) from RFID tag-recapture data based on aggregated data on recaptures and scanned numbers in year 1 and 2 after each release year. Data excluded in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019), release years 2011-2012 and ages 2-4 and 12+, are marked with dotted lines in year class trends. Note that dotted grey lines are showing a total mortality $\mathrm{Z}=0.4$ for comparison with year class trends.


Figure 9. Internal consistency of the of mackerel RFID abundance index from release years 2011 to 2020, based on indices from Figure 8. Ages indicated by white numbers in grey diagonal cells. Statistically significant positive correlations ( $p<0.05$ ) are indicated by regression lines and red cells in upper left half. Correlation coefficients ( r ) are given in the lower right half.


Figure 10. Trends in various age aggregated biomass indices from RFID tag-recapture data compared with the SSB ( $\pm 95$ confidence intervals) from the WGWIDE 2022 stock assessment. Data are based on a combination of estimated numbers by year class from Figure 8 scaled by survival parameter estimated by SAM in WGWIDE 2022 (0.1599) and weight at age in stock form same assessment. Vertical dotted line marks the starting year where RFID tagging experiments are used in the stock assessment based on decisions in the ICES IBPNEAMac 2019. meeting (ICES 2019), and the trend of ages $5-11$ is representing the subset of ages used in updated assessments. Note that final year with data 2020 is only based on recapture year 1 after release, whereas the other years are based on recapture year 1-2 after release, i.e. completed. In recent years (2016-2018) the estimates have tended to increase when adding the second recapture year (See Figures 11-12).


Figure 11. Trends in year class abundance ( $\mathrm{N}=$ numbers released/numbers recaptured*numbers scanned) from RFID tag-recapture data based on different filtering of recapture year included. Upper panels show the difference between basing the estimate on either year 1, 2, 3, or 4 after release, whereas bottom panels show the difference between using year 1 after release versus various intervals of years after release. Note that data are shown for all ages (1-max 16) with data.


Figure 12. Trends in various age aggregated biomass indices from RFID tag-recapture data based on different filtering of recapture year included. Upper panels show the difference between basing the estimate on either year $1,2,3$, or 4 after release, whereas bottom panels show the difference between using year 1 after release versus various intervals of years after release.


Figure 13. Signals of total mortality rate in input data to the mackerel stock assessment. (A)Upper panels show the trends in year class abundance and estimated slope of decrease from the age 4 when it is fully recruited to the spawning stock until age 12 (interpreted as signal of total mortality), of various sources of unscaled input data to the mackerel stock assessment (RFID, IESSNS and catch data) compared with the final trend estimated in the stock assessment (WGWIDE 2022). (B) Bottom panels summarize the year class differences in estimated total mortality rate (with $95 \%$ confidence intervals), and differences between the various data sources.


Figure 14. Comparison of relative year class contributions between RFID-tag estimates, catch data, IESSNS data and the WGWIDE2022 stock assessment it self.


Figure 15. Distribution (summed per ICES rectangle) of recaptures 2012-2022 from an RFID tagging experiment on mackerel in the North Sea at the Norwegian West coast (blue dot) in 2011. This was mainly young mackerel tagged, where $88 \%$ were 1 year olds and $6.5 \% 2$ year olds, using the North Sea/Norwegian coast as nursery.


[^0]:    ICES
    INTERNATIONAL COUNCIL FOR THE EXPLORATION OF THE SEA
    CIEM COUNSEIL INTERNATIONAL POUR L'EXPLORATION DE LA MER

[^1]:    * Official catches by area from Sweden are not included (2012); ~
    ** Official catches by area from Sweden and Greenland are not included (2013);
    *** Grand total includes only 1336 tonnes from UK(England + Wales) (2016 total catch from UK(England + Wales) = $\mathbf{1 3 7 4}$ ton).

[^2]:    * the values of catches inside/outside NEAFC RA have been estimated based on the ICES Preliminary Catch Statistics.
    ** without the Russian preliminary catch data and the unallocated catch data.

[^3]:    ^ Russia 2021 preliminary data (Q1+Q2) submitted to WGWIDE 2021.

[^4]:    ^ Russia 2021 preliminary data (Q1+Q2) submitted to WGWIDE 2021.

[^5]:    *Survey discarded. **No survey

[^6]:    $\overline{1}$ - Southern Horse Mackerel (ICES Division 9) is assessed by ICES WGHANSA since 2011

[^7]:    * Percentage based on ICES estimate with regards to age samples.

[^8]:    year
    age $20162017 \quad 2018 \quad 2019 \quad 2020 \quad 2021$
    $\begin{array}{lllllll}0 & 0.035 & 0.018 & 0.066 & 0.057 & 0.057 & 0.049\end{array}$
    $\begin{array}{llllllll}1 & 0.154 & 0.178 & 0.147 & 0.112 & 0.174 & 0.163\end{array}$
    $\begin{array}{lllllll}2 & 0.240 & 0.266 & 0.247 & 0.260 & 0.285 & 0.277\end{array}$
    $\begin{array}{llllllll}3 & 0.297 & 0.311 & 0.320 & 0.297 & 0.322 & 0.338\end{array}$
    $4 \quad 0.329 \quad 0.3560 .3550 .360 \quad 0.360 \quad 0.374$
    $\begin{array}{llllllll}5 & 0.356 & 0.377 & 0.397 & 0.388 & 0.389 & 0.406\end{array}$
    $\begin{array}{llllllll}6 & 0.383 & 0.397 & 0.410 & 0.429 & 0.417 & 0.441\end{array}$
    $\begin{array}{lllllll}7 & 0.411 & 0.415 & 0.426 & 0.441 & 0.444 & 0.457\end{array}$
    $8 \quad 0.438 \quad 0.444 \quad 0.446 \quad 0.4530 .459 \quad 0.477$
    $9 \quad 0.4530 .4650 .469 \quad 0.4720 .4710 .486$
    $10 \quad 0.4790 .4840 .492 \quad 0.497 \quad 0.495 \quad 0.501$
    $\begin{array}{lllllll}11 & 0.499 & 0.497 & 0.507 & 0.514 & 0.519 & 0.514\end{array}$
    $\begin{array}{lllllll}12 & 0.520 & 0.531 & 0.537 & 0.537 & 0.554 & 0.548\end{array}$

[^9]:    ${ }^{1}$ Institute of Marine Research, Bergen, Norway
    ${ }^{2}$ DTU-Aqua, Denmark
    ${ }^{3}$ Thünen-Institute of Sea Fisheries, Germany
    ${ }^{4}$ Marine and Freshwater Research Institute, Hafnarfjordur, Iceland
    ${ }^{5}$ Faroe Marine Research Institute, Tórshavn, Faroe Islands
    ${ }^{6}$ Wageningen Marine Research, Netherlands
    ${ }^{7}$ CEFAS, United Kingdom

[^10]:    *Updated working document from WD11 for WGWIDE2021 to WGWIDE2022.

