# ARCTIC FISHERIES WORKING GROUP (AFWG; outputs from 2022 meeting) 

Version 2: correction to executive summary description of cap.27.1-2

## VOLUME 5 | ISSUE 18

ICES SCIENTIFIC REPORTS

RAPPORTS
SCIENTIFIQUES DU CIEM


[^0]
## International Council for the Exploration of the Sea Conseil International pour l'Exploration de la Mer

H.C. Andersens Boulevard 44-46<br>DK-1553 Copenhagen V<br>Denmark<br>Telephone (+45) 33386700<br>Telefax (+45) 33934215<br>www.ices.dk<br>info@ices.dk

ISSN number: 2618-1371

This document has been produced under the auspices of an ICES Expert Group or Committee. The contents therein do not necessarily represent the view of the Council.
© 2023 International Council for the Exploration of the Sea

This work is licensed under the Creative Commons Attribution 4.0 International License (CC BY 4.0). For citation of datasets or conditions for use of data to be included in other databases, please refer to ICES data policy.


## ICES Scientific Reports

## Volume 5 | Issue 18

## ARCTIC FISHERIES WORKING GROUP (AFWG)

Recommended format for purpose of citation:

ICES. 2023. Arctic Fisheries Working Group (AFWG; outputs from 2022 meeting). ICES Scientific Reports. 5:18. 507 pp. https://doi.org/10.17895/ices.pub. 20012675

## Editors

Daniel Howell

Authors<br>Jane Aanestad Godiksen • Caroline Aas Tranang • Erik Berg • Matthias Bernreuther • Bjarte Bogstad Olav Nikolai Breivik • José Miguel Casas • Laura Clain • Elise Eidset • Elena Eriksen • Johanna Fall Maria Fossheim • Harald Gjøsæter • Sofie Gundersen • Elvar H. Hallfredsson • Hannes Höffle • Daniel Howell • Edda Johannesen • Kjell Nedreaas • Anders Nielsen • Georg Skaret • Arved Staby • Brian Stock Samuel Subbey • Ross Tallman • John Tyler Trochta • Tone Vollen • Kristin Windsland

## Contents

i Executive summary ..... vi
ii Expert group information ..... viii
1 Introduction and ecosystem considerations ..... 1
1.1 Terms of reference ..... 1
1.2 Additional requests. ..... 1
1.3 Responses to terms of reference ..... 1
1.4 Benchmarks ..... 2
1.5 Total catches ..... 2
1.5.1 Uncertainty in catch data ..... 3
1.5.2 Sampling effort-commercial fishery. ..... 3
1.5.2.1 Cod, haddock, and saithe ..... 4
1.5.2.2 Data issues with S. mentella ..... 4
1.5.2.3 Data issues with $S$. norvegicus ..... 4
1.5.2.4 Data issues with NEA Greenland halibut ..... 4
1.5.3 The percentage of the total catch that has been taken in the NEAFC regulatory areas by year in the last year ..... 5
1.6 Uncertainties in survey data ..... 8
1.7 Age reading ..... 11
1.8 Assessment method issues ..... 13
1.9 Environmental information included in the advice of NEA cod ..... 14
1.10 Proposals for status of assessments in 2022-2023 ..... 14
1.11 Ecosystem information ..... 38
1.11.1 O-group abundance ..... 38
1.11.2 Consumption, natural mortality, and growth ..... 38
1.11.3 Maturation, condition factor, and fisheries-induced evolution. ..... 39
1.11.4 Recruitment prediction for northeast Arctic cod ..... 40
1.11.5 Historic overview ..... 40
1.11.6 Models used in 2021 ..... 42
2 Norwegian coastal cod ..... 55
2.1 Fisheries (both stocks) ..... 56
2.1.1 Revision of catch data ..... 57
2.1.2 Catch sampling ..... 57
2.1.3 Regulations ..... 58
2.2 Northern Norwegian coastal cod ..... 62
2.2.1 Stock status summary ..... 62
2.2.2 The fishery (Table 2.2.1-Table 2.2.4) ..... 63
2.2.3 Survey results ..... 64
2.2.4 Data used in the assessment ..... 67
2.2.5 Final assessment run ..... 68
2.2.6 Reference points ..... 69
2.2.7 Predictions ..... 70
2.2.8 Comments to the assessment and the forecast ..... 70
2.2.9 Tables and figures ..... 71
2.3 Southern Norwegian coastal cod ..... 104
2.3.1 Stock status summary ..... 104
2.3.2 Fisheries (Table 2.3.2-Table 2.3.4) ..... 105
2.3.3 Reference fleet ..... 106
2.3.4 Standardized CPUE index (Table 2.3.6 and Figures 2.3.3-2.3.7) ..... 106
2.3.5 Stochastic LBSPR (Table 2.3.1) ..... 108
2.3.6 Results of the assessment (Figure 2.3.6-Figure 2.3.13) ..... 110
2.3.7 Comments to the assessment ..... 112
2.3.8 Reference points ..... 113
2.3.9 Catch scenarios for 2023 ..... 113
2.3.10 Management considerations ..... 113
2.3.11 Rebuilding plan for coastal cod. ..... 114
2.3.12 Recent ICES advice ..... 114
2.3.13 Figures and tables ..... 115
Northeast Arctic cod ..... 131
3.1 Status of the fisheries ..... 131
3.1.1 Historical development of the fisheries (Table 3.1) ..... 131
3.1.2 Reported catches prior to 2021 (Tables 3.1-3.4, Figure 3.1) ..... 131
3.1.3 Unreported catches of Northeast Arctic cod (Table 3.1) ..... 132
3.1.4 TACs and advised catches for 2021 and 2022 ..... 133
3.2 Status of research ..... 133
3.2.1 Fishing effort and CPUE (Table A1, Figure 3.6a-c) ..... 133
3.2.2 Survey results - abundance and size at age (Tables 3.5, A2-A14) ..... 133
3.2.3 Revision of 2021 survey results ..... 135
3.2.4 Age reading ..... 136
3.3 Data available for use in assessment ..... 136
3.3.1 Catch-at-age (Table 3.6) ..... 136
3.3.2 Survey indices available for use in assessment (Table 3.13, A13) ..... 136
3.3.3 Weight-at-age (Tables 3.7-3.9, A2, A4, A6, A8, A12) ..... 137
3.3.4 Natural mortality including cannibalism (Table 3.12, Table 3.17) ..... 137
3.3.5 Maturity-at-age (Tables 3.10-3.11, Tables 3.10-3.11) ..... 138
4 Northeast Arctic haddock ..... 139
4.1 Introductory note. ..... 139
4.2 Status of the fisheries ..... 139
4.2.1 Historical development of the fisheries ..... 139
4.2.2 Catches prior to 2021 (Table 4.1-Table 4.3, Figure 4.1) ..... 139
4.2.3 Catch advice and TAC for 2021 ..... 140
4.3 Status of research ..... 140
4.3.1 Survey results ..... 140
4.4 Data used in the assessment ..... 141
4.4.1 Catch-at-age (Table 4.4) ..... 141
4.4.2 Catch-weight-at-age (Table 4.5) ..... 141
4.4.3 Stock-weight-at-age (Table 4.6) ..... 141
4.4.4 Maturity-at-age (Table 4.7) ..... 141
4.4.5 Natural mortality (Table 4.8) ..... 141
4.4.6 Data for tuning (Table 4.9) ..... 141
4.4.7 Changes in data from last year (Table 4.6-Table 4.7, Table 4.9) ..... 142
4.5 Assessment models and settings (Table 4.10) ..... 142
4.6 Results of the assessment (Table 4.11-Table 4.14 and Figure 4.1-Figure 4.3) ..... 142
4.7 Comparison with last year's assessment (Figure 4.4) ..... 143
4.8 Additional assessment methods (Table 4.15, Figure 4.5-Figure 4.6) ..... 143
4.8.1 XSA (Figure 4.5) ..... 143
4.8.2 TISVPA (Figure 4.6) ..... 143
4.8.3 Model comparisons (Figure 4.7) ..... 144
4.9 Predictions, reference points and harvest control rules (Table 4.16-Table 4.21) ..... 144
4.9.1 Recruitment (Table 4.16-Table 4.17) ..... 144
4.9.2 Prediction data (Table 4.18, Figure 4.8) ..... 145
4.9.3 Biomass reference points (Figure 4.1) ..... 145
4.9.4 Fishing mortality reference points (Figure 4.1) ..... 145
4.9.5 Harvest control rule ..... 146
4.9.6 Prediction results and catch options for 2021 (Table 4.19-Table 4.21) ..... 146
4.9.7 Comments to the assessment and predictions (Figure 4.2-Figure 4.4 and Figure 4.9) ..... 147
5 Northeast Arctic saithe. ..... 211
5.1 The fishery (Table 5.1 and Table 5.2, Figure 5.1) ..... 211
5.1.1 ICES advice applicable to 2021 and 2022 ..... 211
5.1.2 Management applicable in 2021 and 2022 ..... 211
5.1.3 The fishery in 2021 and expected landings in 2022 ..... 211
5.2 Commercial catch-effort data and research vessel surveys ..... 212
5.2.1 Catch-per-unit-effort ..... 212
5.2.2 Survey results (Figure 5.1-5.2) ..... 212
5.2.3 Recruitment indices ..... 212
5.3 Data used in the assessment ..... 213
5.3.1 Catch numbers-at-age (Table 5.3) ..... 213
5.3.2 Weight-at-age (Table 5.4) ..... 213
5.3.3 Natural mortality ..... 213
5.3.4 Maturity-at-age (Table 5.5) ..... 213
5.3.5 Tuning data (Table 5.6) ..... 213
5.4 SAM runs and settings (Table 5.7) ..... 214
5.5 Final assessment run (Table 5.8 to Table 5.11, Figure 5.3-5.6) ..... 214
5.5.1 SAM F, N, and SSB results (Tables 5.9-5.11, Figures 5.5-5.6) ..... 215
5.5.2 Recruitment (Table 5.10, Figure 5.5) ..... 215
5.6 Reference points (Figure 5.5) ..... 215
5.6.1 Harvest control rule ..... 216
5.7 Predictions ..... 216
5.7.1 Input data (Table 5.12) ..... 216
5.7.2 Catch options for 2022 (short-term predictions; Tables 5.13-14) ..... 216
5.7.3 Comparison of the present and last year's assessment ..... 217
5.8 Comments to the assessment and the forecast (Figure 5.6) ..... 217
5.9 Tables and figures ..... 218
6 Northeast Arctic beaked redfish. ..... 250
6.1 Status of the fisheries ..... 250
6.1.1 Development of the fishery ..... 250
6.1.2 Bycatch in other fisheries ..... 250
6.1.3 Landings prior to 2021 (Tables 6.1-6.7, Figure 6.1) ..... 251
6.1.4 Expected landings in 2022 ..... 251
6.2 Data used in the assessment ..... 252
6.2.1 Length-composition from the fishery (Figure 6.4) ..... 252
6.2.2 Catch-at-age (Tables 6.8-6.11, Figure 6.5) ..... 253
6.2.3 Weight-at-age (Tables 6.12, 6.13, Figures 6.6, 6.7) ..... 253
6.2.4 Maturity-at-age (Table 6.14, Figure 6.8) ..... 254
6.2.5 Natural mortality ..... 254
6.2.6 Scientific surveys ..... 254
6.3 Assessment ..... 255
6.3.1 Results of the assessment (Tables 6.20, 6.21, Figures 6.18-6.24) ..... 255
6.4 Comments to the assessment ..... 257
6.5 Biological reference points ..... 258
6.6 Management advice ..... 258
6.7 Possible future development of the assessment ..... 258
6.8 Tables and figures ..... 260
7 Northeast Arctic golden redfish ..... 329
7.1 Status of the fisheries ..... 329
7.1.1 Recent regulations of the fishery ..... 329
7.1.2 Landings prior to 2022 (Tables 7.1-7.4 and Figures 7.1-7.3) ..... 329
7.1.3 Expected landings in 2022 ..... 330
7.2 Data used in the assessment (Table 0.1 and Figure E1) ..... 330
7.2.1 Catch-at-length and age (Table 7.5 and Figure 7.4) ..... 330
7.2.2 Catch weight-at-age (Table 7.6) ..... 331
7.2.3 Maturity-at-age (Table E1, Figure 7.5a-b) ..... 331
7.2.4 Survey results (Tables E2a,b-E3a,b-E4, Figures 7.6a,b-7.8) ..... 331
7.3 Assessment with the Gadget model ..... 332
7.3.1 Description of the model ..... 332
7.3.2 Data used for tuning ..... 333
7.3.3 Assessment results using the Gadget model (Figures 7.9-7.13) ..... 333
7.3.4 State of the stock ..... 334
7.3.5 Biological reference points ..... 335
7.3.6 Management advice ..... 335
7.3.7 Implementing the ICES Fmsy framework ..... 335
7.4 Tables and figures ..... 337
7.5 Additional tables and figures ..... 368
382
8.1 Status of the fisheries ..... 382
8.1.1 Landings prior to 2022 (Tables 8.1-8.8, Figures 8.1-8.3) ..... 382
8.1.2 ICES advice applicable to 2021-2023 ..... 382
8.1.3 Management ..... 383
8.1.4 Expected landings in 2022 ..... 384
8.2 Status of research ..... 384
8.2.1 Survey results (Tables 8.9-8.13, Figures 8.4-8.14) ..... 384
8.2.2 Commercial catch-per-unit-effort (Table 8.6, Figure 8.15) ..... 385
8.2.3 Age readings ..... 385
8.3 Data used in the assessment ..... 386
8.4 Methods used in the assessment ..... 387
8.4.1 Model settings ..... 387
8.5 Results of the assessment ..... 387
8.5.1 Biological reference points ..... 387
8.6 Comments to the assessment ..... 387
8.6.1 Future work ..... 387
8.7 Tables and figures ..... 388
9 Northeast Arctic anglerfish ..... 435
9.1 General. ..... 435
9.1.1 Species composition ..... 435
9.1.2 Stock description and management units ..... 435
9.1.3 Biology ..... 436
9.1.4 Fishery ..... 437
9.1.5 Scientific surveys ..... 438
9.2 Data ..... 438
9.2.1 Landings data ..... 438
9.2.2 Discards ..... 439
9.2.3 Length composition data ..... 439
9.2.4 Catch per unit effort (CPUE) data ..... 439
9.3 Methods and results ..... 440
9.3.1 The length-based-spawning-potential-ratio (LBSPR) approach ..... 440
9.3.2 CPUE standardization ..... 441
9.3.3 JABBA ..... 443
9.4 Management considerations and future investigations ..... 444
9.5 Tables and figures ..... 445
10 Barents Sea capelin ..... 467
10.1 Regulation of the Barents Sea capelin fishery ..... 467
10.2 TAC and catch statistics (Table 10.1) ..... 467
10.3 Stock assessment ..... 468
10.3.1 Acoustic stock size estimates in 2021 (Table 10.2, Figure 10.1, 10.2 and 10.3) ..... 468
10.3.2 Stock assessment in 2021 (Table 10.3-10.5, Figure 10.4) ..... 468
10.3.3 Recruitment ..... 469
10.3.4 Comments to the assessment ..... 469
10.3.5 Further work on survey and assessment methodology ..... 470
10.3.6 Reference points ..... 471
11 References ..... 487
Annex 1: List of participants ..... 500
Annex 2: Resolutions ..... 502
Annex 3: Working documents ..... 503
Annex 4: Audit reports ..... 504

## i Executive summary

On 30th March 2022, all Russian participation in ICES was suspended. Although the announcement of the suspension stressed the role of ICES as a "multilateral science organization", this suspension applied not only to research activities but also to the ICES work providing fisheries advice for the sustainable management of fish stocks and ecosystems. As a result of the suspension, it is not possible to run ICES stock assessments or provide ICES advice for the Barents Sea stocks of NEA cod, NEA haddock, Sebastes mentella or Greenland Halibut, as management and data collection for these stocks are shared between Norway and Russia. There are therefore no AFWG stock assessments for these stocks this year. This is especially unfortunate as NEA cod is currently declining, and updated assessments are required to ensure an appropriate management response. It is to be hoped that the political decision to exclude Russia from the ICES advice process which underlies our sustainable fisheries management does not lead to mismanagement of the shared stocks and the consequent ecological harm.

It should be noted that bilateral Russian-Norwegian advice is being provided to the managing body outside ICES for the affected stocks, and there is therefore no current management need for ICES advice. This year AFWG is therefore providing advice for saithe, coastal cod north, coastal cod south, and S. norvegicus. In addition, an assessment has been run for anglerfish, although there is no formal request for advice for this stock. The stock trends are as follows:

## Stock-by-stock summaries

Cod in subareas 1 and 2 North of $67^{\circ} \mathrm{N}$ (Norwegian coastal cod North); cod.27.2.coastN.

- The existing coastal cod north assessment and Blim from the 2022 benchmark gives an SSB estimate of 130671 tonnes, up from 116771 tonnes in 2021. An ICES HCR evaluation has been conducted at WKNCCHCR, which proposed slight modifications to the tuning data for the model. WKNCCHCR noted a high degree of uncertainty around any Blim estimate, and therefore proposed a HCR based on a precautionary F0.1 and no formal $\mathrm{B}_{\mathrm{lim}}$. This HCR has been adopted, and the catch advice for 2023 is 29347 tonnes. It should be noted that this stock cannot be directly managed via a quota (as the fish are not visually distinguishable from NEA cod in the same area), and therefore management is based on gear and area regulation.

Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$ (Norwegian coastal cod South); cod.27.2.coastS.

- The catch advice of 9136 tonnes is based on a standardized CPUE index, which increased to such an extent that the $+20 \%$ stability cap was reached. However, this index has high uncertainty, and auxiliary analyses show fairly poor status (SPR $=0.25$ and $\mathrm{F} / \mathrm{M}>1$ ). About half of the catch is immature, and this proportion has increased in the last 10 years.
Saithe in subareas 1 and 2 (Northeast Arctic)
- The NEA saithe stock is currently in good status, with the SSB well above Bpa at 715674 tonnes (up from 568 972) in last year's assessment. Following the HCR (and constrained by a $15 \%$ stability constr, the catch advice is 226794 tonnes (which is constrained by the $15 \%$ annual stability constraint). This stock, together with the associated North Sea saithe stock, is aiming for a benchmark, likely in 2024.
Redfish (Sebastes norvegicus) in subareas 1 and 2 (Northeast Arctic)
- The stock is continuing to be assessed as in a poor status, and with increasing catches is increasingly identified as overfished. A revision in the catch splitting between the two
redfish species resulted in an upwards revision of the catch and therefore SSB history but does not affect the overall downward trend of SSB in the assessment. The catch advice is therefore zero.
- As a result of a move to new age readers, a discrepancy in the age readings for older fish in the last three years compared with previous data was noted. This was dealt with by excluding the data on $30+$ fish in the tuning series, but this feeds into a strong desire for a benchmark for this stock before the next advice is due in 2024.


## Anglerfish (Lophius budegassa, Lophius piscatorius) in subareas 1 and 2 (Northeast Arctic)

- Data-limited model results based on length data from the fishery suggest that the biomass seems to be doing well and that the exploitation pattern is appropriate, while the rate might be near/slightly above the level that would lead to maximum yield. Management is based on technical measures rather than a quota. AFWG does not currently give advice on this stock but considers the current assessment of sufficient quality to base catch advice on if requested by the managers.


## Barents Sea capelin

- Following ToR b), the data on Barents Sea capelin were updated. No assessment is conducted during the spring AFWG meeting, the assessment occurs in autumn following the ecosystem survey (which in 2022 will be conducted outside ICES). An ICES benchmark will be held in late 2022 for this stock together with capelin in the IcelandEast Greenland-Jan Mayen area (WKCAPELIN).


## ii Expert group information

| Expert group name | Arctic Fisheries Working Group (AFWG) |
| :--- | :--- |
| Expert group cycle | Annual |
| Year cycle started | 2021 |
| Reporting year in cycle | $1 / 1$ |
| Chair | Daniel Howell, Norway |
| Meeting venues and dates | $21-27$ April $2022^{1}$, incl. data review meeting 20 April 2022, online (28 participants) |

[^1]
## 1 Introduction and ecosystem considerations

## 2022 report of the Arctic Fisheries Working Group

### 1.1 Terms of reference

## 2021/2/FRSG02

Approved November 2021
The Arctic Fisheries Working Group (AFWG), chaired by Daniel Howell, Norway, will meet online 21-27 April 2022 to:
a) Address generic ToRs for Regional and Species Working Groups, for all stocks except the Barents Sea capelin, which will be addressed at a meeting in autumn;
b) For Barents Sea capelin oversee the process of providing intersessional assessment;
c) Conduct reviews as required of time any series computed using the STOX and ECA open-source software for use in assessment in the Barents Sea.
The assessments will be carried out on the basis of the stock annex. The assessments must be available for audit on the first day of the meeting.

Material and data relevant to the meeting must be available to the group on the dates specified in the 2022 ICES data call.

AFWG will report by 6 May 2022 and October 2022 for Barents Sea capelin ${ }^{2}$ for the attention of the Advisory Committee.

Only experts appointed by national Delegates or appointed in consultation with the national Delegates of the expert's country can attend this Expert Group.

### 1.2 Additional requests

There were no additional requests.

### 1.3 Responses to terms of reference

Under ToR a (address generic ToRs), the stock assessments and advice were conducted according to generic ToRs c and d, while the generic ToR e benchmark review can be found further down in this introduction and the haddock, NEA cod, and coastal cod sections. Work on generic ToRs $a$ and $b$ will be conducted intersessionally as it becomes appropriate.

ToR $b$ is handled in detail by the capelin subgroup of AFWG, held in autumn after the capelin survey. A brief report on the previous capelin assessment is given in this report.

ToR c is to review data changes as required, and this was not required in 2021.

[^2]
### 1.4 Benchmarks

A cod benchmark (WKBARFAR 2021) was conducted in early 2021 (ICES, 2021a). This benchmark resulted in a modification of the existing NEA cod SAM assessment model. For coastal cod, the benchmark resulted in the stock being split into two, a category one northern stock (with a SAM stock assessment) and a category three southern stock (2-over-3 rule based on a CPUE series).

Capelin ${ }^{3}$ is scheduled to have a benchmark in 2022, with HCR revision conducted at the benchmark. Greenland halibut is scheduled for a benchmark in $2023^{4}$, followed by an HCR evaluation.

### 1.5 Total catches

In this report, the terms 'landings' and 'catches' are, somewhat incorrectly, used as synonyms, as discards are in no cases used in the assessments. This does not mean, however, that discards have not occurred, but the WG has no information on the possible extent. In contrast, available information indicates low discard rates at present (less than $5 \%$ of catch), and it is assumed that discards are negligible in the context of the precision of the advice.
In previous years a report from the Norwegian-Russian Analysis group dealing with estimation of total catch of cod and haddock in the Barents Sea in 2021 was available to AFWG. The report presents estimated catches made by Norwegian, Russian and third countries separately. According to that report, the total catches of both cod and haddock reported to AFWG are very close (within 1\%) to the estimates made by the analysis group. Thus, it was decided to set the IUU catches for 2021 to zero.

For further information on under- and misreporting, we refer to the 2016 AFWG report.
Discards estimates (1994-2021) of redfish, cod, haddock, and Greenland halibut juveniles in the commercial shrimp fishery in the Barents Sea are presented in Figure 0.1. These estimates are obtained with a spatio-temporal model based on a procedure elaborated in Breivik et al. (2017). In Breivik et al. (2017) an extensive validation study indicates that the new procedure obtains bycatch estimates with approximately correct uncertainty. Previous estimates for the period 1982-2015 are given in earlier reports (e.g. AFWG 2018), and we have not been able to compare these two time-series in detail. Such a comparison should be performed on a relatively fine spa-tio-temporal resolution. The bycatch estimates illustrated in Figure 0.1 and are available for each quarter in each main statistical area (not shown in report). Note that it is still a work in progress regarding improving the new estimates.

The new time-series in Figure 0.1 are obtained by scaling the estimated bycatch in the Norwegian fishery with the international fishery in each ICES area. The scaling procedure assumes that the Norwegian fishery is representative of the international fishery. This assumption is necessary because the international catch data are available only to a low spatio-temporal resolution. If the international vessels in a relatively high degree trawl at locations not trawled by Norwegian vessels, the bycatch estimates illustrated in figure 0.1 may be biased.

[^3]
### 1.5.1 Uncertainty in catch data

For the Norwegian estimates of catch numbers at-age and mean weight-at-age for cod and haddock methods for estimating the precision have been developed, and the work is still in progress (Aanes and Pennington, 2003; Hirst et al., 2004; Hirst et al., 2005; Hirst et al., 2012). The methods are general and can in principle be used for the total catch, including all countries' catches, and provide estimates both at-age and at-length groups. Typical error coefficients of variation for the catch numbers-at-age are in the range of $5-40 \%$ depending on age and year. It is evident that the estimates of the oldest fish are the most imprecise due to the small numbers in the catches and resulting small number of samples on these age groups. From 2006 onwards, the Norwegian catch-at-age in the assessment has been calculated using the ECA method described by Hirst et al. (2005). The methodology for using ECA to split cod catches into NEA cod and coastal cod is still under development (WKARCT 2015). ECA has now been implemented for saithe, and with partial success for $S$. mentella. A new version of the program (StoX-ECA) is now being tested.

Aging error is another source of uncertainty, which causes increased uncertainty in addition to bias in the estimates: An estimated age distribution appears smoother than it would have been in absence of ageing error. Some data have been analysed to estimate the precision in ageing (Aanes, 2002). If the ageing error is known, this can currently be taken into account for the estimation of catch-at-age described above.

For capelin, the uncertainty in the catch data is not evaluated. The catch data are used, however, only when parameters in the predation model are updated at infrequent intervals, and the uncertainty in the catch data are considered small compared with other types of uncertainties in the estimation.

We note that the SToX survey methodology reviewed by the group can produce uncertainty estimates for the survey time-series.
Additional sources of uncertainty arising from sources beyond sampling or age-reading errors have implications for a number of the stocks assessed here. Coastal cod catches, and to a lesser extent catches of the much larger NEA cod stock, have uncertainty issues due to the difficulty of splitting catches between the two stocks. A similar issue applies to small S. norvegicus stock and the larger $S$. mentella stock, where species misidentification can be a significant source of error. Finally, there is no agreement between Norway and Russia on an age-reading methodology for Greenland halibut, and such data are not used for tuning the model. The absence of age data creates an important (but unquantifiable) source of error on the GHL stock estimate.

### 1.5.2 Sampling effort-commercial fishery

Concerns about commercial sampling: The main Norwegian sampling program for demersal fish in ICES subareas 1 and 2 has been port sampling, carried out onboard a vessel travelling from port to port for approximately 6 weeks each quarter. A detailed description of this sampling program is given in Hirst et al. (2004). However, this program was, for economic reasons, terminated 1 July 2009. Sampling by the 'reference fleet' and the Coast Guard has increased in recent years. However, the reduction in port sampling of many different vessels seems to have increased the uncertainty in the catch-at-age estimates from 2009 onwards (WD6, 2010). A Norwegian port sampling program was restarted in 2011, although with a lower effort, this improved the basis for the 2011-2019 catch-at-age estimates. From 2014 this program is run by 4 -year contracts of a vessel that sails between fish landing sites along the coast from about $66^{\circ} \mathrm{N}$ to Varanger $\left(70^{\circ} \mathrm{N}, 30^{\circ} \mathrm{E}\right)$ three periods a year during the first, second, and fourth quarters, altogether up to 120 days. This is a reduction compared to about 180 days a year before 2009. The catch sampling is done of landed fish, mainly from the fleet fishing in coastal waters, and usually inside the
plant, and the rented vessel acts as a transport, accommodation and working (age reading, data work) platform. AFWG recommends that such sampling is also carried out during the third quarter.

Tables $0.1-0.4$ show the development of the Norwegian, Russian, Spanish and German sampling of commercial catches in the period 2008-2021. The tables show the total sampling effort, but do not show how well the sampling covers the fishery. Indices of coverage should be developed to indicate this. The main reason for the general strong decrease in numbers of Norwegian samples in the first part of this period is the termination of the port sampling program in northern Norway. This program is now up and running again. It should be considered whether catch sampling carried out by different countries fishing by trawl for the same time and area could be coordinated and data shared on a detailed level to a greater extent than is done today. Due to the Russian suspension not all these tables are updated with 2021 data.

### 1.5.2.1 Cod, haddock, and saithe

Available catch-at-age and length data covered the largest portion of catches by the respective fisheries. However, there was a period in spring 2020 when port sampling was at a lower level than usual due to the COVID-19 situation. However, the aggregation level (time and space) used when splitting these catches into Northeast Arctic cod and Norwegian Coastal Cod is also an important issue. Despite the improvement in sampling coverage in 2016-2020, the number of samples should be increased in the coming years, with the aim of covering all quarters and areas contributing the highest catches.

Due to the adopted amendments of the Russian Federal Law "On fisheries and preservation of aquatic biological resources" coming into force, especially concerning the destruction of biological resources caught under scientific research, sampling activities (age sample numbers and length/weight measurements of fish) on board fishing vessels are also reduced, especially in ICES subareas 2.a and 2.b, which may result in greater uncertainty of the stock assessments due to possible biases in the age-length distributions of the commercial catch.

Length measurements of fish and age sampling by Russia have been especially low in ICES subareas 2.a and 2.b in the first half of 2020 due to administrative difficulties in arrangement (stationing) observers onboard fishing vessels (a prolonged procedure via open contest). Available Norwegian data on cod and haddock length measurements onboard Russian vessels made by the Norwegian Coast Guard in the Norwegian economic zone have been used, where possible, in calculations of catch-at-age data by Russia.

### 1.5.2.2 Data issues with S. mentella

There is still a concern about the biological sampling from the fishery and scientific surveys that may have become critically low, however, there is also a lag of several years between collection of age samples and the processing of them. This is elaborated in the section for this stock.

### 1.5.2.3 Data issues with $S$. norvegicus

Despite a recent increase in age-reading for this species, age data are rather poor, and effort in age sampling from the catches is required. The other main source of uncertainty is species misidentification from S. mentella, and consequently, careful monitoring that species composition is being reported correctly is required.

### 1.5.2.4 Data issues with NEA Greenland halibut

There is still a concern about the biological sampling from the fishery that may have become critically low. Age information is not available, due to disagreements on age reading method, and may affect precision in the assessment which now is length-based. Norwegian landings are
split on Greenland halibut by sex for area, gear groups, and quarters. Annual sample level has decreased in the last years and may affect the precision of the catch distribution.

The samples and data basis behind each stock assessment are discussed more in detail under each stock-specific section of this report (e.g. the coastal cod). The number of aged individuals per 1000 t is now well below the standard set by the EU in their Data Collection regulations. For several stocks sampling is inadequate for area/quarter/gear combinations making up considerable proportions of the total catch.

Discontinuation of the Russian autumn survey decreased considerably the biological sampling (age sample numbers, abundance indices evaluations, maturity status of fish definitions, feeding data collections, etc.).

### 1.5.3 The percentage of the total catch that has been taken in the NEAFC regulatory areas by year in the last year

Generic ToR c-iii asks for the percentage of the total catch that has been taken in the NEAFC regulatory area by year in the last year. In the area where AFWG stocks are distributed, there are two areas outside national EEZs which are part of the NEAFC regulatory area: The International area in ICES Subarea 1 in the Barents Sea ("loophole", denoted as 1.a or 27_1_A) and the International area in ICES divisions 2.a and 2.b in the Norwegian Sea ("banana hole", denoted as 2.a. 1 and $2 . b .1$ or $27 \_2 \_A \_1$ and $27 \_2 \_B \_1$ ). In the table below the WG presents the most likely landings from these areas based on the official reports and discussions within the WG. The text table below shows the percentages for S. mentella, Northeast Arctic cod and haddock and Greenland halibut. For the other AFWG stocks, no catches are taken in those areas. The highest precision in these numbers is probably the $S$. mentella figures since these figures have been tabulated each year since 2004, and have been given regular and special attention, also by NEAFC.

|  | ICES 1.a | ICES 2.a.1 | ICES 2.b.1 | Total | \%NEAFC |
| :--- | :--- | :--- | :--- | :--- | :--- |
| NEA cod | 1896 | 2 | 0 | 758383 | $0.25 \%$ |
| Coastal cod (south+north) | 0 | 0 | 0 | 52705 | $0.0 \%$ |
| Commercial catches | 0 | 0 | 0 | 42043 | $0.0 \%$ |
| Recreational catches | 0 | 0 | 0 | 10662 | $0.0 \%$ |
| NEA haddock | 0 | 0 | 0 | 188175 | $0.0 \%$ |
| Sebastes mentella | 0 | 2 | 0 | 63482 | $4.5 \%$ |
| Sebastes norvegicus | 0 | 0 | 0 | 10193 | $0.0 \%$ |
| Greenland halibut | 638 | 0 | 0 | 0 | 0 |


|  | ICES 1.a | ICES 2.a. 1 | ICES 2.b. 1 | Total | \%NEAFC |
| :---: | :---: | :---: | :---: | :---: | :---: |
| NEA cod | 1607 | 9 | 0 | 692903 | 0.23\% |
| Coastal cod | 0 | 0 | 0 | 56653 | 0.0\% |
| NEA haddock | 0 | 0 | 0 | 182468 | 0.0\% |
| NEA saithe | 0 | 3 | 0 | 169405 | <0.1\% |
| Sebastes mentella | 0 | 5469 | 0 | 53631 | 10.2\% |
| Sebastes norvegicus | 0 | 0 | 0 | 9646 | 0.0\% |
| Greenland halibut | 450 | 0 | 0 | 28713 | 1.5\% |
| Capelin | 0 | 0 | 0 | 0 | 0.0\% |
| Anglerfish | 0 | 0 | 0 | 2280 | 0.0\% |
| 2019 |  |  |  |  |  |
| NEA cod | 1094 | 0 | 0 | 692609 | 0.16\% |
| Coastal cod | 0 | 0 | 0 | 52807 | 0.0\% |
| NEA haddock | 394 | 0 | 0 | 175402 | 0.225\% |
| NEA saithe | 250 | 7 | 0 | 163180 | 0.001\% |
| Sebastes mentella | 0 | 6060 | 0 | 45954 | 13.2\% |
| Sebastes norvegicus | 0 | 0 | 0 | 8285 | 0.0\% |
| Greenland halibut | 1108 | 3 | 0 | 28832 | 3.8\% |
| Capelin | 0 | 0 | 0 | 0 | 0.0\% |
| Anglerfish | 0 | 0 | 0 | 2809 | 0.0\% |
| 2018 |  |  |  |  |  |
| NEA cod | 1724 | 2 | 0 | 778627 | 0.22\% |
| Coastal cod | 0 | 0 | 0 | 49075 | 0.0\% |
| NEA haddock | 24.1 | 0 | 0 | 191276 | 0.013\% |
| NEA saithe | 2.4 | 0 | 0 | 181280 | 0.001\% |
| Sebastes mentella | 3 | 7823 | 0 | 38765 | 20.2\% |
| Sebastes norvegicus | 0 | 0 | 0 | 6647 | 0.0\% |
| Greenland halibut | 798 | 0 | 0 | 28544 | 2.80\% |
| Capelin | 0 | 0 | 0 | 0 | 0.0\% |
| Anglerfish | 0 | 0 | 0 | 1903 | 0.0\% |


|  | ICES 1.a | ICES 2.a. 1 | ICES 2.b. 1 | Total | \%NEAFC |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 |  |  |  |  |  |
| NEA cod | 1212 | 12 | 0 | 868276 | 0.14\% |
| Coastal cod | 0 | 0 | 0 | 51053 | 0.0\% |
| NEA haddock | 90 | 0 | 0 | 227588 | 0. $0004 \%$ |
| NEA saithe | 70 | 11 | 0 | 145403 | 0.06\% |
| Sebastes mentella | 0 | 6463 | 0 | 31200 | 20.7\% |
| Sebastes norvegicus | 5 | 0 | 0 | 5340 | 0.1\% |
| Greenland halibut | 592 | 6 | 0 | 26380 | 2.3\% |
| Capelin | 0 | 0 | 0 | 0 | 0.0\% |
| Anglerfish | 0 | 0 | 0 | 1478 | 0.0\% |
| 2016 |  |  |  |  |  |
| NEA cod | 3619 | 0 | 0 | 849422 | 0.4\% |
| Coastal cod | 0 | 0 | 0 | 54767 | 0.0\% |
| NEA haddock | 7 | 0 | 0 | 233416 | 0.003\% |
| NEA saithe | 81 | 0 | 0 | 140392 | 0.06\% |
| Sebastes mentella | 0 | 7170 | 0 | 35429 | 20.2\% |
| Sebastes norvegicus | 10 | 0 | 0 | 4674 | 0.2\% |
| Greenland halibut | 363 | 5 | 0 | 24972 | 1.5\% |
| Capelin | 0 | 0 | 0 | 0 | 0.0\% |
| Anglerfish | 0 | 0 | 0 | 1435 | 0.0\% |
| 2015 |  |  |  |  |  |
| NEA cod | 9 | 0 | 0 | 864384 | 0.001\% |
| Coastal cod | 0 | 0 | 0 | 35843 | 0.0\% |
| NEA haddock | 702 | 0 | 0 | 194756 | 0.4\% |
| NEA saithe | 30 | 0 | 0 | 131765 | 0.0\% |
| Sebastes mentella | 0 | 4752 | 0 | 25856 | 18.4\% |
| Sebastes norvegicus | 13 | 0 | 0 | 3632 | 0.4\% |
| Greenland halibut | 55 | 0 | 0 | 24748 | 0.2\% |
| Capelin | 0 | 0 | 0 | 115044 | 0.0\% |


|  | ICES 1.a | ICES 2.a. 1 | ICES 2.b. 1 | Total | \%NEAFC |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Anglerfish | 0 | 0 | 0 | 1043 | 0.0\% |
| 2014 |  |  |  |  |  |
| NEA cod | 534 | 0 | 0 | 986449 | 0.1\% |
| Coastal cod | 0 | 0 | 0 | 33660 | 0.0\% |
| NEA haddock | 0 | 0 | 0 | 177522 | 0.0\% |
| NEA saithe | 0 | 0 | 0 | 132005 | 0.0\% |
| Sebastes mentella | 0 | 4020 | 0 | 18780 | 21.4\% |
| Sebastes norvegicus | 0 | 0 | 0 | 4438 | 0.0\% |
| Greenland halibut | 211 | 0 | 0 | 23025 | 0.9\% |
| Capelin | 0 | 0 | 0 | 66000 | 0.0\% |
| Anglerfish | 0 | 0 | 0 | 1657 | 0.0\% |

### 1.6 Uncertainties in survey data

This section is retained for information, although 2021 data were not available to ICES due to the decision to suspend Russian participation. This section is therefore not updated for 2021.

While the area coverage of the winter surveys for demersal fish was incomplete in 1997 and 1998, the coverage was normal for these surveys in 1999-2002. In autumn 2002, 2006 and winter 2003, 2007, 2016 and 2017 however, surveys were again incomplete due to lack of access to both the Norwegian and Russian Economic Zones. This affects the reliability of some of the most important survey time-series for cod and haddock and consequently also the quality of the assessments.

It is very important that the Norwegian and Russian authorities give each other's research vessels full access to the respective economic zones when assessing the joint resources, as was the case for Joint winter surveys (BS-NoRu-Q1 (Btr) and BS-NoRu-Q1 (Aco)) in 2004-2005, 2008-2011 and 2013, for example. This is the case regardless of if advice is conducted within or outside ICES.

The area coverage in the winter survey was extended from 2014 onwards (Figure 0.2, Table 3.5). With the recent expansion of the cod distribution, it is likely that in years before 2014 the coverage in the February survey (BS-NoRu-Q1 (BTr) and BS-NoRu-Q1 (Aco)) has been incomplete, in particular for the younger ages. This could cause a bias in the assessment, but the magnitude is unknown. The 2014-2021 surveys covered considerably larger areas than earlier winter surveys and showed that cod, haddock and Greenland halibut was distributed far outside the standard survey area. The 2017 and 2018 surveys were restricted by ice Northeast of Hopen Island, and the survey did not extend quite as far as in the years 2014-2016. In 2019 the coverage was almost as extensive as in 2014. Coverage in 2020-2022 was less extensive mainly due to increased ice cover in the east. For all stocks except Greenland halibut, mainly younger age groups are found in the northern area. It should however be noted that the survey index from this survey is currently not used in the assessment of Greenland halibut.

The survey estimates within the new, extended area are now used for the tuning data for cod, but with the bottom trawl series split in 2014, as decided at the WKBARFAR 2021 benchmark.

For haddock, the new northern area is also included as decided at the WKDEM benchmark in 2020.

There are also other issues with incomplete survey coverage of stocks, e.g. haddock off the Norwegian coast south of Finnmark is not covered in the winter survey and the $S$. mentella survey in the Norwegian Sea does not cover the entire distribution area.

From 2004 onwards, a joint Norwegian-Russian survey has been conducted in August-September. This is a multi-purpose survey termed an "ecosystem survey" because most of the ecosystem is covered; including an acoustic survey for the pelagic species, which is used for capelin assessment, and a bottom trawl survey which includes non-commercial species. The ecosystem survey is now included in both cod and haddock assessments. The survey is also utilized in the assessment of redfish and Greenland halibut.

In 2018, a large area in the eastern Barents Sea was not covered due to technical problems with one vessel, while in 2019, most of the Barents Sea was covered except parts of the International waters and the Northeastern most part. In 2020 the spatial coverage was good, but for COVID19 related reasons, the survey was less synoptic than usual as the time between the start and end of the survey was 13 weeks while the normal is about 8 weeks. Also, one of the vessels used had not previously been used in this type of bottom trawl surveys. The bottom trawl survey indices for cod and haddock from this survey in 2020 were considerably lower than expected, in particular for cod, but it was decided to include them in the assessment. Also, the survey coverage for capelin was not complete at the time assessment and advice had to be provided. Although this did not affect the advice this year, which would have been zero catch even when using the final estimate for the entire area, that may not be the case in future. Spatial coverage in 2021 was good except that the International waters ("Loophole") was not covered.

It is very important that this survey should be continued with complete spatial coverage and as synoptic as possible. In addition to being the only survey used in capelin assessment and being used in assessment of demersal stocks, it has been shown to be valuable for sampling of synoptic ecosystem information, cover the entire area of fish distribution in the Barents Sea, and provide additional data on geographical distribution of demersal fish, which could prove valuable in future inclusion of more ecosystem information in the fish stock assessments.

The Norwegian coastal survey (NOcoast-Aco-4Q) has in its current design been conducted since 2002. The survey covers the coastal area, including most fjords, and shelf area, including banks, between Kirkenes in northern Norway and Stadt off central Norway. The survey area is divided into seventeen strata, each containing several substrata, and is generally covered by two vessels, which collect acoustic data along defined transects and catch and biological data from both fixed bottom trawl stations and trawl stations identifying acoustic registrations. The coverage of the area has been fairly consistent throughout the time-series. In 2020 bad weather prevented the coverage of three substrata in the southern part of the survey area. Historically the contribution of these areas to the saithe and coastal cod survey index has been low, and it is therefore assumed that the lack of coverage of these areas in the 2020 estimate will not affect the final survey index.


Bycatch of redfish



Figure 0.1. Estimated bycatch of cod, haddock, redfish and Greenland halibut in the Barents Sea shrimp fishery. Intervals are $90 \%$ confidence intervals.


Figure 0.2. Strata (1-26) and main areas ( $A, B, C, D, D^{\prime}, E$ and $S$ ) used for swept-area estimations and acoustic estimations with StoX. Strata (24-26, main area $N$ ) are covered since 2014, and are now included in the standard time-series.

After AFWG 2021 minor errors were discovered in the Norwegian SToX dataseries for 2021 for NEA cod and haddock. The advice has been updated and reflects the corrected data. However the values presented in this report are prior to the correction. More detail is given in the relevant stock sections.

### 1.7 Age reading

In 1992, PINRO, Murmansk and IMR, Bergen began a routine exchange program of cod otoliths to validate age readings and ensure consistency in age interpretations (Yaragina et al., 2009b, AFWG 2008, WD 20). Later, a similar exchange program has been established for haddock, capelin, and S. mentella otoliths. Once a year (now every second year, no exchanges of redfish age readers so far) the age readers have come together and evaluated discrepancies, which are seldom more than 1 year, and the results show an improvement over the period, despite still observing discrepancies for cod in the magnitude of $15-30 \%$. An observation that is supported by the results of an NEA cod otolith exchange between Norway, Russia, and Germany (Høie et al., 2009; AFWG 2009, WD 6). 100 cod otoliths were read by three Norwegian, two Russian and one German reader, reaching nearly $83 \%$ agreement (coefficient of variation $8 \%$ ). The age reading comparisons of these 100 cod otoliths show that there are no reading biases between readers within each country. However, there is a clear trend of bias between the readers from different countries, Russian age readers assign higher ages than the Norwegian and German age readers. This systematic difference is a source of concern and is also discussed in Yaragina et al. (2009b). This seems to be a persistent trend and will be revealed in the following annual otolith and age reader exchanges.

From 2009 onwards, it was decided to have meetings between cod and haddock otolith readers only every second year. The overall percentage agreement for the 2017-2018 exchange was $87.7 \%$ for cod (WD 08), which was a little lower than at the previous meeting. The general trend is that
the Russian readers assigned slightly higher ages than the Norwegian readers compared to the modal age for age group 7 years and older. The main reason for cod ageing discrepancies between Russian and Norwegian specialists was still a result of different interpretations of the false zones. This can partly be caused by different reading techniques, i.e. IMR reading opaque zones and PINRO reading translucent zones. For haddock, the main reason for discrepancies between PINRO and IMR readers was a different interpretation of the otolith summer structures in the first and second year of fish' life due to false zones. Sometimes discrepancies were caused by a different interpretation of the latest increments that were very thin in some cases.

For both species, the samples collected in autumn appeared to be the hardest to interpret. The main reason for that seems to be difficulties in determining if the marginal increment represents summer (opaque) or winter (translucent) growth.
A positive development is seen for haddock age readings showing that the frequency of a different reading (usually $\pm 1$ year) has decreased from above $25 \%$ in 1996-1997 to about $10 \%$ at present. The discrepancies are always discussed and a final agreement on the exchanged cod and haddock otoliths is achieved for all otoliths at present, except ca. $2-5 \%$. For haddock, the overall percentage agreement for recent data (2017-2018) was $88.1 \%$ and the precision CV was $3.0 \%$, the same values for cod totalled $87.7 \%$ and $3.7 \%$ accordingly and considered to be satisfactory.

The workshop on cod and haddock otolith reading planned for May-June of 2021 was delayed and the date for the next workshop in uncertain.

As the EU catches only make up a few percent ( $<10 \%$ ) of the total, the German and Spanish length and age data do not have a major impact on the assessment of the relevant stocks. But to use consistent datasets, regular age-reading comparisons should be made. EU age readers could be invited to the NOR-RUS exchanges and workshops.

To determine the effects of changes in age reading protocols between contemporary and historical practices, randomly chosen cod otolith material from each decade for the period 1940s-1980s has been re-read by experts (Zuykova et al., 2009). Although some year-specific differences in age determination were seen between historical and contemporary readers, there was no significant effect on length-at-age for the historical period. A small systematic bias in the number spawning zones detection was observed, demonstrating that the age at first maturation in the historic material as determined by the contemporary readers is younger than that determined by historical readers. The difference was largest in the first sampled years constituting approximately 0.6 years in 1947 and 1957. Then it decreased with time and was found to be within the range of $0.0-0.28$ years in the 1970-1980s. The study also shows that cod otoliths could be used for age and growth studies even after long storage.

For capelin otoliths, there is a very good correspondence between the Norwegian and Russian age readings, with a discrepancy in less than $5 \%$ of the otoliths. This was confirmed at the Nor-wegian-Russian age reading workshop on capelin in October 2011 (WD 13, 2012).

For some of the samples, a very high agreement was reached after the initial reading by the different experts. In other cases, some disagreement was evident after the first reading. After the initial reading, the results were analysed. The otoliths that caused disagreement were read again and discussed among the readers. After discussions about the reasons for disagreement, some readers wanted to change their view on some of the otoliths. When the samples were read once more, the agreement was $95 \%$.

It was concluded that experts from all laboratories normally interpret capelin otoliths equally. Difficult otoliths are sometimes interpreted differently, but these samples are few, and should not cause large problems for common work on capelin biology and stock assessment. All participants noted the great value of conducting joint work on otolith reading, and it was decided to continue the programme of capelin otolith exchange and to involve the labs at Iceland and

Newfoundland in the exchange program. Readers from Norway and Russia should continue to meet at Workshops every second year. A capelin age reading Workshop was held in Murmansk in April 2016, and the report from that meeting was presented to the capelin assessment meeting in October 2016. An age reading Workshop for capelin was held in Murmansk in October 2019.

In order to achieve the most accurate age estimates, ICES recommends methods and best practices for age reading of both redfish and Greenland halibut. Still there continue to be differences in opinion between PINRO and IMR regarding age reading methods for these species. It is recommended to start an annual or biannual exchange of otoliths and age reading experts on these species in order to identify the differences in interpretation and to discuss possibilities for a common approach.

The report from the Workshop on Age Reading of Greenland Halibut (WKARGH; ICES CM 2011/ACOM:41) described and evaluated several age reading methods for Greenland Halibut. A second workshop (WKARGH 2) was conducted in August 2016 and worked on further validation on new age reading methods. The workshop recommended that two new methods can be used to provide age estimations for stock assessments. Further, recognizing some bias and low precision in methods, the WKARGH2 recommends that an ageing error matrix or growth curve with error be provided for use in future stock assessments (WKARGH2 report 2016, ICES CM 2016/SSGIEOM:16). WKARGH2 recommends regular inter-lab calibration exercises to improve precision (i.e. exchange of digital images between readers for each method and between methods). The new age readings are not comparable with older data or the Russian age readings, and the new methods show that the species is more slow-growing and vulnerable than the previous age readings suggest. AFWG suggests that Russian and Norwegian scientists and age readers meet to work out issues of disagreements on Greenland halibut aging.

From 2009 onwards, an exchange of Sebastes mentella otoliths is conducted annually between the Norwegian and Russian laboratories (see section 6.2.2). In 2011 ICES/PGCCDBS identified differences in the interpretation of age structure by different national laboratories and recommended that international exchanges of otoliths be conducted (ICES C.M. 2011/ACOM:40). The work was conducted during 2011 (Heggebakken, 2011) with participation from Canada, Iceland, Norway, Poland and Spain. Unfortunately, Russia did not respond to the invitation to participate. The agreement in age determination was $79.2 \%$ (with allowance for $\pm 1$ years) for all ages combined, but $38.6 \%$ when only fish older than 20 years were considered. It is recommended that 1) future exchanges be conducted every $3-5$ years, 2 ) that these should primarily focus on $20+$-year-old fish and 3) that Russian scientists contribute to future exchanges. A meeting between S. mentella age readers from Norway and Russia was held in 2013. Otolith exchanges took place in 2014. It is recommended that such meetings and otolith exchanges be conducted regularly in future.

### 1.8 Assessment method issues

For coastal cod, the benchmark has resulted in a split into two stocks. For the northern (north of 67 degrees) part there is now a SAM assessment model. There is also a newly adopted HCR to provide target fishing mortality, however there was not sufficient information to provide a reliable Blim. In addition, since this is the first assessment model it is likely that there will be a need for a revision once we accumulate some years' experience running the model. The southern (between 62 and 67 degrees north) now gives advice based on a 2-over- 3 rule. A surplus production, based on the reference fleet CPUE, was developed. However, the CPUE time-series was too short to adequately tune the model. This should be investigated further as the time-series is extended, with a view to an eventual benchmark and adoption of the production model for assessment purposes.

Work is in progress on revising the capelin assessment methodologies, with a planned benchmark (in conjunction with Iceland) in 2022. Greenland halibut also has a benchmark (again jointly with Iceland) in 2022, planned to be followed by an HCR evaluation. For Greenland halibut the target F is the key issue, with the previous $\mathrm{F}_{\mathrm{pa}}$ being rejected by the Advice Drafting Group. A revised $\mathrm{F}_{\mathrm{pa}}$ has therefore been submitted. Although both capelin and Greenland halibut are being benchmarked through ICES, these are joint Norwegian-Russian stocks, and these models will not be used for ICES advice until the Russian suspension is lifted.

### 1.9 Environmental information included in the advice of NEA cod

For the fourteenth time, environmental information has been applied in the advice from AFWG. In this year's assessment ecosystem information was directly used in the projection of NEA cod. A combination of regression models, which is based on both climate and stock parameters, were used for the prediction of recruitment-at-age 3 , see section 1.11.4.

In addition, the temperature is part of the NEA cod consumption calculations that goes into the historical back-calculations of the number of cod, haddock, and capelin eaten by cod.

### 1.10 Proposals for status of assessments in 2022-2023

For anglerfish there is currently no advice, however following the benchmark in 2018 we are now able to conduct an assessment and provide advice if requested to do so. Greenland halibut is assessed this year and will be benchmarked in 2022, although following the Russian suspension there will be no ICES advice in 2022. AFWG is providing advice for Sebastes norvegicus, but the next advice here will be in 2024, it is to be hoped following a benchmark.

Therefore we anticipate providing ICES assessments in 2022 for northern and southern coastal cod, saithe, and background information for managers on anglerfish. Given an absence of tuning data and the presence of external advice used by managers, there no plans to produce ICES advice for NEA cod, NEA haddock, Sebastes mentella, Greenland halibut and capelin until the Russian suspension is lifted.

For saithe the plan is a benchmark in 2024 together with North Sea saithe.

Table 0.1. Age and length sampling by Norway of commercial catches in 2008-2021. Number of samples and average number of fish per sample. Also, number of age samples and aged individuals per 1000 t caught. For comparison, also the EU DCF requirements are shown.

| Year | No of unique vessels | No of length samples | No of lengthmeasured individuals | No of unique vessels (***) | No of age samples | No of aged individuals | Landing tonnes | Lengthsamples per 1000 t | Age samples per 1000 t | Aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |

NEA-cod + coastal cod

| 2008 | 336 | 2526 | 51263 |  | 464 | 16026 | 196067 | 12.9 | 2.4 | 81.7 | 125 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 272 | 2669 | 53350 |  | 417 | 14170 | 224816 | 11.9 | 1.9 | 63.0 | 125 |
| 2010 | 175 | 2542 | 39733 |  | 338 | 7671 | 263816 | 9.6 | 1.3 | 29.1 | 125 |
| 2011 | 273 | 2305 | 46227 |  | 434 | 10043 | 331535 | 7.0 | 1.3 | 30.3 | 125 |
| 2012 | 356 | 3132 | 57954 |  | 618 | 14710 | 363207 | 8.6 | 1.7 | 40.5 | 125 |
| 2013 | 266 | 2917 | 81583 | 84 | 1275 | 13940 | 464258 | 6.3 | 2.7 | 30.0 | 125 |
| 2014 | 556 | 2063 | 254627 | 306 | 1170 | 14815 | 465554 | 4.4 | 2.5 | 31.8 | 125 |
| 2015 | 498 | 1654 | 130514 | 89 | 1392 | 16500 | 413741 | 4.0 | 3.4 | 39.9 | 125 |
| 2016 | 482 | 2500 | 91590 | 401 | 1398 | 17027 | 403907 | 6.2 | 3.5 | 42.2 | 125 |
| 2017 | 413 | 2615 | 91366 | 348 | 1458 | 15471 | 408423 | 6.4 | 3.6 | 37.9 | 125 |
| 2018 | 873 | 3163 | 122788 | 346 | 1545 | 15535 | 369897 | 8.6 | 4.2 | 42.0 | 125 |
| 2019 | 842 | 3093 | 135375 | 337 | 1457 | 12519 | 322233 | 9.6 | 4.5 | 38.9 | 125 |
| 2020 | 389 | 1869 | 53587 | 259 | 653 | 12431 | 334773 | 5.6 | 2.0 | 37.1 | 125 |

NEA-haddock

| Year | No of <br> unique ves- <br> sels | No of <br> length sam- <br> ples | No of <br> length- <br> measured <br> individuals | No of <br> unique ves- <br> sels (**) | No of age <br> samples | No of aged <br> individuals | Land- <br> ing tonnes | Length- <br> samples per <br> 1000 $t$ | Age sam- <br> ples per <br> 1000 $t$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| Year | No of unique vessels | No of length samples | No of lengthmeasured individuals | No of unique vessels (***) | No of age samples | No of aged individuals | Landing tonnes | Lengthsamples per 1000 t | Age samples per 1000 t | Aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 152 | 1210 | 17412 |  | 215 | 4843 | 143314 | 8.4 | 1.5 | 33.8 | 125 |
| 2012 | 209 | 1474 | 19191 |  | 204 | 4113 | 143104 | 10.3 | 1.4 | 28.7 | 125 |
| 2013 | 87 | 1570 | 69469 | 69 | 788 | 5507 | 111981 | 14.0 | 7.0 | 49.2 | 125 |
| 2014 | 192 | 697 | 54365 | 94 | 575 | 5390 | 115880 | 6.0 | 5.0 | 46.5 | 125 |
| 2015 | 206 | 839 | 69375 | 43 | 614 | 6484 | 114830 | 7.3 | 5.3 | 56.5 | 125 |
| 2016 | 226 | 1448 | 52376 | 151 | 737 | 7278 | 121710 | 11.9 | 6.1 | 59.8 | 125 |
| 2017 | 195 | 1416 | 42812 | 141 | 788 | 6348 | 128651 | 11.0 | 6.1 | 49.3 | 125 |
| 2018 | 388 | 1665 | 43938 | 148 | 823 | 6937 | 162454 | 10.2 | 5.1 | 42.7 | 125 |
| 2019 | 380 | 1629 | 43503 | 136 | 817 | 6552 | 144133 | 11.3 | 5.7 | 45.5 | 125 |
| 2020 |  |  |  |  |  |  |  |  |  |  |  |
| Beaked redfish (S. Norvegicus) |  |  |  |  |  |  |  |  |  |  |  |
| 2008 | 104 | 1093 | 18305 |  | 98 | 2281 | 6180 | 176.9 | 15.9 | 369.1 | 125 |
| 2009 | 66 | 1131 | 17386 |  | 96 | 2302 | 6215 | 182.0 | 15.4 | 370.4 | 125 |
| 2010 | 49 | 1050 | 19339 |  | 97 | 2164 | 6515 | 161.2 | 14.9 | 332.2 | 125 |
| 2011 | 75 | 1064 | 16347 |  | 106 | 2310 | 4645 | 229.1 | 22.8 | 497.3 | 125 |
| 2012 | 78 | 993 | 12994 |  | 76 | 1297 | 4250 | 39.1 | 3.1 | 56.7 | 125 |
| 2013 | 35 | 654 | 627 | 17 | 74 | 1122 | 4244 | 154.1 | 17.4 | 264.4 | 125 |


| Year | No of unique vessels | No of length samples | No of lengthmeasured individuals | No of unique vessels (***) | No of age samples | No of aged individuals | Landing tonnes | Lengthsamples per 1000 t | Age samples per 1000 t | Aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2014 | 24 | 66 | 919 | 24 | 24 | 365 | 3053 | 21.6 | 7.9 | 119.6 | 125 |
| 2015 | 28 | 121 | 3497 | 22 | 405 | 1281 | 2492 | 48.6 | 162.5 | 514.0 | 125 |
| 2016 | 54 | 642 | 2376 | 36 | 517 | 1585 | 4606 | 139.4 | 112.2 | 344.1 | 125 |
| 2017 | 69 | 695 | 6177 | 44 | 571 | 1633 | 3354 | 207.2 | 170.2 | 486.9 | 125 |
| 2018 | 64 | 778 | 7354 | 32 | 629 | 1252 | 4287 | 181.5 | 146.7 | 292.0 | 125 |
| 2019 | 34 | 850 | 10007 | 34 | 226 | 1819 | 5951 | 142.8 | 38.0 | 305.7 | 125 |
| 2020 | 37 | 822 | 10176 | 37 | 193 | 1537 | 6503 | 126.4 | 29.7 | 236.3 | 125 |
| 2021 | 31 | 916 | 11069 | 31 | 0 | 0 | 7701 | 118.9 | 0 | 0 | 125 |
| Golden redfish (S. mentella) ** |  |  |  |  |  |  |  |  |  |  |  |
| 2008 | 13 | 178 | 1038 |  | 0 | 0 | 2214 | 80.4 | 0.0 | 0.0 | 125 |
| 2009 | 12 | 319 | 1841 |  | 2 | 40 | 2567 | 124.3 | 0.8 | 15.6 | 125 |
| 2010 | 11 | 284 | 3664 |  | 11 | 320 | 2245 | 126.5 | 4.9 | 142.5 | 125 |
| 2011 | 9 | 255 | 3210 |  | 11 | 298 | 2690 | 94.8 | 4.1 | 110.8 | 125 |
| 2012 | 13 | 166 | 2187 |  | 13 | 241 | 2098 | 79.1 | 6.2 | 114.9 | 125 |
| 2013 | 14 | 184 | 383 | 5 | 13 | 390 | 1361 | 135.2 | 9.6 | 286.6 | 125 |
| 2014 | 11 | 36 | 4664 | 12 | 49 | 5 | 13402 | 2.7 | 3.7 | 0.4 | 125 |
| 2015 | 21 | 166 | 23794 | 10 | 21 | 184 | 19700 | 8.4 | 1.1 | 9.3 | 125 |


| Year | No of unique vessels | No of length samples | No of lengthmeasured individuals | No of unique vessels (***) | No of age samples | No of aged individuals | Landing tonnes | Lengthsamples per 1000 t | Age samples per 1000 t | Aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 | 23 | 285 | 5470 | 9 | 22 | 169 | 19083 | 15.0 | 1.2 | 8.9 | 125 |
| 2017 | 30 | 256 | 3196 | 24 | 211 | 24 | 17280 | 14.8 | 12.2 | 1.4 | 125 |
| 2018 | 39 | 409 | 8782 | 20 | 364 | 25 | 19287 | 21.2 | 18.9 | 1.3 | 125 |
| 2019 | 17 | 352 | 5897 | 17 | 38 | 329 | 23844 | 14.8 | 1.6 | 13.8 | 125 |
| 2020 | 19 | 494 | 10963 | 19 | 76 | 694 | 32950 | 15.0 | 2.3 | 21.1 | 125 |
| 2021 | 16 | 627 | 17161 | 16 | 0 | 0 | 43797 | 14.3 | 0 | 0 | 125 |
| Greenland halibut |  |  |  |  |  |  |  |  |  |  |  |
| 2008 | 53 | 580 | 9074 |  | 0 | 0 | 7394 | 78.4 | 0.0 | 0.0 | 125 |
| 2009 | 36 | 922 | 12853 |  | 0 | 0 | 8446 | 109.2 | 0.0 | 0.0 | 125 |
| 2010 | 26 | 519 | 8395 |  | 0 | 0 | 7685 | 67.5 | 0.0 | 0.0 | 125 |
| 2011 | 29 | 463 | 8204 |  | 0 | 0 | 8273 | 56.0 | 0.0 | 0.0 | 125 |
| 2012 | 34 | 610 | 7716 |  | 0 | 0 | 10074 | 60.6 | 0.0 | 0.0 | 125 |
| 2013 | 26 | 597 | 4930 |  | 0 | 0 | 12613 | 47.3 | 0.0 | 0.0 | 125 |
| 2014 | 33 | 236 | 2559 | 10 | 0 | 0 | 10876 | 21.7 | 0.0 | 0.0 | 125 |
| 2015 | 31 | 273 | 8769 | 11 | 0 | 0 | 10704 | 25.5 | 0.0 | 0.0 | 125 |
| 2016 | 83 | 384 | 2304 | 60 | 0 | 0 | 12573 | 30.5 | 0.0 | 0.0 | 125 |
| 2017 | 67 | 556 | 10022 | 43 | 317 | 0 | 13194 | 42.1 | 24.0 | 0.0 | 125 |


| Year | No of unique vessels | No of length samples | No of lengthmeasured individuals | No of unique vessels (***) | No of age samples | No of aged individuals | Land- <br> ing tonnes | Length- <br> samples per <br> 1000 t | Age samples per 1000 t | Aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2018 | 96 | 582 | 11720 | 63 | 342 | 0 | 14876 | 39.1 | 23.0 | 0.0 | 125 |
| 2019 | 61 | 394 | 9286 | 47 | 80 | 0 | 14813 | 26.6 | 5.4 | 0.0 | 125 |
| 2020 | 80 | 429 | 9110 | 52 | 80 | 0 | 14532 | 29.5 | 5.5 | 0.0 |  |
| Anglerfish**** |  |  |  |  |  |  |  |  |  |  |  |
| 2013 | 8 | 55 | 1551 | 0 | 0 | 0 | 2988 | 18 | 36.5 | 0.0 | 125 |
| 2014 | 8 | 33 | 836 | 0 | 0 | 0 | 1655 | 19 | 18.1 | 24.8 | 125 |
| 2015 | 8 | 74 | 2054 | 0 | 0 | 0 | 933 | 82 | 35.3 | 0.0 | 125 |
| 2016 | 8 | 57 | 1339 | 0 | 0 | 0 | 1355 | 41 | 17.9 | 0.0 | 125 |
| 2017 | 8 | 88 | 3604 | 0 | 0 | 0 | 1473 | 59 | 23.8 | 0.7 | 125 |
| 2018 | 8 | 94 | 3233 | 0 | 0 | 0 | 1884 | 49 | 24.4 | 1.1 | 125 |
| 2019 | 8 | 68 | 3223 | 0 | 0 | 0 | 2750 | 24 | 22.5 | 0.0 | 125 |
| 2020 | 8 | 89 | 4129 | 0 | 0 | 0 | 2258 | 39 | 0 | 0.0 |  |
| Capelin |  |  |  |  |  |  |  |  |  |  |  |
| 2008 | 4 | 3 | 150 |  | 0 | 0 | 5000 | 0.6 | 0.0 | 0.0 | 125 |
| 2009 | 18 | 97 | 7039 |  | 39 | 1039 | 233000 | 0.4 | 0.2 | 4.5 | 125 |
| 2010 | 75 | 230 | 6191 |  | 47 | 1291 | 246000 | 0.9 | 0.2 | 5.2 | 125 |
| 2011 | 115 | 315 | 8346 |  | 48 | 1313 | 273000 | 1.2 | 0.2 | 4.8 | 125 |


| Year | No of unique vessels | No of length samples | No of lengthmeasured individuals | No of unique vessels (***) | No of age samples | No of aged individuals | Landing tonnes | Lengthsamples per 1000 t | Age samples per 1000 t | Aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | 84 | 308 | 9337 |  | 29 | 843 | 181328 | 1.7 | 0.2 | 4.6 | 125 |
| 2013 | 12 | 213 | 12215 | 47 | 47 | 773 | 156340 | 1.4 | 0.3 | 4.9 | 125 |
| 2014 | 27 | 113 | 9054 | 1 | 8 | 1086 | 40021 | 2.8 | 0.2 | 27.1 | 125 |
| 2015 | 65 | 722 | 83776 | 65 | 722 | 5393 | 71435 | 10.1 | 10.1 | 75.5 | 125 |
| 2016 | 7 | 27 | 1863 | 7 | 27 | 649 |  |  |  |  | 125 |
| 2017 | 21 | 43 | 2294 | 14 | 25 | 305 |  |  |  |  | 125 |
| 2018 | 68 | 207 | 15022 | 33 | 76 | 823 | 123461 | 1.7 | 0.6 | 6.7 | 125 |
| 2019 | 4 | 26 | 260 | 2 | 13 | 0 | 0 |  |  |  | 125 |
| 2020 |  |  |  |  |  |  | 0 |  |  |  |  |

## **In addition to age the otoliths are also used for identification of coastal cod.

**Age samples from surveys with commercial trawl come in addition.
***From 2013 No. of unique vessels are split by length and age samples.
**** Only from large, meshed gillnets as basis for assessment.

Table 0.2. Age and length sampling by Russia of commercial catches and age sampling of surveys in 2008-2020. Also length-measured individuals and aged individuals per $\mathbf{1 0 0 0} \mathbf{t}$ caught. For comparison also the EU DCF requirements are shown.

| Year | No of lengthmeasured individuals (commercial catches) | No of aged individuals (commercial catches) | No of aged individuals (surveys) | Total no of aged individuals | Landings tonnes | Length-measured individuals per 1000 t | Aged individuals per 1000 t (commercial catches) | Total aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| NEA-cod* |  |  |  |  |  |  |  |  |  |
| 2008 | 380592 | 3097 | 7565 | 10662 | 190225 | 2001 | 16.3 | 56.0 | 125 |
| 2009 | 178038 | 1075 | 7426 | 8501 | 229291 | 776 | 4.7 | 37.1 | 125 |
| 2010 | 126502 | 1828 | 7670 | 9498 | 267547 | 473 | 6.8 | 35.5 | 125 |
| 2011 | 122623 | 2376 | 5783 | 8159 | 310326 | 395 | 7.7 | 26.3 | 125 |
| 2012*** | 140028 | 2040 | 7742 | 9782 | 329943 | 424 | 6.2 | 29.6 | 125 |
| 2013 | 131455 | 1999 | 8103 | 10102 | 432314 | 304 | 4.6 | 23.4 | 125 |
| 2014 | 114538 | 3110 | 7154 | 10264 | 433479 | 264 | 7.2 | 23.7 | 125 |
| 2015*** | 105721 | 2486 | 6095 | 8581 | 381188 | 277 | 6.5 | 22.5 | 125 |
| 2016 | 158006 | 5090 | 2704 | 7794 | 394107 | 401 | 12.9 | 19.8 | 125 |
| 2017 | 161192 | 4918 | 6121 | 11039 | 396195 | 407 | 12.4 | 27.9 | 125 |
| 2018 | 157048 | 3129 | 1982 | 5111 | 340364 | 461 | 9.2 | 15.0 | 125 |
| 2019*** | 83018 | 2093 | 3737 | 5830 | 316813 | 262 | 6.6 | 18.4 | 125 |
| 2020*** | 112950 | 3105 | 3858 | 6963 | 312683 | 361 | 9.9 | 22.3 | 125 |

NEA-haddock

| Year | No of lengthmeasured individuals (commercial catches) | No of aged individuals (commercial catches) | No of aged individuals (surveys) | Total no of aged individuals | Landings tonnes | Length-measured individuals per 1000 t | Aged individuals per 1000 t (commercial catches) | Total aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 216959 | 2498 | 5677 | 8175 | 68792 | 3154 | 36.3 | 118.8 | 125 |
| 2009 | 43254 | 489 | 5421 | 5910 | 85514 | 506 | 5.7 | 69.1 | 125 |
| 2010 | 85445 | 834 | 5060 | 5894 | 111372 | 767 | 7.5 | 52.9 | 125 |
| 2011 | 61990 | 1570 | 3584 | 5154 | 139912 | 443 | 11.2 | 36.8 | 125 |
| 2012*** | 87880 | 1545 | 5034 | 6579 | 143886 | 611 | 10.7 | 45.7 | 125 |
| 2013 | 42927 | 1205 | 4021 | 5226 | 85668 | 501 | 14.1 | 61.0 | 125 |
| 2014 | 45447 | 899 | 3796 | 4695 | 78725 | 577 | 11.4 | 59.6 | 125 |
| 2015*** | 31009 | 914 | 2972 | 3886 | 91864 | 338 | 9.9 | 42.3 | 125 |
| 2016 | 55598 | 2691 | 1884 | 4575 | 115710 | 480 | 23.3 | 39.5 | 125 |
| 2017 | 74297 | 3554 | 2614 | 6168 | 106714 | 696 | 33.3 | 57.8 | 125 |
| 2018 | 61360 | 2274 | 1136 | 3410 | 90486 | 678 | 25.1 | 37.7 | 125 |
| 2019*** | 44728 | 1923 | 1778 | 3701 | 76125 | 588 | 25.3 | 48.6 | 125 |
| 2020*** | 69301 | 2356 | 1575 | 3931 | 89030 | 778 | 26.5 | 44.2 | 125 |
| NEA-saithe |  |  |  |  |  |  |  |  |  |
| 2008 | 8865 | 479 | 175 | 654 | 11577 | 766 | 41.4 | 56.5 | 125 |
| 2009 | 5279 | 7 | 68 | 75 | 11899 | 444 | 0.6 | 6.3 | 125 |
| 2010 | 422 | 112 | 249 | 361 | 14664 | 29 | 7.6 | 24.6 | 125 |


| Year | No of lengthmeasured individuals (commercial catches) | No of aged individuals (commercial catches) | No of aged individuals (surveys) | Total no of aged individuals | Landings tonnes | Length-measured individuals per 1000 t | Aged individuals per 1000 t (commercial catches) | Total aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 88 | 9 | 27 | 36 | 10007 | 9 | 0.9 | 3.6 | 125 |
| 2012 | 4062 | 145 | 104 | 249 | 13607 | 299 | 10.7 | 18.3 | 125 |
| 2013 | 17124 | 402 | 76 | 478 | 14796 | 1157 | 27.2 | 32.3 | 125 |
| 2014 | 2302 | 278 | 26 | 304 | 12396 | 186 | 22.4 | 24.5 | 125 |
| 2015 | 1505 | 104 | 131 | 235 | 13181 | 114 | 7.9 | 17.8 | 125 |
| 2016 | 4233 | 272 | 16 | 288 | 15203 | 278 | 17.9 | 18.9 | 125 |
| 2017 | 1762 | 228 | 110 | 338 | 14551 | 121 | 15.7 | 23.2 | 125 |
| 2018 | 4758 | 454 | 9 | 463 | 14171 | 336 | 32.0 | 32.7 | 125 |
| 2019 | 4528 | 94 | 0 | 94 | 13990 | 324 | 6.7 | 6.7 | 125 |
| 2020 | 83 | 17 | 96 | 113 | 14082 | 6 | 1.2 | 8.0 | 125 |
| S. norvegicus |  |  |  |  |  |  |  |  |  |
| 2008 | 1196 | 45 | 17 | 62 | 749 | 1597 | 60.1 | 82.8 | 125 |
| 2009 | 241 | 2 | 27 | 29 | 698 | 345 | 2.9 | 41.5 | 125 |
| 2010 | 486 | 25 | 199 | 224 | 806 | 603 | 31.0 | 277.9 | 125 |
| 2011 | 885 | 77 | 62 | 139 | 919 | 963 | 83.8 | 151.3 | 125 |
| 2012 | 1564 | 58 | 54 | 112 | 681 | 2297 | 85.2 | 164.5 | 125 |
| 2013 | 770 | 22 | 142 | 164 | 797 | 966 | 27.6 | 205.8 | 125 |


| Year | No of lengthmeasured individuals (commercial catches) | No of aged individuals (commercial catches) | No of aged individuals (surveys) | Total no of aged individuals | Landings tonnes | Length-measured individuals per 1000 t | Aged individuals per 1000 t (commercial catches) | Total aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2014 | 589 | 25 | 33 | 58 | 806 | 731 | 31.0 | 72.0 | 125 |
| 2015 | 120 |  | 20 | 20 | 664 | 181 | 0.0 | 30.1 | 125 |
| 2016 | 1113 | 147 | 34 | 181 | 776 | 1434 | 189.4 | 233.2 | 125 |
| 2017 | 1426 | 86 | 101 | 187 | 1131 | 1261 | 76.0 | 165.3 | 125 |
| 2018 | 1877 | 30 | 21 | 51 | 1546 | 1214 | 19.4 | 33.0 | 125 |
| 2019 | 1015 | 150 | 0 | 150 | 1804 | 563 | 83.2 | 83.2 | 125 |
| 2020 | 2107 | 47 | 31 | 78 | 2492 | 846 | 18.9 | 31.3 | 125 |
| S. mentella |  |  |  |  |  |  |  |  |  |
| 2008 | 21446 | 471 | 3379 | 3850 | 7117 | 3013 | 66.2 | 541.0 | 125 |
| 2009 | 29435 | 761 | 1447 | 2208 | 3843 | 7659 | 198.0 | 574.6 | 125 |
| 2010 | 2776 | 100 | 2295 | 2395 | 6414 | 433 | 15.6 | 373.4 | 125 |
| 2011 | 917 | 7 | 640 | 647 | 5037 | 182 | 1.4 | 128.4 | 125 |
| 2012 | 7802 | 422 | 1146 | 1568 | 4101 | 1902 | 102.9 | 382.3 | 125 |
| 2013 | 19092 | 1253 | 1625 | 2878 | 3677 | 5192 | 340.8 | 782.7 | 125 |
| 2014 | 817 | 25 | 1297 | 1322 | 1704 | 479 | 14.7 | 775.8 | 125 |
| 2015 | 771 |  | 1818 | 1818 | 1142 | 675 | 0.0 | 1591.9 | 125 |
| 2016 | 27765 | 1076 | 85 | 1161 | 8419 | 3298 | 127.8 | 137.9 | 125 |


| Year | No of lengthmeasured individuals (commercial catches) | No of aged individuals (commercial catches) | No of aged individuals (surveys) | Total no of aged individuals | Landings tonnes | Length-measured individuals per 1000 t | Aged individuals per 1000 t (commercial catches) | Total aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | 958 | 99 | 1000 | 1099 | 4952 | 193 | 20.0 | 221.9 | 125 |
| 2018 | 21004 | 845 | 39 | 884 | 10497 | 2001 | 80.5 | 84.2 | 125 |
| 2019 | 6881 | 400 | 469 | 869 | 13164 | 523 | 30.4 | 66.0 | 125 |
| 2020 | 8718 | 340 | 612 | 952 | 13997 | 623 | 24.3 | 68.0 | 125 |
| Greenland halibut |  |  |  |  |  |  |  |  |  |
| 2008 | 106411 | 1519 | 3366 | 4885 | 5294 | 20100 | 286.9 | 922.7 | 125 |
| 2009 | 77554 | 819 | 2282 | 3101 | 3335 | 23255 | 245.6 | 929.8 | 125 |
| 2010 | 32090 | 416 | 2784 | 3200 | 6888 | 4659 | 60.4 | 464.6 | 125 |
| 2011 | 9892 | 115 | 1541 | 1656 | 7053 | 1403 | 16.3 | 234.8 | 125 |
| 2012 | 82943 | 2140 | 2506 | 4646 | 10041 | 8260 | 213.1 | 462.7 | 125 |
| 2013 | 12608 | 555 | 2756 | 3311 | 10310 | 1223 | 53.8 | 321.1 | 125 |
| 2014 | 24346 | 633 | 2106 | 2739 | 10061 | 2420 | 62.9 | 272.2 | 125 |
| 2015 | 22116 | 575 | 2489 | 3064 | 12953 | 1707 | 44.4 | 236.5 | 125 |
| 2016 | 11818 | 574 | 221 | 795 | 10576 | 1117 | 54.3 | 75.2 | 125 |
| 2017 | 24061 | 1205 | 1579 | 2784 | 10713 | 2246 | 112.5 | 259.9 | 125 |
| 2018 | 21893 | 954 | 308 | 1262 | 12072 | 1814 | 79.0 | 104.5 | 125 |
| 2019 | 861 | 125 | 1552 | 1677 | 12198 | 71 | 10.2 | 137.5 | 125 |


|  | Year | No of lengthmeasured individuals (commercial catches) | No of aged individuals (commercial catches) | No of aged individuals (surveys) | Total no of aged individuals | Landings tonnes | Length-measured individuals per 1000 t | Aged individuals per 1000 t (commercial catches) | Total aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2020 | 1387 | 165 | 1853 | 2018 | 12266 | 113 | 13.5 | 164.5 | 125 |
| Capelin |  |  |  |  |  |  |  |  |  |  |
|  | 2008** | 82625 | 1644 | 2341 | 3985 | 5000 | 16525 | 328.8 | 797.0 | 125 |
|  | 2009 | 94541 | 900 | 2511 | 3411 | 73000 | 1295 | 12.3 | 46.7 | 125 |
|  | 2010 | 67265 | 1072 | 4043 | 5115 | 77000 | 874 | 13.9 | 66.4 | 125 |
|  | 2011 | 63784 | 1273 | 2271 | 3544 | 86531 | 737 | 14.7 | 41.0 | 125 |
|  | 2012 | 20023 | 1130 | 1783 | 2913 | 68182 | 294 | 16.6 | 42.7 | 125 |
|  | 2013 | 54708 | 1565 | 1007 | 2572 | 60413 | 906 | 25.9 | 42.6 | 125 |
|  | 2014 | 13206 | 850 | 1249 | 2099 | 25720 | 513 | 33.0 | 81.6 | 125 |
|  | 2015 | 27200 | 1000 | 1004 | 2004 | 115 |  |  |  | 125 |
|  | 2016 | 8669 | 3954 | 1047 | 5001 | 0 |  |  |  | 125 |
|  | 2017 |  |  | 4115 | 4115 | 6 |  |  |  | 125 |
|  | 2018 | 14491 | 250 | 1050 | 1300 | 65934 | 220 | 3.8 | 19.7 | 125 |
|  | 2019 |  |  | 1498 | 1498 | 34 |  |  |  | 125 |
|  | 2020 |  |  | 1245 | 1245 | 19 |  |  |  | 125 |

## *In addition also used long-term mean age-length keys.

${ }^{* *}$ Age samples from surveys with commercial trawl come in addition.
*** In addition used samples from Russian vessels, sampled by the Norwegian Coast Guard in 2012, 2015, 2019 and 2020.
Table 0.3. Age and length sampling by Spain ${ }^{5}$ of commercial catches and length sampling of surveys in 2008-2021. Also length-measured individuals and aged individuals per 1000 t caught. For comparison also the EU DCF requirements are shown.

| Stock | Year | No of vessels | No of lengthmeasured individuals (commercial catches) | No of aged individuals (commercial catches) | No of aged individuals (surveys) | Total no of aged individuals | Landings tonnes | Lengthmeasured individuals per 1000 t | Aged individuals per 1000 t (commercial catches) | Total aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |

NEA-cod

| 2008 | 2 | 10108 | 610 | 610 | 9658 | 1047 | 63 | 63 | 125 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 2 | 8733 | 1834 | 1834 | 12013 | 727 | 153 | 153 | 125 |
| 2010 | 2 | 28297 | 1735 | 1735 | 12657 | 2236 | 137 | 137 | 125 |
| 2011 | 2 | 11633 | 964 | 964 | 13291 | 875 | 73 | 73 | 125 |
| 2012 | 2 | 9849 | 998 | 998 | 12814 | 769 | 78 | 78 | 125 |
| 2013 | 2 | 30295 | 2381 | 2381 | 15041 | 2014 | 158 | 158 | 125 |
| 2014 | 2 | 27828 | 2306 | 2306 | 16479 | 1689 | 140 | 140 | 125 |
| 2015 | 2 | 18568 | 1445 | 1445 | 18772 | 989 | 77 | 77 | 125 |

${ }^{5}$ The onshore and the at-sea sampling programs coordinated by the IEO were suspended in most of 2020, due notably to administrative problems and to a lesser extend to COVID-19. This affected all stocks. Both sampling programmes are hired by IEO through call for tenders addressed to specialized companies. The public tender launched in 2019 (to start in 2020 ) was declared void, having to be re-launched again. This second launch was delayed as a result of the paralysis of public activity during the state of alarm due to the COVID-19 pandemic and could only be reopened in June-July. Given that the process of awarding the contract by public tender takes three-four months under normal conditions, it was finally resolved in December 2020 and signed in January 2021. Since then all activities have been resumed. The sampling to obtain the biological variables of the population (mainly reproduction and growth) is normally carried out in the IEO laboratories. This activity has also faced problems in 2020. On the one hand the administrative and financial difficulties of the IEO prevented the purchasing of samples in the market and on the other hand the three months closure of the labs ( 15 March to 21 June) due to COVID-19 did not allow for a normal activity.

| Stock | Year | No of vessels | No length- <br> measured in- <br> dividuals <br> (commercial <br> catches) | Noged <br> (commercial <br> catches) | No of aged <br> individuals <br> (surveys) | Total no of <br> aged individ- <br> uals | Landings <br> tonnes |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| Stock | Year | No of vessels | No of lengthmeasured individuals (commercial catches) | No of aged individuals (commercial catches) | No of aged individuals (surveys) | Total no of aged individuals | Landings tonnes | Lengthmeasured individuals per 1000 t | Aged individuals per 1000 t (commercial catches) | Total aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2019 |  |  |  |  |  | 280 |  |  |  |  |
|  | 2020 |  |  |  |  |  | 45 |  |  |  |  |
|  | 2021 |  |  |  |  |  | 131 |  |  |  |  |
| NEA-saithe |  |  |  |  |  |  |  |  |  |  |  |
|  | 2009 | 1 | 123 |  |  |  | 2 |  |  |  |  |
|  | 2013 | 1 |  |  |  |  | 5 |  |  |  |  |
|  | 2014 | 1 |  |  |  |  | 13 |  |  |  |  |
|  | 2015 | 1 |  |  |  |  | 33 |  |  |  |  |
|  | 2016 |  |  |  |  |  | 25 |  |  |  |  |
|  | 2017 |  |  |  |  |  | 85 |  |  |  |  |
|  | 2018 |  |  |  |  |  | 60 |  |  |  |  |
|  | 2019 |  |  |  |  |  | 199 |  |  |  |  |
|  | 2020 |  |  |  |  |  | 0 |  |  |  |  |
|  | 2021 |  |  |  |  |  | 3 |  |  |  |  |
| S. mentella |  |  |  |  |  |  |  |  |  |  |  |
|  | 2008** | 1 | 2275 | 28 |  |  | 987 | 2304 | 28 | 0 | 125 |
|  | 2011* | 1 | 86 |  |  |  | 1237 |  |  |  |  |


| Stock | Year | No of vessels | No of lengthmeasured individuals (commercial catches) | No of aged individuals (commercial catches) | No of aged individuals (surveys) | Total no of aged individuals | Landings tonnes | Lengthmeasured individuals per 1000 t | Aged individuals per 1000 t (commercial catches) | Total aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2012** | 2 | 11579 | 476 |  |  | 1612 | 7183 | 295 | 0 | 125 |
|  | 2014** | 1 | 6177 |  |  |  | 1146 | 5390 |  |  |  |
|  | 2015** | 1 | 6117 |  |  |  | 2371 | 2580 |  |  |  |
|  | 2016** | 1 | 11806 |  |  |  | 3133 | 3768 |  |  |  |
|  | 2017** | 1 | 5015 |  |  |  | 2624 | 1911 |  |  |  |
|  | 2018** | 1 | 11638 |  |  |  | 2399 | 4851 |  |  |  |
|  | 2019** | 1 | 11952 |  |  |  | 1908 | 6265 |  |  |  |
|  | 2020** |  |  |  |  |  | 737 |  |  |  |  |
|  | 2021** | 1 | 2074 | 157 |  |  | 280 | 7396 |  |  |  |
| Greenland halibut |  |  |  |  |  |  |  |  |  |  |  |
|  | 2008 | 2 | 11662 |  |  |  | 112 | 103826 |  |  |  |
|  | 2009 | 1 | 3383 |  |  |  | 210 | 16143 |  |  |  |
|  | 2010 | 1 | 5783 |  |  |  | 182 | 31800 |  |  |  |
|  | 2011 | 1 | 8541 |  |  |  | 169 | 50600 |  |  |  |
|  | 2012 | 1 | 4809 |  |  |  | 186 | 25907 |  |  |  |
|  | 2013 | 1 | 11988 |  |  |  | 190 | 63019 |  |  |  |
|  | 2014 | 1 | 12002 |  |  |  | 206 | 58262 |  |  |  |


| Stock | Year | No of vessels | No of lengthmeasured individuals (commercial catches) | No of aged individuals (commercial catches) | No of aged individuals (surveys) | Total no of aged individuals | Landings tonnes | Lengthmeasured individuals per 1000 t | Aged individuals per 1000 t (commercial catches) | Total aged individuals per 1000 t | EU DCF for comparison per 1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2015 | 1 | 17552 |  |  |  | 111 | 158126 |  |  |  |
|  | 2016 | 1 | 15031 |  |  |  | 218 | 68837 |  |  |  |
|  | 2017 |  |  |  |  |  |  |  |  |  |  |
|  | 2018 |  |  |  |  |  |  |  |  |  |  |
|  | 2019 | 1 |  |  |  |  | 49 |  |  |  |  |
|  | 2020 |  |  |  |  |  | 96 |  |  |  |  |
|  | 2021 |  |  |  |  |  | 125 |  |  |  |  |

*Sampling from bycatch in cod fishery.
**Sampling from pelagic redfish fishery.
***Sampling from Spanish Greenland halibut survey.

Table 0.4. Age and length sampling by Germany of commercial catches and age sampling of surveys in 2008-2021. Also length-measured individuals and aged individuals per 1000 t caught. For comparison also the EU DCF requirements are shown.

| Year | No of unique vessels | No of length samples | No of lengthmeasured individuals | No of aged individuals | Landings tonnes | Length-measured individuals per 1000 t | Age-sampled individuals per 1000 t | EU DCF for comparison |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| NEA cod |  |  |  |  |  |  |  |  |
| 2008 | 5 | 3 | 65800 | 2033 | 4955 | 13280 | 410 | 125 |
| 2009 | 5 | 2 | 43107 | 2419 | 8585 | 5021 | 282 | 125 |
| 2010 | 5 | 2 | 51923 | 3075 | 8442 | 6151 | 364 | 125 |


| Year | No of unique vessels | No of length samples | No of lengthmeasured individuals | No of aged individuals | Landings tonnes | Length-measured individuals per 1000 t | Age-sampled individuals per 1000 t | EU DCF for comparison |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 4 | 1 | 7318 | 769 | 4621 | 1584 | 166 | 125 |
| 2012 | 4 | 2 | 16315 | 1924 | 8500 | 1919 | 226 | 125 |
| 2013 | 4 | 2 | 29281 | 2043 | 7939 | 3688 | 257 | 125 |
| 2014 | 4 | 1 | 23137 | 1291 | 6225 | 3717 | 207 | 125 |
| 2015 | 4 | 1 | 39335 | 886 | 6427 | 6120 | 138 | 125 |
| 2016 | 3 | 1 | 22109 | 1060 | 6636 | 3332 | 160 | 125 |
| 2017 | 4 | 1 | 19942 | 785 | 5969 | 3341 | 132 | 125 |
| 2018 | 4 | 2 | 43371 | 2283 | 7774 | 5579 | 294 | 125 |
| 2019 | 2 | 1 | 17954 | 1444 | 8535 | 2104 | 169 | 125 |
| 2020 | 2 | 1 | 21716 | 1021 | 9786 | 2219 | 104 | 125 |
| 2021 | 2 | 1 | 21548 | 1393 | 5470 | 3939 | 255 | 125 |
| NEA haddock |  |  |  |  |  |  |  |  |
| 2008 | 5 | 3 | 5548 | 442 | 535 | 10370 | 826 | 125 |
| 2009 | 5 | 2 | 23348 | 958 | 1957 | 11931 | 490 | 125 |
| 2010 | 5 | 2 | 54704 | 1039 | 3539 | 15457 | 294 | 125 |
| 2011 | 4 | 1 | 1925 | 160 | 1724 | 1117 | 93 | 125 |
| 2012 | 4 | 2 | 4088 | 502 | 1111 | 3680 | 452 | 125 |
| 2013 | 4 | 1 | 7040 | 478 | 501 | 14052 | 954 | 125 |


|  | Year | No of unique vessels | No of length samples | No of lengthmeasured individuals | No of aged individuals | Landings tonnes | Length-measured individuals per 1000 t | Age-sampled individuals per 1000 t | EU DCF for comparison |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2014 | 4 | 1 | 3113 | 261 | 340 | 9156 | 768 | 125 |
|  | 2015 | 4 | 1 | 616 | 325 | 124 | 4968 | 2621 | 125 |
|  | 2016 | 3 | 1 | 4807 | 544 | 170 | 28276 | 3200 | 125 |
|  | 2017 | 4 | 1 | 3464 | 527 | 155 | 22348 | 3400 | 125 |
|  | 2018 | 4 | 2 | 4345 | 497 | 391 | 11113 | 1271 | 125 |
|  | 2019 | 2 | 1 | 5031 | 393 | 208 | 24188 | 1889 | 125 |
|  | 2020 | 2 | 1 | 2979 | 356 | 283 | 10527 | 1258 | 125 |
|  | 2021 | 2 | 1 | 2808 | 344 | 368 | 7630 | 935 | 125 |
| NEA saithe |  |  |  |  |  |  |  |  |  |
|  | 2008 | 5 | 3 | 10210 | 605 | 2263 | 4512 | 267 | 125 |
|  | 2009 | 6 | 2 | 8667 | 1091 | 2021 | 4288 | 540 | 125 |
|  | 2010 | 7 | 2 | 11424 | 1001 | 1592 | 7176 | 629 | 125 |
|  | 2011 | 4 | 1 | 4863 | 530 | 1371 | 3547 | 387 | 125 |
|  | 2012 | 7 | 2 | 14193 | 1202 | 1371 | 10356 | 877 | 125 |
|  | 2013 | 4 | 1 | 1190 | 414 | 1212 | 982 | 342 | 125 |
|  | 2014 | 3 | 1 | 25 | 0 | 259 | 97 | 0 | 125 |
|  | 2015 | 4 | 0 | 0 | 0 | 424 | 0 | 0 | 125 |
|  | 2016 | 3 | 1 | 13981 | 909 | 951 | 14701 | 956 | 125 |


|  | Year | No of unique vessels | No of length samples | No of lengthmeasured individuals | No of aged individuals | Landings tonnes | Length-measured individuals per 1000 t | Age-sampled individuals per 1000 t | EU DCF for comparison |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2017 | 4 | 1 | 15734 | 603 | 1154 | 13634 | 523 | 125 |
|  | 2018 | 4 | 1 | 19718 | 473 | 1651 | 11943 | 286 | 125 |
|  | 2019 | 2 | 1 | 9465 | 1521 | 1387 | 6824 | 1097 | 125 |
|  | 2020 | 2 | 1 | 11900 | 745 | 1573 | 7565 | 474 | 125 |
|  | 2021 | 2 | 1 | 3707 | 784 | 597 | 6209 | 1313 | 125 |
| Redfish |  |  |  |  |  |  |  |  |  |
|  | 2008 | 5 | 3 | 330 | 0 | 46 | 7174 | 0 | 125 |
|  | 2009 | 8 | 2 | 0 | 0 | 100 | 0 | 0 | 125 |
|  | 2010 | 6 | 2 | 0 | 0 | 52 | 0 | 0 | 125 |
|  | 2011 | 6 | 1 | 7937 | 0 | 844 | 9404 | 0 | 125 |
|  | 2012 | 9 | 2 | 4036 | 0 | 584 | 6911 | 0 | 125 |
|  | 2013 | 4 | 1 | 1315 | 0 | 81 | 16235 | 0 | 125 |
|  | 2014 | 4 | 1 | 571 | 0 | 451 | 1266 | 0 | 125 |
|  | 2015 | 4 | 1 | 76 | 0 | 266 | 286 | 0 | 125 |
|  | 2016 | 3 | 1 | 6095 | 0 | 497 | 12264 | 0 | 125 |
|  | 2017 | 4 | 1 | 977 | 0 | 770 | 1269 | 0 | 125 |
|  | 2018 | 4 | 2 | 3438 | 0 | 2508 | 1371 | 0 | 125 |
|  | 2019 | 2 | 1 | 8958 | 0 | 1741 | 5145 | 0 | 125 |


| Year | No of unique vessels | No of length samples | No of lengthmeasured individuals | No of aged individuals | Landings tonnes | Length-measured individuals per 1000 t | Age-sampled individuals per 1000 t | EU DCF for comparison |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2020 | 3 | 1 | 4248 | 0 | 1998 | 2126 | 0 | 125 |
| 2021 | 2 | 1 | 2261 | 0 | 743 | 3043 | 0 | 125 |
| Greenland halibut |  |  |  |  |  |  |  |  |
| 2008 | 5 | 2 | 0 | 0 | 5 | 0 | 0 | 125 |
| 2009 | 3 | 2 | 0 | 0 | 19 | 0 | 0 | 125 |
| 2010 | 2 | 2 | 0 | 0 | 14 | 0 | 0 | 125 |
| 2011 | 3 | 1 | 0 | 0 | 81 | 0 | 0 | 125 |
| 2012 | 4 | 2 | 0 | 0 | 40 | 0 | 0 | 125 |
| 2013 | 3 | 1 | 1298 | 0 | 49 | 26544 | 0 | 125 |
| 2014 | 4 | 1 | 1076 | 0 | 34 | 31647 | 0 | 125 |
| 2015 | 4 | 1 | 658 | 0 | 32 | 20563 | 0 | 125 |
| 2016 | 3 | 1 | 365 | 0 | 9 | 40556 | 0 | 125 |
| 2017 | 4 | 1 | 0 | 0 | 21 | 0 | 0 | 125 |
| 2018 | 4 | 1 | 257 | 0 | 52 | 4942 | 0 | 125 |
| 2019 | 2 | 1 | 511 | 0 | 45 | 11356 | 0 | 125 |
| 2020 | 2 | 1 | 305 | 0 | 74 | 4122 | 0 | 125 |
| 2021 | 2 | 1 | 160 | 0 | 72 | 2222 | 0 | 125 |



Figure 0.3. Proportion of swept-area biomass in the Joint winter survey found in the new northern area ( N ), by year and species. For 2020 the indices for redfish and Greenland halibut have not yet been calculated.


Figure 0.4. Barents Sea Ecosystem survey (BESS) 2019, realized vessel tracks with pelagic and bottom trawl sampling stations.

### 1.11 Ecosystem information

The aim of this section is to collect important ecosystem information influencing the assessment of fish stocks handled by AFWG. In general, such information is collected and updated by the ICES WGIBAR group, here we only provide information that is directly relevant to the assessment of the AFWG stocks as well as information that is updated after the 2021 WGIBAR report was finished.

### 1.11.1 0-group abundance

The recruitment of the Barents Sea fish species measured as 0-group has shown a large year-toyear variability. The most important reasons for this variability are variations in the spawning biomass, hydrographic conditions, changes in circulation pattern, food availability and predator abundance, and distribution. In 2018 and 2020, 0-group indices were strongly affected by incomplete area coverage in the Barents Sea, but attempts have been made to correct for this (Prozorkevitch and Van der Meeren, 2021).

### 1.11.2 Consumption, natural mortality, and growth

Cod is the most important predator among fish species in the Barents Sea. It feeds on a wide range of prey, including larger zooplankton, most available fish species, including own juveniles and shrimp (Tables 1.1-1.2). Cod prefer capelin as a prey, and fluctuations of the capelin stock may have a strong effect on growth, maturation, and fecundity of cod, as well as on cod recruitment because of cannibalism. The role of euphausiids for cod feeding increases in the years when capelin stock is at a low level (Ponomarenko and Yaragina, 1990). Also, according to Ponomarenko (1973; 1984), interannual changes of euphausiid abundance are important for the survival rate of cod during the first year of life.

The food consumption by NEA cod in 1984-2020, based on data from the Joint Russian-Norwegian stomach content database, is presented in Tables 1.1-1.2. The Norwegian (IMR) calculations are based on the method described by Bogstad and Mehl (1997). The main prey items in 2020 were capelin (about 2 million tonnes), followed by krill, amphipods and polar cod of which the consumption was about 500 thousand tonnes of each category. Shrimp, long rough dab, cod, herring, haddock and snow crab were all less important (between 90 and 180 thousand tonnes for each species). The increase in consumption of polar cod from 2019 to 2020 is consistent with the markedly increased abundance of this species. The decrease in consumption of young cod and haddock is consistent with the low abundance of age 0 and 1 of these species in 2020 . The consumption calculations made by The consumption per cod by cod age-groups are shown in Tables 1.3-1.4 (IMR and PINRO estimates), while the proportion of cod and haddock in the diet by cod age-group (IMR estimates) is given in Tables 1.5 and Table 1.6. IMR show that the total consumption by age 1 and older cod in 2020 was 5.2 million tonnes. For technical reasons, PINRO estimates (Table 1.2 and 1.4) were not updated this year.

Growth of cod as calculated from weight at age in the winter survey has shown a declining trend in the last years, but this decline has now been halted, and for age 6 and older the trend seems to have been reversed. However, weight at age 3 and 4 was the lowest in this survey series from 1994-present, and for ages 3 and 6-8 it was among the three lowest values in the same period. The trends in consumption per cod by age group in recent years seem consistent with the trends in size at age.

Weight at age in the Lofoten survey was stable from 2019 to 2021, while weight-at-age in catch of cod decreased slightly for ages 3-9 from 2018-2020.

How is the outlook for cod food abundance in 2021? Total abundance of pelagic fish stocks is at an average level, for the most important pelagic species, capelin, the abundance of immature capelin in 2020 was intermediate due to a very strong 2019 year class (the strongest since 2000). Polar cod abundance in 2020 was close to the highest value observed in the 35 -year time-series due to the 2019 year class being the strongest ever observed. However, the herring abundance in the Barents Sea is now low as the strong 2016 year class has left the Barents Sea and the following year classes, which still are found in the Barents Sea, are weak. Also, age 1-2 cod and haddock abundance in 2021 is low. On the positive side, shrimp abundance is high, while the abundance of other prey species is around average. Altogether there seems to be reasonable consistency between growth, consumption and feeding data.
One direct application for the management of results from the trophic investigations in the Barents Sea is the inclusion of predator's consumption into fish stock assessment. Predation on cod and haddock by cod has since 1995 been included in the assessment of these two species. These data, summarized in Tables 1.1, 1.3 and 1.5, are used for estimation of cod and haddock consumed by cod and further for estimation of their natural mortality within the SAM model (see sections 3.3.3 and 4.5.5). The average natural mortality for the last years is used as predicted M for the coming years for cod and haddock.

Cod consumption was used in capelin assessment for the first time in 1990, to account for natural mortality due to cod predation on mature capelin in the period January-March (Bogstad and Gjøsæter, 1994). This methodology has been developed further using the Bifrost and CapTool models (Gjøsæter et al., 2002; Tjelmeland, 2005; ICES CM 2009/ACOM:34). CapTool is a tool (in Excel with @RISK) for implementing results from Bifrost in the short term (half-year) prognosis used for determining the quota.

In recent years the abundance of large cod and haddock has been very high, and it is still at a high level for cod. There are a limited number of predators on such large fish. As predation is likely to be a major source of natural mortality, it could thus be considered whether the natural mortality in older age groups should be reduced in such a situation. The assumption of reduced natural mortality on older cod was explored by IBPCOD 2017, but no evidence of this was found based on available catch and survey data. To investigate this further, analyses on predator consumption and biomass flow at higher trophic levels like those done by Bogstad et al. (2000) should be updated, and such work is ongoing for marine mammals. For cod, in particular, the fishing mortality since 2008 has been so much lower than before that the relative impact of the natural mortality on the survival of older fish has increased considerably.

The amount of commercially important prey consumed by other fish predators (haddock, Greenland halibut, long rough dab, and thorny skate), has also been calculated (Dolgov et al., 2007), but these consumption estimates have not been used in assessment for any prey stocks yet. Marine mammals are not included in the current fish stock assessments. However, it has been attempted to extend the stock assessment models of Barents Sea capelin (Bifrost) by including the predatory effects of minke whales, and harp seals (Tjelmeland and Lindstrøm, 2005).

### 1.11.3 Maturation, condition factor, and fisheries-induced evolution

Data on maturity-at-age are one of the basic components for spawning-stock biomass (SSB) estimates. There have been substantial changes observed in maturity-at-age of NEA cod over a large historical period (since 1946) showing an acceleration in maturity rates, especially in the 1980s. They are thought to be connected both with compensatory density-dependence mechanisms and genetic changes in individuals (Heino et al., 2002; Jørgensen et al., 2008; Kovalev and Yaragina, 2009; Eikeset et al., 2013; Kuparinen et al., 2014) resulted from strong fishing pressure.

Studies on possible evolutionary effects for this stock should be updated with data for recent years to investigate the effects on population dynamics, including growth, maturation, and evolutionary effects, of a prolonged period with low fishing mortality and high stock size.

Recent laboratory and fieldwork have shown that skipped spawning does occur in NEA cod stock (Skjæraasen et al., 2009; Yaragina, 2010). Experimental work on captive fish has demonstrated that skipped spawning is strongly influenced by individual energy reserves (Skjæraasen et al., 2009). This is supported by the field data, which suggest that gamete development could be interrupted by a poor liver condition especially. Fish that will skip spawning seem to remain in the Barents Sea and do not migrate to the spawning grounds. These fish need to be identified and excluded when estimating the stock-recruitment potential as currently they are included in the estimate of SSB. However, more work needs to be undertaken to improve our knowledge of skipped spawning in cod (e.g. comparisons and intercalibration of Norwegian and Russian databases on maturity stages should be done) and other species in order to quantify its influence on the stock reproductive potential.

### 1.11.4 Recruitment prediction for northeast Arctic cod

Prediction of recruitment in fish stocks is essential to harvest prognosis. Traditionally, prediction methods have been based on spawning-stock biomass and survey indices of juvenile fish and have not included effects of ecosystem drivers. Multiple linear regression models can be used to incorporate both environmental and parental fish stock parameters. In order for such models to give predictions, there need to be a time-lag between the predictor and response variables. In this section, a model for Northeast Arctic cod which is in use in assessment is presented. Note that a recruitment model for Barents Sea capelin with similar features also was presented to the group (WD 13).

### 1.11.5 Historic overview

Several statistical models, which use multiple linear regressions, have been developed for the recruitment of northeast Arctic cod. All models try to predict recruitment-at-age 3 (at 1 January), as calculated from the assessment model, with cannibalism included. This quantity is denoted as R3. A collection of the most relevant models previously presented to AFWG is described below.

Stiansen et al. (2005) developed a model (JES1) with 2-year prediction possibility:

$$
\begin{aligned}
& \text { JES1: R3~ Temp(-3) + Age1(-2) + MatBio(-2) } \\
& \text { JES2: R3~ Temp(-3) + Age2(-1) + MatBio(-2) } \\
& \text { JES3: R3~ Temp(-3) + Age3(0) + MatBio(-2) }
\end{aligned}
$$

Temp is the Kola annual temperature ( $0-200$ m, station 3-7), Age1 is the winter survey bottom trawl index for cod age 1, and MatBio the maturing biomass of capelin on 1 October. The number in parentheses is the time-lag in years. Two other similar models (JES2, JES3) can be made by substituting the winter index term Age1(-2) with Age2(-1) and Age3(0), giving 1 and 0 -year predictions, respectively.
Svendsen et al. (2007) used a model (SV) based only on data from the ROMS numerical hydrodynamical model, with 3-year prognosis possibility:
SV: R3~ Phyto(-3) + Inflow(-3)

Where Phyto is the modelled phytoplankton production in the whole Barents Sea and Inflow is the modelled inflow through the western entrance to the Barents Sea in autumn. The number in parentheses is the time-lag in years. The model has not been updated since 2007.

The recruitment model (TB) suggested by T. Bulgakova (AFWG 2005, WD14) is a modification of Ricker's model for stock-recruitment defined by:
TB: R3~ m(-3) exp[-SSB(-3) + N(-3)]

Where R3 is the number of age 3 recruits for NEA cod, $m$ is an index of population fecundity, SSB is the spawning-stock biomass and N is equal to the number of months with positive temperature anomalies (TA) on the Kola Section in the birth year for the year class. The number in parentheses is the time-lag in years. For the years before 1998 TA was calculated relative to monthly average for the period 1951-2000. For intervals after 1998, the TA was calculated with relatively linear trend in the temperature for the period 1998-present. The model was run using two-time intervals (using cod year classes 1984-2000 and year classes 1984-2004) for estimating the model coefficients. The models have not been updated since 2009.

Titov (Titov, AFWG 2010, WD 22) and Titov et al. (AFWG 2005, WD 16) developed models with 1 to 4-year prediction possibility (TITOV0, TITOV1, TITOV2, TITOV3, TITOV4, respectively), based on the oxygen saturation at bottom layers of the Kola section stations 3-7 (OxSat), air temperature at the Murmansk station (Ta), water temperature: 3-7 stations of the Kola section (layer $0-200 \mathrm{~m}$; Tw), ice coverage in the Barents Sea (I), spawning-stock biomass (SSB), annual values of 0 -group cod abundance index, corrected for capture efficiency (CodC0) and the bottom-trawl swept-area abundance of cod at the age 1 and 2, 3 derived from the joint winter Barents Sea acoustic survey (CodB1, CodB2, CodB3). At the 2010 AFWG assessment it was suggested (Dingsør et al., 2010, WD 19, and related discussions in the working group to try to simplify these models).

Hjermann et al., (2007) developed a model with a one-year prognosis, which has been modified by Dingsør et al. (AFWG 2010, WD19) to four models with 2-year projection possibility.

$$
\begin{aligned}
& \text { H1: } \log (\text { R3 }) \sim \text { Temp }(-3)+\log (\text { Age0 })(-3)+\text { BM cod3-6 } / \text { ABM }_{\text {capelin }}(-2,-1) \\
& \text { H2: } \log (\text { R3 }) \sim \text { Temp(-2) }+ \text { I(surv })+ \text { Age1(-2) }+ \text { BM }_{\text {cod3-6 }} / \text { ABMcapelin }(-2,-1) \\
& \text { H3: } \log (\text { R3 }) \sim \text { Temp(-1) }+ \text { Age2(-1) }+ \text { BM }_{\text {cod3-6 }} / \text { ABM }_{\text {capelin }}(-1) \\
& \text { H4: } \log (\text { R3 }) \sim \text { Temp(-1) }+ \text { Age3(0) }
\end{aligned}
$$

Temp is the Kola yearly temperature ( $0-200 \mathrm{~m}$ ), Age0 is the 0 -group index of cod, Age1, Age2 and Age3 are the winter survey bottom trawl index for cod age 1, 2 and 3, respectively, $\mathrm{BM}_{\text {cod } 3-6}$ is the biomass of cod between age 3 and 6 , and ABM is the maturing biomass of capelin. The number in parentheses is the time-lag in years. The models were not updated this year.

At AFWG 2008, Subbey et al. presented a comparative study (AFWG 2008, WD27) on the ability of some of the above models in predicting stock-recruitment for NEA cod (Age 3). At the assessment in 2010, a WD by Dingsør et al. (AFWG 2010, WD19) was presented, which investigated the performance of some of the mentioned recruitment models. It was strongly recommended by the working group that a Study Group should be appointed to look at criteria for choosing/rejecting recruitment models suitable for use in stock assessment.

The "Study Group on Recruitment Forecasting" (SGRF; ICES CM 2011/ACOM:31, ICES CM 2012/ACOM:24, ICES CM 2013/ACOM:24) have had three meetings (in October 2011 and 2012, and November 2013). Their mandate is to give a "best practice" (Standards and guidelines) for choosing recruitment models after their next meeting, which may be implemented at the next AFWG.

The SGRF 2012 report addressed the problem of combining several model predictions to obtain a recruitment estimate with minimum variance. The method (involving a weighted average of individual model predictions) was proposed as a replacement for the hybrid method of Subbey et al. (2008). One major issue not addressed in ICES SGRF (2012) was how to choose the initial ensemble of models, whose weighted average is sought. There are practical constraints (with respect to time and personnel), which stipulates that not all plausible models can be included in the calculation of the hybrid recruitment value. A methodology for choosing models to include in the calculation of a hybrid, representative recruitment forecast was addressed in SGRF 2013. Details can be found in the SGRF 2013 ICES report.

### 1.11.6 Models used in 2021

The model approach taken in 2021 was the same as in 2018-2020. Some changes were made in 2018, they are described below.

In 2018 at the meeting of the AFWG, the correction and simplification of models were continued. Since in 2017-2018 there was a significant correction of the initial biological data, which caused significant changes in the results of the prognostic models, in 2018 a complete audit of both prognostic models and the hybrid model combining the results of their work was carried out. The main purpose of the model revision was to increase the stability of the models, that is, to reduce the possibility of potential correction of the models due to correction of the biological data included in the model. The solution to the problem was found by increasing the retrospective database backwards in time, that is, from the beginning of the 1980s to the beginning of the 1960s. Accordingly, sets of predictor sets have been revised. The number of models was reduced from 5 to 2 and the names of the models were changed from Titov $0(1,2,3,4)$ to TitovES (environment, short prediction) and TitovEL (environment, long prediction).
This has been conducted and has improved the statistical performance (details are shown in Titov, AFWG 2018, WD23):

TitovES: R32 ~ DOxSat2(t-13) + ITw(t-43) + expIce(t-40) + Ice(t-15)
TitovEL: R34~OxSat(t-39)+ITw(t-43)
Where DOxSat $(\mathrm{t}-13) \sim \operatorname{expOxSat}(\mathrm{t}-13)+\mathrm{OxSat}(\mathrm{t}-39), \operatorname{ITw}(\mathrm{t}-43) \sim \mathrm{I}(\mathrm{t}-43)+\mathrm{Tw}(\mathrm{t}-46)$. The number in parentheses is the time-lag in months, relative to April in the year when the prediction is carried out.

At the 2018 AFWG assessment, a hybrid model (i.e. an average combination) of the best functioning statistical recruitment models were repeated. A statistical analysis of the accuracy of the model's work was carried out, which consisted in estimating the errors in the recovery of data on the number of NEA cod recruitment. Accuracy of the model's work was verified by calculation of standard deviations of the NEA cod recruitment predicted values from the SAM values for the period 2005-2015 when the model was adjusted for data from 1983 to 2004, which consisted in estimating the errors in the recovery of data on the number of NEA cod recruitment.

Figure 1.1 shows the standard deviations of the NEA cod recruitment prediction. The addition of biological parameters (CodB1, CodB2, CodB3, CodC0, SSB) to environmental models (TitovES, TitovEL) substantially increases the error.

Based on these calculations, after comparing the results of constructing independent retrospective forecasts using the methodology previously used in ICES SGRF (ICES CM 2013/ACOM:24), it was decided to abandon the use of biological predictors and to use only environmental data in the NEA cod recruitment forecasting models. It was also found that all models (TitovES, TitovEL, RCT3) satisfy the quality conditions with respect to the forecast for the mean values accepted as the criterion for entering into the calculation of the hybrid model adopted earlier (ICES CM

2013/ACOM:24). It was decided that all biological data will be included in calculations based on the RCT3 model, and the remaining two models (TitovES, TitovEL) will be used only to account for the effect of environmental conditions on NEA cod recruitment.

In AFWG 2021 the procedure for estimating weights for various models (TitovES, TitovEL, RCT3) was repeated using the same method as was made on Study Group on Recruitment Forecasting (SGRF) in 2013. The input data for the models are given below in Tables 1.7 (TitovES, TitovEL) and 1.8 (RCT3).

In summary, the SAM estimate for age 3 from the AFWG 2021 assessment was used as historical R3. The recruitment forecast for 2021-2024 are based on a hybrid model with weighting estimated at AFWG 2021. The weights and forecasts for the 2021 AFWG assessment can be found in Table 1.9.

It was noted that the oceanographic dataset for the Titov ES and EL models cover the year classes from 1959 onwards, while the survey data used in the RCT3 model only cover the year classes from 1991 onwards, although those survey dataseries started in 1981. Further, the area covered in the surveys was extended in 2014, which is accounted for in the cod assessment by splitting the bottom trawl survey series in that year, while no such split was made in the RCT3 model. It should be investigated how this area expansion in the survey best could be accounted for in the recruitment model.

New software in R was presented during AFWG 2021 for predicting cod recruitment using the hybrid model (WD 20) including the automatic procedure for the submodel's weight estimation. A comparison of predicted values with "old" software (WD 21) was done and the results were identical.

Table 1.1. The North-east arctic COD stock's consumption of various prey species in 1984-2020 (1000 tonnes) based on Norwegian consumption calculations

| Year | Other | Amphipods | Krill | Shrimp | Capelin | Herring | Polar cod | Cod | Haddock | Redfish | G. halibut | Blue whiting | Long rough c | Snow crab | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 494 | 27 | 119 | 447 | 739 | 82 | 16 | 23 | 52 | 374 | 0 | 0 | 25 | 0 | 2398 |
| 1985 | 1252 | 188 | 64 | 179 | 1780 | 214 | 3 | 31 | 54 | 244 | 0 | 2 | 48 | 0 | 4058 |
| 1986 | 679 | 1426 | 133 | 165 | 961 | 162 | 156 | 74 | 110 | 340 | 0 | 0 | 66 | 0 | 4273 |
| 1987 | 813 | 1372 | 89 | 233 | 295 | 38 | 225 | 26 | 6 | 340 | 1 | 0 | 11 | 0 | 3449 |
| 1988 | 447 | 1419 | 337 | 151 | 382 | 8 | 99 | 11 | 2 | 259 | 0 | 5 | 6 | 0 | 3126 |
| 1989 | 679 | 823 | 245 | 123 | 589 | 3 | 37 | 8 | 10 | 222 | 0 | 0 | 67 | 0 | 2805 |
| 1990 | 1149 | 123 | 80 | 162 | 1409 | 7 | 5 | 16 | 14 | 188 | 0 | 81 | 86 | 0 | 3320 |
| 1991 | 688 | 63 | 71 | 164 | 2441 | 7 | 10 | 22 | 16 | 264 | 7 | 8 | 240 | 0 | 4002 |
| 1992 | 826 | 97 | 154 | 354 | 2266 | 275 | 92 | 46 | 88 | 172 | 23 | 2 | 94 | 0 | 4487 |
| 1993 | 709 | 242 | 669 | 305 | 2873 | 155 | 269 | 261 | 69 | 92 | 2 | 2 | 27 | 0 | 5674 |
| 1994 | 611 | 552 | 693 | 506 | 1060 | 146 | 599 | 223 | 48 | 76 | 0 | 1 | 43 | 0 | 4558 |
| 1995 | 827 | 972 | 527 | 358 | 607 | 117 | 245 | 367 | 114 | 194 | 2 | 0 | 36 | 0 | 4366 |
| 1996 | 604 | 620 | 1166 | 345 | 548 | 46 | 101 | 536 | 67 | 95 | 0 | 10 | 37 | 0 | 4173 |
| 1997 | 466 | 404 | 545 | 350 | 978 | 5 | 115 | 350 | 44 | 33 | 0 | 34 | 15 | 0 | 3340 |
| 1998 | 448 | 411 | 513 | 375 | 836 | 104 | 174 | 163 | 36 | 9 | 0 | 14 | 18 | 0 | 3100 |
| 1999 | 422 | 166 | 306 | 300 | 2047 | 151 | 258 | 67 | 30 | 18 | 1 | 35 | 9 | 0 | 3808 |
| 2000 | 427 | 188 | 492 | 503 | 1935 | 61 | 218 | 83 | 58 | 8 | 0 | 41 | 21 | 0 | 4035 |
| 2001 | 721 | 176 | 382 | 291 | 1836 | 76 | 264 | 68 | 51 | 6 | 1 | 157 | 32 | 0 | 4060 |
| 2002 | 376 | 96 | 260 | 241 | 2004 | 86 | 280 | 108 | 127 | 1 | 0 | 239 | 16 | 0 | 3834 |
| 2003 | 545 | 285 | 545 | 238 | 2152 | 216 | 275 | 110 | 166 | 3 | 0 | 74 | 53 | 0 | 4662 |
| 2004 | 626 | 560 | 347 | 246 | 1253 | 216 | 358 | 126 | 198 | 3 | 11 | 56 | 65 | 1 | 4065 |
| 2005 | 781 | 579 | 527 | 274 | 1399 | 132 | 388 | 118 | 324 | 2 | 5 | 115 | 53 | 0 | 4697 |
| 2006 | 870 | 225 | 1078 | 353 | 1737 | 170 | 108 | 80 | 361 | 12 | 2 | 163 | 130 | 0 | 5287 |
| 2007 | 1259 | 310 | 1091 | 428 | 2140 | 285 | 266 | 88 | 378 | 46 | 0 | 44 | 75 | 0 | 6411 |
| 2008 | 1578 | 160 | 931 | 385 | 2865 | 105 | 514 | 187 | 293 | 59 | 13 | 18 | 93 | 0 | 7201 |
| 2009 | 1495 | 243 | 635 | 265 | 3978 | 123 | 730 | 196 | 252 | 28 | 3 | 5 | 115 | 2 | 8072 |
| 2010 | 1616 | 415 | 1049 | 281 | 3900 | 52 | 334 | 241 | 267 | 142 | 10 | 14 | 133 | 7 | 8462 |
| 2011 | 1556 | 254 | 902 | 221 | 4120 | 84 | 424 | 286 | 279 | 115 | 0 | 26 | 122 | 9 | 8398 |
| 2012 | 1975 | 316 | 842 | 345 | 3641 | 51 | 519 | 373 | 220 | 51 | 34 | 8 | 125 | 7 | 8506 |
| 2013 | 1774 | 261 | 566 | 267 | 3660 | 51 | 137 | 380 | 200 | 111 | 1 | 21 | 167 | 15 | 7612 |
| 2014 | 1409 | 326 | 475 | 202 | 3713 | 72 | 31 | 358 | 88 | 31 | 11 | 18 | 106 | 9 | 6849 |
| 2015 | 1595 | 619 | 637 | 243 | 3278 | 126 | 147 | 213 | 178 | 140 | 43 | 59 | 85 | 33 | 7396 |
| 2016 | 1691 | 530 | 745 | 299 | 2210 | 95 | 346 | 198 | 222 | 57 | 6 | 87 | 120 | 10 | 6617 |
| 2017 | 1053 | 126 | 582 | 251 | 2950 | 193 | 88 | 315 | 272 | 45 | 4 | 24 | 139 | 53 | 6097 |
| 2018 | 1032 | 267 | 644 | 180 | 2886 | 203 | 246 | 246 | 276 | 34 | 70 | 47 | 52 | 44 | 6227 |
| 2019 | 779 | 212 | 415 | 308 | 2600 | 181 | 168 | 188 | 212 | 44 | 0 | 2 | 99 | 50 | 5258 |
| 2020 | 919 | 523 | 535 | 172 | 2021 | 107 | 467 | 115 | 92 | 30 | 14 | 13 | 150 | 90 | 5247 |


|  | NOT UPDA | ATED THIS YE | EAR |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Other | Amphipods | Krill | Shrimp | Capelin | Herring | Polar cod | Cod | Haddock | Redfish | G. halibut | Blue whitin! | Long rough | Snow crab | Total |
| 1984 | 560 | 31 | 94 | 353 | 593 | 34 | 18 | 14 | 50 | 197 | 0 | 5 | 52 |  | 2000 |
| 1985 | 767 | 441 | 31 | 211 | 1041 | 26 | 0 | 89 | 36 | 100 | 0 | 18 | 22 |  | 2779 |
| 1986 | 615 | 949 | 66 | 159 | 855 | 51 | 169 | 26 | 99 | 166 | 1 | 3 | 26 |  | 3186 |
| 1987 | 541 | 593 | 79 | 233 | 175 | 9 | 118 | 23 | 2 | 119 | 1 | 10 | 5 |  | 1908 |
| 1988 | 544 | 196 | 239 | 146 | 348 | 21 | 0 | 21 | 76 | 133 | 0 | 0 | 22 |  | 1745 |
| 1989 | 496 | 324 | 190 | 117 | 767 | 4 | 37 | 35 | 2 | 178 | 0 | 0 | 64 |  | 2213 |
| 1990 | 278 | 31 | 105 | 266 | 1264 | 65 | 8 | 24 | 15 | 237 | 0 | 39 | 79 |  | 2409 |
| 1991 | 289 | 81 | 55 | 277 | 3204 | 25 | 45 | 52 | 22 | 141 | 5 | 6 | 46 |  | 4248 |
| 1992 | 788 | 38 | 211 | 258 | 2021 | 335 | 196 | 82 | 37 | 117 | 1 | 0 | 42 |  | 4125 |
| 1993 | 563 | 174 | 184 | 220 | 2743 | 170 | 170 | 144 | 148 | 40 | 5 | 4 | 47 |  | 4611 |
| 1994 | 447 | 296 | 359 | 458 | 1276 | 102 | 486 | 383 | 72 | 55 | 0 | 1 | 40 |  | 3976 |
| 1995 | 502 | 455 | 396 | 533 | 670 | 192 | 191 | 541 | 130 | 110 | 3 | 0 | 52 |  | 3775 |
| 1996 | 674 | 346 | 957 | 195 | 469 | 74 | 74 | 451 | 57 | 67 | 0 | 9 | 45 |  | 3415 |
| 1997 | 463 | 134 | 510 | 257 | 511 | 52 | 111 | 383 | 35 | 29 | 2 | 17 | 17 |  | 2520 |
| 1998 | 311 | 220 | 645 | 286 | 916 | 73 | 134 | 131 | 23 | 15 | 0 | 24 | 20 |  | 2797 |
| 1999 | 179 | 81 | 458 | 268 | 1540 | 80 | 177 | 49 | 16 | 14 | 0 | 27 | 9 |  | 2898 |
| 2000 | 243 | 122 | 437 | 394 | 1800 | 53 | 167 | 59 | 32 | 4 | 0 | 28 | 21 |  | 3360 |
| 2001 | 384 | 75 | 411 | 322 | 1522 | 93 | 148 | 62 | 52 | 4 | 2 | 145 | 31 |  | 3250 |
| 2002 | 225 | 45 | 286 | 202 | 2400 | 55 | 302 | 100 | 80 | 4 | 0 | 110 | 17 |  | 3825 |
| 2003 | 400 | 171 | 547 | 227 | 1219 | 153 | 221 | 132 | 331 | 2 | 0 | 28 | 51 |  | 3481 |
| 2004 | 496 | 393 | 478 | 256 | 1097 | 129 | 369 | 86 | 144 | 7 | 16 | 48 | 62 |  | 3583 |
| 2005 | 620 | 163 | 688 | 244 | 1023 | 168 | 320 | 112 | 271 | 7 | 2 | 67 | 47 |  | 3731 |
| 2006 | 786 | 86 | 1547 | 274 | 1341 | 268 | 125 | 95 | 285 | 17 | 1 | 103 | 148 |  | 5076 |
| 2007 | 831 | 192 | 1340 | 420 | 1881 | 275 | 289 | 68 | 329 | 29 | 1 | 32 | 73 |  | 5760 |
| 2008 | 1021 | 51 | 1005 | 345 | 3278 | 122 | 664 | 156 | 331 | 60 | 13 | 17 | 121 |  | 7184 |
| 2009 | 1048 | 189 | 938 | 284 | 3360 | 229 | 828 | 142 | 347 | 28 | 0 | 8 | 285 |  | 7687 |
| 2010 | 973 | 330 | 1843 | 255 | 4120 | 143 | 512 | 181 | 246 | 163 | 1 | 16 | 136 |  | 8918 |
| 2011 | 1251 | 202 | 831 | 226 | 4473 | 85 | 422 | 259 | 359 | 143 | 2 | 57 | 170 |  | 8479 |
| 2012 | 1771 | 164 | 600 | 273 | 2986 | 97 | 439 | 291 | 415 | 41 | 7 | 33 | 133 |  | 7251 |
| 2013 | 1366 | 210 | 648 | 334 | 3676 | 45 | 146 | 447 | 272 | 178 | 2 | 40 | 216 |  | 7581 |
| 2014 | 1391 | 121 | 744 | 208 | 3340 | 56 | 98 | 390 | 170 | 20 | 7 | 27 | 154 |  | 6726 |
| 2015 | 1122 | 301 | 1160 | 442 | 2675 | 69 | 159 | 175 | 180 | 87 | 14 | 39 | 117 |  | 6539 |
| 2016 | 1542 | 654 | 775 | 216 | 2221 | 86 | 248 | 239 | 158 | 48 | 3 | 51 | 328 |  | 6568 |
| 2017 | 1042 | 85 | 681 | 316 | 2709 | 99 | 75 | 271 | 315 | 188 | 3 | 26 | 249 |  | 6060 |
| 2018 | 1153 | 146 | 1541 | 178 | 1624 | 271 | 117 | 352 | 479 | 41 | 41 | 41 | 121 |  | 6105 |
| 2019 | 751 | 97 | 98 | 189 | 2103 | 379 | 131 | 415 | 292 | 47 | 0 | 15 | 159 |  | 5075 |

Table $1.3 \quad$ Consumption per cod by cod age group (kg/year), based on Norwegian consumption calculations.

| Year/Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0.247 | 0.815 | 1.683 | 2.521 | 3.951 | 5.208 | 8.009 | 8.524 | 9.180 | 9.912 | 9.954 |
| 1985 | 0.304 | 0.761 | 1.833 | 3.105 | 4.675 | 7.360 | 11.246 | 11.972 | 12.497 | 13.751 | 13.869 |
| 1986 | 0.161 | 0.498 | 1.343 | 3.152 | 5.669 | 6.884 | 11.018 | 11.944 | 12.749 | 13.513 | 13.768 |
| 1987 | 0.219 | 0.602 | 1.290 | 2.051 | 3.532 | 5.489 | 7.077 | 8.107 | 8.923 | 9.343 | 9.301 |
| 1988 | 0.164 | 0.702 | 1.150 | 2.149 | 3.743 | 5.877 | 10.098 | 11.222 | 12.575 | 13.127 | 13.373 |
| 1989 | 0.223 | 0.715 | 1.606 | 2.714 | 3.980 | 5.611 | 7.678 | 8.499 | 9.597 | 10.198 | 10.628 |
| 1990 | 0.363 | 0.906 | 1.909 | 3.058 | 4.218 | 5.447 | 6.527 | 6.877 | 7.075 | 7.455 | 7.955 |
| 1991 | 0.293 | 0.972 | 2.178 | 3.536 | 5.318 | 7.073 | 9.470 | 10.238 | 11.292 | 12.339 | 12.037 |
| 1992 | 0.215 | 0.665 | 2.100 | 3.135 | 4.142 | 5.093 | 7.868 | 9.023 | 9.402 | 10.124 | 10.156 |
| 1993 | 0.112 | 0.529 | 1.548 | 3.045 | 4.823 | 6.292 | 9.413 | 11.272 | 11.798 | 12.288 | 12.880 |
| 1994 | 0.130 | 0.406 | 0.924 | 2.523 | 3.508 | 4.544 | 6.404 | 8.844 | 9.716 | 9.988 | 10.232 |
| 1995 | 0.103 | 0.299 | 0.918 | 1.824 | 3.359 | 5.261 | 7.726 | 10.425 | 12.300 | 12.770 | 13.191 |
| 1996 | 0.108 | 0.359 | 0.938 | 1.855 | 3.055 | 4.434 | 7.409 | 11.124 | 14.591 | 15.048 | 15.432 |
| 1997 | 0.140 | 0.327 | 0.952 | 1.778 | 2.717 | 3.537 | 5.261 | 8.128 | 12.659 | 13.389 | 13.205 |
| 1998 | 0.117 | 0.400 | 0.991 | 1.953 | 2.922 | 4.188 | 5.751 | 8.078 | 11.375 | 12.071 | 12.113 |
| 1999 | 0.163 | 0.505 | 1.095 | 2.720 | 3.719 | 5.444 | 6.975 | 9.193 | 10.953 | 12.063 | 12.181 |
| 2000 | 0.170 | 0.499 | 1.239 | 2.467 | 4.262 | 5.650 | 7.975 | 9.405 | 12.679 | 13.401 | 13.542 |
| 2001 | 0.171 | 0.448 | 1.308 | 2.435 | 3.688 | 5.305 | 7.550 | 11.238 | 13.477 | 14.400 | 14.674 |
| 2002 | 0.199 | 0.553 | 1.163 | 2.443 | 3.382 | 4.721 | 6.366 | 9.069 | 10.301 | 11.513 | 11.098 |
| 2003 | 0.207 | 0.648 | 1.316 | 2.391 | 4.002 | 5.958 | 8.438 | 10.435 | 12.903 | 13.576 | 14.443 |
| 2004 | 0.222 | 0.476 | 1.298 | 2.285 | 3.339 | 5.568 | 7.444 | 11.468 | 17.366 | 19.237 | 18.956 |
| 2005 | 0.203 | 0.659 | 1.380 | 2.746 | 4.247 | 6.365 | 7.670 | 10.284 | 13.851 | 14.895 | 15.610 |
| 2006 | 0.204 | 0.626 | 1.584 | 2.811 | 4.241 | 6.316 | 7.868 | 11.626 | 14.023 | 15.100 | 15.929 |
| 2007 | 0.256 | 0.653 | 1.738 | 3.092 | 4.471 | 6.237 | 8.277 | 10.287 | 12.786 | 13.554 | 13.988 |
| 2008 | 0.204 | 0.724 | 1.469 | 2.877 | 4.082 | 7.111 | 8.407 | 11.463 | 15.655 | 16.348 | 16.617 |
| 2009 | 0.192 | 0.618 | 1.494 | 2.769 | 4.434 | 5.759 | 8.470 | 11.487 | 12.793 | 13.632 | 13.821 |
| 2010 | 0.203 | 0.635 | 1.357 | 2.504 | 3.989 | 5.709 | 8.447 | 12.078 | 15.363 | 16.040 | 16.394 |
| 2011 | 0.219 | 0.663 | 1.419 | 2.627 | 4.033 | 5.351 | 7.272 | 9.663 | 15.139 | 16.314 | 16.304 |
| 2012 | 0.231 | 0.763 | 1.503 | 2.688 | 4.103 | 5.077 | 7.312 | 10.038 | 15.400 | 16.594 | 16.518 |
| 2013 | 0.182 | 0.674 | 1.447 | 2.531 | 3.908 | 4.999 | 5.954 | 7.582 | 11.489 | 12.510 | 13.450 |
| 2014 | 0.224 | 0.648 | 1.308 | 2.549 | 3.763 | 4.253 | 5.837 | 8.010 | 10.796 | 11.514 | 12.026 |
| 2015 | 0.218 | 0.662 | 1.426 | 2.528 | 4.254 | 5.695 | 7.376 | 8.628 | 13.081 | 13.892 | 15.034 |
| 2016 | 0.252 | 0.722 | 1.578 | 2.769 | 3.919 | 5.514 | 7.201 | 8.040 | 12.056 | 12.652 | 14.479 |
| 2017 | 0.248 | 0.791 | 1.529 | 2.653 | 3.977 | 5.628 | 7.031 | 8.143 | 11.271 | 14.168 | 16.982 |
| 2018 | 0.194 | 0.775 | 1.566 | 2.813 | 4.391 | 5.208 | 6.811 | 10.602 | 12.879 | 17.074 | 15.980 |
| 2019 | 0.191 | 0.515 | 1.343 | 2.288 | 3.517 | 4.417 | 6.219 | 8.963 | 12.186 | 11.715 | 12.973 |
| 2020 | 0.175 | 0.465 | 1.086 | 2.461 | 3.503 | 4.926 | 6.796 | 10.080 | 11.988 | 13.655 | 15.837 |
| Average | 0.201 | 0.613 | 1.406 | 2.590 | 3.969 | 5.500 | 7.639 | 9.785 | 12.275 | 13.221 | 13.647 |


| Table 1.4 | Consumption per cod by cod age group (kg/year), based on Russian consumption calculations. |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | NOT UPDATED THIS YEAR |  |  |  |  |  |  |  |  |  |  |  |
| Year/Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13+ |
| 1984 | 0.262 | 0.895 | 1.611 | 2.748 | 3.848 | 5.486 | 6.992 | 8.561 | 10.572 | 13.166 | 13.200 | 15.547 | 17.153 |
| 1985 | 0.295 | 0.753 | 1.658 | 2.681 | 4.264 | 6.599 | 8.241 | 9.745 | 10.974 | 14.448 | 17.327 | 17.391 | 19.186 |
| 1986 | 0.179 | 0.526 | 1.455 | 3.455 | 5.001 | 5.991 | 6.458 | 8.157 | 9.766 | 11.457 | 13.188 | 14.621 | 16.134 |
| 1987 | 0.145 | 0.432 | 0.852 | 1.558 | 3.073 | 4.380 | 7.357 | 9.667 | 12.705 | 14.481 | 15.899 | 16.616 | 18.318 |
| 1988 | 0.183 | 0.704 | 1.075 | 1.628 | 2.391 | 4.386 | 8.207 | 9.978 | 10.868 | 16.536 | 14.639 | 16.046 | 17.000 |
| 1989 | 0.282 | 0.909 | 1.465 | 2.207 | 3.243 | 4.798 | 6.578 | 8.725 | 11.134 | 15.798 | 16.313 | 18.436 | 18.041 |
| 1990 | 0.288 | 1.006 | 1.694 | 2.693 | 3.278 | 3.833 | 5.583 | 6.870 | 10.715 | 11.426 | 13.555 | 15.964 | 17.595 |
| 1991 | 0.241 | 0.936 | 2.670 | 4.472 | 6.037 | 7.844 | 9.590 | 11.543 | 14.969 | 19.292 | 18.590 | 21.720 | 23.960 |
| 1992 | 0.178 | 0.969 | 2.475 | 2.866 | 3.995 | 5.137 | 6.723 | 7.414 | 8.755 | 12.303 | 14.288 | 15.184 | 16.745 |
| 1993 | 0.133 | 0.476 | 1.512 | 2.865 | 3.944 | 5.108 | 7.372 | 8.945 | 10.343 | 11.600 | 14.835 | 16.536 | 18.249 |
| 1994 | 0.180 | 0.512 | 1.212 | 2.402 | 3.517 | 5.359 | 7.560 | 10.001 | 11.818 | 12.896 | 14.499 | 17.656 | 19.469 |
| 1995 | 0.194 | 0.497 | 0.962 | 1.801 | 3.204 | 4.847 | 7.332 | 9.688 | 13.835 | 15.247 | 16.899 | 19.273 | 21.254 |
| 1996 | 0.170 | 0.498 | 1.028 | 1.916 | 3.059 | 4.189 | 6.987 | 10.212 | 12.185 | 13.614 | 14.529 | 16.275 | 17.945 |
| 1997 | 0.119 | 0.341 | 0.992 | 1.908 | 2.668 | 3.503 | 4.954 | 7.980 | 12.174 | 16.762 | 16.710 | 18.410 | 20.308 |
| 1998 | 0.232 | 0.528 | 1.081 | 2.016 | 2.823 | 4.089 | 5.469 | 7.346 | 9.586 | 13.012 | 14.404 | 15.640 | 17.243 |
| 1999 | 0.261 | 0.431 | 1.128 | 2.490 | 3.676 | 5.222 | 6.398 | 8.220 | 9.194 | 13.364 | 15.268 | 16.990 | 18.727 |
| 2000 | 0.186 | 0.545 | 1.288 | 2.551 | 4.387 | 6.559 | 8.833 | 10.483 | 11.522 | 15.132 | 17.090 | 19.793 | 21.822 |
| 2001 | 0.150 | 0.413 | 1.163 | 2.110 | 3.430 | 5.571 | 6.835 | 10.233 | 12.457 | 15.130 | 17.341 | 19.307 | 21.345 |
| 2002 | 0.252 | 0.677 | 1.303 | 2.699 | 3.847 | 5.591 | 7.846 | 10.796 | 13.238 | 18.787 | 17.836 | 20.278 | 22.359 |
| 2003 | 0.228 | 0.618 | 1.296 | 2.028 | 3.547 | 4.716 | 6.684 | 8.905 | 13.418 | 14.492 | 19.480 | 19.309 | 21.292 |
| 2004 | 0.250 | 0.654 | 1.412 | 2.567 | 3.857 | 5.660 | 7.730 | 11.126 | 15.907 | 20.770 | 21.607 | 24.940 | 27.503 |
| 2005 | 0.255 | 0.687 | 1.514 | 2.504 | 3.896 | 5.264 | 7.192 | 9.395 | 13.163 | 15.981 | 20.628 | 21.448 | 23.639 |
| 2006 | 0.354 | 0.925 | 1.881 | 2.813 | 4.019 | 5.332 | 7.450 | 10.328 | 13.111 | 17.759 | 19.488 | 22.322 | 24.609 |
| 2007 | 0.234 | 0.681 | 1.874 | 3.128 | 4.459 | 5.893 | 7.563 | 9.178 | 12.032 | 15.919 | 19.961 | 21.644 | 23.863 |
| 2008 | 0.223 | 0.719 | 1.697 | 2.959 | 4.194 | 6.073 | 7.809 | 10.464 | 13.627 | 17.254 | 21.590 | 23.373 | 25.779 |
| 2009 | 0.217 | 0.624 | 1.495 | 2.526 | 4.304 | 5.623 | 7.855 | 11.490 | 13.341 | 15.988 | 18.770 | 21.866 | 24.111 |
| 2010 | 0.235 | 0.651 | 1.401 | 2.577 | 4.065 | 5.757 | 8.312 | 11.805 | 16.090 | 16.844 | 20.129 | 23.023 | 25.387 |
| 2011 | 0.248 | 0.721 | 1.497 | 2.513 | 3.859 | 4.963 | 6.848 | 9.213 | 13.799 | 19.074 | 20.784 | 23.791 | 26.241 |
| 2012 | 0.207 | 0.588 | 1.203 | 2.292 | 3.266 | 4.461 | 5.862 | 7.629 | 11.713 | 16.211 | 19.345 | 21.032 | 23.190 |
| 2013 | 0.190 | 0.656 | 1.641 | 2.552 | 3.809 | 4.952 | 5.791 | 7.757 | 10.881 | 14.989 | 19.785 | 22.386 | 24.691 |
| 2014 | 0.242 | 0.622 | 1.321 | 2.340 | 3.608 | 4.387 | 5.560 | 7.447 | 9.017 | 12.547 | 16.044 | 18.854 | 20.781 |
| 2015 | 0.234 | 0.745 | 1.390 | 2.406 | 3.915 | 4.922 | 5.960 | 7.505 | 10.265 | 12.116 | 16.245 | 19.978 | 22.023 |
| 2016 | 0.307 | 0.870 | 1.722 | 2.813 | 3.474 | 4.740 | 6.754 | 9.117 | 10.665 | 14.810 | 19.921 | 24.195 | 26.683 |
| 2017 | 0.244 | 0.779 | 1.582 | 2.531 | 3.748 | 4.943 | 6.601 | 9.180 | 11.302 | 16.016 | 20.086 | 23.464 | 25.870 |
| 2018 | 0.316 | 0.867 | 1.846 | 2.699 | 3.736 | 5.000 | 6.489 | 9.170 | 11.166 | 14.577 | 18.672 | 21.848 | 24.091 |
| 2019 | 0.269 | 0.655 | 1.383 | 2.204 | 3.316 | 4.500 | 6.415 | 9.078 | 13.251 | 15.509 | 19.423 | 22.635 | 24.958 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Average | 0.227 | 0.670 | 1.466 | 2.514 | 3.743 | 5.158 | 7.005 | 9.260 | 11.932 | 15.147 | 17.455 | 19.661 | 21.599 |

Table 1.5 Proportion of cod in cod diet, based on Norwegian consumption calculations

| Year/age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0.0000 | 0.0000 | 0.0032 | 0.0000 | 0.0432 | 0.0262 | 0.0332 | 0.0361 | 0.0371 | 0.0392 | 0.0394 |
| 1985 | 0.0015 | 0.0009 | 0.0014 | 0.0017 | 0.0312 | 0.0074 | 0.0822 | 0.0826 | 0.0833 | 0.0835 | 0.0840 |
| 1986 | 0.0000 | 0.0022 | 0.0015 | 0.0004 | 0.0130 | 0.1743 | 0.1760 | 0.1761 | 0.1758 | 0.1749 | 0.1745 |
| 1987 | 0.0000 | 0.0000 | 0.0007 | 0.0050 | 0.0103 | 0.0244 | 0.0383 | 0.0395 | 0.0412 | 0.0409 | 0.0443 |
| 1988 | 0.0000 | 0.0000 | 0.0000 | 0.0002 | 0.0059 | 0.0014 | 0.0037 | 0.0036 | 0.0031 | 0.0035 | 0.0031 |
| 1989 | 0.0000 | 0.0006 | 0.0016 | 0.0019 | 0.0027 | 0.0039 | 0.0036 | 0.0036 | 0.0039 | 0.0038 | 0.0040 |
| 1990 | 0.0000 | 0.0000 | 0.0000 | 0.0007 | 0.0010 | 0.0010 | 0.0165 | 0.0172 | 0.0181 | 0.0179 | 0.0178 |
| 1991 | 0.0000 | 0.0005 | 0.0000 | 0.0003 | 0.0032 | 0.0020 | 0.0222 | 0.0227 | 0.0230 | 0.0231 | 0.0231 |
| 1992 | 0.0000 | 0.0021 | 0.0037 | 0.0129 | 0.0248 | 0.0475 | 0.0119 | 0.0160 | 0.0232 | 0.0232 | 0.0231 |
| 1993 | 0.0000 | 0.0410 | 0.0370 | 0.0515 | 0.0541 | 0.1135 | 0.0498 | 0.0795 | 0.0797 | 0.0796 | 0.0802 |
| 1994 | 0.0000 | 0.0037 | 0.0927 | 0.0349 | 0.0285 | 0.0785 | 0.1248 | 0.1330 | 0.2659 | 0.2674 | 0.2668 |
| 1995 | 0.0069 | 0.0812 | 0.0747 | 0.0803 | 0.0923 | 0.1118 | 0.1387 | 0.2526 | 0.2542 | 0.2539 | 0.2545 |
| 1996 | 0.0000 | 0.1500 | 0.2566 | 0.2051 | 0.1321 | 0.1263 | 0.1874 | 0.2091 | 0.2436 | 0.2447 | 0.2437 |
| 1997 | 0.0000 | 0.0687 | 0.0762 | 0.1137 | 0.1558 | 0.1555 | 0.2315 | 0.2269 | 0.2919 | 0.2850 | 0.2916 |
| 1998 | 0.0000 | 0.0134 | 0.0272 | 0.0418 | 0.1037 | 0.0978 | 0.1090 | 0.1498 | 0.2722 | 0.2741 | 0.2718 |
| 1999 | 0.0000 | 0.0000 | 0.0048 | 0.0136 | 0.0147 | 0.0338 | 0.0618 | 0.1114 | 0.1902 | 0.1907 | 0.1843 |
| 2000 | 0.0000 | 0.0000 | 0.0287 | 0.0148 | 0.0134 | 0.0266 | 0.0497 | 0.0570 | 0.2682 | 0.2699 | 0.2594 |
| 2001 | 0.0000 | 0.0160 | 0.0116 | 0.0082 | 0.0131 | 0.0241 | 0.0498 | 0.0375 | 0.3250 | 0.3233 | 0.3268 |
| 2002 | 0.0000 | 0.0385 | 0.0597 | 0.0142 | 0.0187 | 0.0284 | 0.0357 | 0.0623 | 0.1582 | 0.1560 | 0.1555 |
| 2003 | 0.0000 | 0.0190 | 0.0198 | 0.0199 | 0.0206 | 0.0188 | 0.0451 | 0.1030 | 0.2194 | 0.2219 | 0.2228 |
| 2004 | 0.0081 | 0.0234 | 0.0280 | 0.0269 | 0.0296 | 0.0319 | 0.0380 | 0.0663 | 0.1062 | 0.1062 | 0.1077 |
| 2005 | 0.0000 | 0.0266 | 0.0230 | 0.0266 | 0.0145 | 0.0277 | 0.0436 | 0.0779 | 0.1484 | 0.1462 | 0.1437 |
| 2006 | 0.0000 | 0.0103 | 0.0007 | 0.0128 | 0.0288 | 0.0158 | 0.0392 | 0.0368 | 0.0810 | 0.0821 | 0.0820 |
| 2007 | 0.0000 | 0.0000 | 0.0011 | 0.0117 | 0.0119 | 0.0304 | 0.0282 | 0.0901 | 0.1407 | 0.1413 | 0.1383 |
| 2008 | 0.0000 | 0.0559 | 0.0257 | 0.0101 | 0.0157 | 0.0098 | 0.0764 | 0.0873 | 0.0975 | 0.0959 | 0.0981 |
| 2009 | 0.0116 | 0.0225 | 0.0262 | 0.0251 | 0.0152 | 0.0139 | 0.0219 | 0.0945 | 0.1078 | 0.1082 | 0.1076 |
| 2010 | 0.0000 | 0.0327 | 0.0580 | 0.0270 | 0.0243 | 0.0243 | 0.0203 | 0.0383 | 0.1367 | 0.1369 | 0.1353 |
| 2011 | 0.0129 | 0.0152 | 0.0492 | 0.0170 | 0.0361 | 0.0300 | 0.0238 | 0.0575 | 0.1279 | 0.1279 | 0.1278 |
| 2012 | 0.0274 | 0.0608 | 0.0640 | 0.0618 | 0.0274 | 0.0432 | 0.0410 | 0.0373 | 0.0685 | 0.0691 | 0.0681 |
| 2013 | 0.0214 | 0.0303 | 0.0459 | 0.0389 | 0.0276 | 0.0224 | 0.0478 | 0.0538 | 0.1166 | 0.1171 | 0.1335 |
| 2014 | 0.0824 | 0.0363 | 0.0450 | 0.0342 | 0.0213 | 0.0456 | 0.0661 | 0.0787 | 0.0658 | 0.0658 | 0.0752 |
| 2015 | 0.0000 | 0.0088 | 0.0308 | 0.0283 | 0.0266 | 0.0192 | 0.0233 | 0.0281 | 0.0555 | 0.0553 | 0.0539 |
| 2016 | 0.0157 | 0.0192 | 0.0063 | 0.0393 | 0.0146 | 0.0172 | 0.0266 | 0.0137 | 0.0906 | 0.0914 | 0.0910 |
| 2017 | 0.0419 | 0.0354 | 0.0386 | 0.0470 | 0.0436 | 0.0400 | 0.0560 | 0.0913 | 0.0686 | 0.1015 | 0.1409 |
| 2018 | 0.0000 | 0.0186 | 0.0680 | 0.0480 | 0.0351 | 0.0378 | 0.0567 | 0.0310 | 0.0243 | 0.0076 | 0.0252 |
| 2019 | 0.0000 | 0.0000 | 0.0328 | 0.0296 | 0.0339 | 0.0228 | 0.0366 | 0.0741 | 0.0934 | 0.0252 | 0.0792 |
| 2020 | 0.0000 | 0.0227 | 0.0013 | 0.0041 | 0.0110 | 0.0177 | 0.0311 | 0.0504 | 0.0683 | 0.0649 | 0.1118 |

Table 1.6 Proportion of haddock in cod diet, based on Norwegian consumption calculations

| Year/age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0.0443 | 0.0175 | 0.0053 | 0.0225 | 0.0455 | 0.0215 | 0.0022 | 0.0020 | 0.0019 | 0.0018 | 0.0017 |
| 1985 | 0.0205 | 0.0227 | 0.0052 | 0.0076 | 0.0207 | 0.0109 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1986 | 0.0000 | 0.0187 | 0.0015 | 0.0866 | 0.0005 | 0.0530 | 0.0249 | 0.0248 | 0.0257 | 0.0286 | 0.0301 |
| 1987 | 0.0000 | 0.0052 | 0.0003 | 0.0025 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1988 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0003 | 0.0034 | 0.0034 | 0.0034 | 0.0039 | 0.0035 | 0.0039 |
| 1989 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0002 | 0.0339 | 0.0338 | 0.0349 | 0.0347 | 0.0356 |
| 1990 | 0.0000 | 0.0000 | 0.0000 | 0.0024 | 0.0021 | 0.0007 | 0.0130 | 0.0124 | 0.0117 | 0.0118 | 0.0119 |
| 1991 | 0.0000 | 0.0000 | 0.0098 | 0.0079 | 0.0045 | 0.0051 | 0.0031 | 0.0030 | 0.0029 | 0.0028 | 0.0028 |
| 1992 | 0.0000 | 0.0000 | 0.0014 | 0.0683 | 0.0208 | 0.0271 | 0.0278 | 0.0317 | 0.0462 | 0.0462 | 0.0461 |
| 1993 | 0.0000 | 0.0000 | 0.0204 | 0.0073 | 0.0149 | 0.0144 | 0.0278 | 0.0261 | 0.0261 | 0.0261 | 0.0263 |
| 1994 | 0.0000 | 0.0000 | 0.0065 | 0.0131 | 0.0069 | 0.0141 | 0.0298 | 0.0491 | 0.0456 | 0.0452 | 0.0453 |
| 1995 | 0.0000 | 0.0354 | 0.0030 | 0.0429 | 0.0260 | 0.0241 | 0.0393 | 0.0956 | 0.1617 | 0.1615 | 0.1619 |
| 1996 | 0.0000 | 0.0000 | 0.0592 | 0.0155 | 0.0098 | 0.0170 | 0.0376 | 0.0485 | 0.0925 | 0.1016 | 0.0981 |
| 1997 | 0.0000 | 0.0000 | 0.0242 | 0.0189 | 0.0245 | 0.0158 | 0.0127 | 0.0175 | 0.0561 | 0.0569 | 0.0539 |
| 1998 | 0.0000 | 0.0000 | 0.0115 | 0.0120 | 0.0227 | 0.0192 | 0.0106 | 0.0323 | 0.0161 | 0.0166 | 0.0160 |
| 1999 | 0.0000 | 0.0000 | 0.0028 | 0.0078 | 0.0158 | 0.0124 | 0.0120 | 0.0139 | 0.0224 | 0.0225 | 0.0217 |
| 2000 | 0.0000 | 0.0000 | 0.0233 | 0.0102 | 0.0178 | 0.0116 | 0.0158 | 0.0525 | 0.0286 | 0.0285 | 0.0287 |
| 2001 | 0.0000 | 0.0081 | 0.0052 | 0.0163 | 0.0147 | 0.0171 | 0.0194 | 0.0198 | 0.0337 | 0.0330 | 0.0345 |
| 2002 | 0.0000 | 0.0000 | 0.0185 | 0.0339 | 0.0353 | 0.0471 | 0.0747 | 0.0761 | 0.1830 | 0.1793 | 0.1785 |
| 2003 | 0.0000 | 0.0000 | 0.0145 | 0.0311 | 0.0595 | 0.0436 | 0.0553 | 0.1215 | 0.1079 | 0.1078 | 0.1078 |
| 2004 | 0.0044 | 0.0418 | 0.0745 | 0.0388 | 0.0575 | 0.0501 | 0.0564 | 0.0996 | 0.0910 | 0.0911 | 0.0924 |
| 2005 | 0.0000 | 0.0853 | 0.1047 | 0.0595 | 0.0621 | 0.0646 | 0.1038 | 0.1082 | 0.1115 | 0.1101 | 0.1085 |
| 2006 | 0.0000 | 0.0409 | 0.0829 | 0.0872 | 0.0604 | 0.0897 | 0.0716 | 0.1063 | 0.0962 | 0.0957 | 0.0958 |
| 2007 | 0.0000 | 0.0035 | 0.0462 | 0.0415 | 0.0833 | 0.0980 | 0.1335 | 0.1152 | 0.1631 | 0.1627 | 0.1648 |
| 2008 | 0.0000 | 0.0045 | 0.0106 | 0.0156 | 0.0383 | 0.0753 | 0.1148 | 0.1327 | 0.2329 | 0.2346 | 0.2321 |
| 2009 | 0.0000 | 0.0218 | 0.0241 | 0.0182 | 0.0142 | 0.0362 | 0.1090 | 0.0595 | 0.1881 | 0.1868 | 0.1891 |
| 2010 | 0.0000 | 0.0031 | 0.0279 | 0.0182 | 0.0178 | 0.0217 | 0.0362 | 0.1420 | 0.1819 | 0.1806 | 0.1810 |
| 2011 | 0.0000 | 0.0049 | 0.0362 | 0.0285 | 0.0087 | 0.0204 | 0.0411 | 0.0924 | 0.1633 | 0.1630 | 0.1625 |
| 2012 | 0.0000 | 0.0000 | 0.0113 | 0.0282 | 0.0337 | 0.0271 | 0.0368 | 0.0335 | 0.0859 | 0.0848 | 0.0872 |
| 2013 | 0.0000 | 0.0073 | 0.0309 | 0.0112 | 0.0314 | 0.0233 | 0.0147 | 0.0363 | 0.0615 | 0.0615 | 0.0916 |
| 2014 | 0.0000 | 0.0089 | 0.0037 | 0.0255 | 0.0080 | 0.0047 | 0.0022 | 0.0340 | 0.0143 | 0.0143 | 0.0194 |
| 2015 | 0.0000 | 0.0175 | 0.0409 | 0.0254 | 0.0172 | 0.0166 | 0.0258 | 0.0197 | 0.0384 | 0.0385 | 0.0399 |
| 2016 | 0.0000 | 0.0051 | 0.0799 | 0.0771 | 0.0265 | 0.0259 | 0.0323 | 0.0420 | 0.0342 | 0.0343 | 0.0339 |
| 2017 | 0.0106 | 0.0429 | 0.0153 | 0.0450 | 0.0462 | 0.0568 | 0.0466 | 0.0528 | 0.0795 | 0.0677 | 0.0867 |
| 2018 | 0.0000 | 0.0000 | 0.0434 | 0.0365 | 0.0590 | 0.0661 | 0.0551 | 0.0588 | 0.0821 | 0.0304 | 0.1164 |
| 2019 | 0.0000 | 0.0000 | 0.0284 | 0.0564 | 0.0422 | 0.0491 | 0.0513 | 0.0401 | 0.0345 | 0.0644 | 0.2709 |
| 2020 | 0.0000 | 0.0000 | 0.0011 | 0.0063 | 0.0037 | 0.0096 | 0.0257 | 0.0707 | 0.0514 | 0.0816 | 0.0287 |
| Average | 0.0022 | 0.0107 | 0.0236 | 0.0277 | 0.0257 | 0.0296 | 0.0378 | 0.0516 | 0.0706 | 0.0706 | 0.0785 |

Table 1.7. Parameters of TitovES and TitovEL models (subscripts correspond to the time-lag in months before the start of the year to which the value Cod3 is attributed).

| Year | Cod3 | OxSatt 39 | DOxSatt ${ }_{13}$ | ITwt $_{43}$ | Icet $_{15}$ | explcet $_{40}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1962 | 1252375 | -0.19 | -6.6 | 1.86 | 0.5 | 0 |
| 1963 | 900621 | -0.94 | -2.37 | 1.59 | 1.5 | 0 |
| 1964 | 468028 | 1.63 | 1.23 | 2.47 | 9 | 0 |
| 1965 | 870506 | 0.88 | -0.2 | 3.91 | 15.7 | 0 |
| 1966 | 1842715 | -1.09 | -3.98 | 7.97 | 5.3 | 0 |
| 1967 | 1311586 | -0.23 | -2.84 | 8.23 | 5 | 9.3 |
| 1968 | 183717 | 1.5 | -0.13 | 3.78 | 15.5 | 0 |
| 1969 | 110450 | 0.85 | 0.63 | 1.77 | 15.9 | 0 |
| 1970 | 205641 | -0.17 | -0.23 | 3.51 | 19.8 | 7.9 |
| 1971 | 402577 | 0.06 | -0.12 | -0.13 | 18.8 | 2.7 |
| 1972 | 1045979 | -3.32 | -6.59 | 14.55 | -0.6 | 428.9 |
| 1973 | 1723668 | -2.1 | -10.37 | 19.14 | 1.8 | 768.6 |
| 1974 | 568211 | 1.06 | -1.73 | 2.4 | 2 | 0 |
| 1975 | 608710 | 1.9 | 0.78 | -2.64 | -1.2 | 0 |
| 1976 | 607084 | 1.33 | -1.28 | -3.07 | -1.9 | 0 |
| 1977 | 372778 | -0.07 | -1.84 | -2.44 | 2.5 | 0 |
| 1978 | 622679 | 1.19 | 0.1 | 1.05 | -1 | 0 |
| 1979 | 202675 | 0.5 | -1.48 | -0.12 | 3.5 | 0 |
| 1980 | 130292 | -0.31 | -2.72 | 1.98 | 12.9 | 0 |
| 1981 | 143781 | 0.76 | -0.18 | 1.94 | 14.7 | 0 |
| 1982 | 183737 | 0.8 | 0.61 | -3.15 | 8 | 0.1 |
| 1983 | 141514 | 0.78 | 0.22 | 1.87 | 12.2 | 8.5 |
| 1984 | 442251 | -2.21 | -2.35 | -3.08 | 12.9 | 0 |
| 1985 | 534310 | -0.1 | -1.17 | 3.59 | -1.2 | 0.1 |
| 1986 | 1374917 | -2.14 | -4.39 | 1.39 | -8.5 | 2.9 |
| 1987 | 360087 | -0.33 | -1.69 | 2.12 | 0.6 | 0 |
| 1988 | 335536 | 0.87 | -1.4 | -2.34 | 3.8 | 0 |
| 1989 | 157635 | 0.32 | -3.42 | -5.17 | 10.5 | 0 |


| Year | Cod3 | OxSatt ${ }_{39}$ | DOxSatt ${ }_{13}$ | ITwt $_{43}$ | Icet $_{15}$ | explcet $_{40}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 130130 | 1.11 | -1.32 | -4.21 | 10.5 | 0 |
| 1991 | 295846 | 0.88 | 0.7 | 2.42 | 6.5 | 0 |
| 1992 | 715916 | 1.34 | 0.48 | 1.37 | -0.9 | 0 |
| 1993 | 988150 | -1.98 | -3.86 | 6.12 | -0.6 | 0 |
| 1994 | 752473 | -0.5 | -2.26 | 8.25 | -4.9 | 0 |
| 1995 | 539384 | 0.83 | -2.42 | 4.36 | 1.8 | 0 |
| 1996 | 407389 | 0.86 | -0.08 | 0.55 | 0.7 | 0 |
| 1997 | 785420 | 0.88 | 0.17 | 3.11 | -7.3 | 0 |
| 1998 | 1063528 | 0.3 | -6.08 | -2.32 | -2.5 | 0 |
| 1999 | 632034 | -0.72 | -2.4 | -6.81 | 2.9 | 0 |
| 2000 | 749727 | 1.86 | 1.55 | -2.29 | 13.6 | 0 |
| 2001 | 593152 | 0.62 | 0.05 | -6.04 | 2.3 | 0 |
| 2002 | 374202 | -0.88 | -0.98 | 3.63 | -9.9 | 0.8 |
| 2003 | 756675 | -0.39 | -0.64 | 8.5 | -5.8 | 0 |
| 2004 | 242069 | -2.2 | -2.53 | -4.62 | -1.4 | 0 |
| 2005 | 693264 | -1.65 | -1.82 | $-1.45$ | 4.9 | 0 |
| 2006 | 536630 | -1.18 | -1.65 | -4 | -6 | 0 |
| 2007 | 1243906 | -1.39 | -4.42 | 7.42 | -12.3 | 0 |
| 2008 | 1002761 | -1.14 | -1.59 | 3.39 | -18 | 0 |
| 2009 | 581758 | 0.79 | -1.83 | -1.61 | -17.5 | 0 |
| 2010 | 201832 | -0.38 | -2.6 | -8.94 | -9 | 0 |
| 2011 | 358117 | 0.83 | -0.07 | -5 | -4.3 | 0 |
| 2012 | 503017 | 0.91 | -0.13 | -5.05 | -4.3 | 0 |
| 2013 | 464921 | 0.04 | -0.09 | 1.44 | -10.5 | 0 |
| 2014 | 852202 | -0.46 | -1 | 1.43 | -17.8 | 0 |
| 2015 | 452019 | -1.26 | -1.62 | -2.22 | -10.5 | 0 |
| 2016 | 286334 | -1.31 | -1.92 | -7.52 | -5.8 | 0 |
| 2017 | 781901 | -0.33 | -0.64 | -1.69 | -14.4 | 0 |
| 2018 | 508296 | -1.24 | -1.41 | 0.1 | -20.9 | 0 |


| Year | Cod3 | OxSatt 39 | DOxSatt $_{13}$ | ITwt $_{43}$ | Icet $_{15}$ | explcet $4_{40}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2019 | 659091 | -0.63 | -1.08 | -1.71 | -13.2 | 0 |
| 2020 | 572413 | -2.02 | -2.19 | -6.35 | -13.6 | 0 |
| 2021 | NA | -0.8 | -1.08 | -1.33 | -9.2 | 0 |
| 2022 | NA | -1.55 | -2.1 | -2.47 | -12.8 | 0 |
| 2023 | NA | -1.52 | NA | -4.18 | $N A$ | 0 |
| 2024 | -0.31 | $N A$ | -5.63 | $N A$ | 0 |  |

Table 1.8 Initial data for RCT3 model.

| year class | Recruitment | BST1 | BST2 | BST3 | BSA1 | BSA2 | BSA3 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1982 | 534 | NA | NA | NA | NA | NA | NA |
| 1983 | 1375 | $N A$ | $N A$ | $N A$ | $N A$ | $N A$ | $N A$ |
| 1984 | 360 | 336 | $N A$ | $N A$ | $N A$ | $N A$ | $N$ |


| year class | Recruitment | BST1 | BST2 | BST3 | BSA1 | BSA2 | BSA3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 537 | 323 | 217 | 116 | 161 | 139 | 65 |
| 2004 | 1244 | 853 | 289 | 361 | 500 | 158 | 59 |
| 2005 | 1003 | 674 | 370 | 194 | 411 | 47 | 200 |
| 2006 | 582 | 595 | 102 | 126 | 85 | 94 | 108 |
| 2007 | 202 | 69 | 36 | 37 | 51 | 26 | 23 |
| 2008 | 358 | 389 | 95 | 85 | 205 | 44 | 40 |
| 2009 | 503 | 1028 | 226 | 76 | 620 | 91 | 83 |
| 2010 | 465 | 617 | 100 | 69 | 266 | 40 | 61 |
| 2011 | 852 | 703 | 143 | 227 | 497 | 89 | 287 |
| 2012 | 452 | 436 | 191 | 144 | 313 | 211 | 139 |
| 2013 | 286 | 1246 | 343 | 99 | 1759 | 211 | 56 |
| 2014 | 782 | 1642 | 306 | 179 | 1904 | 202 | 112 |
| 2015 | 508 | 312 | 129 | 139 | 241 | 73 | 109 |
| 2016 | 659 | 645 | 501 | 282 | 439 | 280 | 204 |
| 2017 | 572 | 2714 | 559 | 238 | 2058 | 362 | 117 |
| 2018 | NA | 1791 | 274 | 115 | 1437 | 158 | 70 |
| 2019 | NA | 165 | 33 | NA | 93 | 17 | NA |
| 2020 | NA | 88 | NA | NA | 44 | NA | NA |

Table 1.9. Overview available prognoses of NEA cod recruitment (in million individuals of age 3) from different models.

| Model | Parameter | Years of <br> prediction | 2021 <br> Prognosis | Prognosis | Prognosis | Prognosis |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| TitovEL | R at age 3 | 4 | 590 | 614 | 548 | 386 |
|  | Model weight | 0.34 | 0.47 | 1 | 1 |  |
| TitovES | R at age 3 | 2 | 559 | 627 | 0 | 0 |
|  | Model weight |  | 0.42 | 0.53 | 384 |  |
| RCT3 | R at age 3 | 3 | 525 | 301 | 0 | 384 |



Figure 1.1. Standard errors of the NEA cod recruitment predicted values from the SAM values.

## 2 Norwegian coastal cod ${ }^{1}$

A benchmark assessment (WKBARFAR) was conducted in February 2021 to address the failure of the current management plan to reduce fishing mortality on Norwegian coastal cod (NCC; ICES, 2021a). The main outcome of the benchmark was that from assessment year 2021 onwards, Norwegian coastal cod (former stock code: cod.27.1-2coast) was split into two stocks/components by 67 degrees latitude (Figure 2.0.1); a data-rich one in the north: cod.27.1-2coastN (northern Norwegian coastal cod); and a data-limited one in the south: cod.27.2coastS (southern Norwegian coastal cod).

The majority (approximately $80-90 \%$ ) of NCC catches are taken north of $67^{\circ} \mathrm{N}$ (Table 2.1.1), and this is also where the coastal survey has the best coverage. Genetic studies have revealed a genetic gradient in cod along the Norwegian coast without areas of distinct breaks in population connectivity (Dahle et al., 2018). However, NCC in northern Norway have more genetic material in common with the Northeast Arctic cod (NEAC; cod.27.1-2), compared to Norwegian coastal cod further south (Dahle et al., 2018).

Recent updates of the catch series, a revision of the acoustic survey index and a new swept-area index have improved the data basis for assessment in the northern area. The data for northern Norwegian coastal cod were considered of high enough quality to support an age-based analytical assessment. Southern Norwegian coastal cod $\left(62-67^{\circ} \mathrm{N}\right)$ represents the remaining commercial catches of NCC north of $62^{\circ} \mathrm{N}$ (approximately $10-20 \%$ ) and is not as consistently covered by the main survey relevant to monitoring cod. Current data availability and quality cannot support a full analytical assessment, and a data-limited approach has therefore been developed to support management of this stock.


Figure 2.0.1 Norwegian catch reporting areas used to define stock distribution areas for northern Norwegian coastal cod (left) and southern Norwegian coastal cod (right).

[^4]
### 2.1 Fisheries (both stocks)

Coastal cod is fished throughout the year and within nearly all the distribution areas in the Norwegian statistical areas $03,04,05,00,06,07$ (Figure 2.0.1). Most of the coastal cod catches are taken as a bycatch in fisheries aimed at Northeast Arctic cod during its spawning and feeding migrations to coastal waters. The main fishery for coastal cod, therefore, takes place in the first half of the year. The main fishing areas are along the coast from Varangerfjord to Lofoten (areas $03,04,05,00$ ).

Recreational and tourist fisheries take an important fraction of the total catches in some local areas, especially near the coastal cities, and in some fjords where commercial fishing activity is low. Recreational catches are a much larger proportion of the total for the southern stock than for the northern stock, respectively about $50 \%$ vs. $15 \%$. However, there are few reports trying to assess the amount in certain years. In 2010, these reports were used to construct a time-series of recreational catches (ICES 2010). These catch estimates are quite uncertain. No additional information was included during 2010-2019, and the annual recreational catch during this period has been assumed equal to the one estimated for 2009 (12 700 t ).

A new project was conducted in the period 2017-2020 by IMR in collaboration with several Norwegian institutions (NINA, Akvaplan-niva, NMBU and Nordland Research), and a number of international partners. Three study areas Troms, Hordaland, and Oslofjord, were chosen because they represent contrasts in recreational fishing. The project is currently being finished and reports will follow, but some preliminary results were presented at the benchmark assessment (WKBARFAR WD13, ICES 2021a), and further used in the present coastal cod assessments.

Historically there has been no reporting system for NCC taken by recreational or tourist fishers in Norway. In 2019, the Norwegian Directorate for Fisheries established a web portal for obligatory catch reporting (both kept and released fish) by all registered fishing businesses. Tourist fishing effort related to tourist fishing businesses has about doubled from 2009 to 2019. The total quantity of cod caught by tourists staying in tourist businesses has also more than doubled from 1586 tonnes in 2009 (Vølstad et al., 2011) to about 3455 tonnes in 2019.

The current (2019) documented estimate of about 9000 tonnes (WKBARFAR WD13, ICES 2021a) is clearly an underestimate as tourists outside registered tourist businesses and residents fishing with fixed gears are not included. In the estimate of 9000 tonnes is also a share of the catch taken by anglers and released again. Based on investigations in other countries, the AFWG anticipates a mortality rate of $100 \%$ of fish caught by rod from land, and $20 \%$ of released cod caught by rod and handline at sea (e.g. Weltersbach and Strehlow, 2013; Capizzano et al., 2016). Until there is a better quantification of the missing recreational segments, the benchmark WK proposed to keep the quantity of 12700 tonnes recreational catch of Norwegian coastal cod north of $62^{\circ} \mathrm{N}$ on top of the commercial reported landings, with 7900 tonnes north of $67^{\circ} \mathrm{N}$ and 4800 tonnes between $62-67^{\circ} \mathrm{N}$ (Table 2.1.1).

The catches reported (both kept and released fish) by registered fishing businesses to the Norwegian Directorate of Fisheries in the COVID-19 years 2020-2021 were only $23 \%$ and $41 \%$ of 2019 catches, respectively. In the current assessment, the WG has taken this into account and reduced the rod and line catches from tourist boats accordingly and kept the other, Norwegian resident, recreational catches unchanged at the 2019 level. This results in 10039 and 10661 tonnes of recreational NCC catch north of $62^{\circ} \mathrm{N}$ in 2020 and 2021 (Table 2.1.1). The proportion of the recreational total caught north of $67^{\circ} \mathrm{N}$ vs. between $62-67^{\circ} \mathrm{N}$ is assumed to be the same in all years.

The total recreational catch numbers-at-age have been upscaled from the estimated catch-at-age proportions in the commercial landings (Tables 2.2.3c and Table 2.3.3).

It is necessary to update the recreational catch with a better estimate as soon as this is available.

### 2.1.1 Revision of catch data

The benchmark assessment (WKBARFAR, ICES 2021a) tested and analysed two major catch data revisions: i) using the ECA model to separate the Norwegian coastal cod and the Northeast Arctic cod in the commercial catches by the structure of the otoliths in commercial samples, and ii) revising the catch in tonnes since 1992 using recommended seasonal product-round fish conversion factors instead of fixed factors for the whole year.

Until 1992, Norway used seasonal conversion factors to convert the weight of "headed-and-gutted" cod to round weight ( 1.6 during winter and 1.4 during the rest of the year). From 1992 onwards, this factor was set to 1.50 for the same product in all Norwegian cod fisheries all year around. From 2000 onwards, this factor was also agreed upon by the Joint Norwegian-Russian Fisheries Commission (JNRFC). From 2000, it hence became constant for all cod fisheries at all times of the year, although there is a larger difference between "headed-and-gutted" weight and round weight in the winter season when at least the Norwegian coastal fisheries for cod are dominated by mature fish with gonads.

Based on a report published by the Norwegian Directorate of Fisheries (Blom, 2015), and summaries of this previously reported to the AFWG as WD 15 in 2017 and as WD 09 in 2020 (Nedreaas, 2017; Fotland and Nedreaas, 2020), ICES advice for NEA cod in 2018 states that "The use of constant conversion factors between round and gutted weight for all seasons and areas introduces a bias to the catch statistics". During the benchmark meeting (WKBARFAR, ICES 2021a) the Norwegian landings of cod by vessels below 28 m in January-April, all gears, were hence corrected by using 1.311 and 1.671 for the products "gutted with head" and "gutted without head", respectively, for each year since 1994.
Catch numbers-at-age are estimated for both stocks of NCC (i.e. northern and southern) by the ECA model. Commercial and recreational total catches have now been calculated back to 1977 for both stocks (Table 2.1.1, WD 03). In addition, catch-at-age in the years 1977-1993 have been estimated for the northern stock (WD 03), though it is not yet included in the assessment model.

### 2.1.2 Catch sampling

The basis for estimating Norwegian coastal cod catches is the total landings of cod from fisheries operating within the Norwegian statistical areas $03,04,05,00,06,07$ (ref. Figure 2.0.1), combined with the catch samplings of these fisheries. Commercial catches of cod are separated into types of cod by the structure of the otoliths in the commercial catch samples. Figure 2.1.2 illustrates the main difference between the two types: The figure and the following text is from Berg et al. (2005):

> Coastal cod has a smaller and more circular first translucent zone than northeast Arctic cod, and the distance between the first and the second translucent zone is larger. The shape of the first translucent zone in northeast Arctic cod is similar to the outer edge of the broken otolith and to the subsequent established translucent zones. This pattern is established at an age of 2 years, and error in differentiating between the two major types does not increase with age since the established growth zones do not change with age.

The precision and accuracy of the separation method for categorizing cod-type was investigated by comparing the results of different otolith reads to the results of genetic analyses, and the investigation determined that the results from the otolith method are high in accuracy (Berg et al., 2005). Nevertheless, in cases with a low percentage misclassification of large catches of pure NEA cod, the catches of coastal cod could be severely overestimated.


Figure 2.1.2. An image of a Norwegian coastal cod otolith (top) and a Northeast Arctic cod otolith (bottom). The two first translucent zones are highlighted. (from Berg et al., 2005).

Since the catches are separated by type of cod by the structure of the otoliths, the numbers of age samples are critical for the estimated catch of coastal cod. Table 2.1.2 shows the sampling of the cod fisheries by quarters, split by NCC and NEAC. The Norwegian sampling program changed in 2010, which led to poor sampling in that year. The sampling in later years gradually improved, and the number of samples (but not the number of otoliths) is now well above the level prior to 2010.

The number of otoliths sampled in 2020 was lower than previous recent years due to reduced access to fish landing sites because of COVID-19, but the proportion of NCC in samples was similar. In 2021, the number of otoliths were nearly back to pre-pandemic levels; a total of 10612 fish were aged in 2021, whereof $35 \%$ were classified as Norwegian coastal cod (Table 2.1.2).

### 2.1.3 Regulations

The Norwegian cod TAC is a combined TAC for both the NEAC stock and NCC stocks. Landings of cod are counted against the overall cod TAC for Norway, where the expected catch of NCC (North and South) is in the order of $10 \%$. The NCC part of this combined quota was set 40000 t in 2003 and earlier years. In 2004, it was set to 20000 t , and in the following years to 21000 t . There are no separate quotas given for the coastal cod for the different groups within the fishing fleet. Catches of coastal cod are thereby not effectively restricted by quotas.

Since the coastal cod is fished under a merged Norwegian coastal cod/Northeast Arctic cod quota, the main objective of these regulations is to move the traditional coastal fishery from areas with high fractions of NCC to areas where the proportion of NEAC is higher. Most regulation measures for NEAC also applies to NCC; minimum catch size, minimum mesh size, maximum
bycatch of undersized fish, closure of areas having high densities of juveniles, and some seasonal and area restrictions. A number of regulations contribute to some protection of NCC, e.g. a ban on trawl fishing inside 6 nautical miles from the baseline and "fjord-lines" that were drawn along the coast to close the fjords for direct cod fishing with vessels larger than 15 metres. For more details about the technical regulations, see ICES (2020).

Table 2.1.1. Left: estimated commercial catches of Norwegian coastal cod North of $67^{\circ} \mathrm{N}$ (NCC North) and between 62$67^{\circ} \mathrm{N}$ (NCC South), and Northeast Arctic cod between $62-67^{\circ} \mathrm{N}$ (NEAC South). Middle: estimated recreational catches of cod north of $67^{\circ} \mathrm{N}$ and between $62-67^{\circ} \mathrm{N}$, all assumed to be coastal cod. Right: Recreational catches of NCC North and South that were sold and included in the commercial catch statistics. Note that an initial unlikely low share of NCC vs. NEAC in the 2001 commercial landings compared to years before/after was replaced by an average of the 2000 and 2002 NCC values.

| Year | Commercial catch (tonnes) |  |  | Recreational catch (tonnes) |  |  | Sold recreational catch included in commercial catch (tonnes)* |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | NCC <br> North | NCC South | NEAC <br> South | NCC <br> North | NCC South | Total | NCC North | NCC South | Total |
| 1977 | 33735 | 9776 | 13831 | 7789 | 4774 | 12563 |  |  |  |
| 1978 | 36413 | 6272 | 8982 | 7855 | 4814 | 12669 |  |  |  |
| 1979 | 31929 | 8194 | 10745 | 7921 | 4855 | 12776 |  |  |  |
| 1980 | 29792 | 8923 | 12948 | 8003 | 4905 | 12909 |  |  |  |
| 1981 | 36161 | 10117 | 16551 | 8054 | 4936 | 12990 |  |  |  |
| 1982 | 33361 | 5883 | 19361 | 8121 | 4977 | 13098 |  |  |  |
| 1983 | 46297 | 5562 | 10616 | 8188 | 5019 | 13207 |  |  |  |
| 1984 | 63305 | 5621 | 9442 | 8256 | 5060 | 13316 |  |  |  |
| 1985 | 56944 | 7424 | 5786 | 8324 | 5102 | 13425 |  |  |  |
| 1986 | 37359 | 3319 | 10742 | 8392 | 5143 | 13535 |  |  |  |
| 1987 | 39630 | 5147 | 7731 | 8424 | 5163 | 13588 |  |  |  |
| 1988 | 55602 | 5153 | 4069 | 8457 | 5183 | 13640 |  |  |  |
| 1989 | 38174 | 6993 | 4277 | 8551 | 5241 | 13792 |  |  |  |
| 1990 | 16707 | 3687 | 8055 | 9035 | 5538 | 14573 |  |  |  |
| 1991 | 22863 | 3823 | 12331 | 9524 | 5837 | 15361 |  |  |  |
| 1992 | 30110 | 3923 | 20156 | 10018 | 6140 | 16157 |  |  |  |
| 1993 | 39681 | 6202 | 22814 | 9181 | 5627 | 14809 |  |  |  |
| 1994 | 52579 | 6381 | 23430 | 9144 | 5556 | 14700 |  |  |  |
| 1995 | 56907 | 8936 | 16981 | 9144 | 5556 | 14700 |  |  |  |
| 1996 | 41820 | 6207 | 13250 | 9020 | 5480 | 14500 |  |  |  |
| 1997 | 46605 | 4746 | 12695 | 9020 | 5480 | 14500 |  |  |  |


| Year | Commercial catch (tonnes) |  |  | Recreational catch (tonnes) |  |  | Sold recreational catch included in commercial catch (tonnes)* |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | NCC North | NCC South | NEAC South | NCC North | NCC South | Total | NCC North | NCC South | Total |
| 1998 | 45462 | 6200 | 9389 | 9082 | 5518 | 14600 |  |  |  |
| 1999 | 38743 | 5522 | 7101 | 8646 | 5254 | 13900 |  |  |  |
| 2000 | 33081 | 5838 | 4329 | 8460 | 5140 | 13600 |  |  |  |
| 2001 | 24470 | 5250 | 3499 | 8335 | 5065 | 13400 |  |  |  |
| 2002 | 32188 | 6937 | 4266 | 8460 | 5140 | 13600 |  |  |  |
| 2003 | 29253 | 8905 | 3943 | 8646 | 5254 | 13900 |  |  |  |
| 2004 | 31198 | 6866 | 3941 | 8335 | 5065 | 13400 |  |  |  |
| 2005 | 30097 | 8005 | 1462 | 8211 | 4989 | 13200 |  |  |  |
| 2006 | 36884 | 8612 | 1175 | 8087 | 4913 | 13000 |  |  |  |
| 2007 | 26200 | 7695 | 2250 | 8087 | 4913 | 13000 |  |  |  |
| 2008 | 27711 | 9889 | 1376 | 7962 | 4838 | 12800 |  |  |  |
| 2009 | 22988 | 7145 | 2474 | 7900 | 4800 | 12700 |  |  |  |
| 2010 | 34804 | 7634 | 2685 | 7900 | 4800 | 12700 |  |  |  |
| 2011 | 27982 | 7128 | 7474 | 7900 | 4800 | 12700 |  |  |  |
| 2012 | 26778 | 8187 | 4942 | 7900 | 4800 | 12700 | 1425 | 239 | 1665 |
| 2013 | 21376 | 5131 | 8395 | 7900 | 4800 | 12700 | 450 | 167 | 617 |
| 2014 | 22750 | 6244 | 6682 | 7900 | 4800 | 12700 | 774 | 229 | 1003 |
| 2015 | 34483 | 5004 | 5424 | 7900 | 4800 | 12700 | 618 | 226 | 844 |
| 2016 | 49503 | 5962 | 2006 | 7900 | 4800 | 12700 | 810 | 332 | 1142 |
| 2017 | 54273 | 4159 | 1242 | 7900 | 4800 | 12700 | 772 | 307 | 1078 |
| 2018 | 34532 | 4436 | 1822 | 7900 | 4800 | 12700 | 889 | 326 | 1215 |
| 2019 | 35861 | 2965 | 1677 | 7900 | 4800 | 12700 | 1603 | 339 | 1943 |
| 2020 | 43133 | 3481 | 987 | 6233 | 3806 | 10039 | 1789 | 347 | 2136 |
| 2021 | 38347 | 3696 | 578 | 6623 | 4039 | 10661 | 565 | 321 | 885 |

*Source: Norwegian Directorate of Fisheries. All reported recreational cod assumed to be coastal cod.

Table 2.1.2. Number of otoliths sampled by quarter from commercial catches. NCC: Norwegian coastal cod. NEAC: Northeast Arctic cod. The table includes all otoliths from the Norwegian catch sampling areas $\mathbf{0}$ and 3-7 (covering both Norwegian coastal cod stocks).

| Year | Quarter 1 |  | Quarter 2 |  | Quarter 3 |  | Quarter 4 |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | NCC | NEAC | NCC | NEAC | NCC | NEAC | NCC | NEAC | NCC | NEAC | \%NCC |
| 1985 | 1451 | 3852 | 777 | 1540 | 1277 | 1767 | 1966 | 730 | 5471 | 7889 | 41 |
| 1986 | 940 | 1594 | 1656 | 2579 | 0 | 0 | 669 | 966 | 3265 | 5139 | 39 |
| 1987 | 1195 | 2322 | 937 | 3051 | 638 | 1108 | 1122 | 1137 | 3892 | 7618 | 34 |
| 1988 | 257 | 546 | 160 | 619 | 87 | 135 | 55 | 44 | 559 | 1344 | 29 |
| 1989 | 556 | 1387 | 72 | 374 | 65 | 501 | 97 | 663 | 790 | 2925 | 21 |
| 1990 | 731 | 2974 | 61 | 689 | 252 | 97 | 265 | 674 | 1309 | 4434 | 23 |
| 1991 | 285 | 1168 | 92 | 561 | 77 | 96 | 279 | 718 | 733 | 2543 | 22 |
| 1992 | 152 | 619 | 281 | 788 | 79 | 82 | 272 | 672 | 784 | 2161 | 27 |
| 1993 | 314 | 1098 | 172 | 1046 | 0 | 0 | 310 | 541 | 796 | 2685 | 23 |
| 1994 | 317 | 1605 | 179 | 923 | 21 | 31 | 126 | 674 | 643 | 3233 | 17 |
| 1995 | 188 | 1591 | 232 | 1682 | 2095 | 1057 | 752 | 1330 | 3267 | 5660 | 37 |
| 1996 | 861 | 5486 | 591 | 1958 | 1784 | 1076 | 958 | 2256 | 4194 | 10776 | 28 |
| 1997 | 1106 | 5429 | 367 | 2494 | 1940 | 894 | 1690 | 1755 | 5103 | 10572 | 33 |
| 1998 | 608 | 4930 | 552 | 1342 | 489 | 1094 | 2999 | 2217 | 4648 | 9583 | 33 |
| 1999 | 1277 | 4702 | 493 | 2379 | 202 | 717 | 961 | 1987 | 2933 | 9785 | 23 |
| 2000 | 1283 | 4918 | 365 | 2112 | 386 | 1295 | 472 | 668 | 2506 | 9993 | 20 |
| 2001 | 1102 | 5091 | 352 | 2295 | 126 | 786 | 432 | 983 | 2012 | 9155 | 18 |
| 2002 | 823 | 5818 | 321 | 1656 | 503 | 831 | 897 | 1355 | 2544 | 9660 | 21 |
| 2003 | 821 | 4197 | 445 | 2850 | 790 | 936 | 1112 | 1286 | 3168 | 9269 | 25 |
| 2004 | 1511 | 7539 | 758 | 2565 | 532 | 685 | 531 | 1317 | 3332 | 12106 | 22 |
| 2005 | 1583 | 6219 | 767 | 4383 | 473 | 258 | 877 | 1258 | 3700 | 12188 | 23 |
| 2006 | 2244 | 5087 | 1329 | 2819 | 590 | 271 | 119 | 71 | 4282 | 8248 | 34 |
| 2007 | 1867 | 5895 | 944 | 2496 | 503 | 648 | 637 | 1163 | 3951 | 10202 | 28 |
| 2008 | 1450 | 4162 | 1116 | 3122 | 626 | 515 | 693 | 999 | 3885 | 8798 | 31 |
| 2009 | 1114 | 5109 | 558 | 2592 | 126 | 253 | 842 | 465 | 2640 | 8419 | 24 |
| 2010 | 736 | 2000 | 572 | 992 | 464 | 195 | 325 | 270 | 2097 | 3457 | 38 |
| 2011 | 643 | 2271 | 789 | 2548 | 412 | 296 | 732 | 443 | 2576 | 5558 | 32 |


| Year | Quarter 1 | Quarter 2 | Quarter 3 | Quarter 4 | Total |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | NCC | NEAC | NCC | NEAC | NCC | NEAC | NCC | NEAC | NCC | NEAC | \%NCC |
| 2012 | 1294 | 6283 | 749 | 1864 | 379 | 85 | 324 | 185 | 2746 | 8417 | 25 |
| 2013 | 966 | 5389 | 832 | 3155 | 216 | 88 | 1115 | 385 | 3129 | 9017 | 26 |
| 2014 | 1019 | 4470 | 869 | 3312 | 338 | 29 | 1060 | 524 | 3286 | 8335 | 28 |
| 2015 | 746 | 7770 | 618 | 3619 | 327 | 354 | 511 | 547 | 2202 | 12290 | 15 |
| 2016 | 2465 | 5581 | 1073 | 2445 | 616 | 207 | 1501 | 727 | 5655 | 8960 | 39 |
| 2017 | 2276 | 4568 | 879 | 2742 | 810 | 151 | 1231 | 475 | 5196 | 7936 | 40 |
| 2018 | 2007 | 4927 | 924 | 1882 | 498 | 104 | 1143 | 435 | 4572 | 7348 | 40 |
| 2019 | 1830 | 4594 | 759 | 1969 | 838 | 260 | 1284 | 445 | 4711 | 7268 | 39 |
| 2020 | 1926 | 3551 | 587 | 1688 | 424 | 85 | 434 | 317 | 3371 | 5641 | 37 |
| 2021 | 1731 | 4060 | 956 | 2219 | 459 | 291 | 580 | 316 | 3726 | 6886 | 35 |
| 21 | 1126 | 4022 | 627 | 2091 | 525 | 467 | 794 | 838 | 3054 | 7446 | 29 |

### 2.2 Northern Norwegian coastal cod

### 2.2.1 Stock status summary

The assessment is based on the decisions of the 2021 WKBARFAR benchmark (ICES 2021a), with updates from the 2022 WKNCCHCR workshop on evaluation of Norwegian coastal cod harvest control rules (ICES 2022). The latter included changes to the assessment model as a follow-up to the benchmark in addition to reference point and HCR evaluations based on a request from the Norwegian managers.

The changes to the model included replacing the acoustic survey index by age with an aggregated biomass index due to uncertain age information, and a change to the Fbar from ages 4-7 to $4-8$ to better reflect fishing pressure on the stock.

The evaluation of reference points led to the conclusion that it was not possible to set a Blim with the certainty required to use it as a basis for estimating reference points in the ICES AR. Therefore, the requested HCRs (based on Blim) could not be considered precautionary. As an alternative, the workshop proposed a constant fishing mortality HCR without a Blim. In this HCR, target F was set to $\mathrm{F}_{0.1}$, a conservative proxy for $\mathrm{F}_{\text {msy. }}$. This HCR was evaluated as precautionary for all stock sizes above Bloss (lowest SSB observed in last c. 20 years) at WKNCCHCR.

The HCR advice was released on 7 June 2022 and the $\mathrm{F}_{0.1} \mathrm{HCR}$ was adopted by the managers shortly thereafter. The revised model and new HCR were used as basis for the advice released on 15 June 2022.

The 2022 assessment shows that SSB declined from a relatively high level at the start of the assessment period (1994) to a low level in 1999. Between 1999-2002, SSB increased, but to a level lower than the one observed at the start of the assessment period. After 2002, SSB stayed at a similar level until 2010, after which it increased to approximately 50000 t lower than the 1994 level. After 2016, there has been a declining trend back towards the level estimated in 2003-2010,
followed by an increase from 2019 to 2020 of approximately 10000 t and a slight decrease from 2020 to 2021 ( 3500 t). Fishing mortality mainly follows the trend in SSB, with highest F in the period with lowest estimated SSB. However, F increased from 2019 to 2020 despite increasing SSB, and decreased from 2020 to 2021 despite a small decrease in SSB. Recruitment-at-age 3 has been relatively stable over time, with somewhat higher values in the early period. There is a weak relationship between SSB and recruitment-at-age 3 despite low fishing pressure on this age.

Stock numbers-at-age 2 in 2020 were the lowest observed in the time-series, and the estimate of this cohort in 2021 is also one of the lowest in the time-series. TSB in 2021 is about 30000 t lower than in 2020 and the lowest observed since 2006-2007. This is mainly driven by the low age 3 numbers, which were also seen in 2006-2007.
The 2021 advice for this stock was revised two times due to errors in data input, with the final quota advice released 15 June 2022 advising that 2022 catches should not exceed 12143 t (commercial and recreational catches combined). Total landings in 2021 were $\sim 45000 \mathrm{t}$, and it is likely that 2022 landings will be at a similar level, exceeding the quota advice.

Further details on the stock assessment procedure can be found in the Stock Annex.

### 2.2.2 The fishery (Table 2.2.1-Table 2.2.4)

Commercial landings of northern Norwegian coastal cod in 2021 were 38347 t , down c. 5000 t from 2020. Of the total landings, $22 \%$ were taken in ICES Division 1.b and the rest in Division 2.a (Table 2.2.1). The highest landings were made in the Norwegian catch reporting area 05 , using Danish seine and gillnet (Table 2.2.2). Compared to 2020, catch proportions were higher in area 05 and lower in areas 03, 04 and 00 . In total, $40 \%$ of the landings were taken in gillnet fisheries and $32 \%$ in Danish seine, while longline/jig made up $16 \%$ of the landings and trawl $12 \%$.
The estimate of recreational catch (fixed at 7900 t) was adjusted in 2020 and 2021 based on reports from tourist businesses to reflect reduced fishing tourism due to the COVID-19 pandemic.
Catch-at-age (commercial + recreational) of ages 3, 4 and 6-10+ were lower compared to 2020, while catches of ages 2 and 5 increased. The total catch in tonnes decreased by 4400 t compared to 2020.

The level of discarding and misreporting from coastal vessels has been investigated for three periods: 2000 and 2002-2003 (WD 14 at 2002 WG), and 2012-2018 (Berg and Nedreaas 2021). The report from the 2000 -investigation concluded that there was both discarding and misreporting by species in 2000. In the gillnet fishery for cod, discarding and misreporting represented approximately $8-10 \%$ relative to reported catch, and $1 / 3$ of this was probably coastal cod. Data from 2002-2003 showed that misreporting in the coastal gillnet fisheries had been reduced significantly since 2000. A recent work by Berg and Nedreaas (2021) estimating discards of cod in the coastal gillnet fisheries during 2012-2018 showed that discarding (as percentage of total catch in weight including discards) decreased from less than $1 \%$ at the beginning of the period to less than $0.5 \%$ during 2016-2018. In weight, this corresponds to a decrease from more than 500 tonnes-per-year to about 180 tonnes-per-year. The reason for discarding seems to be highgrading by size (and price) during the first half of the year, and damaged fish (same size as landed fish) in the second half of the year.

Tourist fishing businesses reporting to the Norwegian Directorate of Fisheries in 2019 showed that about $42 \%$ of the reported rod and line catch was released, and with an assumed mortality of $20 \%$ of the released cod from the boat (see section 2.1 ), this corresponds to about $8 \%$ discards (dead fish) in the rod and line sector of the recreational fishery.

In the stock assessment, discarding is not included in the commercial landings, i.e. commercial catches are assumed equal to landings, but discarding in the rod and line (from boat) sector of the recreational fishery is included in the recreational catch estimate.

### 2.2.3 Survey results

A trawl-acoustic survey for coastal cod along the Norwegian coast from the Russian border to $62^{\circ} \mathrm{N}$ was started in autumn 1995. In 2003, this survey was combined with a saithe survey conducted at the coastal banks and moved from September to October-November (ICES acronym for the combined survey: A6335). Since 2003, the survey therefore covered an extended area and had a more consistent design with fixed bottom trawl stations in addition to trawl hauls set out on acoustic registrations. The seabed along the Norwegian coast is rugged, with sharp drops and peaks over short distances. This makes it difficult to get reliable survey indices both with acoustics and bottom trawl sampling. Acoustics can reach areas where the seabed is too uneven to perform bottom trawling, but species detection and discrimination can be hindered by dead zones and acoustic shadows. Acoustics and bottom trawl data therefore contain both independent and overlapping information.

For the 2021 benchmark, one acoustic and one swept-area index was prepared (WD 06 to AFWG 2021), and it was decided to include them both in the assessment. At the WKNCCHCR 2022 workshop, further quality control of the survey indices were done, resulting in a decision to change the acoustic index from an index by age to an aggregated biomass index (ICES, 2022). This was due to the index by age poorly tracking age classes, particularly after the coastal cod survey merged with the saithe survey, and that the uncertain age 2 estimates from this index had a large influence on model estimates (particularly the shape of the stock-recruit relationship). The swept-area index has generally higher internal consistency and is still included in the model as an age disaggregated index. It should be noted that the uncertainties associated with these indices are rather large and increasing with age.

The survey indices are calculated with the software StoX (Johnsen et al., 2019), developed at the Institute of Marine Research in Norway. Instead of conventional age-length keys, StoX uses an imputation algorithm to assign age information to individuals that have been length measured but not aged. Crucial to coastal cod, the software also imputes other biological information, particularly otolith type, which is used to split the index on NEAC and NCC. The underlying assumption is that the proportion of NCC in length samples are representative of the proportion in the environment. StoX also estimates coefficients of variation using a bootstrap routine. The bootstrapping consists of two parts; resampling of primary sampling units (trawl stations or acoustic transects) with replacement, and the imputation of missing ages by random draw from individuals in the same length group. Primarily, age information is drawn from individuals in the same length group sampled in the same trawl haul. Should there be none, the draw extends to all trawl hauls within the same survey strata, and lastly, to the entire survey area. The CV is the variability resulting from both parts of the bootstrap routine.

The results of the 2021 survey north of $67^{\circ} \mathrm{N}$ are presented in Tables 2.2.5-2.2.12. Box 2.1 below details a decision that was made at AFWG to exclude the acoustic index data point from 2021 (last survey year) due to an inconsistently high value. This decision must be revisited next year when another data point is available.

## Box 2.1. Decision to exclude the acoustic 2021 survey index from the 2022 stock assessment.

The 2021 acoustic survey index came in very high in relation to most previous years and compared to the 2021 swept-area index. To evaluate whether the high index reflected an actual increase in the stock or was due to an error, several checks were made before and during AFWG.

These are summarized below, and further details can be found in the presentation "Survey_data_acoustics_NCCNorth" on the AFWG 2022 SharePoint.

High acoustic values compared to trawl catches were evident both when the acoustic index was expressed as total abundance and as total biomass (Figure B2.1.1), and were seen for all age groups. Looking at internal consistency, the 2021 data point fell outside or on the very edge of the "cloud" of points for nearly all age groups. This was not the case in the swept-area index, where the 2021 data point were more consistent with previous observations.


Figure B2.1.1. Swept-area (blue) and acoustic (orange) abundance indices (left, in millions) and biomass indices (right, in tonnes) for northern coastal cod. The error bars represent $95 \%$ confidence intervals around the mean. The figures show the indices after correction of stock discrimination and rescrutinization of the acoustics.

It was mainly two strata in the northeast that contributed the high acoustic values, called "Østhavet" and "Hjelmsøy Loppa". While the acoustic values were high in these strata, catch rates in the trawl were not particularly high compared to other areas. Østhavet is the strata with the largest proportion of Northeast Arctic cod in the survey, increasing the risk of misclassification between the two stocks. The area also have high abundances of other demersal fish such as saithe and haddock, increasing the risk of misclassification in the acoustic scrutinization process. Therefore, both the stock discrimination by otolith typing and the acoustic scrutinization in this area were closely examined.

In the rescrutinization, some obvious scrutinization errors were found and corrected. These mainly concerned situations where saithe had been misclassified as cod, or too large proportions of cod had been assigned to mixed demersal fish aggregations. This correction reduced the acoustic index somewhat (c. $-20 \%$ for age 3, c. $-10 \%$ on average).

To examine possible misclassification of stock types, a subsample of the otoliths was reread. This resulted in approximately $9 \%$ of otoliths in the sample being reclassified from coastal cod to Northeast Arctic cod. Coastal cod otoliths come in many variations (some specific to local fjords), and this level of misclassification is not unexpected even among experienced readers. Correcting the stock discrimination further reduced the survey index for some ages, but the total change compared to pre-rescrutinization of the acoustics and pre-rereading of otoliths did not exceed c. $-20 \%$ for any age and the acoustic index was still unusually high compared to the swept-area index.

Next, commercial catches from the approximate time and area with high acoustic survey registrations were examined. While cod catches in October-November 2021 were actually lower than in the previous four years, catches of saithe were higher, particularly in November (Figure B2.1.2). The survey covered the area in October, not November, but the catch statistics nevertheless indicate that saithe abundance rather than cod abundance was high in the area within few weeks of the survey.

In summary, the acoustic survey index in 2021 fell outside the internal consistency "cloud" for several ages, also compared to the one other year (2007) when the acoustic index was much higher than the swept-area index. Two strata with high saithe abundances contributed the most to the coastal cod index. The high coastal cod index was caused by large acoustic registrations classified as cod that were not reflected in large survey trawl catches, or in higher commercial catches around the time of the survey. However, commercial catches of saithe were relatively high in the area at the time of the survey and some weeks after. In conclusion, some degree of misclassification between cod and saithe was strongly suspected and the expert group was not confident enough in the data to include the 2021 cod index in the assessment. Excluding it and re-evaluating this decision again when another data point is available next year was considered to be more appropriate, and a precautionary decision given that the high acoustic index had a large influence on model results ( $20 \%$ increase in terminal year SSB when including the 2021 acoustic data point).


Figure B2.1.2. Commercial catch (kilograms) of cod (left panels), saithe (middle panels) and haddock (right panels) in Norwegian statistical fishing area 03 by month in the years 2017-2021. The survey for coastal cod and saithe takes place in months 10 and 11 (October-November). High saithe catches in this period in 2021 are circled in red.

### 2.2.3.1 Indices of abundance and survey mortality (Tables 2.2.5-2.2.8, Figures 2.2.2-2.2.5)

As detailed in Box 2.1, the acoustic survey index in 2021 were much higher than the swept-area index, for all age groups, and the total acoustic biomass index was nearly three times higher than the total swept-area biomass index (Tables 2.2.5 and 2.2.7).

The 2020 age 1 and 2 swept-area abundance indices were particularly low. In 2021, age 2 indices were higher than expected from the low 2020 estimate of the same year class, while the age 3 estimate were consistent with the low index for this year class in the previous year (Table 2.2.7). The age 1 index in 2021 was higher than in 2021, but still among the lowest in the time-series. Note, however, that age 1 cod are too small to be representatively sampled in the survey and that their distribution extends to shallow habitats not accessible to the research vessels. Indices
for the oldest fish (age 10+) were slightly higher in 2021 than in 2020, but the 2020-2021 indices are nevertheless much lower than those seen in 2009-2019 (Table 2.2.7).

The coefficients of variation (CVs) in both indices are generally higher for ages 8 and above where there is less data (Tables 2.2.6 and 2.2.8).

Survey mortality for age 1-2 decreased sharply in 2021 relative to 2020 as a result of the unexpectedly high estimate of age 2 in 2021 relative to age 1 in 2020 (Figure 2.2.5). Survey mortality for age 9-10 showed an opposite trend with a sharp increase. All other ages in the acoustic index had lower survey Z this year due to the high acoustic estimates, while in the swept-area index, the trends were more variable. Generally, internal consistencies are rather low in both survey indices, and consequently, the survey mortality is highly variable between years (Figure 2.2.5).

### 2.2.3.2 Age reading and stock separation (Table 2.2.9)

About 2400 cod otoliths were sampled north of $67^{\circ} \mathrm{N}$ during the 2021 survey, which slightly down from 2500 in 2020 but well above the long-term average (Table 2.2.9). The proportions of NCC at age among those otoliths were higher for older fish (age $6+$ ) compared to the long-term average and the previous year, but within ranges previously observed (Table 2.2.9).

### 2.2.3.3 Length and weights-at-age (Tables 2.2.10-2.2.11, Figure 2.2.6)

There has been a trend of increasing mean length and, particularly, weight at age over the timeseries for most ages, though the trend has levelled off or even reversed in the last few years. Mean lengths-at-age in 2021 were similar to previous years (Table 2.2.10), while mean weights at age decreased compared to 2020 for all ages except age 6 where it was similar (Table 2.2.11). For ages 8 and older the mean lengths and weights show larger variations, probably caused by few fish sampled in some years (Figure 2.2.6).

### 2.2.3.4 Maturity-at-age (Table 2.2.12, Figure 2.2.7)

The fraction of mature fish in the autumn survey (Table 2.2.12) show rather large variation between years. While some of the variation is likely related to variation in stock size and size at age, it may also be partly caused by the difficulty of distinguishing mature and immature cod in autumn. Coastal cod spawn in February-June and many mature individuals are therefore in a resting state at the time of the survey in October-November. The maturity ogive therefor includes spent/resting individuals, which gives an ogive similar to that estimated from a smaller fishery-dependent dataset, collected during the spawning season (ICES 2021a). No large changes in maturity-at-age were observed between 2020 and 2021 (Figure 2.2.7).

### 2.2.4 Data used in the assessment

### 2.2.4.1 Catch numbers-at-age (Table 2.2.3c)

The estimated total catch-at-age ( $2-10+$ ) for the period 1994-2021, including both commercial and recreational catches, is used in the assessment (Table 2.2.3c). Tables 2.2.3a and 2.2.3b show the commercial and recreational catches separately.

### 2.2.4.2 Catch weight-at-age (Table 2.2.4)

Weight-at-age in catches is derived from the commercial sampling and is shown in Table 2.2.4. The same weight-at-age is assumed for recreational and tourist catches. Weight of the plus group is an average for the ages included in the plus group, weighted by abundance-at-age.

### 2.2.4.3 Tuning data (Table 2.2.13)

The acoustic total biomass index (ages $2+$ ) and the swept-area survey index by age ( $2-10+$ ) are used in the assessment (Table 2.2.13). The acoustic index is split in two parts; 1995-2002 and

2003- due to a change in catchability when the saithe and coastal cod surveys were combined in 2003.

### 2.2.4.4 Stock weight-at-age (Table 2.2.14)

The weight-at-age for ages $2-7$ in the stock (Table 2.2.14) is obtained from the Norwegian coastal survey (Table 2.2.11), while catch weight-at-age (Table 2.2.4) is used for ages 8-10+ due to large uncertainty for these ages in survey data (Figure 2.2.6). The survey weights are assumed to be relevant to the weight-at-age in the stock at survey time (October). These weights will, however, overestimate the stock biomass at the start of the year, and in the assessment model, SSB is therefore calculated after applying $80 \%$ of the year's fishing and natural mortality, corresponding to the survey timing.

### 2.2.4.5 Maturity-at-age (Table 2.2.12, Figure 2.2.7)

Annual maturity-at-age observed in the survey is used in the assessment (Table 2.2.12). Maturity of the plus group is an average for the ages included in the plus group, weighted by abundance-at-age.

### 2.2.4.6 Natural mortality (Table 2.2.15, Figure 2.2.8)

In Northeast Arctic cod, cannibalism has been documented to be a significant source of mortality that varies in relation to alternative food and in relation to the abundance of large cod. This might also be the case for the coastal cod (Pedersen and Pope 2003a and b). In the 2005 coastal cod survey 1125 cod stomachs were analysed (Mortensen 2007). The observed average frequency of occurrence of cod in cod stomachs was around $4 \%$. Other important predators on cod in coastal waters are cormorants, harbour porpoises and otters (Anfinsen 2002; Pedersen et al., 2007; Mortensen 2007). Young saithe (ages 2-4) has also been observed to consume post-larvae and 0group cod during summer/autumn (Aas 2007). As detailed data on consumption of coastal cod is lacking, natural mortality in the assessment is assumed dependent on cod size; M is calculated based on stock weight-at-age, following the method by Lorenzen (1996). With this method, M ranges from approximately 0.6 for age 2 to 0.2 for the plus group (Table 2.2.15).

### 2.2.5 Final assessment run

The 2022 assessment was run with the configuration decided upon at the 2021 benchmark (Table 2.2.16), with the necessary updates following decisions from WKNCCHCR (ICES, 2022). These decisions included replacing the acoustic index by age with a total biomass index, including age 8 in the Fbar range (previously F4-7, now F4-8), and reporting recruitment-at-age 3 (model starts at age 2).

The main features of the configuration are: 1) Coupling of fishing mortality states for ages 7-9, 2) Coupling of survey catchability parameters for ages 5-9 in the swept-area index, 3) Separate variance parameter for age 2 in the catch, 4) $\mathrm{AR}(1)$-correlation between ages in the swept-area index, and 5) Recruitment modelled as random walk.

The log-likelihood, number of parameters and AIC of the final run are presented in the table below. There were no problems with model convergence. The "base" model presented below refers to last year's model, which differed from the current model in using acoustic indices by age instead of aggregated biomass indices.

| Model | Log(L) | \#par | AIC |
| :--- | :--- | :--- | :--- |
| Current | -185.44 | 19 | 408.88 |
| base | -180.17 | 37 | 434.33 |

The estimated survey catchabilities at age are presented in Table 2.2.17.

### 2.2.5.1 Model diagnostics (Figure 2.2.9-Figure 2.2.11)

A 5-year retrospective peel indicated no large problems with the estimates of SSB and Fbar, while the model have a low precision in the recruitment (age 2) estimate from 2013 onwards (Figure 2.2.9). The second half of the model period has larger uncertainty as there is an additional survey index (from bottom trawl) that gives generally higher abundance estimates compared to the acoustic index, though this pattern was inversed in the last years. Mohn's rho (average 5-year retrospective bias) was 0.2 for SSB, -0.15 for Fbar, and 0.32 for recruitment. Thus, the model would have overestimated recruitment, particularly from 2013 and onwards, had it been run in previous years.

The process residuals were improved at the benchmark by splitting the acoustic index in two parts and show no concerning patterns (Figure 2.2.10). The one-step-ahead residuals (Figure 2.2.11) were also improved by introducing correlations between ages in the survey indices. Evaluation of this correlation structure should be done at the next benchmark to see if the residuals can be further improved, particularly since the correlation structure has recently been removed from the acoustic index due to the removal of age information.

### 2.2.5.2 Model results (Table 2.2.18-2.2.20, Figure 2.2.1)

Stock numbers-at-age 2 in 2020 were the lowest observed in the time-series, and the estimate of this cohort (age 3 recruits) in 2021 was also one of the lowest in the time-series (Table 2.2.18). SSB decreased with 3500 t from 2020 to 2021, but Fbar also decreased somewhat reflecting the decreased catches of most ages included in the 4-8 Fbar range (Table 2.2.18 and Table 2.2.3c). Fishing mortality for ages $2-5$ in 2021 were slightly higher than in 2019 and 2020, while Fs for ages 6 and above were lower (Table 2.2.19). Abundances of ages 9 and $10+$ in 2021 were low and slightly down from last year (Table 2.2.20). Abundances of ages 2, 5 and 8 increased compared to 2020.

### 2.2.6 Reference points

Reference points were evaluated at the 2021 benchmark (ICES 2021a). The estimated stock-recruitment (age 2) relationship showed increasing recruitment with increasing SSB throughout the model period, and the same pattern resulted from adding 2020 data in the assessment (ICES, 2021d). At the benchmark, Blim was therefore set near the highest SSB observed, based on the reasoning that the lack of plateau in the SSB-recruit relationship indicated that the stock was below full reproductive capacity.

At the 2022 evaluation of reference points and harvest control rules, this decision was re-evaluated by looking closer at assessment data input and historical catch data. An extension of the assessment model back in time indicated that the stock had not experienced severe recruitment failure in the period examined. The stock also appeared to swiftly respond to decreased $F$, which would not be expected from a severely depleted stock. At the same time, simulations demonstrated a high sensitivity of the stock-recruit relationship, and therefore also Blim, to small changes in the assessment model, though the estimates of SSB and F were rather consistent. The workshop therefore concluded that it was not possible to set a Blim with the certainty required to use it as a basis for estimating reference points in the ICES AR. Lacking such reference points, the managers adopted a constant fishing mortality HCR (see below) in 2022.

### 2.2.6.1 Management plan

The Norwegian management plan was implemented in June 2022 and forms the basis for the 2022 advice (ICES, 2022). The target F in the plan is set to $\mathrm{F}_{0.1}$, a conservative proxy for $\mathrm{F}_{\mathrm{msy}}$ that is expected to drive the stock towards and above $B_{m s y}$. This HCR was evaluated as precautionary
for all stock sizes above SSBlowerbound $^{\text {(lowest }}$ SSB observed in last c. 20 years) at WKNCCHCR (ICES, 2022). No adjustment of target F is thus applied as long as SSB is above this value. The HCR requires re-evaluation should the stock fall below $\mathrm{SSB}_{\text {lowerbound. }}$

### 2.2.7 Predictions

### 2.2.7.1 Input data (Tables 2.2.21a-b)

The built-in forecast option in SAM is used for short-term prediction. Since the fishery is not quota regulated, status quo fishing is assumed for the interim year, i.e. same F as in the final year of assessment (Table 2.2.21a). Process noise is included in the prediction (i.e. processNoiseF=FALSE). Averages from the last 5 years of the assessment are used for stock weights, catch weights, maturity, and natural mortality-at-age (Table 2.2.21b). Recruitment is the median resampled from the last 10 years (Table 2.2.21a).

### 2.2.7.2 Catch options for 2021 (Table 2.2.22, Figure 2.2.12)

The ICES advice basis for northern Norwegian coastal cod is the Norwegian management plan. This leads to catch advice of no more than 29347 tonnes in 2023. This catch level is expected to lead to an $8 \%$ increase in SSB relative to SSB estimated for 2022, while the same level of fishing in 2023 as in 2021 is expected to give a $1.5 \%$ decrease in SSB. Zero catch in 2023 is expected to give a $26 \%$ increase in SSB (Table 2.2.21, Figure 2.2.12).

### 2.2.7.3 Comparison of the present and last year's assessments

Due to the updates to the assessment model following WKNCCHCR (ICES, 2022), last year's assessment is not directly comparable to this year's assessment. However, for exploratory reasons both the old and new models were run with 1994-2020 and 1994-2021 data and the results indicated a downwards revision of SSB (and corresponding increase in F) approximately five years back in time when adding 2021 data. For 2020, the downwards revision was approximately 7000 t.

### 2.2.8 Comments to the assessment and the forecast

The assessment model performs rather well despite uncertainties in survey data. However, as both the stock and model are new, the assessment has so far been tested in a limited number of situations. Both the data input and configuration should be improved leading up to the next benchmark. Some areas of research that can potentially reduce uncertainty in the assessment include:

- Examining whether survey index uncertainty can be improved, e.g. by adjusting the survey design or the post-stratification applied to calculate indices.
- Rereading subsamples of otoliths from the first part of the survey (1995-2002) as these readings are expected to be less precise.
- Extending the swept-area index back to 1995.
- Re-examining the coupling of ages applied in the swept-area index observation correlation in SAM.
- Consider the new option of modelling natural mortality, stock weights, proportion mature and catch weights as processes with error (as opposed to fixed values) in SAM


### 2.2.9 Tables and figures

Table 2.2.1. Northern Norwegian coastal cod. Total commercial catch ( $t$ ) by fishing areas in the last two years.

| Year | $\mathbf{0 3}$ | $\mathbf{0 4}$ | $\mathbf{0 5}$ | $\mathbf{0 0}$ | Total in Di- <br> vision 1.b <br> (NOR area <br> 03) | Total in Di- <br> vision 2.a <br> (NOR areas <br> 04+00+05) | Total |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2020 | 12245 | 12393 | 10832 | 7652 | 12245 | 30877 | 43122 |
| 2021 | 8244 | 6548 | 18542 | 4640 | 8244 | 29730 | 37974 |

Table 2.2.2. Commercial catch of northern Norwegian coastal cod ( $t$ ) in 2021 by gear and Norwegian statistical fishing area.

| Year | 2021 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Area | 03 | 04 | 05 | 00 | Total north of $67^{\circ} \mathrm{N}$ | \% by gear |
| Gillnet | 1007 | 2985 | 7667 | 3352 | 15011 | 40 |
| L.line/Jig | 3578 | 621 | 1436 | 382 | 6017 | 16 |
| Danish seine | 2568 | 1633 | 7178 | 892 | 12272 | 32 |
| Trawl | 1083 | 1303 | 2258 | 14 | 4658 | 12 |
| Others* | 7.2 | 6.1 | 2.8 | - | 16 | <0.1 |
| Total | 8244 | 6548 | 18542 | 4640 | 37974 |  |

Table 2.2.3a. Northern Norwegian coastal cod. Estimated commercial landings in numbers ('000) at-age and total tonnes by year.

|  | Age |  | $\mathbf{2}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1994 | 11 | 98 | 978 | 4394 | 3760 | 2756 | 1119 | 304 | 675 | 52579 |
| 1995 | 21 | 228 | 814 | 2743 | 4796 | 3164 | 1815 | 943 | 612 | 56907 |
| 1996 | 41 | 768 | 1415 | 2035 | 3130 | 3086 | 1210 | 542 | 584 | 41820 |
| 1997 | 57 | 1111 | 2106 | 1956 | 2344 | 2721 | 1856 | 565 | 746 | 46605 |
| 1998 | 436 | 1631 | 6433 | 4391 | 2784 | 835 | 779 | 377 | 393 | 45462 |
| 1999 | 79 | 912 | 3395 | 4938 | 2037 | 783 | 527 | 394 | 425 | 38743 |
| 2000 | 30 | 534 | 2549 | 3925 | 2240 | 826 | 376 | 112 | 273 | 33081 |
| 2001 | 10 | 330 | 1863 | 2242 | 1641 | 961 | 305 | 104 | 493 | 24470 |
| 2002 | 42 | 308 | 1551 | 2585 | 2391 | 1057 | 630 | 183 | 363 | 32188 |
| 2003 | 120 | 350 | 952 | 1859 | 2173 | 1206 | 582 | 308 | 252 | 29253 |


| Year | Age |  |  |  |  |  |  |  |  | Tonnes <br> Landed |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |  |
| 2004 | 23 | 179 | 1067 | 1520 | 2189 | 1570 | 784 | 328 | 371 | 31198 |
| 2005 | 13 | 241 | 924 | 1984 | 2003 | 1463 | 716 | 255 | 345 | 30097 |
| 2006 | 23 | 222 | 1276 | 1977 | 2619 | 1735 | 1017 | 402 | 396 | 36884 |
| 2007 | 36 | 376 | 1198 | 1667 | 1327 | 1088 | 477 | 277 | 279 | 26200 |
| 2008 | 63 | 387 | 997 | 1909 | 1549 | 1005 | 576 | 278 | 287 | 27711 |
| 2009 | 21 | 456 | 667 | 1177 | 1194 | 812 | 419 | 431 | 211 | 22988 |
| 2010 | 29 | 530 | 754 | 2832 | 1947 | 1055 | 528 | 283 | 857 | 34804 |
| 2011 | 65 | 465 | 1209 | 1318 | 1239 | 1081 | 568 | 343 | 583 | 27982 |
| 2012 | 374 | 1017 | 1126 | 1118 | 1287 | 760 | 364 | 177 | 596 | 26778 |
| 2013 | 131 | 503 | 1024 | 1038 | 909 | 704 | 478 | 219 | 340 | 21376 |
| 2014 | 88 | 505 | 824 | 1258 | 839 | 676 | 523 | 297 | 397 | 22750 |
| 2015 | 331 | 1106 | 1411 | 1251 | 1700 | 1040 | 639 | 437 | 873 | 34483 |
| 2016 | 75 | 937 | 1988 | 1582 | 1723 | 2119 | 1174 | 640 | 1073 | 49503 |
| 2017 | 846 | 1577 | 2071 | 2323 | 2087 | 1491 | 1331 | 700 | 903 | 54273 |
| 2018 | 171 | 563 | 1465 | 1634 | 1525 | 1416 | 747 | 518 | 497 | 34532 |
| 2019 | 49 | 953 | 1299 | 1776 | 1585 | 1260 | 985 | 318 | 519 | 35861 |
| 2020 | 40 | 534 | 2205 | 2116 | 2538 | 1615 | 906 | 354 | 309 | 43133 |
| 2021 | 162 | 408 | 1914 | 3023 | 1801 | 1270 | 644 | 177 | 251 | 38347 |

Table 2.2.3b. Northern Norwegian coastal cod. Estimated catch number ('000) at-age in recreational and tourist catches.

| Year | Age |  |  |  |  |  |  |  |  | Tonnes <br> landed |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |  |
| 1994 | 2 | 17 | 170 | 764 | 654 | 479 | 195 | 53 | 117 | 9144 |
| 1995 | 3 | 37 | 131 | 441 | 771 | 508 | 292 | 151 | 98 | 9144 |
| 1996 | 9 | 166 | 305 | 439 | 675 | 666 | 261 | 117 | 126 | 9020 |
| 1997 | 11 | 215 | 408 | 378 | 454 | 527 | 359 | 109 | 144 | 9020 |
| 1998 | 87 | 326 | 1285 | 877 | 556 | 167 | 156 | 75 | 78 | 9082 |
| 1999 | 18 | 204 | 758 | 1102 | 455 | 175 | 118 | 88 | 95 | 8646 |
| 2000 | 8 | 136 | 652 | 1004 | 573 | 211 | 96 | 29 | 70 | 8460 |


| Year | Age |  |  |  |  |  |  |  |  | Tonnes <br> landed |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |  |
| 2001 | 3 | 112 | 635 | 764 | 559 | 327 | 104 | 36 | 168 | 8335 |
| 2002 | 11 | 81 | 408 | 679 | 628 | 278 | 166 | 48 | 95 | 8460 |
| 2003 | 36 | 104 | 281 | 549 | 642 | 356 | 172 | 91 | 74 | 8646 |
| 2004 | 6 | 48 | 285 | 406 | 585 | 419 | 209 | 88 | 99 | 8335 |
| 2005 | 4 | 66 | 252 | 541 | 546 | 399 | 195 | 69 | 94 | 8211 |
| 2006 | 5 | 49 | 280 | 433 | 574 | 380 | 223 | 88 | 87 | 8087 |
| 2007 | 11 | 116 | 370 | 514 | 410 | 336 | 147 | 85 | 86 | 8087 |
| 2008 | 18 | 111 | 287 | 549 | 445 | 289 | 165 | 80 | 82 | 7962 |
| 2009 | 7 | 157 | 229 | 405 | 410 | 279 | 144 | 148 | 73 | 7900 |
| 2010 | 7 | 120 | 171 | 643 | 442 | 240 | 120 | 64 | 194 | 7900 |
| 2011 | 18 | 131 | 341 | 372 | 350 | 305 | 160 | 97 | 165 | 7900 |
| 2012 | 110 | 300 | 332 | 330 | 380 | 224 | 107 | 52 | 176 | 7900 |
| 2013 | 48 | 186 | 379 | 383 | 336 | 260 | 177 | 81 | 126 | 7900 |
| 2014 | 31 | 175 | 286 | 437 | 291 | 235 | 181 | 103 | 138 | 7900 |
| 2015 | 76 | 253 | 323 | 287 | 389 | 238 | 146 | 100 | 200 | 7900 |
| 2016 | 12 | 150 | 317 | 253 | 275 | 338 | 187 | 102 | 171 | 7900 |
| 2017 | 123 | 230 | 301 | 338 | 304 | 217 | 194 | 102 | 131 | 7900 |
| 2018 | 39 | 129 | 335 | 374 | 349 | 324 | 171 | 119 | 114 | 7900 |
| 2019 | 11 | 210 | 286 | 391 | 349 | 278 | 217 | 70 | 114 | 7900 |
| 2020 | 6 | 77 | 319 | 306 | 367 | 233 | 131 | 51 | 45 | 6233 |
| 2021 | 28 | 71 | 331 | 522 | 311 | 219 | 111 | 31 | 43 | 6623 |

Table 2.2.3c. Northern Norwegian coastal cod. Total estimated catch number ('000) at age, including recreational and tourist catches.

|  | Age |  |  |  |  |  |  | Tonnes |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ | landed |
| 1994 | 13 | 115 | 1148 | 5158 | 4414 | 3235 | 1313 | 356 | 793 | 61723 |
| 1995 | 24 | 264 | 945 | 3183 | 5567 | 3672 | 2106 | 1094 | 711 | 66051 |
| 1996 | 50 | 934 | 1720 | 2473 | 3805 | 3752 | 1471 | 659 | 709 | 50840 |
| 1997 | 68 | 1326 | 2514 | 2334 | 2797 | 3248 | 2215 | 674 | 890 | 55624 |


|  | Age |  |  |  |  |  |  |  |  | Tonnes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | landed |
| 1998 | 523 | 1957 | 7718 | 5268 | 3341 | 1002 | 935 | 452 | 471 | 54544 |
| 1999 | 97 | 1116 | 4152 | 6040 | 2492 | 957 | 644 | 482 | 520 | 47390 |
| 2000 | 38 | 670 | 3201 | 4929 | 2812 | 1037 | 472 | 141 | 342 | 41541 |
| 2001 | 13 | 442 | 2497 | 3006 | 2199 | 1288 | 409 | 140 | 661 | 32806 |
| 2002 | 53 | 389 | 1959 | 3265 | 3019 | 1335 | 796 | 231 | 459 | 40648 |
| 2003 | 156 | 454 | 1234 | 2408 | 2815 | 1562 | 754 | 399 | 326 | 37900 |
| 2004 | 30 | 227 | 1352 | 1926 | 2774 | 1989 | 993 | 415 | 470 | 39533 |
| 2005 | 17 | 307 | 1176 | 2525 | 2550 | 1862 | 911 | 324 | 440 | 38308 |
| 2006 | 28 | 271 | 1556 | 2410 | 3193 | 2115 | 1240 | 490 | 482 | 44970 |
| 2007 | 47 | 492 | 1567 | 2181 | 1737 | 1423 | 624 | 362 | 365 | 34287 |
| 2008 | 81 | 498 | 1284 | 2458 | 1994 | 1294 | 741 | 358 | 369 | 35674 |
| 2009 | 28 | 612 | 896 | 1582 | 1605 | 1091 | 563 | 579 | 284 | 30888 |
| 2010 | 35 | 651 | 925 | 3474 | 2388 | 1295 | 647 | 347 | 1051 | 42704 |
| 2011 | 83 | 597 | 1550 | 1690 | 1588 | 1386 | 728 | 440 | 747 | 35882 |
| 2012 | 484 | 1317 | 1458 | 1447 | 1666 | 984 | 471 | 229 | 772 | 34678 |
| 2013 | 179 | 689 | 1403 | 1421 | 1245 | 965 | 655 | 300 | 466 | 29276 |
| 2014 | 119 | 680 | 1110 | 1695 | 1130 | 911 | 704 | 400 | 534 | 30650 |
| 2015 | 407 | 1360 | 1734 | 1537 | 2089 | 1278 | 785 | 537 | 1072 | 42383 |
| 2016 | 86 | 1086 | 2305 | 1835 | 1998 | 2458 | 1362 | 743 | 1244 | 57403 |
| 2017 | 969 | 1806 | 2373 | 2661 | 2391 | 1707 | 1525 | 802 | 1035 | 62173 |
| 2018 | 210 | 691 | 1800 | 2007 | 1873 | 1740 | 918 | 637 | 611 | 42432 |
| 2019 | 60 | 1163 | 1585 | 2167 | 1934 | 1537 | 1202 | 387 | 633 | 43761 |
| 2020 | 45 | 612 | 2524 | 2422 | 2905 | 1849 | 1037 | 405 | 353 | 49366 |
| 2021 | 190 | 479 | 2245 | 3545 | 2112 | 1490 | 755 | 207 | 294 | 44970 |

Table 2.2.4. Northern Norwegian coastal cod. Mean catch weight at age (kg).

|  | Age |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| 1994 | 0.910 | 1.422 | 1.987 | 2.649 | 3.479 | 4.343 | 5.245 | 6.487 | 8.825 |


| Year | Age |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 2.2.5. Northern Norwegian coastal cod. Acoustic abundance indices by age (in thousands) and total biomass (t) from the Coastal survey (A6335). The split between coastal cod and Northeast Arctic cod is uncertain for age 1.

|  | Age |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | Sum | Biomass |
| 1995 | 26495 | 8774 | 4974 | 6382 | 6440 | 4373 | 1309 | 532 | 319 | 132 | 59729 | 55126 |
| 1996 | 17580 | 9025 | 8592 | 4576 | 5306 | 2723 | 1022 | 213 | 32 | 24 | 49093 | 39263 |
| 1997 | 16567 | 15358 | 16930 | 7710 | 4484 | 2316 | 716 | 328 | 59 | 33 | 64502 | 45756 |
| 1998 | 8360 | 6757 | 8524 | 8261 | 3717 | 1530 | 700 | 102 | 122 | 45 | 38118 | 39474 |
| 1999 | 2494 | 3486 | 3387 | 2788 | 2498 | 751 | 172 | 30 | 22 | 20 | 15648 | 16167 |
| 2000 | 5028 | 7439 | 5831 | 3939 | 3853 | 2825 | 622 | 258 | 71 | 32 | 29899 | 35602 |
| 2001 | 2711 | 4551 | 4246 | 3776 | 2184 | 1499 | 974 | 149 | 29 | 93 | 20211 | 27250 |
| 2002 | 1188 | 2071 | 2532 | 2926 | 2075 | 970 | 596 | 293 | 106 | 124 | 12882 | 21203 |
| 2003 | 3276 | 2168 | 3026 | 3303 | 1838 | 1519 | 651 | 364 | 190 | 69 | 16403 | 23978 |
| 2004 | 3046 | 2643 | 2819 | 2589 | 1686 | 1094 | 371 | 213 | 104 | 72 | 14639 | 18237 |
| 2005 | 904 | 1201 | 2228 | 1816 | 1490 | 843 | 234 | 233 | 127 | 79 | 9156 | 14690 |
| 2006 | 4981 | 1836 | 2587 | 2210 | 1453 | 1612 | 1046 | 130 | 89 | 27 | 15970 | 22116 |
| 2007 | 2458 | 3037 | 2778 | 3794 | 2437 | 1632 | 1215 | 441 | 120 | 41 | 17952 | 33314 |
| 2008 | 2344 | 1739 | 1684 | 1511 | 985 | 761 | 399 | 225 | 97 | 74 | 9821 | 15491 |
| 2009 | 3907 | 1502 | 2084 | 2596 | 1373 | 605 | 386 | 378 | 140 | 64 | 13035 | 18716 |
| 2010 | 5509 | 2503 | 2853 | 2240 | 1679 | 583 | 309 | 432 | 229 | 195 | 16531 | 21966 |
| 2011 | 2104 | 2542 | 1869 | 2372 | 1469 | 1215 | 394 | 278 | 137 | 150 | 12529 | 23115 |
| 2012 | 3561 | 2170 | 3546 | 1832 | 1154 | 791 | 503 | 254 | 107 | 224 | 14142 | 20913 |
| 2013 | 4694 | 3084 | 1597 | 1770 | 1287 | 838 | 657 | 430 | 216 | 252 | 14825 | 21105 |
| 2014 | 6030 | 4171 | 3066 | 2137 | 2904 | 1609 | 1151 | 429 | 462 | 326 | 22286 | 37127 |
| 2015 | 3421 | 3122 | 2465 | 1802 | 1017 | 1128 | 477 | 363 | 303 | 265 | 14362 | 23144 |
| 2016 | 2921 | 3341 | 3667 | 2349 | 2308 | 841 | 669 | 452 | 222 | 308 | 17078 | 30763 |
| 2017 | 1018 | 3289 | 3202 | 2335 | 1764 | 1122 | 450 | 256 | 181 | 183 | 13800 | 25998 |
| 2018 | 4977 | 2847 | 1837 | 2376 | 1246 | 946 | 494 | 246 | 136 | 169 | 15274 | 22602 |
| 2019 | 2607 | 2992 | 3724 | 2221 | 2149 | 1272 | 656 | 212 | 262 | 266 | 16360 | 29992 |
| 2020 | 481 | 1618 | 3378 | 3739 | 2025 | 890 | 522 | 319 | 85 | 125 | 12701 | 26878 |
| 2021 | 3735 | 4806 | 3597 | 4923 | 3935 | 2102 | 1143 | 747 | 231 | 243 | 21727 | 43863 |

Table 2.2.6. Northern Norwegian coastal cod. Acoustic abundance index coefficient of variation (CV, in \%) by age.

| Year | Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1995 | 17 | 13 | 9 | 12 | 14 | 21 | 19 | 40 | 51 | 41 |
| 1996 | 20 | 11 | 15 | 17 | 14 | 26 | 54 | 39 | 52 | 156 |
| 1997 | 24 | 25 | 16 | 16 | 14 | 25 | 26 | 47 | 90 | 81 |
| 1998 | 26 | 19 | 12 | 16 | 16 | 31 | 69 | 40 | 87 | 104 |
| 1999 | 24 | 10 | 11 | 20 | 17 | 23 | 19 | 47 | 40 | 92 |
| 2000 | 14 | 16 | 12 | 10 | 9 | 10 | 15 | 29 | 49 | 89 |
| 2001 | 18 | 31 | 18 | 16 | 19 | 18 | 21 | 41 | 72 | 69 |
| 2002 | 25 | 17 | 21 | 16 | 14 | 15 | 23 | 36 | 72 | 67 |
| 2003 | 27 | 26 | 14 | 14 | 14 | 16 | 18 | 22 | 26 | 35 |
| 2004 | 17 | 15 | 14 | 12 | 13 | 17 | 17 | 25 | 69 | 33 |
| 2005 | 18 | 23 | 18 | 10 | 14 | 20 | 23 | 30 | 40 | 61 |
| 2006 | 108 | 68 | 15 | 14 | 15 | 27 | 22 | 23 | 31 |  |
| 2007 | 21 | 20 | 19 | 15 | 16 | 16 | 21 | 31 | 45 | 97 |
| 2008 | 24 | 19 | 14 | 13 | 12 | 14 | 20 | 24 | 39 | 37 |
| 2009 | 22 | 20 | 15 | 12 | 17 | 14 | 18 | 19 | 31 | 25 |
| 2010 | 41 | 18 | 16 | 13 | 12 | 22 | 22 | 22 | 21 | 21 |
| 2011 | 22 | 17 | 16 | 15 | 15 | 15 | 27 | 21 | 19 | 35 |
| 2012 | 20 | 20 | 13 | 14 | 15 | 11 | 19 | 16 | 24 | 18 |
| 2013 | 14 | 16 | 14 | 15 | 14 | 13 | 17 | 20 | 31 | 37 |
| 2014 | 16 | 19 | 12 | 15 | 15 | 13 | 15 | 14 | 23 | 43 |
| 2015 | 21 | 16 | 11 | 10 | 12 | 12 | 16 | 16 | 16 | 27 |
| 2016 | 29 | 15 | 10 | 8 | 11 | 16 | 17 | 21 | 39 | 31 |
| 2017 | 34 | 16 | 12 | 16 | 14 | 18 | 23 | 28 | 43 | 25 |
| 2018 | 18 | 17 | 17 | 16 | 18 | 9 | 18 | 60 | 20 | 35 |
| 2019 | 18 | 20 | 15 | 13 | 12 | 15 | 18 | 28 | 33 | 35 |
| 2020 | 28 | 16 | 16 | 12 | 14 | 14 | 19 | 27 | 39 | 57 |
| 2021 | 18 | 16 | 13 | 12 | 13 | 13 | 16 | 19 | 32 | 45 |

Table 2.2.7. Northern Norwegian coastal cod. Swept-area abundance indices by age (in thousands) and total biomass (t) from the Coastal survey (A6335). The split between coastal cod and Northeast Arctic cod is uncertain for age 1.

| Age |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | Sum | Biomass |
| 2003 | 5254 | 3268 | 3763 | 4521 | 2700 | 2319 | 863 | 489 | 220 | 69 | 23467 | 33861 |
| 2004 | 2837 | 2201 | 2396 | 2602 | 1463 | 722 | 359 | 181 | 46 | 63 | 12868 | 15980 |
| 2005 | 665 | 1042 | 1988 | 1478 | 1268 | 746 | 157 | 107 | 68 | 54 | 7574 | 11379 |
| 2006 | 1802 | 2156 | 2623 | 2946 | 1554 | 1026 | 941 | 171 | 107 | 23 | 13349 | 22526 |
| 2007 | 446 | 911 | 853 | 1071 | 789 | 465 | 394 | 114 | 75 | 29 | 5146 | 11943 |
| 2008 | 2463 | 1822 | 2795 | 1883 | 1419 | 1145 | 580 | 348 | 161 | 94 | 12710 | 23090 |
| 2009 | 6642 | 2251 | 3570 | 3716 | 1584 | 868 | 712 | 466 | 204 | 160 | 20172 | 24986 |
| 2010 | 7412 | 2353 | 3268 | 3385 | 2397 | 784 | 383 | 733 | 317 | 328 | 21360 | 29875 |
| 2011 | 2322 | 3471 | 2498 | 2866 | 2095 | 1445 | 292 | 315 | 213 | 310 | 15827 | 27845 |
| 2012 | 4299 | 3218 | 4485 | 2784 | 1537 | 1042 | 930 | 411 | 200 | 346 | 19251 | 28587 |
| 2013 | 6382 | 4101 | 1706 | 2666 | 1887 | 1575 | 890 | 578 | 297 | 419 | 20502 | 32875 |
| 2014 | 5696 | 5448 | 4026 | 3034 | 3521 | 2016 | 1388 | 465 | 364 | 337 | 26296 | 43823 |
| 2015 | 4298 | 4733 | 4154 | 3727 | 2068 | 1818 | 902 | 506 | 397 | 222 | 22827 | 40385 |
| 2016 | 3944 | 4433 | 4522 | 2610 | 1995 | 746 | 735 | 413 | 203 | 210 | 19810 | 31320 |
| 2017 | 768 | 2891 | 2407 | 1563 | 1151 | 715 | 308 | 200 | 147 | 157 | 10308 | 18682 |
| 2018 | 4070 | 3197 | 1916 | 1879 | 1049 | 748 | 323 | 183 | 128 | 168 | 13661 | 18815 |
| 2019 | 2234 | 2114 | 2470 | 1508 | 1460 | 839 | 490 | 148 | 129 | 211 | 11601 | 19974 |
| 2020 | 560 | 1670 | 2599 | 2416 | 1188 | 611 | 291 | 177 | 49 | 72 | 9072 | 16780 |
| 2021 | 1412 | 2531 | 1367 | 1589 | 1367 | 732 | 289 | 239 | 82 | 81 | 8277 | 14699 |

Table 2.2.8. Northern Norwegian coastal cod. Swept-area abundance index coefficient of variation (CV, in \%).
$\left.\begin{array}{lllllllllll}\hline & \text { Age } & \mathbf{1} & \mathbf{2} & \mathbf{3} & \mathbf{4} & \mathbf{5} & \mathbf{6} & \mathbf{7} & 8 & 9\end{array}\right\} 10$

| Year | Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 2008 | 15 | 26 | 21 | 13 | 11 | 17 | 15 | 20 | 37 | 36 |
| 2009 | 16 | 16 | 18 | 14 | 14 | 18 | 15 | 21 | 24 | 27 |
| 2010 | 9 | 16 | 19 | 21 | 16 | 18 | 26 | 27 | 21 | 16 |
| 2011 | 20 | 24 | 27 | 19 | 23 | 17 | 25 | 23 | 23 | 35 |
| 2012 | 9 | 37 | 24 | 13 | 12 | 13 | 16 | 17 | 23 | 20 |
| 2013 | 14 | 17 | 15 | 23 | 20 | 21 | 16 | 17 | 31 | 38 |
| 2014 | 17 | 30 | 17 | 16 | 17 | 26 | 14 | 15 | 22 | 39 |
| 2015 | 19 | 17 | 18 | 27 | 29 | 22 | 30 | 19 | 19 | 23 |
| 2016 | 20 | 13 | 13 | 10 | 9 | 13 | 16 | 24 | 20 | 20 |
| 2017 | 30 | 20 | 17 | 15 | 9 | 17 | 18 | 39 | 30 | 27 |
| 2018 | 15 | 19 | 16 | 15 | 12 | 11 | 15 | 27 | 19 | 19 |
| 2019 | 15 | 16 | 16 | 13 | 10 | 9 | 12 | 17 | 25 | 30 |
| 2020 | 28 | 14 | 16 | 13 | 13 | 16 | 15 | 19 | 31 | 41 |
| 2021 | 19 | 19 | 21 | 16 | 21 | 18 | 13 | 16 | 25 | 35 |

Table 2.2.9. Proportion Norwegian coastal cod by age among all aged cod in the Norwegian coastal survey north of $67^{\circ} \mathrm{N}$. The split between coastal cod and Northeast Arctic cod is uncertain for age 1.

| Year | Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | Total <br> number <br> aged |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1995 | 0.92 | 0.98 | 0.94 | 0.86 | 0.60 | 0.54 | 0.60 | 0.56 | 0.90 | 1.00 | 2236 |
| 1996 | 0.87 | 0.96 | 0.89 | 0.81 | 0.68 | 0.60 | 0.41 | 0.42 | 0.27 | 0.25 | 2289 |
| 1997 | 0.88 | 0.91 | 0.86 | 0.79 | 0.71 | 0.64 | 0.43 | 0.26 | 0.14 | 0.75 | 1774 |
| 1998 | 0.89 | 0.85 | 0.80 | 0.74 | 0.80 | 0.69 | 0.50 | 0.34 | 0.32 | 0.60 | 2639 |
| 1999 | 0.88 | 0.90 | 0.81 | 0.64 | 0.58 | 0.62 | 0.52 | 0.20 | 0.22 | 0.13 | 2911 |
| 2000 | 0.97 | 0.91 | 0.85 | 0.76 | 0.65 | 0.57 | 0.42 | 0.46 | 0.18 | 0.08 | 4325 |
| 2001 | 0.88 | 0.84 | 0.74 | 0.71 | 0.65 | 0.55 | 0.45 | 0.41 | 0.21 | 0.31 | 3282 |
| 2002 | 0.84 | 0.86 | 0.78 | 0.68 | 0.54 | 0.34 | 0.32 | 0.29 | 0.10 | 0.18 | 2265 |
| 2003 | 0.90 | 0.94 | 0.87 | 0.88 | 0.85 | 0.75 | 0.65 | 0.59 | 0.52 | 0.57 | 2953 |
| 2004 | 0.86 | 0.76 | 0.77 | 0.59 | 0.67 | 0.57 | 0.60 | 0.49 | 0.41 | 0.63 | 2287 |
| 005 | 0.81 | 0.76 | 0.76 | 0.65 | 0.59 | 0.48 | 0.56 | 0.50 | 0.44 | 1209 |  |


| Year | Age |  |  |  |  |  |  |  |  |  | Total number aged |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |  |
| 2006 | 0.98 | 0.93 | 0.94 | 0.83 | 0.75 | 0.71 | 0.68 | 0.68 | 0.57 | 0.00 | 1419 |
| 2007 | 0.73 | 0.81 | 0.76 | 0.82 | 0.73 | 0.61 | 0.69 | 0.43 | 0.83 | 0.50 | 1021 |
| 2008 | 0.99 | 0.99 | 0.99 | 0.83 | 0.89 | 0.84 | 0.78 | 0.67 | 0.94 | 0.75 | 1448 |
| 2009 | 0.94 | 0.94 | 0.83 | 0.69 | 0.55 | 0.58 | 0.75 | 0.76 | 0.73 | 0.72 | 1944 |
| 2010 | 0.94 | 0.94 | 0.89 | 0.75 | 0.66 | 0.49 | 0.60 | 0.86 | 0.90 | 0.97 | 2093 |
| 2011 | 0.90 | 0.93 | 0.91 | 0.89 | 0.77 | 0.66 | 0.52 | 0.73 | 0.80 | 0.83 | 1577 |
| 2012 | 0.94 | 0.89 | 0.90 | 0.82 | 0.83 | 0.73 | 0.71 | 0.61 | 0.88 | 0.84 | 1831 |
| 2013 | 0.93 | 0.94 | 0.88 | 0.77 | 0.79 | 0.83 | 0.74 | 0.79 | 0.73 | 1.00 | 1920 |
| 2014 | 0.99 | 0.99 | 0.99 | 0.96 | 0.93 | 0.90 | 0.93 | 0.87 | 0.87 | 0.88 | 2361 |
| 2015 | 0.89 | 0.93 | 0.89 | 0.86 | 0.75 | 0.73 | 0.65 | 0.73 | 0.82 | 0.96 | 1859 |
| 2016 | 0.99 | 0.98 | 0.99 | 0.90 | 0.84 | 0.69 | 0.75 | 0.80 | 0.71 | 0.83 | 2041 |
| 2017 | 1.00 | 0.98 | 0.95 | 0.93 | 0.86 | 0.74 | 0.78 | 0.68 | 0.84 | 1.00 | 1732 |
| 2018 | 0.99 | 0.97 | 0.91 | 0.86 | 0.88 | 0.82 | 0.72 | 0.68 | 0.87 | 0.90 | 2395 |
| 2019 | 0.95 | 0.99 | 0.97 | 0.88 | 0.84 | 0.83 | 0.84 | 0.76 | 0.82 | 0.91 | 2107 |
| 2020 | 1.00 | 0.84 | 0.85 | 0.81 | 0.71 | 0.70 | 0.75 | 0.83 | 0.78 | 0.64 | 2504 |
| 2021 | 0.97 | 0.93 | 0.85 | 0.84 | 0.76 | 0.79 | 0.81 | 0.83 | 0.84 | 0.83 | 2405 |
| Average 95-21 | 0.91 | 0.91 | 0.87 | 0.80 | 0.74 | 0.67 | 0.63 | 0.60 | 0.62 | 0.65 | 2179 |

Table 2.2.10. Northern Norwegian coastal cod. Mean length (cm) at-age from Coastal survey data (A6335). Mean lengths of ages $>\mathbf{7}$ have higher uncertainty due to few samples. The split between coastal cod and Northeast Arctic cod is uncertain for age 1. For the plus group, mean length is the average mean length for ages 10+, weighted by abundance-at-age.

| Age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 1995 | 18.9 | 31.4 | 42.1 | 51.8 | 58.8 | 64.3 | 77.5 | 82.4 | 87.1 | 105.7 |
| 1996 | 16.7 | 28.3 | 41.3 | 51.9 | 58.1 | 65.2 | 74.8 | 86.7 | 99.6 | 115.0 |
| 1997 | 16.6 | 29.6 | 40.7 | 52.0 | 58.1 | 66.9 | 66.8 | 68.6 | 102.0 | 92.0 |
| 1998 | 17.8 | 30.3 | 44.0 | 52.0 | 60.3 | 67.8 | 74.9 | 82.2 | 83.8 | 107.8 |
| 1999 | 19.4 | 31.2 | 44.1 | 54.1 | 58.7 | 65.4 | 74.0 | 89.0 | 88.2 | 72.7 |
| 2000 | 20.0 | 32.5 | 44.0 | 54.0 | 61.4 | 64.5 | 73.8 | 81.9 | 80.3 | 90.3 |
| 2001 | 20.0 | 33.7 | 45.7 | 55.4 | 61.1 | 65.2 | 67.6 | 76.1 | 87.2 | 109.7 |


| Year | Age | 1 |  |  | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | 7 | 8 | 9 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 2.2.11. Northern Norwegian coastal cod. Mean weight (g) at-age from Coastal survey data (A6335). Mean weights of ages $>\mathbf{7}$ have higher uncertainty due to few samples. The split between coastal cod and Northeast Arctic cod is uncertain for age 1. For the plus group, mean weight is the average mean weight for ages 10+, weighted by abundance-at-age.

| Age |  |  |  |  |  |  |  | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $10+$ |  |  |
| 1995 | 58 | 282 | 719 | 1395 | 2091 | 2767 | 4693 | 5905 | 7211 | 13022 |
| 1996 | 41 | 216 | 672 | 1349 | 1939 | 2779 | 4223 | 6638 | 11146 | 20000 |
| 1997 | 41 | 244 | 655 | 1393 | 1914 | 2921 | 2988 | 3768 | 9600 | 7779 |
| 1998 | 49 | 259 | 840 | 1406 | 2261 | 3173 | 4320 | 5275 | 5896 | 15476 |


|  | Age |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 2.2.12. Northern Norwegian coastal cod. Maturity-at-age as determined from maturity stages observed in the coastal survey (A6335). Maturity for age 10+ is the average proportion mature for ages 10 and above, weighted by abun-dance-at-age. The split between coastal cod and Northeast Arctic cod is uncertain for age 1.

| Age |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 9 | $10+$ |  |
| 1995 | 0.00 | 0.00 | 0.13 | 0.51 | 0.60 | 0.78 | 0.86 | 0.99 | 1.00 | 1.00 |


|  | Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 1996 | 0.00 | 0.02 | 0.14 | 0.38 | 0.74 | 0.84 | 0.92 | 1.00 | 1.00 | 1.00 |
| 1997 | 0.03 | 0.06 | 0.25 | 0.36 | 0.64 | 0.93 | 0.92 | 0.86 | 1.00 | 1.00 |
| 1998 | 0.01 | 0.03 | 0.13 | 0.24 | 0.56 | 0.70 | 0.98 | 0.93 | 0.88 | 1.00 |
| 1999 | 0.00 | 0.02 | 0.06 | 0.27 | 0.52 | 0.69 | 0.74 | 1.00 | 0.57 | 1.00 |
| 2000 | 0.00 | 0.00 | 0.06 | 0.20 | 0.51 | 0.68 | 0.80 | 0.92 | 1.00 | 1.00 |
| 2001 | 0.00 | 0.00 | 0.04 | 0.27 | 0.76 | 0.96 | 0.97 | 0.97 | 1.00 | 1.00 |
| 2002 | 0.00 | 0.01 | 0.11 | 0.30 | 0.78 | 0.89 | 0.98 | 0.94 | 1.00 | 1.00 |
| 2003 | 0.00 | 0.00 | 0.03 | 0.28 | 0.55 | 0.88 | 0.95 | 0.93 | 1.00 | 1.00 |
| 2004 | 0.00 | 0.01 | 0.11 | 0.30 | 0.78 | 0.92 | 0.94 | 1.00 | 1.00 | 1.00 |
| 2005 | 0.00 | 0.00 | 0.11 | 0.37 | 0.56 | 0.83 | 0.94 | 0.97 | 1.00 | 1.00 |
| 2006 | 0.00 | 0.01 | 0.19 | 0.53 | 0.72 | 0.93 | 0.90 | 0.96 | 1.00 | 1.00 |
| 2007 | 0.00 | 0.00 | 0.16 | 0.54 | 0.72 | 0.93 | 0.96 | 1.00 | 1.00 | 1.00 |
| 2008 | 0.00 | 0.02 | 0.10 | 0.30 | 0.73 | 0.88 | 0.97 | 1.00 | 1.00 | 1.00 |
| 2009 | 0.00 | 0.00 | 0.05 | 0.21 | 0.39 | 0.64 | 0.77 | 0.90 | 0.97 | 0.94 |
| 2010 | 0.00 | 0.00 | 0.03 | 0.27 | 0.57 | 0.78 | 0.92 | 0.99 | 0.98 | 1.00 |
| 2011 | 0.02 | 0.00 | 0.05 | 0.31 | 0.63 | 0.74 | 0.89 | 0.90 | 0.88 | 1.00 |
| 2012 | 0.00 | 0.01 | 0.04 | 0.28 | 0.57 | 0.86 | 0.89 | 1.00 | 0.96 | 1.00 |
| 2013 | 0.00 | 0.00 | 0.02 | 0.22 | 0.57 | 0.86 | 0.99 | 0.94 | 0.96 | 1.00 |
| 2014 | 0.00 | 0.00 | 0.03 | 0.15 | 0.56 | 0.78 | 0.90 | 0.98 | 1.00 | 1.00 |
| 2015 | 0.00 | 0.01 | 0.04 | 0.19 | 0.48 | 0.74 | 0.78 | 0.93 | 0.95 | 1.00 |
| 2016 | 0.00 | 0.00 | 0.06 | 0.28 | 0.61 | 0.85 | 0.91 | 0.98 | 1.00 | 1.00 |
| 2017 | 0.00 | 0.00 | 0.05 | 0.29 | 0.60 | 0.83 | 0.95 | 1.00 | 0.91 | 1.00 |
| 2018 | 0.00 | 0.00 | 0.07 | 0.24 | 0.60 | 0.79 | 0.94 | 1.00 | 1.00 | 1.00 |
| 2019 | 0.00 | 0.00 | 0.05 | 0.23 | 0.50 | 0.73 | 0.89 | 1.00 | 0.97 | 1.00 |
| 2020 | 0.00 | 0.02 | 0.07 | 0.33 | 0.60 | 0.88 | 0.97 | 0.98 | 1.00 | 1.00 |
| 2021 | 0.00 | 0.00 | 0.07 | 0.29 | 0.58 | 0.88 | 0.89 | 0.96 | 1.00 | 1.00 |

## Table 2.2.13. Northern Norwegian coastal cod. Tuning data used in the final SAM run.

Norwegian Coastal cod
101
Norw-Coast-Ac-Q4-1995 (Aco)
19952002

| 1 | 1 | 0.75 | 0.85 |
| :--- | :--- | :--- | :--- | -2

153586
138553
145079
139064

116012
135255
127051
121098

Norw-Coast-Ac-Q4-2003 (Aco)
20032020

| 1 | 1 | 0.75 | 0.85 |
| :--- | :--- | :--- | :--- |

-2
123749
17968
114601
121748
133075
15266
18428

121637
122991
120654
120705
136710

122892
130551
125918
122347
129829
126833

```
Norw-Coast-Ac-Q4 (BTr)
        2003 2021
\begin{tabular}{lrrrrrrrrr}
1 & 1 & 0.75 & 0.85 & & & & & \\
2 & 10 & & & & & & & \\
1 & 3.268 & 3.763 & 4.521 & 2.700 & 2.319 & 0.863 & 0.489 & 0.220 & 0.069 \\
1 & 2.201 & 2.396 & 2.602 & 1.463 & 0.722 & 0.359 & 0.181 & 0.046 & 0.063 \\
1 & 1.042 & 1.988 & 1.478 & 1.268 & 0.746 & 0.157 & 0.107 & 0.068 & 0.054
\end{tabular}
```

| 1 | 2.156 | 2.623 | 2.946 | 1.554 | 1.026 | 0.941 | 0.171 | 0.107 | 0.023 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :--- | :--- |
| 1 | 0.911 | 0.853 | 1.071 | 0.789 | 0.465 | 0.394 | 0.114 | 0.075 | 0.029 |
| 1 | 1.822 | 2.795 | 1.883 | 1.419 | 1.145 | 0.580 | 0.348 | 0.161 | 0.094 |
| 1 | 2.251 | 3.570 | 3.716 | 1.584 | 0.868 | 0.712 | 0.466 | 0.204 | 0.160 |
| 1 | 2.353 | 3.268 | 3.385 | 2.397 | 0.784 | 0.383 | 0.733 | 0.317 | 0.328 |
| 1 | 3.471 | 2.498 | 2.866 | 2.095 | 1.445 | 0.292 | 0.315 | 0.213 | 0.310 |
| 1 | 3.218 | 4.485 | 2.784 | 1.537 | 1.042 | 0.930 | 0.411 | 0.200 | 0.346 |
| 1 | 4.101 | 1.706 | 2.666 | 1.887 | 1.575 | 0.890 | 0.578 | 0.297 | 0.419 |
| 1 | 5.448 | 4.026 | 3.034 | 3.521 | 2.016 | 1.388 | 0.465 | 0.364 | 0.337 |
| 1 | 4.733 | 4.154 | 3.727 | 2.068 | 1.818 | 0.902 | 0.506 | 0.397 | 0.222 |
| 1 | 4.433 | 4.522 | 2.610 | 1.995 | 0.746 | 0.735 | 0.413 | 0.203 | 0.210 |
| 1 | 2.891 | 2.407 | 1.563 | 1.151 | 0.715 | 0.308 | 0.2 | 0.147 | 0.157 |
| 1 | 3.197 | 1.916 | 1.879 | 1.049 | 0.748 | 0.323 | 0.183 | 0.128 | 0.168 |
| 1 | 2.114 | 2.470 | 1.508 | 1.460 | 0.839 | 0.490 | 0.148 | 0.129 | 0.211 |
| 1 | 1.670 | 2.599 | 2.416 | 1.188 | 0.611 | 0.291 | 0.177 | 0.049 | 0.072 |
| 1 | 2.531 | 1.367 | 1.589 | 1.367 | 0.732 | 0.289 | 0.239 | 0.082 | 0.081 |

Table 2.2.14. Northern Norwegian coastal cod. Stock mean weight-at-age (kg) was used in the assessment model. Mean weights at age in the catch are used in place of stock weights for ages 8-10+. Mean weights in 1994, when the survey had not yet started, are means of stock weights in the years 1995-1997 for ages 2-7 and set to weight in catch for ages 8-10+.

| Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 1994 | 0.247 | 0.682 | 1.379 | 1.981 | 2.822 | 3.968 | 5.245 | 6.487 | 8.825 |
| 1995 | 0.282 | 0.719 | 1.395 | 2.091 | 2.767 | 4.693 | 5.228 | 6.121 | 9.469 |
| 1996 | 0.216 | 0.672 | 1.349 | 1.939 | 2.779 | 4.223 | 4.544 | 5.462 | 7.814 |
| 1997 | 0.244 | 0.655 | 1.393 | 1.914 | 2.921 | 2.988 | 4.738 | 5.616 | 7.768 |
| 1998 | 0.259 | 0.840 | 1.406 | 2.261 | 3.173 | 4.320 | 4.786 | 5.389 | 9.584 |
| 1999 | 0.272 | 0.793 | 1.508 | 1.964 | 2.759 | 4.257 | 4.923 | 5.415 | 8.339 |
| 2000 | 0.322 | 0.826 | 1.561 | 2.363 | 2.811 | 4.260 | 5.553 | 5.834 | 9.781 |
| 2001 | 0.377 | 0.933 | 1.660 | 2.320 | 2.998 | 3.338 | 4.498 | 4.794 | 7.711 |
| 2002 | 0.357 | 0.918 | 1.595 | 2.377 | 3.468 | 4.415 | 5.268 | 6.236 | 9.943 |
| 2003 | 0.361 | 0.820 | 1.427 | 2.269 | 3.127 | 4.114 | 5.417 | 5.713 | 9.07 |
| 2004 | 0.338 | 0.877 | 1.646 | 2.153 | 3.197 | 3.810 | 5.367 | 5.93 | 7.991 |
| 2005 | 0.436 | 0.878 | 1.727 | 2.205 | 2.542 | 3.666 | 5.233 | 5.981 | 8.32 |
| 2006 | 0.400 | 0.989 | 1.649 | 2.231 | 3.502 | 3.992 | 5.806 | 6.638 | 9.71 |
| 2007 | 0.486 | 1.066 | 1.865 | 2.579 | 3.168 | 4.520 | 5.781 | 6.871 | 9.771 |
| 2008 | 0.427 | 1.109 | 1.971 | 3.327 | 3.393 | 4.543 | 5.844 | 6.279 | 9.239 |


| Year | Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 2009 | 0.357 | 1.032 | 1.878 | 2.695 | 3.803 | 4.599 | 6.178 | 6.516 | 9.248 |
| 2010 | 0.502 | 1.088 | 1.872 | 2.745 | 3.586 | 4.684 | 5.437 | 6.185 | 7.599 |
| 2011 | 0.401 | 1.165 | 2.279 | 3.109 | 3.702 | 5.163 | 5.941 | 6.422 | 8.346 |
| 2012 | 0.355 | 1.141 | 2.026 | 2.907 | 3.690 | 4.688 | 6.448 | 6.914 | 9.446 |
| 2013 | 0.384 | 0.918 | 1.817 | 3.041 | 3.438 | 3.963 | 5.892 | 6.800 | 10.104 |
| 2014 | 0.359 | 1.122 | 1.894 | 2.929 | 3.690 | 4.646 | 5.791 | 6.461 | 9.643 |
| 2015 | 0.406 | 1.115 | 2.145 | 2.987 | 3.774 | 4.839 | 5.601 | 6.482 | 9.044 |
| 2016 | 0.347 | 1.101 | 1.904 | 3.327 | 3.928 | 4.689 | 5.893 | 6.850 | 8.928 |
| 2017 | 0.504 | 1.058 | 1.969 | 2.943 | 3.997 | 4.676 | 5.977 | 6.933 | 9.356 |
| 2018 | 0.522 | 1.109 | 2.094 | 3.206 | 3.763 | 5.391 | 5.711 | 6.581 | 9.333 |
| 2019 | 0.372 | 1.131 | 1.984 | 2.983 | 3.815 | 5.141 | 5.748 | 6.562 | 8.561 |
| 2020 | 0.380 | 1.012 | 1.932 | 2.963 | 3.741 | 4.908 | 5.655 | 6.387 | 9.024 |
| 2021 | 0.348 | 0.853 | 1.704 | 2.542 | 3.756 | 4.421 | 6.391 | 7.285 | 8.998 |

Table 2.2.15. Northern Norwegian coastal cod. Natural mortality-at-age is used in the assessment model. Estimated from mean weights at age (Table 2.2.14) by the Lorenzen (1996) method.

| Year | Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 1994 | 0.687 | 0.504 | 0.407 | 0.364 | 0.327 | 0.295 | 0.271 | 0.254 | 0.231 |
| 1995 | 0.661 | 0.496 | 0.405 | 0.358 | 0.329 | 0.280 | 0.271 | 0.258 | 0.226 |
| 1996 | 0.716 | 0.507 | 0.410 | 0.367 | 0.329 | 0.289 | 0.283 | 0.267 | 0.240 |
| 1997 | 0.690 | 0.511 | 0.406 | 0.368 | 0.324 | 0.321 | 0.279 | 0.265 | 0.240 |
| 1998 | 0.677 | 0.473 | 0.404 | 0.350 | 0.316 | 0.287 | 0.278 | 0.268 | 0.225 |
| 1999 | 0.668 | 0.482 | 0.396 | 0.365 | 0.329 | 0.288 | 0.276 | 0.268 | 0.235 |
| 2000 | 0.634 | 0.476 | 0.392 | 0.345 | 0.327 | 0.288 | 0.266 | 0.262 | 0.224 |
| 2001 | 0.604 | 0.458 | 0.384 | 0.347 | 0.321 | 0.311 | 0.284 | 0.278 | 0.241 |
| 2002 | 0.615 | 0.461 | 0.389 | 0.345 | 0.307 | 0.285 | 0.270 | 0.257 | 0.223 |
| 2003 | 0.612 | 0.477 | 0.403 | 0.350 | 0.317 | 0.292 | 0.268 | 0.264 | 0.229 |
| 2004 | 0.625 | 0.467 | 0.386 | 0.355 | 0.315 | 0.298 | 0.269 | 0.261 | 0.238 |
| 2005 | 0.578 | 0.467 | 0.380 | 0.353 | 0.338 | 0.302 | 0.271 | 0.260 | 0.235 |


|  | Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 2006 | 0.594 | 0.450 | 0.385 | 0.351 | 0.306 | 0.294 | 0.262 | 0.252 | 0.224 |
| 2007 | 0.559 | 0.440 | 0.371 | 0.336 | 0.316 | 0.283 | 0.263 | 0.249 | 0.224 |
| 2008 | 0.582 | 0.435 | 0.365 | 0.311 | 0.309 | 0.283 | 0.262 | 0.256 | 0.228 |
| 2009 | 0.614 | 0.444 | 0.370 | 0.332 | 0.299 | 0.282 | 0.258 | 0.253 | 0.228 |
| 2010 | 0.554 | 0.437 | 0.371 | 0.330 | 0.304 | 0.280 | 0.268 | 0.257 | 0.242 |
| 2011 | 0.593 | 0.428 | 0.349 | 0.318 | 0.301 | 0.272 | 0.261 | 0.255 | 0.235 |
| 2012 | 0.615 | 0.431 | 0.362 | 0.324 | 0.301 | 0.280 | 0.254 | 0.249 | 0.226 |
| 2013 | 0.601 | 0.461 | 0.374 | 0.320 | 0.308 | 0.295 | 0.261 | 0.250 | 0.222 |
| 2014 | 0.613 | 0.433 | 0.369 | 0.323 | 0.301 | 0.281 | 0.263 | 0.254 | 0.225 |
| 2015 | 0.591 | 0.434 | 0.356 | 0.321 | 0.299 | 0.277 | 0.265 | 0.254 | 0.229 |
| 2016 | 0.620 | 0.436 | 0.369 | 0.311 | 0.296 | 0.280 | 0.261 | 0.250 | 0.230 |
| 2017 | 0.553 | 0.441 | 0.365 | 0.323 | 0.294 | 0.280 | 0.260 | 0.249 | 0.227 |
| 2018 | 0.547 | 0.435 | 0.358 | 0.315 | 0.300 | 0.268 | 0.264 | 0.253 | 0.227 |
| 2019 | 0.607 | 0.432 | 0.364 | 0.322 | 0.298 | 0.272 | 0.263 | 0.253 | 0.233 |
| 2020 | 0.603 | 0.447 | 0.367 | 0.322 | 0.300 | 0.276 | 0.265 | 0.255 | 0.229 |
| 2021 | 0.619 | 0.471 | 0.381 | 0.338 | 0.300 | 0.285 | 0.255 | 0.245 | 0.230 |

Table 2.2.16. Northern Norwegian coastal cod. SAM configuration.
Model used: SAM (State-space assessment model; https://www.stockassessment.org; Nielsen and Berg 2014).
Software used: Template Model Builder (TMB) and R.

Age range of assessment: $2-10$, where 10 is a plus group.

## Start year of assessment: 1994

## Last change of configuration: WKNCCHCR 2022

The assessment is available at www.stockassessment.org under the name NCCN67_acotsb_2022_Excl2021acou
\# Configuration saved: Thu Oct 21 15:33:05 2021
\# Where a matrix is specified rows corresponds to fleets and columns to ages. Same number indicates same parameter \# used. Numbers (integers) starts from zero and must be consecutive. Negative numbers indicate that the parameter is not \# included in the model
\$minAge
\# The minimium age class in the assessment 2

## \$maxAge

\# The maximum age class in the assessment

## \$maxAgePlusGroup

\# Is last age group considered a plus group for each fleet (1 yes, or 0 no).
1001

## \$keyLogFsta

\# Coupling of the fishing mortality states processes for each age (normally only the first row (= fleet) is used). Sequential \# numbers indicate that the fishing mortality is estimated individually for those ages; if the same number is used for two or \# more ages, F is bound for those ages (assumed to be the same). Binding fully selected ages will result in a flat selection \# pattern for those ages.
$\begin{array}{lllllllll}0 & 1 & 2 & 3 & 4 & 5 & 5 & 5 & 6\end{array}$
-1 -1 -1 -1 -1 -1 -1 -1 -1
-1 -1 -1 -1 -1 -1 -1 -1 -1
-1 -1 -1 $-1 \begin{array}{lllll}1 & -1 & -1 & -1 & -1\end{array}$

## \$corFlag

\# Correlation of fishing mortality across ages ( 0 independent, 1 compound symmetry, 2 AR(1), 3 separable AR(1).
\# 0 : independent means there is no correlation between $F$ across age 1 : compound symmetry means that all ages are equally \# correlated; 2: AR(1) first order autoregressive - similar ages are more highly correlated than ages that are further apart, \# so similar ages have similar F patterns over time. if the estimated correlation is high, then the F pattern over time for each \# age varies in a similar way. E.g if almost one, then they are parallel (like a separable model) and if almost zero then they \# are independent. 3: Separable AR - Included for historic reasons . . . more later 2

## \$keyLogFpar

\# Coupling of the survey catchability parameters (nomally first row is not used, as that is covered by fishing mortality).
-1
0 -1
1 -1 -1 $-1 \begin{array}{lllll}1 & -1 & -1 & -1 & -1\end{array}$
234555556
\$keyQpow
\# Density dependent catchability power parameters (if any).
-1
-1 -1 -1 -1 -1 $-1 \begin{array}{llll}1 & -1 & -1\end{array}$
-1 -1 -1 -1 -1 -1 -1 -1 -1
-1 -1 -1 -1 -1 -1 -1 -1 -1

## \$keyVarF

\# Coupling of process variance parameters for $\log (F)$-process (Fishing mortality normally applies to the first (fishing) fleet; \# therefore only first row is used)
000000000
-1 -1 -1 -1 -1 -1 -1 -1 -1
-1
-1

## \$keyVarLogN

\# Coupling of the recruitment and survival process variance parameters for the $\log (\mathrm{N})$-process at the different ages. It is \# advisable to have at least the first age class (recruitment) separate, because recruitment is a different process than \# survival.
011111111

## \$keyVarObs

\# Coupling of the variance parameters for the observations. First row refers to the coupling of the variance parameters for \# the catch data observations by age. Second and further rows refers to coupling of the variance parameters for the index \# data observations by age
000000000

```
1 -1 -1 -1 -1 -1 -1 -1 -1
2 -1 -1 -1 -1 \(-1 \begin{array}{llll}-1 & -1 & -1\end{array}\)
\(\begin{array}{lllllllll}3 & 3 & 3 & 3 & 3 & 3 & 3 & 3 & 3\end{array}\)
```


## \$obsCorStruct

\# Covariance structure for each fleet ("ID" independent, "AR" AR(1), or "US" for unstructured). | Possible values are: "ID", \# "AR", "US"
"ID" "ID" "ID" "AR"

```
$keyCorObs
#2-3 3-4 4-5 5-6 6-7 7-8 8-9 9-10
NA NA NA NA NA NA NA NA
-1 -1 -1 -1 -1 -1 -1 -1
-1 -1 -1 -1 -1 -1 -1 -1
0 1 1 1 1 2 3 3 3
```

\# Coupling of correlation parameters can only be specified if the AR(1) structure is chosen above. NA's indicate where
\# correlation parameters can be specified ( -1 where they cannot).

## \$stockRecruitmentModelCode

\# Stock recruitment code ( 0 for plain random walk, 1 for Ricker, 2 for Beverton-Holt, and 3 piece-wise constant). 0

## \$noScaledYears

\# Number of years where catch scaling is applied.
0
\$keyScaledYears
\# A vector of the years where catch scaling is applied.
\$keyParScaledYA
\# A matrix specifying the couplings of scale parameters (nrow = no scaled years, ncols = no ages).

## \$fbarRange

\# lowest and higest age included in Fbar
48
\$keyBiomassTreat
\# To be defined only if a biomass survey is used ( 0 SSB index, 1 catch index, 2 FSB index, 3 total catch, 4 total landings and \# 5 TSB index).
-155-1
\$obsLikelihoodFlag
\# Option for observational likelihood | Possible values are: "LN" "ALN"
"LN" "LN" "LN" "LN"
\$fixVarToWeight
\# If weight attribute is supplied for observations this option sets the treatment ( 0 relative weight, 1 fix variance to weight). 0

## \$fracMixF

\# The fraction of $\mathrm{t}(3)$ distribution used in logF increment distribution
0
\$fracMixN
\# The fraction of $\mathrm{t}(3)$ distribution used in $\operatorname{logN}$ increment distribution
0

## \$fracMixObs

\# A vector with same length as number of fleets, where each element is the fraction of $t(3)$ distribution used in the

```
# distribution of that fleet
0000
$constRecBreaks
# Vector of break years between which recruitment is at constant level. The break year is included in the left interval. (This
# option is only used in combination with stock-recruitment code 3)
$predVarObsLink
# Coupling of parameters used in a prediction-variance link for observations.
    -1 -1 -1 -1 -1 -1 -1 -1 -1
NA NA NA NA NA NA NA NA NA
NA NA NA NA NA NA NA NA NA
-1 -1 -1 -1 -1 -1 -1 -1 -1
$hockeyStickCurve
#
20
$stockWeightModel
# Integer code describing the treatment of stock weights in the model (0 use as known, 1 use as observations to inform
# stock weight process (GMRF with cohort and within year correlations))
O
$keyStockWeightMean
# Coupling of stock-weight process mean parameters (not used if stockWeightModel==0)
NA NA NA NA NA NA NA NA NA
$keyStockWeightObsVar
# Coupling of stock-weight observation variance parameters (not used if stockWeightModel==0)
NA NA NA NA NA NA NA NA NA
$catchWeightModel
# Integer code describing the treatment of catch weights in the model (0 use as known, 1 use as observations to inform
# catch weight process (GMRF with cohort and within year correlations))
O
$keyCatchWeightMean
# Coupling of catch-weight process mean parameters (not used if catchWeightModel==0)
NA NA NA NA NA NA NA NA NA
$keyCatchWeightObsVar
# Coupling of catch-weight observation variance parameters (not used if catchWeightModel==0)
NA NA NA NA NA NA NA NA NA
$matureModel
# Integer code describing the treatment of proportion mature in the model (0 use as known, 1 use as observations to inform
# proportion mature process (GMRF with cohort and within year correlations on logit(proportion mature)))
O
$keyMatureMean
# Coupling of mature process mean parameters (not used if matureModel==0)
NA NA NA NA NA NA NA NA NA
$mortalityModel
# Integer code describing the treatment of natural mortality in the model (0 use as known, 1 use as observations to inform
# natural mortality process (GMRF with cohort and within year correlations))
O
```

\#
NA NA NA NA NA NA NA NA NA
\$keyMortalityObsVar
\# Coupling of natural mortality observation variance parameters (not used if mortalityModel==0)
NA NA NA NA NA NA NA NA NA
\$keyXtraSd
\# An integer matrix with 4 columns (fleet year age coupling), which allows additional uncertainty to be estimated for the specified observations

Table 2.2.17. Northern Norwegian coastal cod. SAM output. Estimated catchability at age for each fleet. The two parts of the acoustic biomass index have one catchability parameter each as the biomass index is not split by age. In the sweptarea index, catchabilities are coupled (set equal) in the SAM configuration for ages 5-9.

| Fleet/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | 2+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Acoustic biomass index pt. 1 | - | - | - | - | - | - | - | - | - | 0.131 |
| Acoustic biomass index pt. 2 | - | - | - | - | - | - | - | - | - | 0.084 |
| Sweptarea index | 0.000059 | 0.000099 | 0.000141 | 0.000155 | 0.000155 | 0.000155 | 0.000155 | 0.000155 | 0.000206 | - |

Table 2.2.18. Northern Norwegian coastal cod. SAM output. Estimated recruitment (1000's), Spawning-stock biomass (SSB, t), average fishing mortalities for ages 4-8 (Fbar(4-8)), and Totalstock biomass (TSB, t).

| Year/Age | R (age 3) | Low | High | SSB | Low | High | Fbar (4-8) | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 34992 | 27728 | 44159 | 130993 | 96944 | 177001 | 0.28 | 0.22 | 0.35 | 319804 | 274928 | 372005 |
| 1995 | 41348 | 33622 | 50850 | 111126 | 83352 | 148154 | 0.37 | 0.30 | 0.45 | 304438 | 265793 | 348702 |
| 1996 | 50989 | 42740 | 60829 | 88896 | 71071 | 111192 | 0.36 | 0.30 | 0.45 | 250491 | 223931 | 280201 |
| 1997 | 61719 | 51220 | 74369 | 68453 | 55548 | 84356 | 0.46 | 0.38 | 0.56 | 225985 | 203998 | 250343 |
| 1998 | 53189 | 45264 | 62500 | 57631 | 46109 | 72033 | 0.47 | 0.39 | 0.57 | 239763 | 217051 | 264852 |
| 1999 | 54709 | 46415 | 64485 | 46563 | 38713 | 56004 | 0.43 | 0.35 | 0.53 | 218210 | 198852 | 239451 |
| 2000 | 53310 | 45246 | 62810 | 51115 | 43978 | 59411 | 0.33 | 0.27 | 0.41 | 229435 | 209257 | 251560 |
| 2001 | 45938 | 39048 | 54044 | 66803 | 59523 | 74974 | 0.27 | 0.22 | 0.33 | 234635 | 214160 | 257068 |
| 2002 | 46742 | 39710 | 55019 | 80746 | 71999 | 90555 | 0.31 | 0.25 | 0.37 | 252166 | 229999 | 276470 |
| 2003 | 47784 | 40608 | 56228 | 67357 | 59689 | 76011 | 0.30 | 0.25 | 0.36 | 235077 | 214066 | 258152 |
| 2004 | 42553 | 36652 | 49404 | 76698 | 67828 | 86727 | 0.33 | 0.27 | 0.40 | 237280 | 215168 | 261663 |
| 2005 | 43888 | 37773 | 50993 | 68735 | 60330 | 78311 | 0.29 | 0.24 | 0.35 | 229863 | 208034 | 253982 |
| 2006 | 35580 | 30541 | 41451 | 86978 | 75614 | 100050 | 0.34 | 0.27 | 0.41 | 239023 | 215690 | 264880 |
| 2007 | 33123 | 28438 | 38579 | 93704 | 80558 | 108995 | 0.24 | 0.189 | 0.29 | 247700 | 222464 | 275799 |
| 2008 | 42627 | 36580 | 49673 | 93726 | 79841 | 110026 | 0.22 | 0.180 | 0.28 | 264965 | 237448 | 295670 |
| 2009 | 40757 | 35217 | 47169 | 72516 | 60529 | 86876 | 0.19 | 0.149 | 0.23 | 258190 | 230778 | 288857 |
| 2010 | 37846 | 32781 | 43693 | 83864 | 70567 | 99665 | 0.23 | 0.185 | 0.28 | 270820 | 243344 | 301398 |


| Year/Age | R (age 3) | Low | High | SSB | Low | High | Fbar (4-8) | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 36999 | 31855 | 42973 | 94782 | 80225 | 111980 | 0.21 | 0.167 | 0.26 | 292032 | 262578 | 324790 |
| 2012 | 45225 | 39106 | 52300 | 99826 | 84178 | 118383 | 0.166 | 0.135 | 0.21 | 286213 | 257303 | 318372 |
| 2013 | 34861 | 30026 | 40475 | 100983 | 85441 | 119352 | 0.143 | 0.116 | 0.176 | 275339 | 247640 | 306137 |
| 2014 | 40524 | 35046 | 46858 | 105791 | 90454 | 123730 | 0.139 | 0.114 | 0.170 | 295707 | 267055 | 327433 |
| 2015 | 40680 | 35009 | 47270 | 97679 | 83546 | 114202 | 0.20 | 0.164 | 0.24 | 315537 | 285400 | 348856 |
| 2016 | 42547 | 35926 | 50387 | 102587 | 88477 | 118947 | 0.28 | 0.24 | 0.34 | 308133 | 276832 | 342974 |
| 2017 | 42932 | 35446 | 52000 | 86205 | 73511 | 101091 | 0.37 | 0.31 | 0.44 | 299472 | 264379 | 339224 |
| 2018 | 41939 | 33397 | 52665 | 82113 | 68806 | 97994 | 0.32 | 0.26 | 0.39 | 297859 | 254656 | 348391 |
| 2019 | 52186 | 39653 | 68680 | 71717 | 57570 | 89341 | 0.34 | 0.27 | 0.43 | 284703 | 235074 | 344811 |
| 2020 | 42704 | 30808 | 59193 | 83705 | 62620 | 111890 | 0.36 | 0.26 | 0.49 | 273050 | 215481 | 346000 |
| 2021 | 34086 | 22950 | 50628 | 80421 | 55386 | 116771 | 0.28 | 0.185 | 0.41 | 245947 | 182979 | 330584 |

Table 2.2.19. Northern Norwegian coastal cod. SAM output. Estimated fishing mortalities at age. F for ages 7-9 are coupled (set equal) in the SAM configuration.

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 0.000 | 0.005 | 0.038 | 0.162 | 0.327 | 0.432 | 0.432 | 0.432 | 0.330 |
| 1995 | 0.000 | 0.008 | 0.055 | 0.181 | 0.388 | 0.602 | 0.602 | 0.602 | 0.424 |
| 1996 | 0.001 | 0.018 | 0.091 | 0.227 | 0.396 | 0.551 | 0.551 | 0.551 | 0.424 |
| 1997 | 0.001 | 0.025 | 0.119 | 0.274 | 0.535 | 0.680 | 0.680 | 0.680 | 0.559 |
| 1998 | 0.004 | 0.054 | 0.243 | 0.470 | 0.631 | 0.502 | 0.502 | 0.502 | 0.425 |
| 1999 | 0.001 | 0.027 | 0.169 | 0.382 | 0.540 | 0.535 | 0.535 | 0.535 | 0.463 |
| 2000 | 0.001 | 0.016 | 0.127 | 0.323 | 0.403 | 0.403 | 0.403 | 0.403 | 0.414 |
| 2001 | 0.000 | 0.010 | 0.085 | 0.223 | 0.342 | 0.359 | 0.359 | 0.359 | 0.660 |
| 2002 | 0.001 | 0.012 | 0.082 | 0.211 | 0.379 | 0.428 | 0.428 | 0.428 | 0.870 |
| 2003 | 0.001 | 0.013 | 0.066 | 0.177 | 0.329 | 0.456 | 0.456 | 0.456 | 0.869 |
| 2004 | 0.001 | 0.008 | 0.050 | 0.144 | 0.326 | 0.555 | 0.555 | 0.555 | 1.001 |
| 2005 | 0.000 | 0.008 | 0.054 | 0.150 | 0.279 | 0.481 | 0.481 | 0.481 | 1.121 |
| 2006 | 0.001 | 0.011 | 0.068 | 0.190 | 0.328 | 0.550 | 0.550 | 0.550 | 1.634 |
| 2007 | 0.001 | 0.016 | 0.077 | 0.181 | 0.248 | 0.334 | 0.334 | 0.334 | 0.981 |
| 2008 | 0.001 | 0.018 | 0.073 | 0.205 | 0.259 | 0.287 | 0.287 | 0.287 | 0.650 |
| 2009 | 0.001 | 0.015 | 0.046 | 0.153 | 0.240 | 0.244 | 0.244 | 0.244 | 0.413 |
| 2010 | 0.001 | 0.018 | 0.055 | 0.185 | 0.303 | 0.299 | 0.299 | 0.299 | 0.576 |
| 2011 | 0.002 | 0.021 | 0.063 | 0.141 | 0.221 | 0.306 | 0.306 | 0.306 | 0.543 |
| 2012 | 0.006 | 0.038 | 0.078 | 0.129 | 0.184 | 0.220 | 0.220 | 0.220 | 0.449 |
| 2013 | 0.003 | 0.026 | 0.061 | 0.106 | 0.149 | 0.200 | 0.200 | 0.200 | 0.366 |
| 2014 | 0.003 | 0.022 | 0.061 | 0.101 | 0.141 | 0.196 | 0.196 | 0.196 | 0.386 |
| 2015 | 0.005 | 0.040 | 0.095 | 0.142 | 0.206 | 0.273 | 0.273 | 0.273 | 0.592 |
| 2016 | 0.003 | 0.030 | 0.097 | 0.155 | 0.282 | 0.443 | 0.443 | 0.443 | 0.810 |
| 2017 | 0.009 | 0.057 | 0.145 | 0.215 | 0.352 | 0.560 | 0.560 | 0.560 | 0.856 |
| 2018 | 0.003 | 0.025 | 0.085 | 0.160 | 0.271 | 0.536 | 0.536 | 0.536 | 0.678 |
| 2019 | 0.001 | 0.020 | 0.082 | 0.164 | 0.289 | 0.583 | 0.583 | 0.583 | 0.703 |
| 2020 | 0.001 | 0.017 | 0.090 | 0.192 | 0.357 | 0.579 | 0.579 | 0.579 | 0.598 |
| 2021 | 0.002 | 0.021 | 0.104 | 0.203 | 0.281 | 0.397 | 0.397 | 0.397 | 0.477 |

Table 2.2.20. Northern Norwegian coastal cod. SAM output. Estimated stock numbers-at-age (1000’s).

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 83306 | 34992 | 39108 | 35936 | 17710 | 10156 | 4859 | 1124 | 3081 |
| 1995 | 99046 | 41348 | 21156 | 25103 | 21136 | 9196 | 4904 | 2426 | 2391 |
| 1996 | 122445 | 50989 | 24666 | 13420 | 14705 | 10279 | 3800 | 2040 | 2278 |
| 1997 | 106813 | 61719 | 29781 | 14775 | 7417 | 7169 | 4422 | 1665 | 2104 |
| 1998 | 109788 | 53189 | 37075 | 17441 | 7753 | 3140 | 2647 | 1673 | 1590 |
| 1999 | 102990 | 54709 | 31344 | 19796 | 7616 | 2989 | 1429 | 1217 | 1595 |
| 2000 | 87319 | 53310 | 32472 | 17792 | 9540 | 3182 | 1313 | 631 | 1351 |
| 2001 | 84360 | 45938 | 32944 | 19070 | 9052 | 4668 | 1581 | 661 | 1043 |
| 2002 | 87770 | 46742 | 28354 | 20764 | 10684 | 4636 | 2421 | 840 | 787 |
| 2003 | 80865 | 47784 | 29702 | 17439 | 12024 | 5330 | 2266 | 1202 | 671 |
| 2004 | 79982 | 42553 | 30041 | 18718 | 10327 | 6229 | 2509 | 1083 | 819 |
| 2005 | 63046 | 43888 | 25902 | 19644 | 11604 | 5441 | 2650 | 1088 | 718 |
| 2006 | 60940 | 35580 | 27501 | 16663 | 12008 | 6298 | 2525 | 1254 | 696 |
| 2007 | 72916 | 33123 | 22655 | 17391 | 9725 | 6490 | 2681 | 1128 | 661 |
| 2008 | 72708 | 42627 | 20802 | 14229 | 10443 | 5543 | 3553 | 1480 | 827 |
| 2009 | 68818 | 40757 | 27790 | 13391 | 8383 | 5924 | 3135 | 2090 | 1208 |
| 2010 | 64863 | 37846 | 25556 | 18703 | 8278 | 4860 | 3508 | 1883 | 1938 |
| 2011 | 79524 | 36999 | 24213 | 16582 | 11184 | 4537 | 2725 | 2004 | 1967 |
| 2012 | 67552 | 45225 | 23859 | 15880 | 10497 | 6602 | 2574 | 1518 | 2049 |
| 2013 | 74721 | 34861 | 28813 | 15504 | 10251 | 6472 | 3989 | 1602 | 1965 |
| 2014 | 75747 | 40524 | 21186 | 18945 | 10261 | 6479 | 3964 | 2525 | 2094 |
| 2015 | 78329 | 40680 | 25643 | 13703 | 12444 | 6653 | 3986 | 2523 | 2721 |
| 2016 | 78398 | 42547 | 25151 | 16357 | 8609 | 7530 | 3883 | 2323 | 2673 |
| 2017 | 76621 | 42932 | 25648 | 15970 | 10107 | 4889 | 3645 | 1921 | 2091 |
| 2018 | 89067 | 41939 | 26384 | 15213 | 9415 | 5216 | 2140 | 1603 | 1555 |
| 2019 | 77253 | 52186 | 25915 | 16956 | 9522 | 5308 | 2320 | 971 | 1358 |
| 2020 | 63412 | 42704 | 33048 | 16728 | 10302 | 5326 | 2260 | 989 | 947 |
| 2021 | 79169 | 34086 | 26829 | 20881 | 10098 | 5286 | 2294 | 960 | 842 |

Table 2.2.21a. Northern Norwegian coastal cod. Assumptions for the interim year and in the forecast: Fbar, recruitment, SSB and catch.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| Fages 4-7 (2022) 0.280 | $F_{\text {sq }}=$ median fishing mortality in 2021. |  |
| SSB (2022) | 86899 | Short-term forecast fishing at status quo ( $F_{\text {sq }}$ ); Tonnes. |
| $R_{\text {Rage 2 }}(2022$ <br> and 2023) | 77253 | Median resampled recruitment (2012-2021). The youngest age in the model is age 2. Other <br> reported recruitments are at age 3 when the fish enter the fishery; thousands. |
| Total catch <br> (2022) | 43688 | Short-term forecast fishing at $F_{\text {sq }}$; Tonnes. |

Table 2.2.21b. Northern Norwegian coastal cod. Assumptions for the interim year and in the forecast: mean weights in catch and stock, maturity-at-age, and natural mortality-at-age (last 5 year averages).

| Age | Weight in catch (kg) | Weight in stock (kg) | Proportion mature | Natural mortality |
| :--- | :--- | :--- | :--- | :--- |
| 2 | 1.313 | 0.425 | 0.005 | 0.586 |
| 3 | 1.877 | 1.032 | 0.060 | 0.445 |
| 4 | 2.545 | 1.936 | 0.275 | 0.367 |
| 5 | 3.289 | 2.927 | 0.576 | 0.324 |
| 6 | 5.135 | 4.907 | 0.822 | 0.929 |
| 8 | 5.896 | 5.896 | 0.987 | 0.276 |
| 9 | 9.749 | 9.054 | 0.976 | 0.261 |
| $10+$ |  |  |  | 0.229 |

Table 2.2.22. Northern Norwegian coastal cod. Catch scenarios.

| Basis | Total catch (2023) | $\mathrm{F}_{\text {total }}(2023)$ | SSB (2023)* | $\begin{aligned} & \text { \% SSB } \\ & \text { change** } \end{aligned}$ | \% advice change*** | \% probability of SSB falling below SSB ${ }_{\text {lower }}$ bound in 2023 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |
| Management plan^ | 29347 | 0.176 | 93809 | 8.0 | 142 | 8.8 |
| Other scenarios |  |  |  |  |  |  |
| $F=0$ | 0 | 0 | 109399 | 26 | -100 | 0.9 |
| $F=F_{2021}$ | 44278 | 0.28 | 85568 | -1.5 | 265 | 18.5 |

* For this stock, SSB is calculated at the time of survey (October) as maturity ogives and stock weights are from the survey. Thus SSB is influenced by fisheries between 1 January and 1 October. The actual spawning time is MarchJune.
** SSB in October 2023 relative to SSB in October 2022 (86 899 tonnes).
*** Advice for 2023 relative to advice for 2022 (12 146 tonnes).
^According to the harvest control rule (HCR) in the MP (ICES, 2022a). The advice basis has changed compared to last year following the adoption of the harvest control rule (HCR) evaluated in ICES (2022a) by the managing body.


Recruitment (age 3)


Catches


Figure 2.2.1. Northern Norwegian coastal cod. Standard figures. SAM estimates of a) SSB, b) Fbar(4-8), c) recruitment (age 2,), and d) catch input data.


Figure 2.2.2. Northern Norwegian coastal cod. Acoustic abundance index by age (colours) from the Coastal survey in October-November (survey code A6335). Note that starting in 2022, the acoustic index is included in the assessment model as a total biomass index rather than numbers-at-age.

Acoustic survey index, ages $2+$


Figure 2.2.3. Northern Norwegian coastal cod. Acoustic biomass index (ages 2+) from the Coastal survey in OctoberNovember. Biomass for ages 1+ are reported in Table 2.2.5, but it is biomass for ages 2+ that goes into the assessment model due to the difficulty of distinguishing between coastal and Northeast Arctic cod for age 1. Note that the final data point (2021) was excluded from this year's assessment (see Box 2.1).


Figure 2.2.4. Northern Norwegian coastal cod. Swept-area abundance index by age (colours) from the coastal survey in October-November (survey code A6335).


Figure 2.2.5. Northern Norwegian coastal cod. Survey mortality $(Z)$ at age (colours) in the acoustic index (top) and sweptarea index (bottom). $Z$ was estimated as $-\log \left(A_{a+1, y+a} / A_{a, y}\right)$, where $A_{a, y}$ is abundance of age $a$ in year $y$.


Figure 2.2.6. Northern Norwegian coastal cod. Mean weight-at-age in the coastal survey. Few individuals of ages 10+ were sampled at the beginning of the time-series, leading to extremely large variation in mean weights. In the stock assessment model, stock weights for ages 8-10+ are set equal to mean weight of these ages in the catch.


Figure 2.2.7. Northern Norwegian coastal cod. Proportions mature-at-age as observed in the Coastal survey. Since the survey takes place in October-November and the main spawning season is in March-April, spent/resting individuals are included as mature when calculating these proportions.


Figure 2.2.8. Northern Norwegian coastal cod. Natural mortality-at-age estimated from stock weights-at-age by the Lorenzen (1996) method.


Figure 2.2.9. Northern Norwegian coastal cod. Northern Norwegian coastal cod. 5-year retrospective peel: a) SSB, b) Fbar, c) recruitment, and d) catch. The Mohn's rho value (5-year average retrospective bias) is indicated in the upper right corner of each panel.


Figure 2.2.10. Northern Norwegian coastal cod. Residuals for the $\log (N)($ top $)$ and $\log (F)$ (bottom) process from the final SAM run.


Figure 2.2.11. Northern Norwegian coastal cod. One-step-ahead residuals by fleet from the final SAM run. Blue circles indicate positive residuals and red circles indicate negative residuals. Top left: catch, top right: acoustic index pt. 2, bottom left: acoustic index pt. 1, bottom right: swept-area index.


Figure 2.2.12. Northern Norwegian coastal cod. Short-term prediction. Predicted SSB (top panels), Fbar (middle panels) and recruitment (bottom panels) at status quo fishing (top left), status quo then zero fishing (top right), and fishing according to the management plan ( $F_{0.1}=0.176$ ). In the forecast, recruitment is the same for all scenarios (resampled from the period 2012-2021).

### 2.3 Southern Norwegian coastal cod

### 2.3.1 Stock status summary

An assessment based on the decisions of the 2021 WKBARFAR benchmark (ICES 2021b) is presented for this stock.

Commercial catches have decreased since 2010-2012 (Figure 2.3.1). To some extent this is explained by decreasing effort until 2013, but catches have continued to decrease after 2013 when the effort has been stable or increasing (Figures 2.3.8 and 2.3.9). The recreational fishery by tourists and Norwegian residents is assumed to catch similar amounts as the commercial fishery (Figure 2.3.1 and Table 2.3.3), and a prerequisite for more accurate future assessments is a better estimation of the recreational catches.

Catch advice for southern Norwegian coastal cod $\left(62-67^{\circ} \mathrm{N}\right)$ follows the " rfb " rule for category 3 stocks (ICES, 2020, 2022). The " rfb " rule is primarily driven by the trend in the coastal reference fleet gillnet CPUE index (more controlled than a full fleet CPUE, Section 2.3.3). Thus, the advice depends heavily on the representativeness of the CPUE index (Fischer et al., 2020). The CPUE index has increased enough that the $+20 \%$ stability cap was reached (Section 2.3.9, Figure 2.3.7, and Table 2.3.7).

A stochastic length-based spawning potential ratio (LBSPR) model and other length-based indicators are presented as additional information. In the previous assessment, the LBSPR was used to assess the need for a $20 \%$ precautionary buffer in the " 2 over 3 " rule, although ICES lacks a framework for using the LBSPR directly as a basis for catch advice. ICES recommends the use of the surplus production model SPiCT for category 3 stocks, but the SPiCT fit was determined to be unsatisfactory in the 2021 benchmark and has not been updated here (ICES 2021b).

The LBSPR model estimates that stock size is below, and fishing pressure is above, possible MSY reference points (Figures 2.3.10 and 2.3.11). From 2010-2021, the "spawning potential ratio" (SPR), i.e. the ratio between the spawning potential of the current stock and the theoretical spawning potential without fishing, fluctuated between $20-35 \%$ with an overall downward trend. SPR in 2021 was estimated as $25 \%$ ( $95 \%$ CI: 21-29\%), which places the stock below generally accepted target values ( $30-40 \%$ SPR).

Additional length-based indicators depict a somewhat depleted and worsening stock status. For example, mean length and the mean length of the largest $5 \%$ of caught fish have decreased over the past decade (Figure 2.3.12). The length at $50 \%$ selectivity, i.e. first capture, has decreased from ca. 57.6 to 48.4 cm (Figure 2.3.13). About half of the catch is immature, and this proportion has increased in the last decade (Figure 2.3.14). The minimum legal size ( 44 cm ) is well below the length at $50 \%$ maturity ( 62.8 cm ).

Priorities for more accurate future assessments are 1) better estimation of recreational catches, and 2) re-evaluation of available survey data that could be used as indices. Possible model improvements include 1) accounting for uncertainty in the index, and 2) combining index and length data in one model.

The catch advice for 2022 was 7613 tonnes. The advice for 2023 is that catches should be no more than 9136 tonnes. Assuming recreational catches of 4420 tonnes, this implies a commercial catch of no more than 4716 tonnes.

### 2.3.2 Fisheries (Table 2.3.2-Table 2.3.4)

Coastal cod is fished throughout the year but the main (about 70\%) commercial fishery for coastal cod in the area between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$ takes place during February-April. The main fishing areas are along the coast of Helgeland including Træna and Lovund, Vikna, Halten bank, and further along the coast of Trøndelag and Møre and Romsdal counties. Except for the Borgundfjord at Møre, the quantities fished inside fjords are quite low.

In the 1990s the average percentage share between gear types in the estimated coastal cod commercial landings was around $65 \%$ gillnet, $26 \%$ longline/handline, $8 \%$ Danish seine, and $1 \%$ bottom trawl. In 2021 this share was 50\% gillnet, $15 \%$ longline/handline, $27 \%$ Danish seine, and 5\% bottom trawl (Table 2.3.4).

Recreational and tourist fisheries take an important fraction of the total catches in some local areas, especially near the coastal cities, and in some fjords where commercial fishing activity is low. However, there are a few reports trying to assess the amount in certain years (see section 2.1). The current split of the recreational catches between the area north of $67^{\circ} \mathrm{N}$ and between $62-$ $67^{\circ} \mathrm{N}$ in 2019-2021 is done based on the tourist fishing businesses' reporting to the Norwegian Directorate of Fisheries by county. Since the $67^{\circ} \mathrm{N}$ latitude goes through the Nordland county, the splitting north and south of $67^{\circ} \mathrm{N}$ for this county is done proportional to the number of tourist fishing businesses north and south of this latitude. The same area proportion ( $37.8 \%$ south and $62.2 \%$ north) of the recreational fishery is used for the whole time-series back to 1994, and this is a very rough assumption that should be further investigated and better documented. In 2021, the recreational cod catches between $62-67^{\circ} \mathrm{N}$ are estimated to about $52 \%$ of total cod catches in this region (Table 2.3.3).
Discarding is known to take place. There have previously been conducted two investigations trying to estimate the level of discarding and misreporting from coastal fishing vessels in two periods (2000 and 2002-2003, WD 14 at 2002 WG). The amount of discards was calculated, and the report from the 2000-investigation concluded there was both discard and misreporting by species in 2000, in the gillnet fishery approximately $8-10 \%$ relative to reported catch. One-third of this was probably coastal cod. The last report concluded that misreporting in the Norwegian coastal gillnet fisheries have been reduced significantly since 2000.

According to Berg and Nedreaas (2021), between 2-5\% was discarded in the commercial gillnet fishery in the area $62-67^{\circ} \mathrm{N}$ during $2012-2018$, and about $7 \%$ in the rod and line sector of the recreational fishery. The latter estimate is based on reporting to the Directorate of Fisheries in 2019 showing that about $35 \%$ of the reported rod and line catch was released with an assumed mortality of $20 \%$ of the released cod (Section 2.1). Discarding is not included in the commercial catch in this report but discarding in the rod and line (from boat) sector of the recreational fishery is included in the recreational catch estimate.

### 2.3.2.1 Estimated catches and catch-at-age (Table 2.3.2-Table 2.3.4, and Figure 2.1.1 and Figure 2.3.1-Figure 2.3.2)

The current coastal cod assessments include all coastal cod caught within the coastal statistical areas 600, 601, 700 and 701 which extend beyond the 12 nautical mile zone (see Figure 2.1.1). Estimated commercial and recreational catches of coastal cod and Northeast Arctic (NEA) cod in these statistical areas between $62-60^{\circ} \mathrm{N}$ are shown in Table 2.1.1 and Figures 2.3.1-2.3.2.

The estimated commercial catch-at-age ( $2-10+$ ) for the period 1994-2021 is given in Table 2.3.2. Table 2.3.3 shows the total catch numbers-at-age when recreational and tourist fishing is included, where the proportions-at-age for the recreational catch are assumed equal to those from the commercial catch. The commercial catch in 2021 by gear and Norwegian statistical fishing areas is presented in Table 2.3.4.

### 2.3.2.2 Catch weights-at-age (Table 2.3.5)

Mean weight-at-age in catches is derived from the commercial sampling and is shown in Table 2.3.5. The same weight-at-age is assumed for the recreational and tourist catches.

### 2.3.2.3 Recreational catches in 2023

To split the 2023 catch advice into commercial and recreational components, we assume continued recovery of the tourist/recreational catch towards the pre-Covid level. The assumed recreational catch in 2021 was 4039 t , and for 2022 we assume halfway between this and the pre-Covid level ( 4800 t ), which is 4420 t .

### 2.3.3 Reference fleet

The Norwegian Reference Fleet is a group of active fishing vessels paid and tasked with providing information about catches (self-sampling) and general fishing activity to the Institute of Marine Research. The fleet consists of both high seas and coastal vessels that cover most of the Norwegian waters. The Highseas Reference Fleet began in 2000 and was expanded to include coastal vessels in 2005 (Clegg and Williams, 2020). The Coastal reference fleet has reported catch-pergillnet soaking time (CPUE) from their daily catch operations (WD 07 in ICES 2021b).

These fleets catch cod from both coastal and NEA populations, which can be discriminated based on their otolith shape (Section 2.1.2). Size distribution of individuals is sampled from a subset of fishing events and, within the size samples, individuals are sampled for otolith in a presumably random way.

To determine the origin of the cod, we use all reference fleet data from north of $62^{\circ} \mathrm{N}$ (i.e. ICES Subarea 2.a.2; Norwegian statistical areas $3,4,5,0,6,7)$ with information on otolith type. In this update assessment, we used the models selected in the benchmark (ICES 2021b), after confirming that model diagnostics were satisfactory (Figures 2.3.3 and 2.3.5). To calculate the CPUE index between $62-67^{\circ} \mathrm{N}$ we only use quarters 3-4 because at that time of year there are fewer issues with mixing coastal and NEA cod (Figure 2.3.4).

### 2.3.4 Standardized CPUE index (Table 2.3.6 and Figures 2.3.3-2.3.7)

Raw CPUE data are seldom proportional to population abundance as many factors (e.g. changes in fish distribution, catch efficiency, effort, etc) potentially affect its value. Therefore, CPUE standardization is an important step that attempts to derive an index that tracks relative population dynamics.

The first step in the CPUE standardization is to estimate the proportion of Norwegian coastal vs. Northeast Arctic (NEA) cod in the catch, as these two cod stocks (ecotypes) mix in the Norwegian Sea. Our goal is to derive an index of only coastal cod abundance. We follow these steps:

1. Fit a binomial GLM to estimate the probability that cod caught between $62-67^{\circ} \mathrm{N}$ are coastal vs. NEA cod during the time frame of interest (quarters 3-4).
2. Fit a lognormal GLM to standardize total cod CPUE, taking into account year, gear, area, and quarter.
3. Combine the output from the previous two steps to create an index of abundance for only coastal cod.

Here we define important terms used in the CPUE standardization:

Standardized effort (gillnet day) = gear count x soaking time (hours) / 24 hours
CPUE (per gillnet day) = catch weight / standardized effort

## Step 1: Proportion coastal vs. NEA cod

We used all data from above $62^{\circ} \mathrm{N}$ (i.e. areas $3,4,5,0,6,7$ ) with information on otolith type. The latter is the source of identification that helps separate coastal vs. NEA cod (Section 2.1.2). Otolith types 1 and 2 were categorized as coastal cod and types 3-5 as NEA cod. Around 2500 otolith samples have been read per year since 2010. A total of 30828 samples between 2007-2021 were included in the binomial GLM, after removing covariates that had less than three observations to ensure estimability.

We then fit a binomial model with logit link using four categorical explanatory variables: year, area, quarter, and gear, with an area-year interaction effect. In other words, the probability that individual $\operatorname{cod} i$ is classified as coastal, $\pi_{i}$, is given by:

$$
\begin{equation*}
Z_{i} \sim \operatorname{Bernoulli}\left(\pi_{i}\right), \tag{eq1}
\end{equation*}
$$

$\operatorname{logit}\left(\pi_{i}\right)=\alpha+\sum_{a} \beta_{a}$ Area $_{i}+\sum_{y} \beta_{y}$ Year $_{i}+\sum_{g} \beta_{g} \operatorname{Gear}_{i}+\sum_{q} \beta_{q}$ Quarter $_{i}+\sum_{y} \sum_{a} \beta_{a, y}$ Area $_{i}$ Year $_{i}$
where $Z_{i}$ is a binary variable that equals 1 if cod $i$ was coastal and 0 if not. Likewise, Area $_{i}$, Year $_{i}$, Gear $_{i}$, and Quarter ${ }_{i}$ are 1 if cod $i$ was caught in that area, year, gear, and quarter and 0 if not.

There were no issues with the diagnostics (Figure 2.3.3). We then predicted the proportion of coastal cod that would be expected in areas 6 and 7, during quarters 3 and 4, between 2007-2021 (Figure 2.3.4).

## Step 2: Total cod CPUE standardization

The final lognormal GLMM selected in the benchmark was fitted on all cod CPUE data (no distinction between coastal and NEA cod) in areas 6-7 and quarters 3-4 between 2007-2021 (ICES 2021b). As in the benchmark, data were filtered to remove gears with less than 3 observations or only used in one year. There were only three zero catch observations out of 747, and these were removed, resulting in a final sample size of $N=744$. We fit the model:

$$
\begin{gather*}
\log \left(Y_{j}\right) \sim N\left(\mu_{j}=\alpha+\sum_{a} \beta_{a} \text { Area }_{j}+\sum_{y} \beta_{y} \text { Year }_{j}+\sum_{g} \beta_{g} \operatorname{Gear}_{j}+\sum_{q} \beta_{q} \text { Quarter }_{j}+\right. \\
b_{\text {AreaYear }_{j}} \text { AreaYear }_{j}+b_{\text {QuarterYear } \left._{j} \text { QuarterYear }_{j}\right)} \\
b_{\text {AreaYear }_{j}} \sim N\left(0, \sigma_{\text {AreaYear } \left.^{2}\right)},\right.  \tag{eq2}\\
b_{\text {QuarterYear }_{j}} \sim N\left(0, \sigma_{\text {QuarterYear } \left._{2}^{2}\right) .}\right.
\end{gather*}
$$

where $Y_{j}$ is the CPUE of gillnet set $j, \beta$ are categorical fixed effect terms for each area, year, gear, and quarter (as in equation 1), and $b$ are random effect intercept terms for area-year and quarteryear interactions. The AreaYear ${ }_{j}$ indicates that the area and year variables were concatenated into a single variable and considered as a random effect acting on the intercept, and likewise for QuarterYear ${ }_{j}$. The total cod CPUE model showed reasonable diagnostics (Figure 2.3.5).

## Step 3: Joining steps 1-2 to create a standardized coastal cod CPUE

The predicted proportion coastal cod, $\hat{\pi}_{y, q, a}$, and total cod CPUE, $\hat{Y}_{y, q, a}$, for each year $y$, quarter $q$, and area $a$ combination were calculated from the two models above and combined to estimate the standardized coastal cod CPUE index, $I_{y, q, a}$ :

$$
\begin{equation*}
I_{y, q, a}=\hat{\pi}_{y, q, a} * \hat{Y}_{y, q, a} \tag{eq3}
\end{equation*}
$$

The variance of $I_{y, q, a}$ was calculated as:

$$
\begin{equation*}
V\left(I_{y, q, a}\right)=\left(\hat{\pi}_{y, q, a}\right)^{2} V\left(Y_{y, q, a}\right)+\left(\hat{Y}_{y, q, a}\right)^{2} V\left(\pi_{y, q, a}\right) \tag{eq4}
\end{equation*}
$$

The resulting standardized coastal cod CPUE indices for areas 6 and 7 are shown in Figure 2.3.6, where quarters 3 and 4 are weighted equally. To combine the indices for areas 6 and 7, we weighted the indices in proportion to the surface area within $12 \mathrm{~nm}(0.587$ for area $6,0.413$ for area 7). The composite standardized CPUE index for coastal cod in the entire area between 62$67^{\circ} \mathrm{N}$, is shown in Figure 2.3.7 and Table 2.3.6.

### 2.3.5 Stochastic LBSPR (Table 2.3.1)

Given the uncertainty in parameters and the demonstrated sensitivity of the length-based spawning potential ratio (LBSPR) model to input parameters (Hordyk et al., 2015b, 2015a), the AFWG developed a stochastic LBSPR approach at the last benchmark (ICES 2021b), similar to the one developed for anglerfish (Section 9). While the LBSPR assumes that key life history parameters (growth, natural mortality, and maturity; described below) are known, our approach includes uncertainty and correlation in these parameters by fitting the LBSPR model 1000 times using randomly sampled values from their estimated distributions. Observation uncertainty of the annual length distributions is also included by random resampling (bootstrapping) the length data.

Most of the parameters estimated during the benchmark do not need to be re-evaluated on an annual basis and could be randomly generated using the reported mean and standard deviation values. However, we re-estimated each of the life history parameter models selected in the benchmark with data updated through 2021 (Table 2.3.1). All parameter estimates and residual diagnostics were very similar to those from the benchmark.

### 2.3.5.1 Growth ( $k$, $L_{\text {inf }}$ )

The von Bertalanffy growth model parameters Linf (asymptotic length) and $k$ (growth coefficient) were estimated using non-least-squares fit to length and decimal age data from the reference fleet. The value for the theoretical age when size is zero, $t_{0}=-0.0387$, was borrowed from northern coastal cod (north of $67^{\circ} \mathrm{N}$ ). To account for biases from size selective sampling, we used composite weights based on the product of 1) calibrated weights (size-selective ageing among individuals sampled for size; Perreault et al., 2020) and 2) weights correcting for size selectivity-at-age in the catch (loosely based on model 1 in Taylor et al., 2005), using selectivity parameters estimated using LBSPR and parameters borrowed from northern coastal cod.

### 2.3.5.2 Natural mortality ( $M$ )

One of the most critical parameters for the performance of LBSPR is $M / k$. For southern coastal cod we had a reasonable estimate of $k$ but no $a$ priori information on $M / k$. The benchmark evaluated four methods of estimating $M$ based on life history and selected the size-varying $M$ following Lorenzen (1996) due to its consistency with cannibalism-driven mortality in the partially sympatric NEA cod and that it estimated similar SPR and $F / M$ to assuming $M=0.2$.

### 2.3.5.3 Maturity ( $L M_{50}, L M_{95}$ )

The maturity parameters $L M_{50}$ and $L M_{95}$ (length at $50 \%$ and $95 \%$ maturity) were estimated by fitting a binomial GLM with covariate length to yearly bootstrapped maturity data from the autumn coastal survey. All data north of $62^{\circ} \mathrm{N}$ were used because biological samples from the area between $62-67^{\circ} \mathrm{N}$ were scarce. For consistency with the choices made for the northern stock, resting individuals (stage 4) were considered mature.

Table 2.3.1. Life history parameter distributions estimated using data through 2021, used as inputs in the LBSPR model. Other required LBSPR parameter values not included here were left at their default values.

| Parameter | Mean (sd) | Description |
| :---: | :---: | :---: |
| M | $\begin{aligned} & 0.230 \\ & (0.001) \end{aligned}$ | Natural mortality (year ${ }^{-1}$ ) at asymptotic length ( $L_{\text {inf }}$ ). Size-varying $M$ following Lorenzen (1996) fit to resampled reference fleet commercial sampling data. |
| $M_{\text {pow }}$ | $\begin{aligned} & 0.959 \\ & (0.005) \end{aligned}$ | aka exponent c, eqn. 17 in Hordyk et al. (2016): parameterization of the size-varying $M$ in LBSPR, following Lorenzen (1996) fit to resampled reference fleet commercial sampling data. |
| $k$ | $\begin{aligned} & 0.255 \\ & (0.003)^{*} \end{aligned}$ | von Bertalanffy growth coefficient |
| $M / k$ | $\begin{aligned} & 0.900 \\ & (0.007) \end{aligned}$ | $M / k$ at $L_{\text {inf }}$, derived from the above estimates |
| $L_{\text {inf }}$ | $\begin{aligned} & 94.1 \\ & (0.455)^{*} \end{aligned}$ | Asymptotic length (cm) as defined in the von Bertalanffy growth function |
| $t_{0}$ | -0.0388 | Theoretical age when length $=0$ in the von Bertalanffy growth function. Not used in the LBSPR model, but used in the estimation of $k$ and $L_{\text {inf }}$ (above). Borrowed from northern coastal cod. |
| CVL ${ }_{\text {inf }}$ | $\begin{aligned} & 0.155 \\ & (0.001) \end{aligned}$ | Coefficient of variation of $L_{\text {inf, }}$, encompasses all inter-individual growth variability of LBSPR. The values used are borrowed from northern coastal cod, estimated and randomly generated on the log scale (mean $=-1.862$; s.d. $=0.0039$ ). |
| $L M_{50}$ | $\begin{aligned} & 62.8 \\ & (1.842)^{\dagger} \end{aligned}$ | Length (cm) at 50\% maturity. Estimated from resampled coastal survey data (2010-2021, all data north of $67^{\circ} \mathrm{N}$ ) using a binomial glm. |
| LM ${ }_{95}$ | $\begin{aligned} & 79.6 \\ & (3.816)^{\dagger} \end{aligned}$ | Length (cm) at 95\% maturity. Estimated from resampled coastal survey data (2010-2021, all data north of $67^{\circ} \mathrm{N}$ ) using a binomial glm. |

*randomly generated preserving the correlation structure between $k$ and Linf using a multinormal distribution. ${ }^{\dagger}$ pairs (LM50, LM95) estimated from a same bootstrapped dataset and year drawn together to preserve the correlation between the two parameters and avoid using a parameterization based on the distribution of $\Delta L M=L M 95-$ LM50.

### 2.3.5.4 Length distribution resampling

The LBSPR model is fitted to 1000 bootstrapped length data and parameter sets. While input parameters were randomly generated/drawn as per Table 2.3.1, the generation of the randomized datasets is twofold:

1. random attribution of unclassified individuals as coastal and NEA cod, using a binomial random generator based on the GAM,
```
gam(is_coastal ~ s(length) + factor(area) * factor(year) + factor(quarter) +
    factor(gear), family=binomial(link = "logit"))
```

2. bootstrap of the length composition within each year, i.e. draw the number of individuals sampled within each year of data from step 1, with replacement.
For each of the 1000 randomized data and parameter sets, the LBSPR model estimates SPR, F/M, and the lengths at $50 \%$ and $95 \%$ selectivity, SL50 and SL95.

### 2.3.6 Results of the assessment (Figure 2.3.6-Figure 2.3.13)

### 2.3.6.1 Standardized CPUE index

In recent years, the standardized CPUE index for coastal cod based on the reference fleet gillnet data has generally increased in area 6 (northern subarea, $64-67^{\circ} \mathrm{N}$ ) and decreased in area 7 (southern subarea, $62-64^{\circ} \mathrm{N}$; Figure 2.3.6). The composite CPUE index combining areas 6 and 7 decreased from 2007-2013 and has increased since 2013, with large uncertainty ( $95 \%$ CIs extend to 0 in all years; Figure 2.3.7). The composite CPUE index in 2020-2021 was higher than from 2017-2019, and so the " 2 over 3 " ratio that largely determines the catch advice increased from last year's assessment (red lines in Figure 2.3.7). CPUE in 2020-2021 was similar to 2007-2008, the beginning of the time-series.

### 2.3.6.2 Effort and CPUE from official landings statistics

We have also calculated CPUE from the full fleet, although this is less controlled for fishing behaviour and uses a less precise measure of effort than the reference fleet CPUE. Still, it is valuable to consider because it covers the entire commercial fleet instead of just a few boats in the reference fleet.
Calculating fishing effort for the full fleet is much less precise than for the reference fleet, where we can calculate kg cod caught per gillnet per day. The number of sales notes has been shown to give an overestimation of the fishing effort, since a trip can give several sales notes by splitting the entire trip catch into several sales, each with its own sales note. We therefore consider a "trip" by combining the vessel's "Registration mark" in the sales note statistics with "Last catch date", and define effort as the number of sales note trips.

| Vessel size group | 2018 |  | 2019 |  | 2020 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number of trips | Landed round weight ( t ) | Number of trips | Landed round weight ( t ) | Number of trips | Landed round weight ( t ) |
| LG1: (blank) | 680 | 29 | 605 | 30 | 603 | 33 |
| LG2: < 11 m | 4203 | 229 | 3814 | 191 | 4311 | 298 |
| $\begin{aligned} & \text { LG3: 11-14.99 } \\ & \mathrm{m} \end{aligned}$ | 1107 | 129 | 1221 | 145 | 1125 | 114 |
| $\begin{aligned} & \text { LG4: 15-20.99 } \\ & \mathrm{m} \end{aligned}$ | 89 | 24 | 99 | 20 | 71 | 19 |
| $\begin{aligned} & \text { LG5: 21-27.99 } \\ & \mathrm{m} \end{aligned}$ | 3 | 2 | 1 | 1 | 32 | 15 |
| LG6: >= 28 m | 1 | 3 | 1 | 0 | 8 | 1 |

The table above shows the number of trips and cod landings (round weight in tonnes) from inside 12 nautical miles during the second half-year during 2018-2020, per vessel size group, all gears. This shows that the vessel size groups $<11$ and $11-14.99 \mathrm{~m}$, represented by the coastal reference fleet (Section 2.3.3), are responsible for most of the effort and cod landings. The 9-15 m vessels in the reference fleet represent the gear and vessel size group responsible for about $60 \%$ of the total annual cod commercial catches in the stock area, and $88 \%$ of the effort (fishing trips) and $86 \%$ of cod catches in the second half of the year.

Figures 2.3.8 and 2.3.9 show the effort and CPUE from official landings statistics from 2007-2020. The recent gillnet CPUE trends differ by vessel size group, with some increasing and some decreasing (Figure 2.3.9).

### 2.3.6.3 Stochastic LBSPR outputs and interpretation

Between 2010-2021, the mean SPR fluctuated between 20 and $35 \%$, with an overall downward trend (Figure 2.3.10). In most years SPR was estimated below common target values (30-40\%) and in 2019-2020 SPR was near the limit reference point (generally accepted to be $20 \%$ in the absence of further information on the stock dynamics; ICES 2018; Prince et al., 2020; Mace and Sissenwine, 1993). SPR in 2021 was estimated as 0.25 ( $95 \%$ CI: 0.21-0.29). In all years 2010-2021, the relative fishing mortality $\mathrm{F} / \mathrm{M}$ was estimated above the value which achieve long-term SPR $=40 \%$, or the more usual proxy $\mathrm{F} / \mathrm{M}=1$ (Figure 2.3.11). $\mathrm{F} / \mathrm{M}$ in 2021 was estimated as $1.28(95 \%$ CI: 1.07-1.52). Concomitant with the decrease in SPR, the size-based indicators $\mathrm{L}_{\text {max5 }}$ (mean length of the largest $5 \%$ of individuals) and $\bar{L}$ (mean length) also declined from 2010-2021 (Figure 2.3.12). These all together depict a somewhat depleted and worsening stock status.

In the absence of clear information on the stock-recruitment relationship, a more legitimate reference point cannot be estimated and even a SPR of $30 \%$ should be considered as a potentially non-precautionary level, with SPR $=40 \%$ preferred as BMSy proxy (Clark, 2002; Hordyk et al., 2015a). In conformity with ICES guidelines (ICES, 2018) and commonly used SPR-based proxies (Prince et al., 2020; Mace and Sissenwine, 1993), the corresponding limit reference point (proxy for $\mathrm{Blim}_{\lim }=\mathrm{BMSY}_{\mathrm{M}} / 2$ ) should be $\mathrm{SPR}=20 \%$. A simulation function in the LBSPR package also allowed us to estimate $\mathrm{F}_{\mathrm{SPR} 40 \%} / \mathrm{M}=0.81$ ( $95 \% \mathrm{CI}$ : $0.74-0.88$ ), which is the $\mathrm{F} / \mathrm{M}$ that leads to $\mathrm{SPR}=40 \%$ given equilibrium and the parameter values (Figure 2.3.11). This also produces the expected mean length at $S P R=40 \%, \bar{L}_{S P R=40 \%}$, which could be evaluated for use as a target/reference length in the fishing pressure proxy part of the ICES ' rfb ' rule (Figure 2.3.12).

### 2.3.6.4 Catch lengths in relation to maturity

Averaged across all years, the length at which $50 \%$ of southern coastal cod are mature, $L M_{50}$, was estimated as 62.8 cm ( $95 \% \mathrm{CI}$ : 59.4-66.9). This is substantially higher than the minimum legal size $(44 \mathrm{~cm})$ or the estimated length at $50 \%$ selectivity ( $S_{50}$; Figure 2.3.13). In addition, $S_{50}$ has decreased in the last decade, i.e. the fishery is catching smaller fish, closer to the minimum size. This has led the proportion of immature fish in the catch to increase from about $25 \%$ in 2010 to about 50\% in 2021 (Figure 2.3.14).

### 2.3.6.5 Total mortality ( $Z$ ) from catch curves

Since catch numbers-at-age data are available for this stock for a longer period (1994-2021; Tables 2.3.2 and 2.3.3) it is possible to estimate the total mortality from catch-curve analyses. The assumptions usually made for catch-curve analysis are that (1) there are no errors in the estimation of age composition, (2) recruitment is constant or at least varies without trend over time, (3) Z is constant over time and across ages, and (4) above some determined age, all animals are equally available and vulnerable to the fishery and the sampling process. The catch-curve estimates a single total mortality rate for all years/ages that compose its synthetic cohort, and this total mortality estimate is generally similar to the average of the true total mortality rate.

We estimated the average total mortality of ages 5-14 for the years 1994-2020, not updated with 2021 data. Note that Tables 2.3.2 and 2.3.3 only present data up to age group 10+ but catch-at-age data were available to the AFWG up to age group 15+. Figure 2.3.15 shows a very stable level of the total mortality during the entire time-series, varying without trend around the long-term average of $Z=0.75$. With $M=0.23$ (Table 2.3.1), this implies fishing mortality around 0.5.

### 2.3.6.6 Additional indices: coastal survey

The last benchmark considered and rejected indices calculated from the main survey covering coastal cod, the autumn coastal survey (Nocoast-Aco-4Q), due to concerns about poor and inconsistent coverage south of $67^{\circ} \mathrm{N}$ (WD33 in ICES 2021b). The reference fleet CPUE index was used instead. The reviewers commented that it was "not entirely clear that this was justified" (ICES 2021b). Given the high uncertainty in the CPUE index ( $95 \%$ CIs extend to 0 in all years; Figure 2.3.7), we calculated swept-area indices from the coastal survey trawl data between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$ for comparison (methods described for northern coastal cod in Section 2.2.3). It is possible that the coastal survey data may not provide reliable abundance-at-age indices, yet still produce a useable aggregate (across ages) biomass index.

Three alternative swept-area indices from the coastal survey are shown in Figure 2.3.16: total age- $2+$ biomass, total numbers age- $2+$, and spawning-stock biomass. There are several notable differences from the reference fleet CPUE index: 1) the survey indices extend back to 2003, whereas the CPUE index starts in 2007; 2) the $95 \%$ CIs are much smaller for the survey indices; and 3) the survey indices are relatively stable from 2003-2013 and then decline from 2013-2021, whereas the CPUE index declines from 2007-2013 and then increases. The coefficient of variation (CV) of the CPUE index is $0.7-0.85$ in most years, and the survey indices CV is $0.2-0.4$ (Figure 2.3.17). The correlations between the CPUE and survey indices are negative, whereas the correlations between the survey indices and SPR estimated from the LBSPR model are positive (Figure 2.3.18). In contrast to the age-aggregated swept-area indices, the index-at-age probably is too uncertain to be useful (CVs > 0.3-0.4 for most ages and years; Figure 2.3.19).

Further exploration of how to produce indices from the coastal survey data is warranted. The survey index CVs reported here may not be reliable as they do not take into account variable spatial coverage by year. Still, the consistency between the survey indices and SPR, and the lower CV of the survey indices, indicates that an age-aggregated swept-area index calculated from the coastal survey may be useful for assessing southern coastal cod.

### 2.3.6.7 Additional indices: shallow water survey

IMR established a shallow water survey using small, passive meshed gear in 2013 in the hope that it would provide information on fish abundance in nearshore habitat not sampled by the main coastal survey, especially for young cod ages 1-3 (Eidset 2019; WD 13).

The shallow water survey appears to provide precise enough estimates of abundance-at-ages 13 to generate useful indices, with CVs between 0.15-0.20 (Figure 2.3.19). CVs for ages 0 and 4 were about 0.30 , and the CV for age 5 was 0.40 . The survey can reasonably track cohorts-the correlations from one age/year to the next were about $0.45-0.60$ for ages $0-5$, with the exception of age- 2 to age-3, which was about 0.15 (Figure 2.3.20). Indices for ages 2 and 3 were somewhat consistent between the coastal survey swept-area and the shallow water survey ( $r=0.82$ and 0.32 , respectively), but not for other ages.

Both surveys estimate declining trends for all ages 1-5 over the period 2013-2021, with the coastal survey estimating steeper declines for all ages (Figure 2.3.21). The coastal survey sweptarea indices-at-age were stable or increasing for all ages in the decade before the shallow water survey was initiated, 2003-2012 (Figure 2.3.21). For further details, see WD 13.

### 2.3.7 Comments to the assessment

The assessment remains rather uncertain. The reasons for this include highly uncertain data for the recreational catch and uncertainty in the catch split between Northeast Arctic cod and coastal cod, although the CPUE series is calculated for the second half of the year to minimize the mixing of the two stocks in the dataseries. The assessment also depends on the representativeness of the
coastal reference fleet gillnet CPUE index. Gillnets are responsible for most of the catches, and the $9-15 \mathrm{~m}$ vessels in the reference fleet represent the gear and vessel size category responsible for about $60 \%$ of the total annual cod commercial catches in the area, and $88 \%$ of the effort (fishing trips) and $86 \%$ of cod catches in the second half of the year. Still, the reference fleet CPUE increasing trend in recent years is not consistent with decreases in the SPR, coastal survey sweptarea index, or shallow water survey index.

ICES catch advice is based on the " rfb " rule for Category 3 stocks, which relies primarily on the reference fleet CPUE. While the reference fleet CPUE has increased since 2013, the SPR, coastal survey swept-area index, and shallow water survey index have decreased and are presented as additional information.

Priorities for more accurate future assessments are 1) better estimation of recreational catches, and 2) re-evaluation of available survey data that could be used as indices. Possible model improvements include 1) accounting for index uncertainty in the ' rfb ' rule, and 2) combining index and length data in one model.

### 2.3.8 Reference points

No biological reference points are established except the SPR and F/M reference levels often referred to in literature. See section 2.3.6.1 above.

### 2.3.9 Catch scenarios for 2023

The ICES Guidance for completing single-stock advice for category 3 stocks was applied (ICES, 2020, 2022). A standardized CPUE index from the coastal reference fleet ( $9-15 \mathrm{~m}$ vessel length) in coastal waters between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$ during quarters 3 and 4 , between 2007-2021, is used as the stock biomass index (Table 2.3.6). The advice is the previous year's catch advice multiplied by four modifiers: 1) ratio of the two latest index values (Index A) to the three preceding values (Index B), 2) length-based proxy of fishing pressure (f), 3) biomass safeguard (not applicable here), and 4) life history multiplier (m). The advice is estimated to have increased by more than $20 \%$ and thus the stability cap was applied. Discarding (of dead fish) is known to take place (2$5 \%$ in the commercial fishery and about $7 \%$ in the rod and line sector of the recreational fishery; Berg and Nedreaas, 2021), but ICES cannot quantify the corresponding catch.
The catch advice for 2023 is estimated to 9136 tonnes (Table 2.3.7). Assuming recreational catches at 4420 tonnes, this implies a commercial catch of no more than 4716 tonnes.

### 2.3.10 Management considerations

Applying the official ICES Guidance for catch advice results in an increase of $20 \%$. Several caveats should be considered:

- Uncertainty of the CPUE index used in the ' rfb ' rule is high, with $95 \%$ confidence intervals extending to 0 in all years (Figure 2.3.7). This is not taken into account when calculating the advice.
- The CPUE index increase is driven by area 6. The index is lower and has decreased in area 7 (Figure 2.3.6).
- The LBSPR results indicate fairly poor status: SPR $=0.25$ ( $95 \% \mathrm{CI}: 0.21-0.29$ ) and $\mathrm{F} / \mathrm{M}=$ 1.28 ( $95 \%$ CI: 1.07-1.52; Figures 2.3.10 and 2.3.11).
- Length-based indicators in the reference fleet data have declined over the past decade (Figures 2.3.12 and 2.3.13). Mean length has decreased from ca. 70.9 to 63.2 cm and the
length at $50 \%$ selectivity, i.e. first capture, has decreased from ca. 57.6 to 48.4 cm (averages 2010-2013 vs. 2018-2021).
- The minimum legal size ( 44 cm ) is well below the length at $50 \%$ maturity $(62.8 \mathrm{~cm})$. About half of the catch is immature, and this proportion has increased in the last decade (Figure 2.3.14).
- Commercial catches have decreased over the last 10-15 years while effort has probably remained stable or increased since 2013 (Figures 2.3.1, 2.3.8, and 2.3.9).
- The coastal survey swept-area and shallow water survey indices decreased from 20132021, the opposite trend as in the CPUE index (Figure 2.3.21).
ICES finds it difficult to give precise catch advice when the recreational catches, likely contributing more than $50 \%$ of total catches, are poorly estimated. A prerequisite for more accurate future assessments is a better estimation of the recreational catches.

The substantial and increasing proportion of immature fish in the catch is concerning, as well as the length at $50 \%$ selectivity being below the length at $50 \%$ maturity (Figures 2.3.13 and 2.3.14). Increasing the size of first capture closer to or above the size of maturity is worth considering, especially given the current difficulties of estimating catch and controlling fishing pressure with a quota (Prince and Hordyk, 2018).

Norwegian coastal cod is taken as part of a mixed fishery with Northeast Arctic cod (cod.27.1-2), from which it cannot be visually distinguished. Without the option of setting a direct TAC, the coastal cod stocks are managed by technical regulatory measures. Despite management actions, the previous management plan has not led to significantly reduced fishing mortality. A new plan is therefore required, with regulations better targeted to areas and seasons where catches of coastal cod are high. The split of the coastal cod stock in two units - one data rich in the north and one data poor in the south - combined with improved genetic stock identification techniques improves the spatial resolution of the assessment and allows development of more targeted management measures.

### 2.3.11 Rebuilding plan for coastal cod

The Norwegian Ministry of Fisheries is working on a new rebuilding plan. Fisheries scientists need to discuss with managers, how to facilitate rebuilding of the stock, evaluate rebuilding targets and measures to avoid high fishing pressure in areas with high fractions of coastal cod. Stronger restrictions are required in all areas where coastal cod is distributed.

### 2.3.12 Recent ICES advice

For the years 2004-2011, the advice was; No catch should be taken from this stock and a recovery plan should be developed and implemented.

For 2012, and later the advice has been to follow the rebuilding plan. The latest ICES advice strongly recommends a new rebuilding plan.
The catch advice for 2022 was 7613 tonnes (ICES, 2021a).

### 2.3.13 Figures and tables

Table 2.3.2. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. Estimated commercial landings in numbers ('000) at-age, and total tonnes by year.

|  | Age |  |  |  |  |  |  |  |  | Tonnes <br> Landed |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |  |
| 1994 | 1 | 7 | 111 | 288 | 361 | 279 | 158 | 71 | 112 | 6381 |
| 1995 | 3 | 32 | 210 | 399 | 491 | 467 | 267 | 114 | 96 | 8936 |
| 1996 | 2 | 64 | 242 | 384 | 304 | 253 | 130 | 36 | 44 | 6207 |
| 1997 | 2 | 117 | 171 | 212 | 189 | 185 | 131 | 44 | 33 | 4746 |
| 1998 | 20 | 177 | 446 | 496 | 332 | 109 | 82 | 22 | 23 | 6200 |
| 1999 | 3 | 116 | 313 | 308 | 255 | 123 | 53 | 66 | 26 | 5522 |
| 2000 | 2 | 242 | 697 | 411 | 159 | 57 | 51 | 17 | 37 | 5838 |
| 2001 | 2 | 94 | 423 | 457 | 304 | 149 | 52 | 17 | 86 | 5250 |
| 2002 | 9 | 88 | 360 | 409 | 441 | 138 | 52 | 12 | 16 | 6937 |
| 2003 | 23 | 204 | 237 | 571 | 398 | 380 | 112 | 22 | 53 | 8905 |
| 2004 | 5 | 112 | 334 | 260 | 400 | 232 | 139 | 35 | 26 | 6866 |
| 2005 | 2 | 65 | 381 | 522 | 445 | 262 | 122 | 37 | 19 | 8005 |
| 2006 | 10 | 48 | 308 | 617 | 565 | 179 | 99 | 54 | 50 | 8612 |
| 2007 | 11 | 154 | 364 | 497 | 379 | 113 | 51 | 23 | 29 | 7695 |
| 2008 | 31 | 103 | 893 | 665 | 195 | 265 | 69 | 38 | 47 | 9889 |
| 2009 | 1 | 224 | 663 | 259 | 311 | 107 | 74 | 42 | 20 | 7145 |
| 2010 | 5 | 115 | 400 | 434 | 245 | 260 | 50 | 36 | 45 | 7634 |
| 2011 | 3 | 59 | 310 | 484 | 267 | 194 | 65 | 36 | 35 | 7128 |
| 2012 | 28 | 113 | 268 | 501 | 317 | 279 | 73 | 36 | 36 | 8187 |
| 2013 | 5 | 54 | 239 | 214 | 248 | 169 | 80 | 27 | 16 | 5131 |
| 2014 | 1 | 56 | 166 | 390 | 265 | 226 | 79 | 43 | 38 | 6244 |
| 2015 | 21 | 149 | 257 | 229 | 263 | 120 | 69 | 37 | 41 | 5004 |
| 2016 | 1 | 83 | 248 | 313 | 206 | 200 | 121 | 66 | 83 | 5962 |
| 2017 | 13 | 73 | 275 | 279 | 157 | 97 | 70 | 24 | 34 | 4159 |
| 2018 | 9 | 57 | 131 | 298 | 255 | 141 | 90 | 36 | 32 | 4436 |
| 2019 | 4 | 34 | 85 | 101 | 128 | 121 | 77 | 21 | 24 | 2965 |


|  | Age |  |  |  |  |  |  | Tonnes |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $10+$ | Landed |
| 2020 | 1 | 46 | 164 | 140 | 144 | 79 | 84 | 37 | 16 | 3481 |
| 2021 | 34 | 173 | 198 | 228 | 114 | 78 | 50 | 27 | 33 | 3696 |

Table 2.3.3. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. Total estimated catch number ('000) at age, including recreational and tourist catches.

|  | Age |  |  |  |  |  |  |  |  | Tonnes <br> landed | Hereof <br> rec. ( t ) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |  |  |
| 1994 | 2 | 14 | 207 | 538 | 676 | 523 | 296 | 132 | 210 | 11937 | 5556 |
| 1995 | 4 | 51 | 341 | 647 | 797 | 757 | 433 | 184 | 155 | 14492 | 5556 |
| 1996 | 3 | 120 | 455 | 723 | 572 | 476 | 245 | 68 | 82 | 11687 | 5480 |
| 1997 | 5 | 253 | 369 | 456 | 407 | 399 | 283 | 95 | 72 | 10226 | 5480 |
| 1998 | 38 | 334 | 842 | 937 | 628 | 207 | 155 | 42 | 43 | 11718 | 5518 |
| 1999 | 5 | 226 | 610 | 600 | 497 | 240 | 103 | 128 | 51 | 10776 | 5254 |
| 2000 | 3 | 456 | 1311 | 773 | 299 | 107 | 96 | 32 | 69 | 10979 | 5140 |
| 2001 | 3 | 184 | 832 | 897 | 598 | 293 | 101 | 34 | 169 | 10315 | 5065 |
| 2002 | 15 | 153 | 627 | 711 | 768 | 240 | 91 | 22 | 28 | 12077 | 5140 |
| 2003 | 36 | 325 | 377 | 907 | 633 | 605 | 178 | 35 | 85 | 14159 | 5254 |
| 2004 | 9 | 194 | 581 | 451 | 695 | 403 | 242 | 60 | 45 | 11931 | 5065 |
| 2005 | 3 | 105 | 619 | 848 | 722 | 426 | 197 | 61 | 31 | 12994 | 4989 |
| 2006 | 16 | 76 | 484 | 968 | 888 | 282 | 156 | 84 | 79 | 13525 | 4913 |
| 2007 | 18 | 252 | 597 | 814 | 620 | 185 | 83 | 38 | 47 | 12609 | 4913 |
| 2008 | 46 | 153 | 1330 | 990 | 290 | 395 | 103 | 56 | 71 | 14727 | 4838 |
| 2009 | 1 | 375 | 1109 | 433 | 519 | 178 | 124 | 70 | 34 | 11945 | 4800 |
| 2010 | 7 | 187 | 651 | 706 | 398 | 423 | 81 | 58 | 74 | 12434 | 4800 |
| 2011 | 5 | 98 | 518 | 811 | 447 | 325 | 109 | 59 | 58 | 11928 | 4800 |
| 2012 | 45 | 179 | 425 | 795 | 502 | 442 | 115 | 57 | 58 | 12987 | 4800 |
| 2013 | 9 | 105 | 463 | 414 | 480 | 327 | 154 | 52 | 31 | 9931 | 4800 |
| 2014 | 1 | 100 | 293 | 690 | 469 | 400 | 140 | 76 | 68 | 11044 | 4800 |
| 2015 | 41 | 293 | 503 | 449 | 515 | 234 | 135 | 72 | 80 | 9804 | 4800 |
| 2016 | 2 | 151 | 448 | 566 | 371 | 360 | 218 | 120 | 150 | 10762 | 4800 |


|  | Age |  | $\mathbf{5}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{7}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ | landed |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | rec. (t)

Table 2.3.4. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. Commercial catch in 2021 by gear and Norwegian statistical fishing area. Both fishing areas lie within ICES Division 2.a.

| Gear | Area 06 | Area 07 | Total 62-67 $\mathbf{N}$ | \% by gear |
| :--- | :--- | :--- | :--- | :--- |
| Gillnet | 996.0 | 835.6 | 1831.6 | 49.8 |
| Longline/Handline | 291.9 | 248.1 | 540.0 | 14.7 |
| Danish seine | 0.1 | 1004.6 | 1004.7 | 27.3 |
| Trawl | 85.7 | 109.7 | 195.4 | 5.3 |
| Others | 1.2 | 103.2 | 104.4 | 2.8 |
| Total | 1374.9 | 2301.2 | 3676.1 |  |

Table 2.3.5. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. Mean weight at age in the catch.

| CWT | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1994 | 1.028 | 1.537 | 2.206 | 2.985 | 3.822 | 4.908 | 5.954 | 7.468 | 9.571 |
| 1995 | 0.845 | 1.392 | 1.950 | 2.603 | 3.649 | 4.811 | 6.076 | 7.404 | 10.566 |
| 1996 | 1.177 | 1.975 | 2.554 | 3.392 | 4.186 | 5.242 | 6.429 | 7.283 | 11.591 |
| 1997 | 1.348 | 2.004 | 2.611 | 3.439 | 4.282 | 5.387 | 6.563 | 7.467 | 10.828 |
| 1998 | 1.007 | 1.737 | 2.454 | 3.373 | 4.483 | 5.484 | 6.914 | 7.825 | 14.092 |
| 1999 | 1.459 | 2.231 | 2.927 | 3.800 | 4.854 | 6.032 | 7.009 | 8.257 | 12.088 |
| 2000 | 1.344 | 1.971 | 2.811 | 3.568 | 4.610 | 5.588 | 6.860 | 7.815 | 11.806 |
| 2001 | 0.565 | 0.981 | 1.533 | 2.250 | 3.129 | 4.160 | 5.375 | 6.722 | 16.118 |
| 2002 | 1.372 | 2.330 | 3.302 | 4.199 | 5.225 | 6.290 | 7.226 | 9.768 | 13.031 |
| 2003 | 1.312 | 2.143 | 2.962 | 3.899 | 4.702 | 5.648 | 6.616 | 7.425 | 11.376 |
| 2004 | 1.368 | 2.124 | 2.758 | 3.684 | 4.705 | 5.858 | 6.874 | 7.901 | 11.117 |
| 2005 | 1.488 | 2.332 | 2.990 | 3.701 | 4.562 | 5.637 | 6.699 | 7.703 | 10.364 |


| CWT | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 1.526 | 2.158 | 2.866 | 3.790 | 4.703 | 5.769 | 6.725 | 7.876 | 10.103 |
| 2007 | 1.613 | 2.295 | 3.285 | 4.337 | 5.744 | 7.105 | 8.397 | 9.991 | 12.359 |
| 2008 | 1.455 | 2.221 | 3.179 | 3.932 | 5.443 | 6.533 | 7.990 | 8.341 | 11.107 |
| 2009 | 1.667 | 2.135 | 3.234 | 4.207 | 5.279 | 6.527 | 7.568 | 7.606 | 11.305 |
| 2010 | 1.480 | 2.262 | 3.325 | 4.431 | 5.534 | 6.335 | 7.598 | 9.048 | 9.543 |
| 2011 | 1.381 | 2.127 | 3.172 | 4.263 | 5.511 | 6.510 | 8.012 | 9.032 | 11.065 |
| 2012 | 1.214 | 2.012 | 3.011 | 4.302 | 5.520 | 6.686 | 8.188 | 9.569 | 11.635 |
| 2013 | 1.269 | 2.027 | 3.092 | 4.024 | 5.268 | 6.370 | 7.524 | 8.918 | 12.241 |
| 2014 | 1.304 | 2.194 | 3.047 | 3.998 | 4.959 | 6.115 | 7.181 | 8.234 | 11.537 |
| 2015 | 1.219 | 1.832 | 2.726 | 3.797 | 4.627 | 5.845 | 7.009 | 8.195 | 10.981 |
| 2016 | 1.339 | 1.930 | 2.617 | 3.578 | 4.471 | 5.421 | 6.429 | 7.445 | 9.132 |
| 2017 | 1.529 | 2.022 | 2.750 | 3.663 | 4.543 | 5.612 | 6.542 | 7.489 | 9.678 |
| 2018 | 1.190 | 1.848 | 2.547 | 3.434 | 4.265 | 5.301 | 6.375 | 7.333 | 9.393 |
| 2019 | 1.662 | 2.283 | 3.120 | 3.895 | 4.840 | 5.796 | 6.743 | 7.737 | 9.548 |
| 2020 | 1.660 | 2.395 | 3.150 | 3.922 | 4.707 | 5.505 | 6.313 | 7.130 | 8.993 |
| 2021 | 1.325 | 2.049 | 2.827 | 3.696 | 4.692 | 5.835 | 6.755 | 7.672 | 11.064 |

Table 2.3.6. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. Composite standardized CPUE index from the coastal reference fleet during quarters 3 and 4, between 2007-2021. SE = standard error. $95 \%$ confidence intervals (CI) calculated using the approximation CPUE +/-1.96 SE.

| Year | CPUE index | SE | CI low (2.5\%) | Cl high (97.5\%) |
| :---: | :---: | :---: | :---: | :---: |
| 2007 | 0.30 | 0.27 | 0 | 0.84 |
| 2008 | 0.39 | 0.28 | 0 | 0.93 |
| 2009 | 0.25 | 0.17 | 0 | 0.57 |
| 2010 | 0.16 | 0.11 | 0 | 0.37 |
| 2011 | 0.24 | 0.18 | 0 | 0.60 |
| 2012 | 0.24 | 0.21 | 0 | 0.65 |
| 2013 | 0.06 | 0.04 | 0 | 0.13 |
| 2014 | 0.13 | 0.09 | 0 | 0.30 |
| 2015 | 0.26 | 0.18 | 0 | 0.62 |
| 2016 | 0.29 | 0.20 | 0 | 0.68 |


| Year | CPUE index | SE | Cl low (2.5\%) | CI high (97.5\%) |
| :--- | :--- | :--- | :--- | :--- |
| 2017 | 0.37 | 0.32 | 0 | 0.99 |
| 2018 | 0.14 | 0.11 | 0 | 0.36 |
| 2019 | 0.17 | 0.13 | 0 | 0.42 |
| 2020 | 0.39 | 0.31 | 0 | 1.00 |
| 2021 | 0.30 | 0.25 | 0 | 0.79 |

Table 2.3.7. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. Values used for calculating catch advice under the ICES "rfb" rule (ICES, 2022a).*

| Quantity | Value |
| :--- | :---: |
| Ay: Previous year catch advice | 7613 t |
| Stock biomass trend | 0.342 |
| Index A (average CPUE 2020-2021) | 0.225 |
| Index B (average CPUE 2017-2019) | 1.52 |
| $r:$ Stock biomass trend (ratio A/B) |  |

Fishing pressure proxy

| Mean catch length $\left(L_{\text {mean }}=L_{2021}\right)^{* *}$ | 67.7 cm |
| :--- | :--- |
| MSY proxy length $\left.\left(L_{\mathrm{F}=\mathrm{m}}\right)\right)^{* *}$ | 66.2 cm |
| f: Fishing pressure proxy relative to MSY proxy $\left(\mathrm{L}_{2021} / \mathrm{L}_{\mathrm{F}=\mathrm{M}}\right)$ | 1.02 |

Biomass safeguard

| Last index value $\left(I_{2021}\right)$ | 0.297 |
| :--- | :---: |
| Index trigger value $\left(I_{\text {trigger }}=I_{\text {los }} \times 1.4\right)$ | 0.058 |
| $b:$ index relative to trigger value, $\min \left\{I_{2021} / I_{\text {trigger }}, 1\right\}$ | 1 |

Precautionary multiplier to maintain biomass above $B_{\text {lim }}$ with $95 \%$ probability

| m: multiplier (generic multiplier based on life history) | 0.9 |
| :--- | :---: |
| rfb rule catch advice**** | 10643 t |
| Stability cap (+20\%/-30\% compared to Ay, only applied if $\mathrm{b} \geq 1$ ) | Applied |
| Discard rate | Not quantified |
| Catch advice for $\mathbf{2 0 2 3}$ | $\mathbf{9 1 3 6} \mathbf{t}$ |
| $\%$ advice change^ | $+20 \%$ |

* The figures in the table are rounded. Calculations were done with unrounded inputs, and computed values may not match exactly when calculated using the rounded figures in the table.
** Calculated as per ICES (2022a), only using lengths greater than Lc.
*** Equation A. 3 in Jardim et al. (2015).
**** $\left[\mathbf{A}_{\mathbf{y}} \times \mathbf{r} \times \mathbf{f} \times \mathbf{b} \times \mathbf{m}\right]$
^ Advice value for 2023 relative to the advice value for 2022.


Figure 2.3.1. Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$, Southern Norwegian coastal cod. Commercial landings and recreational catches. Recreational catches are fixed from 2009-2019 at 4800 tonnes and then reduced from 2020-2021 due to Covid-19 impacts on tourist fishing.


Figure 2.3.2. Estimated commercial landings of Northeast Arctic cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$.


Figure 2.3.3. Residual diagnostic plots for the final binomial model to differentiate coastal cod vs. NEAC. The panel on the left is a standard output from the residual diagnostics using the $R$ package DHARMa. The panel on the right plots the model standardized residuals against available covariates. Both panels indicate no significant issues with the final model.


Figure 2.3.4. Predicted probability of cod being classified as coastal instead of Northeast Arctic, based on the quarter (vertical panels), area (horizontal panels), and year (x-axis within each panel). The grey shaded polygon represents the 95\% confidence interval.


Figure 2.3.5. Residual diagnostic plots for the final CPUE model fitted to cod data in area 6 and 7, and quarters 3 and 4. Top panel left: standard output from the residual diagnostics using the R package DHARMa. Top panel right: normal QQplot. Bottom panel: model standardized residuals vs. available covariates. All panels indicate no significant (though some) issues with the final model.


Figure 2.3.6. Standardized reference fleet CPUE (kg per gillnet per day) index for coastal cod in areas 6 and 7 during quarters 3 and 4, between 2007-2021. The grey shaded polygon represents the $95 \%$ confidence interval (calculated using the approximation: mean +/-1.96 SE).


Figure 2.3.7. Composite reference fleet CPUE (kg cod per gillnet per day) index for southern Norwegian coastal cod, areas 6 and 7 combined. 95\% confidence intervals are calculated using the approximation: mean +/-1.96 SE. Red horizontal lines indicate the averages for the last 2 years (2020-2021) and previous 3 (2017-2019) used in the 'rfb' rule for catch advice (Table 2.3.7).

## Distinct count of trips $\mathrm{N}^{\circ}$, only length groups 2 and 3, linear y -axis scaling



Figure 2.3.8. Full commercial fleet fishing effort presented as the number of sales note trips for two boat sizes, LG2 = <11 m and $\mathrm{LG} 3=11-14.99 \mathrm{~m}$, for areas $62-67^{\circ} \mathrm{N}$ in the second half of the year. Left panel: all gears; right panel: gillnet only. Note different y -axes.


Figure 2.3.9. Full commercial fleet CPUE ( kg cod per sales note trip) per boat size (LG1-LG6) for area $62-67^{\circ} \mathrm{N}$ in the second half of the year. Left panel: all gears; right panel: gillnet only.


Figure 2.3.10. Spawning potential ratio (SPR) per year estimated by the length based spawning potential ratio (LBSPR) model. Mean (black line) and confidence intervals (dark shaded area, $95 \%$ interquartile range [IQR]), based on the stochastic LBSPR. The light shaded area delimits the SPR ${ }_{30 \%-40 \%}$ zone (common targets) and the red dashed horizontal line the SPR $_{20 \%}$ limit reference point.


Figure 2.3.11. Estimated fishing mortality relative to natural mortality ( $F / M$ ) per year estimated by the length based spawning potential ratio (LBSPR) model. Mean (black line) and confidence intervals (dark shaded area, 95\% IQR), based on the stochastic LBSPR. Red dashed line indicates $F / M=1$, and grey dashed line indicates $F_{40 \% \text { SPR }} / M$ (with $95 \%$ IQR, light shaded area), common target reference points.


Figure 2.3.12. Length-based indicators $L_{\text {max5\% }}$ and mean catch length ( $\bar{L}$ ) in relation to their reference points (mean and $95 \% \mathrm{CI}$ ). The reference points were estimated using the LBSPR simulation model together with the stochastic parameters detailed in Table 2.3.1 (mortality scenario following Lorenzen, 1996) and SPRs of 40\% and 100\% (unfished).


Figure 2.3.13. Length-based indicators, mean catch length $(\bar{L})$ and length at $50 \%$ selectivity $\left(S_{50}\right)$, in relation to the minimum legal size ( 44 cm ) and length at $50 \%$ maturity $\left(M_{50}\right) . M_{50}$ is estimated with uncertainty by bootstrapping data from the coastal survey. $\mathrm{S}_{50}$ is estimated by the length based spawning potential ratio (LBSPR) model, independently by year. $\bar{L}$ is calculated from the coastal reference fleet biological samples.


Figure 2.3.14. Proportion of the catch that is immature, southern Norwegian coastal cod. Linetype shows the proportion of cod in each year that are smaller than the yearly length at 50\% maturity ( $\mathrm{M}_{50}$, dotted line), yearly $\mathrm{M}_{50}$ times 1.1 (dashed line), and average $M_{50}$ ( $\bar{M}_{50}=62.8 \mathrm{~cm}$, solid line).


Figure 2.3.15. Total mortality $(Z)$ estimated from catch curves (average over ages 5-14 in commercial and recreational catches) 1994-2020.


Figure 2.3.16. Coastal survey trawl swept-area indices in relation to the reference fleet CPUE index and SPR, each standardized to its mean. Three alternative indices are calculated from the coastal survey: total age-2+ biomass (Survey B 2+), numbers age-2+ (Survey N 2+), and spawning-stock biomass (Survey SSB). Shading depicts 95\% confidence intervals.


Figure 2.3.17. Coefficient of variation (CV) from the coastal survey trawl swept-area indices, reference fleet CPUE index, and SPR. Three alternative indices are calculated from the coastal survey: total age-2+ biomass (Survey B 2+), numbers age-2+ (Survey N 2+), and spawning-stock biomass (Survey SSB).


Figure 2.3.18. Correlation between the coastal survey trawl swept-area indices, reference fleet CPUE index, and SPR. Three alternative indices are calculated from the coastal survey: total age-2+ biomass (Survey B 2+), numbers age-2+ (Survey N 2+), and spawning-stock biomass (Survey SSB).

- Coastal survey (swept area) - Shallow water survey


Figure 2.3.19. Coefficient of variation (CV) for additional survey indices-at-age, by year. Green: coastal survey swept-area (trawl). Orange: shallow water (garn ruse) survey. Dashed horizontal line indicates CV = 0.3, a commonly used upper threshold for considering indices to be informative on stock trends. See WD 13 for more details.


Figure 2.3.20. Correlation between the shallow water survey index-at-age in the previous age/year to the next, i.e. consistency, or the ability to track cohorts. Error bars indicate bootstrapped $95 \%$ confidence intervals. The delta-lognormal model was selected. See WD 13 for more details.


Figure 2.3.21. Southern Norwegian coastal cod indices-at-age from two available surveys, standardized to their means (horizontal dashed lines). Green: coastal survey swept-area (trawl). Orange: shallow water (garn ruse) survey. Lines are linear model fits from 2013-2021. See WD 13 for more details.

## 3 Northeast Arctic cod ${ }^{1}$

On 30 March 2022, all Russian participation in ICES was suspended. As a result of this decision, it is not possible to run ICES stock assessments or provide ICES advice for the Barents Sea stocks of NEA cod, NEA haddock, Sebastes mentella or Greenland Halibut, as management and data collection for these stocks are shared between Norway and Russia. There is therefore no stock assessment for NEA cod this year, but input data to the assessment are updated as far as possible.

The tables and figures updated are the following: Tables 3.1-3.5, 3.7, 3.13, Tables A1-A8, A13A15 and Figure 3.6b. The numbering of tables and figures is unchanged from AFWG 2021 so there are some 'holes' in the numbering.

Figures and tables can be found in the Data/NEA cod folder. Tables A9-A12 are not updated and will not be included in the report, but for completeness, they are also uploaded in the AFWG SharePoint folder.

### 3.1 Status of the fisheries

### 3.1.1 Historical development of the fisheries (Table 3.1)

From a level of about 900000 t in the mid-1970s, the total catch declined steadily to around 300000 t in 1983-1985 (Table 3.1). Catches increased to above 500000 t in 1987 before dropping to 212000 t in 1990, the lowest level recorded in the post-war period. The catches increased rapidly from 1991 onwards, stabilized around 750000 t in 1994-1997 but decreased to about 414000 t in 2000. From 2000-2009, the reported catches were between 400000 and 520000 t , in addition, there were unreported catches (see below). Catches have been above the long-term average since 2011 and have decreased from a peak of 986449 tonnes in 2014 to 693000 tonnes in 2019-2020 before increasing to 758000 tonnes in 2021. The fishery is conducted both with an international trawler fleet and with coastal vessels using traditional fishing gears. Quotas were introduced in 1978 for the trawler fleets and in 1989 for the coastal fleets. In addition to quotas, the fishery is regulated by a minimum catch size, a minimum mesh size in trawls and Danish seines, a maximum bycatch of undersized fish, closure of areas having high densities of juveniles and seasonal and area restrictions.

### 3.1.2 Reported catches prior to 2021 (Tables 3.1-3.4, Figure 3.1)

The provisional catch of cod in Subarea 1 and divisions 2.a and $2 . b$ for 2021 reported to the working group is 800427 t (including both NEA cod and NCC catches).
Reported catch figures used for the assessment of Northeast Arctic cod:
The historical practice (considering catches between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$ for the whole year and catches between $67^{\circ} \mathrm{N}$ and $69^{\circ} \mathrm{N}$ for the second half of the year to be Norwegian coastal cod) has been used for estimating the Norwegian landings of Northeast Arctic cod up to and including 2011 (Table 3.2). The catches of coastal cod subtracted from total cod catches in Subarea 1 and divisions 2.a and 2.b for the period 1960-2021 are given in Table 3.2. For 2012-2021 the

[^5]Norwegian catches have been analysed by an ECA-version designed for simultaneously providing estimates of catch numbers-at-age for each of the two stocks.

Coastal cod catches in 2021 for the southern and northern areas combined were 42044 tonnes using the current conversion factors between round and gutted weight, and this amount was as in previous years subtracted from the total cod catch north of $62^{\circ} \mathrm{N}$ to get the figure for NEA cod used in that assessment (Table 3.1 and 3.2). The figure for the total coastal cod catch in 2021 using the revised conversion factors, as decided at WKBARFAR 2021 and used in the coastal cod assessment was 32043 tonnes (Table 2.1a), which is $3.9 \%$ above the value using the current conversion factors.

These values for coastal cod are now inconsistent with the coastal cod catches presented in Chapter 2 , as the coastal cod catch time-series were revised at WKBARFAR, but not the NEA cod timeseries. At WKBARFAR, the proposal for revision of NEA cod catch dataseries was rejected, as Norwegian data for many years and age groups (especially ages 12+ in years prior to 2013) were changed considerably and the reason for this was not sufficiently explained. WKBARFAR recommended that when the revision of the historical Norwegian catch data are ready it should be submitted to ICES for review, ideally by a review attached to the AFWG.

The catch by area is shown in Table 3.1, and further split into trawl and other gears in Table 3.3. The distribution of catches by areas and gears in 2021 was similar to 2020. The nominal landings by country are given in Table 3.4.

There is information on cod discards (see section 0.4) but it was not included in the assessment because these data are fragmented and different estimates are in contradiction with each other. Moreover, the level of discards is relatively small in the recent period and the inclusion of these estimates in the assessment should not change our perception on NEA cod stock size.

In summer/autumn 2018, a Norwegian vessel caught 441 t of cod in the Jan Mayen EEZ, which is a part of ICES area 2a, mostly by longline. Cod is known to occasionally occur in this area, but rarely in densities which are suitable for commercial fisheries. The cod caught in this area in 2018 was large ( $65-110 \mathrm{~cm}$ ), and otolith readings and genetics both showed this cod to be a mix of Northeast Arctic and Icelandic cod. Norway did in 2019-2020 carry out an experimental longline fishery during four different periods each year in order to investigate further the occurrence of cod in this area in space and time as well as stock identity. The size distribution and genetic composition of the cod caught in this area in 2019-2021 were similar to that in 2018, although there was somewhat smaller cod ( $<65 \mathrm{~cm}$ ) in 2020-2021 than in 2019. Most of the cod caught in April-May 2019 was spawning or spent, while most cod caught in March 2020 had not started spawning. Cod spawning in this area has not been observed prior to 2019. Total catches in 2019 amounted to 628 t , in 2020 to 522 t and in 2021 to 146 t . The 2018 catches in this area were partly counted against the Norwegian TAC for cod north of $62^{\circ} \mathrm{N}$, while the 2019 and 2020 TAC for this area comes in addition to the Norwegian TAC for cod as agreed by JNRFC. There have been varying practice considering including those catches in the assessment, they were included in 2020 but the plan is to exclude them for all years in future assessments. Regulations for the fishery in this area for 2022 have not yet been decided upon.

### 3.1.3 Unreported catches of Northeast Arctic cod (Table 3.1)

In the years 2002-2008, certain quantities of unreported catches (IUU catches) have been added to the reported landings. More details on this issue are given in the Working group reports for that period.

There are no reliable data on the level of IUU catches outside the periods 1990-1994 and 20022008, but it is believed that their level was not substantial enough to influence historical stock assessment.

According to reports from the Norwegian-Russian analysis group on estimation of total catches the total catches of cod since 2009 were very close to officially reported landings.

### 3.1.4 TACs and advised catches for 2021 and 2022

The Joint Norwegian-Russian Fisheries Commission (JNRFC) agreed on a cod TAC of 885600 t for 2020 and in addition 21000 t Norwegian coastal cod. The total reported catch of 800427 t in 2021 was 106173 t below the agreed TAC. Since 2015 JNRFC has decided that Norway and Russia can transfer to next year or borrow from last year $10 \%$ of the cod country's quota. That may lead to some deviation between agreed TAC and reported catch. As an extraordinary measure due to expected underfishing of the TAC in 2021, JNRFC decided that it should be possible to transfer $15 \%$ of the TAC between 2021 and 2022.

The advice for 2022 given by ACOM in 2021 was 708480 t based on the agreed harvest control rule. The quota established by JNRFC for 2022 was set equal to the advice. In addition, the TAC for Norwegian Coastal Cod was set to the same value for 2022 as for 2021: 21000 t .

ICES will not give advice for this stock for 2023.

### 3.2 Status of research

### 3.2.1 Fishing effort and CPUE (Table A1, Figure 3.6a-c)

CPUE series of the Norwegian and Russian trawl fisheries are given in Table A1. Russian CPUE data for 2021 were not available. The data reflect the total trawl effort (Figure 3.6a), both for Norway and Russia. The Norwegian series is given as a total for all areas. Norwegian data for 2011-2021 are not necessarily compatible with data for 2007 and previous years. Norwegian CPUE declined from 2020 to 2021 and reached the lowest level in the 2011-2021 time-series (Figure 3.6b).

### 3.2.2 Survey results - abundance and size at age (Tables 3.5, A2-A14)

Some survey results for 2021 were revised since AFWG 2021, for a summary of this, see section 3.2.3.

### 3.2.2.1 Joint Barents Sea winter survey (bottom trawl and acoustics) Acronyms: BS-NoRu-Q1 (BTr) and BS-NoRu-Q1 (Aco)

Results from this survey were not available as Russian data have not been exchanged, but the survey was carried out as planned with good spatial coverage.

Before 2000 this survey was made without participation from Russian vessels, while in 20012005, 2008-2016 and 2018-2022 Russian vessels have covered important parts of the Russian zone. In 2006-2007 the survey was carried out only by Norwegian vessels. In 2007, 2016, 2021 and 2022 the Norwegian vessels were not allowed to cover the Russian EEZ. The method for adjustment for incomplete area coverage in 2007 is described in the 2007 report. The same method was used to adjust the 1997-1998 survey indices in the 2016 revision (Mehl et al. 2016). Table 3.5 shows areas covered in the time-series and the additional areas implied in the method used to adjust for missing coverage in the Russian Economic Zone. In 5 of the 8 adjusted years
(including 2021) the adjustments were not based on area ratios, but the "index ratio by age" was used. This means that the index by age for the covered area was scaled by the observed ratio between total index and the index for the same area observed in the years prior to the survey. The adjustments for 2017 were based on average index rations by age for 2014-2016. Adjustments were also made in 2020-2021 using the average index ratios by age for 2018-2019 and 2019-2020, respectively.

Regarding the older part of this time-series it should be noted that the survey prior to 1993 covered a smaller area (Jakobsen et al. 1997), and the number of young cod (particularly 1- and 2year old fish) was probably underestimated. Other changes in the survey methodology through time are described by Jakobsen et al. (1997), while the surveys for the years 2007-2012 and 20132018 are reported in Mehl et al. (2013, 2014, 2015, 2016, 2017a). Note that the change from 35 to 22 mm mesh size in the codend in 1994 is not corrected for in the time-series. This mainly affects the age 1 indices.

With the recent expansion of the cod distribution it is likely that in recent years the coverage in the February survey (BS-NoRu-Q1 (BTr) and BS-NoRu-Q1 (Aco)) has been incomplete, in particular for the younger ages. This could cause a bias in the assessment, but the magnitude is unknown. The 2014-2021 surveys covered considerably larger areas than earlier winter surveys, and showed that most age groups of cod (particularly ages 1 and 2 ) were distributed far outside the standard survey area. The bottom trawl survey estimates including the extended area for 2014-2021 were used in the tuning data separately from the same index before 2014, as decided at WKBARFAR 2021.

### 3.2.2.2 Lofoten acoustic survey on spawners Acronym: Lof-Aco-Q1

The estimated abundance indices from the Norwegian acoustic survey off Lofoten and Vesterålen (the main spawning area for this stock) in March/April are given in Table A4. A description of the survey, sampling effort and details of the estimation procedure can be found in Korsbrekke (1997). The 2022 survey results in biomass terms was 182 thousand tonnes, this is $21 \%$ below the 2021 level and the lowest since 2006.

### 3.2.2.3 Russian autumn survey Acronym: RU-BTr-Q4

Abundance estimates from the Russian autumn survey (November-December) are given in Table A9 (acoustic estimates) and Table A10 (bottom trawl estimates). The entire bottom trawl timeseries was in 2007 revised backwards to 1982 (Golovanov et al., 2007, WD3), using the same method as in the revision presented in 2006, which went back to 1994. The new swept-area indices reflect Northeast Arctic cod stock dynamics more precisely compared to the previous one catch per hour trawling. The Russian autumn survey in 2006 was carried out with reduced area coverage. Divisions 2a and 2 b were adequately investigated in the survey in contrast to Subarea 1, where the survey covered approximately $40 \%$ of the long-term average area coverage. The Subarea 1 survey indices were calculated based on actual covered area ( 40541 sq . miles). The 2007 AFWG decided to use the "final" year-class indices without any correction because of satisfactory internal correspondence between year-class abundances at age $2-9$ years according to the 2006 survey and ones due to the previous surveys.

This survey was not conducted in 2016, but was carried out in 2017, when $79 \%$ of the standard survey area was covered (Sokolov et al 2018, WD 11). The index shows a reliable internal consistence and it was decided to use it in the assessment. This survey was not carried out in 2018-2021 and will likely be discontinued.

### 3.2.2.4 Joint Ecosystem survey Acronym: Eco-NoRu-Q3 (Btr)

Swept-area bottom trawl estimates from the joint Norwegian-Russian ecosystem survey in Au-gust-September for the period 2004-2021 are given in Table A14. This survey normally covers the entire distribution area of cod at that time of the year.

In 2014 this survey had an essential problem with area coverage in the northwest region because of difficult ice conditions. In the area covered by ice in 2014 a substantial part of population was distributed during 2013 survey. So, based on those observations AFWG decided in 2015 to exclude 2014 year from that tuning series in current assessment. In 2016 there was incomplete coverage in the international waters and close to the Murmansk coast. An adjustment for this incomplete coverage was made based on interpolation from adjacent areas (Kovalev et al 2017, WD 12). At this time of the year, usually a relatively small part of the cod stock is found in the area which was not covered in 2016. In 2017 and 2019 the coverage was close to complete, although the far northeastern part of the survey area (west of the north island of Novaya Zemlya) was not covered due to military restrictions. In 2018, a large area in the eastern part of the Barents Sea was not covered Thus it was decided not to include 2018 data from this survey in the assessment.

The coverage in 2020 was less synoptic than usual, as explained in Section 0.6. As the survey indices from the BESS 2020 showed an unexplainable large decline compared to the 2019 indices, it was considered to exclude 2020 indices from this survey, but it was decided to keep them in and re-evaluate next year whether they should still be included in the assessment. The 2021 coverage was good, although as in several previous years, most of the international waters in the Barents Sea was not covered. The mentioned re-evaluation has not been carried out.

The survey indices are calculated both the BioFox and StoX calculation methods, and as in earlier years, the Biofox series was used in the tuning. A research recommendation from WKBARFAR was to unify these two methods for estimating indices from ecosystem survey. However, the benchmark decided to use weight at age from the StoX in calculations of weight at age used in the assessment.

### 3.2.2.5 Survey results - length and weight-at-age (Tables A5-A8, A11-A12, A15)

Length-at-age is shown in Table A5 for the Norwegian survey in the Barents Sea in winter, in Table A7 for the Lofoten survey and in Table A11 for the Russian survey in October-December. Weight-at-age is shown in Table A6 for the Norwegian survey in the Barents Sea in winter, in Table A8 for the Lofoten survey, Table A12 for the Russian survey in October-December and Table A15 for the BESS survey (calculated using StoX).

Length and weight at age in the Lofoten survey increased from 2021 to 2022 for age groups 5-6 and 8-11. The size at age in the BESS survey was about the same in 2021 as in 2020.

### 3.2.3 Revision of 2021 survey results

Some errors in StoX software were found in summer 2021, affecting the 2021 winter survey results (bottom trawl and acoustic) for cod and haddock and thus a revised assessment was carried out in September 2021 for both stocks (as described in the AFWG 2021 report executive summary). Also an error in calculating the $12+$ group for the bottom trawl survey for use in the tuning was corrected. After that some additional errors in StoX software have been found and corrected, final estimates for 2021 are in the survey report which is now published (Fall et al. 2022). In addition, the 2020 ecosystem survey indices and weight at age as well as the 2021 Lofoten survey indices and weight at age have been revised.

### 3.2.4 Age reading

The joint Norwegian-Russian work on cod otolith reading has continued, with regular exchanges of otoliths and age readers (see chapter 0.7). The results of fifteen years of annual comparative age readings are described in Yaragina et al. (2009). Zuykova et al. (2009) re-read old otoliths and found no significant difference in contemporary and historical age determination and subsequent length-at-age. However, age at first maturation in the historical material as determined by contemporary readers is younger than that determined by historical readers. Taking this difference into account would thus have effect on the spawning stock-recruitment relationship and thus on the biological reference points. The overall percentage agreement for the 2017-2018 exchange was $87.7 \%$ (WD 8, AFWG 2020). The main reason for cod ageing discrepancies between Russian and Norwegian specialists remains the same, representing the latest summer growth zone, and different interpretations of the false zones. The general trend is that the Russian readers assign slightly lower ages than the Norwegian readers compared to the modal age for all age groups. This is opposite of what we have seen in previous readings, where the Russian readers has tended to be slightly overestimating the age compared to the Norwegian readers. More details can be found in section 0.7.

The trend with bias in NEA cod age determination registered for some years of the period 19922018 between experts of both countries is a solid argument to continue comparative cod age reading between PINRO and IMR to monitor the situation. The German participant has expressed an intention to join the age reading cooperation in future.

### 3.3 Data available for use in assessment

Data for the period 1946-1983 are taken from the AFWG 2001 report (ICES CM 2001/ACFM:19) and were not revised at the WKBARFAR benchmark in 2021.

### 3.3.1 Catch-at-age (Table 3.6)

For 2021, age compositions from all areas were available from Norway, Spain and Germany. Russian data were not available and thus total catch-at-age was not calculated.

There is still a concern about the biological sampling from parts of the Norwegian fishery that may be too low. Also the split between NEA cod and coastal cod may be affected by the sampling coverage.

### 3.3.2 Survey indices available for use in assessment (Table 3.13, A13)

The following survey dataseries were available:

| Fleet <br> code | Name | Place | Season | Age | Years |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Fleet 15* | Joint bottom trawl survey | Barents Sea | Feb-Mar | 3-12+ | 1981-2013, 2014-2021 |
| Fleet 16 | Joint acoustic survey | Barents Sea+Lofoten | Feb-Mar | $3-12+$ | 1985-2021 |
| Fleet 18 | Russian bottom trawl surv. | Total area | Oct-Dec | $3-12+$ | 1982-2017 |
| Fleet 007 | Ecosystem surv. | Total area | Aug-Sep | $3-12+$ | $2004-2021$ |

*Survey indices for Fleet 15 were divided by two series (before and after 2014) in model tuning as decided at WKBARFAR 2021.

The tuning fleet file is shown in Table 3.13. Note that the joint acoustic survey (sum of Barents Sea and Lofoten acoustic survey indices) is given in Table A13.

Survey indices for Fleet 15 have been multiplied by a factor 100, while survey indices for Fleets 007, 16 and 18 have been multiplied by a factor 10 . This is done to keep the dynamics of the surveys even for very low indices, because some models (e.g. XSA) adds 1.0 to the indices before the logarithm is taken.

### 3.3.3 Weight-at-age (Tables 3.7-3.9, A2, A4, A6, A8, A12).

### 3.3.3.1 Catch weights

For 2021, weight-at-age in the catch for areas $1,2 a$ and $2 b$ was provided by Norway, Spain and Germany (Table 3.7). Russian data were not available and thus total weight at age was not calculated. For ages up to and including 11, observations are used. Following the WKBARFAR 2021 decision, weight at age in catch for the years 1983-present for ages 12-15+ are calculated by a cohort-based von Bertalanffy approach used to replace previous fixed values.

### 3.3.3.2 Stock weights

Weight at age in the stock for 2022 were not calculated as winter survey data were not available.
For ages 1-11 stock weights-at-age at the start of year y ( $\mathrm{W}_{\mathrm{a}, \mathrm{y}}$ ) for 1983-2021 are calculated combining, when available, weight at age from the Winter, Lofoten, Russian autumn and ecosystem surveys. The details are given in the stock annex. For ages 12-15+ a similar approach as for weight at age in the catch was used.

### 3.3.4 Natural mortality including cannibalism (Table 3.12, Table 3.17)

A natural mortality (M) of $0.2+$ cannibalism was used. Cannibalism is assumed to only affect natural mortality of ages 3-6.

2021 data are available and 2020 data have been updated, but tables with results based on these data are not included (Tables 3.12 and 3.17 in the 2021 AFWG report) as no assessment was done.

The method used for calculation of the prey consumption by cod described by Bogstad and Mehl (1997) is used to calculate the consumption of cod by cod for use in cod stock assessment. The consumption is calculated based on cod stomach content data taken from the joint PINRO-IMR stomach content database (methods described in Mehl and Yaragina 1992). On average about 9000 cod stomachs from the Barents Sea have been analysed annually in the period 1984-2021.

These data are used to calculate the per capita consumption of cod by cod for each half-year (by prey age groups $0-6$ and predator age groups $1-11+$. It was assumed that the mature part of the cod stock is found outside the Barents Sea for three months during the first half of the year. Thus, consumption by cod in the spawning period was omitted from the calculations.
An iterative procedure was applied to include the per capita consumption data in the SAM run. It is described in detail in Stock Annex.

For the cod assessment data from annual sampling of cod stomachs has been used for estimating cannibalism, since the 1995 assessment. The argument has been raised that the uncertainty in such calculations are so large that they introduce too much noise in the assessment. A rather comprehensive analysis of the usefulness of this was presented in Appendix 1 in the 2004 AFWG report. The conclusion was that it improves the assessment.

The data on cod cannibalism for the historical period (1946-1983) was included in assessment during the benchmark to make the time-series consistent (ICES 2015, WKARCT 2015). These estimates were based on hindcasted values of NEA cod natural mortality-at-ages 3-5 using PINRO database on food composition from cod stomach for the historical period (Yaragina et al. 2018).

### 3.3.5 Maturity-at-age (Tables 3.10-3.11, Tables 3.10-3.11)

Since data from the winter survey 2022 were not available, ogives for 2022 could not be calculated.

Historical (pre-1982) Norwegian and Russian time-series on maturity ogives were reconstructed by the 2001 AFWG meeting (ICES CM 2001/ACFM:19). The Norwegian maturity ogives were constructed using the Gulland method for individual cohorts, based on information on age at first spawning from otoliths. For the period 1946-1958 only the Norwegian data were available. The Russian proportions mature-at-age, based on visual examinations of gonads, were available from 1959.

Since 1982 Russian and Norwegian survey data have been used (Table 3.10). For the years 19852021, Norwegian maturity-at-age ogives have been obtained by combining the Barents Sea winter survey and the Lofoten survey. Russian maturity ogives from the autumn survey as well as from commercial fishery for November-February are available from 1984 until present. The Norwegian maturity ogives tend to give a higher percent mature-at-age compared to the Russian ogives, which is consistent with the generally higher growth rates observed in cod sampled by the Norwegian surveys. The percent mature-at-age for the Russian and Norwegian surveys have been arithmetically averaged for all years, except 1982-1983 when only Norwegian observations were used and 1984 when only Russian observations were used.
Russian data for the autumn survey for 2018 and later years were not available as the survey was not conducted. In WD15, 2019, updated correction factors to allow for this when calculating the combined maturity-at-age in 2019 were calculated, based on historical differences between Norwegian and Russian data. These correction factors were then applied to the Norwegian data for 2020-2021.

The approach used for calculating maturity-at-age is the same as previously used and consistent with the approach used to estimate the weight-at-age in the stock, except that no data from the BESS survey are used. However, since survey data, both abundance indices and proportion mature, have been revised, the entire time-series of ogives back to 1994 was revised at the benchmark. The proportions of mature cod for age 13-15 are set to 1 for the period 1984-present.

Maturity-at-age for cod has been variable the last five years, particularly for ages 6-9. According to the combined data, maturity-at-age decreased in 2015-2016, then increased, but decreased again from 2019 to 2021 (Table 3.11).

## 4 Northeast Arctic haddock ${ }^{1}$

### 4.1 Introductory note

On 30 March 2022 all Russian participation in ICES was suspended. As a result of this decision, it is not possible to run ICES stock assessments or provide ICES advice for the Barents Sea stocks of NEA cod, NEA haddock, Sebastes mentella or Greenland Halibut, as management and data collection for these stocks are shared between Norway and Russia. There is therefore no stock assessment for NEA haddock this year.

The following tables were updated: Tables 4.1-4.5. Except for these tables, the text, tables and figures are unchanged from last year's report.

The data folder at the SharePoint will be updated when more data becomes available.

### 4.2 Status of the fisheries

### 4.2.1 Historical development of the fisheries

Haddock is mainly fished by trawl as bycatch in the fishery for cod. Also, a directed trawl fishery for haddock is conducted. The proportion of the total catches taken by direct fishery varies between years. On average approximately $30 \%$ of the catch is with conventional gears, mostly longline, which in the past was used almost exclusively by Norway. Some of the longline catches are from a directed fishery, which is restricted by national quotas. In the Norwegian management, the quotas are set separately for trawl and other gears. The fishery is also regulated by a minimum landing size, a minimum mesh size in trawls and Danish seine, a maximum bycatch of undersized fish, closure of areas with high density/catches of juveniles and other seasonal and area restrictions.

The exploitation rate of haddock has been variable. The highest fishing mortalities for haddock have occurred at low to intermediate stock levels and historically show little relationship with the exploitation rate of cod, despite haddock being primarily caught as bycatch in the cod fishery. However, the more restrictive quota regulations introduced around 1990 have resulted in a more stable pattern in the exploitation rate.

The exceptionally strong year classes 2005-2006 contributed to the strong increase to all-time high stock levels and high levels in the last decade. Their importance in the catches is currently minimal.

### 4.2.2 Catches prior to 2021 (Table 4.1-Table 4.3, Figure 4.1)

The highest landings of haddock historically were 322 kt in 1973 . Since 1973 the highest catches observed were about 316 kt in 2012. In 2013-2015 the stock biomass started to decline and the landings in 2018, 2019 and 2020 were below 200 kt (Figure 4.1).

In 2006 it was decided to include reported Norwegian landings of haddock from the Norwegian statistical areas 06 and 07 (i.e. between $62^{\circ} \mathrm{N}$ and Lofoten Islands). These areas were not

[^6]previously included in the total landings of NEA haddock as input for this stock assessment (ICES CM 2006/ACFM:19; ICES CM 2006/ACFM:25).

Provisional official landings for 2020 are about 183 kt , which is $15 \%$ below agreed TAC ( 215 kt ).
Estimates of unreported catches (IUU catches) of haddock have been added to reported landings for the years from 2002 to 2008. Two estimates of IUU catches were available, one Norwegian and one Russian. At the benchmark in 2011 it was decided to base the final assessment on the Norwegian IUU estimates (ICES CM 2011/ACOM:38; Table 4.1).

We continue to include the estimates of IUU catches 2002-2008, but the IUU are assumed to be negligible for 2009-2020 and therefore set to zero.

### 4.2.3 Catch advice and TAC for 2021

The catch advice for 2021 was 233 kt and the Joint Norwegian-Russian Fisheries Commission set the TAC in accordance with the HCR. Furthermore, Russia and Norway can transfer the unused part of their own quota, restricted to a maximum of $10 \%$ of own quotas from 2020 to 2021.

### 4.3 Status of research

### 4.3.1 Survey results

Russia provided indices for 1982-2015 and 2017 for the Barents Sea trawl and acoustic survey (TAS) which was carried out in October-December (FLT01, RU-BTr-Q4). The survey was discontinued in 2018.

The Joint Barents Sea winter survey provides two index series used for tuning and recruitment forecast (bottom trawl: FLT02, NoRu-BTr-Q1 and acoustics: FLT04, NoRu-Aco-Q1). The survey area has been extended from 2014 with additional northern areas $(\mathrm{N})$ covered. The extended area is now included in total and standard survey index calculations for haddock (WKDEM 2020). Overall, this survey tracks both strong and poor year classes well. The indices from the Joint winter survey of cod and haddock in the Barents Sea 1994-2021 are given in WD 2. The spatial survey coverage in 2021 was relatively good. Note that since the AFWG was conducted, minor errors were discovered in the winter survey index for 2021 (both acoustic and bottom trawl). These had minimal ( $<1 \%$ ) impact on the assessment of SSB for NEA haddock. This report is not updated to account for correcting these errors.

Both the acoustic and swept indices of all ages were lower in 2021 compared to 2020.
The Joint Barents Sea ecosystem survey provides indices by age from bottom-trawl data (FLT007, Eco-NoRu-Q3 Btr) used for tuning and recruitment forecast. At the benchmark in 2011 it was decided to include this survey as tuning series. Tuning indices by age from the joint ecosystem survey are presented in WD 1 (2004-2020 except 2018). The survey coverage in 2020 was good, but the survey covered the eastern Barents Sea much later than the western Barents Sea (almost three months), which might have influenced the results in an unknown way. The distribution of haddock was reduced in 2020 compared 2019, especially on the Novaya Zemlya bank, where haddock was almost absent. The indices were much lower for the youngest and oldest haddock in 2020 compared to 2019.

### 4.4 Data used in the assessment

### 4.4.1 Catch-at-age (Table 4.4)

Age and length composition of the landings in 2020 were available from Norway and Russia in Subarea 1 and Division 2.b, and from Norway, Russia, and Germany in Division 2.a. The biological sampling of NEA haddock catches is considered good for the most important ages in the fisheries (see section 1).

Relevant data of estimated catch-at-age obtained from InterCatch for the period 2008-2020 and historical values from 1950-2007 is listed in Table 4.4.

### 4.4.2 Catch-weight-at-age (Table 4.5)

The mean weight-at-age in the catch was obtained from InterCatch as a weighted average of the weight-at-age in the catch for Norway, Russia, and Germany.

### 4.4.3 Stock-weight-at-age (Table 4.6)

Since 1983 the stock weights-at-age (Table 4.6) are calculated using the average of the weight-atage estimate from the Joint Barents Sea winter survey and the Russian bottom trawl survey. These averages are assumed to give representative values for the beginning of the year (see stock annex for details). However, the Russian bottom trawl survey has been discontinued and therefor stock weights-at-age were calculated using a correction factor (WKDEM 2020). Since the benchmark in 2006 stock weight at age has been smoothed (ICES 2006, see stock annex for details).

### 4.4.4 Maturity-at-age (Table 4.7)

Since the benchmark 2006, smoothed estimates were produced separately for the Russian autumn survey and the joint winter survey and then combined using arithmetic average. These averages are assumed to give representative values for the beginning of the year. However, the Russian bottom trawl survey has been discontinued and therefore stock weights-at-age were calculated using a correction factor (see WKDEM 2020 and stock annex).

### 4.4.5 Natural mortality (Table 4.8)

Natural mortality used in the assessment was 0.2 . For ages $3-6$ mortality predation by cod are added (see stock annex). For the period from 1984 and onwards actual estimates of predation by cod was used. For the years 1950-1983 the average natural mortality for 1984-2020 was used (age groups 3-6). Estimated mortality from predation by cod in this year's assessment is based on the 'final run' cod assessment. The proportion of F and M before spawning was set to zero.

### 4.4.6 Data for tuning (Table 4.9)

The following survey series are included in the data for tuning for SAM, the last age for all surveys is the plus group. Data are lacking (no survey) for FLT01 in 2016, and for FLT007 in 2018 (not included due to poor coverage).

| Name | ICES Acronym | Place | Season | Age | Year | prior weight |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FLT01: Russian bottom trawl | RU-BTr-Q4 | Barents <br> Sea | October-December | 3-8 | $\begin{aligned} & 1991- \\ & 2017 \end{aligned}$ | 1 |
| FLTO2: Joint Barents Sea surveyacoustic | BS-NoRU-Q1(Aco) | Barents <br> Sea | FebruaryMarch | 3-9 | $\begin{aligned} & 1993- \\ & 2021 \end{aligned}$ | 1 |
| FLT04: Joint Barents Sea surveybottom trawl | BS-NoRu-Q1 (BTr) | Barents <br> Sea | FebruaryMarch | 3-10 | $\begin{aligned} & 1994- \\ & 2021 \end{aligned}$ | 1 |
| FLT007: Joint Russian-Norwegian ecosystem autumn survey in the Barents Sea-bottom trawl | Eco-NoRu-Q3 (Btr) | Barents <br> Sea | August-September | 3-9 | $\begin{aligned} & 2004- \\ & 2020 \end{aligned}$ | 1 |

### 4.4.7 Changes in data from last year (Table 4.6-Table 4.7, Table 4.9)

At the benchmark (WKDEM 2020) it was decided that historic values (1950-1993) of stock weight and maturity should not be updated in the following years. Due to the smoothing procedure (see stock annex) the stock weight and maturity ate at age back to 1994 are updated every year.

Natural mortality includes cod predation for the ages 3-6. The data from 1984 and onwards are updated every year after the update of the cod assessment. This year, the change in consumption estimates back to 1984 were larger than usual due to the revision of the cod stock undertaken at the cod benchmark held in early 2021. The averages used for the historic period (1950-1983) were updated and used in the assessment.

### 4.5 Assessment models and settings (Table 4.10)

At the benchmark in 2020 it was decided to continue using the SAM model as the main model and XSA, with revised settings, will be used as additional model for comparison. This year the TISVPA model is also used as an additional model for comparison.

The SAM configuration was revised during the benchmark in 2020. The main changes were 1) to include age group 3 in the winter survey indices (Fleet 02 and 04 ), 2 ) include a plus group in all survey series (new option in SAM), 3) include a prediction variance link for the observation variances (new option in SAM, Breivik et al., in prep) 4) correlation structure in observation variance for the surveys (Berg and Nielsen, 2016).

The configuration, settings and tuning of SAM that were decided on during the benchmark (WKDEM 2020) were used in the current assessment. The configuration file is given in Table 4.10 and in the stock annex.

### 4.6 Results of the assessment (Table 4.11-Table 4.14 and Figure 4.1-Figure 4.3)

The dominating feature of the assessment is that the stock reached an all-time high level around 2011 due to the strong 2004-2006 year classes, and since declined (Table 4.11; Figure 4.1)

Fishing mortality has increased since 2013 (Table 4.12). The estimate of fishing mortality of main ages (4-7) in 2020 was 0.43 and above $\mathrm{F}_{\text {msY }}=0.35$.

The SSB has decreased since the peak in 2013, and the estimate for 2021201 kt and is still well above MSY Btrigger $=80 \mathrm{kt}$ (Figure 4.1).

Most of last year residuals are negative while catch observation close to predicted values, which means survey tends to underestimate stock. Retrospective estimates confirms that stock going down only based on last year surveys data (Figure 4.2 and Figure 4.3)

### 4.7 Comparison with last year's assessment (Figure 4.4)

The text table below compares this year's estimates with last year's estimates. Compared to last year's assessment the current estimates by SAM model of the total stock (TSB) and spawning stock (SSB) are lower for 2020. The F in 2019 is estimated a higher. Estimates for all ages except ages 4 and 5 (2015 and 2016 year classes) were reduced.

| Year of assessment , model | F (2019) | Numbers 2020 (ages) |  |  |  |  |  |  |  |  |  |  | $\begin{aligned} & \text { SSB } \\ & \text { (2020) } \end{aligned}$ | TSB <br> (2020) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13+ |  |  |
| 2020 SAM | 0.38 | 497 | 532 | 171 | 60 | 29 | 11 | 10 | 4 | 4 | 2 | 5 | 243 | 798 |
| 2021 SAM | 0.43 | 442 | 530 | 164 | 48 | 24 | 9 | 8 | 3 | 3 | 2 | 3 | 205 | 723 |
| $\begin{aligned} & \text { Ratio } \\ & \text { 2021/2020 } \end{aligned}$ | 1.1 | 0.9 | 1.0 | 1.0 | 0.8 | 0.8 | 0.8 | 0.8 | 0.9 | 0.7 | 1.0 | 0.7 | 0.8 | 0.9 |

### 4.8 Additional assessment methods (Table 4.15, Figure 4.5-Figure 4.6)

### 4.8.1 XSA (Figure 4.5)

The Extended Survivors Analysis (XSA) was used to tune the VPA by available index series. As last years, FLR was used for the assessment of haddock (see stock annex), and thus all results concerning XSA are obtained using FLR. The settings used were the same as set in the benchmark in 2015 (WKARCT 2015). The biomass estimates of XSA with these settings significantly deviated from estimates of main model SAM. During the WKDEM 2020 it was found that changing S.E. of the mean survivor estimates shrinkage F from 1.5 to 0.5 gives estimates of biomass dynamics close to SAM estimates. Furthermore, this change improved XSA retrospective pattern. At AFWG 2021 this comparison was also done and confirmed that usage of survivor estimates shrinkage 0.5 gave the similar result with SAM estimates.

The estimated consumption of NEA haddock by NEA cod is incorporated into the XSA analysis by first constructing a catch number-at-age matrix, adding the numbers of haddock eaten by cod to the catches for the years where such data are available (1984-2020). The summary of XSA stock estimates with shrinkage value 0.5 are presented in Table 4.15. A retrospective estimate for XSA gave same signals as for main model SAM (Figure 4.5).

### 4.8.2 TISVPA (Figure 4.6)

The TISVPA (Triple Instantaneous Separable VPA) model (Vasilyev, 2005; 2006) represents fishing mortality coefficients (more precisely, exploitation rates) as a product of three parameters: $\mathrm{f}(\text { year })^{*} \mathrm{~s}(\text { age })^{*} \mathrm{~g}$ (cohort). The generation-dependent parameters, which are estimated within the model, are intended to adapt traditional separable representation of fishing mortality to
situations when several year classes may have peculiarities in their interaction with fishing fleets caused by different spatial distribution, higher attractiveness of more abundant schools to fishers, or by some other reasons. To NEA haddock stock the TISVPA model was at benchmark group for arctic stocks (WKARCT) in 2015 and this year it was decided to apply to NEA haddock using the same data as SAM except that natural mortality values from cannibalism were taken from the SAM runs. All the input data, including catch-at-age, weight-at-age in stock and in catches, maturity-at-age were taken the same as for stock assessment by means of SAM. During AFWG 2021 the results of runs using the TISVPA model were presented in WD\#22. Generally biomass estimates of this model were higher than SAM estimates, which can be explained by different assumptions about indices catchability. A retrospective assessment for TISVPA shows same trends as for both another models (Figure 4.6).

### 4.8.3 Model comparisons (Figure 4.7)

Results from SAM, XSA and TISVPA are compared in Figure 4.7. Comparison of results of SAM, TISVPA and XSA with previous year settings shows that the models estimate similar trends. The TSVPA model is more flexible for settings than the others and considering a possible decrease in survey data consistency, it was attempted to do tuning of surveys not at abundance but to age proportions because the probable change in effective survey catchability.

### 4.9 Predictions, reference points and harvest control rules (Table 4.16-Table 4.21)

### 4.9.1 Recruitment (Table 4.16-Table 4.17)

SAM was used to estimate the recruitment-at-age 3 of the 2018 year class in 2021. The RCT3 program translation in R was used to estimate the recruiting year classes 2019-2020 in 2022 and 2023 with survey data from the ecosystem survey and winter survey. Input data and results are shown in Tables 4.16 and 4.17, respectively.

The text table below shows the recruitment estimates for the year classes 2000-2018 from assessments and RCT3 (shaded cells). Overall, there is a good agreement with the year-class strength estimate from RCT3 and the assessments, for the year classes 2014-2018, the correlation between the initial estimate from RCT3 and the estimate in SAM is $98 \%$. For the 2004-2017 year classes the estimate from SAM was on average $80 \%$ of the initial estimate, whereas the SAM estimate of the recruitment-at-age 3 of the 2018 year class was less than $50 \%$ from the initial estimate from RCT3 calculated in 2019.

| Year | Year of assessment, base model (XSA 2005-2014) |  |  |  |  |  |  |  |  |  | XSA | SAM | SAM | SAM | SAM | SAM | SAM | SAM |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Class | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| 2000 | 197 | 237 | 236 | 249 | 246 | 222 | 232 | 232 | 232 | 229 | 237 | 179 | 231 | 247 | 244 | 247 | 352 | 340 |
| 2001 | 176 | 219 | 224 | 257 | 245 | 237 | 241 | 239 | 239 | 236 | 247 | 184 | 239 | 222 | 218 | 220 | 268 | 260 |
| 2002 | 295 | 313 | 339 | 367 | 365 | 371 | 352 | 359 | 359 | 352 | 368 | 275 | 352 | 351 | 349 | 353 | 377 | 366 |
| 2003 | 156 | 183 | 135 | 161 | 171 | 185 | 189 | 183 | 186 | 181 | 197 | 169 | 208 | 165 | 161 | 164 | 161 | 158 |
| 2004 | 462 | 755 | 672 | 665 | 668 | 610 | 765 | 743 | 725 | 698 | 768 | 687 | 930 | 898 | 869 | 879 | 557 | 543 |
| 2005 |  | 521 | 731 | 943 | 975 | 1029 | 1193 | 1301 | 1317 | 1303 | 1415 | 996 | 1456 | 1330 | 1241 | 1251 | 1149 | 1113 |
| 2006 |  |  | 463 | 832 | 1036 | 811 | 1057 | 1187 | 1264 | 1267 | 1366 | 827 | 1254 | 1083 | 1027 | 1030 | 1063 | 1025 |
| 2007 |  |  |  | 202 | 208 | 212 | 284 | 330 | 370 | 384 | 411 | 211 | 355 | 307 | 305 | 308 | 249 | 241 |
| 2008 |  |  |  |  | 149 | 101 | 120 | 151 | 155 | 169 | 178 | 89 | 157 | 107 | 109 | 110 | 122 | 117 |
| 2009 |  |  |  |  |  | 303 | 315 | 320 | 345 | 357 | 363 | 230 | 351 | 294 | 291 | 293 | 356 | 340 |
| 2010 |  |  |  |  |  |  | 188 | 146 | 137 | 146 | 150 | 100 | 133 | 105 | 105 | 106 | 124 | 119 |
| 2011 |  |  |  |  |  |  |  | 483 | 513 | 482 | 398 | 298 | 397 | 340 | 329 | 332 | 425 | 411 |
| 2012 |  |  |  |  |  |  |  |  | 124 | 145 | 104 | 78 | 73 | 79 | 70 | 68 | 75 | 72 |
| 2013 |  |  |  |  |  |  |  |  |  | 394 | 290 | 197 | 235 | 184 | 174 | 177 | 219 | 213 |
| 2014 |  |  |  |  |  |  |  |  |  |  | 279 | 198 | 247 | 189 | 145.96 | 148 | 202 | 194 |
| 2015 |  |  |  |  |  |  |  |  |  |  |  |  | 422 | 398 | 333 | 336 | 384 | 368 |
| 2016 |  |  |  |  |  |  |  |  |  |  |  |  |  | 1067 | 933 | 930 | 875 | 822 |
| 2017 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 577 | 629 | 497 | 442 |
| 2018 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 344 | 294 | 154 |
| 2019 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 39 | 31 |
| 2020 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 95 |

### 4.9.2 Prediction data (Table 4.18, Figure 4.8)

The input data for the prediction are presented in Table 4.18.
Stock numbers for 2021-2022 at age 3 are taken from RCT3, and abundance-at-ages 3-13+ in 2020 from the SAM assessment. The average fishing pattern observed in 2018-2020 scaled to F in 2020 was used for distribution of fishing mortality-at-age for 2021-2023 (Figure 4.8). The proportion of M and F before spawning was set to 0 .

Input data to projection of weight at age in the stock, weight at age in the catch, maturity and mortality followed the stock annex.

### 4.9.3 Biomass reference points (Figure 4.1)

Biological and fisheries reference points for NEA haddock were last set following a thorough analysis as part of the WKNEAMP-2 (ICES, 2016) Harvest Control Rule evaluation in 2016. The revised model developed during the 2020 benchmark produced better fits to the data but only a small change in the reconstructed stock (WKDEM 2020). A brief analysis at WKDEM 2020 indicated that the reference points from the current model are very similar to the previously estimated values. Given the more thorough analysis at WKNEAMP-2 (ICES, 2016), this is taken as indicating that there was no evidence to deviate from the reference points set in 2016.

At the last benchmark (WKDEM 2020) it was proposed to keep $B_{\lim }=50000 \mathrm{t}$ and $\mathrm{B}_{\mathrm{pa}}=80000 \mathrm{t}$ with the rationale that $B_{l i m}$ is equal to $B_{l o s s,}$ and $B_{p a}=B_{\lim }{ }^{*} \exp \left(1.645^{*} \sigma\right)$, where $\sigma=0.3$. This gives a $95 \%$ probability of maintaining SSB above Blim considering the uncertainty in the assessments and stock dynamics. Bmsy trigger was proposed equal $\mathrm{B}_{\mathrm{pa}}$, $\mathrm{B}_{\text {trigger }}$ was then selected as a biomass that is encountered with low probability if $\mathrm{F}_{\text {MSY }}$ is implemented, as recommended by WKFRAME2 (ICES CM 2011/ACOM:33). Values of reference points compared with current stock values are reflected in Figure 4.1.

### 4.9.4 Fishing mortality reference points (Figure 4.1)

Biological and fisheries reference points for NEA haddock were last set following a thorough analysis as part of the WKNEAMP-2 (ICES, 2016) Harvest Control Rule evaluation in 2016. The revised model developed during the 2020 benchmark produced better fits to the data but only a small change in the reconstructed stock (WKDEM 2020). A brief analysis at WKDEM 2020
indicated that the reference points from the current model are very similar to the previously estimated values. Given the more thorough analysis at WKNEAMP-2 (ICES, 2016), this is taken as indicating that there was no evidence to deviate from the reference points set in 2016.

There is no standard method of estimating $\mathrm{F}_{\text {lim }}$ nor $\mathrm{F}_{\mathrm{pa}}$, and ACOM accepted to use geometric mean recruitment ( 146 million) and $\mathrm{Blim}_{\mathrm{lim}}$ as basis for the Flim estimate. Flim is then based on the slope of line from origin at $\mathrm{SSB}=0$ to the geometric mean recruitment ( 146 million) and $\mathrm{SSB}=\mathrm{B}_{\text {lim }}$. The SPR value of this slope give Flim value on SPR curve; $\mathrm{F}_{\mathrm{lim}}=0.77$ (found using Pasoft). Using the same approach as for $\mathrm{B}_{\mathrm{pa}} ; \mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\lim }{ }^{*} \exp \left(-1.645^{*} \sigma\right)=0.47$.

FMSY $=0.35$ has been estimated by long-term stochastic simulations. Values of reference points compared with current stock values are reflected in Figure 4.1.
The estimates of cod's consumption of haddock were revised following the cod benchmark in early 2021. At the AFWG 2021 meeting, the haddock $\mathrm{F}_{\text {MSY }}$ was checked with the new updated mortality estimates and found to still be valid and precautionary.

### 4.9.5 Harvest control rule

The harvest control rule (HCR) was evaluated by ICES in 2007 (ICES CM 2007/ACFM:16) and found to be in agreement with the precautionary approach. The agreed HCR for haddock with last modifications is as follows (Protocol of the $40^{\text {th }}$ Session of The Joint Norwegian Russian Fisheries Commission (JNRFC), 14 October 2011):

- TAC for the next year will be set at level corresponding to $\mathrm{F}_{\mathrm{m}} \mathrm{y}$.
- The TAC should not be changed by more than $+/-25 \%$ compared with the previous year TAC.
- If the spawning stock falls below $B_{p a}$, the procedure for establishing TAC should be based on a fishing mortality that is linearly reduced from $F_{M S Y}$ at $B_{p a}$ to $F=0$ at $\operatorname{SSB}$ equal to zero. At SSB-levels below $\mathrm{B}_{\mathrm{pa}}$ in any of the operational years (current year and a year ahead) there should be no limitations on the year-to-year variations in TAC.

As mentioned above $\mathrm{F}_{\lim }$ and $\mathrm{F}_{\mathrm{pa}}$ were revised in 2011. The new values of $\mathrm{F}_{\lim }=0.77$ and $\mathrm{F}_{\mathrm{pa}}=0.47$ are higher than the previous values ( 0.49 and 0.35 , respectively). In the 2012 meeting of the JNRFC the proposals of ICES were accepted, and the current HCR management is based on FMSY instead of $\mathrm{F}_{\mathrm{pa}}$. This corresponds to the goal of the management strategy for this stock and should provide maximum sustainable yield.

In 2014, JNRFC decided that from 2015 onwards, Norway and Russia can transfer to next year or borrow from last year maximum $10 \%$ of the country's quota. At its $45^{\text {th }}$ session in October 2015, the Joint Norwegian-Russian Fisheries Commission (JNRFC) decided that a number of alternative harvest control rules (HCRs) for Northeast Arctic haddock should be evaluated by ICES. This was done by WKNEAMP (ICES 2015/ACOM:60, ICES C. M. 2016/ACOM:47). Six HCRs for NEA haddock including the existing one were tested. At its $46^{\text {th }}$ session in October 2016, the JNRFC decided not to change the HCR.

### 4.9.6 Prediction results and catch options for 2021 (Table 4.19-Table 4.21)

The projection shows a slight increase in SSB from 203 kt in 2021 to 205 kt in 2022 (Table 4.19). TAC constraint F is used for 2021. The TAC for 2022 is established using the current one-year HCR, in accordance of the management plan. $\mathrm{F}_{\mathrm{MSY}}=0.35$ would give a quota for 2022 of 180 kt , this is a $23 \%$ decrease from the TAC and advice for 2021. Yield-per-recruit is given in Table 4.21.

Catch options for 2021 are shown in the text table below (weights in tonnes).

| Basis | Total catch (2022) | $\begin{aligned} & \text { F ages 4-7 } \\ & \text { (2022) } \end{aligned}$ | SSB (2023) | \% SSB change | \% TAC change ** | \% Advice change *** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |
| Management plan | 180003 | 0.35 | 201485 | -1.6 | -22.6 | -22.6 |
| Other scenarios |  |  |  |  |  |  |
| MSY approach: <br> $\mathrm{F}_{\mathrm{MSY}}$ | 180003 | 0.35 | 201485 | -1.6 | -22.6 | -22.6 |
| $F=0$ | 0 | 0 | 309362 | 51.1 | -100.0 | -100.0 |
| $F=F 2021$ | 214185 | 0.44 | 181739 | -11.2 | -7.9 | -7.9 |
| $\mathrm{F}_{\mathrm{pa}}$ | 227071 | 0.47 | 174372 | -14.8 | -2.4 | -2.4 |
| $\mathrm{F}_{\text {lim }}$ | 320921 | 0.77 | 122248 | -40.3 | 38.0 | 38.0 |

* SSB 2023 relative to SSB 2022.
** Catch in 2022 relative to TAC in 2021
*** Catch value for 2022 relative to advice value for 2021
Detailed information about expected catches by following HCR in 2022 and 2023 is given in Table 4.20. This catch forecast covers all catches. It is then implied that all types of catches are to be counted against this TAC. It also means that if any overfishing is expected to take place, the above calculated TAC should be reduced by the expected amount of overfishing.


### 4.9.7 Comments to the assessment and predictions (Figure 4.2-Figure 4.4 and Figure 4.9)

Haddock was benchmarked prior to last year's assessment (WKDEM 2020). The motivation for the benchmark was the poor retrospective (text table below).

| Retrospective bias (Mohn's Rho), 5-year peel | R | SSB | F | TSB |
| :--- | :--- | :--- | :--- | :--- |
| AFWG 2018 | $-3 \%$ | $24 \%$ | $-7 \%$ | $14 \%$ |
| AFWG 2019 | $-5 \%$ | $18 \%$ | $-7 \%$ | $7 \%$ |
| WKDEM 2020 | $-2 \%$ | $3 \%$ | $-3 \%$ | $1 \%$ |
| AFWG 2020 | $-4 \%$ | $-3 \%$ | $0 \%$ | $-5 \%$ |
| AFWG 2021 | $1 \%$ | $6 \%$ | $-7 \%$ | $3 \%$ |

The one step ahead residuals showed no clear pattern (Figure 4.2). This year, we also used model simulations and jitter analysis, as diagnostics of SAM model performance. No problems were detected.

By adding a new year of data, the analytical retrospective bias increased for SSB and F and decreased for R and TSB (Figure 4.3). The increased bias was mainly due to the low survey indices from the ecosystem survey 2020 and winter survey 2021, pulling the stock estimate down. Compared to last year's assessment, except for the ages 4 and 5, estimates of all ages in 2020 was
estimated lower at this year's assessment. This is mainly due to the low survey indices from the ecosystem survey of 2020 and winter survey 2021, but also due to update of the data, especially of the predation from cod, following the benchmark of the cod stock in 2021.

According to this year's assessment, the 2016 year class is the sixth strongest year class in the time-series back to 1950 and the 2017 year class is also above average, whereas the 2018 year class is weak. The 2019-2020 year classes are predicted to be well below average, the 2019 year class as the weakest since 1990.

As for the last two assessments F was above Fmsy in 2020 (Figure 4.4). This appears to be due to a too optimistic estimate of the stock in the assessment in 2019, and consequently too high TAC set for 2020. There was less fishing on youngest fish than initially assumed. Also, the weight in the catch in 2020 was considerably lower than was assumed in the forecast, especially for the 4year olds (Figure 4.9).

The retrospective trend indicates that the catch advice given in 2020 for 2021 is likely biased high. The catch in 2020 was $15 \%$ lower than TAC and the catch is expected to be below the TAC also in 2021, especially since the TAC in 2021 was higher than the 2020 TAC.

Table 4.1. Northeast Arctic haddock. Total nominal catch (t) by fishing areas.

| Year | Subarea 1 | Division 2.a | Division 2.b | un-reported ${ }^{2}$ ) | Total ${ }^{3}$ | Norwegian statistical areas 06 \& $07^{4}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | 125026 | 27781 | 1844 | - | 154651 | 6000 |
| 1961 | 165156 | 25641 | 2427 | - | 193224 | 4000 |
| 1962 | 160561 | 25125 | 1723 | - | 187409 | 3000 |
| 1963 | 124332 | 20956 | 936 | - | 146224 | 4000 |
| 1964 | 79262 | 18784 | 1112 | - | 99158 | 6000 |
| 1965 | 98921 | 18719 | 943 | - | 118583 | 6000 |
| 1966 | 125009 | 35143 | 1626 | - | 161778 | 5000 |
| 1967 | 107996 | 27962 | 440 | - | 136398 | 3000 |
| 1968 | 140970 | 40031 | 725 | - | 181726 | 3000 |
| 1969 | 89948 | 40306 | 566 | - | 130820 | 2000 |
| 1970 | 60631 | 27120 | 507 | - | 88258 | - |
| 1971 | 56989 | 21453 | 463 | - | 78905 | - |
| 1972 | 221880 | 42111 | 2162 | - | 266153 | - |
| 1973 | 285644 | 23506 | 13077 | - | 322227 | - |
| 1974 | 159051 | 47037 | 15069 | - | 221157 | 10000 |
| 1975 | 121692 | 44337 | 9729 | - | 175758 | 6000 |


| Year | Subarea 1 | Division 2.a | Division 2.b | un-reported ${ }^{\text {2 }}$ | Total ${ }^{3}$ | Norwegian statistical areas 06 \& 074 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1976 | 94054 | 37562 | 5648 | - | 137264 | 2000 |
| 1977 | 72159 | 28452 | 9547 | - | 110158 | 2000 |
| 1978 | 63965 | 30478 | 979 | - | 95422 | 2000 |
| 1979 | 63841 | 39167 | 615 | - | 103623 | 6000 |
| 1980 | 54205 | 33616 | 68 | - | 87889 | 5098 |
| 1981 | 36834 | 39864 | 455 | - | 77153 | 4767 |
| 1982 | 17948 | 29005 | 2 | - | 46955 | 3335 |
| 1983 | 5837 | 16859 | 1904 | - | 24600 | 3112 |
| 1984 | 2934 | 16683 | 1328 | - | 20945 | 3803 |
| 1985 | 27982 | 14340 | 2730 | - | 45052 | 3583 |
| 1986 | 61729 | 29771 | 9063 | - | 100563 | 4021 |
| 1987 | 97091 | 41084 | 16741 | - | 154916 | 3194 |
| 1988 | 45060 | 49564 | 631 | - | 95255 | 3756 |
| 1989 | 29723 | 28478 | 317 | - | 58518 | 4701 |
| 1990 | 13306 | 13275 | 601 | - | 27182 | 2912 |
| 1991 | 17985 | 17801 | 430 | - | 36216 | 3045 |
| 1992 | 30884 | 28064 | 974 | - | 59922 | 5634 |
| 1993 | 46918 | 32433 | 3028 | - | 82379 | 5559 |
| 1994 | 76748 | 50388 | 8050 | - | 135186 | 6311 |
| 1995 | 75860 | 53460 | 13128 | - | 142448 | 5444 |
| 1996 | 112749 | 61722 | 3657 | - | 178128 | 5126 |
| 1997 | 78128 | 73475 | 2756 | - | 154359 | 5987 |
| 1998 | 45640 | 53936 | 1054 | - | 100630 | 6338 |
| 1999 | 38291 | 40819 | 4085 | - | 83195 | 5743 |
| 2000 | 25931 | 39169 | 3844 | - | 68944 | 4536 |
| 2001 | 35072 | 47245 | 7323 | - | 89640 | 4542 |
| 2002 | 40721 | 42774 | 12567 | 18736/5310 | 114798/101372 | 6898 |
| 2003 | 53653 | 43564 | 8483 | 33226/9417 | 138926/115117 | 4279 |


| Year | Subarea 1 | Division 2.a | Division 2.b | un-reported ${ }^{\text {2 }}$ | Total ${ }^{3}$ | Norwegian statistical areas 06 \& $07^{4}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | 64873 | 47483 | 12146 | 33777/8661 | 158279/133163 | 3743 |
| 2005 | 53518 | 48081 | 16416 | 40283/9949 | 158298/127964 | 5538 |
| 2006 | 51124 | 47291 | 33291 | 21451/8949 | 153157/140655 | 5410 |
| 2007 | 62904 | 58141 | 25927 | 14553/3102 | 161525/150074 | 7110 |
| 2008 | 58379 | 60178 | 31219 | 5828/- | 155604/149776 | 6629 |
| 2009 | 57723 | 66045 | 76293 | 0 | 200061 | 4498 |
| 2010 | 62604 | 86279 | 100318 | 0 | 249200 | 3661 |
| 2011 | 86931 | 99307 | 123546 | 0 | 309785 | 4169 |
| 2012 | 90141 | 96807 | 128679 | 0 | 315627 | 3869 |
| 2013 | 68416 | 64810 | 60520 | 0 | 193744 | 4000 |
| 2014 | 61537 | 58320 | 57665 | 0 | 177522 | 3433 |
| 2015 | 75195 | 61567 | 57993 | 0 | 194756 | 3902 |
| 2016 | 78714 | 95140 | 59561 | 0 | 233416 | 3233 |
| 2017 | 94772 | 75455 | 57362 | 0 | 227589 | 2987 |
| 2018 | 80902 | 58522 | 51853 | 0 | 191276 | 4437 |
| 2019 | 87446 | 50967 | 36989 | 0 | 175402 | 2812 |
| $2020{ }^{1)}$ | 98341 | 57397 | 26730 | 0 | 182468 | 3196 |
| 2021 ${ }^{1)}$ | 107907 | 58097 | 37025 | 0 | 203118 | 2363 |

1) Provisional figures
2) Figures based on Norwegian/Russian IUU estimates. From 2009, IUU estimates are made by a Joint Russian-Norwegian analysis group under the Russian-Norwegian Fisheries Commission.
3) In 2002-2008, the Norwegian IUU estimates were used in final assessment.
4) Included in total landings and in landings in region 2.a.

Table 4.2. Northeast Arctic haddock. Total nominal catch ('000 t) by trawl and other gear for each area.

|  | Subarea 1 |  | Division 2.a | Division 2.b | Unreported ${ }^{2}$ |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | Trawl | Others | Trawl | Others | Trawl | Others |  |
| 1967 | 73.7 | 34.3 | 20.5 | 7.5 | 0.4 | - | - |
| 1968 | 98.1 | 42.9 | 31.4 | 8.6 | 0.7 | - | - |
| 1969 | 41.4 | 47.8 | 33.2 | 7.1 | 1.3 | - | - |


| Year | Subarea 1 |  | Division 2.a |  | Division 2.b |  | Unreported ${ }^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Trawl | Others | Trawl | Others | Trawl | Others |  |
| 1970 | 37.4 | 23.2 | 20.6 | 6.5 | 0.5 | - | - |
| 1971 | 27.5 | 29.2 | 15.1 | 6.7 | 0.4 | - | - |
| 1972 | 193.9 | 27.9 | 34.5 | 7.6 | 2.2 | - | - |
| 1973 | 242.9 | 42.8 | 14 | 9.5 | 13.1 | - | - |
| 1974 | 133.1 | 25.9 | 39.9 | 7.1 | 15.1 | - | - |
| 1975 | 103.5 | 18.2 | 34.6 | 9.7 | 9.7 | - | - |
| 1976 | 77.7 | 16.4 | 28.1 | 9.5 | 5.6 | - | - |
| 1977 | 57.6 | 14.6 | 19.9 | 8.6 | 9.5 | - | - |
| 1978 | 53.9 | 10.1 | 15.7 | 14.8 | 1 | - | - |
| 1979 | 47.8 | 16 | 20.3 | 18.9 | 0.6 | - | - |
| 1980 | 30.5 | 23.7 | 14.8 | 18.9 | 0.1 | - | - |
| 1981 | 18.8 | 17.7 | 21.6 | 18.5 | 0.5 | - | - |
| 1982 | 11.6 | 11.5 | 23.9 | 13.5 | - | - | - |
| 1983 | 3.6 | 2.2 | 8.7 | 8.2 | 0.2 | 1.7 | - |
| 1984 | 1.6 | 1.3 | 7.6 | 9.1 | 0.1 | 1.2 | - |
| 1985 | 24.4 | 3.5 | 6.2 | 8.1 | 0.1 | 2.6 | - |
| 1986 | 51.7 | 10.1 | 14 | 15.8 | 0.8 | 8.3 | - |
| 1987 | 79 | 18.1 | 23 | 18.1 | 3 | 13.8 | - |
| 1988 | 28.7 | 16.4 | 34.3 | 15.3 | 0.6 | 0 | - |
| 1989 | 20 | 9.7 | 13.5 | 15 | 0.3 | 0 | - |
| 1990 | 4.4 | 8.9 | 5.1 | 8.2 | 0.6 | 0 | - |
| 1991 | 9 | 8.9 | 8.9 | 8.9 | 0.2 | 0.2 | - |
| 1992 | 21.3 | 9.6 | 11.9 | 16.1 | 1 | 0 | - |
| 1993 | 35.3 | 11.6 | 14.5 | 17.9 | 3 | 0 | - |
| 1994 | 58.6 | 18.2 | 26.1 | 24.3 | 7.9 | 0.2 | - |
| 1995 | 63.9 | 12 | 29.6 | 23.8 | 12.1 | 1 | - |
| 1996 | 98.3 | 14.4 | 36.5 | 25.2 | 3.4 | 0.3 | - |
| 1997 | 57.4 | 20.7 | 44.9 | 28.6 | 2.5 | 0.3 | - |


| Year | Subarea 1 |  | Division 2.a |  | Division 2.b |  | Unreported ${ }^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Trawl | Others | Trawl | Others | Trawl | Others |  |
| 1998 | 26 | 19.6 | 27.1 | 26.9 | 0.7 | 0.3 | - |
| 1999 | 29.4 | 8.9 | 19.1 | 21.8 | 4 | 0.1 | - |
| 2000 | 20.1 | 5.9 | 18.8 | 20.4 | 3.7 | 0.1 | - |
| 2001 | 28.4 | 6.7 | 23.4 | 23.8 | 7 | 0.3 | - |
| 2002 | 30.5 | 10.2 | 19.5 | 23.3 | 12.5 | 0.1 | 18.7/5.3 |
| 2003 | 42.7 | 10.9 | 21.9 | 21.7 | 8.1 | 0.4 | 33.2/9.4 |
| 2004 | 52.4 | 12.5 | 27 | 20.5 | 11.5 | 0.6 | 33.8/8.7 |
| 2005 | 38.5 | 15 | 24.9 | 20.9 | 13 | 1.6 | 40.3/9.9 |
| 2006 | 40.1 | 11 | 22 | 25.3 | 30.1 | 3.2 | 21.5/8.9 |
| 2007 | 51.8 | 11.1 | 30.5 | 27.7 | 20.4 | 5.5 | 14.6/3.1 |
| 2008 | 46.8 | 11.6 | 30.9 | 29.3 | 24.9 | 6.3 | 5.8/- |
| 2009 | 49 | 8.8 | 40.1 | 25.3 | 67.1 | 7.8 | 0 |
| 2010 | 43.6 | 19 | 50 | 35.7 | 87 | 10.4 | 0 |
| 2011 | 55.8 | 31.1 | 61.1 | 38.9 | 107.7 | 14.3 | 0 |
| 2012 | 58.8 | 31.3 | 57.5 | 39.2 | 103.2 | 24.8 | 0 |
| 2013 | 40.1 | 28.3 | 37.7 | 26.9 | 52.1 | 8.1 | 0 |
| 2014 | 35.2 | 26.3 | 32.5 | 25.8 | 49 | 8.6 | 0 |
| 2015 | 49.1 | 26.1 | 34.6 | 27 | 48.5 | 9.4 | 0 |
| 2016 | 56.4 | 22.3 | 62.5 | 32.5 | 45.4 | 14.1 | 0 |
| 2017 | 65 | 29.8 | 50.7 | 24.7 | 47.1 | 10.3 | 0 |
| 2018 | 51.7 | 29.2 | 36.9 | 21.6 | 43.2 | 8.6 | 0 |
| 2019 | 53.9 | 33.5 | 30.4 | 20.4 | 31.0 | 5.9 | 0 |
| 2020 | 66.7 | 31.6 | 35.1 | 22.3 | 23.2 | 3.5 | 0 |
| 2021 ${ }^{1)}$ | 80.5 | 27.4 | 41.4 | 16.7 | 31.5.2 | 5.5 | 0 |

## 1) Provisional

2) Figures based on Norwegian/Russian IUU estimates.

Table 4.3 Northeast Arctic haddock. Nominal catch (t) by countries. Subarea 1 and divisions 2.a and 2.b combined. (Data provided by Working Group members).

| Year | Faroe Islands | France | GDR (- <br> 1990) <br> and <br> Green- <br> land <br> (1992-) | Germany | Norway ${ }^{4}$ | Poland | UK | Russia ${ }^{2}$ | Others | Total ${ }^{3}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | 172 | - | - | 5597 | 46263 | - | 45469 | 57025 | 125 | 154651 |
| 1961 | 285 | 220 | - | 6304 | 60862 | - | 39650 | 85345 | 558 | 193224 |
| 1962 | 83 | 409 | - | 2895 | 54567 | - | 37486 | 91910 | 58 | 187408 |
| 1963 | 17 | 363 | - | 2554 | 59955 | - | 19809 | 63526 | - | 146224 |
| 1964 | - | 208 | - | 1482 | 38695 | - | 14653 | 43870 | 250 | 99158 |
| 1965 | - | 226 | - | 1568 | 60447 | - | 14345 | 41750 | 242 | 118578 |
| 1966 | - | 1072 | 11 | 2098 | 82090 | - | 27723 | 48710 | 74 | 161778 |
| 1967 | - | 1208 | 3 | 1705 | 51954 | - | 24158 | 57346 | 23 | 136397 |
| 1968 | - | - | - | 1867 | 64076 | - | 40129 | 75654 | - | 181726 |
| 1969 | 2 | - | 309 | 1490 | 67549 | - | 37234 | 24211 | 25 | 130820 |
| 1970 | 541 | - | 656 | 2119 | 37716 | - | 20423 | 26802 | - | 88257 |
| 1971 | 81 | - | 16 | 896 | 45715 | 43 | 16373 | 15778 | 3 | 78905 |
| 1972 | 137 | - | 829 | 1433 | 46700 | 1433 | 17166 | 196224 | 2231 | 266153 |
| 1973 | 1212 | 3214 | 22 | 9534 | 86767 | 34 | 32408 | 186534 | 2501 | 322226 |
| 1974 | 925 | 3601 | 454 | 23409 | 66164 | 3045 | 37663 | 78548 | 7348 | 221157 |
| 1975 | 299 | 5191 | 437 | 15930 | 55966 | 1080 | 28677 | 65015 | 3163 | 175758 |
| 1976 | 536 | 4459 | 348 | 16660 | 49492 | 986 | 16940 | 42485 | 5358 | 137264 |
| 1977 | 213 | 1510 | 144 | 4798 | 40118 | - | 10878 | 52210 | 287 | 110158 |
| 1978 | 466 | 1411 | 369 | 1521 | 39955 | 1 | 5766 | 45895 | 38 | 95422 |
| 1979 | 343 | 1198 | 10 | 1948 | 66849 | 2 | 6454 | 26365 | 454 | 103623 |
| 1980 | 497 | 226 | 15 | 1365 | 66501 | - | 2948 | 20706 | 246 | 92504 |
| 1981 | 381 | 414 | 22 | 2402 | 63435 | Spain | 1682 | 13400 | - | 81736 |
| 1982 | 496 | 53 | - | 1258 | 43702 | - | 827 | 2900 | - | 49236 |
| 1983 | 428 | - | 1 | 729 | 22364 | 139 | 259 | 680 | - | 24600 |
| 1984 | 297 | 15 | 4 | 400 | 18813 | 37 | 276 | 1103 | - | 20945 |
| 1985 | 424 | 21 | 20 | 395 | 21272 | 77 | 153 | 22690 | - | 45052 |


| Year | Faroe Islands | France |  | Germany | Norway ${ }^{4}$ | Poland | UK | Russia ${ }^{2}$ | Others | Total ${ }^{3}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 893 | 12 | 75 | 1079 | 52313 | 22 | 431 | 45738 | - | 100563 |
| 1987 | 464 | 7 | 83 | 3105 | 72419 | 59 | 563 | 78211 | 5 | 154916 |
| 1988 | 1113 | 116 | 78 | 1323 | 60823 | 72 | 435 | 31293 | 2 | 95255 |
| 1989 | 1217 | - | 26 | 171 | 36451 | 1 | 590 | 20062 | - | 58518 |
| 1990 | 705 | - | 5 | 167 | 20621 | - | 494 | 5190 | - | 27182 |
| 1991 | 1117 | - | Greenland | 213 | 22178 | - | 514 | 12177 | 17 | 36216 |
| 1992 | 1093 | 151 | 1719 | 387 | 36238 | 38 | 596 | 19699 | 1 | 59922 |
| 1993 | 546 | 1215 | 880 | 1165 | 40978 | 76 | 1802 | 35071 | 646 | 82379 |
| 1994 | 2761 | 678 | 770 | 2412 | 71171 | 22 | 4673 | 51822 | 877 | 135186 |
| 1995 | 2833 | 598 | 1097 | 2675 | 76886 | 14 | 3111 | 54516 | 718 | 142448 |
| 1996 | 3743 | 6 | 1510 | 942 | 94527 | 669 | 2275 | 74239 | 217 | 178128 |
| 1997 | 3327 | 540 | 1877 | 972 | 103407 | 364 | 2340 | 41228 | 304 | 154359 |
| 1998 | 1903 | 241 | 854 | 385 | 75108 | 257 | 1229 | 20559 | 94 | 100630 |
| 1999 | 1913 | 64 | 437 | 641 | 48182 | 652 | 694 | 30520 | 92 | 83195 |
| 2000 | 631 | 178 | 432 | 880 | 42009 | 502 | 747 | 22738 | 827 | 68944 |
| 2001 | 1210 | 324 | 553 | 554 | 49067 | 1497 | 1068 | 34307 | 1060 | 89640 |
| 2002 | 1564 | 297 | 858 | 627 | 52247 | 1505 | 1125 | 37157 | 682 | 114798 |
| 2003 | 1959 | 382 | 1363 | 918 | 56485 | 1330 | 1018 | 41142 | 1103 | 138926 |
| 2004 | 2484 | 103 | 1680 | 823 | 62192 | 54 | 1250 | 54347 | 1569 | 158279 |
| 2005 | 2138 | 333 | 15 | 996 | 60850 | 963 | 1899 | 50012 | 1262 | 158298 |
| 2006 | 2390 | 883 | 1830 | 989 | 69272 | 703 | 1164 | 53313 | 1162 | 153157 |
| 2007 | 2307 | 277 | 1464 | 1123 | 71244 | 125 | 1351 | 66569 | 2511 | 161525 |
| 2008 | 2687 | 311 | 1659 | 535 | 72779 | 283 | 971 | 68792 | 1759 | 155604 |
| 2009 | 2820 | 529 | 1410 | 1957 | 104354 | 317 | 1315 | 85514 | 1845 | 200061 |
| 2010 | 3173 | 764 | 1970 | 3539 | 123384 | 379 | 1758 | 111372 | 2862 | 249201 |
| 2011 | 1759 | 268 | 2110 | 1724 | 158202 | 502 | 1379 | 139912 | 4763 | 310619 |
| 2012 | 2055 | 322 | 3984 | 1111 | 159602 | 441 | 833 | 143886 | 3393 | 315627 |


| Year | Faroe Islands | France | GDR (- <br> 1990) <br> and <br> Green- <br> land <br> (1992-) | Germany | Norway ${ }^{4}$ | Poland | UK | Russia ${ }^{2}$ | Others | Total ${ }^{3}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2013 | 1886 | 342 | 1795 | 500 | 99215 | 439 | 639 | 85668 | 3260 | 193744 |
| 2014 | 1470 | 198 | 1150 | 340 | 91306 | 187 | 355 | 78725 | 3791 | 177522 |
| 2015 | 2459 | 145 | 1047 | 124 | 95094 | 246 | 450 | 91864 | 3327 | 194756 |
| 2016 | 2460 | 340 | 1401 | 170 | 108718 | 200 | 575 | 115710 | 3838 | 233412 |
| 2017 | 2776 | 108 | 1810 | 170 | 113132 | 228 | 372 | 106714 | 2279 | 227588 |
| 2018 | 2333 | 183 | 1317 | 385 | 93839 | 169 | 453 | 90486 | 2111 | 191276 |
| 2019 | 1515 | 143 | 1208 | 204 | 93860 | 280 | 456 | 76125 | 1611 | 175402 |
| 2020 | 1392 | 96 | 910 | 282 | 88108 | 45 | 320 | 89030 | 2286 | 182468 |
| 2021 ${ }^{1)}$ | 1722 | 102 | 1101 | 365 | 100673 | 13 | 78 | 98282 | 705 | 203041 |

1) Provisional figures.
2) USSR prior to 1991.
3) Figures based on Norwegian IUU estimates in 2002-2008 (see table 4.1)
4) Included landings in Norwegian statistical areas 06 and 07 (from 1983)

Table 4.4. Northeast Arctic haddock. Catch numbers-at-age (numbers, '000).

| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1950 | 0 | 4446 | 3189 | 37949 | 35344 | 18849 | 28868 | 9199 | 1979 | 1093 | 853 | 867 | 1257 |
| 1951 | 4069 | 222 | 65643 | 9178 | 18014 | 13551 | 6808 | 6850 | 3322 | 1182 | 734 | 178 | 436 |
| 1952 | 0 | 13674 | 6012 | 151996 | 13634 | 9850 | 4693 | 3237 | 2434 | 606 | 534 | 185 | 161 |
| 1953 | 392 | 8031 | 64528 | 13013 | 70781 | 5431 | 2867 | 1080 | 424 | 315 | 393 | 202 | 410 |
| 1954 | 1726 | 493 | 6563 | 154696 | 5885 | 27590 | 3233 | 1302 | 712 | 319 | 126 | 68 | 349 |
| 1955 | 0 | 989 | 1154 | 10689 | 176678 | 4993 | 28273 | 1445 | 271 | 100 | 50 | 30 | 20 |
| 1956 | 97 | 3012 | 16437 | 5922 | 14713 | 127879 | 3182 | 8003 | 450 | 200 | 80 | 60 | 45 |
| 1957 | 828 | 243 | 2074 | 24704 | 7942 | 12535 | 46619 | 1087 | 1971 | 356 | 17 | 40 | 119 |
| 1958 | 153 | 2312 | 1727 | 5914 | 31438 | 5820 | 12748 | 17565 | 822 | 1072 | 226 | 79 | 296 |
| 1959 | 169 | 2425 | 20318 | 7826 | 7243 | 14040 | 3154 | 2237 | 5918 | 285 | 316 | 71 | 113 |
| 1960 | 2319 | 3613 | 39910 | 70912 | 13647 | 7101 | 6236 | 1579 | 2340 | 2005 | 497 | 70 | 42 |
| 1961 | 362 | 5531 | 15429 | 56855 | 63351 | 8706 | 3578 | 4407 | 788 | 527 | 1287 | 67 | 80 |
| 1962 | 0 | 4524 | 39503 | 30868 | 48903 | 33836 | 3201 | 1341 | 1773 | 242 | 247 | 483 | 28 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1963 | 3 | 2143 | 28466 | 72736 | 18969 | 13579 | 9257 | 1239 | 559 | 409 | 80 | 84 | 212 |
| 1964 | 149 | 834 | 22363 | 49290 | 30672 | 5815 | 3527 | 2716 | 833 | 104 | 206 | 235 | 190 |
| 1965 | 0 | 3498 | 5936 | 46356 | 40201 | 12631 | 1679 | 974 | 897 | 123 | 204 | 123 | 471 |
| 1966 | 0 | 2577 | 26345 | 22631 | 63176 | 29048 | 5752 | 582 | 438 | 189 | 186 | 25 | 30 |
| 1967 | 0 | 53 | 15907 | 41346 | 13496 | 25719 | 8872 | 1616 | 218 | 175 | 155 | 75 | 41 |
| 1968 | 0 | 33 | 657 | 67632 | 41267 | 7748 | 15599 | 5292 | 655 | 182 | 101 | 115 | 70 |
| 1969 | 0 | 1061 | 1524 | 1968 | 44634 | 19002 | 3620 | 4937 | 1628 | 316 | 43 | 43 | 23 |
| 1970 | 480 | 281 | 23444 | 2454 | 1906 | 22417 | 8100 | 2012 | 2016 | 740 | 166 | 26 | 96 |
| 1971 | 15 | 3535 | 1978 | 24358 | 1257 | 918 | 9279 | 3056 | 826 | 1043 | 369 | 130 | 35 |
| 1972 | 133 | 9399 | 230942 | 22315 | 42981 | 3206 | 1611 | 6758 | 2638 | 900 | 989 | 538 | 120 |
| 1973 | 0 | 5956 | 70679 | 260520 | 24180 | 6919 | 422 | 426 | 1692 | 529 | 147 | 339 | 95 |
| 1974 | 281 | 3713 | 9685 | 41706 | 88120 | 5829 | 4138 | 382 | 618 | 2043 | 935 | 276 | 659 |
| 1975 | 1321 | 4355 | 10037 | 14088 | 33871 | 49711 | 2135 | 1236 | 92 | 131 | 500 | 147 | 287 |
| 1976 | 3475 | 7499 | 13994 | 13454 | 6810 | 20796 | 40057 | 1247 | 1350 | 193 | 280 | 652 | 671 |
| 1977 | 184 | 18456 | 55967 | 22043 | 7368 | 2586 | 7781 | 11043 | 311 | 388 | 96 | 101 | 182 |
| 1978 | 46 | 2033 | 47311 | 18812 | 4076 | 1389 | 1626 | 2596 | 6215 | 162 | 258 | 3 | 139 |
| 1979 | 0 | 48 | 17540 | 35290 | 10645 | 1429 | 812 | 546 | 1466 | 2310 | 181 | 87 | 55 |
| 1980 | 0 | 0 | 627 | 22878 | 21794 | 2971 | 250 | 504 | 230 | 842 | 1299 | 111 | 50 |
| 1981 | 1 | 68 | 486 | 2561 | 22124 | 10685 | 1034 | 162 | 162 | 72 | 330 | 564 | 69 |
| 1982 | 2 | 29 | 883 | 900 | 3372 | 12203 | 2625 | 344 | 75 | 80 | 91 | 321 | 238 |
| 1983 | 3 | 351 | 1173 | 2636 | 1360 | 2394 | 2506 | 1799 | 267 | 37 | 60 | 100 | 132 |
| 1984 | 7 | 754 | 1271 | 1019 | 1899 | 657 | 950 | 2619 | 352 | 87 | 2 | 22 | 53 |
| 1985 | 4 | 2952 | 29624 | 1695 | 564 | 1009 | 943 | 886 | 1763 | 588 | 124 | 64 | 93 |
| 1986 | 506 | 650 | 23113 | 68429 | 1565 | 783 | 896 | 393 | 702 | 1144 | 443 | 130 | 414 |
| 1987 | 9 | 83 | 5031 | 87170 | 64556 | 960 | 597 | 376 | 212 | 230 | 419 | 245 | 73 |
| 1988 | 7 | 139 | 1439 | 12478 | 47890 | 20429 | 397 | 178 | 74 | 88 | 168 | 198 | 80 |
| 1989 | 611 | 221 | 2157 | 4986 | 16071 | 25313 | 3198 | 147 | 1 | 28 | 28 | 53 | 96 |
| 1990 | 2 | 446 | 1015 | 2580 | 2142 | 4046 | 6221 | 840 | 134 | 42 | 14 | 13 | 44 |
| 1991 | 23 | 533 | 4421 | 3564 | 2416 | 3299 | 4633 | 3953 | 461 | 83 | 9 | 18 | 27 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 49 | 2793 | 11571 | 11567 | 4099 | 2642 | 2894 | 3327 | 3498 | 486 | 35 | 32 | 18 |
| 1993 | 498 | 272 | 13487 | 19457 | 13704 | 4103 | 1747 | 1886 | 2105 | 1965 | 201 | 96 | 25 |
| 1994 | 95 | 187 | 3374 | 47821 | 36333 | 13264 | 2057 | 903 | 1453 | 2769 | 1802 | 259 | 49 |
| 1995 | 2 | 85 | 2003 | 16109 | 72644 | 19145 | 6417 | 746 | 361 | 770 | 655 | 804 | 116 |
| 1996 | 35 | 478 | 1662 | 6818 | 36473 | 73579 | 13426 | 2944 | 573 | 365 | 533 | 598 | 767 |
| 1997 | 70 | 94 | 2280 | 5633 | 12603 | 32832 | 49478 | 5636 | 778 | 245 | 126 | 158 | 463 |
| 1998 | 547 | 1476 | 1701 | 11304 | 9258 | 8633 | 13801 | 19469 | 2113 | 330 | 59 | 54 | 377 |
| 1999 | 104 | 568 | 16839 | 8039 | 15365 | 6073 | 4466 | 6355 | 6204 | 647 | 117 | 109 | 220 |
| 2000 | 46 | 692 | 1520 | 29986 | 6496 | 5149 | 2406 | 1657 | 1570 | 1744 | 183 | 70 | 184 |
| 2001 | 374 | 1758 | 12971 | 5230 | 32049 | 5279 | 2941 | 1137 | 1161 | 1169 | 747 | 169 | 288 |
| 2002 | 59 | 603 | 7132 | 46335 | 11084 | 21985 | 2602 | 1602 | 482 | 448 | 581 | 349 | 98 |
| 2003 | 123 | 611 | 6803 | 31448 | 56480 | 11736 | 14541 | 1637 | 2178 | 858 | 411 | 413 | 395 |
| 2004 | 58 | 1295 | 7993 | 21116 | 41310 | 41226 | 4939 | 4914 | 598 | 1252 | 296 | 139 | 465 |
| 2005 | 102 | 865 | 11452 | 19369 | 22887 | 37067 | 24461 | 2393 | 2997 | 990 | 201 | 263 | 1059 |
| 2006 | 271 | 2496 | 4539 | 35040 | 27571 | 15033 | 16023 | 8567 | 1259 | 1298 | 222 | 175 | 321 |
| 2007 | 575 | 3914 | 30707 | 15213 | 45992 | 18516 | 10642 | 7889 | 2570 | 678 | 605 | 197 | 185 |
| 2008 | 440 | 2089 | 14536 | 44192 | 15926 | 31173 | 9145 | 4520 | 2846 | 1181 | 274 | 214 | 166 |
| 2009 | 483 | 1364 | 15379 | 55013 | 52498 | 13679 | 15382 | 3800 | 1669 | 887 | 285 | 353 | 321 |
| 2010 | 457 | 620 | 6545 | 52006 | 80622 | 50306 | 9273 | 5324 | 1954 | 1114 | 533 | 242 | 621 |
| 2011 | 909 | 806 | 1277 | 8501 | 90394 | 100522 | 39496 | 4397 | 2340 | 668 | 437 | 269 | 708 |
| 2012 | 268 | 611 | 7814 | 4206 | 18007 | 93055 | 82721 | 14445 | 1325 | 448 | 217 | 216 | 568 |
| 2013 | 402 | 904 | 1778 | 12780 | 3805 | 12297 | 58024 | 29930 | 4976 | 957 | 331 | 212 | 535 |
| 2014 | 528 | 649 | 6948 | 4503 | 14563 | 6833 | 16304 | 39620 | 16439 | 2431 | 619 | 440 | 545 |
| 2015 | 303 | 1334 | 1645 | 27317 | 8526 | 16624 | 7950 | 20538 | 25534 | 6677 | 1556 | 295 | 312 |
| 2016 | 294 | 655 | 5774 | 3482 | 33177 | 9563 | 18045 | 12030 | 21875 | 13492 | 4757 | 876 | 248 |
| 2017 | 724 | 1898 | 30744 | 46463 | 16895 | 48927 | 10518 | 14992 | 9485 | 8447 | 6640 | 1872 | 317 |
| 2018 | 679 | 1438 | 9424 | 16291 | 34060 | 8466 | 18882 | 5123 | 8902 | 4125 | 3564 | 4504 | 1040 |
| 2019 | 797 | 968 | 13908 | 28572 | 24171 | 32555 | 6278 | 6803 | 2601 | 3618 | 1225 | 1715 | 1400 |
| 2020 | 122 | 1298 | 10797 | 62206 | 46715 | 18137 | 10773 | 3051 | 2839 | 1445 | 996 | 915 | 1092 |

Table 4.5. Northeast Arctic haddock. Catch weights-at-age (kg).

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1951 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1952 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1953 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1954 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1955 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1956 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1957 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1958 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1959 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1960 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1961 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1962 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1963 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1964 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1965 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1966 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1967 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1968 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1969 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1970 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1971 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1972 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1973 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1974 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1975 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1976 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1977 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1978 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1980 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1981 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1982 | 0.299 | 0.519 | 0.75 | 1.038 | 1.321 | 1.617 | 1.873 | 2.147 | 2.418 | 2.698 | 2.931 | 3.094 | 3.461 |
| 1983 | 0.188 | 0.689 | 1.033 | 1.408 | 1.71 | 2.149 | 2.469 | 2.748 | 3.069 | 3.687 | 4.516 | 3.094 | 3.461 |
| 1984 | 0.408 | 0.805 | 1.218 | 1.632 | 2.038 | 2.852 | 2.845 | 3.218 | 3.605 | 4.065 | 4.407 | 4.734 | 5.099 |
| 1985 | 0.319 | 0.383 | 0.835 | 1.29 | 1.816 | 2.174 | 2.301 | 2.835 | 3.253 | 3.721 | 4.084 | 4.137 | 4.926 |
| 1986 | 0.218 | 0.325 | 0.612 | 1.064 | 1.539 | 1.944 | 2.362 | 2.794 | 3.25 | 3.643 | 4.14 | 4.559 | 5.927 |
| 1987 | 0.143 | 0.221 | 0.497 | 0.765 | 1.179 | 1.724 | 2.135 | 2.551 | 3.009 | 3.414 | 3.84 | 4.415 | 5.195 |
| 1988 | 0.279 | 0.551 | 0.55 | 0.908 | 1.097 | 1.357 | 1.537 | 1.704 | 2.403 | 2.403 | 2.486 | 2.531 | 2.834 |
| 1989 | 0.258 | 0.55 | 0.684 | 0.84 | 0.998 | 1.176 | 1.546 | 1.713 | 1.949 | 2.14 | 2.389 | 2.522 | 2.797 |
| 1990 | 0.319 | 0.601 | 0.793 | 1.172 | 1.397 | 1.624 | 1.885 | 2.112 | 2.653 | 3.102 | 3.18 | 3.438 | 3.319 |
| 1991 | 0.216 | 0.616 | 0.941 | 1.281 | 1.556 | 1.797 | 2.044 | 2.079 | 2.311 | 2.788 | 3.408 | 2.896 | 3.274 |
| 1992 | 0.055 | 0.458 | 0.906 | 1.263 | 1.535 | 1.747 | 2.043 | 2.2 | 2.298 | 2.494 | 2.49 | 2.673 | 2.923 |
| 1993 | 0.381 | 0.64 | 0.94 | 1.204 | 1.487 | 1.748 | 1.994 | 2.237 | 2.417 | 2.654 | 2.906 | 3.184 | 3.363 |
| 1994 | 0.278 | 0.521 | 0.614 | 0.906 | 1.287 | 1.602 | 1.968 | 2.059 | 2.39 | 2.545 | 2.881 | 2.918 | 3.222 |
| 1995 | 0.258 | 0.446 | 0.739 | 0.808 | 1.107 | 1.556 | 1.838 | 2.234 | 2.416 | 2.602 | 2.965 | 3.163 | 3.786 |
| 1996 | 0.287 | 0.427 | 0.683 | 0.868 | 1.045 | 1.363 | 1.71 | 1.886 | 2.214 | 2.37 | 2.438 | 2.707 | 2.896 |
| 1997 | 0.408 | 0.575 | 0.682 | 1.028 | 1.151 | 1.369 | 1.637 | 1.856 | 2.073 | 2.5 | 2.279 | 2.532 | 2.609 |
| 1998 | 0.409 | 0.593 | 0.748 | 0.974 | 1.262 | 1.433 | 1.641 | 1.863 | 2.069 | 2.335 | 2.511 | 2.8 | 2.849 |
| 1999 | 0.435 | 0.695 | 0.826 | 1.079 | 1.261 | 1.485 | 1.634 | 1.798 | 2.032 | 2.237 | 2.339 | 2.611 | 2.865 |
| 2000 | 0.378 | 0.577 | 0.853 | 1.186 | 1.395 | 1.588 | 1.808 | 1.989 | 2.264 | 2.415 | 2.587 | 2.647 | 3.098 |
| 2001 | 0.391 | 0.647 | 0.751 | 1.104 | 1.459 | 1.709 | 1.921 | 2.182 | 2.331 | 2.609 | 2.757 | 3.376 | 3.338 |
| 2002 | 0.159 | 0.407 | 0.687 | 1.001 | 1.363 | 1.643 | 1.975 | 2.086 | 2.294 | 2.487 | 2.612 | 2.847 | 3.501 |
| 2003 | 0.198 | 0.384 | 0.594 | 0.875 | 1.113 | 1.364 | 1.361 | 1.972 | 1.636 | 1.877 | 2.088 | 2.351 | 2.842 |
| 2004 | 0.328 | 0.429 | 0.636 | 0.886 | 1.183 | 1.508 | 1.821 | 2.075 | 2.339 | 2.58 | 2.527 | 3.153 | 3.197 |
| 2005 | 0.285 | 0.492 | 0.722 | 0.906 | 1.121 | 1.343 | 1.619 | 2.036 | 2.177 | 2.382 | 2.527 | 2.496 | 2.81 |
| 2006 | 0.311 | 0.567 | 0.745 | 1.041 | 1.287 | 1.504 | 1.72 | 2.082 | 2.377 | 2.738 | 3.082 | 3.02 | 3.43 |
| 2007 | 0.329 | 0.431 | 0.652 | 0.899 | 1.197 | 1.435 | 1.722 | 1.99 | 2.309 | 2.715 | 2.987 | 2.947 | 3.591 |


| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2008 | 0.383 | 0.484 | 0.658 | 0.901 | 1.242 | 1.515 | 1.781 | 2.18 | 2.33 | 2.664 | 3.019 | 3.326 | 3.829 |
| 2009 | 0.378 | 0.508 | 0.707 | 1.024 | 1.28 | 1.538 | 1.806 | 2.107 | 2.398 | 2.531 | 2.606 | 3.089 | 3.541 |
| 2010 | 0.317 | 0.499 | 0.642 | 0.887 | 1.137 | 1.396 | 1.702 | 1.907 | 2.095 | 2.404 | 2.534 | 3.064 | 3.249 |
| 2011 | 0.423 | 0.513 | 0.811 | 0.953 | 1.093 | 1.254 | 1.462 | 1.715 | 1.978 | 2.328 | 2.305 | 2.55 | 2.76 |
| 2012 | 0.271 | 0.506 | 0.756 | 1.004 | 1.174 | 1.371 | 1.514 | 1.715 | 2.051 | 2.444 | 2.414 | 2.615 | 2.932 |
| 2013 | 0.469 | 0.542 | 0.821 | 1.014 | 1.217 | 1.401 | 1.571 | 1.714 | 1.914 | 2.168 | 2.24 | 2.516 | 2.807 |
| 2014 | 0.469 | 0.645 | 0.792 | 1.033 | 1.253 | 1.417 | 1.625 | 1.793 | 1.941 | 2.081 | 2.479 | 2.703 | 3.011 |
| 2015 | 0.473 | 0.647 | 0.876 | 1.054 | 1.327 | 1.571 | 1.777 | 1.934 | 2.025 | 2.216 | 2.481 | 2.99 | 3.455 |
| 2016 | 0.497 | 0.743 | 0.882 | 1.115 | 1.369 | 1.662 | 1.917 | 2.089 | 2.301 | 2.567 | 3.076 | 3.286 | 3.331 |
| 2017 | 0.449 | 0.608 | 0.874 | 1.088 | 1.378 | 1.666 | 1.879 | 2.146 | 2.258 | 2.476 | 2.72 | 2.98 | 3.713 |
| 2018 | 0.443 | 0.663 | 0.820 | 1.051 | 1.339 | 1.629 | 1.927 | 2.156 | 2.372 | 2.588 | 2.728 | 2.773 | 3.175 |
| 2019 | 0.341 | 0.508 | 0.729 | 0.955 | 1.275 | 1.581 | 1.834 | 2.151 | 2.378 | 2.607 | 2.868 | 2.934 | 3.382 |
| 2020 | 0.364 | 0.523 | 0.629 | 0.788 | 1.131 | 1.489 | 1.821 | 2.126 | 2.426 | 2.651 | 2.771 | 3.147 | 3.359 |

Table 4.6. Northeast Arctic haddock. Stock weights-at-age (kg). The data from 1950-1993 is unchanged AFWG 2019, the data from 1994 and onward have been updated this year. The ages 3-13 are adjusted to account for the lack of the Russian survey as described in the stock annex, age 1-2 are unadjusted smoothed estimates based on winter survey data.

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1951 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1952 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1953 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1954 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1955 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1956 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1957 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1958 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1959 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1960 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1961 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1962 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1963 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1964 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1965 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1966 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1967 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1968 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1969 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1970 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1971 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1972 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1973 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1974 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1975 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1976 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1977 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1978 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1979 | 0.031 | 0.145 | 0.354 | 0.653 | 1.016 | 1.427 | 1.867 | 2.327 | 2.771 | 3.195 | 3.597 | 3.597 | 3.597 |
| 1980 | 0.063 | 0.262 | 0.454 | 0.878 | 1.159 | 1.675 | 2.292 | 3.134 | 3.31 | 3.553 | 3.792 | 3.792 | 3.792 |
| 1981 | 0.051 | 0.274 | 0.603 | 0.805 | 1.315 | 1.582 | 2.118 | 2.728 | 3.51 | 3.679 | 3.904 | 3.904 | 3.904 |
| 1982 | 0.036 | 0.224 | 0.631 | 1.049 | 1.217 | 1.782 | 2.017 | 2.553 | 3.14 | 3.853 | 4.016 | 4.016 | 4.016 |
| 1983 | 0.035 | 0.164 | 0.524 | 1.098 | 1.558 | 1.663 | 2.255 | 2.448 | 2.97 | 3.524 | 4.165 | 4.165 | 4.165 |
| 1984 | 0.028 | 0.158 | 0.391 | 0.926 | 1.632 | 2.093 | 2.121 | 2.718 | 2.865 | 3.363 | 3.878 | 3.878 | 3.878 |
| 1985 | 0.03 | 0.127 | 0.379 | 0.700 | 1.394 | 2.195 | 2.626 | 2.572 | 3.158 | 3.261 | 3.728 | 3.728 | 3.728 |
| 1986 | 0.035 | 0.136 | 0.311 | 0.682 | 1.069 | 1.898 | 2.761 | 3.138 | 3.005 | 3.568 | 3.632 | 3.632 | 3.632 |
| 1987 | 0.042 | 0.161 | 0.331 | 0.569 | 1.047 | 1.473 | 2.411 | 3.307 | 3.616 | 3.412 | 3.946 | 3.946 | 3.946 |
| 1988 | 0.039 | 0.189 | 0.383 | 0.603 | 0.887 | 1.452 | 1.895 | 2.915 | 3.822 | 4.054 | 3.787 | 3.787 | 3.787 |
| 1989 | 0.037 | 0.175 | 0.445 | 0.689 | 0.936 | 1.248 | 1.878 | 2.317 | 3.395 | 4.297 | 4.449 | 4.449 | 4.449 |
| 1990 | 0.031 | 0.169 | 0.413 | 0.789 | 1.054 | 1.312 | 1.635 | 2.308 | 2.728 | 3.844 | 4.73 | 4.73 | 4.73 |
| 1991 | 0.025 | 0.141 | 0.402 | 0.737 | 1.193 | 1.458 | 1.714 | 2.035 | 2.732 | 3.122 | 4.256 | 4.256 | 4.256 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.023 | 0.114 | 0.34 | 0.721 | 1.119 | 1.63 | 1.881 | 2.127 | 2.437 | 3.142 | 3.491 | 3.491 | 3.491 |
| 1993 | 0.025 | 0.107 | 0.279 | 0.616 | 1.100 | 1.537 | 2.08 | 2.308 | 2.54 | 2.831 | 3.531 | 3.531 | 3.531 |
| 1994 | 13.8 | 22.1 | 0.25 | 0.502 | 0.936 | 1.646 | 2.17 | 2.713 | 2.866 | 2.817 | 2.978 | 3.64 | 4.181 |
| 1995 | 14.9 | 22.6 | 0.261 | 0.465 | 0.795 | 1.311 | 2.113 | 2.633 | 3.166 | 3.295 | 3.228 | 3.163 | 3.955 |
| 1996 | 14.9 | 24.3 | 0.278 | 0.485 | 0.744 | 1.132 | 1.714 | 2.568 | 3.092 | 3.61 | 3.719 | 3.419 | 3.481 |
| 1997 | 15.2 | 24.3 | 0.343 | 0.512 | 0.766 | 1.06 | 1.49 | 2.122 | 3.021 | 3.546 | 4.044 | 3.887 | 3.738 |
| 1998 | 14 | 24.8 | 0.343 | 0.622 | 0.813 | 1.096 | 1.412 | 1.873 | 2.546 | 3.466 | 3.957 | 4.181 | 4.199 |
| 1999 | 14.2 | 23 | 0.363 | 0.627 | 0.97 | 1.154 | 1.447 | 1.772 | 2.263 | 2.956 | 3.888 | 4.111 | 4.49 |
| 2000 | 13.7 | 23.3 | 0.293 | 0.657 | 0.976 | 1.36 | 1.517 | 1.822 | 2.147 | 2.655 | 3.365 | 4.059 | 4.416 |
| 2001 | 13.2 | 22.5 | 0.301 | 0.538 | 1.023 | 1.36 | 1.774 | 1.905 | 2.205 | 2.539 | 3.05 | 3.56 | 4.361 |
| 2002 | 13.9 | 21.8 | 0.273 | 0.556 | 0.848 | 1.428 | 1.774 | 2.191 | 2.299 | 2.603 | 2.921 | 3.252 | 3.871 |
| 2003 | 13.9 | 22.8 | 0.248 | 0.502 | 0.873 | 1.2 | 1.844 | 2.191 | 2.61 | 2.695 | 2.993 | 3.119 | 3.56 |
| 2004 | 14.1 | 22.8 | 0.283 | 0.461 | 0.795 | 1.238 | 1.572 | 2.284 | 2.623 | 3.043 | 3.093 | 3.178 | 3.434 |
| 2005 | 12.7 | 23.1 | 0.283 | 0.528 | 0.732 | 1.132 | 1.618 | 1.968 | 2.702 | 3.043 | 3.444 | 3.282 | 3.497 |
| 2006 | 12.6 | 20.9 | 0.293 | 0.524 | 0.831 | 1.053 | 1.49 | 2.023 | 2.371 | 3.145 | 3.46 | 3.624 | 3.608 |
| 2007 | 13.2 | 20.9 | 0.219 | 0.542 | 0.831 | 1.177 | 1.395 | 1.873 | 2.432 | 2.776 | 3.555 | 3.64 | 3.938 |
| 2008 | 14 | 21.7 | 0.219 | 0.415 | 0.855 | 1.177 | 1.553 | 1.761 | 2.263 | 2.845 | 3.168 | 3.738 | 3.955 |
| 2009 | 14.1 | 22.9 | 0.248 | 0.411 | 0.664 | 1.207 | 1.544 | 1.936 | 2.135 | 2.669 | 3.242 | 3.373 | 4.041 |
| 2010 | 15.3 | 23.1 | 0.286 | 0.461 | 0.664 | 0.957 | 1.581 | 1.936 | 2.335 | 2.526 | 3.05 | 3.434 | 3.689 |
| 2011 | 14.8 | 24.9 | 0.295 | 0.528 | 0.732 | 0.951 | 1.279 | 1.979 | 2.335 | 2.749 | 2.908 | 3.252 | 3.754 |
| 2012 | 15.7 | 24.3 | 0.366 | 0.546 | 0.836 | 1.053 | 1.271 | 1.626 | 2.383 | 2.735 | 3.137 | 3.105 | 3.56 |
| 2013 | 15.1 | 25.5 | 0.339 | 0.667 | 0.861 | 1.184 | 1.395 | 1.617 | 1.981 | 2.79 | 3.137 | 3.327 | 3.419 |
| 2014 | 15.2 | 24.6 | 0.391 | 0.617 | 1.03 | 1.215 | 1.563 | 1.761 | 1.97 | 2.352 | 3.183 | 3.327 | 3.64 |
| 2015 | 14.9 | 24.8 | 0.353 | 0.704 | 0.962 | 1.437 | 1.59 | 1.946 | 2.135 | 2.34 | 2.728 | 3.373 | 3.64 |
| 2016 | 14.2 | 24.3 | 0.363 | 0.642 | 1.087 | 1.351 | 1.865 | 1.99 | 2.346 | 2.513 | 2.715 | 2.921 | 3.689 |
| 2017 | 13.8 | 23.2 | 0.343 | 0.662 | 0.996 | 1.516 | 1.763 | 2.296 | 2.395 | 2.749 | 2.908 | 2.907 | 3.237 |
| 2018 | 13.6 | 22.7 | 0.298 | 0.622 | 1.023 | 1.394 | 1.948 | 2.179 | 2.729 | 2.803 | 3.153 | 3.105 | 3.222 |
| 2019 | 13.4 | 22.3 | 0.278 | 0.55 | 0.97 | 1.428 | 1.804 | 2.393 | 2.597 | 3.159 | 3.197 | 3.342 | 3.419 |
| 2020 | NA | 22.1 | 0.266 | 0.516 | 0.866 | 1.36 | 1.854 | 2.238 | 2.838 | 3.028 | 3.572 | 3.388 | 3.656 |


| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2021 | NA | NA | 0.259 | 0.494 | 0.813 | 1.222 | 1.774 | 2.284 | 2.663 | 3.279 | 3.444 | 3.754 | 3.705 |

Table 4.7. Northeast Arctic haddock. Proportion mature-at-age. The data from 1950-1993 is unchanged since AFWG 2019, the data from 1994 and onward have been updated this year, ages 11-13+ is set to 1 (not shown)

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1951 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1952 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1953 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1954 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1955 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1956 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1957 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1958 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1959 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1960 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1961 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1962 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1963 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1964 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1965 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1966 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1967 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1968 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1969 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1970 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1971 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1972 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1973 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1974 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1975 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1976 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1977 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1978 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1979 | 0 | 0 | 0.027 | 0.101 | 0.311 | 0.622 | 0.845 | 0.944 | 0.982 | 0.994 |
| 1980 | 0 | 0 | 0.026 | 0.076 | 0.243 | 0.649 | 0.86 | 0.95 | 0.984 | 0.995 |
| 1981 | 0 | 0 | 0.056 | 0.104 | 0.303 | 0.549 | 0.857 | 0.948 | 0.984 | 0.995 |
| 1982 | 0 | 0 | 0.053 | 0.161 | 0.332 | 0.577 | 0.77 | 0.947 | 0.983 | 0.995 |
| 1983 | 0 | 0 | 0.057 | 0.183 | 0.472 | 0.665 | 0.8 | 0.906 | 0.983 | 0.995 |
| 1984 | 0 | 0 | 0.044 | 0.196 | 0.51 | 0.801 | 0.862 | 0.921 | 0.967 | 0.995 |
| 1985 | 0 | 0 | 0.027 | 0.149 | 0.522 | 0.796 | 0.928 | 0.953 | 0.973 | 0.989 |
| 1986 | 0 | 0 | 0.021 | 0.103 | 0.454 | 0.758 | 0.928 | 0.977 | 0.984 | 0.991 |
| 1987 | 0 | 0 | 0.021 | 0.076 | 0.294 | 0.713 | 0.918 | 0.976 | 0.993 | 0.994 |
| 1988 | 0 | 0 | 0.025 | 0.074 | 0.24 | 0.576 | 0.898 | 0.975 | 0.993 | 0.998 |
| 1989 | 0 | 0 | 0.032 | 0.09 | 0.25 | 0.534 | 0.822 | 0.966 | 0.993 | 0.998 |
| 1990 | 0 | 0 | 0.046 | 0.127 | 0.305 | 0.578 | 0.798 | 0.937 | 0.99 | 0.997 |
| 1991 | 0 | 0 | 0.041 | 0.164 | 0.358 | 0.623 | 0.82 | 0.925 | 0.98 | 0.997 |
| 1992 | 0 | 0 | 0.03 | 0.147 | 0.449 | 0.704 | 0.855 | 0.936 | 0.976 | 0.994 |
| 1993 | 0 | 0 | 0.018 | 0.113 | 0.396 | 0.741 | 0.878 | 0.95 | 0.979 | 0.992 |
| 1994 | 0 | 0 | 0.028 | 0.083 | 0.263 | 0.627 | 0.838 | 0.941 | 0.958 | 0.957 |
| 1995 | 0 | 0 | 0.029 | 0.074 | 0.204 | 0.49 | 0.825 | 0.932 | 0.975 | 0.98 |
| 1996 | 0 | 0 | 0.031 | 0.079 | 0.184 | 0.408 | 0.716 | 0.925 | 0.972 | 0.99 |
| 1997 | 0 | 0 | 0.042 | 0.086 | 0.192 | 0.373 | 0.634 | 0.858 | 0.968 | 0.988 |
| 1998 | 0 | 0 | 0.042 | 0.117 | 0.211 | 0.391 | 0.602 | 0.803 | 0.931 | 0.986 |
| 1999 | 0 | 0 | 0.046 | 0.119 | 0.277 | 0.418 | 0.616 | 0.776 | 0.898 | 0.964 |
| 2000 | 0 | 0 | 0.033 | 0.128 | 0.279 | 0.512 | 0.645 | 0.789 | 0.88 | 0.946 |
| 2001 | 0 | 0 | 0.035 | 0.092 | 0.3 | 0.512 | 0.735 | 0.81 | 0.889 | 0.937 |
| 2002 | 0 | 0 | 0.03 | 0.097 | 0.225 | 0.542 | 0.735 | 0.871 | 0.902 | 0.942 |
| 2003 | 0 | 0 | 0.027 | 0.083 | 0.235 | 0.44 | 0.757 | 0.871 | 0.937 | 0.949 |
| 2004 | 0 | 0 | 0.032 | 0.073 | 0.204 | 0.457 | 0.666 | 0.886 | 0.938 | 0.969 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 0 | 0 | 0.032 | 0.09 | 0.179 | 0.408 | 0.683 | 0.826 | 0.945 | 0.969 |
| 2006 | 0 | 0 | 0.033 | 0.089 | 0.218 | 0.37 | 0.634 | 0.837 | 0.911 | 0.973 |
| 2007 | 0 | 0 | 0.023 | 0.094 | 0.218 | 0.429 | 0.594 | 0.803 | 0.919 | 0.954 |
| 2008 | 0 | 0 | 0.023 | 0.063 | 0.228 | 0.429 | 0.659 | 0.772 | 0.898 | 0.958 |
| 2009 | 0 | 0 | 0.027 | 0.062 | 0.154 | 0.443 | 0.655 | 0.818 | 0.878 | 0.947 |
| 2010 | 0 | 0 | 0.032 | 0.073 | 0.154 | 0.325 | 0.67 | 0.818 | 0.907 | 0.936 |
| 2011 | 0 | 0 | 0.035 | 0.09 | 0.179 | 0.322 | 0.543 | 0.828 | 0.907 | 0.952 |
| 2012 | 0 | 0 | 0.046 | 0.095 | 0.22 | 0.37 | 0.54 | 0.731 | 0.913 | 0.951 |
| 2013 | 0 | 0 | 0.041 | 0.131 | 0.23 | 0.433 | 0.594 | 0.728 | 0.851 | 0.955 |
| 2014 | 0 | 0 | 0.051 | 0.116 | 0.303 | 0.447 | 0.662 | 0.772 | 0.848 | 0.918 |
| 2015 | 0 | 0 | 0.043 | 0.142 | 0.274 | 0.545 | 0.673 | 0.82 | 0.878 | 0.917 |
| 2016 | 0 | 0 | 0.046 | 0.123 | 0.327 | 0.509 | 0.762 | 0.831 | 0.908 | 0.935 |
| 2017 | 0 | 0 | 0.042 | 0.129 | 0.288 | 0.578 | 0.732 | 0.888 | 0.914 | 0.952 |
| 2018 | 0 | 0 | 0.035 | 0.117 | 0.3 | 0.527 | 0.785 | 0.868 | 0.947 | 0.956 |
| 2019 | 0 | 0 | 0.031 | 0.096 | 0.277 | 0.542 | 0.744 | 0.903 | 0.936 | 0.974 |
| 2020 | 0 | 0 | 0.03 | 0.087 | 0.233 | 0.512 | 0.76 | 0.879 | 0.956 | 0.968 |
| 2021 |  |  | 0.029 | 0.081 | 0.211 | 0.45 | 0.735 | 0.886 | 0.942 | 0.979 |

Table 4.8. Northeast Arctic haddock. Consumption of Haddock by NEA Cod (min. spec) age 0-6, and total biomass ages 0-6 consumed.

| Age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | Biomass |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1984 | 1975.1 | 990.1 | 15.3 | 0.1 | 0.0 | 0.0 | 0.0 | 51.7 |
| 1985 | 2027.1 | 1378.0 | 5.1 | 0.0 | 0.0 | 0.0 | 0.0 | 53.5 |
| 1986 | 92.8 | 624.2 | 224.5 | 168.5 | 0.0 | 0.0 | 0.0 | 109.8 |
| 1987 | 0.0 | 1058.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 5.8 |
| 1988 | 21.3 | 221.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 9.9 |
| 1989 | 47.9 | 135.9 | 33.9 | 3.3 | 0.0 | 0.0 | 0.0 | 13.9 |
| 1990 | 0.0 | 352.4 | 12.9 | 0.0 | 0.0 | 0.0 | 0.0 | 15.5 |
| 1991 | 132.1 | 1737.1 | 123.0 | 0.9 | 0.0 | 0.0 | 0.0 | 87.7 |
| 1992 | 824.9 | 1441.6 | 143.6 | 32.2 | 3.1 | 2.6 | 0.0 | 69.3 |
| 1993 |  |  | 0.0 | 0.0 | 0.0 | 0.0 |  |  |


| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | Biomass |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 1348.5 | 1483.4 | 73.6 | 23.9 | 6.9 | 0.8 | 0.0 | 48.4 |
| 1995 | 181.8 | 2868.8 | 167.3 | 12.4 | 28.2 | 27.8 | 0.3 | 113.6 |
| 1996 | 359.6 | 1549.9 | 154.2 | 38.2 | 5.2 | 2.5 | 3.2 | 66.6 |
| 1997 | 0.0 | 947.0 | 38.9 | 26.4 | 1.7 | 0.8 | 0.5 | 44.0 |
| 1998 | 0.0 | 1739.4 | 27.5 | 1.7 | 2.6 | 0.4 | 0.0 | 36.0 |
| 1999 | 0.0 | 1041.9 | 25.3 | 0.4 | 0.0 | 0.0 | 0.0 | 29.6 |
| 2000 | 813.4 | 1412.0 | 71.6 | 2.2 | 1.1 | 0.2 | 0.1 | 58.3 |
| 2001 | 1047.9 | 593.6 | 53.3 | 4.7 | 0.1 | 0.0 | 0.0 | 51.2 |
| 2002 | 456.0 | 2437.4 | 240.6 | 39.5 | 2.3 | 0.4 | 0.2 | 127.0 |
| 2003 | 1140.2 | 3568.0 | 214.3 | 39.3 | 12.7 | 1.2 | 0.0 | 165.8 |
| 2004 | 5395.1 | 2862.8 | 303.7 | 39.8 | 9.9 | 2.5 | 0.0 | 198.1 |
| 2005 | 7703.0 | 6674.7 | 276.3 | 55.4 | 9.3 | 2.3 | 0.9 | 324.5 |
| 2006 | 12706.3 | 8410.2 | 375.2 | 5.5 | 4.4 | 1.2 | 0.5 | 360.5 |
| 2007 | 1204.2 | 10143.7 | 660.2 | 71.9 | 3.9 | 2.2 | 0.2 | 377.6 |
| 2008 | 1354.5 | 964.7 | 894.3 | 227.7 | 44.3 | 5.7 | 3.3 | 293.3 |
| 2009 | 5607.2 | 1854.7 | 274.1 | 262.0 | 69.0 | 22.3 | 1.5 | 252.4 |
| 2010 | 1968.7 | 5687.7 | 180.0 | 66.9 | 68.5 | 62.2 | 11.6 | 266.8 |
| 2011 | 2316.3 | 2622.4 | 451.4 | 56.1 | 75.1 | 86.7 | 19.4 | 279.0 |
| 2012 | 231.9 | 7132.1 | 134.3 | 107.3 | 15.0 | 6.7 | 4.3 | 219.5 |
| 2013 | 2172.4 | 1581.6 | 376.4 | 31.6 | 22.4 | 5.5 | 4.2 | 200.4 |
| 2014 | 1195.0 | 1991.3 | 140.6 | 27.5 | 1.8 | 0.6 | 0.0 | 87.6 |
| 2015 | 4931.7 | 2579.5 | 131.3 | 13.6 | 44.5 | 1.5 | 0.2 | 177.8 |
| 2016 | 8067.8 | 2654.8 | 276.8 | 22.6 | 2.5 | 7.7 | 1.8 | 222.0 |
| 2017 | 4421.9 | 7602.9 | 229.3 | 22.9 | 12.7 | 6.2 | 13.7 | 271.8 |
| 2018 | 2348.7 | 7041.1 | 583.6 | 65.0 | 6.9 | 0.6 | 0.0 | 276.1 |
| 2019 | 542.7 | 4542.6 | 411.3 | 119.2 | 8.1 | 0.3 | 0.0 | 211.8 |
| 2020 | 2008.8 | 450.9 | 72.5 | 63.7 | 80.4 | 4.2 | 0.1 | 91.7 |
| Av.1984-2020 | 2017.4 | 2713.4 | 199.9 | 44.9 | 14.7 | 6.9 | 1.8 | 142.5 |

Table 4.9. Northeast Arctic haddock. Survey indices for SAM tuning (see section 4.4.6). The last age is a plus group.
Northeast Arctic haddock
RU-BTr-Q4
104
2020 \#Russian trawl and acoustic survey bottom trawl index

BS-NoRU-Q1(Aco)
19942021
110.0770 .189

39

| 1 | 348.7 | 626.6 | 121.4 | 8.55 | 0.7 | 0.33 | 2.71 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 41.5 | 121.5 | 395.4 | 47.6 | 2.8 | 0.05 | 0.83 |
| 1 | 30 | 22.1 | 68.7 | 143.7 | 5.67 | 0.94 | 0.07 |
| 1 | 57.3 | 22.2 | 15.5 | 56.1 | 62.8 | 4.68 | 0.19 |


| 1 | 33.8 | 58.8 | 24.2 | 7.7 | 14.1 | 20.7 | 1.62 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 83.7 | 21.6 | 22.1 | 6.17 | 1.55 | 3.88 | 2.77 |
| 1 | 36.4 | 75.5 | 14 | 12.6 | 1.57 | 0.53 | 3.02 |
| 1 | 233.5 | 40.2 | 41.4 | 2.2 | 1.61 | 0.16 | 0.71 |
| 1 | 255.2 | 201.8 | 18.5 | 11.7 | 1.59 | 0.29 | 0.56 |
| 1 | 203.7 | 184.6 | 136 | 12.3 | 6.01 | 0.26 | 0.9 |
| 1 | 151 | 101.8 | 107.8 | 57.7 | 7.62 | 1.15 | 0.55 |
| 1 | 221.3 | 115.7 | 57.4 | 56.7 | 12.7 | 0.38 | 0.33 |
| 1 | 56.3 | 123.8 | 47.4 | 19.3 | 13.6 | 3.23 | 0.35 |
| 1 | 209.3 | 46.1 | 80.6 | 28.9 | 10 | 5.05 | 2.79 |
| 1 | 812.4 | 303 | 90 | 74.1 | 7.41 | 12.8 | 2.11 |
| 1 | 883.7 | 630 | 266.6 | 38.9 | 14.6 | 1.26 | 1.71 |
| 1 | 128.1 | 631 | 604 | 167 | 12.1 | 2.94 | 2.11 |
| 1 | 54.2 | 84.2 | 313 | 292.2 | 54.9 | 1.72 | 1.47 |
| 1 | 191.6 | 48.8 | 88.1 | 310.6 | 172.5 | 30.1 | 1.01 |
| 1 | 67.3 | 146.8 | 35.4 | 53 | 223.8 | 102.7 | 14.35 |
| 1 | 334.8 | 39.12 | 108.71 | 23.2 | 34.76 | 86.34 | 38.8 |
| 1 | 24.31 | 189.4 | 26.6 | 46.17 | 9.22 | 22.41 | 31.97 |
| 1 | 71.82 | 12.06 | 59.67 | 12.5 | 17.31 | 7.48 | 33.27 |
| 1 | 81.13 | 65.08 | 4.8 | 34.8 | 6.24 | 7.93 | 17.73 |
| 1 | 170.4 | 62.87 | 64.18 | 6.88 | 15.77 | 2.75 | 14.52 |
| 1 | 507.61 | 146.22 | 31.73 | 21.88 | 4.9 | 3.27 | 4.11 |
| 1 | 290.483 | 302.908 | 81.912 | 23.057 | 11.49 | 1.804 | 6.219 |
| 1 | 43.1 | 114.3 | 173.8 | 17.1 | 6.28 | 0.48 | 1.12 |
| BS-NoRu-Q1 (BTr) |  | \# Join | Barent | ea win | r surv | botto | trawl |

19942021
110.0770 .189

310

| 314.533 | 436.251 | 46.176 | 3.54 | 0.163 | 0.13 | 0.2 | 0.651 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 54.857 | 167.104 | 343.38 | 29.623 | 1.441 | 0.025 | 0.043 | 0.404 |
| 55.843 | 31.334 | 150.768 | 238.108 | 16.131 | 1.15 | 0 | 0.069 |
| 79.632 | 39.855 | 18.255 | 61.566 | 88.411 | 3.277 | 0.082 | 0.043 |
| 21.681 | 36.749 | 11.844 | 1.294 | 9.203 | 7.212 | 0.648 | 0.092 |
| 56.92 | 15.874 | 9.418 | 2.831 | 0.807 | 1.282 | 0.771 | 0.034 |
| 24.08 | 35.241 | 6.789 | 4.134 | 0.684 | 0.083 | 0.802 | 0.288 |
| 293.996 | 26.252 | 22.997 | 1.634 | 0.752 | 0.058 | 0.06 | 0.329 |
| 312.87 | 185.453 | 12.417 | 8.04 | 0.846 | 0.218 | 0.009 | 0.325 |
| 352.236 | 174.452 | 72.708 | 5.104 | 1.682 | 0.119 | 0.104 | 0.217 |
| 173.132 | 100.516 | 77.021 | 51.281 | 7.409 | 0.912 | 0.133 | 0.228 |
| 317.889 | 141.058 | 50.664 | 61.191 | 10.082 | 0.249 | 0.08 | 0.009 |
| 78.798 | 130.76 | 46.048 | 20.874 | 16.208 | 3.184 | 0.094 | 0.265 |
| 443.266 | 81.784 | 84.667 | 26.279 | 5.411 | 2.197 | 1.376 | 0.896 |
| 1591.031 | 583.606 | 53.079 | 54.732 | 6.794 | 10.248 | 0.23 | 0.167 |
| 1230.426 | 751.012 | 368.33 | 25.414 | 12.437 | 0.851 | 0.09 | 0.363 |
| 102.451 | 510.449 | 443.759 | 139.316 | 7.988 | 1.016 | 0.386 | 0.574 |
| 52.883 | 123.634 | 469.482 | 290.036 | 65.236 | 1.416 | 1.121 | 0.184 |


| 1 | 316.077 | 28.785 | 74.714 | 267.945 | 154.601 | 24.766 | 3.115 | 0.391 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 57.444 | 143.984 | 22.019 | 33.624 | 191.145 | 69.385 | 6.114 | 0.076 |
| 1 | 381.173 | 32.729 | 104.397 | 23.257 | 50.035 | 97.536 | 38.692 | 2.425 |
| 1 | 30.615 | 187.035 | 43.601 | 39.44 | 14.668 | 18.735 | 30.744 | 10.2 |
| 1 | 163.385 | 34.342 | 115.597 | 22.406 | 41.948 | 12.437 | 32.396 | 33.161 |
| 1 | 134.9 | 105.5 | 7.553 | 55.338 | 9.692 | 15.6 | 2.527 | 23.861 |
| 1 | 336.307 | 86.656 | 65.764 | 7.771 | 15.59 | 3.621 | 2.564 | 11.931 |
| 1 | 1075.552 | 187.224 | 49.399 | 16.996 | 4.038 | 2.948 | 0.736 | 1.91 |
| 1 | 424.225 | 586.985 | 99.123 | 22.08 | 6.057 | 2.605 | 1.042 | 2.827 |
| 1 | 118.428 | 194.033 | 302.978 | 20.677 | 4.628 | 0.848 | 0.204 | 0.93 |
| FLT007: Eco-NoRu-Q3 (Btr) |  |  | \# Joint Barents Sea ecosystem survey bottom trawl index |  |  |  |  |  |
| 20042020 |  |  |  |  |  |  |  |  |
| 110.650 .75 |  |  |  |  |  |  |  |  |
| 39 |  |  |  |  |  |  |  |  |
| 1 | 123.368 | 70.303 | 69.118 | 31.482 | 2.989 | 1.721 | 0.22 |  |
| 1 | 324.56 | 89.531 | 30.44 | 32.246 | 15.035 | 0.472 | 1.116 |  |
| 1 | 107.467 | 124.64 | 41.597 | 18.98 | 17.482 | 7.289 | 1.384 |  |
| 1 | 1282.94 | 88.498 | 90.369 | 19.227 | 5.881 | 7.102 | 3.209 |  |
| 1 | 1154.869 | 405.999 | 43.133 | 35.517 | 4.94 | 2.514 | 2.539 |  |
| 1 | 650.742 | 619.088 | 305.883 | 21.045 | 6.549 | 0.87 | 0.576 |  |
| 1 | 184.001 | 865.318 | 666.439 | 147.72 | 15.84 | 2.73 | 0.589 |  |
| 1 | 40.446 | 73.802 | 392.93 | 301.368 | 37.357 | 2.972 | 0.514 |  |
| 1 | 92.468 | 20.348 | 67.607 | 214.052 | 152.03 | 12.739 | 2.003 |  |
| 1 | 25.779 | 65.228 | 19.575 | 50.846 | 150.131 | 76.427 | 7.561 |  |
| 1 | 261.631 | 40.768 | 70.161 | 25.781 | 60.452 | 85.771 | 19.646 |  |
| 1 | 42.148 | 213.636 | 25.132 | 37.111 | 20.577 | 47.868 | 42.903 |  |
| 1 | 209.303 | 34.43 | 184.09 | 47.965 | 56.787 | 40.367 | 125.907 |  |
| 1 | 70.313 | 70.306 | 11.47 | 20.537 | 3.963 | 4.025 | 15.265 |  |
| 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |
| 1 | 896.982 | 160.736 | 38.067 | 15.133 | 5.303 | 5.037 | 11.56 |  |
| 1 | 204.059 | 341.372 | 58.813 | 4.918 | 1.959 | 0.802 | 1.483 |  |

Table 4.10 Northeast Arctic haddock. SAM model configuration used. Updated at WKDEM 2020
\#Configuration saved: Wed Feb 12 12:57:09 2020
\# Where a matrix is specified rows corresponds to fleets and columns to ages.
\# Same number indicates same parameter used
\# Numbers (integers) starts from zero and must be consecutive
\$minAge
\# The minimum age class in the assessment
3
\$maxAge
\# The maximum age class in the assessment
13
\$maxAgePlusGroup
\# Is last age group considered a plus group for each fleet (1 yes, or 0 no).
11111

```
$keyLogFsta
# Coupling of the fishing mortality states (nomally only first row is used).
    0}1
    -1 -1 -1 -1 -1 -1 -1 -1 -1 -1 -1
    -1 -1 -1 -1 -1 -1 -1 -1 -1 -1 -1
    -1 -1 -1 -1 -1 -1 -1 -1 -1 -1 -1
    -1 
```

\$corFlag
\# Correlation of fishing mortality across ages ( 0 independent, 1 compound symmetry, 2 AR(1), 3
separable AR(1).
2
\$keyLogFpar
\# Coupling of the survey catchability parameters (nomally first row is not used, as that is covered
by fishing mortality).
-1 $-1 \begin{array}{llllllll}1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}-1$
$\begin{array}{lllllllllll}0 & 1 & 1 & 1 & 1 & 1 & -1 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{lllllllllll}2 & 3 & 3 & 3 & 3 & 4 & 4 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{llllllllll}5 & 6 & 6 & 6 & 6 & 7 & 7 & 7 & -1 & -1\end{array}-1$
$8 \quad 9 \quad 9 \quad 9 \quad 9 \quad 9 \quad 9 \quad-1-1-1-1$
\$keyQpow
\# Density dependent catchability power parameters (if any).

| -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | -1 | -1 | -1 | -1 | -1 |
| 1 | 1 | 1 | 1 | 1 | 2 | 2 | -1 | -1 | -1 | -1 |
| 3 | 3 | 3 | 3 | 3 | 4 | 4 | 4 | -1 | -1 | -1 |
| 5 | 5 | 5 | 5 | 5 | 5 | 5 | -1 | -1 | -1 | -1 |

\$keyVarF
\# Coupling of process variance parameters for $\log (\mathrm{F})$-process (nomally only first row is used)
$\begin{array}{lllllllllll}0 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1\end{array}$
-1
-1 -1 -1 -1 -1 -1 -1 $-1 \begin{array}{llll}1 & -1 & -1\end{array}$
-1 -1 -1 -1 -1 $-1 \begin{array}{lllll}1 & -1 & -1 & -1 & -1\end{array}$

\$keyVarLogN
\# Coupling of process variance parameters for $\log (\mathrm{N})$-process
01111111111
\$keyVarObs
\# Coupling of the variance parameters for the observations.
$\begin{array}{lllllllllll}0 & 1 & 2 & 2 & 2 & 2 & 2 & 2 & 2 & 2 & 2\end{array}$
$\begin{array}{lllllllllll}3 & 3 & 3 & 3 & 3 & 3 & -1 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{llllllllll}4 & 4 & 4 & 4 & 4 & 4 & 4 & -1 & -1 & -1 \\ -1\end{array}$
$\begin{array}{lllllllllll}5 & 5 & 5 & 5 & 5 & 5 & 5 & 5 & -1 & -1 & -1\end{array}$
$\begin{array}{lllllllllll}6 & 6 & 6 & 6 & 6 & 6 & 6 & -1 & -1 & -1 & -1\end{array}$
\$obsCorStruct
\# Covariance structure for each fleet ("ID" independent, "AR" AR(1), or "US" for unstructured). ।
Possible values are: "ID" "AR" "US"
"ID" "AR" "AR" "AR" "AR"
\$keyCorObs
\# Coupling of correlation parameters can only be specified if the $\operatorname{AR}(1)$ structure is chosen above.
\# NA's indicate where correlation parameters can be specified ( -1 where they cannot).
\#V1 V2 V3 V4 V5 V6 V7 V8 V9 V10

```
NA NA NA NA NA NA NA NA NA NA
    \(\begin{array}{llllllllll}0 & 1 & 1 & 1 & 2 & -1 & -1 & -1 & -1 & -1\end{array}\)
    \(\begin{array}{lllllllll}3 & 3 & 3 & 3 & 3 & 4 & -1 & -1 & -1 \\ -1\end{array}\)
    \(\begin{array}{llllllllll}5 & 5 & 5 & 5 & 5 & 6 & 6 & -1 & -1 & -1\end{array}\)
    \(\begin{array}{llllllllll}7 & 7 & 7 & 7 & 7 & 7 & -1 & -1 & -1 & -1\end{array}\)
```

\$stockRecruitmentModelCode
\# Stock recruitment code ( 0 for plain random walk, 1 for Ricker, 2 for Beverton-Holt, and 3 piece-
wise constant).
0
\$noScaledYears
\# Number of years where catch scaling is applied.
0
\$keyScaledYears
\# A vector of the years where catch scaling is applied.
\$keyParScaledYA
\# A matrix specifying the couplings of scale parameters (nrow = no scaled years, ncols = no ages).
\$fbarRange
\# lowest and higest age included in Fbar
47
\$keyBiomassTreat
\# To be defined only if a biomass survey is used ( 0 SSB index, 1 catch index, 2 FSB index, 3 total
catch, 4 total landings and 5 TSB index).
-1-1-1-1-1
\$obsLikelihoodFlag
\# Option for observational likelihood I Possible values are: "LN" "ALN"
"LN" "LN" "LN" "LN" "LN"
\$fixVarToWeight
\# If weight attribute is supplied for observations this option sets the treatment (0 relative weight,
1 fix variance to weight).
0
\$fracMixF
\# The fraction of $\mathrm{t}(3)$ distribution used in $\log \mathrm{F}$ increment distribution
0
\$fracMixN
\# The fraction of $\mathrm{t}(3)$ distribution used in $\log \mathrm{N}$ increment distribution
0
\$fracMixObs
\# A vector with same length as number of fleets, where each element is the fraction of $t(3)$ distri-
bution used in the distribution of that fleet
00000
\$constRecBreaks
\# This option is only used in combination with stock-recruitment code 3)
\$predVarObsLink
\# Coupling of parameters used in a mean-variance link for observations.
$\begin{array}{lllllllllll}0 & 1 & 2 & 2 & 2 & 2 & 2 & 2 & 2 & 2 & 2\end{array}$
$\begin{array}{lllllllllll}3 & 3 & 3 & 3 & 3 & 3 & -1 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{lllllllllll}4 & 4 & 4 & 4 & 4 & 4 & 4 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{llllllllll}5 & 5 & 5 & 5 & 5 & 5 & 5 & 5 & -1 & -1\end{array}-1$


## Table 4.11. Northeast Arctic haddock. SAM model. Estimated recruitment, spawning-stock biomass (SSB), and average fishing mortality

| Year | R(age 3) | Low | High | SSB | Low | High | Fbar(4-7) | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 72387 | 46062 | 113757 | 214451 | 191896 | 239657 | 0.755 | 0.637 | 0.894 | 387984 | 347732 | 432897 |
| 1951 | 657549 | 421933 | 1024740 | 126198 | 111962 | 142244 | 0.683 | 0.574 | 0.812 | 433412 | 338704 | 554603 |
| 1952 | 88651 | 56447 | 139228 | 101722 | 88677 | 116687 | 0.712 | 0.595 | 0.851 | 425163 | 337716 | 535254 |
| 1953 | 1235085 | 805743 | 1893203 | 120624 | 103993 | 139915 | 0.536 | 0.443 | 0.650 | 733145 | 558302 | 962743 |
| 1954 | 133361 | 85029 | 209168 | 174452 | 147488 | 206344 | 0.430 | 0.353 | 0.524 | 826557 | 650141 | 1050844 |
| 1955 | 58610 | 36972 | 92912 | 313927 | 267217 | 368803 | 0.445 | 0.368 | 0.537 | 849059 | 713766 | 1009997 |
| 1956 | 229244 | 145866 | 360280 | 368382 | 313148 | 433358 | 0.470 | 0.390 | 0.567 | 690111 | 591624 | 804993 |
| 1957 | 60266 | 38168 | 95158 | 253706 | 217108 | 296473 | 0.425 | 0.353 | 0.512 | 435085 | 377199 | 501855 |
| 1958 | 72860 | 46450 | 114287 | 182036 | 157918 | 209837 | 0.517 | 0.428 | 0.623 | 315294 | 277030 | 358844 |
| 1959 | 389171 | 254295 | 595585 | 125360 | 108680 | 144599 | 0.445 | 0.366 | 0.540 | 333166 | 273423 | 405963 |
| 1960 | 320748 | 208438 | 493573 | 112847 | 99388 | 128128 | 0.540 | 0.450 | 0.648 | 418829 | 348061 | 503987 |
| 1961 | 145185 | 94620 | 222773 | 124852 | 111078 | 140333 | 0.663 | 0.560 | 0.786 | 402474 | 349320 | 463715 |
| 1962 | 294861 | 192640 | 451325 | 125250 | 111167 | 141117 | 0.791 | 0.670 | 0.933 | 376991 | 323928 | 438745 |
| 1963 | 315359 | 207593 | 479068 | 94365 | 82948 | 107352 | 0.757 | 0.634 | 0.905 | 353624 | 295169 | 423655 |
| 1964 | 353500 | 231399 | 540029 | 84511 | 74143 | 96329 | 0.632 | 0.523 | 0.763 | 386037 | 318642 | 467687 |
| 1965 | 126853 | 81897 | 196486 | 103153 | 89857 | 118418 | 0.524 | 0.432 | 0.635 | 386407 | 325823 | 458256 |
| 1966 | 313477 | 203773 | 482241 | 145776 | 126683 | 167746 | 0.557 | 0.463 | 0.671 | 451214 | 384496 | 529509 |


| Year | R(age 3) | Low | High | SSB | Low | High | Fbar(4-7) | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1967 | 341190 | 221107 | 526492 | 151263 | 130129 | 175829 | 0.441 | 0.363 | 0.535 | 464389 | 389441 | 553759 |
| 1968 | 18013 | 11107 | 29212 | 168174 | 145329 | 194610 | 0.482 | 0.397 | 0.586 | 426984 | 361320 | 504581 |
| 1969 | 20599 | 12799 | 33151 | 167949 | 143974 | 195917 | 0.411 | 0.335 | 0.504 | 316968 | 270836 | 370956 |
| 1970 | 209787 | 134801 | 326485 | 155435 | 131552 | 183655 | 0.383 | 0.309 | 0.474 | 286902 | 241277 | 341154 |
| 1971 | 109545 | 69787 | 171952 | 127588 | 107314 | 151692 | 0.327 | 0.261 | 0.409 | 263556 | 223617 | 310629 |
| 1972 | 1052876 | 667948 | 1659631 | 128490 | 111420 | 148176 | 0.653 | 0.533 | 0.799 | 601810 | 452127 | 801049 |
| 1973 | 310449 | 202458 | 476042 | 125203 | 107368 | 146001 | 0.534 | 0.435 | 0.655 | 637223 | 507838 | 799570 |
| 1974 | 66135 | 42760 | 102289 | 153690 | 133714 | 176650 | 0.504 | 0.415 | 0.612 | 462911 | 398743 | 537405 |
| 1975 | 59421 | 38424 | 91892 | 194817 | 166555 | 227875 | 0.497 | 0.414 | 0.597 | 378920 | 328264 | 437393 |
| 1976 | 61869 | 39371 | 97225 | 196331 | 168410 | 228881 | 0.721 | 0.606 | 0.857 | 296386 | 259233 | 338863 |
| 1977 | 120514 | 75884 | 191393 | 118795 | 99987 | 141140 | 0.735 | 0.606 | 0.893 | 201315 | 172466 | 234989 |
| 1978 | 214589 | 140083 | 328722 | 81208 | 67119 | 98254 | 0.623 | 0.505 | 0.768 | 199556 | 164222 | 242492 |
| 1979 | 161504 | 105201 | 247938 | 62610 | 52588 | 74542 | 0.580 | 0.466 | 0.722 | 206831 | 171527 | 249400 |
| 1980 | 22094 | 13599 | 35894 | 62985 | 53381 | 74317 | 0.471 | 0.377 | 0.589 | 213487 | 177892 | 256205 |
| 1981 | 10280 | 6095 | 17337 | 73069 | 61627 | 86634 | 0.432 | 0.345 | 0.540 | 168620 | 141915 | 200351 |
| 1982 | 16749 | 10277 | 27298 | 68801 | 56759 | 83398 | 0.379 | 0.301 | 0.479 | 122917 | 102645 | 147193 |
| 1983 | 8656 | 5087 | 14729 | 58364 | 47816 | 71239 | 0.351 | 0.275 | 0.449 | 87932 | 73504 | 105192 |
| 1984 | 13271 | 8149 | 21611 | 53199 | 43258 | 65423 | 0.315 | 0.244 | 0.406 | 71822 | 59820 | 86232 |


| Year | R(age 3) | Low | High | SSB | Low | High | Fbar(4-7) | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 358813 | 233153 | 552199 | 49169 | 40822 | 59223 | 0.395 | 0.309 | 0.504 | 191524 | 140182 | 261671 |
| 1986 | 478572 | 311663 | 734868 | 54924 | 46468 | 64919 | 0.535 | 0.425 | 0.675 | 374796 | 293890 | 477975 |
| 1987 | 90214 | 57751 | 140923 | 77959 | 66517 | 91369 | 0.628 | 0.504 | 0.783 | 356744 | 297363 | 427982 |
| 1988 | 38984 | 24377 | 62344 | 80099 | 67250 | 95402 | 0.509 | 0.407 | 0.637 | 253948 | 214793 | 300241 |
| 1989 | 28853 | 17865 | 46599 | 84610 | 69520 | 102976 | 0.372 | 0.294 | 0.470 | 193201 | 161348 | 231341 |
| 1990 | 37125 | 23767 | 57992 | 85901 | 69709 | 105854 | 0.211 | 0.165 | 0.270 | 153622 | 127998 | 184377 |
| 1991 | 111048 | 77956 | 158188 | 100647 | 84303 | 120159 | 0.239 | 0.190 | 0.300 | 186699 | 159043 | 219165 |
| 1992 | 328727 | 233077 | 463631 | 111090 | 95809 | 128808 | 0.294 | 0.237 | 0.365 | 291322 | 243904 | 347959 |
| 1993 | 848769 | 613008 | 1175203 | 125741 | 110626 | 142922 | 0.316 | 0.257 | 0.389 | 526073 | 433781 | 638001 |
| 1994 | 396614 | 318970 | 493159 | 153834 | 137161 | 172532 | 0.371 | 0.306 | 0.451 | 650312 | 566914 | 745978 |
| 1995 | 100060 | 77811 | 128671 | 186134 | 165514 | 209324 | 0.298 | 0.250 | 0.356 | 643113 | 566516 | 730065 |
| 1996 | 99507 | 77719 | 127404 | 215730 | 192019 | 242370 | 0.366 | 0.310 | 0.431 | 557155 | 495314 | 626717 |
| 1997 | 119084 | 93193 | 152169 | 186891 | 166282 | 210055 | 0.445 | 0.376 | 0.527 | 400459 | 358952 | 446765 |
| 1998 | 63240 | 48775 | 81995 | 130850 | 115668 | 148025 | 0.452 | 0.378 | 0.541 | 266478 | 238448 | 297802 |
| 1999 | 151245 | 120741 | 189455 | 94816 | 83809 | 107270 | 0.462 | 0.383 | 0.557 | 233978 | 208477 | 262597 |
| 2000 | 83258 | 65021 | 106611 | 78075 | 68910 | 88460 | 0.341 | 0.279 | 0.417 | 214801 | 189585 | 243371 |
| 2001 | 367666 | 300041 | 450533 | 91259 | 81229 | 102526 | 0.366 | 0.303 | 0.442 | 318048 | 280668 | 360407 |
| 2002 | 395448 | 321892 | 485812 | 108683 | 96817 | 122003 | 0.351 | 0.292 | 0.423 | 436563 | 384807 | 495280 |


| Year | R(age 3) | Low | High | SSB | Low | High | Fbar(4-7) | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 340113 | 272564 | 424403 | 136879 | 122623 | 152791 | 0.424 | 0.358 | 0.503 | 506909 | 450642 | 570201 |
| 2004 | 260359 | 212216 | 319424 | 155689 | 139461 | 173805 | 0.387 | 0.328 | 0.456 | 493539 | 441891 | 551224 |
| 2005 | 366492 | 300172 | 447466 | 166962 | 149621 | 186313 | 0.404 | 0.344 | 0.476 | 510380 | 457657 | 569177 |
| 2006 | 157564 | 127155 | 195244 | 151329 | 135466 | 169050 | 0.369 | 0.312 | 0.437 | 439168 | 393891 | 489649 |
| 2007 | 543223 | 441281 | 668715 | 153562 | 137718 | 171230 | 0.384 | 0.323 | 0.455 | 504466 | 450324 | 565117 |
| 2008 | 1112513 | 913961 | 1354200 | 163092 | 145133 | 183272 | 0.314 | 0.262 | 0.377 | 738154 | 647137 | 841971 |
| 2009 | 1025284 | 845638 | 1243094 | 183533 | 163348 | 206213 | 0.260 | 0.216 | 0.311 | 996702 | 871947 | 1139306 |
| 2010 | 240955 | 195431 | 297083 | 248053 | 220499 | 279050 | 0.244 | 0.206 | 0.291 | 1130768 | 991062 | 1290169 |
| 2011 | 117224 | 92480 | 148588 | 355613 | 315855 | 400375 | 0.255 | 0.217 | 0.301 | 1178847 | 1040816 | 1335183 |
| 2012 | 340386 | 276667 | 418780 | 475908 | 419566 | 539815 | 0.220 | 0.186 | 0.260 | 1175999 | 1040560 | 1329067 |
| 2013 | 119057 | 94420 | 150121 | 523943 | 460492 | 596137 | 0.148 | 0.124 | 0.177 | 1005601 | 890548 | 1135517 |
| 2014 | 411335 | 336043 | 503497 | 523619 | 463357 | 591718 | 0.154 | 0.128 | 0.185 | 983944 | 880258 | 1099843 |
| 2015 | 72464 | 56494 | 92950 | 497402 | 444871 | 556135 | 0.190 | 0.159 | 0.227 | 874947 | 787488 | 972120 |
| 2016 | 212760 | 170769 | 265075 | 489847 | 438583 | 547104 | 0.261 | 0.219 | 0.310 | 803199 | 722937 | 892372 |
| 2017 | 194179 | 156196 | 241399 | 410620 | 369903 | 455820 | 0.351 | 0.296 | 0.416 | 702033 | 634303 | 776994 |
| 2018 | 367841 | 295751 | 457503 | 303265 | 271126 | 339214 | 0.404 | 0.339 | 0.481 | 617524 | 553251 | 689263 |
| 2019 | 821773 | 668831 | 1009689 | 234446 | 206986 | 265549 | 0.433 | 0.355 | 0.527 | 695945 | 612581 | 790655 |
| 2020 | 441844 | 354723 | 550361 | 204484 | 175372 | 238429 | 0.438 | 0.347 | 0.554 | 722596 | 623367 | 837621 |


| Year | R(age 3) | Low | High | SSB | Low | High | Fbar(4-7) | Low | High | TSB | Low |  | 648860 | 532298 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2021 | 153680 | 110687 | 213373 | 200849 | 162390 | 248417 |  | 790945 |  |  |  |  |  |  |

Table 4.12. Northeast Arctic haddock. SAM model estimated fishing mortality-at-age. SAM model.

| Year age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0.096 | 0.412 | 0.706 | 0.849 | 1.052 | 0.886 | 0.886 | 0.886 | 0.886 | 0.886 | 0.886 |
| 1951 | 0.086 | 0.359 | 0.617 | 0.773 | 0.981 | 0.884 | 0.884 | 0.884 | 0.884 | 0.884 | 0.884 |
| 1952 | 0.092 | 0.380 | 0.641 | 0.797 | 1.029 | 0.933 | 0.933 | 0.933 | 0.933 | 0.933 | 0.933 |
| 1953 | 0.067 | 0.282 | 0.473 | 0.588 | 0.802 | 0.737 | 0.737 | 0.737 | 0.737 | 0.737 | 0.737 |
| 1954 | 0.048 | 0.207 | 0.357 | 0.468 | 0.689 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 |
| 1955 | 0.046 | 0.199 | 0.368 | 0.502 | 0.710 | 0.600 | 0.600 | 0.600 | 0.600 | 0.600 | 0.600 |
| 1956 | 0.050 | 0.210 | 0.389 | 0.549 | 0.733 | 0.621 | 0.621 | 0.621 | 0.621 | 0.621 | 0.621 |
| 1957 | 0.047 | 0.198 | 0.367 | 0.492 | 0.643 | 0.547 | 0.547 | 0.547 | 0.547 | 0.547 | 0.547 |
| 1958 | 0.058 | 0.235 | 0.450 | 0.601 | 0.781 | 0.690 | 0.690 | 0.690 | 0.690 | 0.690 | 0.690 |
| 1959 | 0.059 | 0.228 | 0.409 | 0.521 | 0.620 | 0.566 | 0.566 | 0.566 | 0.566 | 0.566 | 0.566 |
| 1960 | 0.089 | 0.317 | 0.537 | 0.633 | 0.672 | 0.616 | 0.616 | 0.616 | 0.616 | 0.616 | 0.616 |
| 1961 | 0.117 | 0.406 | 0.682 | 0.782 | 0.783 | 0.694 | 0.694 | 0.694 | 0.694 | 0.694 | 0.694 |
| 1962 | 0.147 | 0.502 | 0.853 | 0.941 | 0.867 | 0.722 | 0.722 | 0.722 | 0.722 | 0.722 | 0.722 |
| 1963 | 0.133 | 0.471 | 0.805 | 0.909 | 0.845 | 0.681 | 0.681 | 0.681 | 0.681 | 0.681 | 0.681 |
| 1964 | 0.097 | 0.360 | 0.634 | 0.769 | 0.765 | 0.647 | 0.647 | 0.647 | 0.647 | 0.647 | 0.647 |


| Year age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 0.077 | 0.292 | 0.513 | 0.635 | 0.656 | 0.566 | 0.566 | 0.566 | 0.566 | 0.566 | 0.566 |
| 1966 | 0.090 | 0.328 | 0.563 | 0.667 | 0.670 | 0.555 | 0.555 | 0.555 | 0.555 | 0.555 | 0.555 |
| 1967 | 0.072 | 0.268 | 0.446 | 0.515 | 0.535 | 0.465 | 0.465 | 0.465 | 0.465 | 0.465 | 0.465 |
| 1968 | 0.084 | 0.297 | 0.490 | 0.554 | 0.588 | 0.513 | 0.513 | 0.513 | 0.513 | 0.513 | 0.513 |
| 1969 | 0.079 | 0.267 | 0.428 | 0.469 | 0.481 | 0.416 | 0.416 | 0.416 | 0.416 | 0.416 | 0.416 |
| 1970 | 0.082 | 0.262 | 0.402 | 0.428 | 0.439 | 0.381 | 0.381 | 0.381 | 0.381 | 0.381 | 0.381 |
| 1971 | 0.073 | 0.233 | 0.351 | 0.355 | 0.366 | 0.324 | 0.324 | 0.324 | 0.324 | 0.324 | 0.324 |
| 1972 | 0.193 | 0.503 | 0.759 | 0.696 | 0.654 | 0.545 | 0.545 | 0.545 | 0.545 | 0.545 | 0.545 |
| 1973 | 0.199 | 0.486 | 0.641 | 0.530 | 0.477 | 0.381 | 0.381 | 0.381 | 0.381 | 0.381 | 0.381 |
| 1974 | 0.179 | 0.431 | 0.547 | 0.515 | 0.522 | 0.460 | 0.460 | 0.460 | 0.460 | 0.460 | 0.460 |
| 1975 | 0.195 | 0.459 | 0.548 | 0.494 | 0.487 | 0.417 | 0.417 | 0.417 | 0.417 | 0.417 | 0.417 |
| 1976 | 0.289 | 0.647 | 0.785 | 0.723 | 0.728 | 0.640 | 0.640 | 0.640 | 0.640 | 0.640 | 0.640 |
| 1977 | 0.322 | 0.713 | 0.852 | 0.719 | 0.658 | 0.559 | 0.559 | 0.559 | 0.559 | 0.559 | 0.559 |
| 1978 | 0.223 | 0.546 | 0.726 | 0.644 | 0.576 | 0.505 | 0.505 | 0.505 | 0.505 | 0.505 | 0.505 |
| 1979 | 0.160 | 0.443 | 0.670 | 0.652 | 0.557 | 0.502 | 0.502 | 0.502 | 0.502 | 0.502 | 0.502 |
| 1980 | 0.101 | 0.316 | 0.525 | 0.563 | 0.481 | 0.459 | 0.459 | 0.459 | 0.459 | 0.459 | 0.459 |
| 1981 | 0.085 | 0.273 | 0.472 | 0.538 | 0.444 | 0.428 | 0.428 | 0.428 | 0.428 | 0.428 | 0.428 |
| 1982 | 0.075 | 0.244 | 0.411 | 0.477 | 0.385 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 | 0.380 |


| Year age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1983 | 0.077 | 0.247 | 0.388 | 0.428 | 0.342 | 0.341 | 0.341 | 0.341 | 0.341 | 0.341 | 0.341 |
| 1984 | 0.069 | 0.226 | 0.347 | 0.376 | 0.308 | 0.293 | 0.293 | 0.293 | 0.293 | 0.293 | 0.293 |
| 1985 | 0.075 | 0.257 | 0.412 | 0.481 | 0.429 | 0.412 | 0.412 | 0.412 | 0.412 | 0.412 | 0.412 |
| 1986 | 0.088 | 0.315 | 0.541 | 0.666 | 0.619 | 0.588 | 0.588 | 0.588 | 0.588 | 0.588 | 0.588 |
| 1987 | 0.097 | 0.359 | 0.644 | 0.786 | 0.724 | 0.658 | 0.658 | 0.658 | 0.658 | 0.658 | 0.658 |
| 1988 | 0.071 | 0.278 | 0.511 | 0.655 | 0.592 | 0.537 | 0.537 | 0.537 | 0.537 | 0.537 | 0.537 |
| 1989 | 0.055 | 0.219 | 0.388 | 0.466 | 0.414 | 0.362 | 0.362 | 0.362 | 0.362 | 0.362 | 0.362 |
| 1990 | 0.029 | 0.126 | 0.214 | 0.255 | 0.248 | 0.231 | 0.231 | 0.231 | 0.231 | 0.231 | 0.231 |
| 1991 | 0.031 | 0.136 | 0.243 | 0.291 | 0.285 | 0.262 | 0.262 | 0.262 | 0.262 | 0.262 | 0.262 |
| 1992 | 0.032 | 0.146 | 0.291 | 0.367 | 0.372 | 0.341 | 0.341 | 0.341 | 0.341 | 0.341 | 0.341 |
| 1993 | 0.026 | 0.128 | 0.291 | 0.407 | 0.439 | 0.398 | 0.398 | 0.398 | 0.398 | 0.398 | 0.398 |
| 1994 | 0.024 | 0.124 | 0.305 | 0.476 | 0.579 | 0.544 | 0.544 | 0.544 | 0.544 | 0.544 | 0.544 |
| 1995 | 0.019 | 0.099 | 0.231 | 0.366 | 0.497 | 0.489 | 0.489 | 0.489 | 0.489 | 0.489 | 0.489 |
| 1996 | 0.024 | 0.123 | 0.286 | 0.439 | 0.614 | 0.620 | 0.620 | 0.620 | 0.620 | 0.620 | 0.620 |
| 1997 | 0.032 | 0.158 | 0.374 | 0.534 | 0.716 | 0.683 | 0.683 | 0.683 | 0.683 | 0.683 | 0.683 |
| 1998 | 0.038 | 0.178 | 0.402 | 0.552 | 0.677 | 0.676 | 0.676 | 0.676 | 0.676 | 0.676 | 0.676 |
| 1999 | 0.045 | 0.203 | 0.432 | 0.560 | 0.652 | 0.624 | 0.624 | 0.624 | 0.624 | 0.624 | 0.624 |
| 2000 | 0.033 | 0.159 | 0.325 | 0.412 | 0.468 | 0.438 | 0.438 | 0.438 | 0.438 | 0.438 | 0.438 |


| Year age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 0.034 | 0.162 | 0.355 | 0.455 | 0.491 | 0.449 | 0.449 | 0.449 | 0.449 | 0.449 | 0.449 |
| 2002 | 0.031 | 0.151 | 0.321 | 0.453 | 0.481 | 0.423 | 0.423 | 0.423 | 0.423 | 0.423 | 0.423 |
| 2003 | 0.036 | 0.169 | 0.366 | 0.531 | 0.629 | 0.570 | 0.570 | 0.570 | 0.570 | 0.570 | 0.570 |
| 2004 | 0.034 | 0.158 | 0.329 | 0.483 | 0.578 | 0.547 | 0.547 | 0.547 | 0.547 | 0.547 | 0.547 |
| 2005 | 0.037 | 0.163 | 0.336 | 0.494 | 0.624 | 0.603 | 0.603 | 0.603 | 0.603 | 0.603 | 0.603 |
| 2006 | 0.036 | 0.159 | 0.316 | 0.443 | 0.558 | 0.549 | 0.549 | 0.549 | 0.549 | 0.549 | 0.549 |
| 2007 | 0.037 | 0.158 | 0.319 | 0.465 | 0.592 | 0.572 | 0.572 | 0.572 | 0.572 | 0.572 | 0.572 |
| 2008 | 0.025 | 0.112 | 0.230 | 0.383 | 0.532 | 0.524 | 0.524 | 0.524 | 0.524 | 0.524 | 0.524 |
| 2009 | 0.020 | 0.088 | 0.178 | 0.307 | 0.465 | 0.479 | 0.479 | 0.479 | 0.479 | 0.479 | 0.479 |
| 2010 | 0.020 | 0.084 | 0.168 | 0.287 | 0.438 | 0.489 | 0.489 | 0.489 | 0.489 | 0.489 | 0.489 |
| 2011 | 0.021 | 0.088 | 0.184 | 0.303 | 0.446 | 0.489 | 0.489 | 0.489 | 0.489 | 0.489 | 0.489 |
| 2012 | 0.020 | 0.082 | 0.159 | 0.264 | 0.373 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 |
| 2013 | 0.015 | 0.061 | 0.108 | 0.171 | 0.252 | 0.311 | 0.311 | 0.311 | 0.311 | 0.311 | 0.311 |
| 2014 | 0.017 | 0.069 | 0.121 | 0.178 | 0.249 | 0.345 | 0.345 | 0.345 | 0.345 | 0.345 | 0.345 |
| 2015 | 0.022 | 0.089 | 0.160 | 0.223 | 0.288 | 0.396 | 0.396 | 0.396 | 0.396 | 0.396 | 0.396 |
| 2016 | 0.029 | 0.115 | 0.224 | 0.312 | 0.392 | 0.509 | 0.509 | 0.509 | 0.509 | 0.509 | 0.509 |
| 2017 | 0.037 | 0.150 | 0.305 | 0.439 | 0.511 | 0.590 | 0.590 | 0.590 | 0.590 | 0.590 | 0.590 |
| 2018 | 0.037 | 0.155 | 0.348 | 0.523 | 0.590 | 0.640 | 0.640 | 0.640 | 0.640 | 0.640 | 0.640 |


| Year age | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2019 | 0.035 | 0.155 | 0.374 | 0.596 | 0.604 | 0.600 | 0.600 | 0.600 | 0.600 | 0.600 | 0.600 |
| 2020 | 0.035 | 0.156 | 0.385 | 0.598 | 0.615 | 0.579 | 0.579 | 0.579 | 0.579 | 0.579 | 0.579 |
| 2021 |  |  |  |  |  |  |  |  |  |  |  |

Table 4.13. Northeast Arctic haddock. SAM model. Estimated stock numbers-at-age.

| Year age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 72387 | 101009 | 76017 | 37150 | 46935 | 16676 | 4880 | 2688 | 1381 | 1458 | 2057 |
| 1951 | 657549 | 47705 | 46081 | 27475 | 12803 | 12509 | 5437 | 1943 | 1014 | 446 | 1091 |
| 1952 | 88651 | 438929 | 30695 | 19192 | 9000 | 4349 | 3848 | 1638 | 740 | 358 | 506 |
| 1953 | 1235085 | 52138 | 209525 | 14008 | 6354 | 2642 | 1334 | 1051 | 533 | 255 | 309 |
| 1954 | 133361 | 913544 | 26058 | 91355 | 6875 | 2330 | 1091 | 550 | 387 | 198 | 228 |
| 1955 | 58610 | 84501 | 631189 | 14601 | 52376 | 3092 | 919 | 454 | 237 | 160 | 168 |
| 1956 | 229244 | 40701 | 55883 | 324913 | 7240 | 17802 | 1441 | 402 | 215 | 114 | 153 |
| 1957 | 60266 | 151466 | 27728 | 36033 | 111034 | 3106 | 6150 | 704 | 168 | 100 | 131 |
| 1958 | 72860 | 39770 | 92930 | 15488 | 20893 | 40149 | 1644 | 2509 | 354 | 84 | 120 |
| 1959 | 389171 | 51295 | 26037 | 40026 | 7337 | 7294 | 14884 | 731 | 899 | 148 | 88 |
| 1960 | 320748 | 266359 | 35741 | 15664 | 16981 | 3484 | 3678 | 6151 | 365 | 369 | 109 |
| 1961 | 145185 | 192859 | 145259 | 17681 | 6976 | 8042 | 1598 | 1508 | 2792 | 158 | 204 |
| 1962 | 294861 | 86481 | 92421 | 59752 | 6747 | 2709 | 3285 | 659 | 610 | 1159 | 139 |


| Year age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1963 | 315359 | 177692 | 37947 | 26417 | 17576 | 2650 | 1088 | 1226 | 273 | 244 | 536 |
| 1964 | 353500 | 199644 | 75558 | 12273 | 7678 | 5842 | 1227 | 440 | 508 | 123 | 346 |
| 1965 | 126853 | 240169 | 115011 | 30342 | 4168 | 2789 | 2265 | 536 | 199 | 218 | 212 |
| 1966 | 313477 | 82668 | 159195 | 62307 | 12375 | 1706 | 1278 | 942 | 273 | 92 | 187 |
| 1967 | 341190 | 201060 | 43604 | 72639 | 24821 | 4868 | 791 | 602 | 450 | 133 | 132 |
| 1968 | 18013 | 248132 | 118431 | 21878 | 36202 | 12529 | 2349 | 410 | 314 | 233 | 138 |
| 1969 | 20599 | 11699 | 142453 | 55382 | 10694 | 15788 | 5755 | 1164 | 197 | 157 | 175 |
| 1970 | 209787 | 12601 | 7442 | 70596 | 25187 | 5928 | 8046 | 3010 | 645 | 106 | 186 |
| 1971 | 109545 | 135078 | 7121 | 4480 | 33447 | 12303 | 3367 | 4542 | 1695 | 372 | 163 |
| 1972 | 1052876 | 80012 | 82395 | 4549 | 3103 | 17570 | 6739 | 2020 | 2777 | 1031 | 316 |
| 1973 | 310449 | 611103 | 46689 | 23226 | 1698 | 1550 | 7634 | 2898 | 926 | 1381 | 612 |
| 1974 | 66135 | 168872 | 250030 | 16572 | 10670 | 885 | 1018 | 4471 | 1685 | 549 | 1231 |
| 1975 | 59421 | 37507 | 90353 | 140384 | 6794 | 4948 | 449 | 564 | 2145 | 815 | 939 |
| 1976 | 61869 | 33814 | 16493 | 44274 | 79181 | 3147 | 2774 | 247 | 336 | 1149 | 973 |
| 1977 | 120514 | 31955 | 13774 | 6432 | 17629 | 30320 | 1281 | 1184 | 103 | 150 | 807 |
| 1978 | 214589 | 55473 | 9805 | 4432 | 2903 | 7738 | 15125 | 627 | 564 | 45 | 431 |
| 1979 | 161504 | 118148 | 23372 | 3261 | 2038 | 1408 | 4103 | 7088 | 338 | 273 | 226 |
| 1980 | 22094 | 103045 | 58844 | 8328 | 1152 | 1050 | 718 | 2169 | 3494 | 175 | 240 |


| Year age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 10280 | 15556 | 63778 | 26434 | 3456 | 551 | 560 | 381 | 1144 | 1721 | 215 |
| 1982 | 16749 | 6731 | 11059 | 31900 | 10551 | 1721 | 278 | 308 | 219 | 627 | 960 |
| 1983 | 8656 | 11414 | 4623 | 6826 | 13527 | 5614 | 984 | 146 | 178 | 128 | 805 |
| 1984 | 13271 | 5143 | 6723 | 2738 | 3892 | 8834 | 2874 | 577 | 80 | 105 | 519 |
| 1985 | 358813 | 8928 | 2896 | 3609 | 1787 | 2574 | 5370 | 1840 | 369 | 51 | 399 |
| 1986 | 478572 | 277557 | 5190 | 1600 | 1853 | 994 | 1477 | 2795 | 1027 | 206 | 263 |
| 1987 | 90214 | 251326 | 157099 | 2536 | 656 | 793 | 470 | 680 | 1205 | 471 | 209 |
| 1988 | 38984 | 69536 | 135665 | 46741 | 1070 | 233 | 319 | 205 | 302 | 507 | 280 |
| 1989 | 28853 | 25825 | 49166 | 71076 | 12181 | 553 | 95 | 152 | 99 | 146 | 365 |
| 1990 | 37125 | 21055 | 17098 | 26048 | 32816 | 5474 | 358 | 59 | 87 | 57 | 277 |
| 1991 | 111048 | 25165 | 13652 | 14116 | 20258 | 20295 | 3130 | 252 | 40 | 57 | 205 |
| 1992 | 328727 | 84057 | 16045 | 10130 | 10434 | 12634 | 12657 | 1883 | 167 | 26 | 158 |
| 1993 | 848769 | 223913 | 57735 | 10760 | 5933 | 6253 | 7669 | 7276 | 1047 | 103 | 107 |
| 1994 | 396614 | 587436 | 154930 | 31942 | 4717 | 3143 | 3765 | 4809 | 4340 | 594 | 117 |
| 1995 | 100060 | 226590 | 435698 | 78166 | 14754 | 2118 | 1430 | 1880 | 2211 | 2156 | 341 |
| 1996 | 99507 | 61789 | 169995 | 248671 | 32136 | 7295 | 1100 | 713 | 945 | 1113 | 1277 |
| 1997 | 119084 | 55471 | 38253 | 96439 | 103120 | 13962 | 2515 | 500 | 315 | 419 | 1105 |
| 1998 | 63240 | 80491 | 34945 | 18197 | 36788 | 39134 | 5215 | 991 | 213 | 133 | 718 |


| Year age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 151245 | 48598 | 47807 | 17437 | 8943 | 15880 | 13968 | 1913 | 411 | 95 | 395 |
| 2000 | 83258 | 120846 | 31027 | 21381 | 6915 | 4355 | 6581 | 5478 | 813 | 189 | 237 |
| 2001 | 367666 | 68635 | 94932 | 16897 | 10167 | 3556 | 2621 | 3527 | 2687 | 439 | 242 |
| 2002 | 395448 | 300091 | 52067 | 48539 | 9168 | 5544 | 1920 | 1468 | 1939 | 1411 | 359 |
| 2003 | 340113 | 261408 | 196328 | 34543 | 25078 | 4620 | 3530 | 1249 | 843 | 1100 | 1007 |
| 2004 | 260359 | 172273 | 166036 | 112867 | 16305 | 11103 | 2162 | 1680 | 629 | 400 | 1083 |
| 2005 | 366492 | 171572 | 94829 | 110334 | 51502 | 6674 | 5666 | 1165 | 744 | 318 | 809 |
| 2006 | 157564 | 219442 | 109811 | 52161 | 45104 | 21091 | 3242 | 2875 | 569 | 352 | 551 |
| 2007 | 543223 | 121375 | 168189 | 61734 | 26885 | 19538 | 8239 | 1776 | 1508 | 293 | 455 |
| 2008 | 1112513 | 468268 | 98184 | 105061 | 22152 | 14209 | 7305 | 3341 | 914 | 737 | 371 |
| 2009 | 1025284 | 728429 | 383448 | 62880 | 40729 | 10451 | 5495 | 3239 | 1485 | 513 | 620 |
| 2010 | 240955 | 691017 | 611521 | 237174 | 32624 | 15444 | 4886 | 2807 | 1654 | 800 | 679 |
| 2011 | 117224 | 194409 | 563046 | 432721 | 124025 | 14466 | 6299 | 2164 | 1383 | 855 | 862 |
| 2012 | 340386 | 73679 | 139426 | 404212 | 273255 | 55692 | 6248 | 2614 | 1060 | 724 | 988 |
| 2013 | 119057 | 202072 | 58279 | 96150 | 278419 | 130583 | 24094 | 3248 | 1443 | 609 | 1036 |
| 2014 | 411335 | 74058 | 147167 | 50176 | 89044 | 149208 | 62995 | 11011 | 1919 | 910 | 1046 |
| 2015 | 72464 | 289943 | 66054 | 93229 | 40823 | 70958 | 75588 | 26121 | 5433 | 1045 | 1069 |
| 2016 | 212760 | 49328 | 170881 | 46203 | 62182 | 33887 | 50649 | 38657 | 13022 | 2602 | 1031 |


| Year age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | 194179 | 178302 | 34064 | 111000 | 28140 | 36803 | 19145 | 22167 | 18206 | 5675 | 1498 |
| 2018 | 367841 | 136644 | 126603 | 24863 | 44110 | 14515 | 18040 | 9062 | 9354 | 8647 | 3170 |
| 2019 | 821773 | 245167 | 89266 | 64306 | 16672 | 17748 | 6762 | 7639 | 3760 | 3946 | 4324 |
| 2020 | 441844 | 529584 | 163506 | 48047 | 23575 | 8750 | 7624 | 3230 | 3185 | 1816 | 3452 |
| 2021 | 153680 | 259641 | 362981 | 65434 | 24257 | 9282 | 3972 | 3483 | 1479 | 1459 | 2410 |

Table 4.14. Northeast Arctic haddock. SAM model. Natural mortality estimated.

| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1951 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1952 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1953 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1954 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1955 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1956 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1957 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1958 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1959 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1960 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |


| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1962 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1963 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1964 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1965 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1966 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1967 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1968 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1969 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1970 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1971 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1972 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1973 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1974 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1975 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1976 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1977 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1978 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |


| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1980 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1981 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1982 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1983 | 0.347 | 0.258 | 0.245 | 0.242 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1984 | 0.216 | 0.224 | 0.214 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1985 | 0.209 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1986 | 0.640 | 0.262 | 0.200 | 0.210 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1987 | 0.200 | 0.207 | 0.421 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1988 | 0.379 | 0.200 | 0.200 | 0.393 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1989 | 0.200 | 0.200 | 0.200 | 0.232 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1990 | 0.328 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1991 | 0.202 | 0.216 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1992 | 0.216 | 0.205 | 0.203 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1993 | 0.253 | 0.248 | 0.274 | 0.260 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1994 | 0.289 | 0.216 | 0.295 | 0.227 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1995 | 0.379 | 0.341 | 0.319 | 0.291 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1996 | 0.724 | 0.319 | 0.253 | 0.283 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |


| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 0.503 | 0.267 | 0.255 | 0.284 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1998 | 0.230 | 0.291 | 0.265 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 1999 | 0.200 | 0.207 | 0.278 | 0.260 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2000 | 0.214 | 0.200 | 0.215 | 0.245 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2001 | 0.210 | 0.200 | 0.226 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2002 | 0.323 | 0.213 | 0.200 | 0.204 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2003 | 0.417 | 0.250 | 0.208 | 0.203 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2004 | 0.414 | 0.301 | 0.201 | 0.228 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2005 | 0.396 | 0.302 | 0.231 | 0.270 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2006 | 0.223 | 0.214 | 0.275 | 0.211 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2007 | 0.297 | 0.200 | 0.239 | 0.320 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2008 | 0.371 | 0.279 | 0.266 | 0.338 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2009 | 0.402 | 0.248 | 0.284 | 0.256 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2010 | 0.358 | 0.249 | 0.273 | 0.285 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2011 | 0.529 | 0.468 | 0.310 | 0.227 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2012 | 0.593 | 0.313 | 0.204 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2013 | 0.460 | 0.340 | 0.248 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2014 | 0.283 | 0.206 | 0.219 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |


| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 0.344 | 0.402 | 0.211 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2016 | 0.305 | 0.200 | 0.248 | 0.229 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2017 | 0.330 | 0.296 | 0.233 | 0.412 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2018 | 0.442 | 0.250 | 0.265 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2019 | 0.361 | 0.269 | 0.200 | 0.276 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2020 | 0.412 | 0.360 | 0.323 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |
| 2021 | 0.412 | 0.360 | 0.323 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 | 0.200 |

Table 4.15. Northeast Arctic haddock. Summary XSA (p-shrinkage not applied, F shrinkage= 0.5). Thu Apr 23 16:16:08 2020.

| YEAR | RECR_a3 | TOTBIO | TOTSPB | LANDINGS | YIELDSSB | SOPCOFAC | FBAR 4-7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 82517 | 242696 | 134602 | 132125 | 0.9816 | 1.5897 | 0.8305 |
| 1951 | 669592 | 356206 | 101130 | 120077 | 1.1874 | 1.2272 | 0.6238 |
| 1952 | 76993 | 235716 | 57527 | 127660 | 2.2191 | 1.7404 | 0.7243 |
| 1953 | 1276811 | 512541 | 82624 | 123920 | 1.4998 | 1.4279 | 0.5157 |
| 1954 | 152912 | 538732 | 117456 | 156788 | 1.3349 | 1.474 | 0.3802 |
| 1955 | 68791 | 486182 | 178951 | 202286 | 1.1304 | 1.536 | 0.5112 |
| 1956 | 208993 | 475286 | 243778 | 213924 | 0.8775 | 1.2623 | 0.4328 |
| 1957 | 66305 | 326559 | 186324 | 123583 | 0.6633 | 1.2455 | 0.4322 |
| 1958 | 87212 | 277194 | 157018 | 112672 | 0.7176 | 1.1252 | 0.5185 |


| YEAR | RECR_a3 | TOTBIO | TOTSPB | LANDINGS | YIELDSSB | SOPCOFAC | FBAR 4-7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1959 | 398937 | 365304 | 133348 | 88211 | 0.6615 | 0.9405 | 0.3672 |
| 1960 | 289884 | 401516 | 114703 | 154651 | 1.3483 | 1.0411 | 0.484 |
| 1961 | 130882 | 391762 | 130068 | 193224 | 1.4856 | 0.9942 | 0.6362 |
| 1962 | 291125 | 346736 | 118945 | 187408 | 1.5756 | 1.0518 | 0.8 |
| 1963 | 341475 | 311066 | 82694 | 146224 | 1.7683 | 1.1458 | 0.8645 |
| 1964 | 398845 | 302301 | 63902 | 99158 | 1.5517 | 1.3572 | 0.6522 |
| 1965 | 124503 | 358459 | 95547 | 118578 | 1.241 | 1.1507 | 0.4935 |
| 1966 | 294241 | 388088 | 127654 | 161778 | 1.2673 | 1.1621 | 0.583 |
| 1967 | 362769 | 468419 | 154643 | 136397 | 0.882 | 0.9984 | 0.4147 |
| 1968 | 23990 | 421753 | 169593 | 181726 | 1.0715 | 0.9976 | 0.503 |
| 1969 | 21471 | 342797 | 184231 | 130820 | 0.7101 | 0.882 | 0.3972 |
| 1970 | 202641 | 286838 | 156150 | 88257 | 0.5652 | 0.9762 | 0.3575 |
| 1971 | 122645 | 345853 | 168613 | 78905 | 0.468 | 0.7638 | 0.2465 |
| 1972 | 1252757 | 619817 | 123068 | 266153 | 2.1626 | 1.0883 | 0.6918 |
| 1973 | 342252 | 604302 | 114785 | 322226 | 2.8072 | 1.1656 | 0.5362 |
| 1974 | 69287 | 604427 | 200945 | 221157 | 1.1006 | 0.8946 | 0.4315 |
| 1975 | 60222 | 493447 | 256440 | 175758 | 0.6854 | 0.8957 | 0.4268 |
| 1976 | 66905 | 307480 | 206755 | 137264 | 0.6639 | 1.12 | 0.5705 |


| YEAR | RECR_a3 | тотвIO | TOTSPB | LANDINGS | YIELDSSB | SOPCOFAC | FBAR 4-7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 134417 | 229040 | 141828 | 110158 | 0.7767 | 1.09 | 0.6832 |
| 1978 | 213614 | 256138 | 130603 | 95422 | 0.7306 | 0.9219 | 0.5112 |
| 1979 | 176286 | 318567 | 129566 | 103623 | 0.7998 | 0.7684 | 0.5515 |
| 1980 | 34826 | 343544 | 133268 | 87889 | 0.6595 | 0.7568 | 0.3978 |
| 1981 | 13441 | 293155 | 148313 | 77153 | 0.5202 | 0.7174 | 0.4012 |
| 1982 | 17394 | 212027 | 127285 | 46955 | 0.3689 | 0.7224 | 0.3093 |
| 1983 | 9563 | 104393 | 71491 | 24600 | 0.3441 | 1.0373 | 0.2715 |
| 1984 | 13434 | 83502 | 64118 | 20945 | 0.3267 | 1.0547 | 0.2498 |
| 1985 | 288300 | 182799 | 62012 | 45052 | 0.7265 | 0.9761 | 0.32 |
| 1986 | 529936 | 343817 | 62309 | 100563 | 1.6139 | 1.0484 | 0.4388 |
| 1987 | 109761 | 333920 | 75055 | 154916 | 2.064 | 0.992 | 0.5958 |
| 1988 | 54817 | 260029 | 78423 | 95255 | 1.2146 | 0.9955 | 0.499 |
| 1989 | 26591 | 212726 | 91989 | 58518 | 0.6361 | 0.9774 | 0.3892 |
| 1990 | 36885 | 170781 | 95306 | 27182 | 0.2852 | 1.0159 | 0.1562 |
| 1991 | 104289 | 195374 | 110525 | 36216 | 0.3277 | 1.0374 | 0.2082 |
| 1992 | 207573 | 269180 | 125749 | 59922 | 0.4765 | 0.9797 | 0.2838 |
| 1993 | 661827 | 442193 | 130412 | 82379 | 0.6317 | 1.0031 | 0.359 |
| 1994 | 292252 | 542649 | 144884 | 135186 | 0.9331 | 1.0056 | 0.425 |


| YEAR | RECR_a3 | TOTBIO | TOTSPB | LANDINGS | YIELDSSB | SOPCOFAC | FBAR 4-7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 97799 | 538481 | 158892 | 142448 | 0.8965 | 1.0247 | 0.3825 |
| 1996 | 102077 | 472118 | 184556 | 178128 | 0.9652 | 1.0175 | 0.4235 |
| 1997 | 115566 | 349254 | 162754 | 154359 | 0.9484 | 1.0519 | 0.4862 |
| 1998 | 58271 | 249707 | 124288 | 100630 | 0.8097 | 1.0113 | 0.4235 |
| 1999 | 230876 | 252735 | 93038 | 83195 | 0.8942 | 1.021 | 0.4212 |
| 2000 | 89446 | 250625 | 85299 | 68944 | 0.8083 | 1.026 | 0.2802 |
| 2001 | 366245 | 356725 | 110567 | 89640 | 0.8107 | 0.9903 | 0.2795 |
| 2002 | 342709 | 443325 | 128727 | 114798 | 0.8918 | 1.011 | 0.3173 |
| 2003 | 224429 | 474128 | 150713 | 138926 | 0.9218 | 1.019 | 0.4292 |
| 2004 | 225230 | 455037 | 157794 | 158279 | 1.0031 | 1.0192 | 0.3795 |
| 2005 | 347443 | 471039 | 168020 | 158298 | 0.9421 | 1.0029 | 0.49 |
| 2006 | 157072 | 415213 | 142651 | 153157 | 1.0736 | 0.9938 | 0.405 |
| 2007 | 668942 | 496479 | 140120 | 161525 | 1.1528 | 0.9916 | 0.4228 |
| 2008 | 1339631 | 738745 | 146275 | 155604 | 1.0638 | 0.9928 | 0.3902 |
| 2009 | 1454218 | 1075831 | 168600 | 200061 | 1.1866 | 1.0019 | 0.3525 |
| 2010 | 526318 | 1253906 | 233140 | 249200 | 1.0689 | 0.9994 | 0.293 |
| 2011 | 245890 | 1275393 | 336181 | 309785 | 0.9215 | 0.9978 | 0.3175 |
| 2012 | 381957 | 1158133 | 419440 | 315627 | 0.7525 | 0.9994 | 0.266 |


| YEAR | RECR_a3 | TOTBIO | TOTSPB | LANDINGS | YIELDSSB | SOPCOFAC | FBAR 4-7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2013 | 156234 | 988402 | 465852 | 193744 | 0.4159 | 0.9967 | 0.134 |
| 2014 | 389701 | 993569 | 511632 | 177522 | 0.347 | 0.9968 | 0.111 |
| 2015 | 103379 | 934929 | 524799 | 194756 | 0.3711 | 0.9953 | 0.1558 |
| 2016 | 260916 | 846474 | 496913 | 233183 | 0.4693 | 1.0006 | 0.2208 |
| 2017 | 200597 | 729410 | 417225 | 227588 | 0.5455 | 0.994 | 0.3318 |
| 2018 | 368406 | 618897 | 307333 | 191276 | 0.6224 | 0.9943 | 0.3915 |
| 2019 | 871151 | 709103 | 236928 | 175402 | 0.7403 | 0.9963 | 0.4545 |
| 2020 | 415726 | 760305 | 214036 | 182468 | 0.8525 | 0.9962 | 0.4345 |

Table 4.16. Northeast Arctic haddock. Input data for recruitment prediction (RCT3)- recruits as 3 year-olds. Recr: recruitment estimate from SAM 2020 NT1: Norwegian Russian winter bottom trawl survey age 1 NT2: Norwegian Russian winter bottom trawl survey age 2 NT3: Norwegian Russian winter bottom trawl survey age 3 NAK1: Norwegian Russian winter acoustic survey age 1 NAK2: Norwegian Russian winter acoustic survey age 2 NAK3: Norwegian Russian winter acoustic survey age 3 ECO1: Ecosystem survey age 1. ECO2: Ecosystem survey age 2. The Russian survey (RT) was discontinued in 2017 and has not been used for recruitment.

| Year class | Recr. | NT1 | NT2 | NT3 | NAK1 | NAK2 | NAK3 | EC01 | ECO2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 848769 | NA | NA | NA | NA | NA | NA | NA | NA |
| 1991 | 396614 | NA | NA | 315 | NA | NA | 349 | NA | NA |
| 1992 | 100060 | NA | 225 | 55 | NA | 188 | 42 | NA | NA |
| 1993 | 99507 | 604 | 200 | 56 | 888 | 89 | 30 | NA | NA |
| 1994 | 119084 | 1429 | 265 | 80 | 1198 | 95 | 57 | NA | NA |
| 1995 | 63240 | 301 | 91 | 22 | 133 | 27 | 34 | NA | NA |
| 1996 | 151245 | 1118 | 197 | 57 | 509 | 151 | 84 | NA | NA |


| Year class | Recr. | NT1 | NT2 | NT3 | NAK1 | NAK2 | NAK3 | ECO1 | ECO2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 83258 | 248 | 83 | 24 | 211 | 30 | 36 | NA | NA |
| 1998 | 367666 | 1208 | 437 | 294 | 653 | 405 | 234 | NA | NA |
| 1999 | 395448 | 832 | 447 | 313 | 1063 | 266 | 255 | NA | NA |
| 2000 | 340113 | 1231 | 475 | 352 | 753 | 268 | 204 | NA | NA |
| 2001 | 260359 | 1700 | 472 | 173 | 1315 | 362 | 151 | NA | NA |
| 2002 | 366492 | 3327 | 707 | 318 | 2744 | 467 | 221 | NA | 268 |
| 2003 | 157564 | 701 | 386 | 79 | 529 | 144 | 56 | 189 | 114 |
| 2004 | 543223 | 4473 | 1310 | 443 | 2277 | 625 | 209 | 604 | 929 |
| 2005 | 1112513 | 4945 | 1685 | 1591 | 2091 | 954 | 812 | 2270 | 1819 |
| 2006 | 1025284 | 3731 | 2042 | 1230 | 2016 | 1754 | 884 | 988 | 1292 |
| 2007 | 240955 | 853 | 317 | 103 | 778 | 209 | 128 | 322 | 144 |
| 2008 | 117224 | 563 | 80 | 53 | 444 | 86 | 54 | 135 | 65 |
| 2009 | 340386 | 1635 | 354 | 316 | 1559 | 288 | 192 | 274 | 114 |
| 2010 | 119057 | 676 | 137 | 57 | 429 | 95 | 67 | 105 | 42 |
| 2011 | 411335 | 1867 | 490 | 381 | 1583 | 407 | 335 | 591 | 223 |
| 2012 | 72464 | 345 | 124 | 31 | 293 | 110 | 24 | 156 | 75 |
| 2013 | 212760 | 1281 | 342 | 163 | 1839 | 247 | 72 | 265 | 145 |
| 2014 | 194179 | 1134 | 562 | 135 | 1593 | 107 | 81 | 320 | 145 |


| Year class | Recr. | NT1 | NT2 | NT3 | NAK1 | NAK2 | NAK3 | EC01 | ECO2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 367841 | 2299 | 770 | 336 | 1276 | 331 | 170 | 794 | 189 |
| 2016 | 821773 | 5065 | 1676 | 1076 | 3344 | 806 | 508 | 936 | NA |
| 2017 | 441844 | 3823 | 1125 | 424 | 2931 | 688 | 286 | NA | 585 |
| 2018 | 153680 | 1898 | 268 | 118 | 1545 | 261 | 43 | 379 | 58 |
| 2019 | NA | 111 | 31 | NA | 273 | 32 | NA | 27 | NA |
| 2020 | NA | 462 | NA | NA | 435 | NA | NA | NA | NA |

## Table 4.17. Northeast Arctic haddock Analysis by RCT3 ver3.1-R translation

## Analysis by RCT3 ver3.1-R translation

Data for 8 surveys over 31 year classes : 1990-2020
Regression type $=\mathrm{C}$
Tapered time weighting applied
power $=3$ over 20 years
Survey weighting not applied
Final estimates shrunk towards mean
Estimates with S.E.'S greater than that of mean included
Minimum S.E. for any survey taken as 0.2
Minimum of 3 points used for regression

Forecast/Hindcast variance correction used.


```
    0.06460
    0.04973
    0.01810
yearclass:2019
    index slope intercept se rsquare n indices prediction se.pred
    NT1 1.0341 4.886 0.3606 0.8393 20 4.715 9.762 0.5627
    NT2 0.8802 7.128 0.3358}0.859420 3.455 10.170 0.4915
    NT3 NA NA NA NA NA NA NA NA
    NAK1 1.2736 3.396 0.5859 0.6643 20 5.612 10.543 0.7771
    NAK2 0.9857 6.947 0.3531 0.8468 20 3.490 10.388 0.4971
    NAK3 NA NA NA NANA NA NA NA
    EC01 1.1232 5.823 0.4206 0.8056 15 3.326 9.558 0.6831
    ECO2 NA NA NA NANA NA NA NA
VPA Mean NA NA NA NA 29 NA 12.518}0.782
WAP.weights
    0 . 1 8 8 2 1
    0.24677
        NA
    0.09871
    0 . 2 4 1 1 6
        NA
    0.12772
        NA
    0.09743
yearclass:2020
    index slope intercept se rsquare n indices prediction se.pred
    NT1 1.031 4.895 0.3624 0.837419}60.137 11.22 0.4597
    NT2 NA NA NA NANA NA NA NA
    NT3 NA NA NA NANA NA NA NA
    NAK1 1.257 3.489 0.5814 0.666719 6.078 11.13 0.7321
    NAK2 NA NA NA NA NA NA NA NA
```

| NAK3 | NA | NA | NA | NA NA | NA | NA | NA |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :---: |
| EC01 NA | NA | NA | NA NA | NA | NA | NA |  |
| ECO2 NA | NA | NA | NA NA | NA | NA | NA |  |
| VPA Mean NA | NA | NA | NA 29 | NA | 12.51 | 0.7770 |  |
| WAP.weights |  |  |  |  |  |  |  |
| 0.5733 |  |  |  |  |  |  |  |
| NA |  |  |  |  |  |  |  |
| NA |  |  |  |  |  |  |  |
| 0.2260 |  |  |  |  |  |  |  |
| NA |  |  |  |  |  |  |  |
| NA |  |  |  |  |  |  |  |
| NA |  |  |  |  |  |  |  |
| NA |  |  |  |  |  |  |  |
| 0.2006 |  |  |  |  |  |  |  |

WAP $\log$ WAP int.se
yearclass:2018 18887712.150 .09103
yearclass:2019 3073610.330 .24414
yearclass:2020 9470211.460 .34806

## Table 4.18. Northeast Arctic haddock. Prediction with management option table: Input data (based on SAM estimates

"MFDP version 1a"
"Run: 2021"
"Time and date: 22:28 19.04.2021"
"Fbar age range: 4-7"
""


| 9 | . | 0.2 | 0.946 |  | 0 |  | 0 | 2.716 | 0.6257 | 2.066 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 10 | . | 0.2 | 0.971 |  | 0 |  | 0 | 3.085 | 0.6257 | 2.314 |
| 11 | . | 0.2 | 1 |  | 0 |  | 0 | 3.686 | 0.6257 | 2.379 |
| 12 | . | 0.2 | 1 |  | 0 |  | 0 | 3.624 | 0.6257 | 2.799 |
| 13 | . | 0.2 | 1 |  | 0 |  | 0 | 4.059 | 0.6257 | 3.468 |
| 2023 |  |  |  |  |  |  |  |  |  |  |
| Age | $N$ | M | Mat | PF |  | PM |  | sWt | Sel | cWt |
| 3 | 94702 | 0.405 | 0.03 |  | 0 |  | 0 | 0.315 | 0.0368 | 0.753 |
| 4 | . | 0.293 | 0.082 |  | 0 |  | 0 | 0.497 | 0.160 | 0.922 |
| 5 | . | 0.263 | 0.192 |  | 0 |  | 0 | 0.766 | 0.3808 | 1.138 |
| 6 | . | 0.225 | 0.401 |  | 0 |  | 0 | 1.117 | 0.5906 | 1.380 |
| 7 | . | 0.2 | 0.649 |  | 0 |  | 0 | 1.526 | 0.6223 | 1.696 |
| 8 | . | 0.2 | 0.833 |  | 0 |  | 0 | 2.000 | 0.6257 | 1.884 |
| 9 | . | 0.2 | 0.937 |  | 0 |  | 0 | 2.610 | 0.6257 | 2.070 |
| 10 | . | 0.2 | 0.973 |  | 0 |  | 0 | 3.145 | 0.6257 | 2.283 |
| 11 | . | 0.2 | 1 |  | 0 |  | 0 | 3.507 | 0.6257 | 2.334 |
| 12 | . | 0.2 | 1 |  | 0 |  | 0 | 3.854 | 0.6257 | 2.775 |
| 13 | . | 0.2 | 1 |  | 0 |  | 0 | 3.938 | 0.6257 | 3.471 |

Table 4.19. Northeast Arctic haddock. Prediction with management option table for 2021-2023 (TAC constraint applied for intermediate year MFDP version 1 a

Run: 2021
2021MFDP Index file 19.04.2021
Time and date: 22:28 19.04.2021
Fbar age range: 4-7


## Table 4.20. Northeast Arctic haddock. Prediction single option table for 2020-2022 based on HCR

MFDP version 1 a
Run: Fhcr
Time and date: 22:38 19.04.2021
Fbar age range: 4-7

| Year: |  | 2021 | F multiplier: | 0.9932 | Fbar: | 0.4355 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
|  | 3 | 0.0366 | 4541 | 3147 | 153680 | 39803 | 4457 | 1154 | 4457 | 1154 |
|  | 4 | 0.1592 | 33255 | 30561 | 259641 | 128263 | 21031 | 10389 | 21031 | 10389 |
|  | 5 | 0.3782 | 101347 | 119589 | 362981 | 295104 | 76589 | 62267 | 76589 | 62267 |
|  | 6 | 0.5866 | 26289 | 38776 | 65434 | 79960 | 29445 | 35982 | 29445 | 35982 |
|  | 7 | 0.6181 | 10240 | 18872 | 24257 | 43032 | 17829 | 31628 | 17829 | 31628 |
|  | 8 | 0.6215 | 3934 | 7553 | 9282 | 21200 | 8224 | 18783 | 8224 | 18783 |
|  | 9 | 0.6215 | 1683 | 3619 | 3972 | 10577 | 3742 | 9964 | 3742 | 9964 |
|  | 10 | 0.6215 | 1476 | 3562 | 3483 | 11421 | 3410 | 11181 | 3410 | 11181 |
|  | 11 | 0.6215 | 627 | 1560 | 1479 | 5094 | 1479 | 5094 | 1479 | 5094 |
|  | 12 | 0.6215 | 618 | 1770 | 1459 | 5477 | 1459 | 5477 | 1459 | 5477 |
|  | 13 | 0.6215 | 1021 | 3527 | 2410 | 8929 | 2410 | 8929 | 2410 | 8929 |
| Total |  |  | 185031 | 232537 | 888078 | 648860 | 170074 | 200849 | 170074 | 200849 |
| Year: |  | 2022 | F multiplier: | 0.7982 | Fbar: | 0.35 |  |  |  |  |
| Age |  | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
|  | 3 | 0.0294 | 732 | 518 | 30736 | 8391 | 922 | 252 | 922 | 252 |
|  | 4 | 0.128 | 10320 | 9340 | 98822 | 47533 | 7708 | 3708 | 7708 | 3708 |
|  | 5 | 0.304 | 38324 | 44226 | 165188 | 129507 | 32872 | 25772 | 32872 | 25772 |
|  | 6 | 0.4714 | 64912 | 91785 | 191163 | 220602 | 79906 | 92212 | 79906 | 92212 |


|  | 7 | 0.4967 | 10397 | 18142 | 29062 | 46761 | 19733 | 31751 | 19733 | 31751 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 8 | 0.4994 | 3846 | 7426 | 10704 | 23452 | 9323 | 20427 | 9323 | 20427 |
|  | 9 | 0.4994 | 1467 | 3030 | 4082 | 11087 | 3862 | 10488 | 3862 | 10488 |
|  | 10 | 0.4994 | 628 | 1452 | 1747 | 5389 | 1696 | 5233 | 1696 | 5233 |
|  | 11 | 0.4994 | 550 | 1309 | 1532 | 5646 | 1532 | 5646 | 1532 | 5646 |
|  | 12 | 0.4994 | 234 | 654 | 650 | 2357 | 650 | 2357 | 650 | 2357 |
|  | 13 | 0.4994 | 611 | 2120 | 1702 | 6906 | 1702 | 6906 | 1702 | 6906 |
| Total |  |  | 132021 | 180003 | 535387 | 507632 | 159906 | 204751 | 159906 | 204751 |
| Year: |  | 2023 | F multiplier: | 0.7982 | Fbar: | 0.35 |  |  |  |  |
| Age |  | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
|  | 3 | 0.0294 | 2256 | 1699 | 94702 | 29831 | 2841 | 895 | 2841 | 895 |
|  | 4 | 0.128 | 2079 | 1917 | 19907 | 9894 | 1632 | 811 | 1632 | 811 |
|  | 5 | 0.304 | 15050 | 17127 | 64869 | 49690 | 12455 | 9540 | 12455 | 9540 |
|  | 6 | 0.4714 | 31818 | 43909 | 93703 | 104666 | 37575 | 41971 | 37575 | 41971 |
|  | 7 | 0.4967 | 34082 | 57803 | 95269 | 145381 | 61830 | 94352 | 61830 | 94352 |
|  | 8 | 0.4994 | 5202 | 9800 | 14479 | 28958 | 12061 | 24122 | 12061 | 24122 |
|  | 9 | 0.4994 | 1911 | 3955 | 5318 | 13881 | 4983 | 13007 | 4983 | 13007 |
|  | 10 | 0.4994 | 729 | 1664 | 2028 | 6379 | 1973 | 6207 | 1973 | 6207 |
|  | 11 | 0.4994 | 312 | 728 | 868 | 3044 | 868 | 3044 | 868 | 3044 |
|  | 12 | 0.4994 | 273 | 759 | 761 | 2933 | 761 | 2933 | 761 | 2933 |
|  | 13 | 0.4994 | 420 | 1457 | 1169 | 4602 | 1169 | 4602 | 1169 | 4602 |
| Total |  |  | 94131 | 140817 | 393074 | 399259 | 138149 | 201485 | 138149 | 201485 |

## Table 4.21. Northeast Arctic haddock. Yield-per-recruit. Input data and results.

MFYPR version $2 a$
Run: 2021YPR
Time and date: 22:25 19.04.2021
Yield per results

| FMult | Fbar | CatchNos | Yield | StockNos | Biomass | SpwnNosJan | SSBJan | SpwnNosSpwn | SSBSpwn |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 4.2321 | 6.4432 | 1.9203 | 5.0608 | 1.9203 | 5.0608 |
| 0.1 | 0.0495 | 0.1087 | 0.2095 | 3.7039 | 4.7588 | 1.4293 | 3.4316 | 1.4293 | 3.4316 |
| 0.2 | 0.099 | 0.1778 | 0.3169 | 3.3732 | 3.7718 | 1.1326 | 2.4938 | 1.1326 | 2.4938 |
| 0.3 | 0.1485 | 0.2264 | 0.3785 | 3.1444 | 3.1343 | 0.9353 | 1.9004 | 0.9353 | 1.9004 |
| 0.4 | 0.198 | 0.2629 | 0.417 | 2.9753 | 2.6943 | 0.7954 | 1.5002 | 0.7954 | 1.5002 |
| 0.5 | 0.2475 | 0.2917 | 0.4427 | 2.8442 | 2.3754 | 0.6914 | 1.2172 | 0.6914 | 1.2172 |
| 0.6 | 0.297 | 0.3153 | 0.4607 | 2.7389 | 2.1352 | 0.6114 | 1.0098 | 0.6114 | 1.0098 |
| 0.7 | 0.3465 | 0.3351 | 0.4739 | 2.6519 | 1.9487 | 0.5482 | 0.8532 | 0.5482 | 0.8532 |
| 0.8 | 0.396 | 0.3521 | 0.4839 | 2.5784 | 1.8001 | 0.497 | 0.7322 | 0.497 | 0.7322 |
| 0.9 | 0.4455 | 0.367 | 0.4917 | 2.5152 | 1.6793 | 0.4549 | 0.6367 | 0.4549 | 0.6367 |
| 1 | 0.495 | 0.3802 | 0.4979 | 2.4601 | 1.5791 | 0.4196 | 0.56 | 0.4196 | 0.56 |
| 1.1 | 0.5445 | 0.392 | 0.5029 | 2.4114 | 1.4948 | 0.3897 | 0.4974 | 0.3897 | 0.4974 |
| 1.2 | 0.594 | 0.4028 | 0.5071 | 2.3679 | 1.4229 | 0.3641 | 0.4458 | 0.3641 | 0.4458 |
| 1.3 | 0.6435 | 0.4126 | 0.5105 | 2.3287 | 1.3608 | 0.3419 | 0.4026 | 0.3419 | 0.4026 |
| 1.4 | 0.693 | 0.4216 | 0.5135 | 2.2931 | 1.3066 | 0.3226 | 0.3661 | 0.3226 | 0.3661 |
| 1.5 | 0.7425 | 0.4299 | 0.516 | 2.2605 | 1.2588 | 0.3055 | 0.3349 | 0.3055 | 0.3349 |
| 1.6 | 0.792 | 0.4377 | 0.5182 | 2.2306 | 1.2164 | 0.2903 | 0.3081 | 0.2903 | 0.3081 |
| 1.7 | 0.8415 | 0.445 | 0.5201 | 2.2029 | 1.1784 | 0.2768 | 0.2849 | 0.2768 | 0.2849 |
| 1.8 | 0.891 | 0.4519 | 0.5218 | 2.1771 | 1.1442 | 0.2647 | 0.2646 | 0.2647 | 0.2646 |
| 1.9 | 0.9405 | 0.4583 | 0.5234 | 2.1531 | 1.1131 | 0.2538 | 0.2468 | 0.2538 | 0.2468 |
| 2 | 0.99 | 0.4644 | 0.5247 | 2.1306 | 1.0848 | 0.2439 | 0.2311 | 0.2439 | 0.2311 |

## Reference point

| Fbar(3-13) | 1 | 0.495 |
| :--- | :--- | :--- |
| FMax | $>=1000000$ |  |
| F0.1 | 0.4082 | 0.2021 |
| F35\%SPR | 0.3284 | 0.1626 |

Weights in kilograms


Figure 4.1 Landings, fishing mortality, recruitment, and spawning-stock biomass of Northeast Arctic haddock 1950-2021. Fishing mortality and spawning-stock biomass are given with point wise $95 \%$ confidence intervals (shaded areas).


Figure 4.2. Northeast Arctic haddock; on step ahead residuals for the final SAM run. Blue circles indicate positive residuals (observations larger than predicted) and red circles indicate negative residuals.


Figure 4.3. Northeast Arctic haddock. 5 year retrospective plots of SSB (top right), fishing mortality (top left), TSB (bottom left), and recruitment (bottom right) for years 2000-2021 (SAM with 95\% confidence intervals).


Figure 4.4. Results of assessment of NEA haddock. Fbar, TSB, recruits and SSB from AFWG 2020 (last year) and AFWG 2021 from 2001 and onwards. The last red points on the blue lines are forecasts from last year.


Figure 4.5. Northeast Arctic haddock. Retrospective plots of SSB, fishing mortality and recruitment for assessment years 1950-2020 (XSA without P shrinkage, F shrinkage= 0.5 )




Figure 4.6. Northeast Arctic haddock. Retrospective plots of SSB, fishing mortality and recruitment for assessment years 1990-2020 from TSVPA model (see WD 22).


Figure 4.7. Comparison of results of assessment of NEA haddock. Recruits, biomass, spawning biomass and F in 19902020 by different models: medium SAM estimates, XSA with setting mentioned at section 4.9 and TISVPA with settings as mentioned at WDXX.


Figure 4. 8. Standard selection pattern model (red) used for short-term forecasts at AFWG 2021.


Figure 4.9. Comparisons of catch data by age 2020 from InterCatch with forecasts from AFWG 2019 and 2020. Top: catch number of individuals, middle: catch weights, bottom: yield.

## 5 Northeast Arctic saithe ${ }^{1}$

### 5.1 The fishery (Table 5.1 and Table 5.2, Figure 5.1)

Currently, the main fleets targeting saithe are trawl, purse-seine, gillnet, handline, and Danish seine. Landings of saithe were highest in 1970-1976 with an average of 239000 t and a maximum of 265000 t in 1970. This period was followed by a sharp decline to a level of about 160000 t in the years 1978-1984, while in 1985 to 1991 the landings ranged from $67000-123000 \mathrm{t}$. After 1991 landings increased, ranging between 136000 t (in 2000) and 212000 t (in 2006), followed by a decline to 132000 t in 2015. In 2020 landings were 169405 t and 188176 t in 2020.

Discarding, although illegal, occurs in the saithe fishery, but is not considered a major problem in the assessment. Due to its nearshore distribution saithe is virtually inaccessible for commercial gears during the first couple of years of life and there are no reports indicating overall high discard rates in the Norwegian fisheries. There are reported incidents of slipping in the purse-seine fishery, mainly related to minimum landing size. Observations from non-Norwegian commercial trawlers indicate that discarding may occur when vessels targeting other species catch saithe, for which they may not have a quota or have filled it. However, there are no quantitative estimates of the level of discarding available.

### 5.1.1 ICES advice applicable to 2021 and 2022

The advice from ICES for 2021 was as follows:

- ICES advised that catches in 2021 should be no more than 197779 t .

The advice from ICES for 2022 was as follows:

- ICES advised that catches in 2022 should be no more than 197212 t.


### 5.1.2 Management applicable in 2021 and 2022

Management of Saithe in subareas 1 and 2 is by TAC and technical measures. For 2021, The Norwegian Ministry of Trade, Industry and Fisheries set the TAC according to the advice from ICES, i.e. 197779 t.

For 2022, The Norwegian Ministry of Trade, Industry and Fisheries set the TAC according to the advice from ICES, i.e. 197212 t.

### 5.1.3 The fishery in 2021 and expected landings in 2022

Provisional figures show that the landings in 2021 were approximately 188176 t , which is 9603 t lower than the TAC of 197779 t .

Since the WG does not have any prognosis of total landings in 2022 available, the TAC of 197212 t is used in the projections.

[^7]
# 5.2 Commercial catch-effort data and research vessel surveys 

### 5.2.1 Catch-per-unit-effort

The NEA saithe interbenchmark protocol (IBP; ICES CM 2014/ACOM: 53) recommended leaving out the CPUE time-series in the model tuning (see section 5.3.5). A detailed description of the Norwegian trawl CPUE and its previous use is given in the stock annex.

### 5.2.2 Survey results (Figure 5.1-5.2)

An ad hoc subgroup of the AFWG was held to review proposed changes to several survey series using the new "StoX" survey computation methodology on 16 and 17 April 2017 at the JRC, Italy. The survey series reviewed included the coastal survey for saithe for the period 2003 to 2017. StoX is a new program developed at IMR Norway, to produce a more robust, transparent, and automated method of computing survey series. The method is currently used in ICES assessments (for example for NSS herring). For the saithe survey series, a WD was presented to the group (Mehl et al., 2018a), examining the differences between the previous survey series and those resulting from StoX in survey indices by age, as well as mean weight and mean length. During the meeting consistency plots were produced for each survey and showed to have a better fit with the StoX series compared to the old series. The meeting concluded that the new StoX survey series should be used to replace the previous survey series in AFWG stock assessment, but that once the assessment model is run the residuals and fits to the data should be examined to check for unexpected detrimental effects on model performance. The resulting SAM model fits using the old and the StoX survey series (using data for both survey series up to 2016, but excluding the 2003 StoX estimate, as this was considered abnormally high) were practically the same, without any detrimental effects on model performance.

The echo abundance observed in 2021 (Staby et al., in press) increased by 30\% compared to 2020 and was about $20 \%$ higher than the average for 2003-2020. The abundance estimated with StoX increased with $8 \%$ compared to 2020 . This increase is the result of higher estimates of $5-9$-yearold saithe, which were between $24-33 \%$ higher than in 2020 . Only estimates of 3 -year old saithe were below the 2020 estimate. The proportion of saithe in the southern part of the survey area (south of the Lofoten islands between $62^{\circ-670} \mathrm{~N}$ ) increased from about $20 \%$ in 1997 to above $60 \%$ in 2008, decreased in later years and was $20 \%$ in 2021, similar to the 2020 proportion.

### 5.2.3 Recruitment indices

Owing to the nearshore distribution of juvenile saithe, obtaining early estimates of recruitment for ages $0-2$ has not been possible so far. The survey recruitment indices are strongly dependent on the extent to which 2-4 year old saithe have migrated from the coastal areas and become available to the acoustic saithe survey on the banks, and this varies between years. Also, observations from an observer programme, established in 2000 to start a 0-group index series (Borge and Mehl, WD 21 2002) did not seem to reflect the dynamics in year-class strength very well. (Mehl, WD 6 2007; Mehl, WD 7 to WKROUND 2010). The programme was consequently terminated in 2010.

### 5.3 Data used in the assessment

### 5.3.1 Catch numbers-at-age (Table 5.3)

Total Norwegian landings by gear and landings data for all other countries from 2021 were updated based on the official total catch (preliminary) reported to ICES or to Norwegian authorities.

Age composition data for 2021 were available for Norwegian landings. The biological sampling of all gear groups, areas, and quarters was sufficient to produce a reliable catch-at-age matrix for 2021. As in previous years age data from the Danish seine and bottom-trawl fishery were combined to increase the number of samples by area and quarter, thereby improving the estimate of catch-at-age numbers.

Catch-at-age estimates (numbers and mean weight and length-at-age) were produced with StoXReca (version 3.4) for the 2021 assessment ${ }^{2}$. Comparative runs with the older ECA program were not possible for the 2021 data since data in the required format is not available anymore. This is the second year that catch-at-age estimates are produced with StoX-Reca for input in the SAM assessment. In previous years catch-at-age was estimated manually, and until 2020 with ECA.

### 5.3.2 Weight-at-age (Table 5.4)

Constant weights-at-age values for age groups 3-11 are used for the period 1960-1979, whereas estimated values for the $12+$ group vary during this period. For subsequent years, annual estimates of weight-at-age in the catches are used. Weight-at-age in the stock is assumed to be the same as weight-at-age in the catch. Compared to 2020, estimated weight-at-age for age groups 3-12+ differed only slightly in 2021.

### 5.3.3 Natural mortality

A fixed natural mortality of 0.2 for all age groups was used both in the assessment and the forecast.

### 5.3.4 Maturity-at-age (Table 5.5)

A 3-year running average is used for the period from 1985 and onwards (2-year average for the first and last year). Inconsistencies between proportion mature fish and trends in SSB and recruitment since 2008 resulted in the NEA saithe IBP to recommend the use of a constant maturity ogive for the years from 2007 and onwards based on the average 2005-2007 (ICES CM 2014/ACOM: 53). Analysis are currently being done to investigate which method, i.e. macroscopic determination, otolith spawning rings or histological analysis, is the most reliable to determine the maturity stage.

### 5.3.5 Tuning data (Table 5.6)

Until the 2005 WG, the XSA tuning was based on three dataseries: CPUE from Norwegian purseseine and Norwegian trawl and indices from a Norwegian acoustic survey. The 2005 WG found rather large and variable $\log \mathrm{q}$ residuals and large S.E. $\log \mathrm{q}$ for the purse-seine fleet, as well as strong year effects, and in the combined tuning the fleet got low scaled weights. The WG decided

[^8]not to include the purse-seine tuning fleet in the analysis. This was confirmed by new analyses at the 2010 benchmark assessment (ICES CM 2010/ACOM:36). The trawl CPUE series on the other hand did not show the trends in stock size abundance of NEA saithe in later years. In the more recent years there were signs of changes in fishing strategy, with fewer and shorter fishing periods and a smaller proportion of directed saithe fishery (Mehl and Fotland, WD 20 2013).

Analyses of the two remaining tuning series done at the 2010 benchmark assessment indicated that there had been a shift in catchability around year 2002. The survey was redesigned in 2003, and the fishery to a larger degree targeted older ages. Permanent breaks were made in both tuning series in 2002. The acoustic survey, compared with the trawl CPUE time-series, seemed to track the stock changes better, both in abundance and distribution.
The sensitivity runs presented to the IBP (Fotland WD 302014 IBP NEA saithe) clearly showed that the residual pattern got worse (strong year effects) when using both tuning series in SAM. It became obvious that SAM tries to fit something in between both contradicting data sources. Therefore, it had to be decided whether one data source was more reliable or whether both data sources should be considered leading to a fit in between both extremes. Given that CPUE series should not be used when larger changes in fishing patterns occur (selectivity, spatial distribution of the fleet, change between targeted and bycatch fishery) it was recommended to leave out the CPUE time-series in its current form for now (ICES CM 2014/ACOM: 53). Another reason was that the proportion of catches covered by the index had decreased steadily between 2002 and 2011, further questioning the representativeness of the CPUE index. However, it may be worth trying alternative CPUE indices (e.g. one index for the targeted fishery only and one index for the fishery with saithe bycatches) until the next benchmark.
The following two tuning fleets are thus used in the present assessment (by the time this report was written the new ICES name for this survey was not available)

- NOcoast-Aco-4Q: Indices from the Norwegian acoustic survey 1994-2001, age groups 3 to 7 .
- NOcoast-Aco-4Q: Indices from the Norwegian acoustic survey 2002-2021, age groups 3 to 7 .


### 5.4 SAM runs and settings (Table 5.7)

In connection with the NEA saithe IBP a number of exploratory SAM runs were performed. Model settings and results are presented in working documents included in the IBP report (ICES CM 2014/ACOM: 53).

SAM model settings and configuration in 2021 were the same as in previous simulations.

- Tuning data: Acoustic survey series (age 3-7) only, time-series split (1994-2001 and 2002present);
- Maturity data: Ogives for the years 2007 and later based on the average of the 2005-2007 data;
- $\quad$ Flat exploitation pattern for age groups $8+$;
- Correlated Fs between age groups and time;
- Beverton-Holt stock-recruitment relationship used to estimate recent recruitment.


### 5.5 Final assessment run (Table 5.8 to Table 5.11, Figure 5.3-5.6)

The state-space assessment model (SAM) was used for the final run. SAM catchabilities and negative log likelihood values are given in Table 5.8.

Figure 5.3 presents normalized residuals for the total catches and the two parts of the acoustic tuning series. There are both year- and age effects and the second part of the series seems to perform better than the first part. Figure 5.4 shows plots of the stock numbers from the SAM vs. tuning indices.

### 5.5.1 SAM F, N, and SSB results (Tables 5.9-5.11, Figures 5.5-5.6)

The estimated fishing mortality ( $\mathrm{F}_{4-7}$ ) in 2020 was 0.219 (AFWG 2021), which is higher than 0.187 from this year's assessment and below the $\mathrm{F}_{\mathrm{pa}}$ of 0.35 . The fishing mortality ( $\mathrm{F}_{4-7}$ ) in 2021 was estimated at 0.186 . From 1997 to 2009 fishing mortality was below $\mathrm{F}_{\mathrm{pa}}$, but started to increase in 2005 and was above $\mathrm{F}_{\mathrm{pa}}$ in 2010-2012.

Fishing mortality and stock size have in the last decade generally been considerably over- and underestimated respectively. Due to the changes made to the assessment following the benchmark assessment workshop in 2010 (ICES CM 2010/ACOM: 36) and later the NEA saithe IBP in 2014 (ICES CM 2014/ACOM: 53), the retrospective patterns have improved considerably, as is illustrated in Figure 5.7. Based on the 2021 assessment the SSB has in recent years been both slightly over and underestimated while $\mathrm{F}_{4-7}$ has been generally overestimated.

The SAM-estimate of the 2014 year class was considered to be reliable enough to be used in the projections. In previous assessments the value of the 3-year olds in the last data year has been set to the long-term geometrical mean, and the value of the year class at age 4 were obtained by applying Pope's approximation. Since 2007 the 2007, 2010, 2013, and 2016 year classes have been above the long-term geometric mean, while in the other years, year-class strength has been considered average or below.

The total biomass (ages 3+) was above the long-term (1960-2021) average from 1997 to 2008, reached a local maximum in 2005, and declined below the average level between 2011 and 2015. Since 2016 it has been above the long-term average, and in 2021 was estimated at > 1140000 tonnes, the highest estimate in the time-series. The SSB was above the long-term mean from 2000 to 2009, decreased below the average between 2010 to 2013, and has been above the long-term average since 2014. SSB has been above $\mathrm{B}_{\mathrm{pa}}(220000 \mathrm{t}$ ) since 1996 (Figure 5.5).

### 5.5.2 Recruitment (Table 5.10, Figure 5.5)

Catches of age group 3 have varied considerably during the period 2004-2017 (Table 5.10). Until the 2005 WG, RCT3-runs were conducted to estimate the corresponding year classes, with 2 and 3 year olds from the acoustic survey as input together with XSA numbers. However, it was stated several times in the ACOM Technical Minutes that it would be more transparent to use the longterm geometric mean (GM) recruitment. GM values were therefore used in the 2005-2014 since the issue was not discussed at the IBP when SAM was adopted as assessment model. During the 2015 AFWG assessment, analyses were performed to investigate if the last year recruitment value from SAM could be used instead of the long-term GM (for method description refer to Stock Annex). Results from this analysis showed that the retrospective runs of SAM gave better estimates of recruitment than the geometric mean and consequently estimates of the recruiting year class (3 year olds in the last data year) from the SAM were accepted for the last year.

### 5.6 Reference points (Figure 5.5)

In 2010 the age span was expanded from 11+ to 15+ and important XSA parameter settings were changed (ICES CM 2010/ACOM: 36). LIM reference points were re-estimated at the 2010 WG according to the methodology outlined in ICES CM 2003/ACFM: 15, while the PA reference point estimation was based on the old procedure (ICES CM 1998/ACFM: 10). The results were not very much different from the previous analyses performed in 2005 (ICES CM 2005/ACFM: 20), and it
was decided not to change the existing LIM and PA reference points. The shift from XSA to SAM resulted in only minor changes in estimated fishing mortality, spawning-stock-biomass and recruitment and no new reference points were estimated. Reference points were estimated as: Blim 136000 t , Bpa 220000 t , $\mathrm{F}_{\mathrm{MP}} 0.32$ Flim 0.58, and $\mathrm{F}_{\mathrm{pa}} 0.35$.

### 5.6.1 Harvest control rule

In 2007 ICES evaluated the harvest control rule for setting the annual fishing quota (TAC) for Northeast Arctic saithe. ICES concluded that the HCR was consistent with the precautionary approach for all simulated data and settings, including a rebuilding situation under the condition that the assessment uncertainty and error are not greater than those calculated from historic data. This also held true when an implementation error (difference between TAC and catch) equal to the historic level was included. The HCR was implemented the same year. It contains the following elements:

- Estimate the average TAC level for the coming 3 years based on $\mathrm{F}_{\mathrm{mp}}$. TAC for the next year will be set to this level as a starting value for the 3-year period.
- The year after, the TAC calculation for the next 3 years is repeated based on the updated information about the stock development. However, the TAC should not be changed by more than $15 \%$ compared with the previous year's TAC.
- If the spawning-stock-biomass (SSB) at the beginning of the year for which the quota is set (first year of prediction), is below $B_{p a}$, the procedure for establishing TAC should be based on a fishing mortality that is linearly reduced from $F_{m p}$ at $S S B=B_{p a}$ to 0 at SSB equal to zero. At SSB levels below $\mathrm{B}_{\mathrm{pa}}$ in any of the operational years (current year and 3 years of prediction) there should be no limitations on the year-to-year variations in TAC.

In 2011 the evaluation was repeated taking into account the changes made to the assessment after the 2010 benchmark assessment (ICES CM 2010/ACOM: 36). The analyses indicate that the HCR still is in agreement with the precautionary approach (Mehl and Fotland, WD 11 2011).

The fishing mortality used in the harvest control rule ( $\mathrm{F}_{\mathrm{mp}}$ ) was in 2007 set to $\mathrm{F}_{\mathrm{pa}}=0.35$. In June 2013, after the ICES advice for 2014 for this stock had been given, $\mathrm{F}_{\mathrm{mp}}$ was reduced to 0.32 .

### 5.7 Predictions

### 5.7.1 Input data (Table 5.12)

The input data to the predictions based on results from the final model run are given in Table 5.12. The estimates for stock number-at-age in 2022 were taken from the final SAM run for ages $4+$. The geometric mean (GM) for recruitment (age 3) of 161659 thousand was used in 2022 and subsequent year classes. The natural mortality of 0.2 is the same as used in the assessment. For exploitation pattern the average of the 2019-2021 fishing mortalities estimated in the final SAM run for ages 3 to 12 was used, with mortalities for $8+$ being constant. For weight-at-age in stock and catch the average of the last three years (2019-2021) from SAM input file was used. For ma-turity-at-age the average of the 2005-2007 annual ogives was applied.

### 5.7.2 Catch options for 2022 (short-term predictions; Tables 5.13-14)

The management option table (Table 5.13) shows that the expected landings of 197212 t in 2022 will result in a fishing an adjusted mortality $\mathrm{F}_{\mathrm{bar}}$ of 0.207 , which is lower compared to 2021 of 0.265 , but well below the $\mathrm{F}_{\mathrm{pa}}$ of 0.35 . A catch in 2023 corresponding to the $\mathrm{F}_{\text {status quo }}$ level of 0.207 will be 189690 t , while a catch in 2023 corresponding to the evaluated and implemented HCR of 226794 t will result in F of 0.254 (Table 5.14).

For a catch in 2022 corresponding to the TAC of $197212 t$, the SSB is expected to decrease from about 745913 t at the beginning of 2022 to 686937 t at the beginning of 2023. At $\mathrm{F}_{\text {status quo }}$ in 2023 SSB is estimated to decrease to 633154 t at the beginning of 2024 and for a catch corresponding to the HCR it will decrease to about 597899 in 2024.

### 5.7.3 Comparison of the present and last year's assessment

The current assessment estimated the total stock in 2022 to be $20 \%$ higher and the SSB $26 \%$ higher compared to the previous assessment. The F in 2020 from the current assessment is higher than the F from the previous assessment, and the realized F in 2021 is lower compared to the predicted one in 2021 based on the TAC.

|  | Total stock (3+) by 1 January 2021 <br> (tonnes) | SSB by 1 January 2021 <br> (tonnes) | F4-7 in 2021 F4-7 in 2020 |  |
| :---: | :---: | :---: | :---: | :---: |
| WG 2021 | 954114 | 568972 | 0.23 | 0.22 |
| WG 2022 | 1140302 | 715674 | 0.186 | 0.187 |

### 5.8 Comments to the assessment and the forecast (Figure 5.6)

A statistical model is less sensitive to +group setting than XSA. In addition, the results from XSA were more dependent on the input data (use or no use of CPUE, split of the tuning survey timeseries), the shrinkage parameter and whether the number of iterations is capped or not. XSA only converged at a large number of iterations. In contrast, results from SAM are much more robust and depend to a lesser degree on subjective choice of model settings (such as shrinkage). In addition, SAM as a stochastic model is not treating catches as known without error. The fishing mortality rates could be considered correlated in time, and to reflect that neighbouring age groups have more similar fishing mortalities.

The retrospective pattern has been a major concern in the assessment, but due to the changes done at the benchmark assessment in 2010 (ICES CM 2010/ACOM: 36) and later at the NEA saithe IBP in 2014 (ICES CM 2014/ACOM: 53), the assessment has become stable (Figure 5.6)

The biological sampling from the fishery got critically low after the termination of the original Norwegian port-sampling program in 2009. In 2015 this was in particular the case for samples from trawl in quarter two and three in ICES area 1 and age samples from purse-seine fishery south of Lofoten (ICES area 2.a). In 2021 biological sampling from the saithe purse-seine fishery catches in Norwegian waters was adequate.

Lack of reliable recruitment estimates is a major problem. Prediction of catches will still, to a large extent, be dependent on assumptions of average recruitment in the intermediate year and the forecast period, since fish from age four to seven constitute major parts of the catches. Since the saithe HCR is a three-year-rule, the estimation of average $\mathrm{F}_{\mathrm{mp}}$ catch in the HCR will affect stock numbers up to age five, and thereby affect the total prognosis of the fishable stock and the quotas derived from it. The recruitment-at-age 3 estimated by the SAM has on average been at about the long-term geometric mean level since 2005

### 5.9 Tables and figures

Table 5.1. Saithe in subareas 1 and 2 (Northeast Arctic). Nominal catch ( $\mathbf{t}$ ) by countries as officially reported to ICES.

| Year | Faroe <br> Islands | France | Germany (Dem Rep) | Germany <br> (Fed Rep) | Iceland | Norway | Poland | Portugal | Russia ${ }^{3}$ | Spain | UK | Others ${ }^{5}$ | Total: all countries |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | 23 | 1700 |  | 25948 |  | 96050 |  |  |  |  | 9780 | 14 | 133515 |
| 1961 | 61 | 3625 |  | 19757 |  | 77875 |  |  |  |  | 4615 | 18 | 105951 |
| 1962 | 2 | 544 |  | 12651 |  | 101895 |  |  | 912 |  | 4699 | 4 | 120707 |
| 1963 |  | 1110 |  | 8108 |  | 135297 |  |  |  |  | 4112 |  | 148627 |
| 1964 |  | 1525 |  | 4420 |  | 184700 |  |  | 84 |  | 6511 | 186 | 197426 |
| 1965 |  | 1618 |  | 11387 |  | 165531 |  |  | 137 |  | 6746 | 181 | 185600 |
| 1966 |  | 2987 | 813 | 11269 |  | 175037 |  |  | 563 |  | 13078 | 41 | 203788 |
| 1967 |  | 9472 | 304 | 11822 |  | 150860 |  |  | 441 |  | 8379 | 48 | 181326 |
| 1968 |  |  | 1248 | 4753 |  | 96641 |  |  |  |  | 8782 |  | 111424 |
| 1969 | 20 | 193 | 6744 | 4355 |  | 115140 |  |  |  |  | 13585 | 23 | 140060 |
| 1970 | 1097 |  | 29200 | 23466 |  | 151759 |  |  | 43550 |  | 15690 |  | 264924 |
| 1971 | 215 | 14536 | 16840 | 12204 |  | 128499 | 6017 |  | 39397 | 13097 | 10467 |  | 241272 |
| 1972 | 109 | 14519 | 7474 | 24595 |  | 143775 | 1111 |  | 1278 | 9247 | 8348 |  | 210456 |
| 1973 | 7 | 11320 | 12015 | 30338 |  | 148789 | 23 |  | 2411 | 2115 | 6841 |  | 213859 |
| 1974 | 46 | 7119 | 29466 | 33155 |  | 152699 | 2521 |  | 28931 | 7075 | 3104 | 5 | 264121 |


| Year | Faroe <br> Islands | France | Germany (Dem Rep) | Germany <br> (Fed Rep) | Iceland | Norway | Poland | Portugal | Russia ${ }^{3}$ | Spain | UK | Others ${ }^{5}$ | Total: all countries |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 28 | 3156 | 28517 | 41260 |  | 122598 | 3860 | 6430 | 13389 | 11397 | 2763 | 55 | 233453 |
| 1976 | 20 | 5609 | 10266 | 49056 |  | 131675 | 3164 | 7233 | 9013 | 21661 | 4724 | 65 | 242486 |
| 1977 | 270 | 5658 | 7164 | 19985 |  | 139705 | 1 | 783 | 989 | 1327 | 6935 |  | 182817 |
| 1978 | 809 | 4345 | 6484 | 19190 |  | 121069 | 35 | 203 | 381 | 121 | 2827 |  | 155464 |
| 1979 | 1117 | 2601 | 2435 | 15323 |  | 141346 |  |  | 3 | 685 | 1170 |  | 164680 |
| 1980 | 532 | 1016 |  | 12511 |  | 128878 |  |  | 43 | 780 | 794 |  | 144554 |
| 1981 | 236 | 218 |  | 8431 |  | 166139 |  |  | 121 |  | 395 |  | 175540 |
| 1982 | 339 | 82 |  | 7224 |  | 159643 |  |  | 14 |  | 732 |  | 168034 |
| 1983 | 539 | 418 |  | 4933 |  | 149556 |  |  | 206 | 33 | 1251 |  | 156936 |
| 1984 | 503 | 431 | 6 | 4532 |  | 152818 |  |  | 161 |  | 335 |  | 158786 |
| 1985 | 490 | 657 | 11 | 1873 |  | 103899 |  |  | 51 |  | 202 |  | 107183 |
| 1986 | 426 | 308 |  | 3470 |  | 63090 |  |  | 27 |  | 75 |  | 67396 |
| 1987 | 712 | 576 |  | 4909 |  | 85710 |  |  | 426 |  | 57 | 1 | 92391 |
| 1988 | 441 | 411 |  | 4574 |  | 108244 |  |  | 130 |  | 442 |  | 114242 |
| 1989 | 388 | $460{ }^{2}$ |  | 606 |  | 119625 |  |  | 506 | 506 | 726 |  | 122817 |
| 1990 | 1207 | $340^{2}$ |  | 1143 |  | 92397 |  |  | 52 |  | 709 |  | 95848 |
| 1991 | 963 | $77^{2}$ | Greenland | 2003 |  | 103283 |  |  | $504{ }^{4}$ |  | 492 | 5 | 107327 |
| 1992 | 165 | 1980 | 734 | 3451 |  | 119763 |  |  | 964 | 6 | 541 |  | 127604 |


| Year | Faroe <br> Islands | France | Germany (Dem Rep) | Germany <br> (Fed Rep) | Iceland | Norway | Poland | Portugal | Russia ${ }^{3}$ | Spain | UK | Others ${ }^{5}$ | Total: all countries |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 31 | 566 | 78 | 3687 | 3 | 140604 |  | 1 | 9509 | $4^{2}$ | 415 | 5 | 154903 |
| 1994 | $67^{2}$ | 557 | 15 | 1863 | $4^{2}$ | 141589 |  | $1^{2}$ | $1640{ }^{2}$ | $655{ }^{2}$ | 557 | 2 | 146950 |
| 1995 | $172{ }^{2}$ | 358 | 53 | 935 |  | 165001 |  | 5 | 1148 |  | 688 | 18 | 168378 |
| 1996 | $248{ }^{2}$ | 346 | 165 | 2615 |  | 166045 |  | 24 | 1159 | 6 | 707 | 33 | 171348 |
| 1997 | $193{ }^{2}$ | 560 | $363^{2}$ | 2915 |  | 136927 |  | 12 | 1774 | 41 | 799 | 45 | 143629 |
| 1998 | 366 | 932 | $437{ }^{2}$ | 2936 |  | 144103 |  | 47 | 3836 | 275 | 355 | 40 | 153327 |
| 1999 | 181 | $638{ }^{2}$ | $655^{2}$ | 2473 | 146 | 141941 |  | 17 | 3929 | 24 | 339 | 32 | 150375 |
| 2000 | $224{ }^{2}$ | 1438 | $651{ }^{2}$ | 2573 | 33 | 125932 |  | 46 | 4452 | 117 | 454 | $8^{2}$ | 135928 |
| 2001 | 537 | 1279 | $701^{2}$ | 2690 | 57 | 124928 |  | 75 | 4951 | 119 | 514 | 2 | 135853 |
| 2002 | 788 | 1048 | 1393 | 2642 | 78 | 142941 |  | 118 | 5402 | 37 | 420 | 3 | 154870 |
| 2003 | 2056 | 1022 | $929{ }^{2}$ | 2763 | $80^{2}$ | 150400 |  | 147 | 3894 | 18 | 265 | $18^{2}$ | 161592 |
| 2004 | 3071 | 255 | $891^{2}$ | 2161 | 319 | 147975 |  | 127 | 9192 | 87 | 544 | 14 | 164636 |
| 2005 | 3152 | 447 | $817^{2}$ | 2048 | 395 | 162338 |  | 354 | 8362 | 25 | 630 |  | 178568 |
| 2006 | 1795 | 899.7 | $779{ }^{2}$ | 2780 | 255 | 195462 | 88.9 | 101 | 9823 | 0 | 532 | 42 | 212557 |
| 2007 | 2048 | 965.6 | $801^{2}$ | 3019 | 219 | 178644 | 99.3 | 412 | 12168 | 22 | 557 | 11.8 | 198967 |
| 2008 | 2405 | 1008.6 | $513{ }^{2}$ | 2264 | 113 | 165998 | 65.8 | 348 | 11577 | 33 | 506 | 9.7 | 184840 |
| 2009 | 1611 | 378.6 | 697 | 2021 | 69 | 144570 | 30.6 | 184.01 | 11899 | 2 | 379 | 24 | 161865 |
| 2010 | 1632 | 677.2 | 954 | 1592 | 124 | 175246 | 278.9 | 93 | 14664 | 8 | 283 | 2.5 | 195554 |


| Year | Faroe <br> Islands | France | Germany <br> (Dem Rep) | Germany <br> (Fed Rep) | Iceland | Norway | Poland | Portugal | Russia ${ }^{3}$ | Spain | UK | Others ${ }^{5}$ | Total: all countries |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 306 | 504.2 | 445 | 1371 | 66 | 143314 | 0 | 45.34 | 10007 | 2 | 972 | 15.14 | 157048 |
| 2012 | 146 | 780.55 | 658 | 1371 | 126 | 143174 | 0 | 7.65 | 13607 | 4 | 1087 | 0 | 160960 |
| 2013 | 80 | 1900.92 | 972 | 1212 | 245 | 111961 | 2.21 | 17.24 | 14796 | 5 | 415 | 21.93 | 131629 |
| 2014 | 273 | 1674 | 407 | 259 | 659 | 115864 | 0.86 | 8.25 | 12396 | 12 | 518 | 0 | 132070 |
| 2015 | 766 | 515 | 393 | 424 | 248 | 115157 | 1143 | 10.42 | 13181 | 34 | 403 | 0 | 132275 |
| 2016 | 1148 | 526 | 613 | 952 | 702 | 121705 | 530 | 52 | 15203 | 26 | 301 | 10 | 141768 |
| $2017{ }^{1}$ | 639 | 680 | 407 | 865 | 589 | 126947 | 504 | 86 | 14551 | 88 | 439 | 24 | 145819 |
| 2018 | 626 | 937 | 448 | 1642 |  | 162460 | 404 | 51 | 14171 | 60 | 464 | 17 | 181280 |
| 2019 | 618 | 1472 | 424 | 1371 |  | 144076 | 46 | 131 | 13990 | 199 | 419 | 434 | 163180 |
| 2020 |  | 530 | 410 | 1544 |  | 151697 | 1.2 | 132 | 14082 | 0 | 517 | 118 | 169405 |
| 2021 | 573 | 684 | 449 | 600 | 148 | 171836 | 0.3 | 21 | 13836 | 3 | 2 | 23 | 188176 |

1 Provisional figures.
2 As reported to Norwegian authorities.
3 USSR prior to 1991.
4 Includes Estonia.
5 Includes Denmark. Netherlands. Ireland. and Sweden.
6 As reported by Working Group member.

Table 5.2 Saithe in subareas 1 and 2 (Northeast Arctic). Catch ('000) by fishing gear.

| Year | Purse-seine | Trawl | Gillnet | Others | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 75.2 | 69.5 | 19.3 | 12.7 | 176.7 |
| 1978 | 62.9 | 57.6 | 21.1 | 13.9 | 155.5 |
| 1979 | 74.7 | 52.5 | 21.6 | 15.9 | 164.7 |
| 1980 | 61.3 | 46.8 | 21.1 | 15.4 | 144.6 |
| 1981 | 64.3 | 72.4 | 24.0 | 14.8 | 175.5 |
| 1982 | 76.4 | 59.4 | 16.7 | 15.5 | 168.0 |
| 1983 | 54.1 | 68.2 | 19.6 | 15.0 | 156.9 |
| 1984 | 36.4 | 85.6 | 23.7 | 13.1 | 158.8 |
| 1985 | 31.1 | 49.9 | 14.6 | 11.6 | 107.2 |
| 1986 | 7.9 | 36.2 | 12.3 | 8.2 | 64.6 |
| 1987 | 34.9 | 27.7 | 19.0 | 10.8 | 92.4 |
| 1988 | 43.5 | 45.4 | 15.3 | 10.0 | 114.2 |
| 1989 | 49.5 | 45.0 | 16.9 | 11.4 | 122.8 |
| 1990 | 24.6 | 44.0 | 19.3 | 7.9 | 95.8 |
| 1991 | 38.9 | 40.1 | 18.9 | 9.4 | 107.3 |
| 1992 | 27.1 | 67.0 | 22.3 | 11.2 | 127.6 |
| 1993 | 33.1 | 84.9 | 21.2 | 15.7 | 154.9 |
| 1994 | 30.2 | 82.2 | 21.1 | 13.5 | 147.0 |
| 1995 | 21.8 | 103.5 | 26.9 | 16.1 | 168.4 |
| 1996 | 46.9 | 72.5 | 31.6 | 20.3 | 171.3 |
| 1997 | 44.4 | 55.9 | 24.4 | 19.0 | 143.6 |
| 1998 | 44.4 | 57.7 | 27.6 | 23.6 | 153.3 |
| 1999 | 39.2 | 57.9 | 29.7 | 23.6 | 150.4 |
| 2000 | 28.3 | 54.5 | 29.6 | 23.5 | 135.9 |
| 2001 | 28.1 | 58.1 | 28.2 | 21.5 | 135.9 |
| 2002 | 27.4 | 75.5 | 30.4 | 21.5 | 154.8 |
| 2003 | 43.3 | 73.8 | 25.2 | 19.3 | 161.6 |
| 2004 | 41.8 | 74.6 | 26.9 | 21.3 | 164.6 |
| 2005 | 42.1 | 91.8 | 25.6 | 19.1 | 178.6 |


| Year | Purse-seine | Trawl | Gillnet | Others | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 73.5 | 87.1 | 29.7 | 22.5 | 212.8 |
| 2007 | 41.8 | 100.7 | 33.3 | 23.2 | 199.0 |
| 2008 | 39.4 | 91.2 | 37.0 | 17.1 | 184.7 |
| 2009 | 35.5 | 81.1 | 33.2 | 12.1 | 161.9 |
| 2010 | 54.9 | 89.8 | 36.9 | 13.2 | 194.8 |
| 2011 | 45.3 | 67.1 | 32.1 | 12.2 | 156.7 |
| 2012 | 44.2 | 73.9 | 28.3 | 14.5 | 160.9 |
| 2013 | 34.7 | 65.2 | 19.2 | 12.7 | 131.8 |
| 2014 | 29.3 | 54.8 | 26.7 | 21.2 | 132.0 |
| 2015 | 30.4 | 55.4 | 23.5 | 22.5 | 131.8 |
| 2016 | 28.9 | 64.1 | 21.4 | 26.9 | 141.3 |
| $2017{ }^{1}$ | 32.4 | 65.0 | 21.4 | 27.3 | 146.1 |
| 2018 | 36.0 | 83.6 | 28.8 | 33.2 | 181.5 |
| 2019 | 28.7 | 68.6 | 29.4 | 36.6 | 163.1 |
| 2020 | 26.8 | 74 | 30.3 | 38.3 | 169.4 |
| 2021 | 30.9 | 81.6 | 29.5 | 46 | 188 |

${ }^{1}$ Provisional figures.
${ }^{2}$ Unresolved discrepancies between Norwegian catch by gear figures and the total reported to ICES for these years.
${ }^{3}$ Includes 4300 tonnes not categorized by gear. proportionally adjusted.
${ }^{4}$ Reduced by 1200 tonnes not categorized by gear. proportionally adjusted.

Table 5.3 Catch numbers-at-age ('000) of northeast Arctic saithe.

| Age groups |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| 1960 | 13517 | 16828 | 17422 | 6514 | 6281 | 3088 | 1691 | 956 | 481 | 1481 |
| 1961 | 25237 | 12929 | 17707 | 5379 | 1886 | 1371 | 736 | 573 | 538 | 1202 |
| 1962 | 45932 | 13720 | 5449 | 10218 | 2991 | 1262 | 1156 | 556 | 611 | 1518 |
| 1963 | 51171 | 35199 | 7165 | 5659 | 4699 | 1337 | 1308 | 848 | 550 | 1612 |
| 1964 | 10925 | 72344 | 15966 | 3299 | 4214 | 3223 | 1518 | 1482 | 1282 | 3038 |
| 1965 | 42578 | 5737 | 30171 | 11635 | 3282 | 2421 | 3135 | 802 | 1136 | 2986 |
| 1966 | 25127 | 61199 | 14727 | 14475 | 5220 | 1542 | 1047 | 1083 | 530 | 2724 |


| Age groups |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| 1967 | 28457 | 23826 | 34493 | 3957 | 5388 | 2797 | 1356 | 1340 | 814 | 2536 |
| 1968 | 29955 | 21856 | 6065 | 9846 | 936 | 2274 | 1070 | 686 | 465 | 922 |
| 1969 | 76011 | 11745 | 16650 | 4666 | 4716 | 1107 | 1682 | 663 | 199 | 303 |
| 1970 | 43834 | 63270 | 14081 | 16298 | 5157 | 8004 | 2521 | 3722 | 1103 | 1714 |
| 1971 | 61743 | 47522 | 21614 | 7661 | 7690 | 2326 | 3489 | 1760 | 2514 | 1888 |
| 1972 | 55351 | 44490 | 24752 | 8650 | 4769 | 3012 | 1584 | 1817 | 1044 | 1631 |
| 1973 | 62938 | 20793 | 22199 | 13224 | 5868 | 3246 | 2368 | 2153 | 1291 | 1947 |
| 1974 | 36884 | 44149 | 15714 | 20476 | 12182 | 4815 | 3267 | 2512 | 1440 | 2392 |
| 1975 | 70255 | 13502 | 18901 | 5123 | 9018 | 7841 | 3365 | 2714 | 2237 | 2544 |
| 1976 | 135592 | 33159 | 8618 | 9448 | 3725 | 3483 | 2905 | 1870 | 1183 | 1940 |
| 1977 | 105935 | 36703 | 10845 | 2205 | 4633 | 1557 | 1718 | 1030 | 495 | 718 |
| 1978 | 56505 | 31946 | 14396 | 5232 | 1694 | 2132 | 1082 | 1126 | 756 | 1726 |
| 1979 | 75819 | 28545 | 17280 | 5384 | 3550 | 1178 | 1659 | 536 | 373 | 1086 |
| 1980 | 40303 | 36202 | 9100 | 6302 | 3161 | 1322 | 145 | 721 | 406 | 1204 |
| 1981 | 85966 | 22345 | 22044 | 3706 | 2611 | 2056 | 378 | 286 | 258 | 385 |
| 1982 | 35853 | 67150 | 13481 | 8477 | 1088 | 1291 | 476 | 271 | 124 | 338 |
| 1983 | 18216 | 25108 | 34543 | 3408 | 3178 | 1243 | 803 | 261 | 215 | 587 |
| 1984 | 43579 | 34927 | 12679 | 11775 | 1193 | 1862 | 589 | 585 | 407 | 537 |
| 1985 | 48989 | 11992 | 7200 | 5287 | 3746 | 776 | 879 | 134 | 274 | 427 |
| 1986 | 21322 | 12433 | 5845 | 4363 | 2704 | 1349 | 338 | 438 | 123 | 152 |
| 1987 | 18555 | 51742 | 4506 | 3238 | 3624 | 784 | 644 | 267 | 263 | 565 |
| 1988 | 8144 | 35928 | 32901 | 4570 | 2333 | 1222 | 968 | 321 | 73 | 30 |
| 1989 | 12607 | 19400 | 33343 | 18578 | 1762 | 352 | 177 | 189 | 1 | 205 |
| 1990 | 23792 | 16930 | 9054 | 10238 | 7341 | 1076 | 160 | 112 | 150 | 118 |
| 1991 | 68682 | 13630 | 5752 | 4883 | 3877 | 2381 | 383 | 61 | 90 | 89 |
| 1992 | 44627 | 33294 | 5987 | 5412 | 4751 | 3176 | 1462 | 286 | 93 | 350 |
| 1993 | 22812 | 61931 | 31102 | 3747 | 1759 | 1378 | 1027 | 797 | 76 | 71 |
| 1994 | 7063 | 32671 | 49410 | 19058 | 2058 | 724 | 421 | 278 | 528 | 129 |


| Age groups |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| 1995 | 17178 | 52109 | 40145 | 30451 | 4177 | 483 | 125 | 259 | 31 | 263 |
| 1996 | 10510 | 54886 | 18499 | 18357 | 17834 | 2849 | 485 | 214 | 148 | 325 |
| 1997 | 11789 | 11698 | 35011 | 13567 | 13452 | 7058 | 812 | 55 | 48 | 98 |
| 1998 | 3091 | 16215 | 11946 | 31818 | 8376 | 5539 | 2873 | 727 | 111 | 282 |
| 1999 | 9655 | 12236 | 22872 | 10347 | 18930 | 3374 | 3343 | 2290 | 419 | 170 |
| 2000 | 9175 | 22768 | 7747 | 10676 | 6123 | 8303 | 2530 | 2652 | 1022 | 197 |
| 2001 | 3816 | 7946 | 26960 | 8769 | 7120 | 3146 | 4687 | 1935 | 1406 | 528 |
| 2002 | 6582 | 17492 | 11573 | 25671 | 5312 | 4276 | 2382 | 3431 | 965 | 1420 |
| 2003 | 2345 | 50653 | 13600 | 7123 | 9594 | 5494 | 3545 | 2519 | 2327 | 1813 |
| 2004 | 1002 | 6129 | 33840 | 10613 | 7494 | 8307 | 2792 | 3088 | 2377 | 3072 |
| 2005 | 26093 | 12543 | 9841 | 23141 | 10799 | 5659 | 7852 | 2674 | 713 | 1588 |
| 2006 | 1590 | 68137 | 12328 | 10098 | 16757 | 8080 | 5671 | 5127 | 1815 | 2529 |
| 2007 | 3144 | 4115 | 39889 | 15301 | 7963 | 11302 | 7749 | 4138 | 2157 | 849 |
| 2008 | 25259 | 18953 | 5969 | 24363 | 9712 | 5624 | 7697 | 4705 | 1606 | 1572 |
| 2009 | 9050 | 34311 | 9954 | 6628 | 15930 | 4766 | 3021 | 4224 | 2471 | 1426 |
| 2010 | 26382 | 43436 | 28514 | 7988 | 3129 | 12444 | 2749 | 1314 | 1212 | 1431 |
| 2011 | 6239 | 45213 | 13307 | 15157 | 6622 | 2901 | 5934 | 1730 | 647 | 1115 |
| 2012 | 30742 | 17841 | 33911 | 10496 | 7058 | 3522 | 1570 | 2586 | 557 | 890 |
| 2013 | 17151 | 15491 | 15946 | 21980 | 5512 | 3298 | 1149 | 729 | 885 | 653 |
| 2014 | 7650 | 24769 | 13822 | 9343 | 12331 | 3284 | 2130 | 904 | 378 | 763 |
| 2015 | 13185 | 15459 | 30159 | 9271 | 7324 | 7133 | 1697 | 723 | 433 | 620 |
| 2016 | 8278 | 20955 | 13044 | 15532 | 6621 | 4774 | 4363 | 1053 | 718 | 1382 |
| 2017 | 5421 | 34736 | 12901 | 7324 | 9032 | 3885 | 2562 | 1924 | 376 | 1999 |
| 2018 | 5260 | 19260 | 41425 | 12618 | 5903 | 5667 | 2843 | 1956 | 1112 | 1567 |
| 2019 | 12421 | 15078 | 15388 | 25177 | 8327 | 3243 | 2848 | 1357 | 619 | 1171 |
| 2020 | 6216 | 27602 | 13466 | 14054 | 17767 | 5031 | 2034 | 1469 | 564 | 1236 |
| 2021 | 5732 | 7938 | 26311 | 12418 | 11357 | 12295 | 3544 | 1580 | 954 | 1939 |

Table 5.4 Catch weight-at-age (kg) northeast Arctic saithe.

| Age groups |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| 1960 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.55 |
| 1961 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.75 |
| 1962 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.52 |
| 1963 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.33 |
| 1964 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.35 |
| 1965 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.54 |
| 1966 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.43 |
| 1967 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.49 |
| 1968 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.36 |
| 1969 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.16 |
| 1970 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.03 |
| 1971 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 7.87 |
| 1972 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.14 |
| 1973 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.01 |
| 1974 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 7.69 |
| 1975 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 7.73 |
| 1976 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 7.86 |
| 1977 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.05 |
| 1978 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.00 |
| 1979 | 0.71 | 1.11 | 1.63 | 2.33 | 3.16 | 4.03 | 4.87 | 5.63 | 6.44 | 8.28 |
| 1980 | 0.79 | 1.27 | 2.03 | 2.55 | 3.29 | 4.34 | 5.15 | 5.75 | 6.11 | 7.22 |
| 1981 | 0.73 | 1.40 | 2.05 | 2.76 | 3.30 | 4.38 | 5.95 | 6.39 | 6.61 | 7.00 |
| 1982 | 0.77 | 1.12 | 2.02 | 2.61 | 3.27 | 3.91 | 4.69 | 5.63 | 7.18 | 7.69 |
| 1983 | 1.05 | 1.33 | 1.86 | 2.80 | 4.00 | 4.18 | 5.33 | 5.68 | 7.31 | 9.16 |
| 1984 | 0.71 | 1.26 | 2.02 | 2.70 | 3.88 | 4.47 | 5.36 | 6.06 | 6.28 | 7.88 |
| 1985 | 0.75 | 1.33 | 2.07 | 2.63 | 3.28 | 3.96 | 4.54 | 5.55 | 6.88 | 8.74 |
| 1986 | 0.59 | 1.22 | 1.97 | 2.30 | 2.87 | 3.72 | 4.30 | 4.69 | 5.84 | 7.21 |
| 1987 | 0.53 | 0.84 | 1.66 | 2.32 | 2.97 | 4.00 | 4.72 | 5.44 | 5.79 | 7.42 |


| Year | Age groups |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| 1988 | 0.62 | 0.87 | 1.31 | 2.43 | 3.87 | 5.38 | 5.83 | 5.36 | 6.92 | 8.82 |
| 1989 | 0.74 | 0.95 | 1.40 | 1.78 | 2.96 | 3.73 | 4.62 | 4.66 | 8.34 | 7.69 |
| 1990 | 0.71 | 1.00 | 1.45 | 2.09 | 2.49 | 3.75 | 3.90 | 6.74 | 4.94 | 7.34 |
| 1991 | 0.68 | 1.05 | 1.85 | 2.39 | 3.08 | 3.35 | 4.48 | 4.66 | 5.62 | 7.31 |
| 1992 | 0.67 | 1.01 | 1.92 | 2.28 | 2.77 | 3.20 | 3.73 | 6.35 | 6.90 | 7.83 |
| 1993 | 0.61 | 0.99 | 1.65 | 2.46 | 2.85 | 3.03 | 3.71 | 4.49 | 5.56 | 7.13 |
| 1994 | 0.52 | 0.76 | 1.24 | 2.12 | 3.22 | 3.83 | 4.69 | 5.31 | 5.66 | 7.29 |
| 1995 | 0.56 | 0.79 | 1.19 | 1.71 | 2.87 | 3.78 | 4.06 | 5.30 | 6.86 | 7.65 |
| 1996 | 0.59 | 0.82 | 1.33 | 1.84 | 2.48 | 3.73 | 4.32 | 5.34 | 5.98 | 7.58 |
| 1997 | 0.62 | 0.95 | 1.24 | 1.72 | 2.35 | 3.10 | 4.19 | 5.79 | 6.77 | 7.75 |
| 1998 | 0.68 | 1.00 | 1.48 | 1.87 | 2.58 | 3.07 | 4.13 | 5.44 | 6.70 | 8.59 |
| 1999 | 0.67 | 1.05 | 1.45 | 1.93 | 2.27 | 2.97 | 3.61 | 4.10 | 4.93 | 6.97 |
| 2000 | 0.60 | 1.03 | 1.63 | 2.10 | 2.67 | 3.14 | 3.81 | 4.41 | 5.76 | 8.07 |
| 2001 | 0.75 | 1.12 | 1.54 | 2.04 | 2.60 | 3.14 | 3.63 | 4.54 | 5.05 | 6.17 |
| 2002 | 0.69 | 1.01 | 1.50 | 1.97 | 2.54 | 3.25 | 3.77 | 4.31 | 4.91 | 6.11 |
| 2003 | 0.66 | 0.91 | 1.42 | 1.89 | 2.54 | 2.58 | 3.49 | 3.75 | 4.12 | 5.90 |
| 2004 | 0.70 | 1.03 | 1.37 | 1.90 | 2.41 | 2.98 | 3.44 | 3.73 | 4.14 | 5.47 |
| 2005 | 0.59 | 0.89 | 1.49 | 2.09 | 2.16 | 2.99 | 3.24 | 3.82 | 3.92 | 6.19 |
| 2006 | 0.63 | 0.83 | 1.43 | 1.78 | 2.27 | 2.73 | 3.02 | 3.90 | 4.06 | 5.82 |
| 2007 | 0.73 | 1.08 | 1.41 | 1.86 | 2.43 | 2.94 | 3.35 | 3.66 | 4.17 | 5.54 |
| 2008 | 0.63 | 0.98 | 1.38 | 1.92 | 2.31 | 2.83 | 3.16 | 3.43 | 3.82 | 4.75 |
| 2009 | 0.73 | 1.03 | 1.65 | 2.00 | 2.37 | 2.69 | 3.23 | 3.38 | 3.46 | 4.67 |
| 2010 | 0.70 | 0.99 | 1.45 | 2.14 | 2.50 | 3.13 | 3.34 | 3.81 | 3.99 | 5.17 |
| 2011 | 0.70 | 0.82 | 1.42 | 2.07 | 2.68 | 3.25 | 3.62 | 3.97 | 4.52 | 5.84 |
| 2012 | 0.59 | 1.07 | 1.35 | 2.15 | 2.82 | 3.20 | 3.67 | 4.16 | 4.60 | 5.70 |
| 2013 | 0.57 | 1.01 | 1.50 | 1.83 | 2.74 | 3.33 | 3.91 | 4.61 | 4.50 | 6.13 |
| 2014 | 0.66 | 0.92 | 1.58 | 2.12 | 2.54 | 3.49 | 4.01 | 4.22 | 4.71 | 5.80 |
| 2015 | 0.61 | 0.85 | 1.24 | 1.91 | 2.45 | 3.02 | 3.97 | 4.74 | 4.51 | 6.05 |


| Age groups |  | $\mathbf{5}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | $\mathbf{3}$ | $\mathbf{4}$ | 1.04 | 1.46 | 2.02 | 2.36 | 3.12 | 3.53 | 4.14 | 4.65 |
| 2016 | 0.84 | 1.12 | 1.68 | 2.18 | 2.63 | 3.13 | 3.63 | 4.16 | 4.5 | 5.9 |
| 2017 | 0.89 | 1.12 |  |  |  |  |  |  |  |  |
| 2018 | 0.91 | 1.21 | 1.56 | 2.02 | 2.51 | 3.04 | 3.44 | 3.89 | 4.50 | 5.60 |
| 2019 | 0.83 | 1.17 | 1.64 | 2.06 | 2.62 | 3.18 | 3.71 | 4.13 | 4.88 | 6.14 |
| 2020 | 0.74 | 1.06 | 1.57 | 2.01 | 2.53 | 3.13 | 3.75 | 4.36 | 5.05 | 6.80 |
| 2021 | 0.77 | 1.16 | 1.61 | 2.14 | 2.68 | 3.15 | 3.65 | 4.14 | 4.7 | 6.3 |

Table 5.5. 3-year running average maturity ogive 1985-2006. Values for 2007-2020 average of 2005-2007.

| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 0 | 0.02 | 0.5 | 0.92 | 0.99 | 1 | 1 | 1 | 1 | 1 |
| 1986 | 0 | 0.02 | 0.51 | 0.94 | 0.99 | 1 | 1 | 1 | 1 | 1 |
| 1987 | 0 | 0 | 0.35 | 0.98 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1988 | 0 | 0 | 0.25 | 0.96 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1989 | 0 | 0 | 0.15 | 0.92 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1990 | 0 | 0 | 0.2 | 0.85 | 0.99 | 1 | 1 | 1 | 1 | 1 |
| 1991 | 0 | 0.02 | 0.25 | 0.84 | 0.98 | 1 | 1 | 1 | 1 | 1 |
| 1992 | 0 | 0.02 | 0.3 | 0.83 | 0.93 | 0.92 | 0.9 | 0.95 | 1 | 1 |
| 1993 | 0 | 0.02 | 0.26 | 0.88 | 0.92 | 0.89 | 0.87 | 0.89 | 1 | 0.99 |
| 1994 | 0 | 0.02 | 0.26 | 0.84 | 0.9 | 0.82 | 0.87 | 0.89 | 1 | 0.99 |
| 1995 | 0 | 0.02 | 0.22 | 0.8 | 0.92 | 0.9 | 0.97 | 0.94 | 1 | 0.99 |
| 1996 | 0 | 0.03 | 0.21 | 0.65 | 0.91 | 0.93 | 1 | 1 | 1 | 1.00 |
| 1997 | 0 | 0.03 | 0.14 | 0.45 | 0.83 | 0.94 | 0.93 | 0.97 | 1 | 1.00 |
| 1998 | 0 | 0.04 | 0.07 | 0.33 | 0.74 | 0.93 | 0.92 | 0.96 | 1 | 1.00 |
| 1999 | 0 | 0 | 0.08 | 0.32 | 0.74 | 0.92 | 0.92 | 0.96 | 0.99 | 0.98 |
| 2000 | 0 | 0 | 0.08 | 0.46 | 0.82 | 0.96 | 0.98 | 0.99 | 0.97 | 0.95 |
| 2001 | 0 | 0 | 0.11 | 0.64 | 0.93 | 0.97 | 0.98 | 0.99 | 0.97 | 0.94 |
| 2002 | 0 | 0 | 0.13 | 0.78 | 0.95 | 0.98 | 0.98 | 0.99 | 0.98 | 0.97 |
| 2003 | 0 | 0 | 0.14 | 0.82 | 0.96 | 0.98 | 0.98 | 0.99 | 1 | 0.99 |
| 2004 | 0 | 0 | 0.21 | 0.8 | 0.97 | 0.99 | 0.99 | 1 | 1 | 0.98 |


| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 0 | 0.03 | 0.3 | 0.82 | 0.97 | 0.99 | 0.99 | 1 | 1 | 1.00 |
| 2006 | 0 | 0.04 | 0.4 | 0.86 | 0.98 | 0.99 | 1 | 1 | 1 | 1.00 |
| 2007 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 0.99 |
| 2008 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 0.99 |
| 2009 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 0.99 |
| 2010 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 0.99 |
| 2011 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 1.00 |
| 2012 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 1.00 |
| 2013 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 1.00 |
| 2014 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 1.00 |
| 2015 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 1.00 |
| 2016 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 1.00 |
| 2017 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 1.00 |
| 2018 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 1.00 |
| 2019 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 1.00 |
| 2020 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 1.00 |
| 2021 | 0 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 0.98 | 0.97 | 0.97 | 1.00 |

## Table 5.6 Northeast Arctic saithe. Tuning datasets applied in final SAM run

```
North-East Arctic saithe (Sub-areas I and II)
102
FLT13: Norway Ac Survey (Catch: Unknown) (Effort: Unknown)
19942001
1 1 0.75 0.85
3 7
\begin{tabular}{rrrrrr}
1 & 87.1 & 108.9 & 41.4 & 8.1 & 0.7 \\
1 & 166.1 & 86.5 & 46.5 & 16.5 & 2.4 \\
1 & 122.6 & 207.4 & 31.7 & 15.1 & 4.0 \\
1 & 38.0 & 184.8 & 79.8 & 50.6 & 9.6 \\
1 & 96.7 & 202.6 & 69.3 & 84.3 & 6.6 \\
1 & 233.8 & 72.9 & 62.2 & 21.0 & 19.2 \\
1 & 142.5 & 176.3 & 11.6 & 11.5 & 8.0 \\
1 & 275.9 & 45.9 & 53.8 & 5.6 & 6.1
\end{tabular}
FLT14: Norway Ac Survey (Catch: Unknown) (Effort: Unknown)
20022021
1 1 0.75 0.85
3
\begin{tabular}{ccccc}
230.2 & 92.6 & 18.9 & 10.6 & 2.2 \\
87.5 & 151.7 & 26.1 & 6.2 & 6.4 \\
191.2 & 107.6 & 44.3 & 15.2 & 4.25 \\
198.5 & 51.9 & 17.6 & 13.2 & 7.68 \\
40.9 & 129.9 & 14.4 & 4.62 & 9.49 \\
93.5 & 23.9 & 58.5 & 6.51 & 3.95 \\
55.9 & 15.9 & 7.84 & 9.99 & 3.06 \\
96.9 & 61.4 & 6.99 & 4.01 & 7.62 \\
143.0 & 22.5 & 17.1 & 3.95 & 1.68 \\
42.7 & 59.6 & 4.61 & 4.23 & 1.07 \\
69 & 29.7 & 18.8 & 3.48 & 2.83 \\
77.1 & 16.5 & 13.3 & 11.6 & 2.19 \\
40.1 & 70.8 & 8.73 & 5.6 & 5.44 \\
72.4 & 22.7 & 30.1 & 6.08 & 4.22
\end{tabular}
            145.7 32.0 10.5 11.2 4.15
            91.1 63.9 13.3 2.76 5.35
            30.6 61.1 45.4 12.3 4.2
                84.4 50.6 1.4 24.2 
                64.9 33.6 59.3 15.3 8.3
```


## Table 5.7 SAM parameter settings

Model used: State-space assessment model SAM (https://www.stockassessment.org).
Software used: Template Model Builder (TMB) and R.
Visible stock on (https://www.stockassessment.org) "afwg_saithe_2018_001".
Model Options agreed upon at IBP saithe winter 2014.
\$minAge
\# The minimium age class in the assessment
3
\$maxAge
\# The maximum age class in the assessment
12
\$maxAgePlusGroup
\# Is last age group considered a plus group (1 yes, or 0 no).

1
\$keyLogFsta
\# Coupling of the fishing mortality states (nomally only first row is used).
$\begin{array}{llllllllll}0 & 1 & 2 & 3 & 4 & 5 & 5 & 5 & 5 & 5\end{array}$
$\begin{array}{cccccccccc}-1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{cccccccccc}-1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
\$corFlag
\# Correlation of fishing mortality across ages ( 0 independent, 1 compound symmetry, or 2 AR(1)
2
\$keyLogFpar
\# Coupling of the survey catchability parameters (nomally first row is not used, as that is covered by fishing mortality).
$\begin{array}{llllllllll}-1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{llllllllll}0 & 1 & 2 & 3 & 3 & -1 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{llllllllll}4 & 5 & 6 & 7 & 7 & -1 & -1 & -1 & -1 & -1\end{array}$
\$keyQpow
\# Density dependent catchability power parameters (if any).
$\begin{array}{cccccccccc}-1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{ccccccccc}-1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}-1$
$\begin{array}{ccccccccc}-1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}-1$
\$keyVarF
\# Coupling of process variance parameters for $\log (\mathrm{F})$-process (nomally only first row is used)
$\begin{array}{llllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$
$\begin{array}{cccccccccc}-1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{cccccccccc}-1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
\$keyVarLogN
\# Coupling of process variance parameters for $\log (\mathrm{N})$-process
0111111111

## \$keyVarObs

\# Coupling of the variance parameters for the observations.
$\begin{array}{llllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$
$\begin{array}{llllllllll}1 & 1 & 1 & 1 & 1 & -1 & -1 & -1 & -1 & -1\end{array}$
$222222-1-1-1-1-1$

## Table 5.7 SAM parameter settings continued

## \$obsCorStruct

\# Covariance structure for each fleet ("ID" independent, "AR" $\mathrm{AR}(1)$, or "US" for unstructured). I Possible values are: "ID" "AR" "US"
"ID" "ID" "ID"
\$keyCorObs
\# Coupling of correlation parameters can only be specified if the $\operatorname{AR}(1)$ structure is chosen above.
\# NA's indicate where correlation parameters can be specified (-1 where they cannot).
\#3-4 4-5 5-6 6-7 7-8 8-9 9-10 10-11 11-12
NA NA NA NA NA NA NA NA NA
NA NA NA NA -1 -1
NA NA NA NA -1 -1 -1 -1 -1
\$stockRecruitmentModelCode
\# Stock recruitment code ( 0 for plain random walk, 1 for Ricker, and 2 for Beverton-Holt).

2
\$noScaledYears
\# Number of years where catch scaling is applied.
0
\$keyScaledYears
\# A vector of the years where catch scaling is applied.
\$keyParScaledYA
\# A matrix specifying the couplings of scale parameters (nrow = no scaled years, ncols = no ages).
\$fbarRange
\# lowest and higest age included in Fbar
47
\$keyBiomassTreat
\# To be defined only if a biomass survey is used ( 0 SSB index, 1 catch index, and 2 FSB index).
$-1-1-1$

Table 5.8 SAM catchabilities, negative log likelihood values and number of parameters.

| Index | Fleet number | Age | Catchability | Low | High |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 2 | 3 | 0.87 | 0.592 | 1.279 |
| 2 | 2 | 4 | 1.171 | 0.798 | 1.718 |
| 3 | 2 | 5 | 0.606 | 0.413 | 0.89 |
| 4 | 2 | 6 | 0.374 | 0.278 | 0.504 |
| 5 | 3 | 7 | 0.374 | 0.278 | 0.504 |
| 7 | 3 | 4 | 0.585 | 0.48 | 0.713 |
| 8 | 3 | 5 | 0 | 0.486 | 0.232 |
| 9 | 3 | 7 | 0.184 | 0.155 | 0.392 |
| 10 | 3 | 7 | 0.155 | 0.22 |  |

Model fitting.

| Model | $\log (\mathrm{L})$ | \#par | AIC |
| :--- | :--- | :---: | :---: |
| Current | -567.30 | 17 | 1168.61 |
| base | -560.41 | 17 | 1154.81 |

Table 5.9 Estimated fishing mortalities.

| Year <br> Age | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | 10 | 11 | 12 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1960 | 0.236 | 0.284 | 0.321 | 0.278 | 0.222 | 0.164 | 0.164 | 0.164 | 0.164 | 0.164 |
| 1961 | 0.222 | 0.260 | 0.273 | 0.226 | 0.173 | 0.126 | 0.126 | 0.126 | 0.126 | 0.126 |
| 1962 | 0.222 | 0.261 | 0.267 | 0.225 | 0.177 | 0.132 | 0.132 | 0.132 | 0.132 | 0.132 |
| 1963 | 0.224 | 0.272 | 0.281 | 0.238 | 0.194 | 0.153 | 0.153 | 0.153 | 0.153 | 0.153 |
| 1964 | 0.237 | 0.298 | 0.318 | 0.277 | 0.240 | 0.207 | 0.207 | 0.207 | 0.207 | 0.207 |
| 1965 | 0.234 | 0.291 | 0.325 | 0.288 | 0.253 | 0.230 | 0.230 | 0.230 | 0.230 | 0.230 |
| 1966 | 0.260 | 0.320 | 0.344 | 0.289 | 0.244 | 0.223 | 0.223 | 0.223 | 0.223 | 0.223 |
| 1967 | 0.261 | 0.310 | 0.319 | 0.265 | 0.225 | 0.216 | 0.216 | 0.216 | 0.216 | 0.216 |
| 1968 | 0.222 | 0.241 | 0.230 | 0.185 | 0.152 | 0.147 | 0.147 | 0.147 | 0.147 | 0.147 |
| 1969 | 0.231 | 0.241 | 0.222 | 0.175 | 0.143 | 0.131 | 0.131 | 0.131 | 0.131 | 0.131 |
| 1970 | 0.330 | 0.362 | 0.342 | 0.285 | 0.251 | 0.240 | 0.240 | 0.240 | 0.240 | 0.240 |


| $\begin{aligned} & \text { Year } \\ & \text { Age } \end{aligned}$ | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 0.360 | 0.385 | 0.357 | 0.295 | 0.270 | 0.259 | 0.259 | 0.259 | 0.259 | 0.259 |
| 1972 | 0.382 | 0.391 | 0.351 | 0.283 | 0.259 | 0.244 | 0.244 | 0.244 | 0.244 | 0.244 |
| 1973 | 0.421 | 0.428 | 0.386 | 0.317 | 0.299 | 0.284 | 0.284 | 0.284 | 0.284 | 0.284 |
| 1974 | 0.544 | 0.561 | 0.513 | 0.429 | 0.417 | 0.395 | 0.395 | 0.395 | 0.395 | 0.395 |
| 1975 | 0.597 | 0.620 | 0.567 | 0.478 | 0.489 | 0.478 | 0.478 | 0.478 | 0.478 | 0.478 |
| 1976 | 0.653 | 0.683 | 0.612 | 0.499 | 0.497 | 0.471 | 0.471 | 0.471 | 0.471 | 0.471 |
| 1977 | 0.578 | 0.614 | 0.541 | 0.430 | 0.417 | 0.378 | 0.378 | 0.378 | 0.378 | 0.378 |
| 1978 | 0.575 | 0.651 | 0.597 | 0.488 | 0.476 | 0.432 | 0.432 | 0.432 | 0.432 | 0.432 |
| 1979 | 0.554 | 0.677 | 0.639 | 0.529 | 0.509 | 0.454 | 0.454 | 0.454 | 0.454 | 0.454 |
| 1980 | 0.493 | 0.637 | 0.620 | 0.519 | 0.481 | 0.422 | 0.422 | 0.422 | 0.422 | 0.422 |
| 1981 | 0.456 | 0.629 | 0.622 | 0.521 | 0.460 | 0.393 | 0.393 | 0.393 | 0.393 | 0.393 |
| 1982 | 0.422 | 0.620 | 0.623 | 0.527 | 0.449 | 0.375 | 0.375 | 0.375 | 0.375 | 0.375 |
| 1983 | 0.402 | 0.629 | 0.655 | 0.595 | 0.531 | 0.453 | 0.453 | 0.453 | 0.453 | 0.453 |
| 1984 | 0.444 | 0.715 | 0.733 | 0.722 | 0.682 | 0.594 | 0.594 | 0.594 | 0.594 | 0.594 |
| 1985 | 0.351 | 0.589 | 0.611 | 0.648 | 0.679 | 0.592 | 0.592 | 0.592 | 0.592 | 0.592 |
| 1986 | 0.241 | 0.448 | 0.496 | 0.571 | 0.649 | 0.594 | 0.594 | 0.594 | 0.594 | 0.594 |
| 1987 | 0.224 | 0.454 | 0.530 | 0.664 | 0.809 | 0.756 | 0.756 | 0.756 | 0.756 | 0.756 |
| 1988 | 0.214 | 0.456 | 0.537 | 0.660 | 0.772 | 0.663 | 0.663 | 0.663 | 0.663 | 0.663 |
| 1989 | 0.201 | 0.423 | 0.471 | 0.525 | 0.534 | 0.402 | 0.402 | 0.402 | 0.402 | 0.402 |
| 1990 | 0.223 | 0.477 | 0.523 | 0.592 | 0.602 | 0.453 | 0.453 | 0.453 | 0.453 | 0.453 |
| 1991 | 0.191 | 0.426 | 0.477 | 0.551 | 0.568 | 0.43 | 0.43 | 0.43 | 0.43 | 0.43 |
| 1992 | 0.172 | 0.429 | 0.54 | 0.689 | 0.754 | 0.605 | 0.605 | 0.605 | 0.605 | 0.605 |
| 1993 | 0.13 | 0.354 | 0.475 | 0.62 | 0.679 | 0.542 | 0.542 | 0.542 | 0.542 | 0.542 |
| 1994 | 0.1 | 0.297 | 0.419 | 0.568 | 0.629 | 0.505 | 0.505 | 0.505 | 0.505 | 0.505 |
| 1995 | 0.081 | 0.249 | 0.339 | 0.438 | 0.471 | 0.372 | 0.372 | 0.372 | 0.372 | 0.372 |
| 1996 | 0.073 | 0.227 | 0.315 | 0.421 | 0.488 | 0.418 | 0.418 | 0.418 | 0.418 | 0.418 |
| 1997 | 0.053 | 0.163 | 0.226 | 0.297 | 0.338 | 0.291 | 0.291 | 0.291 | 0.291 | 0.291 |
| 1998 | 0.046 | 0.153 | 0.221 | 0.297 | 0.347 | 0.322 | 0.322 | 0.322 | 0.322 | 0.322 |
| 1999 | 0.045 | 0.157 | 0.228 | 0.298 | 0.338 | 0.321 | 0.321 | 0.321 | 0.321 | 0.321 |


| Year <br> Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 0.038 | 0.139 | 0.205 | 0.267 | 0.295 | 0.29 | 0.29 | 0.29 | 0.29 | 0.29 |
| 2001 | 0.029 | 0.115 | 0.177 | 0.237 | 0.264 | 0.272 | 0.272 | 0.272 | 0.272 | 0.272 |
| 2002 | 0.026 | 0.108 | 0.168 | 0.228 | 0.261 | 0.289 | 0.289 | 0.289 | 0.289 | 0.289 |
| 2003 | 0.024 | 0.102 | 0.157 | 0.216 | 0.261 | 0.323 | 0.323 | 0.323 | 0.323 | 0.323 |
| 2004 | 0.022 | 0.095 | 0.148 | 0.206 | 0.261 | 0.348 | 0.348 | 0.348 | 0.348 | 0.348 |
| 2005 | 0.032 | 0.126 | 0.181 | 0.241 | 0.29 | 0.377 | 0.377 | 0.377 | 0.377 | 0.377 |
| 2006 | 0.039 | 0.154 | 0.214 | 0.285 | 0.344 | 0.455 | 0.455 | 0.455 | 0.455 | 0.455 |
| 2007 | 0.046 | 0.171 | 0.229 | 0.299 | 0.356 | 0.463 | 0.463 | 0.463 | 0.463 | 0.463 |
| 2008 | 0.07 | 0.248 | 0.299 | 0.365 | 0.42 | 0.531 | 0.531 | 0.531 | 0.531 | 0.531 |
| 2009 | 0.08 | 0.275 | 0.322 | 0.372 | 0.418 | 0.52 | 0.52 | 0.52 | 0.52 | 0.52 |
| 2010 | 0.097 | 0.328 | 0.373 | 0.405 | 0.431 | 0.505 | 0.505 | 0.505 | 0.505 | 0.505 |
| 2011 | 0.096 | 0.313 | 0.369 | 0.409 | 0.439 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 |
| 2012 | 0.101 | 0.303 | 0.353 | 0.384 | 0.408 | 0.437 | 0.437 | 0.437 | 0.437 | 0.437 |
| 2013 | 0.084 | 0.249 | 0.293 | 0.317 | 0.338 | 0.351 | 0.351 | 0.351 | 0.351 | 0.351 |
| 2014 | 0.074 | 0.219 | 0.265 | 0.288 | 0.313 | 0.325 | 0.325 | 0.325 | 0.325 | 0.325 |
| 2015 | 0.068 | 0.206 | 0.252 | 0.273 | 0.298 | 0.305 | 0.305 | 0.305 | 0.305 | 0.305 |
| 2016 | 0.058 | 0.185 | 0.239 | 0.273 | 0.311 | 0.331 | 0.331 | 0.331 | 0.331 | 0.331 |
| 2017 | 0.05 | 0.159 | 0.209 | 0.249 | 0.294 | 0.319 | 0.319 | 0.319 | 0.319 | 0.319 |
| 2018 | 0.051 | 0.156 | 0.206 | 0.251 | 0.304 | 0.333 | 0.333 | 0.333 | 0.333 | 0.333 |
| 2019 | 0.048 | 0.139 | 0.18 | 0.221 | 0.271 | 0.291 | 0.291 | 0.291 | 0.291 | 0.291 |
| 2020 | 0.045 | 0.129 | 0.164 | 0.203 | 0.254 | 0.274 | 0.274 | 0.274 | 0.274 | 0.274 |
| 2021 | 0.044 | 0.124 | 0.16 | 0.201 | 0.259 | 0.291 | 0.291 | 0.291 | 0.291 | 0.291 |

Table 5.10 Estimated stock numbers.

| Year <br> Age | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1960 | 84026 | 103212 | 54063 | 28175 | 26072 | 14377 | 10474 | 7296 | 3627 | 12068 |
| 1961 | 116162 | 56676 | 68814 | 30180 | 17272 | 15941 | 8956 | 6995 | 5128 | 11294 |
| 1962 | 206835 | 67972 | 36513 | 44590 | 18691 | 12606 | 11345 | 6190 | 5188 | 12543 |
| 1963 | 273837 | 133053 | 38589 | 25467 | 28675 | 11916 | 9860 | 8217 | 4491 | 13419 |
| 1964 | 80835 | 192878 | 77533 | 22463 | 17668 | 18939 | 8050 | 7502 | 6155 | 13840 |


| $\begin{aligned} & \text { Year } \\ & \text { Age } \end{aligned}$ | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 254979 | 49841 | 112584 | 45181 | 14490 | 11634 | 12341 | 5023 | 5213 | 13956 |
| 1966 | 134273 | 182245 | 34470 | 63060 | 26337 | 9323 | 7542 | 7296 | 3184 | 12731 |
| 1967 | 174211 | 83249 | 111131 | 20133 | 36539 | 16005 | 6325 | 5254 | 4560 | 10163 |
| 1968 | 143787 | 116727 | 47168 | 64268 | 12956 | 23786 | 10028 | 4120 | 3369 | 8278 |
| 1969 | 267366 | 88097 | 80560 | 31720 | 42494 | 10731 | 17821 | 6985 | 2677 | 6823 |
| 1970 | 220408 | 169181 | 58085 | 54888 | 22485 | 29924 | 9254 | 14125 | 5140 | 7175 |
| 1971 | 229850 | 143772 | 87223 | 35346 | 32832 | 14314 | 17677 | 6588 | 9316 | 7926 |
| 1972 | 154265 | 138705 | 86049 | 46349 | 22941 | 19544 | 9605 | 10385 | 4330 | 10127 |
| 1973 | 201294 | 80094 | 79530 | 52478 | 27745 | 15439 | 12675 | 6829 | 6372 | 8956 |
| 1974 | 100846 | 110892 | 41709 | 46327 | 32933 | 16776 | 10303 | 8264 | 4297 | 9026 |
| 1975 | 168309 | 44068 | 52917 | 19820 | 23857 | 17932 | 9298 | 6058 | 4796 | 7160 |
| 1976 | 220420 | 75068 | 19305 | 25739 | 10487 | 11393 | 8684 | 4696 | 3080 | 5760 |
| 1977 | 202624 | 90088 | 30935 | 8391 | 13327 | 5454 | 5692 | 4260 | 2303 | 4198 |
| 1978 | 136704 | 89616 | 38553 | 15022 | 4590 | 7301 | 3202 | 3088 | 2395 | 3966 |
| 1979 | 195867 | 60076 | 38732 | 17165 | 7704 | 2359 | 4023 | 1756 | 1536 | 3421 |
| 1980 | 118880 | 94852 | 23529 | 16838 | 8555 | 3654 | 1123 | 2060 | 963 | 2671 |
| 1981 | 232133 | 57025 | 43586 | 9993 | 8254 | 4418 | 1830 | 686 | 1062 | 1820 |
| 1982 | 127952 | 125404 | 24407 | 19554 | 4695 | 4369 | 2237 | 1031 | 397 | 1629 |
| 1983 | 100879 | 68200 | 54301 | 9874 | 9348 | 2589 | 2500 | 1240 | 604 | 1293 |
| 1984 | 94848 | 58223 | 30631 | 20737 | 4285 | 4557 | 1300 | 1334 | 710 | 1061 |
| 1985 | 104305 | 42143 | 23127 | 12843 | 7080 | 1920 | 2089 | 553 | 608 | 830 |
| 1986 | 178608 | 49257 | 17676 | 11011 | 5982 | 2441 | 945 | 954 | 268 | 630 |
| 1987 | 144151 | 132580 | 22527 | 8341 | 5503 | 2777 | 854 | 481 | 424 | 466 |
| 1988 | 80501 | 101647 | 76524 | 11120 | 3458 | 2046 | 1327 | 228 | 201 | 292 |
| 1989 | 78046 | 54928 | 56075 | 39182 | 4874 | 1188 | 817 | 616 | 51 | 290 |
| 1990 | 87261 | 47774 | 29573 | 26563 | 18844 | 2441 | 593 | 458 | 369 | 216 |
| 1991 | 226767 | 48317 | 22071 | 15097 | 11249 | 8480 | 1237 | 296 | 263 | 324 |
| 1992 | 281942 | 142737 | 22449 | 10926 | 7827 | 5050 | 4674 | 646 | 168 | 376 |
| 1993 | 211259 | 213473 | 76372 | 10118 | 4266 | 3120 | 1966 | 2308 | 279 | 238 |


| Year <br> Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 150273 | 162581 | 132562 | 37440 | 4346 | 1719 | 1484 | 755 | 1243 | 269 |
| 1995 | 274143 | 132666 | 112259 | 75546 | 15576 | 1848 | 794 | 777 | 300 | 828 |
| 1996 | 158412 | 244059 | 88297 | 68551 | 40422 | 7966 | 1036 | 484 | 447 | 705 |
| 1997 | 164614 | 120139 | 178524 | 58072 | 40148 | 21591 | 4151 | 503 | 259 | 629 |
| 1998 | 104290 | 135570 | 83800 | 128048 | 32888 | 24116 | 12886 | 2554 | 332 | 632 |
| 1999 | 241011 | 78990 | 95788 | 53536 | 73974 | 18368 | 15028 | 7672 | 1477 | 581 |
| 2000 | 159210 | 193027 | 51166 | 55833 | 31213 | 40670 | 11326 | 9626 | 4373 | 1130 |
| 2001 | 212316 | 106590 | 140257 | 35446 | 33157 | 18987 | 24144 | 7251 | 6082 | 3184 |
| 2002 | 357911 | 178223 | 78274 | 93953 | 23958 | 20586 | 12629 | 15001 | 4495 | 5918 |
| 2003 | 150915 | 317001 | 123912 | 51622 | 56760 | 17215 | 12746 | 8645 | 9148 | 6508 |
| 2004 | 153670 | 121325 | 209984 | 86127 | 35543 | 36407 | 10979 | 7448 | 5520 | 9153 |
| 2005 | 436325 | 119168 | 79064 | 125579 | 56566 | 23815 | 22286 | 6916 | 3815 | 7598 |
| 2006 | 73821 | 345200 | 79938 | 48535 | 73946 | 34788 | 14872 | 12596 | 3933 | 6097 |
| 2007 | 113108 | 53944 | 216276 | 52597 | 29793 | 39911 | 19826 | 8324 | 6281 | 4403 |
| 2008 | 200409 | 76125 | 37793 | 114916 | 30173 | 16518 | 19969 | 10815 | 4207 | 5099 |
| 2009 | 145999 | 154224 | 46054 | 25005 | 62623 | 15713 | 7883 | 9287 | 5323 | 4258 |
| 2010 | 269620 | 98774 | 91079 | 28578 | 14178 | 33189 | 7780 | 3771 | 4152 | 4392 |
| 2011 | 113082 | 199262 | 50578 | 46919 | 15677 | 8135 | 15981 | 3933 | 1849 | 3951 |
| 2012 | 153896 | 91823 | 123730 | 31155 | 24749 | 9056 | 4398 | 7733 | 1916 | 2866 |
| 2013 | 209004 | 92107 | 63809 | 77806 | 18423 | 13253 | 5007 | 2445 | 3911 | 2504 |
| 2014 | 108650 | 170558 | 60363 | 42747 | 46061 | 11056 | 7654 | 3138 | 1446 | 3686 |
| 2015 | 165109 | 80832 | 121163 | 41909 | 28657 | 27070 | 6462 | 4433 | 1956 | 3238 |
| 2016 | 252926 | 119916 | 54779 | 73849 | 27556 | 17665 | 15680 | 3768 | 2878 | 3794 |
| 2017 | 178636 | 220533 | 82076 | 34542 | 41960 | 16410 | 10391 | 8671 | 2096 | 4721 |
| 2018 | 130677 | 151127 | 179261 | 60196 | 24652 | 23598 | 9822 | 6206 | 4995 | 4402 |
| 2019 | 257000 | 124197 | 112305 | 120744 | 36398 | 15055 | 12994 | 5636 | 3480 | 5550 |
| 2020 | 122722 | 233796 | 104193 | 82456 | 74786 | 23057 | 9465 | 7550 | 3336 | 5780 |
| 2021 | 147428 | 91131 | 190540 | 79694 | 56869 | 48629 | 14558 | 6074 | 4587 | 6134 |
| pred |  | 115519 | 65936 | 132944 | 53373 | 35935 | 29775 | 8914 | 3719 | 6565 |

Table 5.11 Estimated recruitment, total-stock biomass (TBS), spawning-stock biomass (SSB), and average fishing mortality for ages 4 to 7 (F4-7).

| Year | $\begin{aligned} & \text { R } \\ & \text { (age 3) } \end{aligned}$ | Low | High | SSB | Low | High | $\begin{aligned} & \text { Fbar } \\ & (4-7) \end{aligned}$ | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | 84026 | 52561 | 134326 | 462688 | 338674 | 632112 | 0.276 | 0.198 | 0.387 | 686916 | 533851 | 883869 |
| 1961 | 116162 | 76540 | 176295 | 454708 | 335633 | 616028 | 0.233 | 0.170 | 0.319 | 661579 | 517022 | 846553 |
| 1962 | 206835 | 137011 | 312245 | 460869 | 343520 | 618305 | 0.233 | 0.172 | 0.315 | 725964 | 576883 | 913571 |
| 1963 | 273837 | 181598 | 412927 | 458340 | 345386 | 608234 | 0.246 | 0.184 | 0.330 | 837994 | 675845 | 1039045 |
| 1964 | 80835 | 53177 | 122880 | 483760 | 370183 | 632184 | 0.283 | 0.213 | 0.377 | 818944 | 659137 | 1017496 |
| 1965 | 254979 | 169260 | 384110 | 523809 | 405297 | 676974 | 0.289 | 0.218 | 0.384 | 858901 | 696125 | 1059738 |
| 1966 | 134273 | 89365 | 201748 | 482581 | 370844 | 627985 | 0.299 | 0.225 | 0.398 | 827172 | 670337 | 1020701 |
| 1967 | 174211 | 115695 | 262323 | 494141 | 382863 | 637762 | 0.280 | 0.210 | 0.373 | 800174 | 649854 | 985264 |
| 1968 | 143787 | 95615 | 216229 | 469782 | 362951 | 608057 | 0.202 | 0.151 | 0.270 | 758020 | 616216 | 932456 |
| 1969 | 267366 | 177243 | 403313 | 509859 | 402401 | 646012 | 0.195 | 0.147 | 0.259 | 869361 | 717880 | 1052806 |
| 1970 | 220408 | 146917 | 330662 | 568159 | 457854 | 705038 | 0.310 | 0.238 | 0.403 | 973772 | 817707 | 1159623 |
| 1971 | 229850 | 153927 | 343220 | 554682 | 452021 | 680661 | 0.327 | 0.253 | 0.422 | 954274 | 806359 | 1129321 |
| 1972 | 154265 | 103449 | 230043 | 535848 | 440342 | 652069 | 0.321 | 0.250 | 0.413 | 878566 | 745483 | 1035406 |
| 1973 | 201294 | 135058 | 300013 | 537224 | 446847 | 645881 | 0.358 | 0.280 | 0.457 | 846588 | 723212 | 991011 |
| 1974 | 100846 | 67415 | 150854 | 493712 | 412902 | 590337 | 0.480 | 0.380 | 0.606 | 736039 | 632146 | 857006 |
| 1975 | 168309 | 112918 | 250872 | 398963 | 334802 | 475420 | 0.539 | 0.429 | 0.677 | 614139 | 527453 | 715071 |
| 1976 | 220420 | 147490 | 329412 | 281331 | 234555 | 337436 | 0.573 | 0.457 | 0.718 | 544141 | 461430 | 641678 |
| 1977 | 202624 | 135938 | 302023 | 208941 | 173586 | 251498 | 0.500 | 0.398 | 0.630 | 478268 | 402502 | 568295 |
| 1978 | 136704 | 91625 | 203960 | 189086 | 158224 | 225968 | 0.553 | 0.442 | 0.692 | 418443 | 354574 | 493817 |
| 1979 | 195867 | 131462 | 291824 | 170439 | 142582 | 203739 | 0.588 | 0.471 | 0.735 | 410417 | 343590 | 490243 |
| 1980 | 118880 | 79764 | 177178 | 150189 | 125504 | 179728 | 0.564 | 0.451 | 0.706 | 391858 | 328104 | 468000 |
| 1981 | 232133 | 155017 | 347614 | 154449 | 128375 | 185819 | 0.558 | 0.446 | 0.698 | 447833 | 368831 | 543756 |
| 1982 | 127952 | 85652 | 191143 | 135715 | 112885 | 163162 | 0.555 | 0.442 | 0.696 | 403434 | 333787 | 487614 |
| 1983 | 100879 | 67291 | 151234 | 164048 | 135411 | 198741 | 0.603 | 0.483 | 0.752 | 410114 | 342736 | 490739 |
| 1984 | 94848 | 63020 | 142751 | 146889 | 121652 | 177361 | 0.713 | 0.575 | 0.885 | 323432 | 272348 | 384098 |
| 1985 | 104305 | 69193 | 157235 | 110715 | 92052 | 133162 | 0.632 | 0.507 | 0.788 | 270744 | 226156 | 324123 |
| 1986 | 178608 | 118543 | 269107 | 83490 | 69335 | 100536 | 0.541 | 0.432 | 0.677 | 266515 | 217251 | 326950 |
| 1987 | 144151 | 96320 | 215734 | 72061 | 59969 | 86591 | 0.614 | 0.495 | 0.761 | 284521 | 232275 | 348519 |


| Year | $\begin{aligned} & \text { R } \\ & \text { (age 3) } \end{aligned}$ | Low | High | SSB | Low | High | $\begin{aligned} & \text { Fbar } \\ & (4-7) \end{aligned}$ | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 80501 | 53240 | 121721 | 88318 | 72923 | 106963 | 0.606 | 0.488 | 0.753 | 302927 | 249183 | 368263 |
| 1989 | 78046 | 51493 | 118292 | 104092 | 80609 | 134415 | 0.488 | 0.388 | 0.615 | 286337 | 236432 | 346777 |
| 1990 | 87261 | 57160 | 133215 | 120178 | 95890 | 150620 | 0.549 | 0.437 | 0.689 | 273010 | 228545 | 326126 |
| 1991 | 226767 | 149759 | 343375 | 114661 | 93974 | 139901 | 0.506 | 0.402 | 0.636 | 355669 | 288542 | 438413 |
| 1992 | 281942 | 186702 | 425764 | 95211 | 80072 | 113212 | 0.603 | 0.483 | 0.753 | 464558 | 373132 | 578385 |
| 1993 | 211259 | 141268 | 315927 | 97293 | 80974 | 116900 | 0.532 | 0.425 | 0.666 | 533627 | 431635 | 659720 |
| 1994 | 150273 | 102344 | 220647 | 148467 | 120525 | 182887 | 0.478 | 0.379 | 0.603 | 485997 | 402285 | 587130 |
| 1995 | 274143 | 185154 | 405903 | 197554 | 158396 | 246391 | 0.374 | 0.294 | 0.476 | 588527 | 488899 | 708456 |
| 1996 | 158412 | 107589 | 233241 | 246590 | 200772 | 302864 | 0.363 | 0.284 | 0.463 | 682201 | 569882 | 816657 |
| 1997 | 164614 | 111990 | 241966 | 246211 | 200966 | 301643 | 0.256 | 0.198 | 0.331 | 725654 | 604353 | 871300 |
| 1998 | 104290 | 71230 | 152695 | 294713 | 240842 | 360634 | 0.254 | 0.196 | 0.329 | 803607 | 669600 | 964433 |
| 1999 | 241011 | 164532 | 353040 | 309916 | 250154 | 383956 | 0.255 | 0.196 | 0.332 | 806148 | 677769 | 958842 |
| 2000 | 159210 | 108688 | 233216 | 368993 | 298161 | 456652 | 0.226 | 0.174 | 0.295 | 825951 | 697965 | 977407 |
| 2001 | 212316 | 146303 | 308116 | 374833 | 307242 | 457293 | 0.198 | 0.153 | 0.257 | 883686 | 751264 | 1039450 |
| 2002 | 357911 | 251930 | 508475 | 450424 | 375437 | 540388 | 0.191 | 0.148 | 0.247 | 1027724 | 880623 | 1199397 |
| 2003 | 150915 | 106003 | 214855 | 437861 | 368459 | 520334 | 0.184 | 0.143 | 0.237 | 1003187 | 858533 | 1172213 |
| 2004 | 153670 | 106851 | 221003 | 518880 | 441074 | 610410 | 0.178 | 0.137 | 0.230 | 1016491 | 870454 | 1187028 |
| 2005 | 436325 | 305892 | 622375 | 602367 | 509925 | 711569 | 0.209 | 0.162 | 0.270 | 1097718 | 941847 | 1279385 |
| 2006 | 73821 | 52104 | 104591 | 535304 | 456254 | 628049 | 0.249 | 0.194 | 0.320 | 942032 | 809282 | 1096558 |
| 2007 | 113108 | 80112 | 159694 | 545628 | 466743 | 637846 | 0.264 | 0.206 | 0.338 | 880826 | 754665 | 1028077 |
| 2008 | 200409 | 142439 | 281971 | 468492 | 394737 | 556028 | 0.333 | 0.262 | 0.424 | 730583 | 629841 | 847439 |
| 2009 | 145999 | 103996 | 204967 | 361785 | 304858 | 429342 | 0.347 | 0.275 | 0.439 | 677268 | 585513 | 783402 |
| 2010 | 269620 | 192575 | 377489 | 327806 | 277143 | 387730 | 0.384 | 0.304 | 0.486 | 698709 | 600271 | 813290 |
| 2011 | 113082 | 80132 | 159583 | 292358 | 246890 | 346200 | 0.382 | 0.301 | 0.485 | 584827 | 501895 | 681463 |
| 2012 | 153896 | 109359 | 216571 | 301256 | 254999 | 355904 | 0.362 | 0.285 | 0.459 | 595305 | 510906 | 693647 |
| 2013 | 209004 | 148918 | 293333 | 323389 | 270270 | 386947 | 0.299 | 0.235 | 0.382 | 608668 | 520886 | 711244 |
| 2014 | 108650 | 77223 | 152865 | 348814 | 291208 | 417817 | 0.271 | 0.212 | 0.347 | 642313 | 549371 | 750979 |
| 2015 | 165109 | 117480 | 232047 | 357938 | 298395 | 429363 | 0.257 | 0.200 | 0.330 | 626740 | 534207 | 735301 |
| 2016 | 252926 | 178613 | 358158 | 391741 | 323101 | 474963 | 0.252 | 0.195 | 0.326 | 794025 | 671466 | 938953 |


| Year | R <br> (age 3) | Low | High | SSB | Low | High | Fbar <br> (4-7) | Low | High | TSB | Low | High |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2017 | 178636 | 126300 | 252660 | 401931 | 329879 | 489720 | 0.228 | 0.175 | 0.297 | 891948 | 751182 | 1059092 |
| 2018 | 130677 | 90890 | 187881 | 464929 | 378166 | 571597 | 0.229 | 0.175 | 0.300 | 940084 | 786349 | 1123876 |
| 2019 | 257000 | 179210 | 368558 | 560109 | 445368 | 704413 | 0.203 | 0.152 | 0.270 | 1055975 | 875840 | 1273157 |
| 2020 | 122722 | 83756 | 179818 | 616956 | 480534 | 792107 | 0.187 | 0.138 | 0.255 | 1069119 | 871949 | 1310875 |
| 2021 | 147428 | 92304 | 235474 | 715674 | 542678 | 943818 | 0.186 | 0.131 | 0.264 | 1140302 | 899686 | 1445270 |

Table 5.12 Northeast Arctic saithe. Prediction input data
rMFDP version
Run: $r$
$F_{\text {bar }}$ age range: 4-7

2022

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 161659 | 0.2 | 0 | 0 | 0 | 0.78 | 0.046 | 0.78 |
| 4 | 115519 | 0.2 | 0.05 | 0 | 0 | 1.128 | 0.131 | 1.128 |
| 5 | 65936 | 0.2 | 0.42 | 0 | 0 | 1.605 | 0.168 | 1.605 |
| 6 | 132944 | 0.2 | 0.87 | 0 | 0 | 2.088 | 0.208 | 2.088 |
| 7 | 53373 | 0.2 | 0.97 | 0 | 0 | 2.651 | 0.261 | 2.651 |
| 8 | 29775 | 0.2 | 0.98 | 0 | 0 | 3.192 | 0.285 | 3.192 |
| 9 | 0914 | 0.2 | 0.98 | 0 | 0 | 3.715 | 0.285 | 3.715 |
| 10 | 0565 | 0.2 | 0.97 | 0 | 0 | 4.16 | 0.285 | 4.16 |
| 11 | 0.97 | 0 | 0 | 0 | 0.834 | 0.285 | 4.833 |  |
| 12 | 0.2 | 0.285 | 6.434 |  |  |  |  |  |

2023

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 161659 | 0.2 | 0 | 0 | 0 | 0.78 | 0.046 | 0.78 |
| 4 | $\cdot$ | 0.2 | 0.05 | 0 | 0 | 1.128 | 0.131 | 1.128 |
| 5 | $\cdot$ | 0.2 | 0.42 | 0 | 0 | 1.605 | 0.168 | 1.605 |
| 6 | $\cdot$ | 0.2 | 0.87 | 0 | 0 | 2.088 | 0.208 | 2.088 |
| 7 | $\cdot$ | 0.2 | 0.97 | 0 | 0 | 2.651 | 0.261 | 2.651 |
| 8 | 0.2 | 0.98 | 0 | 0 | 3.192 | 0.285 | 3.192 |  |


| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 9 | $\cdot$ | 0.2 | 0.98 | 0 | 0 | 3.715 | 0.285 | 3.715 |
| 10 | $\cdot$ | 0.2 | 0.97 | 0 | 0 | 4.16 | 0.285 | 4.16 |
| 11 | $\cdot$ | 0.2 | 0.97 | 0 | 0 | 4.833 | 0.285 | 4.833 |
| 12 | $\cdot$ | 0.2 | 0.99 | 0 | 0 | 6.434 | 0.285 | 6.434 |

2024

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 161659 | 0.2 | 0 | 0 | 0 | 0.78 | 0.046 | 0.78 |
| 4 | . | 0.2 | 0.05 | 0 | 0 | 1.128 | 0.131 | 1.128 |
| 5 | . | 0.2 | 0.42 | 0 | 0 | 1.605 | 0.168 | 1.605 |
| 6 | . | 0.2 | 0.87 | 0 | 0 | 2.088 | 0.208 | 2.088 |
| 7 | . | 0.2 | 0.97 | 0 | 0 | 2.651 | 0.261 | 2.651 |
| 8 | . | 0.2 | 0.98 | 0 | 0 | 3.192 | 0.285 | 3.192 |
| 9 | . | 0.2 | 0.98 | 0 | 0 | 3.715 | 0.285 | 3.715 |
| 10 | . | 0.2 | 0.97 | 0 | 0 | 4.16 | 0.285 | 4.16 |
| 11 | . | 0.2 | 0.97 | 0 | 0 | 4.833 | 0.285 | 4.833 |
| 12 | . | 0.2 | 0.99 | 0 | 0 | 6.434 | 0.285 | 6.434 |

Input units are thousands and kg - output in tonnes

Table 5.13 Northeast Arctic saithe. Short-term prediction
rMFDP version
Run: r
$F_{\text {bar }}$ age range: 4-7

2022

| Biomass | SSB | F Mult | F | Bar |
| :--- | :--- | :--- | :--- | :--- |

2023-2024

| 2023 |  |  |  | 2024 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Biomass | SSB | F Mult | Fandings | Biomass | SSB |  |
| 1050549 | 686937 | 0 | 0 | 0 | 1212129 | 815773 |
| . | 686937 | 0.1 | 0.0192 | 19638 | 1190491 | 796705 |


| 2023 |  |  |  |  | 2024 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | $F_{\text {Mult }}$ | $F_{\text {Bar }}$ | Landings | Biomass | SSB |
| . | 686937 | 0.2 | 0.0384 | 38826 | 1169356 | 778107 |
| . | 686937 | 0.3 | 0.0576 | 57575 | 1148713 | 759967 |
| . | 686937 | 0.4 | 0.0768 | 75895 | 1128549 | 742273 |
| . | 686937 | 0.5 | 0.096 | 93798 | 1108852 | 725014 |
| . | 686937 | 0.6 | 0.1152 | 111294 | 1089611 | 708179 |
| . | 686937 | 0.7 | 0.1344 | 128393 | 1070814 | 691758 |
| . | 686937 | 0.8 | 0.1536 | 145104 | 1052450 | 675738 |
| . | 686937 | 0.9 | 0.1728 | 161438 | 1034509 | 660112 |
| . | 686937 | 1 | 0.192 | 177402 | 1016980 | 644867 |
| . | 686937 | 1.1 | 0.2112 | 193007 | 999852 | 629995 |
| - | 686937 | 1.2 | 0.2304 | 208262 | 983117 | 615487 |
| - | 686937 | 1.3 | 0.2496 | 223174 | 966763 | 601332 |
| - | 686937 | 1.4 | 0.2688 | 237752 | 950783 | 587523 |
| . | 686937 | 1.5 | 0.288 | 252005 | 935166 | 574049 |
| . | 686937 | 1.6 | 0.3072 | 265940 | 919904 | 560904 |
| . | 686937 | 1.7 | 0.3264 | 279566 | 904988 | 548078 |
| . | 686937 | 1.8 | 0.3456 | 292888 | 890410 | 535564 |
| - | 686937 | 1.9 | 0.3648 | 305916 | 876161 | 523353 |
| - | 686937 | 2 | 0.384 | 318656 | 862233 | 511438 |

Input units are thousands and kg - output in tonnes

Table 5.14 Northeast arctic saithe. Short-term projection output HCR landings
rMFDP version

## Run: $r$

$F_{\text {bar }}$ age range: 4-7

2022

| Biomass | SSB | F Mult | F Bar $^{\text {Landings }}$ |  |
| :--- | :--- | :--- | :--- | :--- |
| 1103920 | 745913 | 1.0786 | 0.2071 | 197212 |

2023

| Biomass | SSB | $\mathbf{F}_{\text {Mult }}$ | $\mathbf{F}_{\text {Bar }}$ | Landings |
| :--- | :--- | :--- | :--- | :--- |
| 1050549 | 686937 | 1.3246 | 0.254 | 226794 |

2024

| Biomass | SSB | F Mult | F Bar | Landings |
| :--- | :--- | :--- | :--- | :--- |
| 962796 | 597899 | 1.667 | 0.32 | 246332 |



Figure 5.1. Northeast Arctic saithe. Echo abundance and proportion of saithe in the southern half of the survey area (subarea C+D).


Figure 5.2. Northeast Arctic saithe. acoustic survey tuning indices by age class (3-7). break in 2002 black line.


Figure 5.3. Northeast Arctic saithe. Final run normalized residuals. Blue circles indicate positive residuals (larger than predicted) and filled red circles indicate negative residuals. The top figure shows residuals for the total catch series. the figure in the middle the residuals for the first survey series and the bottom figure the residuals for the survey series from 2002.


Figure 5.4. NEA saithe - Acoustic survey vs. SAM. Green point 2021 data.


Figure 5.5. Northeast Arctic saithe (subareas 1 and 2).



Figure 5.6. Saithe in subareas 1 and 2 (Northeast Arctic) RETROSPECTIVE SAM SSB. F4-7. and recruits.

## 6 Northeast Arctic beaked redfish ${ }^{1}$

On 30 March 2022 all Russian participation in ICES was suspended. As a result of this decision, it is not possible to run ICES stock assessments or provide ICES advice for the Barents Sea stocks of NEA cod, NEA haddock, Sebastes mentella or Greenland Halibut, as management and data collection for these stocks are shared between Norway and Russia. There is therefore no stock assessment for NEA cod this year, but input data to the assessment are updated as far as possible.
The chapter was therefore updated as in a non-assessment year. This comprises tables 6.1 through to 6.7 and tables 6.12 , to 6.19 . Updated figures comprise figures $6.1,6.2,6.6,6.7$ (upper panel), 6.8, 6.10, 6.12, 6.15 and 6.16.

### 6.1 Status of the fisheries

### 6.1.1 Development of the fishery

A description of the historical development of the fishery in subareas 1 and 2 is found in the stock annex for this stock.

An international pelagic fishery for S. mentella in the Norwegian Sea outside EEZs has developed since 2004 (Figure 6.1). This pelagic fishery, which is further described in the stock annex, is managed by the Northeast Atlantic Fisheries Commission (NEAFC). Since 2014 the directed demersal and pelagic fisheries are reopened in the Norwegian Economic Zone, the Fisheries Protection Zone around Svalbard and, for pelagic fisheries only, in the Fishing Zone around Jan Mayen. The spatial regulation for this fishery is illustrated in Figures 6.2 and 6.3. In 2021, most of the catches of S. mentella from the Russian and Norwegian fisheries were taken in the Norwegian Exclusive Economic Zone or as bycatch in the Fisheries Protection Zone around Svalbard. Catches in international waters were mainly taken by EU nations.

Figure 6.2 shows the distribution of catch among national fishing fleets for 2018 to 2021 and the location of Norwegian S. mentella catches in the Norwegian EEZ in 2021 as well as bycatch in other areas. The $44^{\text {th }}$ Session of the Joint Norwegian-Russian Fisheries Commission decided to split the total TAC among countries as follows: Norway: 72\%, Russia: 18\%, Third countries: $10 \%$ (as bycatch in the fishery protection zone at Svalbard (Spitsbergen): $4.1 \%$, and international waters of the Norwegian Sea (NEAFC-area): $5.9 \%$ ). This split was reconducted at the $51^{\text {st }}$ session of the commission in 2021.

### 6.1.2 Bycatch in other fisheries

During 2003-2013, all catches of S. mentella, except the pelagic fishery in the Norwegian Sea outside EEZ, were taken as bycatches in other fisheries. Some of the pelagic catches are taken as bycatches in the blue whiting and herring fisheries. From 2014 onwards most of the catch is taken as targeted catch and no longer as bycatch, following the opening of a targeted fishery in the Norwegian EEZ, Svalbard Fisheries Protection Zone and around Jan Mayen. When fishing for other species it has since 2013 been allowed to have up to $20 \%$ redfish (both species together) in round weight as bycatch outside 12 nautical miles and only $10 \%$ bycatch inside 12 nautical miles to better protect $S$. norvegicus.

[^9]
### 6.1.3 Landings prior to 2021 (Tables 6.1-6.7, Figure 6.1)

Nominal catches of S. mentella by country for subareas 1 and 2 combined are presented in Table 6.1, while they are presented for Subarea 1 and divisions $2 . a$ and $2 . b$ in Tables 6.2-6.4. The pelagic catch of S. mentella in the Norwegian Sea outside EEZs reported to NEAFC and/or ICES amounted to 7739 t in 2018, 6060 t in 2019, 5469 t in 2020 and 2872 t in 2021, and is shown by country in Table 6.5. Nominal catches for both redfish species combined (i.e. S. mentella and S. norvegicus) by country are presented in Table 6.6. The sources of information used are catches reported to ICES, NEAFC, Norwegian and Russian authorities (foreign vessels fishing in the Norwegian and Russian economic zones) or direct reporting to the AFWG. Where catches are reported as Sebastes sp., they are split into S. norvegicus and S. mentella by AFWG experts based on available correlation between official catches of these two species in the considered areas. All tables have been updated for 2020, and new figures presented for 2021. Total international landings in 1952-2021 are also shown in Figure 6.1.

In 2014, ICES advised that the annual catch in 2015, 2016, and 2017 should be set at no more than 30000 t and in 2017, ICES advised that the annual catch in 2018 should not exceed 32658 t . Following the benchmark (WKREDFISH, ICES 2018a) and the subsequent evaluation of a management plan for the stock (WKREBMSE, ICES 2018b) ICES advised an annual catch of no more than 53757 t for 2019 and 55860 t in 2020, corresponding to a fishing mortality of $\mathrm{F}=0.06$. This was continued in 2020, when ICES advised an annual catch of no more than 66158 t in 2021 and 67210 t in 2022, still corresponding to $\mathrm{F}=0.06$. No advice was given in 2022.
Because of the novelty of the situation, related with reopening fisheries after 10 years of its ban, the total landings of S. mentella in subareas 1 and 2 in 2014, demersal and pelagic catches, amounted to only 18426 t . The total landings of the demersal and pelagic fishery increased to 34754 t in 2016, 30783 t in 2017, 38046 t in 2018, 45640 t in 2019, 53631 t in 2020 and 63482 t in 2021. Of this, 2872 t were reported from the pelagic fishery in international waters of the Norwegian Sea. The total landings in 2017 and 2018 were respectively 783 t and 5388 t above the TAC advised by ICES, but were 8117 t, 2229 t and 2676 t below TAC in 2019, 2020 and 2021, respectively. Norway caught the major share of the demersal catches, but Russian demersal catches increased substantially after 2017, particularly in ICES Division 2.b.

The redfish population in Subarea 4 (North Sea) is believed to belong to the Northeast Arctic stock. Since this area is outside the traditional areas handled by this Working Group, the catches are not included in the assessment. The total redfish landings (golden and beaked redfish combined) from Subarea 4 have up to 2003 been 1000-3000 t per year. Since 2005 the annual landings from this area have varied between 90 and 333 t (Table 6.7).

### 6.1.4 Expected landings in 2022

ICES has advised on the basis of precautionary considerations that the annual catch should be set at no more than 67210 t in 2022. The $51^{\text {st }}$ sessions of the Joint Norwegian-Russian Fisheries Commission decided to follow this advice.

In 2022 Norwegian fishing vessels, can catch and land up to 44291 t of redfish in the Norwegian economic zone (NEZ) in a limited area north of $65^{\circ} 20^{\prime} \mathrm{N}$ (see map in Figure 6.3), in international waters and the fisheries zone around Jan Mayen. Of this quantity, 100 t are allocated to cover bycatch in other fisheries and 52 t for research/surveillance and education purposes, while the remaining 43139 t can be taken in a directed fishery. Only vessels with cod and saithe trawl permits can participate in the directed fishery for redfish. Each vessel which has the right to participate is assigned a maximum quota, which can be adjusted during the year, per how much of the national quota is exploited. The fishery may be stopped if the total quota is reached. This
quota must also cover catches of redfish (both species) in other fisheries. It is prohibited to fish for redfish with bottom trawls in the period from 1 March until 10 May. Investigations were conducted in 2015-2016 to see if the protection of females during the main time of larvae release should be improved by extending the period of prohibited fishing until later in May, and to see if the area south of Bear Island (Area 20 in Figure 6.3) can be opened for directed fishing, either with or without sorting grid, and permissions were granted to a small number of vessels of the Norwegian reference fleet for an earlier onset of fishing to gain further data. The hitherto conclusion is that males dominated the catches (more than 70\%) in the main fishing areas south and southwest of Bear Island during the investigations from late April until the directed fishery started on 10 May, and that the area south of Bear Island should stay closed during JanuaryFebruary due to smaller S. mentella inhabiting this area at the beginning of the year.

Since 2015, Russia has had access to the NEZ when fishing their quota share. In 2021 Russia may fish 11908 t ( $18 \%$ ) plus 2000 t transferred from Norway to Russia. Apart from this an additional 2100 t were transferred from Norway to Russia to cover bycatch of redfish (both species) in Russian fisheries targeting other species. The remaining 6616 t are divided between third countries in the NEZ and Svalbard Zone (2713 t) and the NEAFC areas (3903 t). Catch in the NEAFC areas in 2021 amounted to 2872 t while the catch in the national economic zones of Norway and Russia as well as the fisheries protection zone around Svalbard was 60610 t . The total catch in 2021 was 2676 t lower than the advised TAC. It is assumed that the total catch in 2022 should not exceed the TAC of 67210 t set by ICES.

### 6.2 Data used in the assessment

Analytical assessment was conducted for this stock following recommendation from the benchmark assessment working group (WKREDFISH, ICES 2018a). Input datasets were updated with the most recently available data. The analytical assessment, based on a statistical catch-at-age model (SCAA), covers the period 1992-2020. The input data consists of the following tables:

- $\quad$ Total catch in tonnes (Table 6.1)
- $\quad$ Catch in tonnes in the pelagic fishery Norwegian Sea outside EEZs (Table 6.5)
- Total catch numbers-at-age 6-19+ (Table 6.8)
- $\quad$ Catch numbers-at-age $7-19+$ in the pelagic fishery (Table 6.9)
- Weight-at-age $2-19+$ in the population (Table 6.12)
- Maturity-at-age 2-19+ in the population (Table 6.14)
- Russian autumn survey numbers-at-age 0-11 (Table 6.15)
- Ecosystem survey numbers-at-age 2-15 (Table 6.17)
- Winter survey numbers-at-age 2-15 (Table 6.18b)
- Deep pelagic ecosystem survey proportions-at-age (Table 6.19)

There was no direct observation of catch numbers-at-age for the pelagic fishery in the Norwegian Sea outside EEZs in 2012-2021. Instead, numbers-at-age were estimated based on catch-at-age from previous or following year, and weight-at-age and fleet selectivities (section 6.2.2 in AFWG report 2013). In 2013, 2016 and 2019, observations from the scientific survey in the Norwegian Sea were used to derive numbers-at-age in the pelagic fishery. This was considered appropriate given that the survey operates in the area of the fishery, with a commercial pelagic trawl and at the time of the start of the fishery.

### 6.2.1 Length- composition from the fishery (Figure 6.4)

Comparison of length distributions of the Norwegian and Russian catches of S. mentella in 20192020 are shown in Figure 6.4. In 2020, the Russian and Norwegian fleets fished smaller fish than
in 2019, reflecting good year classes due to enter the fishable stock. In 2020 length of beaked redfish in Norwegian catches was larger than in Russian catches. This is probably due to differences in the fishing areas. The Russian fleet largely operated in area $2 b$, and the Norwegian fleet in area $2 a$.

### 6.2.2 Catch-at-age (Tables 6.8-6.11, Figure 6.5)

Catch-at-age in the Norwegian fishery was estimated using ECA for 2014. For 2015, 2016 and 2018, it was not possible to run ECA and the catch-at-age for the Norwegian Fishery was estimated using the older Biomass program in SAS (Table 6.8). Not enough age readings were available to estimate catch-at-age in 2017, 2019 and 2020. For the demersal fisheries 2017, 2019 and 2020 as well as the pelagic fisheries 2017, 2018 and 2020 (Table 6.9) proportions-at-age in the catch were derived from proportions at-age in earlier years, weight-at-age and fleet selectivity (section 6.2.2 in AFWG report 2013).

The procedure for estimating catch-at-age for recent years in which age data are not available is somewhat problematic. This is because the last year of observation has a large effect on the estimated catch-at-age for several years. At the assessment working group in 2017 and at the benchmark assessment in January 2018, the last year of observations for the catch-at-age was 2014 and the values for the years 2015 and 2016 were extrapolated. Once available, the data for 2015 (demersal) and 2016 (pelagic) were substantially different from these earlier extrapolations.

Age composition of the Russian and Norwegian catches in 2020 was calculated using the agelength key, based on Russian age readings. The joint age-length key for the last three years (20182020) was applied. In general, the age distribution in the Norwegian fishery was shifted towards older fish compared to the Russian fishery. In the Russian catches fish at age 15-16 dominated, while in the Norwegian catches 16-17 years old. (Figure 6.5). The proportion (by numbers) of individuals at age 18 and older in the Norwegian catches was almost twice as large as in the Russian ones.

Age-length-keys for S. mentella are uncertain because of the slow growth rate of individuals and therefore these data should be used with caution. They were not used in the current assessment but may be considered in future assessments. Given that age is difficult to derive from length it is important that age readings are available for the most recent years, at the time of the working group.

### 6.2.3 Weight-at-age (Tables 6.12, 6.13, Figures 6.6, 6.7)

In earlier assessment, weight-at-age in the stock was set equal to the weight-at-age in the catch. This turned out to be problematic because of important fluctuations in reported weight-at-age in the catch that cannot be explained biologically (i.e. these are noisy data). In 2015, it was advised to either use a fixed weight-at-age for the $19+$ group, or use a modelled weight-at-age based on catch and survey records (Planque, 2015). The second option was chosen. Weight-at-age in the population was modelled for each year using mixed-effect models of a von Bertalanffy growth function (in weight). In 2018 an attempt was made to model weight-at-age for each cohort (rather than each year of observation). This showed that the growth function is nearly invariant between cohorts. Therefore, it was decided to use a fixed (i.e. common to all years) weight-at-age as input to the Statistical Catch-at-age model. The observed and modelled weight-at-age are presented in Table 6.12 as well as Figures 6.6 and 6.7.

### 6.2.4 Maturity-at-age (Table 6.14, Figure 6.8)

The proportion maturity-at-age was estimated for individual years using a mixed-effect statistical model (Table 6.14, Figure 6.8). The modelled values of maturity-at-age for individual years are used in the analytical assessment models, except in 2008, 2011 and 2016-2020 when the fixed effects only were considered, at least in the two latest years due to a lack of age data.

### 6.2.5 Natural mortality

In previous years, natural mortality for $S$. mentella was set to 0.05 for all ages and all years. This was based on life-history correlates presented in Hoenig (1983). Thirty-nine alternative mortality estimates were explored during the benchmark workshop, based on the review work by Kenchington (2014) and several additional recent papers (Then et al., 2014; Hamel, 2014; Charnov et al., 2013). Overall, the mode of these natural mortality estimates is 0.058 which departs only slightly from the original estimate of 0.050 (Figure 6.9). WKREDFISH 2018 decided to continue using 0.050 as the value of M in the assessment model. These estimates were updated for a peer-reviewed paper submitted in 2022 (Höffle and Planque, in revision) with 44 estimators resulting in a mode of the distribution of 0.07 .

Figure 6.10 shows cod's predation on juvenile ( $5-14 \mathrm{~cm}$ ) redfish during 1984-2020. This timeseries confirms the presence of redfish juveniles and may be used as an indicator of redfish abundance. A clear difference is seen between the abundance/consumption ratio in the 1980s and at present. A change in survey trawl catchability (smaller meshes) from 1993 onwards (Jakobsen et al., 1997) and/or a change in the cod's prey preference may cause this difference. As long as the trawl survey time-series has not been corrected for the change in catchability, the abundance index of juvenile redfish less than 15 cm during the 1980s might have been considerably higher, if this change in catchability had been corrected for. The decrease in the abundance of young redfish in the surveys during the 1990s is consistent with the decline in the consumption of redfish by cod. It is important that the estimation of the consumption of redfish by cod is being continued.

### 6.2.6 Scientific surveys

Following a dedicated review, AFWG approved the use of the new SToX versions of winter and ecosystem surveys for use in the S.s mentella assessment (WD 17 and WD 18 in AFWG 2020). The group recommended that the data be monitored annually to identify if a significant portion of the mentella stock moves east of the strata system. The group further recommended that work continues to investigate redfish-specific strata systems for the winter survey.

The results from the following research vessel survey series were evaluated by the Working Group:

### 6.2.6.1 Surveys in the Barents Sea and Svalbard area (Tables 1.1, 1.2, 6.15-6.18, Figures 6.11, 6.12)

Russian bottom trawl survey in the Svalbard and Barents Sea areas in October-December for 1978-2015 in fishing depths of 100-900 m (Table 6.15, Figure 6.11). ICES acronym: RU-BTr-Q4.

Russian-Norwegian Barents Sea 'Ecosystem survey' (bottom trawl survey, August-September) from 1986-2019 in fishing depths of 100-500 m (Figures 6.11-6.12). Data disaggregated by age for the period 1992-2019 (Tables 6.16b-6.17). ICES acronym: Since 2003 part of Eco-NoRu-Q3 (BTr), survey code: A5216.

Winter Barents Seabed-trawl survey (February) from 1986-2014 (jointly with Russia since 2000, except 2006 and 2007) in fishing depths of 100-500 m (Figures 6.11-6.12). Data disaggregated by age for the period 1992-2011 and 2013 (Table 6.18b). ICES acronym: BS-NoRu-Q1 (BTr), survey code: A6996.

The Norwegian survey initially designed for redfish and Greenland halibut is now part of the ecosystem survey and covers the Norwegian Economic Zone (NEZ) and Svalbard Fisheries Protection Zone incl. north and east of Spitsbergen during August 1996-2012 from less than 100 m to 800 m depth. This survey includes survey no. 2 above, and has been a joint survey with Russia since 2003, and since then called the Ecosystem survey. ICES acronym: Eco-NoRu-Q3 (Btr), survey code: A5216.

### 6.2.6.2 Pelagic survey in the Norwegian Sea (Table 6.19, Figures 6.13, 6.14)

The international deep pelagic ecosystem survey in the Norwegian Sea (WGIDEEPS, ICES 2016, survey code: A3357) monitors deep pelagic ecosystems, focusing on beaked redfish (S. mentella). The latest survey was conducted in the open Norwegian Sea from 11 August until 28 August 2019, following similar surveys in 2008, 2009, 2013 and 2016. The spatial coverage of the survey and the catch rates of beaked redfish in the trawl are presented in Figure 6.13. The survey is scheduled every third year. Estimated numbers-at-age from this survey were presented at the benchmark assessment in 2018 and used in the SCAA model. Data for 2016 was updated in 2019, using additional age readings and numbers-at-age for the 2019 survey were presented during AFWG 2020, used in the assessment and updated for AFWG 2021. The details of the data preparation, using StoX, are available from WD7 of AFWG 2018 (Planque et al., 2018). The data used as input to the analytical assessment consists of proportions-at-age from age 2 to 75 years (Figure 6.14).

### 6.2.6.3 Additional surveys (Figures 6.15-6.17)

The international 0-group survey in the Svalbard and Barents Sea areas in August-September 1980-2021, is now part of the Ecosystem survey (Figures 6.15 and 6.16). ICES acronym: Eco-NoRu-Q3 (Btr), survey code: A5216.

A slope survey "Egga-sør survey" was carried out by IMR from 07 March to 07 April 2020, following similar surveys in 2009, 2012, 2014, 2016 and 2018. The spatial coverage of the 2022 survey and the distribution of beaked redfish registered by acoustic is presented in Figure 6.17. EggaSør and Egga-Nord surveys operate on a biennial basis. The length and age distributions of beaked redfish from these surveys show consistent ageing in the population and gradual incoming of new cohorts after the recruitment failure period. These surveys are considered as candidates for data input to the analytical assessment of S. mentella (see also Planque, 2016).

### 6.3 Assessment

The group performed the analytical assessment using the statistical catch-at-age (SCAA) model reviewed at the benchmark in January 2018 (WKREDFISH, ICES 2018a). The model was configured as the benchmark baseline model which includes 53 parameters to be estimated and the model converged correctly.

### 6.3.1 Results of the assessment (Tables 6.20, 6.21, Figures 6.18-6.24)

### 6.3.1.1 Stock trends

The temporal patterns in recruitment-at-age 2 (Figures $6.18,6.21$ ) confirm the previously reported recruitment failure for the year classes 1996 to 2003 and indicate a return to high levels of recruitment. The estimates of year-class strength for recent years are uncertain due to limited age
data from winter and ecosystem surveys. Modelled spawning-stock biomass (SSB) has increased from 1992 to 2007 (Table 6.21). In the late 2000s the total-stock biomass (TSB) consisted of a larger proportion of mature fish than in the 1990s. This is reversing as individuals from new successful year classes, but still immature, are growing. TSB has increased from about 1.0 to above 1.4 million tonnes in the last 10 years (Table 6.21 and Figures 6.21-6.22). The concurrent decline in SSB from 2007 to 2014 can be attributed to the weak year classes (1996-2003) entering the mature stock. This trend has levelled off and SSB increases again. SSB at the start of 2021 is estimated at 900221 t.

### 6.3.1.2 Fishing mortality (Tables 6.20a,b-6.21, Figure 6.19)

The patterns of fleet selectivity-at-age indicate that most of the fish captured by the demersal fleet in 2020 are of age 8 years and older, while the pelagic fleet mostly captures fish of age 14 and older (Tables 6.20a,b and Figure 6.19). While model results at the benchmark workshop showed a gradual shift in the demersal selectivity towards older ages in recent years, this is no longer observed after the 2015 catch-at-age data were incorporated in the model. The demersal fleet selectivity appears shifted towards later ages only in 2014. In $2020 \mathrm{~F}_{19+}$ is estimated at 0.05 (Table 6.21), with 0.04 for the demersal and 0.008 for the pelagic fleets (Table 6.20a), respectively.

### 6.3.1.3 Survey selectivity patterns (Figure 6.20)

Winter and ecosystem surveys selectivity at age are very similar and show reduced selectivity for age 8 years and older, which is consistent with the known geographical distribution of different life stages of $S$. mentella (Figure 6.20). Conversely, the Russian survey shows a reduced selectivity for age 7 years and younger. This is believed to result from gear selectivity.

### 6.3.1.4 Residual patterns (Figure 6.23)

Residual patterns in catch and survey indices are presented in Figure 6.23a-e. There is generally no visible trend in the residuals for the Russian groundfish survey neither by age nor by year. Trends in residuals are visible in recent years for winter and ecosystem surveys and will need to be investigated further. Alternative methods for the estimation of the survey selectivity patterns will be investigated in the benchmark assessment planned for 2024 and could resolve the issue. Residual patterns for the demersal fleet indicate a similar fit of the model compared to AFWG 2018, when a time varying selectivity-at-age for this fleet was introduced.

### 6.3.1.5 Retrospective patterns (Figure 6.24)

The historical retrospective patterns for the years 2007 to 2016 are presented in Figure 6.24. All model parameters were estimated in each individual run. The most recent model run (last year of data 2020) is consistent with previous runs. As in 2018 the SSB time-series is smoother than before, due to fixed weight-at-age for every year. The new estimates for winter and Ecosystem surveys in 2020 led to an increase in estimated SSB, up to $19 \%$ in the early years and around $7 \%$ to $9 \%$ in later years. Contrarily, the 2021 update revised SSB moderately down, by about $5 \%$ to $6 \%$. Retrospective bias (Mohn's rho) over the last 5 assessments was $-48 \%$ for recruitment, $-2 \%$ for $F(19+)$ and $+7 \%$ for SSB. The benchmark run stands out and this is due to the unavailability of recent catch-at-age data during the benchmark assessment (see section 6.2.2).

### 6.3.1.6 Projections

Fmsy at age 19+ is approximated using $\mathrm{F}_{0.1}$ and estimated at 0.084 (section 1.4 of the WKREBMSE report 2018b).

The estimated fishing mortality in 2020 is: $\mathrm{F}_{19+}=0.05$.

If the fishing mortality is maintained, this is expected to lead to a catch of 57743 t in 2021, well below the advised TAC of 66158 t . This would lead to an SSB of 925932 t in early 2022, catches of 59466 t in 2022 and SSB of 955688 t in 2023.

Raising $\mathrm{F}_{19+}$ to the precautionary approach ( $\mathrm{F}_{19+}=0.06$ ), recommended in the latest advice, in 2022-2024 would lead to average catches of 72263 t during that period and a SSB of 999340 t by 2025 (SSB at the start of 2020 is estimated at 874727 t ).

These projections assume that the selectivity patterns of the demersal and pelagic fleets are identical with those estimated for 2019. It is also assumed that the ratio of fishing mortality between these two fleets remains unchanged.

### 6.3.1.7 Additional considerations

Historical fluctuations in the recruitment-at-age 2 (Figures 6.18 and 6.21) are consistent with the 0-group survey index (Figure 6.16), although the 0-group survey index is not used as an input to the SCAA.

The population age structure derived from the model outputs for the old individuals (beyond $19+$, Figure 6.22) is consistent with the age structure reported from the slopes surveys although these are not yet used as input to the model.

Recent recruitment levels estimated with SCAA are highly uncertain since they rely on only few years of observations and since the age readings from winter survey were not available for years 2014-2021. The use of the autoregressive model for recruitment (random effects in the SCAA) which was introduced in this assessment allows for a projection of the recruitment in recent years, despite the current lack of age data.

### 6.3.1.8 Assessment summary (Table 6.21, Figure 6.21)

The history of the stock as described by the SCAA model for the period 1992-2019 is summarized in Table 6.21 and Figure 6.21. The key elements are as follows:

- upward trend in Total-stock biomass from 1992 to 2006 followed by stabilization until 2011 and a new upward trend until the present,
- upward trend in spawning-stock biomass from 1992 to 2007 followed by stabilization (or slight decline) until 2014 and subsequent increase,
- recruitment failure for year classes 1996-2003 (2y old fish in 1998-2005),
- good (although uncertain) recruitment for year classes born after 2005. Age data for recruits (at age $2 y$ ) after 2014 is limited.
- Annual fishing mortality for the $19+$ group throughout the assessment period varied between 0.003 and 0.05 .


### 6.4 Comments to the assessment

Currently, the survey series used in the SCAA do not appropriately cover the geographical distribution of the adult population. Data from the pelagic survey in the Norwegian Sea has been reviewed in the last benchmark and is now included in the assessment model. Priority should be given to including additional data from the slope surveys that include older age groups, in the analytical assessment in future (WD 5 in 2016).

The SCAA model relies on the availability of reliable age data in surveys and in the catch. Although additional age reading since the last assessment has improved reliability, it requires a continuous effort to keep these data at an appropriate level.

### 6.5 Biological reference points

The proposed reference points estimated during the workshop on the management plan for $S$. mentella in (ICES 2018b) were:

| Reference point | Value |
| :--- | :--- |
| $\mathrm{B}_{\mathrm{lim}}$ | 227000 t |
| $\mathrm{B}_{\mathrm{pa}}$ | 315000 t |
| $\mathrm{F}_{\mathrm{MYY19+}}=\mathrm{F}_{0.1}$ | 0.084 |

Which are revised from those set during the benchmark in the same year (ICES 2018a) which were $B_{p a}=450 \mathrm{kt}, \mathrm{B}_{\lim }=324 \mathrm{kt}$ and $\mathrm{F}_{\mathrm{MSY} 19+}=\mathrm{F}_{0.1}=0.08$.

### 6.6 Management advice

The present report neither assesses the stock nor does it give advice.

### 6.7 Possible future development of the assessment

Many developments suggested in earlier years were presented and evaluated at the benchmark in January 2018. These include integrating a stochastic process model i) for recruitment-at-age 2, ii) for the annual component of fishing mortalities, and iii) to account for annual changes in fleet selectivities-at-age. In addition, iv) a right trapezoid population matrix, v) coding of older ages into flexible predefined age-blocks, and vi) integrating of data from pelagic surveys in the Norwegian Sea were implemented. The purpose of these new features was to reduce the number of parameters to estimate ( $\mathrm{i}, \mathrm{ii}$ ), include new data on the older age fraction of the population (iv, v , vi) and account for possible temporal changes in selectivity linked to changes in the national and international fisheries and their regulations (iii).

Recommendations that have been followed since comprise:

- An increase in the number of age readings from surveys and from the fishery, particularly for recent years.
- Use of a standardized method (StoX) for the determination of numbers-at-age in the surveys. The use of StoX for survey indices was evaluated at the beginning of AFWG 2020.
Future developments for the assessment of $S$. mentella may possibly include:
- Use of a standardized method (ECA) for the determination of numbers-at-age in the catch.
- A genetic-based method for rapidly identifying Sebastes species (S. norvegicus, S. mentella, S. viviparus);
- Direct use of length information (as in GADGET);
- Development of a joint age-length key for calculation of age composition of all S. mentella catches.
- Development of a joint model for S. mentella and S. norvegicus which can include uncertainty in species identification and reporting of catch of Sebastes $s p$.

Implementing the current model in a more generic framework (SAM or XSAM) would provide a set of diagnostic tools and the wider expertise shared by the groups developing these models. The new version of GADGET, running the currently used TMB-package in the background, may provide an opportunity to put both species on the same platform.

Further studies of redfish mortality at young age, including a scientific publication, should be carried out. These studies should also take account of historic estimates of bycatch. Variable M by age and possibly time period could then be incorporated in the assessment.

## 6．8 Tables and figures

Table 6．1．$S$ ．mentella in subareas 1 and 2．Nominal catch（ t ）by countries in Subarea 1，divisions 2．a and 2．b combined．

|  |  |  |  | 凹10 |  |  |  |  | － |  | $\begin{aligned} & \text { n } \\ & \frac{1}{0} \\ & \frac{1}{0} \\ & 0 \\ & \frac{1}{2} \end{aligned}$ | $\begin{aligned} & \text { 㐅} \\ & \text { n } \\ & 0 \\ & \text { Z } \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \frac{\pi}{O} \\ & \hline \end{aligned}$ | $\overline{0}$ 0 0 0 0 | $\begin{aligned} & \stackrel{\pi}{\omega} \\ & \underset{\sim}{3} \end{aligned}$ | $\begin{aligned} & \text { 드̄ } \\ & \text { ĩ } \end{aligned}$ | $\underset{ }{〕}$ | Ј |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 |  | － | 20 | 73 | 100 | 14 | － | 9 | － | － | － | 9733 | 13 | 125 | 3646 | 177 | 134 | 14045 |
| 1999 |  | － | 73 | 26 | 202 | 50 | － | 3 | － | － | － | 7884 | 6 | 65 | 2731 | 29 | 140 | 11209 |
| 2000 |  | － | 50 | 12 | 62 | 29 | 48 | 1 | － | － | － | 6020 | 2 | 115 | 3519 | 87 | 130 | 10075 |
| 2001 |  | － | 74 | 16 | 198 | 17 | 3 | 4 | － | － | － | 13937 | 5 | 179 | 3775 | 90 | 120 | 18418 |
| 2002 |  | 15 | 75 | 58 | 99 | 18 | 41 | 4 | － | － | － | 2152 | 8 | 242 | 3904 | 190 | 188 | 6993 |
| 2003 |  | － | 64 | 22 | 32 | 8 | 5 | 5 | － | － | － | 1210 | 7 | 44 | 952 | 47 | 124 | 2520 |
| 2004 | Sweden－1 | － | 588 | 13 | 10 | 4 | 10 | 3 | － | － | － | 1375 | 42 | 235 | 2879 | 257 | 76 | 5493 |
| 2005 |  | 5 | 1147 | 46 | 33 | 39 | 4 | 4 | － | － | 7 | 1760 | － | 140 | 5023 | 163 | 95 | 8465 |
| 2006 | Canada－ 433 | 396 | 3808 | 215 | 2483 | 63 | 2513 | 4 | 341 | 845 | － | 4710 | 2496 | 1804 | 11413 | 710 | 1027 | 33261 |
| 2007 |  | 684 | 2197 | 234 | 520 | 29 | 1587 | 17 | 349 | 785 | － | 3209 | 1081 | 1483 | 5660 | 2181 | 202 | 20219 |
| 2008 |  | － | 1849 | 187 | 16 | 25 | 9 | 9 | 267 | 117 | 13 | 2220 | 8 | 713 | 7117 | 463 | 83 | 13096 |
| 2009 | EU－ 889 | － | 1343 | 15 | 42 | － | 33 | － | － | － | 3 | 2677 | 338 | 806 | 3843 | 177 | 80 | 10246 |
| 2010 |  | － | 979 | 175 | 21 | 12 | 2 | － | 243 | 457 | － | 2065 | － | 293 | 6414 | 1184 | 79 | 11924 |
| 2011 |  | － | 984 | 175 | 835 | － | 2 | － | 536 | 565 | － | 2471 | 11 | 613 | 5037 | 1678 | 55 | 12962 |


|  | $\begin{aligned} & \text { N} \\ & \stackrel{0}{\overleftarrow{0}} \\ & \text { H } \end{aligned}$ | $\begin{aligned} & \text { n } \\ & \text { C } \\ & \text { त } \\ & \text { ज } \\ & \text { Din } \\ & \stackrel{0}{4} \end{aligned}$ |  |  |  |  | $\begin{aligned} & \text { D } \\ & \text { 프N} \end{aligned}$ | $\sum_{\text {何 }}^{0}$ |  |  | $$ | $\begin{aligned} & \text { 들 } \\ & \frac{\pi}{0} \end{aligned}$ |  | $\begin{aligned} & \stackrel{\pi}{\omega} \\ & \stackrel{y}{3} \end{aligned}$ |  | $\underset{J}{3}$ | - |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | - | 259 | - | 517 | - | 36 | - | 447 | 449 | - | 2114 | 318 | 1038 | 4101 | 1780 | - | 11059 |
| 2013 | - | 697 | - | 80 | 21 | 1 | - | 280 | 262 | - | 1750 | 84 | 1078 | 3677 | 1459 | - | 9389 |
| 2014 | - | 743 | 215 | 446 | 15 | - | - | 215 | 167 | 3 | 13149 | 103 | 505 | 1704 | 1162 | - | 18426 |
| 2015 | - | 657 | 49 | 242 | 48 | 3 | - | 537 | 192 | 3 | 19433 | 5 | 678 | 1142 | 2529 | 52 | 25570 |
| 2016 | - | 502 | 134 | 493 | 74 | 24 | 0 | 1243 | 1065 | - | 18191 | 208 | 1066 | 8419 | 3213 | 122 | 34754 |
| 2017 | 4 | 443 | 45 | 763 | 66 | 3 | - | 562 | 790 | - | 17077 | 102 | 1060 | 6593 | 2838 | 436 | 30783 |
| 2018 | - | 425 | 67 | 2473 | 82 | 10 | - | 1020 | 1010 | 374 | 18594 | 275 | 699 | 10497 | 2457 | 63 | 38046 |
| 2019 | - | 156 | 370 | 1599 | 615 | 10 | - | - | 653 | 244 | 23844 | 471 | 1422 | 13444 | 2222 | 590 | 45640 |
| 2020 | - | 149 | 163 | 1807 | 62 | 5 | - | 2 | 1081 | 1483 | 32950 | 4 | 870 | 13874 | 744 | 437 | 53631 |
| $2021{ }^{1}$ | - | 290 | 218 | 1166 | 85 | 6 | - | - | 1379 | - | 43797 | 2 | 381 | 14887 | 615 | 655 | 63482 |

1-Provisional figures.
Table 6.2. S. mentella in subareas 1 and 2. Nominal catch ( $\mathbf{t}$ ) by countries in Subarea 1.

|  |  | 凹 |  |  |  |  |  |  |  |  |  | $\begin{aligned} & \text { त } \\ & \text { 3 } \\ & \text { 30 } \end{aligned}$ | $\begin{aligned} & \text { ס } \\ & \frac{1}{\Pi} \\ & \mathbf{0} \end{aligned}$ |  | T0 0 00 0 0 |  | $\begin{aligned} & \underset{\sim}{n} \\ & \\ & \end{aligned}$ | ¢ in in |  | $\underset{ }{\text { J }}$ |  | - |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 20 |  | - |  | - |  | - |  | - |  | - | 26 |  | - |  | - | 378 |  | - |  | - | 424 |
| 1999 | 69 |  | - |  | - |  | - |  | - |  | - | 69 |  | - |  | - | 489 |  | - |  | - | 627 |
| 2000 | - |  | - |  | - |  | - |  | 48 |  | - | 47 |  | - |  | - | 406 |  | - |  | - | 501 |


|  |  |  | 凹 |  | त त ® Ej0 U |  |  |  | O ¢ ¢ U |  |  |  | त 3 3 0 |  |  |  | $\overline{0}$ 00 it ¢ |  |  | ¢ in in |  | $\underset{ }{〕}$ | - |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 |  | - |  | - |  | - |  | - |  | 3 |  | - |  | 8 |  | - |  | - | 296 |  | - | - | 307 |
| 2002 |  | - |  | - |  | - |  | - |  | - |  | - |  | 4 |  | - |  | - | 587 |  | - | - | 591 |
| 2003 |  | - |  | - |  | - |  | - |  | - |  | - |  | 6 |  | - |  | - | 292 |  | - | - | 298 |
| 2004 |  | - |  | - |  | - |  | - |  | - |  | - |  | 2 |  | - |  | - | 355 |  | - | - | 357 |
| 2005 |  | - |  | - |  | - |  | - |  | - |  | - |  | 3 |  | - |  | - | 327 |  | - | - | 330 |
| 2006 |  | 2 |  | - |  | - |  | - |  | - |  | - |  | 12 |  | - |  | - | 460 |  | - | 2 | 476 |
| 2007 |  | - |  | - |  | - |  | - |  | 8 |  | - |  | 11 |  | - |  | - | 210 |  | - | 20 | 249 |
| 2008 |  | - |  | - |  | - |  | - |  | - |  | - |  | 5 |  | - |  | - | 155 |  | - | 2 | 162 |
| 2009 |  | - |  | - |  | - |  | - |  | 8 |  | - |  | 3 |  | - |  | - | 80 |  | - | - | 91 |
| 2010 |  | - |  | - |  | - |  | - |  | - |  | - |  | 20 |  | - |  | - | 10 |  | - | - | 30 |
| 2011 |  | - |  | - |  | - |  | - |  | - |  | - |  | 48 |  | - |  | - | 13 |  | - | - | 61 |
| 2012 |  | - |  | - |  | - |  | - |  | - |  | - |  | 34 |  | - |  | - | 17 |  | - | - | 51 |
| 2013 |  | - |  | - |  | - |  | - |  | - |  | - |  | 64 |  | - |  | - | 27 |  | - | - | 91 |
| 2014 |  | - |  | - |  | - |  | - |  | - |  | - |  | 159 |  | - |  | - | 63 |  | - | - | 222 |
| 2015 |  | - |  | - |  | - |  | 18 |  | - |  | - |  | 138 |  | 1 |  | - | 125 |  | - | - | 282 |
| 2016 |  | - |  | - |  | - |  | - |  | - |  | - |  | 225 |  | 1 |  | - | 229 |  | 342 | - | 797 |
| 2017 |  | - |  | - |  | - |  | 12 |  | - |  | - |  | 207 |  | 3 |  | - | 196 |  | - | - | 418 |


|  |  | 凹 |  |  | ¢ ¢ ¢ ¢ |  |  |  | $\begin{aligned} & \text { त } \\ & \text { 3 } \\ & \text { 30 } \end{aligned}$ | $\begin{aligned} & \text { 듬 } \\ & \text { त्व } \end{aligned}$ |  |  |  | $\begin{aligned} & \stackrel{\pi}{n} \\ & \stackrel{y}{x} \\ & \end{aligned}$ | ¢ in in | $\underset{ }{〕}$ | － |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2018 | － | － | 19 | 26 |  | 3 |  | － | 255 |  | － |  | － | 376 | － | － | 679 |
| 2019 | 83 | 4 | － | 13 |  | － |  | 1 | 369 |  | 16 |  | 1 | 206 | 19 | 4 | 715 |
| 2020 | 35 | 12 | 6 | 18 |  | 1 |  | － | 335 |  | 3 |  | 2 | 118 | 1 | － | 532 |
| $2021{ }^{1}$ | 87 | 31 | － | 14 |  | － |  | － | 195 |  | － |  | 4 | 367 | 1 | － | 699 |

1 －Provisional figures．

Table 6．3．S．mentella in subareas 1 and 2．Nominal catch（ t ）by countries in Division 2．a（including landings from the pelagic trawl fishery in the international waters）．

|  |  |  | 凹 |  |  |  | $\begin{aligned} & \text { 들 } \\ & \text { N } \\ & \underline{\underline{0}} \end{aligned}$ |  |  |  |  | $\begin{aligned} & \text { 㐅} \\ & \text { 3 } \\ & \text { 30 } \\ & \mathbf{2} \end{aligned}$ | $\begin{aligned} & \overline{0} \\ & 00 \\ & \text { t. } \\ & 0.0 \end{aligned}$ |  |  | $\underset{\substack{\text { 苟 } \\ \underset{\sim}{x}}}{ }$ |  | $\underset{ }{〕}$ | $\stackrel{\square}{\square}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 |  | － | 73 | 58 | 14 | － | 6 |  | － |  | － | 9186 | 118 |  | － | 2626 | 55 | 106 | 12242 |
| 1999 |  | － | 16 | 160 | 50 | － | 3 |  | － |  | － | 7358 | 56 |  | － | 1340 | 14 | 120 | 9117 |
| 2000 |  | 50 | 11 | 35 | 29 | － | － |  | － |  | － | 5892 | 98 |  | － | 2167 | 18 | 103 | 8403 |
| 2001 |  | 63 | 12 | 161 | 17 | － | 4 |  | － |  | － | 13636 | 105 |  | － | 2716 | 18 | 95 | 16827 |
| 2002 |  | 37 | 54 | 59 | 18 | 41 | 4 |  | － |  | － | 1937 | 124 |  | － | 2615 | 8 | 157 | 5054 |
| 2003 |  | 58 | 18 | 17 | 8 | 5 | 5 |  | － |  | － | 1014 | 17 |  | － | 448 | 8 | 102 | 1700 |
| 2004 | Sweden－1 | 555 | 8 | 4 | 4 | 10 | 3 |  | － |  | － | 987 | 86 |  | － | 2081 | 7 | 18 | 3764 |
| 2005 |  | 1101 | 36 | 17 | 38 | 2 | 4 |  | － |  | － | 1083 | 71 |  | － | 3307 | 20 | 15 | 5694 |


| $\begin{aligned} & \text { 㐫 } \\ & \end{aligned}$ |  |  | $\begin{aligned} & \text { 쓴 } \\ & \text { 든 } \end{aligned}$ |  |  | $\begin{aligned} & \text { 들 } \\ & \underline{\pi} \\ & \underline{U} \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \text { N } \\ & \underline{\underline{\omega}} \end{aligned}$ |  | 年 | त 3 3 0 0 |  |  | $\begin{aligned} & \underset{\sim}{n} \\ & \underset{\sim}{2} \end{aligned}$ | $\begin{aligned} & \frac{.5}{\overline{0}} \\ & \text { in } \end{aligned}$ | こ | $\stackrel{\text { ¢̈ }}{\square}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | Estonia－ 396 <br> Canada－ 433 | 3793 | 199 | 2475 | 52 | 2513 | 3 | 845 | － | 4010 | 1731 | 2467 | 10110 | 589 | 958 | 30574 |
| 2007 | Estonia－684 | 2157 | 226 | 519 | 29 | 1579 | 16 | 785 | 349 | 3043 | 1395 | 1079 | 5061 | 2159 | 120 | 19201 |
| 2008 | Netherlands－ 13 | 1821 | 179 | 9 | 24 | 9 | 9 | 117 | 267 | 1952 | 666 | 1 | 6442 | 430 | 62 | 12001 |
| 2009 | EU－889 | 1316 | 7 | 23 | － | 25 | － | － | － | 2208 | 764 | 338 | 3305 | 137 | 62 | 9074 |
| 2010 |  | 961 | 175 | 13 | 12 | 2 | － | 457 | 243 | 1705 | 246 | － | 5903 | 1183 | 55 | 10955 |
| 2011 |  | 932 | 175 | 697 | － | 2 | － | 561 | 536 | 1682 | 599 | － | 4326 | 1656 | 19 | 11185 |
| 2012 |  | 259 | － | 469 | － | 32 | － | 449 | 447 | 1500 | 1038 | 311 | 3478 | 1770 | － | 9753 |
| 2013 | NL | 675 | － | 24 | 21 | 1 | － | 262 | 280 | 871 | 1055 | 68 | 3293 | 1435 | － | 7985 |
| 2014 | 2 | 728 | 209 | 411 | 15 | － | － | 167 | 215 | 4089 | 505 | 100 | 1334 | 1159 | － | 8934 |
| 2015 | 3 | 657 | 49 | 236 | 25 | 3 | － | 192 | 537 | 11410 | 678 | 3 | 480 | 2508 | 47 | 16828 |
| 2016 |  | 495 | 107 | 493 | 61 | － | 24 | 1065 | 1243 | 8887 | 1052 | 183 | 3949 | 2862 | 71 | 20492 |
| 2017 |  | 425 | 38 | 763 | 44 | 3 | － | 790 | 562 | 7348 | 1059 | 94 | 3922 | 2813 | 429 | 18287 |
| 2018 | 374 | 400 | 47 | 2440 | 51 | 7 | － | 1010 | 876 | 14057 | 699 | 272 | 4721 | 2435 | 62 | 27451 |
| 2019 | 244 | 73 | 363 | 1599 | 59 | 10 | － | 652 | － | 17741 | 1421 | 455 | 7366 | 2184 | 569 | 32736 |
| 2020 | 1483 | 112 | 146 | 1797 | 41 | 4 | － | 1081 | － | 22854 | 868 | － | 6085 | 737 | 403 | 35613 |
| $2021{ }^{1}$ | － | 151 | 182 | 1128 | 70 | 6 | － | 1379 | － | 35799 | 377 | － | 6008 | 535 | 552 | 46187 |

1－Provisional figures．

## Table 6.4. S. mentella in subareas 1 and 2. Nominal catch ( $\mathbf{t}$ ) by countries in Division 2.b.

|  |  |  |  |  |  | 凹 |  |  |  |  | त 3 30 2 | - | ¢0 0 0 0 0 | $\begin{aligned} & \underset{\sim}{\omega} \\ & \underset{\sim}{3} \end{aligned}$ | $\begin{aligned} & \text {.드드N } \\ & \text { ĩ } \end{aligned}$ |  | $\stackrel{\text { V }}{ }$ | $\stackrel{\overline{\text { ® }}}{ }$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 |  |  | - |  | - | - | 42 |  | - | 3 | 521 | 13 | 7 | 642 | 122 | - | 29 | 1379 |
| 1999 |  |  | - |  | 4 | 10 | 42 |  | - | - | 457 | 6 | 9 | 902 | 15 | - | 20 | 1465 |
| 2000 |  |  | - |  | - | 1 | 27 |  | - | 1 | 82 | 2 | 17 | 946 | 69 | - | 27 | 1172 |
| 2001 |  |  | - |  | 11 | 4 | 37 |  | - | - | 293 | 5 | 74 | 763 | 72 | Estonia | 25 | 1284 |
| 2002 |  |  | - |  | 38 | 4 | 40 |  | - | - | 210 | 8 | 118 | 702 | 182 | 15 | 31 | 1348 |
| 2003 |  |  | - |  | 6 | 4 | 15 |  | - | - | 190 | 7 | 27 | 212 | 39 | - | 22 | 522 |
| 2004 |  |  | - |  | 33 | 5 | 6 |  | - | - | 386 | 42 | 149 | 443 | 250 | - | 58 | 1372 |
| 2005 | Iceland-2 |  | 7 |  | 46 | 10 | 17 |  | 1 | - | 673 | - | 69 | 1389 | 143 | 5 | 80 | 2442 |
| 2006 |  |  | - |  | 13 | 16 | 8 |  | 11 | 1 | 688 | 29 | 73 | 843 | 121 | - | 67 | 1870 |
| 2007 |  |  | - |  | 40 | 8 | 1 |  | - | 1 | 155 | 2 | 88 | 389 | 22 | - | 62 | 768 |
| 2008 |  |  | - |  | 28 | 8 | 7 |  | 1 | - | 263 | 6 | 47 | 520 | 33 | - | 19 | 932 |
| 2009 | Canada - 3 |  | 3 |  | 27 | 8 | 19 |  | - | - | 466 | 1 | 42 | 458 | 41 | - | 17 | 1085 |
| 2010 |  |  | - |  | 18 | - | 8 |  | - | - | 339 | - | 47 | 501 | 1 | - | 24 | 938 |
| 2011 | LT-4 |  | - |  | 52 | - | 139 |  | - | - | 741 | 11 | 14 | 698 | 23 | - | 36 | 1717 |
| 2012 | Iceland-4 |  | - |  | - | - | 48 |  | - | - | 581 | 7 | - | 606 | 10 | - | - | 1256 |
| 2013 |  |  | - |  | 22 | - | 56 |  | - | - | 815 | 16 | 23 | 357 | 23 | - | - | 1312 |


|  |  |  |  |  |  | 凹 |  |  |  |  |  | त 3 3 2 | 즈﹎ त O |  | $\begin{aligned} & \stackrel{\pi}{\omega} \\ & \stackrel{y}{\omega} \\ & \underset{x}{2} \end{aligned}$ |  |  |  | $\underset{J}{ }$ | － |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2014 |  |  | 1 |  | 15 | 6 | 34 |  | － |  | － | 8901 | 3 | － | 307 | 3 |  | － | － | 9270 |
| 2015 |  |  | － |  | － | － | 6 |  | 5 |  | － | 7885 | 1 | － | 536 | 21 |  | － | 5 | 8459 |
| 2016 |  |  | － |  | 7 | 27 | － |  | 14 |  | － | 9078 | 24 | 14 | 4241 | 9 |  | － | 50 | 13464 |
| 2017 |  |  | － |  | 18 | 7 | 1 |  | 10 |  | － | 9522 | 5 | 1 | 2476 | 25 |  | 4 | 7 | 12075 |
| 2018 | LT－144 |  | － |  | 25 | 20 | 14 |  | 6 |  | － | 4281 | 3 | － | 5400 | 22 |  | － | 1 | 9915 |
| 2019 |  |  | － |  | － | 4 | － |  | 543 |  | － | 5734 | － | － | 5873 | 19 |  | － | 17 | 12190 |
| $2020^{1}$ | LV－2 |  | － |  | 2 | 5 | 4 |  | 2 |  | － | 9760 | － | － | 7671 | 6 |  | － | 34 | 17486 |
| $2021{ }^{1}$ |  |  | － |  | 52 | 6 | 38 |  | 1 |  |  | 7803 | 2 | － | 8512 | 79 |  | 1 | 103 | 16596 |

1 －Provisional figures．

Table 6．5．S．mentella in subareas 1 and 2．Nominal catch（ $\mathbf{t}$ ）by countries of the pelagic fishery in international waters of the Norwegian Sea（see text for further details）．

| $\begin{aligned} & \text { 厄 } \\ & \text { ভ兀 } \end{aligned}$ |  | $\begin{aligned} & . \frac{0}{ً} \\ & \text { D} \\ & \text { H } \\ & \hline \end{aligned}$ |  |  |  | O ¢ ¢ U | $\sum_{\substack{0}}^{0}$ |  | त 3 3 0 0 | O ¢ त O | ¢0 0 00 0 0 0 |  | ¢ n in | $\underset{ }{\beth}$ | $\stackrel{\square}{\square}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 |  | － | － | － | 9 | － | － | － | － | － | － | － | － | － | 9 |
| 2003 |  | － | － | － | 40 | － | － | － | － | － | － | － | － | － | 40 |
| 2004 |  | － | 500 | － | 2 | － | － | － | － | － | － | 1510 | － | － | 2012 |
| 2005 |  | － | 1083 | － | 20 | － | － | － | － | － | － | 3299 | － | － | 4402 |
| 2006 | CAN－ 433 | 396 | 3766 | 192 | 2475 | 2510 | 341 | 845 | 2862 | 2447 | 1697 | 9390 | 575 | 841 | 28770 |


|  |  | . | $\begin{aligned} & \text { n } \\ & \frac{C}{C} \\ & \frac{\pi}{ज} \\ & \underline{0} \\ & \frac{0}{0} \\ & \stackrel{0}{4} \end{aligned}$ | 凹 |  | O O ¢ U | $\sum_{\substack{0}}^{0}$ |  | त 3 3 ¢ | $\begin{aligned} & \text { ס } \\ & \frac{\overline{1}}{0} \\ & \mathbf{0} \end{aligned}$ | N 0 0 年 0 | $\begin{aligned} & \underset{\sim}{\omega} \\ & \stackrel{y}{\hat{a}} \end{aligned}$ | $\begin{aligned} & \text { 드제 } \\ & \text { in } \end{aligned}$ | $\underset{ }{\text { J }}$ | $\stackrel{\overline{0}}{\square}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 |  | 684 | 1968 | 226 | 497 | 1579 | 349 | 785 | 1813 | 1079 | 1377 | 3645 | 2155 | - | 16157 |
| 2008 |  | - | 1797 | - | - | - | 267 | 117 | 330 | - | 641 | 4901 | 390 | - | 8443 |
| 2009 | EU-889 | - | 1253 | - | - | - | - | - | - | 337 | 701 | 1975 | 135 | - | 5290 |
| 2010 |  | - | 912 | - | - | - | 243 | 457 | 450 | - | 244 | 5103 | 820 | - | 8229 |
| 2011 |  | - | 740 | 175 | 693 | - | 536 | 561 | 342 | - | 595 | 3621 | 1648 | - | 8911 |
| 2012 |  | - | 259 | - | 469 | 31 | 447 | 449 | - | 311 | 1038 | 2714 | 1768 | - | 7486 |
| 2013 |  | 8 | 675 | - | - | - | 280 | 262 | 1 | 68 | 1078 | 2720 | 1435 | - | 6527 |
| 2014 |  | - | 697 | - | 409 | - | 215 | 167 | - | 100 | 505 | 795 | 1146 | - | 4034 |
| 2015 |  | - | 606 | - | 231 | - | 537 | 192 | - | - | 678 | - | 2508 | - | 4752 |
| 2016 |  | - | 393 | - | 493 | - | 1243 | 1065 | 9 | - | 821 | 512 | 2862 | - | 7398 |
| 2017 | NL | - | 296 | - | 761 | - | 562 | 790 | - | 14 | 791 | 1014 | 2624 | - | 6852 |
| 2018 | 374 | - | 400 | - | 2192 | - | 876 | 1010 | - | 116 | 372 | - | 2399 | - | 7739 |
| 2019 | 244 | Greenland | - | 298 | 1157 | - | - | 652 | 1 | 364 | 1096 | 117 | 1908 | 223 | 6060 |
| 2020 | 1366 | 3 | - | 73 | 1380 | - | - | 1081 | - | - | 480 | 25 | 737 | 324 | 5469 |
| $2021{ }^{1}$ | - | - | - | 117 | 514 | - | - | 1379 | - | - | 84 | 498 | 280 | - | 2872 |

1-Provisional figures.

## Table 6.6. REDFISH in subareas 1 and 2. Nominal catch ( t ) by countries in Subarea 1, divisions 2.a and 2.b combined for both S. mentella and S. norvegicus.

|  | $\sum_{\substack{0}}^{\text {T }}$ |  |  |  | 凹 | $\begin{aligned} & \text { Z } \\ & \text { İ } \\ & \text { E } \\ & \text { © } \end{aligned}$ |  | ¢ ¢ ¢ U |  | $\begin{aligned} & \text { 들 } \\ & \text { 끌 } \end{aligned}$ |  |  | $\begin{aligned} & \text { त } \\ & \text { 30 } \\ & \text { Z } \end{aligned}$ |  |  | $\overline{0}$ 0 0 0 0 | $\begin{aligned} & \stackrel{\pi}{\omega} \\ & \stackrel{\pi}{n} \\ & \underset{\sim}{2} \end{aligned}$ | $\begin{aligned} & \stackrel{\subseteq}{\pi} \\ & \text { in } \end{aligned}$ | $\begin{aligned} & \underset{\substack{3 \\ \underset{~}{\sim} \\ \underset{y}{c} \\ \hline}}{ } \end{aligned}$ | $\begin{aligned} & \overline{\ddot{O}} \\ & \dot{W} \\ & \underset{y}{y} \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | - | - | - | - | 2970 | 7457 | - | - | - | - |  | - | 18650 |  | - | 1806 | 69689 | 25 | 716 | - | 101313 |
| 1985 | - | - | - | - | 3326 | 6566 | - | - | - | - |  | - | 20456 |  | - | 2056 | 59943 | 38 | 167 | - | 92552 |
| 1986 | - | DK | - | 29 | 2719 | 4884 | - | - | - | - |  | - | 23255 |  | - | 1591 | 20694 | - | 129 | 14 | 53315 |
| 1987 | - | + | - | $450^{3}$ | 1611 | 5829 | - | - | - | - |  | - | 18051 |  | - | 1175 | 7215 | 25 | 230 | 9 | 34595 |
| 1988 | - | - | - | 973 | 3349 | 2355 | - | - | - | - |  | - | 24662 |  | - | 500 | 9139 | 26 | 468 | 2 | 41494 |
| 1989 | - | - | - | 338 | 1849 | 4245 | - | - | - | - |  | - | 25295 |  | - | 340 | 14344 | $5^{2}$ | 271 | 1 | 46688 |
| 1990 | - | $37^{3}$ | - | 386 | 1821 | 6741 | - | - | - | - |  | - | 34090 |  | - | 830 | 18918 | - | 333 | - | 63156 |
| 1991 | - | 23 | - | 639 | 791 | 981 | - | - | - | - |  | - | 49463 |  | - | 166 | 15354 | 1 | 336 | 13 | 67768 |
| 1992 | CAN | 9 | - | 58 | 1301 | 530 | 614 | - | - | - |  | - | 23451 |  | - | 977 | 4335 | 16 | 479 | 3 | 31773 |
| 1993 | $8^{3}$ | 4 | - | 152 | 921 | 685 | 15 | - | - | - |  | - | 18319 |  | - | 1040 | 7573 | 13 | 734 | 1 | 29465 |
| 1994 | - | 28 | - | 26 | 771 | 1026 | 6 | 4 | 4 | 3 |  | - | 21466 |  | - | 985 | 6220 | 34 | 259 | 13 | 30841 |
| 1995 | - | - | - | 30 | 748 | 693 | 7 | 1 | 1 | 5 |  | 1 | 16162 |  | - | 936 | 6985 | 67 | 252 | 13 | 25900 |
| 1996 | - | - | - | $42^{3}$ | 746 | 618 | 37 | - | - | 2 |  | - | 21675 |  | - | 522 | 1641 | 409 | 305 | 121 | 26118 |
| 1997 | - | - | - | 7 | 1011 | 538 | $39^{2}$ | - | - | 11 |  | - | 18839 |  | 1 | 535 | 4556 | 308 | 235 | 29 | 26109 |
| 1998 | - | - | - | 98 | 567 | 231 | $47^{3}$ | - | - | 28 |  | - | 26273 |  | 13 | 131 | 5278 | 228 | 211 | 94 | 33200 |
| 1999 | - | - | - | 108 | $61^{3}$ | 430 | 97 | 14 |  | 10 |  | - | 24634 |  | 6 | 68 | 4422 | 36 | 247 | 62 | 30195 |


|  | $\sum_{\pi}^{0}$ |  |  |  |  | $\begin{aligned} & \text { 肴 } \\ & \text { N } \\ & \text { EIU } \\ & \text { U } \end{aligned}$ |  |  |  |  | $\begin{aligned} & \text { 㐅} \\ & \text { 3 } \\ & \text { 30 } \end{aligned}$ |  | $\overline{0}$ 0 0 0 |  | $\begin{aligned} & \stackrel{\text { In }}{0} \\ & \stackrel{0}{n} \end{aligned}$ | $\begin{aligned} & \underset{\sim}{\underset{\sim}{x}} \\ & \underset{\sim}{\underset{j}{3}} \\ & \text { un } \end{aligned}$ |  | - |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | - | - | - | $67^{3}$ | 25 | 222 | 51 | 65 | 1 | - | 19052 | 2 | 131 | 4631 | 87 | - | 2036 | 24536 |
| 2001 | - | - | - | $111^{3}$ | 46 | 436 | 34 | 3 | 5 | - | 23071 | 5 | 186 | 4738 | 91 | - | $239{ }^{6}$ | 28965 |
| 2002 | - | - | 15 | $135^{3}$ | 89 | 141 | 49 | 44 | 4 | - | 10713 | $8^{3}$ | 276 | 4736 | $193{ }^{2}$ | - | 2346 | 16636 |
| 2003 | S | - | - | $173^{3}$ | 30 | 154 | $44^{3}$ | 9 | $5^{3}$ | 89 | 8063 | 7 | 50 | 1431 | $47^{2}$ | - | $258{ }^{6}$ | 10360 |
| 2004 | 1 | - | - | 607 | $17^{3}$ | 78 | $24^{3}$ | 40 | 3 | 33 | $7608{ }^{12}$ | 42 | 240 | $3601{ }^{2}$ | $260^{2}$ | - | $145^{6}$ | 12699 |
| 2005 | CAN | LT | 5 | 1194 | 56 | 105 | $75^{3}$ | $12^{2}$ | $4^{3}$ | $55^{2}$ | $7845{ }^{12}$ | - | 196 | 5637 | $171{ }^{3}$ | - | $147^{6}$ | 15502 |
| 2006 | 433 | 845 | 396 | 3919 | 223 | 2518 | $107^{3}$ | $2544{ }^{3}$ | $12^{3}$ | 21 | 11015 | $2496{ }^{2}$ | 1873 | 12126 | $719^{2}$ | - | $1066^{6}$ | 40649 |
| 2007 | LV | 785 | 684 | 2343 | 249 | 587 | $84^{3}$ | $1655^{2}$ | $7^{3}$ | 20 | $8993{ }^{2}$ | $1081^{2}$ | 1708 | 6550 | $2186^{2}$ | - | $257{ }^{6}$ | 27591 |
| 2008 | 267 | 117 | - | $2123^{3}$ | 250 | 46 | $96^{3}$ | $36^{3}$ | $15^{3}$ | 15 | $7436{ }^{1}$ | 8 | 785 | 7866 | $467^{2}$ | $E U^{7}$ | $168{ }^{6}$ | 19695 |
| 2009 | - | - | - | 1413 | 16 | 100 | 81 | 99 | - | 4 | 8128 | 338 | 836 | 4541 | 177 | 889 | $111^{6}$ | 16733 |
| 2010 | $243^{3}$ | 4573 | - | 1150 | 226 | 52 | $84^{3}$ | $24^{3}$ | - | - | 8059 | $1^{3}$ | 321 | 6979 | 1187 | - | $123{ }^{6}$ | 18906 |
| 2011 | 536 | 565 | - | $1008{ }^{2}$ | 228 | 844 | 51 | 24 | - | 1 | 7152 | 59 | 638 | 5956 | $1684^{2}$ | - | $68^{6}$ | 18814 |
| 2012 | 447 | 449 | - | 346 | 182 | 588 | 58 | 59 | 12 | 5 | 6361 | 352 | 1055 | 4782 | $1780^{2}$ | DK | $100^{6}$ | 16576 |
| 2013 | 280 | 262 | - | 780 | 353 | 81 | 66 | 9 | 1 | - | 5606 | 103 | 1114 | 4474 | 1459 | 1 | 4936 | 15082 |
| 2014 | 215 | 167 | - | 810 | 434 | 452 | 35 | 29 | - | 4 | 16556 | 124 | 510 | 2510 | 1162 | - | $211^{6}$ | 23219 |
| 2015 | 537 | 192 | - | 733 | 102 | 266 | 259 | 38 | - | 3 | 22208 | 22 | 678 | 1806 | 2531 | 1 | 1096 | 29485 |
| 2016 | 1243 | 1065 | - | 685 | 164 | 497 | 161 | 79 | - | - | 22322 | 234 | 1066 | 9283 | 32013 | 7 | $198{ }^{6}$ | 40217 |


|  | $\sum_{\substack{\pi}}^{\frac{\pi}{2}}$ |  |  |  | $\begin{aligned} & \text { U } \\ & \text { 든 } \end{aligned}$ |  |  | $\begin{aligned} & \text { 흗 } \\ & \underline{\pi} \\ & \underline{(1} \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \underline{\underline{0}} \\ & \underline{\underline{I}} \end{aligned}$ |  | $\begin{aligned} & \text { 㐅} \\ & \text { 3 } \\ & 0 \\ & \mathbf{0} \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \frac{1}{0} \\ & \mathbf{0} \end{aligned}$ | $\overline{0}$ 0 0 능 |  | $\begin{aligned} & \text {.드제 } \\ & \text { ion } \end{aligned}$ |  | $\begin{aligned} & \mp \\ & \dot{0} \\ & \dot{\omega} \\ & \check{y} \end{aligned}$ | ¢ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | 562 | 790 | 4 | 566 | 62 | 782 | 127 | 68 | - | 2 | 20581 | 129 | 1150 | 7890 | 2882 |  | $596{ }^{6}$ | 36192 |
| 2018 | 1020 | 1010 | - | 571 | 104 | 2539 | 159 | 77 | - | 374 | 23563 | 311 | 766 | 12331 | 2469 |  | $100^{6}$ | 45395 |
| 2019 | - | 656 | - | 392 | 395 | 1692 | 671 | 93 | - | 244 | 29795 | 491 | 1495 | 15373 | 2287 |  | $615^{6}$ | 54199 |
| 2020 | 2 | 1081 | - | 315 | 164 | 1895 | 161 | 57 | - | 1483 | 39453 | 13 | 956 | 16489 | 750 |  | $456{ }^{6}$ | 63277 |
| $2021{ }^{1}$ | - | 1379 | - | 613 | 224 | 1242 | 177 | 78 | - | - | 51498 | 22 | 441 | 16624 | 623 |  | $751{ }^{6}$ | 73675 |

1-Provisional figures.
2 - Working Group figure.
3 - As reported to Norwegian authorities or NEAFC.
4 - Includes former GDR prior to 1991.
5 - USSR prior to 1991.
6 - UK(E\&W) + UK(Scot.)
7 - EU not split on countries.
Table 6.7. REDFISH in Subarea 4 (North Sea). Nominal catch ( t ) by countries as officially reported to ICES. Not included in the assessment.

\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline  \&  \&  \&  \&  \&  \& $$
\begin{aligned}
& \text { ס } \\
& \text { 든 } \\
& \underline{\underline{0}}
\end{aligned}
$$ \&  \& त

3
0
2 \& O
त
त

O \&  \&  \& $$
\begin{aligned}
& \dot{\ddot{0}} \\
& \dot{U} \\
& \text { u}
\end{aligned}
$$ \&  <br>

\hline 1998 \& 2 \& 27 \& 12 \& 570 \& 370 \& 4 \& 21 \& 1113 \& \& - \& - \& 749 \& 2868 <br>
\hline 1999 \& 3 \& 52 \& 1 \& - \& 58 \& 39 \& 16 \& 862 \& \& - \& - \& 532 \& 1563 <br>
\hline
\end{tabular}

| $\begin{aligned} & \text { پ } \\ & \text { ঠ̀ } \end{aligned}$ | $\begin{aligned} & E \\ & \frac{E}{b} \\ & \frac{0}{D 0} \\ & \hline 0 \end{aligned}$ |  |  |  |  | $\begin{aligned} & \text { 들 } \\ & \text { 들 } \\ & \underline{\underline{0}} \end{aligned}$ |  | $\begin{aligned} & \text { त } \\ & \text { 3 } \\ & \text { 30 } \end{aligned}$ |  |  |  |  | - |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 5 | 41 | - | 224 | 19 | 28 | 19 | 443 |  | - | - | 618 | 1397 |
| 2001 | 4 | 96 | - | 272 | 13 | 19 | + | 421 |  | - | - | 538 | 1363 |
| 2002 | 2 | 40 | 2 | 98 | 11 | 7 | + | 241 |  | - | - | 524 | 925 |
| 2003 | 1 | 71 | 2 | 26 | 2 | - | - | 474 |  | - | - | 463 | 1039 |
| 2004 | + | 42 | 3 | 26 | 1 | - | - | 287 |  | - | - | 214 | 578 |
| 2005 | 2 | 34 | - | 10 | 1 | - | - | 84 |  | - | - | 28 | 159 |
| 2006 | 1 | 49 | 1 | 12 | 3 | - | - | 163 | - | 33 | - | 79 | 341 |
| 2007 | + | 27 | - | 8 | 1 | - | - | 116 | 1 | - | - | 77 | 230 |
| 2008 | + | 3 | - | 8 | 1 | - | - | 77 | - | - | 1 | 54 | 144 |
| 2009 | + | 4 | 1 | 38 | + | - | - | 119 | - | - | + | 86 | 248 |
| 2010 | - | 5 | - | 3 | - | - | - | 62 | - | - | + | 150 | 220 |
| 2011 | - | 9 | - | 90 | 1 | - | - | 66 | - | - | + | 71 | 237 |
| 2012 | - | 10 | - | 19 | + | - | - | 71 | - | - | + | 87 | 187 |
| 2013 | - | 7 | - | 40 | + | - | - | 54 | - | - | - | 176 | 277 |
| 2014 | - | - | - | 32 | 1 | - | - | 146 | - | - | + | 93 | 272 |
| 2015 | + | 1 | - | 14 | 1 | - | - | 157 | - | - | + | 61 | 234 |
| 2016 | - | 3 | - | 11 | + | - | - | 180 | - | - | + | 22 | 216 |


| $\begin{aligned} & \text { 厄゙ } \\ & \text { ঠ्र } \end{aligned}$ | $\begin{aligned} & E \\ & \frac{E}{\overline{b 0}} \\ & \stackrel{D}{D} \end{aligned}$ |  | $\begin{aligned} & \text { n } \\ & \frac{0}{\pi} \\ & \frac{\pi}{G} \\ & \frac{0}{0} \\ & \frac{0}{4} \end{aligned}$ |  |  | $\begin{aligned} & \text { 들 } \\ & \text { 들 } \\ & \underline{\underline{0}} \end{aligned}$ |  | $\begin{aligned} & \text { त } \\ & \text { 3 } \\ & \text { 30 } \end{aligned}$ | $\begin{aligned} & \text { ㄷ } \\ & \frac{\bar{\pi}}{0} \\ & \hline \end{aligned}$ | $\begin{aligned} & \overline{5} \\ & 0 \\ & 0 \\ & 0.7 \\ & 00 \end{aligned}$ |  | $\begin{aligned} & \dot{\ddot{0}} \\ & \dot{\sim} \\ & \text { y } \end{aligned}$ | - |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | - | 3 | - | 10 | + | - | - | 168 | - | - | + | 38 | 21 |
| 2018 | - | 10 | - | 4 | - | - | - | 71 | - | - | + | 29 | 114 |
| $2019{ }^{1}$ | - | 7 | + | 10 | + | - | + | 62 | - | - | + | 10 | 89 |
| 2020 | - | 10 | - | 4 | + | - | + | 54 | - | - | + | 27 | 95 |
| $2021{ }^{1}$ | - | 4 | - | 11 | + | - | + | 30 | - | - | + | 123 | 168 |

1 - Provisional figures.

+ denotes less than 0.5 tonnes.
Table 6.8. S. mentella in subareas 1 and 2. Catch numbers-at-age 6 to 18 and 19+ (in thousands) and total landings (in tonnes). For the period 2012-2016 age data are missing from the pelagic fishery. For the period 2015-2018, age data are missing from all fisheries. The numbers-at-age have been estimated following the method outlined in section 6.2.2.

| Year/Age | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | +gp | Total No. | Tonnes Land. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 1873 | 2498 | 1898 | 1622 | 1780 | 1531 | 2108 | 2288 | 2258 | 2506 | 2137 | 1512 | 677 | 9258 | 33946 | 15590 |
| 1993 | 159 | 159 | 174 | 512 | 2094 | 3139 | 2631 | 2308 | 2987 | 1875 | 1514 | 1053 | 527 | 6022 | 25154 | 12814 |
| 1994 | 738 | 730 | 722 | 992 | 2561 | 2734 | 3060 | 1535 | 2253 | 2182 | 3336 | 1284 | 734 | 3257 | 26118 | 12721 |
| 1995 | 662 | 941 | 1279 | 719 | 740 | 1230 | 2013 | 4297 | 3300 | 2162 | 1454 | 757 | 794 | 2404 | 22752 | 10284 |
| 1996 | 223 | 634 | 1699 | 1554 | 1236 | 1078 | 1146 | 1413 | 1865 | 880 | 621 | 498 | 700 | 2247 | 15794 | 8075 |
| 1997 | 125 | 533 | 1287 | 1247 | 1297 | 1244 | 876 | 1416 | 1784 | 1217 | 537 | 1177 | 342 | 3568 | 16650 | 8598 |
| 1998 | 37 | 882 | 2904 | 4236 | 3995 | 2741 | 1877 | 1373 | 1277 | 1595 | 1117 | 784 | 786 | 6241 | 29845 | 14045 |


| Year/Age | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | +gp | Total No. | Tonnes Land. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 9 | 83 | 441 | 1511 | 2250 | 3262 | 1867 | 1454 | 1447 | 1557 | 1418 | 1317 | 658 | 3919 | 21193 | 11209 |
| 2000 | 1 | 24 | 390 | 1235 | 2460 | 2149 | 1816 | 1205 | 1001 | 993 | 932 | 505 | 596 | 5705 | 19012 | 10075 |
| 2001 | 117 | 372 | 542 | 976 | 925 | 1712 | 2651 | 2660 | 1911 | 1773 | 1220 | 714 | 814 | 16234 | 32621 | 18418 |
| 2002 | 2 | 40 | 252 | 572 | 709 | 532 | 1382 | 1893 | 1617 | 855 | 629 | 163 | 237 | 4082 | 12965 | 6993 |
| 2003 | 6 | 37 | 103 | 93 | 132 | 220 | 384 | 391 | 434 | 466 | 513 | 199 | 231 | 1193 | 4402 | 2520 |
| 2004 | 7 | 16 | 70 | 96 | 278 | 429 | 611 | 433 | 1063 | 813 | 830 | 841 | 607 | 3076 | 9170 | 5493 |
| 2005 | 2 | 20 | 57 | 155 | 244 | 262 | 295 | 754 | 783 | 1896 | 817 | 1087 | 1023 | 6065 | 13460 | 8465 |
| 2006 | 0 | 4 | 3 | 38 | 64 | 121 | 423 | 1461 | 1356 | 2835 | 4271 | 3487 | 3969 | 32084 | 50116 | 33261 |
| 2007 | 0 | 1 | 3 | 22 | 33 | 86 | 235 | 631 | 2194 | 2825 | 3657 | 4359 | 3540 | 15824 | 33410 | 20219 |
| 2008 | 0 | 0 | 1 | 10 | 46 | 100 | 197 | 469 | 612 | 1502 | 1384 | 894 | 1886 | 11906 | 19007 | 13095 |
| 2009 | 0 | 1 | 16 | 22 | 42 | 39 | 254 | 258 | 577 | 364 | 823 | 692 | 1856 | 11706 | 16650 | 10246 |
| 2010 | 10 | 4 | 6 | 19 | 34 | 55 | 61 | 241 | 267 | 390 | 566 | 655 | 667 | 13879 | 16854 | 11924 |
| 2011 | 4 | 4 | 4 | 25 | 55 | 114 | 11 | 103 | 286 | 394 | 408 | 479 | 567 | 15223 | 17677 | 12962 |
| 2012 | 4 | 24 | 29 | 24 | 26 | 66 | 69 | 78 | 80 | 279 | 387 | 365 | 409 | 13332 | 15172 | 11056 |
| 2013 | 0 | 3 | 19 | 92 | 88 | 41 | 42 | 42 | 10 | 167 | 144 | 174 | 299 | 11726 | 12847 | 9474 |
| 2014 | 14 | 28 | 346 | 97 | 124 | 96 | 152 | 55 | 111 | 69 | 252 | 293 | 197 | 23744 | 25578 | 18780 |
| 2015 | 43 | 41 | 135 | 569 | 849 | 1362 | 1254 | 721 | 388 | 952 | 291 | 599 | 877 | 29612 | 37693 | 25856 |
| 2016 | 42 | 0 | 1015 | 687 | 3469 | 2670 | 3089 | 2067 | 2037 | 1314 | 1385 | 1288 | 1143 | 37744 | 57950 | 35646 |


| Year/Age | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | +gp | Total No. | Tonnes Land. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | 0 | 84 | 0 | 4479 | 2823 | 11454 | 5380 | 4385 | 2451 | 2235 | 1396 | 1437 | 1290 | 20897 | 58311 | 30934 |
| 2018 | 1173 | 4126 | 4511 | 4873 | 7166 | 4872 | 2339 | 2925 | 3570 | 6944 | 1973 | 2330 | 2677 | 30661 | 80140 | 38739 |
| 2019 | 0 | 4106 | 14968 | 14423 | 12882 | 15533 | 8137 | 2059 | 3499 | 4599 | 10818 | 2992 | 3576 | 11058 | 108650 | 45954 |
| 2020 | 0 | 0 | 8772 | 23581 | 18571 | 15195 | 17516 | 9091 | 2319 | 3883 | 5056 | 11870 | 3273 | 9248 | 128375 | 54686 |

Table 6.9. Pelagic S. mentella in the Norwegian Sea (outside the EEZ). Catch numbers-at-age.

| Numbers $\mathbf{1 0}^{\mathbf{3}}$ |  |  |  | Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19+ |
| 2006 | 0 | 0 | 0 | 0 | 23 | 93 | 1083 | 323 | 1563 | 3628 | 2514 | 3756 | 29704 |
| 2007 | 0 | 0 | 9 | 18 | 25 | 154 | 444 | 1642 | 2302 | 3021 | 3394 | 3156 | 12684 |
| 2008 | 0 | 0 | 0 | 0 | 28 | 146 | 115 | 143 | 214 | 594 | 752 | 753 | 13258 |
| 2009 | 0 | 0 | 0 | 0 | 9 | 1314 | 294 | 471 | 889 | 999 | 869 | 1150 | 2981 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 0 | 130 | 336 | 254 | 466 | 467 | 508 | 11510 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 223 | 83 | 83 | 168 | 136 | 166 | 136 | 13182 |
| $2012{ }^{1}$ | 0 | 0 | 0 | 22 | 29 | 19 | 294 | 146 | 132 | 217 | 288 | 126 | 8939 |
| $2013{ }^{2}$ | 11 | 137 | 98 | 465 | 123 | 158 | 96 | 169 | 246 | 196 | 238 | 598 | 7968 |
| $2014{ }^{3}$ | 0 | 10 | 125 | 88 | 406 | 103 | 125 | 70 | 113 | 151 | 112 | 130 | 4398 |
| $2015{ }^{3}$ | 0 | 0 | 0 | 0 | 169 | 54 | 51 | 0 | 0 | 0 | 85 | 22 | 6345 |
| $2016^{3}$ | 0 | 0 | 154 | 307 | 271 | 276 | 134 | 90 | 107 | 239 | 445 | 229 | 10499 |


| Numbers $10^{3}$ |  |  |  | Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19+ |
| $2017{ }^{3}$ | 0 | 0 | 0 | 237 | 461 | 389 | 370 | 165 | 100 | 109 | 226 | 402 | 8351 |
| $2018{ }^{3}$ | 0 | 0 | 0 | 0 | 687 | 1274 | 1004 | 873 | 352 | 195 | 199 | 393 | 12673 |
| 2019 | 25 | 5 | 200 | 400 | 220 | 242 | 197 | 279 | 183 | 155 | 135 | 161 | 6696 |
| $2020{ }^{4}$ | 0 | 44 | 8 | 344 | 670 | 352 | 361 | 270 | 345 | 207 | 163 | 136 | 5500 |

1 - No age data in 2012, catch numbers-at-age are estimated from proportions at age in 2011 and in 2013.
2 - No age data from the catches in 2013. Age readings from the research survey conducted in September 2013 are used to derive catch numbers-at-age.
3 - No age data in 2014-2018, catch numbers-at-age are estimated from previous year according to protocol described in section 6.2.2.
4 - No age data in 2020, catch numbers-at-age are estimated from previous year according to protocol described in section 6.2.2.

Table 6.10. S. mentella in subareas 1 and 2. Total catch numbers-at-length, in thousands, for 2011-2020.

| Year | Length group |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | N ¢ $\cdots$ | $\begin{aligned} & \text { N } \\ & \text { Ǹ } \end{aligned}$ | $\begin{aligned} & \underset{\sim}{N} \\ & \underset{N}{N} \end{aligned}$ | $\begin{aligned} & \stackrel{0}{N} \\ & \underset{N}{N} \end{aligned}$ | $\begin{gathered} \infty \\ \underset{N}{N} \end{gathered}$ | $\begin{gathered} \text { O} \\ \\ \text { No } \end{gathered}$ | $\underset{N}{N}$ | $\underset{\sim}{\underset{\sim}{N}}$ | $\begin{aligned} & \text { O} \\ & \underset{N}{1} \end{aligned}$ | $\begin{gathered} \infty \\ \substack{0 \\ \hline} \end{gathered}$ | $\begin{aligned} & \text { O} \\ & \mathbf{\infty} \end{aligned}$ | $\begin{aligned} & \text { Y } \\ & \hline \end{aligned}$ | $\pm$ $\pm$ $\sim$ | 0 + 4 | $\underbrace{\infty}_{0}$ | ¢ $\substack{\infty \\ \text { ¢ }}$ | N in in |
| 2011 | 0 | 12 | 0 | 0 | 1 | 8 | 249 | 2544 | 6481 | 6528 | 3620 | 829 | 95 | 18 | 1 | 0 | 0 |
| 2012 | 0 | 0 | 23 | 19 | 26 | 28 | 41 | 287 | 1898 | 5030 | 5385 | 1911 | 451 | 197 | 43 | 23 | 0 |
| 2013 | 0 | 0 | 4 | 32 | 154 | 137 | 90 | 69 | 1382 | 4214 | 4480 | 1633 | 497 | 197 | 0 | 0 | 0 |
| 2014 | 0 | 5 | 0 | 25 | 29 | 235 | 660 | 697 | 3358 | 7667 | 8544 | 3808 | 787 | 34 | 0 | 0 | 0 |
| 2015 | Data not available at the time of the working group |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2016 | Data not available at the time of the working group |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |


Lear

Table 6.12. S. mentella in subareas 1 and 2. Observed mean weights-at-age ( $\mathbf{k g}$ ) from the Norwegian data (Catches and surveys combined). Weights-at-age used in the statistical catch-at-age model are identical for every year and given at the bottom line of the table.

| Year/Age | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.167 | 0.164 | 0.211 | 0.241 | 0.309 | 0.324 | 0.378 | 0.366 | 0.428 | 0.454 | 0.487 | 0.529 | 0.571 | 0.805 |
| 1993 | 0.141 | 0.181 | 0.217 | 0.254 | 0.306 | 0.357 | 0.349 | 0.4 | 0.45 | 0.436 | 0.46 | 0.499 | 0.462 | 0.846 |
| 1994 | 0.174 | 0.188 | 0.235 | 0.298 | 0.361 | 0.396 | 0.415 | 0.48 | 0.492 | 0.562 | 0.642 | 0.636 | 0.72 | 0.846 |
| 1995 | 0.158 | 0.185 | 0.226 | 0.261 | 0.324 | 0.36 | 0.432 | 0.468 | 0.496 | 0.519 | 0.566 | 0.573 | 0.621 | 0.758 |
| 1996 | 0.175 | 0.189 | 0.224 | 0.272 | 0.323 | 0.337 | 0.377 | 0.518 | 0.536 | 0.603 | 0.69 | 0.8 | 0.683 | 0.958 |
| 1997 | 0.152 | 0.191 | 0.228 | 0.28 | 0.324 | 0.367 | 0.435 | 0.492 | 0.521 | 0.615 | 0.601 | 0.611 | 0.671 | 0.911 |
| 1998 | 0.12 | 0.148 | 0.192 | 0.261 | 0.326 | 0.373 | 0.427 | 0.496 | 0.537 | 0.566 | 0.587 | 0.625 | 0.658 | 0.809 |
| 1999 | 0.133 | 0.17 | 0.226 | 0.286 | 0.343 | 0.382 | 0.441 | 0.483 | 0.537 | 0.565 | 0.62 | 0.644 | 0.672 | 0.757 |
| 2000 | 0.109 | 0.144 | 0.199 | 0.276 | 0.332 | 0.392 | 0.437 | 0.49 | 0.54 | 0.585 | 0.631 | 0.65 | 0.671 | 0.872 |


| Year/Age | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 0.115 | 0.137 | 0.183 | 0.262 | 0.31 | 0.356 | 0.4 | 0.434 | 0.484 | 0.534 | 0.581 | 0.615 | 0.624 | 0.819 |
| 2002 | 0.114 | 0.139 | 0.182 | 0.253 | 0.329 | 0.372 | 0.392 | 0.434 | 0.476 | 0.52 | 0.545 | 0.587 | 0.601 | 0.833 |
| 2003 | 0.109 | 0.124 | 0.196 | 0.245 | 0.312 | 0.371 | 0.422 | 0.434 | 0.477 | 0.516 | 0.551 | 0.591 | 0.623 | 0.817 |
| 2004 | 0.104 | 0.129 | 0.18 | 0.264 | 0.308 | 0.376 | 0.413 | 0.444 | 0.478 | 0.521 | 0.579 | 0.614 | 0.688 | 0.835 |
| 2005 | 0.104 | 0.136 | 0.196 | 0.263 | 0.322 | 0.37 | 0.408 | 0.451 | 0.478 | 0.523 | 0.55 | 0.551 | 0.64 | 0.797 |
| 2006 | 0.107 | 0.143 | 0.2 | 0.266 | 0.314 | 0.374 | 0.419 | 0.462 | 0.489 | 0.527 | 0.57 | 0.602 | 0.59 | 0.796 |
| 2007 | 0.115 | 0.131 | 0.18 | 0.252 | 0.305 | 0.364 | 0.409 | 0.449 | 0.485 | 0.513 | 0.523 | 0.554 | 0.569 | 0.737 |
| 2008 | 0 | 0.158 | 0.177 | 0.242 | 0.304 | 0.402 | 0.465 | 0.486 | 0.511 | 0.546 | 0.6 | 0.596 | 0.635 | 0.803 |
| 2009 | 0.129 | 0.179 | 0.206 | 0.249 | 0.326 | 0.394 | 0.51 | 0.55 | 0.542 | 0.583 | 0.609 | 0.594 | 0.595 | 0.809 |
| 2010 | 0.129 | 0.128 | 0.175 | 0.263 | 0.375 | 0.447 | 0.501 | 0.541 | 0.582 | 0.602 | 0.593 | 0.608 | 0.592 | 0.706 |
| 2011 | 0.136 | 0.156 | 0.183 | 0.261 | 0.316 | 0.435 | 0.512 | 0.604 | 0.655 | 0.609 | 0.671 | 0.647 | 0.677 | 0.795 |
| 2012 | 0.135 | 0.178 | 0.225 | 0.246 | 0.249 | 0.356 | 0.474 | 0.582 | 0.53 | 0.626 | 0.654 | 0.73 | 0.699 | 0.833 |
| 2013 | 0.129 | 0.145 | 0.189 | 0.23 | 0.27 | 0.282 | 0.345 | 0.384 | 0.534 | 0.559 | 0.634 | 0.627 | 0.661 | 0.72 |
| 2014 | 0.193 | 0.172 | 0.221 | 0.167 | 0.192 | 0.239 | 0.333 | 0.277 | 0.364 | 0.516 | 0.713 | 0.78 | 0.797 | 0.882 |
| 2015 | 0.167 | 0.168 | 0.232 | 0.294 | 0.346 | 0.383 | 0.457 | 0.436 | 0.474 | 0.538 | 0.665 | 0.69 | 0.724 | 0.824 |
| $2016{ }^{1}$ | 0.11 | 0 | 0.331 | 0.356 | 0.401 | 0.392 | 0.434 | 0.486 | 0.543 | 0.579 | 0.74 | 0.591 | 0.598 | 0.776 |
| 2017 | 0.154 | 0.196 | 0.254 | 0.27 | 0.306 | 0.413 | 0.425 | 0.458 | 0.533 | 0.472 | 0.562 | 0.65 | 0.692 | 0.796 |
| $2018{ }^{1}$ | 0 | 0.233 | 0.135 | 0.371 | 0.323 | 0.28 | 0.379 | 0.452 | 0.524 | 0.633 | 0.483 | 0.589 | 0.457 | 0.821 |


| Year/Age | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5}$ | $\mathbf{1 6}$ | $\mathbf{1 7}$ | $\mathbf{1 8}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $2019^{1}$ | 0.118 | 0.38 | 0.341 | 0.47 | 0.538 | 0.523 | 0.539 | 0.565 | 0.572 | 0.62 | 0.656 | 0.601 | 0.633 |
| Modelled | 0.141 | 0.188 | 0.237 | 0.286 | 0.334 | 0.381 | 0.424 | 0.465 | 0.503 | 0.537 | 0.569 | 0.597 | 0.623 |

1 - Provisional figures.
Table 6.13. Pelagic S. mentella in the Norwegian Sea (outside the EEZ). Catch weights-at-age (kg).

| Year/ Age | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 0.44 | 0.44 | 0.52 | 0.44 | 0.49 | 0.55 | 0.53 | 0.56 | 0.61 |
| 2007 | 0.39 | 0.43 | 0.41 | 0.48 | 0.50 | 0.52 | 0.55 | 0.57 | 0.64 |
| 2008 | 0.36 | 0.47 | 0.56 | 0.50 | 0.56 | 0.54 | 0.56 | 0.55 | 0.64 |
| 2009 | 0.38 | 0.44 | 0.45 | 0.48 | 0.54 | 0.59 | 0.64 | 0.58 | 0.69 |
| 2010 | - | - | 0.62 | 0.56 | 0.54 | 0.59 | 0.59 | 0.56 | 0.61 |
| 2011 | - | 0.48 | 0.54 | 0.54 | 0.64 | 0.59 | 0.54 | 0.59 | 0.59 |
| 2012 | No data | - | - | - | - | - | - | - | - |
| $2013{ }^{2}$ | 0.31 | - | - | - | 0.56 | 0.62 | 0.60 | 0.62 | 0.68 |
| 2014 | No data | - | - | - | - | - | - | - | - |
| 2015 | No data | - | - | - | - | - | - | - | - |
| 2016 | No data | - | - | - | - | - | - | - | - |
| 2017 | No data | - | - | - | - | - | - | - | - |
| 2018 | No data | - | - | - | - | - | - | - | - |
| 2019 | No data | - | - | - | - | - | - | - | - |


| 2020 | No data | - | - | - | - |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $2021^{1}$ | No data | - | - | - | - | - |

1 - Provisional figures.
2 - As observed in the research survey in the Norwegian Sea in September 2013.

Table 6.14. Proportion of maturity-at-age 6-19+ in S. mentella in subareas 1 and 2 derived from Norwegian commercial and survey data. The proportions were derived from samples with at least 5 individuals. a50 w1 and w2 are the annual coefficients for modelled maturity ogives using a double half sigmoid of the form 0.5 ((1+tanh(age- a50)/w1)) for age < a50 and 0.5 (1+tanh ((age-a50)/w2) for age >a50. a50 equals the age at $50 \%$ maturity.

| year/Age | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.00 | 0.01 | 0.02 | 0.04 | 0.07 | 0.14 | 0.26 | 0.42 | 0.53 | 0.59 | 0.65 | 0.70 | 0.75 | 1.00 |
| 1993 | 0.01 | 0.02 | 0.04 | 0.08 | 0.15 | 0.28 | 0.44 | 0.55 | 0.61 | 0.67 | 0.72 | 0.77 | 0.82 | 1.00 |
| 1994 | 0.02 | 0.04 | 0.08 | 0.15 | 0.28 | 0.44 | 0.59 | 0.72 | 0.81 | 0.88 | 0.93 | 0.96 | 0.98 | 1.00 |
| 1995 | 0.03 | 0.07 | 0.13 | 0.24 | 0.39 | 0.57 | 0.71 | 0.83 | 0.90 | 0.95 | 0.97 | 0.98 | 0.99 | 1.00 |
| 1996 | 0.01 | 0.01 | 0.02 | 0.05 | 0.10 | 0.19 | 0.33 | 0.50 | 0.59 | 0.66 | 0.73 | 0.79 | 0.84 | 1.00 |
| 1997 | 0.02 | 0.04 | 0.08 | 0.16 | 0.29 | 0.46 | 0.55 | 0.61 | 0.66 | 0.71 | 0.76 | 0.80 | 0.84 | 1.00 |
| 1998 | 0.02 | 0.04 | 0.08 | 0.15 | 0.26 | 0.43 | 0.56 | 0.65 | 0.73 | 0.80 | 0.85 | 0.90 | 0.93 | 1.00 |
| 1999 | 0.03 | 0.05 | 0.10 | 0.20 | 0.34 | 0.51 | 0.57 | 0.64 | 0.70 | 0.75 | 0.80 | 0.84 | 0.87 | 1.00 |
| 2000 | 0.03 | 0.06 | 0.11 | 0.21 | 0.36 | 0.52 | 0.63 | 0.73 | 0.81 | 0.87 | 0.91 | 0.94 | 0.96 | 1.00 |
| 2001 | 0.01 | 0.02 | 0.04 | 0.09 | 0.17 | 0.30 | 0.47 | 0.56 | 0.62 | 0.68 | 0.74 | 0.79 | 0.83 | 1.00 |
| 2002 | 0.02 | 0.05 | 0.10 | 0.19 | 0.33 | 0.50 | 0.54 | 0.59 | 0.63 | 0.67 | 0.70 | 0.74 | 0.77 | 1.00 |
| 2003 | 0.03 | 0.06 | 0.12 | 0.21 | 0.36 | 0.51 | 0.57 | 0.63 | 0.69 | 0.73 | 0.78 | 0.82 | 0.85 | 1.00 |
| 2004 | 0.03 | 0.06 | 0.12 | 0.22 | 0.37 | 0.51 | 0.55 | 0.59 | 0.63 | 0.67 | 0.70 | 0.73 | 0.76 | 1.00 |


| year/Age | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 0.02 | 0.05 | 0.09 | 0.18 | 0.31 | 0.49 | 0.55 | 0.61 | 0.66 | 0.71 | 0.75 | 0.79 | 0.83 | 1.00 |
| 2006 | 0.01 | 0.02 | 0.03 | 0.07 | 0.13 | 0.24 | 0.39 | 0.53 | 0.59 | 0.64 | 0.70 | 0.75 | 0.79 | 1.00 |
| 2007 | 0.02 | 0.04 | 0.09 | 0.17 | 0.30 | 0.47 | 0.64 | 0.77 | 0.87 | 0.93 | 0.96 | 0.98 | 0.99 | 1.00 |
| $2008{ }^{1}$ | 0.02 | 0.04 | 0.08 | 0.15 | 0.27 | 0.43 | 0.55 | 0.62 | 0.68 | 0.74 | 0.79 | 0.83 | 0.87 | 1.00 |
| 2009 | 0.02 | 0.04 | 0.09 | 0.17 | 0.30 | 0.47 | 0.60 | 0.71 | 0.80 | 0.87 | 0.92 | 0.95 | 0.97 | 1.00 |
| 2010 | 0.02 | 0.04 | 0.08 | 0.16 | 0.28 | 0.45 | 0.54 | 0.60 | 0.66 | 0.71 | 0.76 | 0.80 | 0.83 | 1.00 |
| $2011{ }^{1}$ | 0.02 | 0.04 | 0.08 | 0.15 | 0.27 | 0.43 | 0.55 | 0.62 | 0.68 | 0.74 | 0.79 | 0.83 | 0.87 | 1.00 |
| 2012 | 0.02 | 0.05 | 0.10 | 0.19 | 0.32 | 0.50 | 0.59 | 0.68 | 0.75 | 0.81 | 0.86 | 0.90 | 0.93 | 1.00 |
| 2013 | 0.00 | 0.01 | 0.02 | 0.04 | 0.08 | 0.15 | 0.28 | 0.45 | 0.62 | 0.77 | 0.87 | 0.93 | 0.97 | 1.00 |
| 2014 | 0.00 | 0.00 | 0.01 | 0.02 | 0.03 | 0.06 | 0.12 | 0.23 | 0.38 | 0.53 | 0.61 | 0.68 | 0.74 | 1.00 |
| 2015 | 0.01 | 0.02 | 0.05 | 0.09 | 0.17 | 0.31 | 0.48 | 0.54 | 0.58 | 0.63 | 0.67 | 0.71 | 0.74 | 1.00 |
| 2016 | 0.03 | 0.06 | 0.12 | 0.22 | 0.38 | 0.52 | 0.56 | 0.61 | 0.66 | 0.70 | 0.74 | 0.77 | 0.81 | 1.00 |
| $2017{ }^{1}$ | 0.02 | 0.04 | 0.08 | 0.15 | 0.27 | 0.43 | 0.55 | 0.62 | 0.68 | 0.74 | 0.79 | 0.83 | 0.87 | 1.00 |
| $2018{ }^{1}$ | 0.02 | 0.04 | 0.08 | 0.15 | 0.27 | 0.43 | 0.55 | 0.62 | 0.68 | 0.74 | 0.79 | 0.83 | 0.87 | 1.00 |
| $2019{ }^{1}$ | 0.02 | 0.04 | 0.08 | 0.15 | 0.27 | 0.43 | 0.55 | 0.62 | 0.68 | 0.74 | 0.79 | 0.83 | 0.87 | 1.00 |
| $2020{ }^{1}$ | 0.02 | 0.04 | 0.08 | 0.15 | 0.27 | 0.43 | 0.55 | 0.62 | 0.68 | 0.74 | 0.79 | 0.83 | 0.87 | 1.00 |
| $2021{ }^{1}$ | 0.02 | 0.04 | 0.08 | 0.15 | 0.27 | 0.43 | 0.55 | 0.62 | 0.68 | 0.74 | 0.79 | 0.83 | 0.87 | 1.00 |

1 - Model parameter estimates were unrealistic and replaced by average parameter values.

Table 6.15. S. mentella. Average catch (numbers of specimens) per hour trawling of different ages of S. mentella in the Russian groundfish survey in the Barents Sea and Svalbard areas (19761983 published in Annales Biologiques). The survey was not conducted in 2016 took place in 2017 with insufficient coverage and was terminated after that year.

| Year class | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1974 | - | - | 4.8 | - | 4.9 | 22.8 | 4.8 | 4.8 | - | - | - | 3 |
| 1975 | - | 7.4 | - | 1.7 | 6.4 | 2.4 | 3.5 | 5 | - | - | 4 | - |
| 1976 | 7 | - | 8.1 | 1.2 | 2.5 | 6.8 | 4.9 | 5 | 1 | 13 | - | - |
| 1977 | - | 0.2 | 0.2 | 0.2 | 0.9 | 5.1 | 3.7 | 1 | 19 | 2 | - | - |
| 1978 | 0.8 | 0.02 | 0.9 | 1 | 5 | 3.8 | 2 | 20 | 6 | - | - | - |
| 1979 | - | 1.9 | 1.4 | 3.6 | 2.3 | 9 | 11 | 16 | 1 | - | - | 0.1 |
| 1980 | 0.3 | 0.4 | 2 | 2.5 | 16 | 6 | 11 | 25 | 2 | - | 1.5 | 2 |
| 1981 | - | 2.2 | 3.9 | 20 | 6 | 12 | 47 | 18 | 6.3 | 1.6 | 0.5 | 1 |
| 1982 | 19.8 | 13.2 | 13 | 15 | 34 | 44 | 39 | 32.6 | 4.3 | 3.1 | 4.9 | + |
| 1983 | 12.5 | 3 | 5 | 6 | 31 | 34 | 32.3 | 13.3 | 4 | 4.2 | 0.6 | 1.1 |
| 1984 | - | 10 | 2 | - | 5 | 18.3 | 19 | 2.2 | 2.4 | 0.2 | 1.7 | 2.4 |
| 1985 | 107 | 7 | - | 1 | 5.2 | 16.2 | 1.7 | 1.7 | 0.6 | 2.8 | 3.8 | 0.3 |
| 1986 | 2 | - | 1 | 1.8 | 8.4 | 3.6 | 2.1 | 1.2 | 5.6 | 8.2 | 0.9 | 0.7 |
| 1987 | - | 3 | 37.9 | 1.3 | 8 | 4.1 | 2 | 10.6 | 9.6 | 1.4 | 2 | 1.3 |
| 1988 | 4 | 58.1 | 4.3 | 13.3 | 25.8 | 3.9 | 8.6 | 11.2 | 2.8 | 4.2 | 3 | 4.7 |
| 1989 | 8.7 | 9 | 17 | 23.4 | 4.6 | 5.4 | 4 | 6.6 | 6.6 | 4.1 | 7.7 | 5.3 |
| 1990 | 2.5 | 6.3 | 6.1 | 1 | 4.3 | 1.7 | 11.5 | 6.5 | 5.5 | 6.7 | 7.4 | 3.6 |


| Year class | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 0.3 | 1 | 0.5 | 1.5 | 1.2 | 11.3 | 3.9 | 3.3 | 4.6 | 5.8 | 2.7 | 1.9 |
| 1992 | 0.6 | + | 0.2 | 0.1 | 4.3 | 1.3 | 2 | 2.3 | 4.9 | 2.3 | 1 | 4.1 |
| $1993{ }^{1}$ | - | + | 1.5 | 1.8 | 1 | 1.2 | 3 | 4.2 | 2.6 | 2 | 3.2 | 2.1 |
| 1994 | 0.3 | 3.5 | 1.7 | 1.7 | 0.9 | 3.6 | 5.2 | 4.3 | 3.1 | 3.3 | 1.8 | 1.2 |
| 1995 | 2.8 | 1 | 1.1 | 0.4 | 2.2 | 2.6 | 3.5 | 3.4 | 2.9 | 1.2 | 1 | 8.5 |
| $1996{ }^{2}$ | + | 0.1 | 0.1 | 0.4 | 0.7 | 1.1 | 1 | 1.4 | 1 | 0.8 | 3.7 | 0.6 |
| 1997 | - | - | + | 0.4 | 0.5 | 0.3 | 0.9 | 0.6 | 1 | 1.1 | 0.5 | 0.4 |
| 1998 | - | 0.1 | 0.2 | 0.3 | 0.2 | 1.1 | 0.5 | 0.7 | 1 | 0.4 | 0.4 | 0.7 |
| 1999 | 0.1 | - | 0.1 | + | 0.1 | 0.3 | 0.5 | 0.8 | 0.5 | 0.2 | 0.4 | 0.6 |
| 2000 | - | 0.6 | 0.1 | 0.5 | 0.3 | 0.3 | 0.6 | 0.4 | 0.1 | 0.1 | 0.7 | 0.3 |
| 2001 | - | 0.1 | 0.4 | - | 0.1 | 0.2 | 0.2 | 0.3 | 0.2 | 0.8 | 0.1 | 1 |
| 20023 | 0.1 | 0.5 | 0.1 | - | - | 0.1 | 0.5 | 0.4 | 1.5 | 0.5 | 1 | 1.1 |
| 2003 | - | - | 0.1 | - | 0.3 | 1.0 | 0.5 | 4.8 | 2.1 | 3.7 | 1.3 | 1.9 |
| 2004 | - | 0.2 | 0.3 | 0.5 | 1.5 | 0.9 | 4.4 | 3.7 | 7.5 | 4.1 | 3.1 | 3.3 |
| 2005 | - | - | 1.4 | 1.9 | 1.4 | 2.3 | 3.9 | 7.2 | 6.1 | 6.8 | 3.1 |  |
| $2006{ }^{4}$ | 0.1 | 1.8 | 1.2 | 1.1 | 0.8 | 2.1 | 4.1 | 3.0 | 6.1 | 5.9 |  |  |
| 2007 | 2.5 | 0.4 | 0.1 | 1.2 | 1.7 | 2.4 | 3.6 | 4.3 | 7.4 |  |  |  |
| 2008 | 0.1 | 0.1 | 1.6 | 1.8 | 4.1 | 2.9 | 5.8 | 5.5 |  |  |  |  |


| Year class | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 1.6 | 1.9 | 1.1 | 4.4 | 4.8 | 2.9 | 4.8 |  |  |  |  |  |
| 2010 | 7.5 | 0.7 | 1.2 | 1.5 | 1.9 | 1.6 |  |  |  |  |  |  |
| 2011 | 0.1 | 0.3 | 0.6 | 1.6 | 1.6 |  |  |  |  |  |  |  |
| 2012 | 0.2 | 0.7 | 0.5 | 0.3 |  |  |  |  |  |  |  |  |
| 2013 | 0.1 | 0.1 | 0.4 |  |  |  |  |  |  |  |  |  |
| 2014 | 3.6 | 1.0 |  |  |  |  |  |  |  |  |  |  |
| 2015 | 6.6 |  |  |  |  |  |  |  |  |  |  |  |

1 - Not complete area coverage of Division 2.b.
2 - Area surveyed restricted to Subarea 1 and Division 2.a only.
3 - Area surveyed restricted to Subarea 1 and Division 2.b only.
4 - Area surveyed restricted to divisions 2.a and 2.b only.
Table 6.16a. S. mentella ${ }^{1}$ in Division 2.b. Abundance indices (on length) from the bottom trawl survey in the Svalbard area (Division 2.b) in summer/autumn 1986-2021 (numbers in millions).

| Year | Length group (cm) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 5.0-9.9 | 10.0-14.9 | 15.0-19.9 | 20.0-24.9 | 25.0-29.9 | 30.0-34.9 | 35.0-39.9 | 40.0-44.9 | > 45.0 | Total |
| $1986{ }^{2}$ | 6 | 101 | 192 | 17 | 10 | 5 | 2 | 4 | 0 | 337 |
| $1987{ }^{2}$ | 20 | 14 | 140 | 19 | 6 | 2 | 1 | 2 | 0 | 204 |
| $1988{ }^{2}$ | 33 | 23 | 82 | 77 | 7 | 3 | 2 | 2 | 0 | 229 |
| 1989 | 556 | 225 | 24 | 72 | 17 | 2 | 2 | 8 | 4 | 910 |
| 1990 | 184 | 820 | 59 | 65 | 111 | 23 | 15 | 7 | 3 | 1287 |


| Year | Length group (cm) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 5.0-9.9 | 10.0-14.9 | 15.0-19.9 | 20.0-24.9 | 25.0-29.9 | 30.0-34.9 | 35.0-39.9 | 40.0-44.9 | > 45.0 | Total |
| 1991 | 1533 | 1426 | 563 | 55 | 138 | 38 | 30 | 7 | 1 | 3791 |
| 1992 | 149 | 446 | 268 | 43 | 22 | 15 | 4 | 7 | 4 | 958 |
| 1993 | 9 | 320 | 272 | 89 | 16 | 13 | 3 | 1 | 0 | 723 |
| 1994 | 4 | 284 | 613 | 242 | 10 | 9 | 2 | 2 | 1 | 1167 |
| 1995 | 33 | 33 | 417 | 349 | 77 | 18 | 5 | 1 | 0 | 933 |
| 1996 | 56 | 69 | 139 | 310 | 97 | 8 | 4 | 1 | 1 | 685 |
| 1997 | 3 | 44 | 13 | 65 | 57 | 9 | 5 | 0 | 0 | 195 |
| 1998 | 0 | 37 | 35 | 28 | 132 | 73 | 45 | 2 | 0 | 352 |
| 1999 | 3 | 3 | 124 | 62 | 260 | 169 | 42 | 1 | 0 | 664 |
| 2000 | 0 | 10 | 30 | 59 | 126 | 143 | 21 | 1 | 0 | 391 |
| 2001 | 1 | 5 | 3 | 32 | 57 | 227 | 50 | 3 | 0 | 378 |
| 2002 | 1 | 4 | 6 | 21 | 62 | 266 | 47 | 4 | 0 | 410 |
| 2003 | 1 | 5 | 7 | 11 | 51 | 244 | 45 | 1 | 0 | 364 |
| 2004 | 0 | 2 | 8 | 6 | 14 | 78 | 49 | 2 | 0 | 160 |
| 2005 | 22 | 1 | 4 | 4 | 10 | 70 | 47 | 1 | 0 | 158 |
| 2006 | 85 | 6 | 5 | 7 | 43 | 200 | 108 | 3 | 0 | 457 |
| 2007 | 97 | 68 | 1 | 5 | 11 | 102 | 119 | 3 | 0 | 406 |


| Year | Length group (cm) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 5.0-9.9 | 10.0-14.9 | 15.0-19.9 | 20.0-24.9 | 25.0-29.9 | 30.0-34.9 | 35.0-39.9 | 40.0-44.9 | > 45.0 | Total |
| 2008 | 124 | 47 | 22 | 3 | 8 | 22 | 70 | 3 | 0 | 299 |
| 2009 | 9 | 122 | 88 | 14 | 3 | 27 | 219 | 5 | 0 | 486 |
| 2010 | 96 | 18 | 44 | 37 | 2 | 20 | 91 | 7 | 0 | 315 |
| 2011 | 126 | 91 | 81 | 48 | 10 | 7 | 67 | 5 | 1 | 436 |
| 2012 | 29 | 71 | 65 | 77 | 47 | 8 | 94 | 10 | 0 | 400 |
| 2013 | 33 | 43 | 127 | 106 | 67 | 19 | 89 | 13 | 0 | 497 |
| $2014{ }^{3}$ | 3 | 10 | 59 | 49 | 38 | 24 | 66 | 20 | 0 | 268 |
| 2015 | 85 | 7 | 28 | 157 | 115 | 65 | 69 | 25 | 0 | 552 |
| 2016 | 244 | 33 | 44 | 205 | 138 | 139 | 142 | 48 | 0 | 993 |
| 2017 | 41 | 39 | 8 | 20 | 59 | 76 | 57 | 17 | 0 | 317 |
| 2018 | 66 | 62 | 55 | 35 | 100 | 65 | 80 | 26 | 0 | 489 |
| 2019 | 3 | 25 | 84 | 31 | 59 | 82 | 72 | 25 | 1 | 381 |
| 2020 | 97 | 8 | 57 | 39 | 40 | 115 | 97 | 16 | 0 | 470 |
| 2021 | 492 | 135 | 15 | 39 | 16 | 58 | 88 | 18 | 0 | 860 |

1 - Includes some unidentified Sebastes specimens mostly less than 15 cm .
2 - Old trawl equipment (bobbins gear and 80 m sweep length)
3 - Poor survey coverage in 2014.

Table 6.16b. S. mentella ${ }^{1}$ in Division 2.b. Norwegian bottom trawl survey indices (on age) in the Svalbard area (Division 2.b) in summer/autumn 1992-2019 (numbers in millions).

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 283 | 419 | 484 | 131 | 58 | 45 | 14 | 8 | 5 | 2 | 7 | 2 | 1 | 3 | 1462 |
| 1993 | 2 | 527 | 117 | 202 | 142 | 8 | 23 | 6 | 13 | 1 | 7 | 1 | 1 | 0 | 1050 |
| 1994 | 7 | 280 | 290 | 202 | 235 | 42 | 94 | 1 | 1 | 3 | 4 | 1 | 1 | 0 | 1161 |
| 1995 | 4 | 50 | 365 | 237 | 132 | 61 | 19 | 17 | 11 | 0 | 1 | 3 | 0 | 0 | 900 |
| 1996 | 13 | 32 | 10 | 36 | 103 | 135 | 78 | 16 | 50 | 28 | 32 | 8 | 21 | 2 | 565 |
| 1997 | 8 | 43 | 6 | 7 | 38 | 18 | 29 | 19 | 6 | 2 | 0 | 2 | 1 | 1 | 181 |
| 1998 | 0 | 25 | 27 | 13 | 10 | 12 | 61 | 52 | 41 | 15 | 0 | 5 | 13 | 0 | 276 |
| 1999 | 3 | 16 | 108 | 25 | 28 | 39 | 106 | 59 | 54 | 26 | 35 | 14 | 18 | 12 | 543 |
| 2000 | 4 | 6 | 5 | 13 | 30 | 21 | 28 | 44 | 66 | 48 | 21 | 19 | 9 | 6 | 321 |
| 2001 | 1 | 4 | 2 | 0 | 12 | 15 | 18 | 36 | 28 | 46 | 45 | 80 | 53 | 14 | 354 |
| 2002 | 3 | 2 | 4 | 1 | 5 | 22 | 34 | 23 | 90 | 35 | 54 | 65 | 17 | 22 | 377 |
| 2003 | 0 | 4 | 3 | 3 | 5 | 3 | 29 | 25 | 25 | 25 | 11 | 164 | 55 | 23 | 376 |
| 2004 | 1 | 1 | 4 | 4 | 1 | 4 | 2 | 9 | 4 | 15 | 14 | 17 | 15 | 15 | 108 |
| 2005 | 15 | 1 | 1 | 3 | 1 | 2 | 2 | 8 | 4 | 5 | 14 | 7 | 30 | 21 | 115 |
| 2006 | 35 | 1 | 3 | 3 | 2 | 6 | 5 | 37 | 3 | 20 | 46 | 69 | 8 | 22 | 258 |
| 2007 | 28 | 39 | 0 | 0 | 4 | 1 | 5 | 5 | 7 | 5 | 3 | 7 | 28 | 17 | 150 |
| 2008 | 6 | 24 | 19 | 11 | 3 | 2 | 2 | 4 | 3 | 3 | 3 | 3 | 6 | 8 | 96 |
| 2009 | 9 | 69 | 50 | 29 | 26 | 25 | 7 | 1 | 1 | 1 | 4 | 20 | 11 | 8 | 260 |


| Year/Age | 23 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | No age readings available |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2011 | 125 42 | 61 | 42 | 12 | 49 | 31 | 4 | 1 | 0 | 2 | 0 | 0 | 1 | 369 |
| 2012 | 2754 | 32 | 27 | 34 | 43 | 26 | 34 | 18 | 9 | 0 | 1 | 0 | 0 | 305 |
| 2013 | $30 \quad 4$ | 29 | 36 | 7 | 93 | 72 | 43 | 40 | 7 | 8 | 3 | 3 | 3 | 377 |
| 2014,3 | 03 | 2 | 7 | 21 | 40 | 13 | 27 | 5 | 30 | 13 | 11 | 3 | 2 | 176 |
| 2015 | 63 1 | 10 | 56 | 36 | 54 | 33 | 95 | 28 | 21 | 12 | 4 | 5 | 3 | 421 |
| 2016 | No age readings available |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2017 | $39 \quad 26$ | 10 | 13 | 14 | 20 | 39 | 16 | 29 | 8 | 6 | 19 | 1 | 28 | 269 |
| 2018 | No age readings available |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2019 | 032 | 53 | 0 | 24 | 21 | 21 | 46 | 52 | 76 | 0 | 0 | 0 | 0 | 324 |
| 2020 | No age readings available |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2021 | No age readings available |  |  |  |  |  |  |  |  |  |  |  |  |  |

1 - Includes some unidentified Sebastes specimens mostly less than 15 cm .
2 - Old trawl equipment (bobbins gear and 80 m sweep length).
3 - Poor survey coverage in 2014.

Table 6.17. S. mentella in subareas 1 and 2. Abundance indices (on age) from the Ecosystem survey in August-September 1996-2021 covering the Norwegian Economic Zone (NEZ) and Svalbard incl. the area north and east of Spitsbergen (numbers in thousands and total biomass in thousand tonnes) and the continental slope down to 1000 m .

| Year/ age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ | Total | N | Total B |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 146198 | 112742 | 22353 | 53507 | 165531 | 181980 | 108738 | 43328 | 65310 | 40546 | 38254 | 19843 | 29446 | 10931 | 17414 | 1056120 |  | 171 |


| Year/ age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ | Total | N Total B |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 62682 | 130816 | 12492 | 23452 | 74342 | 55880 | 76607 | 82503 | 17640 | 14274 | 675 | 2238 | 1723 | 633 | 8765 | 564723 | 73 |
| 1998 | 313 | 78767 | 85715 | 39849 | 25805 | 23413 | 84825 | 100332 | 54287 | 24329 | 11334 | 7457 | 15250 | 576 | 25212 | 577464 | 105 |
| 1999 | 5359 | 23240 | 117170 | 47851 | 41608 | 76797 | 128677 | 73306 | 58018 | 64781 | 49890 | 13565 | 18458 | 12171 | 24672 | 755562 | 155 |
| 2000 | 5964 | 23169 | 14336 | 19960 | 52666 | 68081 | 83857 | 77513 | 100442 | 72294 | 71148 | 36599 | 17183 | 20590 | 26501 | 690304 | 178 |
| 2001 | 5026 | 6541 | 10957 | 1093 | 19766 | 25591 | 36594 | 51644 | 44407 | 61704 | 50083 | 86122 | 53952 | 15699 | 31877 | 501057 | 162 |
| 2002 | 9112 | 6646 | 7379 | 3821 | 8635 | 28215 | 47456 | 63903 | 103368 | 49964 | 76133 | 71970 | 25241 | 36765 | 34957 | 573565 | 181 |
| 2003 | 4086 | 8218 | 7368 | 3140 | 7885 | 7983 | 43821 | 62360 | 52015 | 34782 | 61735 | 168703 | 107298 | 39760 | 26882 | 636036 | $257{ }^{2}$ |
| 2004 | 8554 | 15793 | 11443 | 7399 | 3554 | 7560 | 6164 | 11686 | 8566 | 22973 | 25920 | 23199 | 20392 | 19472 | 50960 | 243635 | $91^{2}$ |
| 2005 | 32526 | 6856 | 5546 | 5616 | 3772 | 5980 | 6985 | 13151 | 5803 | 5700 | 16554 | 34393 | 34987 | 34336 | 53165 | 265370 | $101^{2}$ |
| 2006 | 125437 | 4833 | 6844 | 6602 | 4255 | 8486 | 7424 | 38309 | 3983 | 24756 | 48733 | 71491 | 13957 | 37991 | 159909 | 563010 | $199{ }^{2}$ |
| 2007 | 411738 | 213851 | 15844 | 5121 | 11830 | 3234 | 8884 | 10298 | 14652 | 7217 | 4200 | 7925 | 53657 | 19308 | 237861 | 1025620 | $199{ }^{2}$ |
| 2008 | 58894 | 206727 | 142254 | 29386 | 7745 | 3182 | 2895 | 6352 | 6132 | 3538 | 3445 | 5380 | 7018 | 9717 | 95279 | 587944 | $84^{2}$ |
| 2009 | 122459 | 176405 | 231265 | 82701 | 109509 | 45607 | 15812 | 2775 | 5807 | 2950 | 3929 | 22097 | 12431 | 9299 | 331974 | 1175019 | $260^{2}$ |
| 2010 | No age reading |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2011 | 422533 | 390888 | 227693 | 61575 | 56025 | 78022 | 47213 | 12153 | 3176 | 2049 | 2607 | 856 | 85 | 2948 | 103653 | 1411479 | $120^{2}$ |
| 2012 | 353610 | 256305 | 351327 | 173183 | 130446 | 70403 | 58164 | 40645 | 21408 | 12671 | 3553 | 1044 | 1568 | 3374 | 139887 | 1617588 | $184{ }^{2}$ |
| 2013 | 299841 | 203094 | 189851 | 194068 | 164206 | 178236 | 112427 | 103262 | 92160 | 13848 | 13956 | 8579 | 2784 | 2857 | 144033 | 1723202 | $271{ }^{2}$ |
| $2014{ }^{1}$ | 2247 | 20884 | 33295 | 82052 | 52428 | 94324 | 93771 | 68765 | 35193 | 56728 | 40647 | 19047 | 16518 | 3335 | 163869 | 783104 | $239{ }^{2}$ |


| $\begin{gathered} \text { Year/ } \\ \text { age } \end{gathered}$ | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ | Total | N | Total B |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 404973 | 86648 | 53046 | 95737 | 53022 | 109686 | 46714 | 126156 | 73141 | 25441 | 19583 | 6569 | 5284 | 3335 | 119261 | 1228596 |  | $207^{2}$ |
| 2016 | No age reading |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2017 | 534647 | 244469 | 213984 | 215852 | 33595 | 45809 | 61428 | 62449 | 37597 | 33901 | 39670 | 37492 | 10364 | 40052 | 85250 | 1696557 |  | $213^{2}$ |
| 2018 | No age reading |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $2019{ }^{3}$ | 93518 | 77195 | 125457 | 81499 | 62447 | 38668 | 61615 | 91672 | 178887 | 124876 | 0 | 0 | 0 | 0 | 60931 | 996765 |  | $211^{2}$ |
| 2020 | No age reading |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2021 | No age reading |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

1 - Poor survey coverage in 2014.
2 - Calculated using modelled weight-at-age.
3 - Provisional figures.

Table 6.18a. S. mentella ${ }^{1}$. Abundance indices (on length) from the bottom trawl survey in the Barents Sea in winter 1986-2021 (numbers in millions). The area coverage was extended from 1993 onwards. Numbers from 1994 onwards were recalculated while numbers for 1986-1993 are as in previous reports.

| Year | Length group (cm) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 5.0-9.9 | 10.0-14.9 | 15.0-19.9 | 20.0-24.9 | 25.0-29.9 | 30.0-34.9 | 35.0-39.9 | 40.0-44.9 | > 45.0 | Total |
| 1986 | 81 | 152 | 205 | 88 | 169 | 130 | 88 | 24 | 14 | 950 |
| 1987 | 72 | 25 | 227 | 56 | 35 | 11 | 5 | 1 | 0 | 433 |
| 1988 | 587 | 25 | 133 | 182 | 40 | 50 | 48 | 4 | 0 | 1068 |
| 1989 | 623 | 55 | 28 | 177 | 58 | 9 | 8 | 2 | 0 | 961 |
| 1990 | 324 | 305 | 36 | 56 | 80 | 13 | 13 | 2 | 0 | 828 |


| Year | Length group (cm) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 5.0-9.9 | 10.0-14.9 | 15.0-19.9 | 20.0-24.9 | 25.0-29.9 | 30.0-34.9 | 35.0-39.9 | 40.0-44.9 | > 45.0 | Total |
| 1991 | 395 | 449 | 86 | 39 | 96 | 35 | 24 | 3 | 0 | 1127 |
| 1992 | 139 | 367 | 227 | 35 | 55 | 34 | 8 | 2 | 1 | 867 |
| 1993 | 31 | 593 | 320 | 116 | 24 | 25 | 6 | 1 | 0 | 1117 |
| 1994 | 8 | 296 | 479 | 488 | 74 | 74 | 17 | 3 | 0 | 1440 |
| 1995 | 310 | 84 | 571 | 390 | 83 | 58 | 24 | 3 | 0 | 1522 |
| 1996 | 215 | 101 | 198 | 343 | 136 | 42 | 17 | 1 | 0 | 1054 |
| $1997{ }^{2}$ | 38 | 83 | 19 | 198 | 266 | 82 | 39 | 3 | 0 | 728 |
| $1998{ }^{2}$ | 1 | 87 | 62 | 101 | 202 | 40 | 13 | 2 | 0 | 507 |
| 1999 | 2 | 7 | 70 | 37 | 172 | 73 | 22 | 3 | 0 | 386 |
| 2000 | 9 | 13 | 40 | 78 | 143 | 97 | 27 | 7 | 2 | 415 |
| 2001 | 10 | 23 | 7 | 57 | 79 | 75 | 10 | 1 | 0 | 260 |
| 2002 | 17 | 7 | 19 | 36 | 96 | 116 | 24 | 1 | 0 | 317 |
| 2003 | 4 | 4 | 10 | 13 | 70 | 198 | 46 | 6 | 0 | 351 |
| 2004 | 2 | 3 | 7 | 19 | 33 | 86 | 32 | 2 | 0 | 183 |
| 2005 | 0 | 6 | 7 | 11 | 28 | 154 | 86 | 4 | 0 | 296 |
| 2006 | 100 | 2 | 10 | 15 | 23 | 104 | 83 | 3 | 1 | 339 |
| 2007 | 382 | 121 | 3 | 7 | 12 | 121 | 121 | 7 | 0 | 773 |
| 2008 | 858 | 359 | 27 | 5 | 12 | 104 | 165 | 5 | 0 | 1533 |


| Year | Length group (cm) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 5.0-9.9 | 10.0-14.9 | 15.0-19.9 | 20.0-24.9 | 25.0-29.9 | 30.0-34.9 | 35.0-39.9 | 40.0-44.9 | > 45.0 | Total |
| 2009 | 95 | 325 | 136 | 5 | 9 | 67 | 163 | 6 | 0 | 806 |
| 2010 | 652 | 276 | 215 | 64 | 7 | 74 | 191 | 6 | 0 | 1485 |
| 2011 | 501 | 230 | 212 | 149 | 14 | 47 | 157 | 5 | 0 | 1315 |
| 2012 | 129 | 280 | 86 | 125 | 47 | 14 | 154 | 18 | 0 | 855 |
| 2013 | 249 | 227 | 245 | 159 | 143 | 35 | 193 | 27 | 0 | 1279 |
| 2014 | 91 | 174 | 250 | 114 | 125 | 51 | 115 | 14 | 0 | 933 |
| 2015 | 175 | 110 | 215 | 302 | 290 | 215 | 171 | 18 | 0 | 1495 |
| 2016 | 615 | 105 | 149 | 332 | 213 | 163 | 124 | 14 | 1 | 1714 |
| 2017 | 568 | 185 | 68 | 197 | 286 | 310 | 231 | 11 | 0 | 1855 |
| 2018 | 189 | 250 | 83 | 109 | 192 | 270 | 214 | 22 | 1 | 1329 |
| 2019 | 42 | 288 | 263 | 92 | 158 | 255 | 211 | 20 | 0 | 1330 |
| 2020 | 196 | 122 | 207 | 92 | 118 | 231 | 209 | 25 | 1 | 1200 |
| 2021 | 887 | 132 | 142 | 124 | 81 | 186 | 172 | 23 | 1 | 1749 |
| $2022^{3}$ | 640 | 1025 | 45 | 104 | 76 | 87 | 153 | 20 | 0 | 2149 |

1 - Includes some unidentified Sebastes specimens mostly less than 15 cm .
2 - Adjusted indices to account for not covering the Russian EEZ in Subarea 1.
3- Russian data not provided in time for AFWG 2022.

Table 6.18b. S. mentella ${ }^{1}$ in subareas 1 and 2. Preliminary Norwegian bottom trawl indices (on age) from the annual Barents Sea survey in February 1992-2020 (numbers in millions). The area coverage was extended from 1993 onwards. Numbers recalculated.

| Year/Age | 2 |  | 3 |  | 4 |  | 5 |  | 6 |  | 7 |  | 8 |  | 9 |  | 10 |  | 11 |  | 12 |  | 13 |  | 14 |  | 15 |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1993 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1994 |  | 5 |  | 96 |  | 315 |  | 160 |  | 342 |  | 269 |  | 97 |  | 55 |  | 4 |  | 28 |  | 13 |  | 14 |  | 26 |  | 5 |  | 1430 |
| 1995 |  | 315 |  | 49 |  | 148 |  | 251 |  | 343 |  | 238 |  | 67 |  | 25 |  | 7 |  | 19 |  | 21 |  | 9 |  | 11 |  | 10 |  | 1512 |
| 1996 |  | 189 |  | 107 |  | 85 |  | 111 |  | 140 |  | 132 |  | 128 |  | 60 |  | 21 |  | 24 |  | 14 |  | 6 |  | 9 |  | 4 |  | 1029 |
| $1997{ }^{2}$ |  | 41 |  | 65 |  | 30 |  | 33 |  | 92 |  | 83 |  | 103 |  | 100 |  | 30 |  | 67 |  | 29 |  | 13 |  | 7 |  | 3 |  | 697 |
| $1998{ }^{2}$ |  | 1 |  | 72 |  | 45 |  | 25 |  | 11 |  | 50 |  | 108 |  | 112 |  | 36 |  | 17 |  | 7 |  | 6 |  | 3 |  | 2 |  | 496 |
| 1999 |  | 0 |  | 1 |  | 38 |  | 40 |  | 29 |  | 28 |  | 52 |  | 62 |  | 55 |  | 32 |  | 16 |  | 4 |  | 7 |  | 1 |  | 364 |
| 2000 |  | 19 |  | 1 |  | 4 |  | 33 |  | 37 |  | 21 |  | 30 |  | 69 |  | 72 |  | 49 |  | 22 |  | 14 |  | 10 |  | 4 |  | 385 |
| 2001 |  | 1 |  | 17 |  | 8 |  | 2 |  | 7 |  | 25 |  | 36 |  | 30 |  | 41 |  | 18 |  | 22 |  | 28 |  | 5 |  | 3 |  | 243 |
| 2002 |  | 18 |  | 4 |  | 11 |  | 8 |  | 2 |  | 9 |  | 43 |  | 56 |  | 23 |  | 14 |  | 34 |  | 19 |  | 38 |  | 14 |  | 293 |
| 2003 |  | 0 |  | 3 |  | 2 |  | 4 |  | 6 |  | 6 |  | 15 |  | 36 |  | 24 |  | 24 |  | 43 |  | 36 |  | 62 |  | 33 |  | 293 |
| 2004 |  | 2 |  | 1 |  | 4 |  | 2 |  | 4 |  | 10 |  | 11 |  | 16 |  | 14 |  | 12 |  | 14 |  | 25 |  | 24 |  | 13 |  | 152 |
| 2005 |  | 0 |  | 4 |  | 3 |  | 2 |  | 6 |  | 6 |  | 7 |  | 14 |  | 18 |  | 8 |  | 18 |  | 27 |  | 40 |  | 57 |  | 208 |
| 2006 |  | 74 |  | 26 |  | 4 |  | 4 |  | 6 |  | 8 |  | 9 |  | 12 |  | 6 |  | 14 |  | 16 |  | 10 |  | 41 |  | 28 |  | 259 |
| 2007 |  | 237 |  | 75 |  | 4 |  | 1 |  | 2 |  | 2 |  | 5 |  | 8 |  | 9 |  | 6 |  | 8 |  | 21 |  | 33 |  | 72 |  | 485 |
| 2008 |  | 699 |  | 166 |  | 101 |  | 14 |  | 0 |  | 2 |  | 4 |  | 6 |  | 4 |  | 6 |  | 4 |  | 20 |  | 22 |  | 30 |  | 1079 |


| Year/Age | 2 | 3 |  | 4 |  | 5 |  | 6 |  | 7 |  | 8 |  | 9 |  | 10 |  | 11 |  | 12 |  | 13 |  | 14 |  | 15 |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 |  | 104 | 108 |  | 100 |  | 87 |  | 64 |  | 32 |  | 19 |  | 14 |  | 4 |  | 6 |  | 21 |  | 1 |  | 22 |  | 7 | 589 |
| 2010 |  | 160 | 264 |  | 176 |  | 166 |  | 93 |  | 72 |  | 24 |  | 23 |  | 3 |  | 11 |  | 5 |  | 8 |  | 10 |  | 17 | 1031 |
| 2011 |  | 348 | 228 |  | 128 |  | 127 |  | 99 |  | 67 |  | 42 |  | 20 |  | 2 |  | 6 |  | 1 |  | 1 |  | 2 |  | 25 | 1095 |
| 2012 |  | No age readings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2013 |  | 0 | 179 |  | 268 |  | 136 |  | 154 |  | 108 |  | 126 |  | 14 |  | 31 |  | 8 |  | 7 |  | 20 |  | 41 |  | 112 | 1105 |
| 2014 |  | No age readings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2015 |  | No age readings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2016 |  | No age readings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2017 |  | No age readings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2018 |  | No age readings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2019 |  | No age readings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2020 |  | No age readings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2021 |  | No age reading |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2022 |  | No age reading |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

1 - Includes some unidentified Sebastes specimens mostly less than 15 cm .
2 - Adjusted indices to account for not covering the Russian EEZ in Subarea 1.

Table 6.19. Comparison of results on S. mentella from the Norwegian Sea pelagic surveys in 2008, 2009, 2013, 2016, and 2019. Acoustic results for the 2019 survey were not available at the time of AFWG 2021.

|  | 2008 | 2009 | 2013 | 2016 | 2019 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| mean length (cm) All/M/F ${ }^{1}$ | 37.0/36.4/37.5 | 36.6/36.0/37.1 | 37.5/37.0/38.1 | 37.7/37.0/38.3 | 37.6/37.2/38.0 |
| mean length (cm) S/DSL/D2 | 37.2/36.8/39.1 | 37.2/36.5/38.3 | 37.1/37.4/38.9 | 38.1/37.6/38.4 | 37.4/37.6/37.7 |
| mean weight (g) All/M/F | 619/585/648 | 625/609/666 | 659/625/706 | 656/619/694 | 683/644/724 |
| Mean age (y) All/M/F | 25/25/25 | 25/25/24 | 28/29/28 | 27/27/26 | -/-/- |
| Sex ratio (M/F) | 45\% / 55\% | 45\% / 55\% | 59\% / 41\% | 50\% / 50\% | 51\% / 49\% |
| Occurrence | 96\% | 100\% | 95\% | 80\% | 99\% |
| Catch rates | 3.80 t/NM2 | 3.94 t/NM2 | 3.47 t/NM2 | 1.01 t/NM2 | 3.40 t/NM2 |
| mean $\mathrm{s}_{\mathrm{A}}$ | $33 \mathrm{~m}^{2} / \mathrm{NM} 2$ | $34 \mathrm{~m}^{2} / \mathrm{NM} 2$ | $19 \mathrm{~m}^{2} / \mathrm{NM} 2$ | $5.2 \mathrm{~m}^{2} / \mathrm{NM} 2$ | - |
| Total Area | 53720 NM2 | 69520 NM2 | 69520 NM2 | 67150 NM2 | 73364 NM2 |
| Abundance (Acoustics) ${ }^{3}$ | 395000 t | 532000 t | 297000 t | 136000 t | - |
| Abundance (Trawl) ${ }^{4}$ | 406000 t | 548000 t | 482000 t | 116000 t | 499000 t |

$1-\mathrm{M}=$ males only, $\mathrm{F}=$ females only.
2 - S = shallower than DSL, DSL = deep scattering layer, $D=$ deeper than DSL.
3 - The abundance derived from hydroacoustics is calculated assuming a Length-dependent target strength equation of TS=20log(L)-68.0. In 2016 the TS equation used was TS=20log(L)69.6 following recommendation from ICES-WKTAR (2010).

4 - Trawls: Gloria 2048 in 2008 and 2009 Gloria 2560 HO helix in 2013 and Gloria 1024 in 2016. Trawl catchability for redfish set to 0.5 for all trawls based on results from Bethke et al. (2010).

Table 6.20a. S. mentella in subareas 1 and 2. Population matrix with numbers-at-age (in thousands) for each year and separable fishing mortality coefficients for the demersal and pelagic fleet by year (Fy) and selectivity at age for the pelagic fleet (Sa). Numbers are estimated from the statistical catch-at-age model.

| sa (demersal) | Varies over time |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| sa (pelagic) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.011 | 0.021 | 0.040 | 0.072 | 0.128 | 0.218 | 0.345 | 0.500 | 0.654 | 0.781 | 0.871 | 0.927 | 1.00 |


| Fy (demseral) | Fy (pelagic) | Year/ <br> Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.047 | 0 | 1992 | $\begin{aligned} & 40048 \\ & 2 \end{aligned}$ | $\begin{aligned} & 38974 \\ & 1 \end{aligned}$ | $\begin{aligned} & 35323 \\ & 0 \end{aligned}$ | $\begin{aligned} & 22634 \\ & 7 \end{aligned}$ | $\begin{aligned} & 13548 \\ & 8 \end{aligned}$ | 92234 | 89791 | 93917 | $\begin{aligned} & 11766 \\ & 3 \end{aligned}$ | 81823 | 93165 | 68973 | 70343 | 61008 | 44144 | 28865 | 19263 | $\begin{aligned} & 18959 \\ & 4 \end{aligned}$ |
| 0.033 | 0 | 1993 | $\begin{aligned} & 26963 \\ & 5 \end{aligned}$ | $\begin{aligned} & 38103 \\ & 1 \end{aligned}$ | $\begin{aligned} & 37081 \\ & 2 \end{aligned}$ | $\begin{aligned} & 33607 \\ & 4 \end{aligned}$ | $\begin{aligned} & 21261 \\ & 6 \end{aligned}$ | $\begin{aligned} & 12702 \\ & 6 \end{aligned}$ | 86296 | 83827 | 87482 | $\begin{aligned} & 10934 \\ & 9 \end{aligned}$ | 75866 | 86189 | 63671 | 64803 | 56096 | 40520 | 26454 | $\begin{aligned} & 18979 \\ & 0 \end{aligned}$ |
| 0.029 | 0 | 1994 | $\begin{aligned} & 18668 \\ & 4 \end{aligned}$ | $\begin{aligned} & 25654 \\ & 0 \end{aligned}$ | $\begin{aligned} & 36252 \\ & 5 \end{aligned}$ | $\begin{aligned} & 35280 \\ & 2 \end{aligned}$ | $\begin{aligned} & 31969 \\ & 3 \end{aligned}$ | $\begin{aligned} & 20218 \\ & 5 \end{aligned}$ | $\begin{aligned} & 12068 \\ & 4 \end{aligned}$ | 81795 | 79052 | 81844 | $\begin{aligned} & 10150 \\ & 9 \end{aligned}$ | 70092 | 79457 | 58648 | 59674 | 51651 | 37307 | $\begin{aligned} & 19909 \\ & 5 \end{aligned}$ |
| 0.022 | 0 | 1995 | $\begin{aligned} & 17688 \\ & 0 \end{aligned}$ | $\begin{aligned} & 17761 \\ & 8 \end{aligned}$ | $\begin{aligned} & 24408 \\ & 0 \end{aligned}$ | $\begin{aligned} & 34491 \\ & 8 \end{aligned}$ | $\begin{aligned} & 33543 \\ & 5 \end{aligned}$ | $\begin{aligned} & 30366 \\ & 3 \end{aligned}$ | $\begin{aligned} & 19163 \\ & 8 \end{aligned}$ | $\begin{aligned} & 11392 \\ & 2 \end{aligned}$ | 76744 | 73703 | 75960 | 93978 | 64817 | 73442 | 54197 | 55140 | 47725 | $\begin{aligned} & 21843 \\ & 1 \end{aligned}$ |
| 0.015 | 0 | 1996 | $\begin{aligned} & 14161 \\ & 9 \end{aligned}$ | $\begin{aligned} & 16828 \\ & 9 \end{aligned}$ | $\begin{aligned} & 16899 \\ & 1 \end{aligned}$ | $\begin{aligned} & 23222 \\ & 6 \end{aligned}$ | $\begin{aligned} & 32794 \\ & 9 \end{aligned}$ | $\begin{aligned} & 31866 \\ & 2 \end{aligned}$ | $\begin{aligned} & 28796 \\ & 7 \end{aligned}$ | $\begin{aligned} & 18115 \\ & 2 \end{aligned}$ | $\begin{aligned} & 10720 \\ & 6 \end{aligned}$ | 71896 | 68826 | 70808 | 87530 | 60347 | 68366 | 50448 | 51324 | $\begin{aligned} & 24773 \\ & 1 \end{aligned}$ |
| 0.015 | 0 | 1997 | $\begin{aligned} & 10033 \\ & 1 \end{aligned}$ | $\begin{aligned} & 13474 \\ & 1 \end{aligned}$ | $\begin{aligned} & 16011 \\ & 6 \end{aligned}$ | $16078$ | $\begin{aligned} & 22089 \\ & 0 \end{aligned}$ | $\begin{aligned} & 31179 \\ & 0 \end{aligned}$ | $\begin{aligned} & 30257 \\ & 3 \end{aligned}$ | $\begin{aligned} & 27267 \\ & 1 \end{aligned}$ | $\begin{aligned} & 17083 \\ & 5 \end{aligned}$ | $\begin{aligned} & 10074 \\ & 7 \end{aligned}$ | 67437 | 64508 | 66347 | 82008 | 56538 | 64050 | 47263 | $\begin{aligned} & 28017 \\ & 5 \end{aligned}$ |
| 0.021 | 0 | 1998 | 51116 | 95458 | $\begin{aligned} & 12819 \\ & 7 \end{aligned}$ | $\begin{aligned} & 15233 \\ & 9 \end{aligned}$ | $\begin{aligned} & 15294 \\ & 2 \end{aligned}$ | $\begin{aligned} & 21003 \\ & 3 \end{aligned}$ | $\begin{aligned} & 29614 \\ & 8 \end{aligned}$ | $\begin{aligned} & 28670 \\ & 3 \end{aligned}$ | $\begin{aligned} & 25739 \\ & 1 \end{aligned}$ | $\begin{aligned} & 16068 \\ & 4 \end{aligned}$ | 94563 | 63242 | 60476 | 62193 | 76869 | 52995 | 60036 | $\begin{aligned} & 30691 \\ & 7 \end{aligned}$ |
| 0.016 | 0 | 1999 | 44153 | 48634 | 90822 | $\begin{aligned} & 12197 \\ & 1 \end{aligned}$ | $\begin{aligned} & 14492 \\ & 4 \end{aligned}$ | $\begin{aligned} & 14544 \\ & 1 \end{aligned}$ | $\begin{aligned} & 19941 \\ & 9 \end{aligned}$ | $\begin{aligned} & 27982 \\ & 8 \end{aligned}$ | $\begin{aligned} & 26889 \\ & 8 \end{aligned}$ | $\begin{aligned} & 24032 \\ & 2 \end{aligned}$ | $\begin{aligned} & 14981 \\ & 6 \end{aligned}$ | 88137 | 58940 | 56361 | 57961 | 71639 | 49389 | $\begin{aligned} & 34198 \\ & 3 \end{aligned}$ |
| 0.013 | 0 | 2000 | 34755 | 42009 | 46272 | 86411 | $\begin{aligned} & 11604 \\ & 5 \end{aligned}$ | $\begin{aligned} & 13787 \\ & 2 \end{aligned}$ | $\begin{aligned} & 13831 \\ & 4 \end{aligned}$ | $\begin{aligned} & 18935 \\ & 4 \end{aligned}$ | $\begin{aligned} & 26449 \\ & 1 \end{aligned}$ | $\begin{aligned} & 25269 \\ & 2 \end{aligned}$ | $\begin{aligned} & 22522 \\ & 3 \end{aligned}$ | $\begin{aligned} & 14030 \\ & 3 \end{aligned}$ | 82528 | 55187 | 52772 | 54269 | 67077 | $\begin{aligned} & 36644 \\ & 8 \end{aligned}$ |


| Fy (demseral) | Fy (pelagic) | Year/ <br> Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.022 | 0 | 2001 | 37339 | 33067 | 39968 | 44024 | 82214 | $11040$ | $\begin{aligned} & 13115 \\ & 5 \end{aligned}$ | $\begin{aligned} & 13141 \\ & 0 \end{aligned}$ | $\begin{aligned} & 17889 \\ & 8 \end{aligned}$ | $\begin{aligned} & 24872 \\ & 5 \end{aligned}$ | $\begin{aligned} & 23742 \\ & 9 \end{aligned}$ | $\begin{aligned} & 21160 \\ & 0 \end{aligned}$ | $\begin{aligned} & 13181 \\ & 5 \end{aligned}$ | 77535 | 51848 | 49579 | 50986 | $\begin{aligned} & 40729 \\ & 8 \end{aligned}$ |
| 0.008 | 0 | 2002 | 38941 | 35525 | 31461 | 38027 | 41864 | 78123 | $\begin{aligned} & 10474 \\ & 9 \end{aligned}$ | $\begin{aligned} & 12406 \\ & 4 \end{aligned}$ | $\begin{aligned} & 12375 \\ & 0 \end{aligned}$ | $\begin{aligned} & 16766 \\ & 6 \end{aligned}$ | $\begin{aligned} & 23227 \\ & 9 \end{aligned}$ | $\begin{aligned} & 22128 \\ & 3 \end{aligned}$ | $\begin{aligned} & 19702 \\ & 1 \end{aligned}$ | $\begin{aligned} & 12268 \\ & 1 \end{aligned}$ | 72149 | 48243 | 46130 | $\begin{aligned} & 42639 \\ & 8 \end{aligned}$ |
| 0.003 | 0 | 2003 | 43637 | 37050 | 33800 | 29933 | 36180 | 39827 | 74299 | 99503 | $\begin{aligned} & 11753 \\ & 2 \end{aligned}$ | $\begin{aligned} & 11695 \\ & 2 \end{aligned}$ | $\begin{aligned} & 15831 \\ & 4 \end{aligned}$ | $\begin{aligned} & 21927 \\ & 4 \end{aligned}$ | $\begin{aligned} & 20888 \\ & 4 \end{aligned}$ | $\begin{aligned} & 18597 \\ & 9 \end{aligned}$ | $\begin{aligned} & 11580 \\ & 5 \end{aligned}$ | 68106 | 45540 | $\begin{aligned} & 44604 \\ & 6 \end{aligned}$ |
| 0.006 | 0 | 2004 | 57553 | 41518 | 35251 | 32158 | 28476 | 34415 | 37875 | 70629 | 94536 | $\begin{aligned} & 11160 \\ & 3 \end{aligned}$ | $\begin{aligned} & 11100 \\ & 9 \end{aligned}$ | $\begin{aligned} & 15023 \\ & 3 \end{aligned}$ | $\begin{aligned} & 20805 \\ & 6 \end{aligned}$ | $\begin{aligned} & 19818 \\ & 7 \end{aligned}$ | $\begin{aligned} & 17645 \\ & 0 \end{aligned}$ | $\begin{aligned} & 10987 \\ & 1 \end{aligned}$ | 64615 | $\begin{aligned} & 46638 \\ & 9 \end{aligned}$ |
| 0.009 | 0 | 2005 | $\begin{aligned} & 13268 \\ & 2 \end{aligned}$ | 54758 | 39501 | 33539 | 30594 | 27087 | 32725 | 35992 | 67041 | 89612 | $\begin{aligned} & 10567 \\ & 3 \end{aligned}$ | $\begin{aligned} & 10504 \\ & 3 \end{aligned}$ | $\begin{aligned} & 14211 \\ & 5 \end{aligned}$ | $\begin{aligned} & 19678 \\ & 9 \end{aligned}$ | $\begin{aligned} & 18744 \\ & 4 \end{aligned}$ | $\begin{aligned} & 16688 \\ & 2 \end{aligned}$ | $\begin{aligned} & 10391 \\ & 2 \end{aligned}$ | $\begin{aligned} & 50220 \\ & 3 \end{aligned}$ |
| 0.005 | 0.037 | 2006 | $\begin{aligned} & 23245 \\ & 0 \end{aligned}$ | $\begin{aligned} & 12623 \\ & 8 \end{aligned}$ | 52099 | 37583 | 31908 | 29103 | 25760 | 31100 | 34148 | 63456 | 84646 | 99708 | 99070 | $\begin{aligned} & 13401 \\ & 4 \end{aligned}$ | $\begin{aligned} & 18556 \\ & 1 \end{aligned}$ | $\begin{aligned} & 17674 \\ & 6 \end{aligned}$ | $\begin{aligned} & 15735 \\ & 7 \end{aligned}$ | $\begin{aligned} & 57151 \\ & 8 \end{aligned}$ |
| 0.005 | 0.02 | 2007 | $\begin{aligned} & 33451 \\ & 4 \end{aligned}$ | $\begin{aligned} & 22116 \\ & 0 \end{aligned}$ | $\begin{aligned} & 12010 \\ & 7 \end{aligned}$ | 49568 | 35757 | 30357 | 27676 | 24483 | 29525 | 32351 | 59910 | 79557 | 93210 | 92061 | $\begin{aligned} & 12381 \\ & 1 \end{aligned}$ | $\begin{aligned} & 17062 \\ & 2 \end{aligned}$ | $\begin{aligned} & 16197 \\ & 6 \end{aligned}$ | $\begin{aligned} & 66517 \\ & 5 \end{aligned}$ |
| 0.005 | 0.014 | 2008 | $\begin{aligned} & 32929 \\ & 0 \end{aligned}$ | $\begin{aligned} & 31826 \\ & 7 \end{aligned}$ | $\begin{aligned} & 21041 \\ & 9 \end{aligned}$ | $\begin{aligned} & 11427 \\ & 4 \end{aligned}$ | 47161 | 34020 | 28875 | 26317 | 23268 | 28028 | 30647 | 56583 | 74873 | 87407 | 86043 | $\begin{aligned} & 11541 \\ & 1 \end{aligned}$ | $\begin{aligned} & 15875 \\ & 4 \end{aligned}$ | $\begin{aligned} & 76782 \\ & 8 \end{aligned}$ |
| 0.003 | 0.01 | 2009 | $\begin{aligned} & 34773 \\ & 1 \end{aligned}$ | $\begin{aligned} & 31329 \\ & 7 \end{aligned}$ | $\begin{aligned} & 30280 \\ & 9 \end{aligned}$ | $\begin{aligned} & 20019 \\ & 9 \end{aligned}$ | $\begin{aligned} & 10872 \\ & 3 \end{aligned}$ | 44870 | 32362 | 27463 | 25018 | 22092 | 26541 | 28937 | 53298 | 70366 | 81967 | 80545 | $\begin{aligned} & 10790 \\ & 3 \end{aligned}$ | $\begin{aligned} & 86491 \\ & 0 \end{aligned}$ |
| 0.004 | 0.011 | 2010 | $\begin{aligned} & 49962 \\ & 1 \end{aligned}$ | $\begin{aligned} & 33084 \\ & 3 \end{aligned}$ | $\begin{aligned} & 29808 \\ & 1 \end{aligned}$ | $\begin{aligned} & 28810 \\ & 3 \end{aligned}$ | $\begin{aligned} & 19047 \\ & 5 \end{aligned}$ | $\begin{aligned} & 10344 \\ & 1 \end{aligned}$ | 42683 | 30778 | 26103 | 23751 | 20941 | 25123 | 27351 | 50295 | 66297 | 77127 | 75720 | $\begin{aligned} & 91342 \\ & 6 \end{aligned}$ |
| 0.006 | 0.01 | 2011 | $\begin{aligned} & 56485 \\ & 4 \end{aligned}$ | $\begin{aligned} & 47535 \\ & 6 \end{aligned}$ | $\begin{aligned} & 31477 \\ & 4 \end{aligned}$ | $\begin{aligned} & 28360 \\ & 4 \end{aligned}$ | $\begin{aligned} & 27410 \\ & 7 \end{aligned}$ | $\begin{aligned} & 18121 \\ & 9 \end{aligned}$ | 98398 | 40592 | 29255 | 24786 | 22518 | 19820 | 23735 | 25792 | 47347 | 62323 | 72432 | $\begin{aligned} & 92768 \\ & 6 \end{aligned}$ |
| 0.005 | 0.01 | 2012 | $\begin{aligned} & 43051 \\ & 9 \end{aligned}$ | $\begin{aligned} & 53742 \\ & 0 \end{aligned}$ | $\begin{aligned} & 45226 \\ & 9 \end{aligned}$ | $\begin{aligned} & 29948 \\ & 6 \end{aligned}$ | $\begin{aligned} & 26982 \\ & 9 \end{aligned}$ | $\begin{aligned} & 26079 \\ & 2 \end{aligned}$ | $\begin{aligned} & 17239 \\ & 2 \end{aligned}$ | 93587 | 38590 | 27780 | 23486 | 21284 | 18693 | 22343 | 24238 | 44435 | 58434 | $\begin{aligned} & 93650 \\ & 7 \end{aligned}$ |


| Fy (demseral) | Fy (pelagic) | Year/ <br> Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.004 | 0.009 | 2013 | $\begin{aligned} & 26696 \\ & 4 \end{aligned}$ | $\begin{aligned} & 40961 \\ & 0 \end{aligned}$ | $\begin{aligned} & 51131 \\ & 9 \end{aligned}$ | $\begin{aligned} & 43030 \\ & 3 \end{aligned}$ | $\begin{aligned} & 28493 \\ & 8 \end{aligned}$ | $\begin{aligned} & 25671 \\ & 9 \end{aligned}$ | $\begin{aligned} & 24808 \\ & 7 \end{aligned}$ | $\begin{aligned} & 16396 \\ & 8 \end{aligned}$ | 88987 | 36672 | 26371 | 22258 | 20126 | 17635 | 21037 | 22787 | 41732 | $\begin{aligned} & 93320 \\ & 5 \end{aligned}$ |
| 0.016 | 0.01 | 2014 | $\begin{aligned} & 25856 \\ & 0 \end{aligned}$ | $\begin{aligned} & 25399 \\ & 8 \end{aligned}$ | $\begin{aligned} & 38971 \\ & 6 \end{aligned}$ | $\begin{aligned} & 48648 \\ & 5 \end{aligned}$ | $\begin{aligned} & 40940 \\ & 4 \end{aligned}$ | $\begin{aligned} & 27109 \\ & 9 \end{aligned}$ | $\begin{aligned} & 24422 \\ & 4 \end{aligned}$ | $\begin{aligned} & 23599 \\ & 0 \end{aligned}$ | $\begin{aligned} & 15594 \\ & 3 \end{aligned}$ | 84602 | 34842 | 25026 | 21082 | 19016 | 16625 | 19798 | 21423 | $\begin{aligned} & 91542 \\ & 3 \end{aligned}$ |
| 0.027 | 0.009 | 2015 | $\begin{aligned} & 36516 \\ & 6 \end{aligned}$ | $\begin{aligned} & 24600 \\ & 2 \end{aligned}$ | $\begin{aligned} & 24166 \\ & 2 \end{aligned}$ | $\begin{aligned} & 37078 \\ & 8 \end{aligned}$ | $\begin{aligned} & 46284 \\ & 4 \end{aligned}$ | $\begin{aligned} & 38949 \\ & 8 \end{aligned}$ | $\begin{aligned} & 25787 \\ & 6 \end{aligned}$ | $\begin{aligned} & 23226 \\ & 7 \end{aligned}$ | $\begin{aligned} & 22435 \\ & 5 \end{aligned}$ | $\begin{aligned} & 14815 \\ & 8 \end{aligned}$ | 80286 | 33001 | 23636 | 19836 | 17820 | 15521 | 18433 | $\begin{aligned} & 86860 \\ & 2 \end{aligned}$ |
| 0.038 | 0.009 | 2016 | $\begin{aligned} & 45110 \\ & 7 \end{aligned}$ | $\begin{aligned} & 34743 \\ & 0 \end{aligned}$ | $\begin{aligned} & 23405 \\ & 4 \end{aligned}$ | $\begin{aligned} & 22992 \\ & 5 \end{aligned}$ | $\begin{aligned} & 35276 \\ & 9 \end{aligned}$ | $\begin{aligned} & 44033 \\ & 1 \end{aligned}$ | $\begin{aligned} & 37046 \\ & 8 \end{aligned}$ | $\begin{aligned} & 24517 \\ & 5 \end{aligned}$ | $\begin{aligned} & 22060 \\ & 5 \end{aligned}$ | $\begin{aligned} & 21260 \\ & 2 \end{aligned}$ | $\begin{aligned} & 13975 \\ & 4 \end{aligned}$ | 75203 | 30686 | 21858 | 18283 | 16392 | 14262 | $\begin{aligned} & 81396 \\ & 7 \end{aligned}$ |
| 0.029 | 0.009 | 2017 | $\begin{aligned} & 51101 \\ & 2 \end{aligned}$ | $\begin{aligned} & 42919 \\ & 8 \end{aligned}$ | $\begin{aligned} & 33055 \\ & 6 \end{aligned}$ | $\begin{aligned} & 22268 \\ & 7 \end{aligned}$ | $\begin{aligned} & 21874 \\ & 6 \end{aligned}$ | $\begin{aligned} & 33558 \\ & 2 \end{aligned}$ | $\begin{aligned} & 41870 \\ & 1 \end{aligned}$ | $\begin{aligned} & 35191 \\ & 9 \end{aligned}$ | $\begin{aligned} & 23229 \\ & 1 \end{aligned}$ | $\begin{aligned} & 20777 \\ & 2 \end{aligned}$ | $\begin{aligned} & 19826 \\ & 0 \end{aligned}$ | $\begin{aligned} & 12899 \\ & 0 \end{aligned}$ | 68930 | 28016 | 19909 | 16628 | 14894 | $\begin{aligned} & 75159 \\ & 7 \end{aligned}$ |
| 0.031 | 0.009 | 2018 | $\begin{aligned} & 45055 \\ & 9 \end{aligned}$ | $\begin{aligned} & 48619 \\ & 3 \end{aligned}$ | $\begin{aligned} & 40835 \\ & 3 \end{aligned}$ | $\begin{aligned} & 31450 \\ & 2 \end{aligned}$ | $\begin{aligned} & 21186 \\ & 8 \end{aligned}$ | $\begin{aligned} & 20810 \\ & 5 \end{aligned}$ | $\begin{aligned} & 31913 \\ & 0 \end{aligned}$ | $\begin{aligned} & 39761 \\ & 5 \end{aligned}$ | $\begin{aligned} & 33252 \\ & 4 \end{aligned}$ | $\begin{aligned} & 21720 \\ & 3 \end{aligned}$ | $\begin{aligned} & 19258 \\ & 6 \end{aligned}$ | $\begin{aligned} & 18308 \\ & 5 \end{aligned}$ | $\begin{aligned} & 11889 \\ & 6 \end{aligned}$ | 63437 | 25746 | 18275 | 15250 | $\begin{aligned} & 70218 \\ & 4 \end{aligned}$ |
| 0.035 | 0.008 | 2019 | $\begin{aligned} & 43062 \\ & 2 \end{aligned}$ | $\begin{aligned} & 42867 \\ & 6 \end{aligned}$ | $\begin{aligned} & 46258 \\ & 0 \end{aligned}$ | $\begin{aligned} & 38852 \\ & 0 \end{aligned}$ | $\begin{aligned} & 29843 \\ & 4 \end{aligned}$ | $\begin{aligned} & 20055 \\ & 7 \end{aligned}$ | $\begin{aligned} & 19620 \\ & 2 \end{aligned}$ | $\begin{aligned} & 29925 \\ & 1 \end{aligned}$ | $\begin{aligned} & 37067 \\ & 9 \end{aligned}$ | $\begin{aligned} & 30845 \\ & 9 \end{aligned}$ | $\begin{aligned} & 20075 \\ & 5 \end{aligned}$ | $\begin{aligned} & 17753 \\ & 9 \end{aligned}$ | $\begin{aligned} & 16842 \\ & 7 \end{aligned}$ | $\begin{aligned} & 10917 \\ & 2 \end{aligned}$ | 58153 | 23572 | 16717 | $\begin{aligned} & 65548 \\ & 2 \end{aligned}$ |
| 0.042 | 0.008 | 2020 | $\begin{aligned} & 43054 \\ & 4 \end{aligned}$ | $\begin{aligned} & 40970 \\ & 8 \end{aligned}$ | $\begin{aligned} & 40785 \\ & 6 \end{aligned}$ | $\begin{aligned} & 44011 \\ & 3 \end{aligned}$ | $\begin{aligned} & 36880 \\ & 6 \end{aligned}$ | $\begin{aligned} & 28217 \\ & 8 \end{aligned}$ | $\begin{aligned} & 18813 \\ & 5 \end{aligned}$ | $\begin{aligned} & 18226 \\ & 1 \end{aligned}$ | $\begin{aligned} & 27605 \\ & 9 \end{aligned}$ | $\begin{aligned} & 34079 \\ & 2 \end{aligned}$ | $\begin{aligned} & 28313 \\ & 3 \end{aligned}$ | $\begin{aligned} & 18406 \\ & 7 \end{aligned}$ | $\begin{aligned} & 16259 \\ & 4 \end{aligned}$ | $\begin{aligned} & 15405 \\ & 1 \end{aligned}$ | 99729 | 53069 | 21495 | $\begin{aligned} & 61235 \\ & 5 \end{aligned}$ |

Table 6.20b. S. mentella in subareas 1 and 2. Fisheries selectivity at age for the demersal fleet by age (Sa). Numbers are estimated from the statistical catch-at-age model.

| Year/ <br> Age | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5}$ | $\mathbf{1 6}$ |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1992 | 0.000 | 0.000 | 0.000 | 0.274 | 0.315 | 0.359 | 0.406 | 0.454 | 0.503 | 0.553 | 0.601 | 0.647 | 0.691 | 0.731 | 0.768 | 0.802 | 0.831 | 1.000 |
| 1993 | 0.000 | 0.000 | 0.000 | 0.006 | 0.016 | 0.044 | 0.115 | 0.270 | 0.512 | 0.749 | 0.895 | 0.960 | 0.986 | 0.995 | 0.998 | 0.999 | 1.000 | 1.000 |
| 1994 | 0.000 | 0.000 | 0.000 | 0.024 | 0.057 | 0.129 | 0.269 | 0.477 | 0.693 | 0.848 | 0.933 | 0.972 | 0.988 | 0.995 | 0.998 | 0.999 | 1.000 | 1.000 |


| Year/ <br> Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 0.000 | 0.000 | 0.000 | 0.030 | 0.069 | 0.150 | 0.296 | 0.500 | 0.704 | 0.850 | 0.931 | 0.970 | 0.987 | 0.995 | 0.998 | 0.999 | 1.000 | 1.000 |
| 1996 | 0.000 | 0.000 | 0.000 | 0.017 | 0.048 | 0.131 | 0.311 | 0.574 | 0.801 | 0.923 | 0.973 | 0.991 | 0.997 | 0.999 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1997 | 0.000 | 0.000 | 0.000 | 0.014 | 0.041 | 0.113 | 0.274 | 0.528 | 0.768 | 0.908 | 0.967 | 0.989 | 0.996 | 0.999 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1998 | 0.000 | 0.000 | 0.000 | 0.005 | 0.024 | 0.100 | 0.334 | 0.693 | 0.910 | 0.979 | 0.995 | 0.999 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1999 | 0.000 | 0.000 | 0.000 | 0.001 | 0.006 | 0.029 | 0.125 | 0.411 | 0.773 | 0.943 | 0.988 | 0.997 | 0.999 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.013 | 0.112 | 0.556 | 0.925 | 0.992 | 0.999 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2001 | 0.000 | 0.000 | 0.000 | 0.024 | 0.056 | 0.126 | 0.260 | 0.460 | 0.674 | 0.834 | 0.924 | 0.967 | 0.986 | 0.994 | 0.998 | 0.999 | 1.000 | 1.000 |
| 2002 | 0.000 | 0.000 | 0.000 | 0.002 | 0.011 | 0.050 | 0.201 | 0.545 | 0.851 | 0.964 | 0.992 | 0.998 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2003 | 0.000 | 0.000 | 0.000 | 0.037 | 0.081 | 0.165 | 0.309 | 0.503 | 0.696 | 0.838 | 0.921 | 0.964 | 0.984 | 0.993 | 0.997 | 0.999 | 0.999 | 1.000 |
| 2004 | 0.000 | 0.000 | 0.000 | 0.016 | 0.038 | 0.092 | 0.203 | 0.392 | 0.620 | 0.805 | 0.912 | 0.963 | 0.985 | 0.994 | 0.998 | 0.999 | 1.000 | 1.000 |
| 2005 | 0.000 | 0.000 | 0.000 | 0.005 | 0.016 | 0.047 | 0.130 | 0.310 | 0.576 | 0.804 | 0.925 | 0.974 | 0.991 | 0.997 | 0.999 | 1.000 | 1.000 | 1.000 |
| 2006 | 0.000 | 0.000 | 0.000 | 0.002 | 0.007 | 0.018 | 0.051 | 0.134 | 0.306 | 0.558 | 0.783 | 0.912 | 0.967 | 0.988 | 0.996 | 0.999 | 0.999 | 1.000 |
| 2007 | 0.000 | 0.000 | 0.000 | 0.001 | 0.003 | 0.008 | 0.024 | 0.065 | 0.166 | 0.363 | 0.620 | 0.824 | 0.930 | 0.975 | 0.991 | 0.997 | 0.999 | 1.000 |
| 2008 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.003 | 0.012 | 0.053 | 0.204 | 0.540 | 0.844 | 0.961 | 0.991 | 0.998 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2009 | 0.000 | 0.000 | 0.000 | 0.001 | 0.005 | 0.017 | 0.060 | 0.190 | 0.461 | 0.757 | 0.919 | 0.976 | 0.993 | 0.998 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2010 | 0.000 | 0.000 | 0.000 | 0.003 | 0.008 | 0.022 | 0.060 | 0.154 | 0.343 | 0.600 | 0.812 | 0.925 | 0.973 | 0.990 | 0.997 | 0.999 | 1.000 | 1.000 |
| 2011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 | 0.006 | 0.020 | 0.069 | 0.210 | 0.487 | 0.773 | 0.924 | 0.978 | 0.994 | 0.998 | 0.999 | 1.000 | 1.000 |
| 2012 | 0.000 | 0.000 | 0.000 | 0.002 | 0.004 | 0.010 | 0.022 | 0.050 | 0.108 | 0.217 | 0.389 | 0.594 | 0.771 | 0.885 | 0.947 | 0.976 | 0.989 | 1.000 |


| Year/ <br> Age | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5}$ |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2013 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.003 | 0.007 | 0.020 | 0.056 | 0.144 | 0.326 | 0.581 | 0.799 | 0.919 | 0.970 | 0.989 | 1.000 |
| 2014 | 0.000 | 0.000 | 0.000 | 0.002 | 0.003 | 0.007 | 0.013 | 0.024 | 0.045 | 0.083 | 0.147 | 0.248 | 0.387 | 0.548 | 0.699 | 0.816 | 0.895 | 1.000 |
| 2015 | 0.000 | 0.000 | 0.000 | 0.001 | 0.003 | 0.007 | 0.020 | 0.050 | 0.124 | 0.273 | 0.500 | 0.727 | 0.876 | 0.950 | 0.980 | 0.993 | 0.997 | 1.000 |
| 2016 | 0.000 | 0.000 | 0.000 | 0.001 | 0.004 | 0.013 | 0.036 | 0.100 | 0.249 | 0.496 | 0.745 | 0.896 | 0.962 | 0.987 | 0.996 | 0.999 | 0.999 | 1.000 |
| 2017 | 0.000 | 0.000 | 0.000 | 0.001 | 0.003 | 0.013 | 0.059 | 0.228 | 0.581 | 0.867 | 0.969 | 0.993 | 0.999 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2018 | 0.000 | 0.000 | 0.000 | 0.084 | 0.161 | 0.287 | 0.456 | 0.636 | 0.785 | 0.884 | 0.941 | 0.971 | 0.986 | 0.993 | 0.997 | 0.998 | 0.999 | 1.000 |
| 2019 | 0.000 | 0.000 | 0.000 | 0.064 | 0.176 | 0.397 | 0.670 | 0.863 | 0.951 | 0.984 | 0.995 | 0.998 | 0.999 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2020 | 0.000 | 0.000 | 0.000 | 0.061 | 0.185 | 0.441 | 0.733 | 0.905 | 0.971 | 0.991 | 0.998 | 0.999 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |

Table 6.21. Stock summary for S. mentella in subareas 1 and 2 as estimated by the statistical catch-at-age model. Stock biomass is for age $2 \mathrm{y}+$.

| Year | Rec (age 2) in millions | Rec (age 6) in millions | Stock Biomass (tonnes) | SSB (tonnes) | F (12-18) | F(19+) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 400 | 135 | 529902 | 251287 | 0.034 | 0.047 |
| 1993 | 270 | 213 | 572073 | 296819 | 0.032 | 0.033 |
| 1994 | 187 | 320 | 625480 | 372504 | 0.029 | 0.029 |
| 1995 | 177 | 335 | 685167 | 427268 | 0.022 | 0.022 |
| 1996 | 142 | 328 | 745628 | 353633 | 0.015 | 0.015 |
| 1997 | 100 | 221 | 804167 | 434166 | 0.015 | 0.015 |
| 1998 | 51 | 153 | 857764 | 490259 | 0.021 | 0.021 |
| 1999 | 44 | 145 | 900559 | 552753 | 0.016 | 0.016 |
| 2000 | 35 | 116 | 936871 | 640611 | 0.013 | 0.013 |
| 2001 | 37 | 82 | 966732 | 593973 | 0.022 | 0.022 |
| 2002 | 39 | 42 | 978051 | 669920 | 0.008 | 0.008 |
| 2003 | 44 | 36 | 992518 | 739317 | 0.003 | 0.003 |
| 2004 | 58 | 28 | 1004779 | 744162 | 0.006 | 0.006 |
| 2005 | 133 | 31 | 1010390 | 794940 | 0.009 | 0.009 |
| 2006 | 232 | 32 | 1012716 | 782416 | 0.028 | 0.042 |
| 2007 | 335 | 36 | 992659 | 911254 | 0.017 | 0.025 |
| 2008 | 329 | 47 | 987952 | 853677 | 0.014 | 0.019 |
| 2009 | 348 | 109 | 992652 | 886130 | 0.009 | 0.013 |
| 2010 | 500 | 190 | 1006686 | 844048 | 0.01 | 0.014 |
| 2011 | 565 | 274 | 1025073 | 833040 | 0.012 | 0.016 |
| 2012 | 431 | 270 | 1052231 | 827546 | 0.01 | 0.014 |
| 2013 | 267 | 285 | 1095856 | 782106 | 0.008 | 0.013 |
| 2014 | 259 | 409 | 1152683 | 733907 | 0.015 | 0.026 |
| 2015 | 365 | 463 | 1202973 | 757372 | 0.029 | 0.036 |
| 2016 | 451 | 353 | 1244958 | 787325 | 0.041 | 0.047 |
| 2017 | 511 | 219 | 1280146 | 790415 | 0.034 | 0.038 |
| 2018 | 451 | 212 | 1327151 | 811748 | 0.037 | 0.041 |


| Year | Rec (age 2) <br> in millions | Rec (age 6) <br> in millions | Stock Biomass (tonnes) | SSB (tonnes) | F (12-18) | F(19+) |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 2019 | 431 | 298 | 1373398 | 842086 | 0.04 | 0.043 |
| 2020 | 431 | 369 | 1418249 | 874727 | 0.047 | 0.05 |



Figure 6.1. S. mentella in subareas 1 and 2. Total international landings 1952-2020 (thousand tonnes).


Figure 6.2. S. mentella in subareas 1 and 2. Left panel: Catch in tonnes reported by national fleets for the subareas 27.1 and 27.2 and in the NEACF regulatory area. Right panel: Geographical location of the directed Norwegian fishery in 2021 within the Norwegian Exclusive Economic Zone and bycatches by Norwegian vessels in all areas. Directed fishing with bottom trawl is not permitted to the east of the red line. Directed fishing with pelagic trawl is not permitted to the east of the blue line. Directed fishing is not permitted in the Fishery Protection Zone around Svalbard.


Figure 6.3. Delineation of the geographical limits for directed fishing in the Norwegian Economic Zone in 2014-2021. Directed pelagic trawling is only allowed west of the blue line. Directed demersal trawling is only allowed between the blue and the red line. The area east of the stippled line inside NEZ south of Bear Island is only open for directed demersal trawling after 10 May. The other areas for directed fishing are also open during 1 January to last February. Due to high bycatch ratios of golden redfish $72^{\circ} \mathrm{N}$ was suggested as southern limit for directed demersal fishing marked by the red line along that latitude to the Norwegian directorate of fisheries in November 2018.


Figure 6.4. S. mentella in subareas 1 and 2. Length-distributions of the commercial demersal catches by Norway and Russia in 2019-2020.


Figure 6.5. S. mentella in subareas 1 and 2. Upper panels: Catch numbers-at-age for the demersal and pelagic fleets 19922020. Lower panel: Age composition of the commercial demersal catches by Norway and Russia in 2020 (calculated using ALK).


Figure 6.6. Weight-at-age of $S$. mentella per year class in subareas 1 and 2 derived from Norwegian commercial and survey data (Table 6.7). The weights were derived from samples with at least five individuals and are expressed in grammes. The blue and purple lines show the fitted mixed-effect models.


Figure 6.7. S. mentella in subareas 1 and 2. The upper panel shows weight-at-age 19+ as reported from catches (blue) or modelled from catches and survey observations (red) using a mixed effect model (Figure 6.5). AFWG 2017 was the last working group using the annual mixed effect model. The weights-at-age used in the assessment were based on the fixed effects model and are therefore the same for every year. These weights were updated in 2022 and differ only slightly from those estimated in the assessments since 2018. The bottom panel shows comparison of the observed Norwegian and Russian weight by age with the modelled one.


Figure 6.8. Proportion maturity-at-age of S. mentella in subareas 1 and 2 derived from Norwegian commercial and survey data (Table D7). The proportions were derived from samples with at least five individuals. The blue and purple lines show the fitted mixed-effect models. For 2008, 2011 and 2016-2019 the common model (fixed effects blue) was used for other years the annual models (random effects purple) were used. Available data for 2019 was insufficient at the time of the meeting and the fixed effect model was used and there was no age data available for 2020 or 2021.


Natural Mortality rates
$N=30$ Bandwidth $=0.01488$

Figure 6.9. Density distribution of natural mortality rates calculated with 30 of the 39 compared methods. The excluded methods are those based on certain taxa or areas. The broken red line indicates the currently used value; the broken green line the most frequent one and the black dotted lines indicate the beginning and end of the distribution's peak.


Figure 6.10. Abundance of S. mentella (5-14 cm) during the winter survey (February) in the Barents Sea compared with the consumption of redfish (mainly S. mentella) by cod (See Section 1 Table 1.1).


Figure 6.11. S. mentella in subareas 1 and 2. Age disaggregated abundance indices for bottom trawl surveys 1992-2020 in the Barents Sea in winter (winter survey top) in summer (Ecosystem survey middle) and in autumn (Russian groundfish survey bottom).


Figure 6.12. S. mentella in subareas 1 and 2. Abundance indices for individual trawl stations during the ecosystem survey in autumn 2021 (top) and winter survey 2021 (bottom).


Figure 6.13. S. mentella in subareas 1 and 2. Left panel: Survey track of the Deep Pelagic Ecosystem Survey in 2019 and categorized trawls. Only trawls in the category "Standard" served as input for the survey index. Right panel: Catch rates in tonnes per square nautical mile for the surveyed depth layers ( $<=300 \mathrm{~m}, 301-600 \mathrm{~m}$ and $>600 \mathrm{~m}$ ).


Figure 6.14. S. mentella in subareas 1 and 2. Proportions at age during the International Deep Pelagic Ecosystem Survey (WGIDEEPS) in the Norwegian Sea. Bars show proportions at age and dots shows the coefficient of variation for each age. Estimated with RStoX.


Figure 6.15. Map showing the specific pelagic 0 -group trawl stations and the abundance of 0 -group $S$. mentella during the joint Norwegian-Russian Ecosystem survey in the Barents Sea and Svalbard in 2020 (upper panel) and 2021 (lower panel).


Figure 6.16. S. mentella in subareas 1 and 2. Abundance indices (in billions) of 0 -group redfish (believed to be mostly $S$. mentella) in the international 0-group survey in the Barents Sea and Svalbard areas in August-September 1980-2021.


Figure 6.17. S. mentella in subareas 1 and 2. Horizontal distribution of S. mentella hydroacoustic backscattering (sA) during the Norwegian slope survey in spring 2020. The circles are proportional to the sA assigned to redfish along the vessel track.

## Recruitment-at-age 2

## Spawning-stock biomass



Fishing mortality - year component


Figure 6.18. S. mentella in subareas 1 and 2. Results from the statistical catch-at-age assessment run showing the estimated recruitment-at-age 2 spawning-stock biomass from 1992 to 2020 and annual fishing mortality coefficients by year (Fy) from the demersal (blue) and pelagic (red) fleets. Error bars (top) and the coloured envelope (bottom) indicate 95\% confidence limits.

## Fleet selectivity - age component




Figure 6.19. S. mentella in subareas 1 and 2. Results from the statistical catch-at-age assessment run showing the estimated annual fleet selectivity by age (Fa) from the pelagic (top panel) and demersal (lower panels) fleets. Colored envelopes indicate 95\% confidence limits.

## Survey selectivities-at-age



Figure 6.20. S. mentella in subareas 1 and 2. Results from the statistical catch-at-age assessment run showing the selec-tivity-at-age for winter (blue) ecosystem (grey) and Russian groundfish (red) surveys.
S. mentella in ICES subareas 1 and 2 - summary


Figure 6.21. S. mentella in subareas 1 and 2. Results from the statistical catch-at-age model showing the evolution of total biomass (in tonnes light blue left axis) spawning-stock-biomass (in tonnes dark blue, left axis) and recruitment-atage 2 (in numbers yellow, right axis) for the period 1992-2020 for S. mentella in subareas 1 and 2.

Age structure in 2020


Figure 6.22. S. mentella in subareas 1 and 2. Modelled distribution of numbers (yellow bars right $y$-axis) biomass (light blue left $\boldsymbol{y}$-axis) and spawning-stock-biomass (dark blue left $\boldsymbol{y}$-axis) at age 2-45+ in 2020.


Figure 6.23a. Diagnostic plots for the demersal fleet catch-at-age data. Top-left: scatterplot of observed vs. fitted indices the dotted red line indicates 1:1 relationship. Top right: boxplot of residuals (observed-fitted) for each age. Bottom left: boxplot of residuals for each year. Bottom right: bubble plot of residuals for each age/year combination bubble size is proportional to mean residuals blue are positive and red are negative residuals.


Figure 6.23b. Diagnostic plots for the pelagic fleet catch-at-age data. See legend from Figure 6.23a.


Figure 6.23c. Diagnostic plots for winter survey data. See legend from Figure 6.23a.


Figure 6.23d. Diagnostic plots for Ecosystem survey data. See legend from Figure 6.23a.


Figure 6.23e. Diagnostic plots for the Russian groundfish survey data. See legend from Figure 6.23a.


Figure 6.24. The upper panel shows the retrospective patterns of the spawning-stock biomass of S. mentella estimated by the SCAA model for runs up to years 2007-2017 and the baseline model of the $\mathbf{2 0 1 8}$ benchmark. The lower panel presents the baseline model with fixed weights-at-age and the assessment models for 2020 and 2021. Confidence Intervals are shown for the latest assessment.

## 7 Northeast Arctic golden redfish ${ }^{1}$

The advice cycle for golden redfish in subareas 1 and 2 is biennial, following the recommendation of the benchmark assessment for redfish stocks in January 2018 (WKREDFISH, ICES 2018a). Advice was last given in 2020. The age-based GADGET model was then run for the period 19902019, in the configuration approved during the benchmark. The present report therefore updates the assessment and provides advice for the next two years.

### 7.1 Status of the fisheries

### 7.1.1 Recent regulations of the fishery

A description of the historical development of the fishery and regulations is found in the Stock Annex for this stock. The Stock Annex was last updated in February 2018.

Prior to 1 January 2003 there were no regulations particularly for the S. norvegicus fishery, and the regulations aimed at $S$. mentella had only marginal effects on the S. norvegicus stock. After this date, all directed trawl fishery for redfish (both S. norvegicus and S. mentella) outside the permanently closed areas were forbidden in the Norwegian Economic Zone north of $62^{\circ} \mathrm{N}$ and in the Svalbard area. When fishing for other species it was legal to have up to $15 \%$ redfish (both species together) in round weight as bycatch per haul and onboard at any time. Until 14 April 2004, there were no regulations of the other gears/fleets fishing for S. norvegicus. After this date, a minimum legal catch size of 32 cm has been set for all fisheries, with the allowance to have up to $10 \%$ undersized (i.e. less than 32 cm ) specimens of $S$. norvegicus (s) per haul. In addition, a time-limited moratorium (up to 8 months) was enforced in the conventional fisheries (gillnet, longline, handline, Danish seine) except for handline vessels less than 11 metres. From 2016, when trawling outside 12 nm , vessels can have up to $20 \%$ by weight of redfish in each catch and upon landing. When trawling inside 12 nm , it is permitted to have up to $10 \%$ bycatch. Since 2015 it has been prohibited to fish for redfish with conventional gears north of $62^{\circ} \mathrm{N}$. The ban does not, however, apply to vessels less than 15 metres fishing with handline from 1 June to 31 Au gust. When fishing with conventional gears for other species, it is permitted to have up to $10 \%$ by weight of redfish. Vessels less than 21 metres can still have up to $30 \%$ by weight of redfish in the period 1 August to 31 December. Bycatch of redfish is calculated in live weight per week.

### 7.1.2 Landings prior to 2022 (Tables 7.1-7.4 and Figures 7.1-7.3)

Nominal catches of S. norvegicus for the years 1998-2021 by country for subareas 1 and 2 combined, and for each subarea and division are presented in Tables 7.1-7.4. The total landings for both S. norvegicus and S. mentella are presented in section 6 (Tables 6.6 and 6.7). The sources of information used are catches reported to ICES, NEAFC, Norwegian and Russian authorities (foreign vessels fishing in these countries' economic zone) or direct reporting to the AFWG. Where catches are reported as Sebastes sp., they are split into S. norvegicus and S. mentella by AFWG experts based on available correlation between official catches of these two species in the considered areas. Landings of S. norvegicus showed a decrease from a level of 23 000-30 000 t in 19841990 to a stable level of about 16 000-19 000 tin the years 1991-1999. Then the landings decreased further, and the total landings figures for S. norvegicus in 2003-2013 were low but remarkably

[^10]stable, between 5500-8000 $t$. In 2014 the landings decreased to $4825 t$, followed by a further decrease in 2015 with landings of 3873 t , mainly due to stronger regulations. This has since reversed with 8559 tonnes in 2019, 9644 tonnes in 2020 and 10193 tonnes in 2021 (provisional). This increase is likely due to the increased quota for beaked redfish and thereby increased bycatch of golden redfish. The time-series of $S$. norvegicus landings is given in Figure 7.1. A map of $S$. norvegicus catches from Norwegian vessels' logbooks in 2020 is shown in Figure 7.2. Note that species identification from landings and logbooks is not always trusted when the Norwegian final landings data are prepared (see Stock Annex).
The Norwegian landings are presented by gear and month/year in figures 7.3a,b. Reported landings were at the lowest level since World War II in 2015. Since 2015 only bycatches of S. norvegicus are allowed except for a limited amount caught by vessels less than 15 metres fishing with handline from 1 June to 31 August. The increase in landings since 2015 is due to increased bycatch in trawl.

The reported Russian catches of S. norvegicus have been around 600-900 t since 2001, but from 2017 onwards the catches increased steadily to a maximum of 2615 tonnes in 2020 and then decreasing again to 1737 tonnes in 2021 Twelve other countries together usually report catches in the 300-500 t range or less (Table 7.1).

The bycatch of redfish (Sebastes spp.) in the Norwegian Barents Sea shrimp fisheries during the period 1983-2017 were dominated by S. mentella, and hence influenced the S. norvegicus to a much lesser extent. However, these bycatches probably inflicted extra mortality on S. norvegicus in the coastal areas before the sorting grid was enforced in 1990. From 1 January 2006, the maximum legal bycatch of redfish juveniles in the international shrimp fisheries in the northeast Arctic has been reduced from ten to three redfish per 10 kg shrimp.

Information describing the splitting of the redfish landings by species and area is given in the Stock Annex.

### 7.1.3 Expected landings in 2022

New regulations were designed and implemented in the Norwegian coastal fisheries with conventional gears in 2016. No directed fishery is allowed, but the bycatch-regulations are currently rather liberal with vessels less than 21 metres being allowed to have up to $30 \%$ by weight of redfish in the period 1 August-31 December. The bycatch is calculated in live weight per week.

As expected, total landings in 2021 increased due to the raised quota for S. mentella, and thus an increase in bycatch of $S$. norvegicus. The quota for S. mentella in 2021 was not fully exhausted but catches increased by about 10000 t compared to the previous year. With an even higher $S$. mentella quota for 2022, the increase in bycatch of S. norvegicus is expected to continue in 2022.

### 7.2 Data used in the assessment (Table 0.1 and Figure E1)

An example of the sampling levels (by season, area and gear) of the data used in the assessment is presented in Figure E1 for 2013. Although Table 0.1 (see Section 0) shows a reasonably good total sampling level for this stock, the number of different boats sampled, and the gear and area coverage should be improved.

### 7.2.1 Catch-at-length and age (Table 7.5 and Figure 7.4)

The method previously used for calculating catch-at-length and age of Norwegian catches can no longer be used and the procedure was intended to use the new StoX-Reca software. However,
this ran into problems with the bimodal growth pattern exhibited by golden redfish and the large number of length-samples compared with age-samples. Therefore, it was decided to fall back onto the workaround used in the 2020 assessment for catch-at-length and to use the age data from StoX-Reca for 2018 onwards with ages 30+, at which most of the differences occurred, set to missing. Work on the StoX-Reca method will continue towards the benchmark in 2024.

Age composition data were only provided by Norway in the latest years. Other countries were assumed to have the same relative age distribution and mean weight as Norway. The catch num-bers-at-age matrix is shown in Table 7.5. Catch at length data were also only available from Norway (Figure 7.4).

### 7.2.2 Catch weight-at-age (Table 7.6)

Weight-at-age data for ages 7-24+ from Norwegian catches were estimated using StoX-Reca starting with the 2018-catches (Table 7.6). For 2021 weight-at-age-data was not available during the working group, due to a lack of age data from that year. Variations in the weight-at-age of young individuals ( $<10$ years) must be considered with caution as these numbers are derived from only a small number of aged individuals.

### 7.2.3 Maturity-at-age (Table E1, Figure 7.5a-b)

A maturity ogive has previously not been available for S. norvegicus, and knife-edge maturity-at-age 15 (age 15 as $100 \%$ mature) had hence been assumed. Maturity-at-age and length is available from Norwegian surveys and landings up to 2019, as reported in Table E1 and presented in Figure 7.5a. Only the data up to 2018 was considered in the model, due to insufficient age readings in the later years. The maturity ogive modelled by Gadget is presented (Figure 7.5b). This analysis shows that $50 \%$ of the fish at age 12 are mature.

### 7.2.4 Survey results (Tables E2a,b-E3a,b-E4, Figures 7.6a,b-7.8)

Results from the following research vessel survey series are available for S. norvegicus:
Joint Norwegian-Russian Barents Sea winter bottom-trawl survey (A6996 BS-NoRu-Q1 BTr) from 1986 to 2022 in fishing depths of 100-500 m. Length compositions for the years 1986-2022 are shown in Table E2a and Figure 7.6a. Age compositions for the years 1992-2016, 2018 and 2019 are shown in Table E2b and Figure 7.6b. This survey covers important nursery areas for the stock. As described in the stock annex, this survey is used in model tuning.

Norwegian Svalbard (Division 2.b) bottom-trawl survey (August-September) from 1985 to 2020 in fishing depths of 100-500 m (depths down to 800 m incl. in the swept-area). Since 2005 this is part of the Joint Norwegian-Russian Barents Sea Ecosystem survey (A6996 Eco-NoRu-Q3 BTr). Length compositions for the years 1985-2021 and age compositions for the years 1992-2008, 2012, 2013, 2016 and 2018 are shown in Table E3a and E3b, respectively. This survey covers the northernmost part of the species' distribution. Missing age compositions are due to insufficient number of age readings or too few age samples. This survey is not currently included in the model tuning.

Data on length and age from winter and ecosystem surveys have been combined and are shown in Figures 7.7a-b.

Norwegian Coastal and Fjord survey in 1998-2020 from Finnmark to Møre (NOcoast-Aco-Q4). Length composition from catch rates (numbers/ $\mathrm{nm}^{2}$ averaged for all stations within subareas and finally averaged, weighted by subarea, for the total surveyed area) are shown in Figure 7.8 and

Table E4. The survey is an acoustic survey designed to obtain indices of abundance and estimates of length and weight-at-age of saithe and coastal cod north of $62^{\circ} \mathrm{N}$. The index for golden redfish was previously used in the assessment, but was considered unreliable and stopped in 2010. A new index series was recalculated for the benchmark in 2018 (WKREDFISH 2018a). The aggregated survey index varied too much year-to-year to be driven by the population dynamics, but the length distribution was included in the assessment.

SToX versions of winter and ecosystem surveys are used since AFWG 2020. The group recommended that work continues to investigate redfish-specific strata systems for the winter survey and continued monitoring whether the distribution of redfish shifts outside the strata system used for the ecosystem survey. The coastal survey for $S$. norvegicus is in the process of conversion to StoX and adoption of a species-specific strata system, aiming to establish a coherent index of abundance and/or biomass can be obtained for this survey (which is currently only used for annual length distributions).

The bottom-trawl surveys covering the Barents Sea and the Svalbard areas show that the abundance indices over the commercial size range ( $>25 \mathrm{~cm}$ ) were relatively stable up to 1998 but declined to lower levels afterwards. Abundance of pre-recruits ( $<25 \mathrm{~cm}$ ) has steadily decreased since 1991 and has dropped to very low levels after 2000 (Figure 7.6a). An increase in the number of pre-recruits is visible from 2008 onwards. Although this could partly result from taxonomic misidentification, the confirmation of increased numbers for individuals of size 15 cm and greater gives some confidence that at least some of the increasing numbers are S. norvegicus.

### 7.3 Assessment with the Gadget model

### 7.3.1 Description of the model

Since AFWG2005, the GADGET model has been used for this stock, first with experimental runs, and then as analytical assessments following its adoption by WKRED (2012) benchmark (ICES CM 2012/ACOM:48). The model was then approved again at WKREDFISH (2018a), where it was also recommended to switch to a two-year advice cycle. A number of changes have been made to the model at the benchmark WKREDFISH (2018a); the model is moved to a one-year timestep; the fleet structure has been revised to better reflect recent fishing patterns; age-length data are used for tuning in 5 cm (rather than the previous 1 cm ) bins to reduce the extensive noise in this series; proportions (but not absolute abundance) by length in the coastal survey is used for tuning; the model weights have been recalculated; a number of minor errors in the model and data were fixed. Full details are in the WKREDFISH benchmark report (ICES 2018a).

The GADGET model used for the assessment of S. norvegicus in subareas 1 and 2 is closely related to the GADGET model that currently is used by the ICES Northwestern WG on S. norvegicus (Björnsson and Sigurdsson, 2003). The functioning of a Gadget model, including parameter estimation and data used for tuning, is described in Bogstad et al. (2004) and in the stock annex for S. norvegicus. In brief, the model is a single species forward simulation age-length structured model, split into mature and immature components. There are three commercial fleets (a gillnet, a trawl and a combined longline and handline fleet). Prior to 2009 the trawl and longline fleets are combined into one, due to difficulties in obtaining data on a finer resolution. The gillfleet has different selectivity from 2009 compared to 2008 and earlier. There are two surveys used in the model, winter survey and coastal survey. Winter survey tunes to total survey index, the coastal survey to length distributions only. Growth and fishing selectivity within each fleet and survey are assumed constant over time (except for the gilfleet), and recruitment is estimated on annual basis (no SSB-recruit relationship).

The weighting scheme for combining the different datasets into a single likelihood score is a method where weights are selected so that the catch and survey data have approximately equal contribution to the overall likelihood score in the optimized model, and that each dataset within each group gives approximately equal contributions to each other. This ensures that both noise and bias (actually divergence from the consensus) are taken into account in the weighting of datasets. The parameters in the model are estimated using a combination of Simulated Annealing (wide-area search) and Hooke and Jeeves (local search) repeated in sequence until a converged solution is found.

### 7.3.2 Data used for tuning

- Annual catch in tonnes from the commercial fishing fleets, i.e. Norwegian gillnet, and trawl fleet, longline since 2009 and "combined trawl and longline" prior to 2009.
- Annual length distribution of total international commercial landings from the commercial fishing fleets to 2021. Due to late data submissions, there is one-year time-lag in the inclusion of length distributions from other countries than Norway.
- Annual age-length data ( 1 year by 5 cm resolution) from the same fishing fleets, up to 2020. In the last three years (2018-2020) ages above 29 were excluded due to changes in age reading which particularly affected the proportion of fish aged $30+$.
- Length disaggregated frequencies from the Barents Sea (Division 2.a) bottom-trawl survey (February) from 1990-2022 (Table E1a).
- Age-length data and aggregated survey indices from the same survey up to 2019, excluding 2017 (Table E1b).
- Length disaggregated frequencies from the Barents Sea (Division 2.a) coastal survey (February) from 1998-2021 (Table E3, Figure 7.8).


### 7.3.3 Assessment results using the Gadget model (Figures 7.9-7.13)

The general patterns in the stock dynamics of S. norvegicus are similar to those modelled for the past several years, but the recruitment event in 2003 is now beginning to have a noticeable positive effect on the overall stock. The overall stock numbers and biomass have shown a decline over a number of years, but the recent recruitment means that immature and total numbers as well as immature biomass are improving. By now some of the 2003 year class are mature, and the mature stock numbers are therefore stabilizing. The mature biomass is not responding yet, since the maturing fish are still relatively small.

As in previous years, we note that there has been a tendency for some recruitment signal to be reduced in subsequent years, possibly due to misidentification of small $S$. mentella (which is a larger stock and has had good recent recruitment) as $S$. norvegicus, and the model has repeatedly revised down the estimates of this recruitment, although not to zero. The largest fish from the 2003 year class are now entering the mature stock and the fishery, and this is providing multiple sources of information that this was a genuinely good recruitment. The WG stresses that the subsequent recruitment signals (for example the high estimated 2009 year class) should still be treated with extreme caution until they enter the fishery (c. 12-15 years after recruiting).

The most important conclusions to be drawn from the current assessment using the Gadget model are:

- The recruitment to the stock has been very poor for a long period, and especially prior to 2005 (Figure 7.10).
- There has been somewhat better-estimated recruitment in recent years, with a reasonably good recruitment in 2003 (Figure 7.13). Indications of a second pulse of good recruitment
in 2009 have strengthened in the current assessment, but are still highly uncertain, and will need to be tracked for some years to come, to reduce this uncertainty.
- The estimated fishing mortality ( $\mathrm{F}_{15+}$ ) declined between 1990 and 2005 but remained relatively stable until around 2015, (Figure 7.11, Table 7.7). The current mortality is estimated to $\mathrm{F}=0.41$ (Figure 7.11), well above a sustainable level for a redfish species, and above the FMSY $=0.05$ estimated at WKREDFISH (ICES 2018a). Note that the F estimate is based on the 2003 year class being a good one, and the estimate would be higher if this is not the case.

According to the model the total-stock biomass (3+) of S. norvegicus has decreased from about 119000 tonnes in the early 1990s to just under 50000 tonnes in 2021 (Figure 7.12, Table 7.8). Due to the improved recruitment from the 2003 year class, the total biomass is beginning to stabilize, although the SSB is continuing to decline. This reduction is primarily the result of prolonged low recruitment, combined with excessively high fishing pressure.

The average assessment bias (Mohn's Rho) over the last 5 assessments was $1 \%$ for recruitment, $56 \%$ for $\mathrm{F}(15+)$ and $-29 \%$ for SSB. The retrospective plots (Figure 7.13) exhibit a sharp rise in the estimate of mature biomass compared to earlier assessments and a corresponding decline in $\mathrm{F}(15+)$. This can partially be explained by a change in the method of splitting the catch between beaked and golden redfish. However, also in earlier years the retrospectives exhibited a rise in mature biomass for which the reason is unclear and will have to be monitored.

### 7.3.4 State of the stock

Survey observations and the Gadget assessment update confirm previous diagnostics that this stock is currently in a very poor situation. This is confirmed by the production model run as a check at WKRED (ICES 2012) and for the 2020 red list evaluation, which produced similar trends. Indications are that the SSB is continuing to fall. This has led to an upwards trend in F to a level that may place an increasing burden on an already poorly performing stock. Furthermore, in the absence of a substantial population of fish in the 10 to 18 age range, the fishery has become increasingly concentrated on the oldest (18 years and older) individuals, reducing the reproductive capacity of the stock.

There are indications that new recruits from the 2003 year class may have entered the population in recent years as noted in previous AFWG reports. The estimated immature biomass is now beginning to increase, but SSB still declines. However, the total level of this recruitment is still uncertain, and although the 2003 year class is estimated to have been the best since the late 1990s, it is not the largest year class seen in the time-series. Consequently, any rebuilding from this year class is likely to be slow. Rebuilding of this stock is therefore dependent on protecting both the existing SSB and any fish recruiting to it. Note that there are significant uncertainties from misidentification between the redfish species in the Barents Sea, and thus the exact values of both stock and F are uncertain, although the trends are clearly defined.
S. norvegicus is currently on the Norwegian Redlist as a threatened (EN) species according to the criteria given by the International Union for Conservation of Nature (IUCN).

Red-listing is understood to mean that a species (or stock) is at risk of extinction. ICES convened two workshops in 2009. The first Workshop WKPOOR1 (ICES CM 2009/ACOM:29) addressed methods for evaluating extinction risk and outlined approaches that could support advice on how to avoid potential extinction. The second Workshop WKPOOR2 (ICES CM 2009/ACOM:49) applied the results of the first workshop to four stocks selected as being of interest to Norway and ICES.

There are three general methods for evaluating extinction risk: (1) screening methods, such as the IUCN redlisting criteria; (2) simple population viability analysis (PVA) based on time-trends; and (3) age-structured population viability analysis. None of the methods are considered reliable for accurately estimating the absolute probability of extinction, but they may be useful to evaluate the relative probability of extinction between species or between management options.

The fishery is largely concentrated on mature individuals. With a currently estimated SSB of below 30000 tonnes and a FMSY of 0.05 , one would expect a sustainable catch to be in the order of 1000 to 1500 tonnes. The current catches are about ten times as much.

### 7.3.5 Biological reference points

Reference point calculations were conducted at WKREDFISH benchmark (2018a), based on a Bloss with reasonable recruitment, and a forecast with constant recruitment to produce an Fmsy candidate. Note that the benchmark used preliminary data and that the results presented here are slightly changed from those at WKREDFISH (2018). We, therefore, follow the methodology presented at WKREDFISH (2018a) but adjust the Blim based on the revised SSB estimate for 2002. This has the effect of raising the proposed Blim from 44000 tonnes to 49000 tonnes. The Fmsy calculations are unaffected, as these are based on steady-state forecasts.

No stock-recruitment relationship is presented for this stock. Within the model, recruitment is modelled as an annual recruitment value with no relationship with the SSB.

- Blim: Blim is based on the Lowest Observed Stock Size at which reasonable recruitment was observed. This is assumed to be the 2003 year class, at which time the SSB is estimated to be 49000 tonnes (or 44000 tonnes using the benchmark values)
- $\quad B_{p a}$ : Using the ICES default multiplier of 1.4 for $B_{p a}$ gives a $B_{p a}$ value of 68600 tonnes (61 000 tonnes using the benchmark values)

The stock is currently well below the biomass limit reference point, and thus Fmsy is not recommended as the current fishing level. However, it was considered useful to try to estimate a candidate $\mathrm{F}_{\mathrm{msy}}$ reference point, which can be used to compare against management performance. Using yield-per-recruit analysis WKREDFISH (2018a) proposes F0.1 (15+), estimated to be 0.0525, as a candidate Fmsy (Figure E2).

Given the poor state of this stock, management should be based on the need to protect and recover the stock, not on Fmsy.

### 7.3.6 Management advice

AFWG considers that the stock is severely depleted. There are signs that recruitment in 2003 is now beginning to stabilize the population and, for the immature fish, improve the stock status. However, the stock remains in a poor state, and as of now, there are only weak indications that the mature stock is improving. AFWG, therefore, recommends that current area closures and low bycatch limits should be maintained. No directed fishery should be conducted on this stock at the moment, and the percent legal bycatch should be set as low as possible for other fisheries to continue. There will be no directed fishery for S. norvegicus in 2022. It is critical that the bycatch regulations do not allow the catch to increase, as this would impair prospects for recovery.

### 7.3.7 Implementing the ICES FMsy framework

As a long-lived species, S. norvegicus has many year classes contributing to the population, and consequently a relatively stable stock level from year-to year. This makes it relatively simple to
manage to some proxy of MSY (e.g. F0.1) once the biomass has reached close to BMSY, provided adequate measures can be implemented to reduce fishing pressure to an appropriate level. It should be noted that the current fishery is well above the preliminary FMSY for the stock. The main focus should therefore be on reducing total F. The current priority is to stabilize the stock and prevent further decline and allow the recruiting 2003 year class to grow and reproduce. Only then could a recovery strategy and eventually an MSY fishery be implemented. The recent upturn in immature biomass gives some hope that such recovery may be possible, given low fishing pressure.

### 7.4 Tables and figures

Table 7.1. S. norvegicus in subareas 1 and 2. Nominal catch ( $t$ ) by countries in Subarea 1 and divisions 2.a and 2.b combined.

|  |  |  |  |  |  |  | $\begin{aligned} & \text { 들 } \\ & \text { ㅌ } \\ & \underline{\underline{\omega}} \end{aligned}$ |  |  | $\begin{aligned} & \text { 㐅 } \\ & \text { 3 } \\ & \text { 30 } \end{aligned}$ |  |  | $\begin{aligned} & \stackrel{\pi}{\hat{n}} \\ & \underset{\sim}{2} \end{aligned}$ | $\begin{aligned} & . \frac{5}{10} \\ & \text { in } \end{aligned}$ | כ | - |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | - | 78 | 494 | 131 | 33 | - | 19 | - | - | 16540 | - | 6 | 1632 | 51 | 171 | 19155 |
| 1999 | - | 35 | 35 | 228 | 47 | 14 | 7 | - | - | 16750 | - | 3 | 1691 | 7 | 169 | 18986 |
| 2000 | - | 17 | 13 | 160 | 22 | 16 | - | - | - | 13032 | - | 16 | 1112 | - | 73 | 14461 |
| 2001 | - | 37 | 30 | 238 | 17 | - | 1 | - | - | 9134 | - | 7 | 963 | 1 | 119 | 10547 |
| 2002 | - | 60 | 31 | 42 | 31 | 3 | - | - | - | 8561 | - | 34 | 832 | 3 | 46 | 9643 |
| 2003 | - | 109 | 8 | 122 | 36 | 4 | - | - | 89 | 6853 | - | 6 | 479 | - | 134 | 7840 |
| 2004 | - | 19 | 4 | 68 | 20 | 30 | - | - | 33 | 6233 | - | 5 | 722 | 3 | 69 | 7206 |
| 2005 | - | 47 | 10 | 72 | 36 | 8 | - | - | 48 | 6085 | - | 56 | 614 | 8 | 52 | 7036 |
| 2006 | - | 111 | 8 | 35 | 44 | 31 | 3 | - | 21 | 6305 | - | 69 | 713 | 9 | 39 | 7388 |
| 2007 | - | 146 | 15 | 67 | 84 | 68 | 13 | - | 20 | 5784 | - | 225 | 890 | 5 | 55 | 7372 |
| 2008 | - | 274 | 63 | 30 | 71 | 27 | 6 | - | 2 | 5216 | - | 72 | 749 | 4 | 85 | 6599 |
| 2009 | - | 70 | 1 | 58 | 81 | 66 | - | - | 1 | 5451 | - | 30 | 698 | - | 31 | 6487 |
| 2010 | - | 171 | 51 | 31 | 72 | 22 | - | - | - | 5994 | 1 | 28 | 565 | 3 | 44 | 6981 |
| 2011 | - | 24 | 53 | 9 | 51 | 22 | - | - | 1 | 4681 | 48 | 25 | 919 | 6 | 13 | 5852 |


|  |  |  |  |  |  | $\begin{aligned} & \text { 들 } \\ & \text { 플 } \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \text { तِ } \\ & \underline{\underline{N}} \end{aligned}$ |  |  | $\begin{aligned} & \text { 㐅} \\ & \text { ふ̀ } \\ & \text { 30 } \end{aligned}$ | $\begin{aligned} & \text { 듬 } \\ & \frac{\pi}{0} \end{aligned}$ | $\overline{0}$ 0 0 t． 0 | $\begin{aligned} & \text { 苟 } \\ & \\ & \end{aligned}$ | $\begin{aligned} & . \frac{5}{\pi} \\ & \text { in } \end{aligned}$ | $\underset{ }{\text { 〕 }}$ | ָ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | － | 87 | 182 | 71 | 58 | 23 | 12 | － | 5 | 4247 | 34 | 17 | 681 | － | 100 | 5517 |
| 2013 | 1 | 83 | 353 | 1 | 45 | 8 | 1 | － | － | 3836 | 19 | 36 | 797 | － | 493 | 5673 |
| 2014 | － | 67 | 219 | 6 | 20 | 29 | － | － | 1 | 3440 | 21 | 5 | 806 | － | 211 | 4825 |
| 2015 | 1 | 76 | 53 | 24 | 211 | 35 | － | － | － | 2733 | 17 | － | 664 | 2 | 57 | 3873 |
| 2016 | 7 | 183 | 30 | 4 | 87 | 55 | － | － | － | 4131 | 26 | － | 864 | － | 76 | 5463 |
| 2017 | － | 123 | 17 | 19 | 61 | 65 | － | － | 2 | 3567 | 27 | 90 | 1297 | 44 | 160 | 5472 |
| 2018 | 1 | 146 | 37 | 66 | 77 | 67 | － | － | － | 4961 | 36 | 67 | 1834 | 12 | 37 | 7341 |
| 2019 | － | 236 | 25 | 93 | 56 | 83 | － | 3 | － | 5951 | 20 | 73 | 1929 | 65 | 25 | 8559 |
| $2020^{1}$ | － | 166 | 1 | 88 | 99 | 52 | － | － | － | 6503 | 9 | 86 | 2615 | 6 | 19 | 9644 |
| $2021{ }^{1}$ | 2 | 323 | 6 | 76 | 92 | 72 | － | － | － | 7701 | 20 | 60 | 1737 | 8 | 96 | 10193 |

1 －Provisional figures．

## Table 7.2. S. norvegicus in subareas 1 and 2. Nominal catch ( t ) by countries in Subarea 1.



| $\begin{aligned} & \text { 厄゙ } \\ & \text { 厄゙ } \end{aligned}$ |  |  |  |  | O ¢ ¢ U |  |  |  | त 3 3 3 |  |  |  | $\begin{aligned} & \text { 苟 } \\ & \underset{\sim}{\sim} \end{aligned}$ | － in ñ | 〕 | － |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2013 | 28 | 2 | 1 | － | ＋ | － |  | － | 563 |  | － | － | 41 | － | 4 | 639 |
| 2014 | 59 | 10 | 6 | 17 | 4 | － |  | － | 573 |  | 2 | － | 26 | － | 17 | 714 |
| 2015 | 57 | 4 | 9 | 211 | 13 | － |  | － | 624 |  | 2 | － | 51 | 2 | 10 | 983 |
| 2016 | 161 | 7 | 4 | 74 | 51 | － |  | － | 1152 |  | 4 | － | 136 | － | 60 | 1649 |
| 2017 | 81 | 5 | － | 8 | 4 | － |  | － | 970 |  | 2 | 2 | 211 | 2 | 23 | 1308 |
| 2018 | 146 | 28 | 35 | 29 | － | － |  | － | 1151 |  | 5 | 3 | 302 | 5 | 25 | 1729 |
| 2019 | 220 | 10 | 32 | 22 | 30 | － |  | 2 | 1104 |  | 4 | 1 | 422 | 3 | 10 | 1860 |
| 2020 | 143 | － | 14 | 18 | 33 | － |  | － | 1284 |  | 2 | 8 | 708 | 6 | 1 | 2217 |
| $2021{ }^{1}$ | 296 | － | － | 54 | 15 | － |  | － | 1445 |  | － | 12 | 305 | － | － | 2127 |

1 －Provisional figures．
＋denotes less than 0.5 tonnes．

Table 7.3 S. norvegicus in subareas 1 and 2. Nominal catch ( $\mathbf{t}$ ) by countries in Division 2.a.





1 - Provisional figures.

Table 7.4 S. norvegicus in subareas 1 and 2. Nominal catch ( $\mathbf{t}$ ) by countries in Division 2.b.

|  |  |  |  |  |  | $\begin{aligned} & \mathbf{C} \\ & \underline{\Pi} \\ & \underline{\#} \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \underline{\underline{0}} \\ & \underline{\underline{N}} \end{aligned}$ |  | $\begin{aligned} & \text { 㐅} \\ & \text { 3 } \\ & \text { 30 } \\ & \text { 2 } \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \frac{\pi}{0} \\ & \hline \end{aligned}$ |  | $\begin{aligned} & \stackrel{\pi}{\omega} \\ & \underset{\sim}{3} \end{aligned}$ | $\begin{aligned} & \text {.듳 } \\ & \text { io } \end{aligned}$ | こ | $\stackrel{\text { ¢̈ }}{\square}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | - | - | - | 10 | - |  |  |  | 105 | - | - | 246 | - | 3 | 364 |
| 1999 | - | - | - | - | - |  |  |  | 38 | - | - | 355 | - | 2 | 395 |
| 2000 | - | - | - | - | - |  |  |  | 10 | - | - | 308 | - | - | 318 |
| 2001 | - | - | - | - | - |  |  |  | 79 | - | 1 | 223 | - | - | 303 |
| 2002 | - | - | - | - | - |  |  |  | 107 | - | 16 | 420 | 1 | 5 | 549 |
| 2003 | - | - | - | - | - |  |  |  | 68 | - | - | 75 | - | - | 143 |
| 2004 | - | - | - | - | - |  |  |  | 124 | - | - | 113 | - | - | 237 |
| 2005 | - | - | - | 13 | - |  |  |  | 228 | - | - | 288 | - | - | 529 |
| 2006 | - | 5 | - | - | - |  |  |  | 1211 | - | 10 | 284 | - | - | 1510 |
| 2007 | - | 12 | - | - | - |  |  |  | 649 | - | 155 | 242 | - | - | 1058 |
| 2008 | - | - | - | - | - |  |  |  | 126 | - | 1 | 250 | - | - | 377 |
| 2009 | - | - | - | - | - |  |  |  | 207 | - | - | 179 | - | - | 386 |
| 2010 | - | - | - | - | - |  |  |  | 83 | - | 2 | 257 | - | - | 342 |
| 2011 | - | - | 2 | - | - | 1 | - | - | 65 | 48 | 25 | 217 | 4 | - | 362 |
| 2012 | - | 21 | - | 35 | - | 1 | 8 | 3 | 102 | 34 | 16 | 227 | - | 49 | 496 |


| $\begin{aligned} & \text { ஃ } \\ & \text { ভ́ } \end{aligned}$ |  | $\begin{aligned} & n \\ & \frac{n}{c} \\ & \frac{\pi}{n} \\ & \underline{n} \\ & \stackrel{0}{0} \\ & \stackrel{0}{0} \end{aligned}$ |  |  |  |  |  |  | $\begin{aligned} & \text { 㐅 } \\ & \text { 3 } \\ & 0 \\ & \text { z } \end{aligned}$ |  | $\overline{0}$ 0 00 0 0 | $\begin{aligned} & \stackrel{\pi}{3} \\ & \\ & \end{aligned}$ | $\begin{aligned} & \stackrel{\check{1}}{\overline{0}} \\ & \stackrel{0}{n} \end{aligned}$ | 극 | - |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2013 | - | - | 9 | - | - | - | 1 | - | 102 | 19 | 27 | 281 | - | 23 | 462 |
| 2014 | - | - | - | - | - | - | - | - | 135 | 19 | 3 | 221 | - | 16 | 394 |
| 2015 | 1 | - | - | - | - | - | - | - | 28 | 3 | - | 175 | - | - | 207 |
| 2016 | 7 | - | - | - | - | - | - | - | 34 | 14 | - | 183 | - | - | 238 |
| 2017 | - | - | - | - | 18 | - | - | - | 48 | 2 | - | 405 | 4 | - | 477 |
| 2018 | 1 | - | - | 14 | 6 | - | - | - | 64 | 19 | - | 1043 | - | - | 1147 |
| 2019 | - | - | - | - | - | - | - | - | 103 | - | - | 712 | 1 | 1 | 817 |
| 2020 | - | - | - | 13 | - | - | - | - | 381 | 7 | - | 961 | - | 3 | 1365 |
| $2021{ }^{1}$ | 2 | 3 | + | 55 | 2 | - | - | - | 576 | 20 | - | 359 | 6 | 6 | 1030 |

1 - Provisional figures.

+ denotes less than 0.5 tonnes.

Table 7.5. S. norvegicus in subareas 1 and 2. Catch numbers-at-age (in thousands). Since 2018, numbers are from StoX-Reca.

| Year/Age | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | +gp | Total Num. | Tonnes Land. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 5 | 22 | 78 | 114 | 394 | 549 | 783 | 1718 | 3102 | 2495 | 2104 | 1837 | 998 | 858 | 688 | 547 | 268 | 3110 | 19670 | 16185 |
| 1993 | 0 | 24 | 193 | 359 | 406 | 1036 | 1022 | 1523 | 2353 | 1410 | 1655 | 1678 | 745 | 716 | 534 | 528 | 576 | 3482 | 18240 | 16651 |


| Year/Age | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | +gp | Total Num. | Tonnes Land. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 46 | 7 | 292 | 640 | 816 | 1930 | 2096 | 2030 | 1601 | 2725 | 2668 | 1409 | 617 | 733 | 514 | 256 | 177 | 1508 | 20065 | 18120 |
| 1995 | 60 | 85 | 230 | 672 | 908 | 1610 | 2038 | 2295 | 1783 | 1406 | 785 | 563 | 670 | 593 | 419 | 368 | 250 | 3232 | 17967 | 15616 |
| 1996 | 9 | 119 | 313 | 361 | 879 | 1234 | 1638 | 2134 | 1675 | 1614 | 1390 | 952 | 679 | 439 | 560 | 334 | 490 | 3135 | 17955 | 18043 |
| 1997 | 9 | 98 | 156 | 321 | 686 | 1065 | 1781 | 2276 | 2172 | 1848 | 1421 | 851 | 804 | 608 | 511 | 205 | 334 | 2131 | 17277 | 17511 |
| 1998 | 28 | 51 | 206 | 470 | 721 | 968 | 1512 | 1736 | 1582 | 1045 | 1277 | 970 | 1018 | 846 | 443 | 764 | 486 | 3389 | 17512 | 19155 |
| 1999 | 78 | 593 | 855 | 572 | 1006 | 1230 | 1618 | 1480 | 1612 | 1239 | 1407 | 1558 | 1019 | 394 | 197 | 459 | 174 | 2131 | 17622 | 18986 |
| 2000 | 4 | 13 | 70 | 245 | 902 | 958 | 1782 | 1409 | 2121 | 2203 | 1715 | 753 | 483 | 458 | 132 | 230 | 224 | 895 | 14597 | 14460 |
| 2001 | 23 | 23 | 44 | 199 | 347 | 482 | 1120 | 1342 | 1674 | 1653 | 1243 | 568 | 119 | 183 | 154 | 112 | 135 | 254 | 9675 | 10547 |
| 2002 | 14 | 36 | 71 | 143 | 414 | 686 | 1199 | 1943 | 1377 | 1274 | 1196 | 388 | 313 | 99 | 104 | 117 | 113 | 253 | 9740 | 9643 |
| 2003 | 22 | 25 | 30 | 44 | 204 | 359 | 705 | 1687 | 1338 | 1071 | 937 | 481 | 367 | 146 | 84 | 51 | 18 | 69 | 7637 | 7841 |
| 2004 | 19 | 47 | 46 | 65 | 198 | 277 | 504 | 590 | 677 | 963 | 1059 | 787 | 436 | 169 | 183 | 108 | 79 | 186 | 6390 | 7320 |
| 2005 | 40 | 55 | 94 | 80 | 165 | 173 | 393 | 779 | 741 | 916 | 926 | 743 | 376 | 210 | 189 | 129 | 111 | 220 | 6338 | 7037 |
| 2006 | 45 | 32 | 56 | 70 | 245 | 204 | 201 | 809 | 549 | 779 | 794 | 747 | 496 | 332 | 310 | 188 | 165 | 397 | 6419 | 7348 |
| 2007 | 15 | 21 | 31 | 68 | 138 | 306 | 448 | 495 | 523 | 637 | 892 | 616 | 510 | 396 | 225 | 322 | 170 | 630 | 6443 | 7306 |
| 2008 | 1 | 4 | 14 | 12 | 49 | 139 | 265 | 366 | 361 | 443 | 442 | 538 | 547 | 479 | 281 | 223 | 144 | 1032 | 5342 | 6557 |
| 2009 | 0 | 11 | 2 | 4 | 9 | 23 | 144 | 277 | 315 | 248 | 406 | 374 | 509 | 404 | 331 | 323 | 253 | 911 | 4544 | 6487 |
| 2010 | 1 | 0 | 10 | 7 | 4 | 20 | 75 | 261 | 291 | 529 | 359 | 311 | 531 | 502 | 385 | 295 | 247 | 776 | 4605 | 6982 |
| 2011 | 2 | 1 | 3 | 0 | 2 | 5 | 64 | 304 | 466 | 266 | 312 | 223 | 378 | 289 | 247 | 229 | 253 | 985 | 4028 | 5852 |


| Year/Age | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | +gp | Total Num. | Tonnes Land. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | 15 | 10 | 5 | 12 | 0 | 2 | 228 | 226 | 322 | 295 | 191 | 169 | 184 | 283 | 266 | 268 | 262 | 1152 | 3891 | 5517 |
| 2013 | 31 | 88 | 138 | 57 | 10 | 44 | 58 | 202 | 241 | 437 | 321 | 205 | 213 | 270 | 258 | 196 | 322 | 1216 | 4309 | 5608 |
| 2014 | 5 | 4 | 8 | 8 | 8 | 15 | 26 | 49 | 67 | 204 | 197 | 148 | 167 | 184 | 165 | 156 | 213 | 1197 | 2821 | 4438 |
| 2015 | 15 | 16 | 14 | 17 | 26 | 43 | 29 | 96 | 113 | 128 | 170 | 147 | 159 | 115 | 99 | 96 | 220 | 1156 | 2661 | 3628 |
| 2016 | 53 | 59 | 60 | 88 | 88 | 147 | 293 | 217 | 266 | 81 | 178 | 176 | 110 | 162 | 110 | 182 | 191 | 1103 | 3563 | 4674 |
| 2017 | 106 | 82 | 132 | 69 | 132 | 165 | 311 | 455 | 225 | 132 | 105 | 83 | 85 | 102 | 88 | 138 | 182 | 1169 | 3760 | 5257 |
| 2018 | 129 | 65 | 230 | 443 | 246 | 496 | 158 | 170 | 236 | 171 | 145 | 183 | 194 | 232 | 233 | 229 | 249 | 2425 | 6235 | 7341 |
| 2019 | 36 | 98 | 169 | 130 | 318 | 635 | 356 | 282 | 96 | 123 | 71 | 99 | 67 | 57 | 145 | 129 | 93 | 2159 | 5064 | 5951 |
| 2020 | 26 | 14 | 108 | 439 | 472 | 580 | 651 | 324 | 190 | 153 | 55 | 62 | 126 | 49 | 112 | 98 | 90 | 1751 | 5302 | 6503 |

Table 7.6. S. norvegicus in subareas 1 and 2. Catch weights at age (kg). Since 2018, numbers are from StoX-Reca.

| Year/Age | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | +gp |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.18 | 0.29 | 0.48 | 0.42 | 0.50 | 0.59 | 0.58 | 0.65 | 0.65 | 0.71 | 0.82 | 0.84 | 0.94 | 1.02 | 1.03 | 1.15 | 1.27 | 1.27 |
| 1993 | 0.2 | 0.33 | 0.36 | 0.43 | 0.51 | 0.51 | 0.64 | 0.64 | 0.76 | 0.86 | 0.89 | 0.98 | 1 | 1.03 | 1.21 | 1.03 | 1.2 | 1.14 |
| 1994 | 0.25 | 0.37 | 0.38 | 0.49 | 0.51 | 0.64 | 0.74 | 0.76 | 0.86 | 0.95 | 1.03 | 1.07 | 1.11 | 1.16 | 1.15 | 1.13 | 1.02 | 1.36 |
| 1995 | 0.33 | 0.43 | 0.64 | 0.61 | 0.59 | 0.65 | 0.74 | 0.79 | 0.84 | 0.92 | 1.12 | 1.01 | 1.01 | 1.21 | 1.14 | 1.09 | 1.3 | 1.01 |
| 1996 | 0.22 | 0.49 | 0.56 | 0.65 | 0.71 | 0.81 | 0.84 | 0.88 | 0.96 | 1 | 1.02 | 1.01 | 1 | 1.03 | 1.04 | 1.14 | 1.09 | 1.16 |
| 1997 | 0.23 | 0.51 | 0.53 | 0.74 | 0.72 | 0.78 | 0.8 | 0.86 | 0.91 | 0.99 | 1.16 | 1.18 | 1.21 | 1.34 | 1.28 | 1.54 | 1.19 | 1.29 |
| 1998 | 0.37 | 0.21 | 0.47 | 0.62 | 0.67 | 0.77 | 0.77 | 0.85 | 1.05 | 0.96 | 1.25 | 1.28 | 1.3 | 1.23 | 1.87 | 1.46 | 1.73 | 1.29 |


| Year/Age | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | +gp |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 0.14 | 0.26 | 0.44 | 0.57 | 0.69 | 0.78 | 0.86 | 1.04 | 1.07 | 1.12 | 1.18 | 1.71 | 1.09 | 1.18 | 1.04 | 1.34 | 1.18 | 1.34 |
| 2000 | 0.19 | 0.24 | 0.32 | 0.44 | 0.53 | 0.64 | 0.73 | 0.84 | 0.96 | 1.11 | 1.25 | 1.32 | 1.53 | 1.06 | 1.29 | 1.32 | 1.12 | 1.2 |
| 2001 | 0.15 | 0.26 | 0.45 | 0.55 | 0.58 | 0.67 | 0.8 | 0.89 | 1.01 | 1.14 | 1.33 | 1.43 | 1.62 | 1.6 | 1.47 | 2 | 2.7 | 2.31 |
| 2002 | 0.17 | 0.25 | 0.33 | 0.42 | 0.54 | 0.67 | 0.72 | 0.84 | 0.98 | 1.09 | 1.2 | 1.3 | 1.44 | 1.78 | 1.68 | 1.88 | 2.12 | 1.84 |
| 2003 | 0.19 | 0.22 | 0.31 | 0.39 | 0.49 | 0.58 | 0.69 | 0.84 | 0.96 | 1.05 | 1.29 | 1.36 | 1.65 | 1.74 | 2.09 | 1.85 | 2.3 | 2.38 |
| 2004 | 0.21 | 0.26 | 0.36 | 0.45 | 0.51 | 0.59 | 0.68 | 0.8 | 0.96 | 1.07 | 1.22 | 1.34 | 1.57 | 1.67 | 1.75 | 2.09 | 1.9 | 2.04 |
| 2005 | 0.16 | 0.21 | 0.36 | 0.45 | 0.52 | 0.58 | 0.68 | 0.82 | 0.94 | 1.03 | 1.16 | 1.36 | 1.46 | 1.51 | 1.67 | 1.91 | 2.23 | 2.27 |
| 2006 | 0.13 | 0.15 | 0.28 | 0.41 | 0.51 | 0.58 | 0.66 | 0.74 | 0.83 | 1 | 1.14 | 1.27 | 1.39 | 1.46 | 1.37 | 1.47 | 1.64 | 2.03 |
| 2007 | 0.15 | 0.21 | 0.33 | 0.39 | 0.5 | 0.59 | 0.65 | 0.77 | 0.9 | 1 | 1.09 | 1.27 | 1.42 | 1.32 | 1.53 | 1.47 | 1.69 | 1.81 |
| 2008 | 0.41 | 0.55 | 0.55 | 0.57 | 0.52 | 0.58 | 0.65 | 0.81 | 0.9 | 1.07 | 1.14 | 1.36 | 1.51 | 1.81 | 1.99 | 2.01 | 2.26 | 1.93 |
| 2009 | 0.00 | 1.01 | 0.34 | 0.59 | 0.61 | 0.66 | 0.82 | 0.92 | 0.94 | 1.09 | 1.22 | 1.35 | 1.40 | 1.57 | 1.68 | 1.74 | 1.73 | 2.25 |
| 2010 | 0.15 | 0.00 | 0.10 | 0.32 | 0.52 | 0.73 | 0.77 | 0.89 | 0.98 | 1.09 | 1.25 | 1.40 | 1.48 | 1.64 | 1.77 | 1.99 | 1.82 | 1.86 |
| 2011 | 0.16 | 0.20 | 0.21 | 0.00 | 0.54 | 0.52 | 0.72 | 0.91 | 1.08 | 1.14 | 1.20 | 1.45 | 1.40 | 1.43 | 1.54 | 1.60 | 1.74 | 1.93 |
| 2012 | 0.19 | 0.25 | 0.33 | 0.72 | 0.61 | 0.88 | 0.70 | 0.86 | 0.95 | 1.02 | 1.13 | 1.18 | 1.33 | 1.48 | 1.31 | 1.55 | 1.50 | 2.59 |
| 2013 | 0.20 | 0.27 | 0.32 | 0.44 | 0.47 | 0.55 | 0.63 | 0.88 | 0.96 | 1.08 | 1.08 | 1.19 | 1.21 | 1.39 | 1.38 | 1.62 | 1.41 | 1.81 |
| 2014 | 0.20 | 0.26 | 0.39 | 0.41 | 0.56 | 0.61 | 0.71 | 0.87 | 0.95 | 1.07 | 1.14 | 1.28 | 1.46 | 1.35 | 1.51 | 1.62 | 1.69 | 1.84 |
| 2015 | 0.16 | 0.22 | 0.30 | 0.50 | 0.51 | 0.60 | 0.66 | 0.88 | 0.93 | 1.04 | 1.15 | 1.18 | 1.23 | 1.34 | 1.51 | 1.50 | 1.48 | 1.62 |
| 2016 | 0.17 | 0.21 | 0.34 | 0.62 | 0.53 | 0.66 | 0.68 | 0.86 | 0.94 | 1.03 | 1.11 | 1.32 | 1.43 | 1.29 | 1.42 | 1.43 | 1.48 | 2.67 |


| Year/Age | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5}$ | $\mathbf{1 6}$ | $\mathbf{1 7}$ | $\mathbf{1 8}$ | $\mathbf{1 9}$ | $\mathbf{2 0}$ | $\mathbf{2 1}$ | $\mathbf{2 2}$ | $\mathbf{2 3}$ |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2017 | 0.18 | 0.23 | 0.29 | 0.38 | 0.55 | 0.59 | 0.70 | 0.80 | 0.92 | 1.06 | 1.15 | 1.35 | 1.40 | 1.56 | 1.37 | 1.74 | 1.83 | 2.92 |
| 2018 | 0.75 | 0.76 | 0.80 | 0.86 | 0.92 | 1.00 | 1.04 | 1.06 | 1.15 | 1.23 | 1.24 | 1.27 | 1.35 | 1.40 | 1.43 | 1.50 | 1.48 | 2.34 |
| 2019 | 0.93 | 0.98 | 1.07 | 1.12 | 1.20 | 1.26 | 1.28 | 1.34 | 1.38 | 1.33 | 1.36 | 1.43 | 1.44 | 1.45 | 1.43 | 1.50 | 1.48 | 1.95 |
| $2020^{1}$ | 1.71 | 1.13 | 1.28 | 1.14 | 1.31 | 1.28 | 1.39 | 1.49 | 1.56 | 1.59 | 1.52 | 1.59 | 1.64 | 1.68 | 1.67 | 1.69 | 1.64 | 2.09 |

1 - Provisional figures.
Table 7.7. S. norvegicus in subareas 1 and 2. Fishing mortalities as estimated by Gadget.

| Age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 6 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 7 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 8 | 0.04 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 |
| 9 | 0.07 | 0.05 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.01 | 0.01 | 0.01 | 0.01 |
| 10 | 0.10 | 0.08 | 0.07 | 0.04 | 0.04 | 0.03 | 0.04 | 0.04 | 0.04 | 0.04 | 0.03 | 0.02 | 0.02 | 0.02 | 0.02 |
| 11 | 0.13 | 0.11 | 0.10 | 0.09 | 0.07 | 0.05 | 0.06 | 0.06 | 0.06 | 0.06 | 0.05 | 0.04 | 0.04 | 0.03 | 0.03 |
| 12 | 0.17 | 0.13 | 0.12 | 0.12 | 0.12 | 0.08 | 0.09 | 0.08 | 0.09 | 0.09 | 0.07 | 0.06 | 0.05 | 0.04 | 0.04 |
| 13 | 0.22 | 0.17 | 0.14 | 0.14 | 0.15 | 0.12 | 0.11 | 0.11 | 0.12 | 0.12 | 0.10 | 0.08 | 0.07 | 0.06 | 0.06 |
| 14 | 0.28 | 0.20 | 0.17 | 0.16 | 0.17 | 0.14 | 0.16 | 0.13 | 0.15 | 0.16 | 0.13 | 0.10 | 0.09 | 0.07 | 0.07 |


| Age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 15 | 0.34 | 0.24 | 0.19 | 0.18 | 0.19 | 0.16 | 0.18 | 0.17 | 0.18 | 0.19 | 0.15 | 0.11 | 0.10 | 0.09 | 0.08 |
| 16 | 0.41 | 0.29 | 0.22 | 0.21 | 0.21 | 0.17 | 0.20 | 0.19 | 0.22 | 0.21 | 0.17 | 0.13 | 0.12 | 0.10 | 0.09 |
| 17 | 0.48 | 0.33 | 0.25 | 0.23 | 0.24 | 0.19 | 0.21 | 0.21 | 0.24 | 0.25 | 0.19 | 0.15 | 0.13 | 0.11 | 0.10 |
| 18 | 0.52 | 0.38 | 0.29 | 0.26 | 0.26 | 0.21 | 0.23 | 0.22 | 0.25 | 0.27 | 0.22 | 0.16 | 0.14 | 0.12 | 0.11 |
| 19 | 0.55 | 0.40 | 0.31 | 0.28 | 0.28 | 0.22 | 0.25 | 0.24 | 0.27 | 0.28 | 0.23 | 0.17 | 0.15 | 0.13 | 0.12 |
| 20 | 0.58 | 0.42 | 0.32 | 0.30 | 0.30 | 0.24 | 0.26 | 0.25 | 0.28 | 0.29 | 0.24 | 0.17 | 0.16 | 0.13 | 0.12 |
| 21 | 0.61 | 0.43 | 0.33 | 0.31 | 0.31 | 0.25 | 0.27 | 0.26 | 0.29 | 0.30 | 0.24 | 0.18 | 0.16 | 0.13 | 0.12 |
| 22 | 0.62 | 0.44 | 0.33 | 0.31 | 0.31 | 0.25 | 0.27 | 0.26 | 0.29 | 0.31 | 0.25 | 0.18 | 0.16 | 0.13 | 0.12 |
| 23 | 0.62 | 0.43 | 0.33 | 0.30 | 0.30 | 0.24 | 0.27 | 0.26 | 0.29 | 0.31 | 0.25 | 0.18 | 0.16 | 0.12 | 0.11 |
| 24 | 0.61 | 0.42 | 0.32 | 0.29 | 0.29 | 0.23 | 0.26 | 0.25 | 0.29 | 0.30 | 0.24 | 0.17 | 0.15 | 0.12 | 0.11 |
| 25 | 0.58 | 0.40 | 0.29 | 0.27 | 0.27 | 0.22 | 0.25 | 0.24 | 0.27 | 0.29 | 0.23 | 0.17 | 0.15 | 0.12 | 0.11 |
| 26 | 0.55 | 0.36 | 0.26 | 0.24 | 0.24 | 0.20 | 0.22 | 0.22 | 0.25 | 0.26 | 0.21 | 0.16 | 0.14 | 0.11 | 0.10 |
| 27 | 0.50 | 0.33 | 0.23 | 0.21 | 0.22 | 0.17 | 0.20 | 0.20 | 0.22 | 0.23 | 0.18 | 0.14 | 0.13 | 0.10 | 0.09 |
| 28 | 0.46 | 0.30 | 0.21 | 0.19 | 0.19 | 0.15 | 0.17 | 0.17 | 0.20 | 0.20 | 0.16 | 0.12 | 0.11 | 0.09 | 0.09 |
| 29 | 0.42 | 0.27 | 0.19 | 0.16 | 0.16 | 0.13 | 0.15 | 0.15 | 0.17 | 0.17 | 0.14 | 0.10 | 0.09 | 0.08 | 0.08 |
| 30 | 0.34 | 0.20 | 0.13 | 0.11 | 0.13 | 0.11 | 0.12 | 0.11 | 0.13 | 0.14 | 0.10 | 0.08 | 0.07 | 0.04 | 0.04 |
| 15+ | 0.513 | 0.351 | 0.264 | 0.241 | 0.243 | 0.196 | 0.219 | 0.212 | 0.239 | 0.251 | 0.199 | 0.147 | 0.132 | 0.107 | 0.099 |


| Age | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 6 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 7 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 8 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.01 | 0.01 |
| 9 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.02 | 0.03 |
| 10 | 0.02 | 0.02 | 0.02 | 0.02 | 0.01 | 0.02 | 0.01 | 0.01 | 0.02 | 0.01 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 | 0.05 | 0.06 |
| 11 | 0.03 | 0.03 | 0.03 | 0.03 | 0.02 | 0.03 | 0.02 | 0.02 | 0.03 | 0.02 | 0.02 | 0.03 | 0.03 | 0.05 | 0.06 | 0.07 | 0.10 |
| 12 | 0.04 | 0.04 | 0.04 | 0.04 | 0.03 | 0.04 | 0.03 | 0.04 | 0.04 | 0.04 | 0.03 | 0.05 | 0.05 | 0.07 | 0.09 | 0.11 | 0.14 |
| 13 | 0.05 | 0.06 | 0.06 | 0.06 | 0.04 | 0.06 | 0.05 | 0.05 | 0.05 | 0.05 | 0.04 | 0.06 | 0.06 | 0.09 | 0.12 | 0.15 | 0.19 |
| 14 | 0.07 | 0.07 | 0.07 | 0.07 | 0.06 | 0.07 | 0.06 | 0.06 | 0.07 | 0.06 | 0.05 | 0.08 | 0.08 | 0.12 | 0.15 | 0.18 | 0.24 |
| 15 | 0.08 | 0.08 | 0.09 | 0.08 | 0.07 | 0.09 | 0.07 | 0.07 | 0.08 | 0.07 | 0.06 | 0.09 | 0.10 | 0.14 | 0.18 | 0.22 | 0.29 |
| 16 | 0.09 | 0.10 | 0.10 | 0.10 | 0.08 | 0.10 | 0.08 | 0.08 | 0.09 | 0.09 | 0.07 | 0.11 | 0.11 | 0.16 | 0.21 | 0.26 | 0.34 |
| 17 | 0.10 | 0.11 | 0.11 | 0.11 | 0.09 | 0.11 | 0.09 | 0.09 | 0.10 | 0.09 | 0.08 | 0.12 | 0.12 | 0.18 | 0.23 | 0.29 | 0.38 |
| 18 | 0.11 | 0.11 | 0.12 | 0.11 | 0.10 | 0.12 | 0.10 | 0.10 | 0.11 | 0.10 | 0.08 | 0.13 | 0.13 | 0.19 | 0.25 | 0.32 | 0.43 |
| 19 | 0.11 | 0.12 | 0.12 | 0.12 | 0.10 | 0.13 | 0.11 | 0.11 | 0.12 | 0.11 | 0.09 | 0.13 | 0.14 | 0.21 | 0.27 | 0.34 | 0.46 |
| 20 | 0.11 | 0.12 | 0.12 | 0.12 | 0.11 | 0.14 | 0.11 | 0.11 | 0.12 | 0.11 | 0.09 | 0.14 | 0.14 | 0.21 | 0.28 | 0.36 | 0.48 |
| 21 | 0.11 | 0.12 | 0.12 | 0.12 | 0.11 | 0.14 | 0.11 | 0.11 | 0.13 | 0.11 | 0.09 | 0.14 | 0.15 | 0.22 | 0.28 | 0.36 | 0.50 |


| Age | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 22 | 0.11 | 0.12 | 0.12 | 0.12 | 0.11 | 0.14 | 0.11 | 0.12 | 0.13 | 0.11 | 0.09 | 0.14 | 0.15 | 0.21 | 0.28 | 0.36 | 0.50 |
| 23 | 0.11 | 0.12 | 0.12 | 0.11 | 0.11 | 0.14 | 0.11 | 0.11 | 0.12 | 0.11 | 0.09 | 0.14 | 0.15 | 0.21 | 0.28 | 0.36 | 0.49 |
| 24 | 0.10 | 0.11 | 0.12 | 0.11 | 0.11 | 0.14 | 0.11 | 0.11 | 0.12 | 0.11 | 0.09 | 0.13 | 0.14 | 0.21 | 0.27 | 0.35 | 0.47 |
| 25 | 0.10 | 0.10 | 0.11 | 0.10 | 0.11 | 0.13 | 0.11 | 0.11 | 0.12 | 0.11 | 0.09 | 0.13 | 0.14 | 0.20 | 0.26 | 0.33 | 0.45 |
| 26 | 0.09 | 0.10 | 0.10 | 0.09 | 0.10 | 0.13 | 0.11 | 0.11 | 0.12 | 0.10 | 0.09 | 0.13 | 0.13 | 0.19 | 0.25 | 0.32 | 0.43 |
| 27 | 0.09 | 0.09 | 0.10 | 0.08 | 0.10 | 0.13 | 0.10 | 0.10 | 0.11 | 0.10 | 0.08 | 0.12 | 0.13 | 0.18 | 0.23 | 0.30 | 0.40 |
| 28 | 0.08 | 0.09 | 0.09 | 0.08 | 0.09 | 0.12 | 0.10 | 0.10 | 0.11 | 0.10 | 0.08 | 0.12 | 0.12 | 0.17 | 0.22 | 0.28 | 0.37 |
| 29 | 0.08 | 0.08 | 0.08 | 0.07 | 0.09 | 0.11 | 0.09 | 0.10 | 0.10 | 0.09 | 0.08 | 0.11 | 0.12 | 0.16 | 0.21 | 0.26 | 0.35 |
| 30 | 0.04 | 0.04 | 0.04 | 0.04 | 0.07 | 0.09 | 0.08 | 0.08 | 0.09 | 0.08 | 0.06 | 0.09 | 0.09 | 0.13 | 0.16 | 0.19 | 0.25 |
| 15+ | 0.095 | 0.101 | 0.104 | 0.098 | 0.096 | 0.123 | 0.101 | 0.102 | 0.111 | 0.101 | 0.083 | 0.122 | 0.129 | 0.186 | 0.240 | 0.307 | 0.411 |

Table 7.8. S. norvegicus in subareas 1 and 2. Stock numbers, biomass, mean weight and maturity ogives as estimated by GADGET.

| year |  | total stock |  |  | mature |  |  | immature |  |  | recruit |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number | mean wt | biomass | number | mean wt | biomass | number | mean wt | biomass | F(15+) | age 3 |
|  | (millions) | (kg) | (1000t) | (millions) | (kg) |  | (millions) | (kg) | (1000t) |  | (millions) |
| 1986 | 375 | 0.35 | 132.28 | 103 | 0.67 | 69.06 | 271 | 0.23 | 63.22 |  | 4.25 |
| 1987 | 370 | 0.35 | 129.94 | 101 | 0.65 | 65.92 | 268 | 0.24 | 64.01 |  | 3.54 |
| 1988 | 348 | 0.36 | 125.06 | 98 | 0.61 | 60.02 | 250 | 0.26 | 65.04 |  | 1.98 |
| 1989 | 328 | 0.37 | 122.35 | 96 | 0.58 | 56.21 | 231 | 0.29 | 66.14 |  | 1.84 |


| year |  | total stock |  |  | mature |  |  | immature |  |  | recruit <br> age 3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number | mean wt | biomass | number | mean wt | biomass | number | mean wt | biomass | F(15+) |  |
|  | (millions) | (kg) | (1000t) | (millions) | (kg) |  | (millions) | (kg) | (1000t) |  | (millions) |
| 1990 | 305 | 0.37 | 113.79 | 92 | 0.54 | 49.82 | 213 | 0.30 | 63.97 | 0.51 | 1.98 |
| 1991 | 289 | 0.39 | 113.64 | 94 | 0.55 | 51.17 | 195 | 0.32 | 62.47 | 0.35 | 1.83 |
| 1992 | 275 | 0.42 | 115.73 | 96 | 0.57 | 55.39 | 178 | 0.34 | 60.34 | 0.26 | 1.65 |
| 1993 | 260 | 0.45 | 116.56 | 98 | 0.61 | 59.71 | 162 | 0.35 | 56.85 | 0.24 | 1.56 |
| 1994 | 248 | 0.46 | 115.09 | 97 | 0.64 | 62.75 | 151 | 0.35 | 52.33 | 0.24 | 1.91 |
| 1995 | 233 | 0.49 | 115.17 | 97 | 0.69 | 66.78 | 136 | 0.36 | 48.38 | 0.20 | 1.24 |
| 1996 | 213 | 0.52 | 111.60 | 94 | 0.72 | 68.08 | 119 | 0.37 | 43.52 | 0.22 | 0.85 |
| 1997 | 195 | 0.55 | 107.39 | 90 | 0.76 | 68.37 | 105 | 0.37 | 39.02 | 0.21 | 0.85 |
| 1998 | 173 | 0.58 | 100.10 | 84 | 0.79 | 65.81 | 89 | 0.39 | 34.29 | 0.24 | 0.42 |
| 1999 | 151 | 0.60 | 91.59 | 76 | 0.81 | 61.68 | 75 | 0.40 | 29.91 | 0.25 | 0.42 |
| 2000 | 135 | 0.64 | 86.51 | 71 | 0.85 | 59.87 | 64 | 0.41 | 26.64 | 0.20 | 0.35 |
| 2001 | 124 | 0.68 | 84.51 | 67 | 0.90 | 60.37 | 56 | 0.43 | 24.14 | 0.15 | 0.44 |
| 2002 | 113 | 0.73 | 82.75 | 64 | 0.95 | 61.03 | 49 | 0.44 | 21.72 | 0.13 | 0.35 |
| 2003 | 104 | 0.79 | 81.95 | 61 | 1.02 | 62.45 | 43 | 0.46 | 19.51 | 0.11 | 0.32 |
| 2004 | 98 | 0.83 | 81.10 | 59 | 1.09 | 63.66 | 40 | 0.44 | 17.44 | 0.10 | 0.52 |
| 2005 | 92 | 0.87 | 79.89 | 56 | 1.15 | 64.41 | 36 | 0.43 | 15.48 | 0.09 | 0.38 |


| year | total stock |  |  | mature |  |  | immature |  |  |  | recruit <br> age 3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number | mean wt | biomass | number | mean wt | biomass | number <br> (millions) | mean wt(kg) | biomass(1000t) | F(15+) |  |
|  | (millions) | (kg) | (1000t) | (millions) | (kg) |  |  |  |  |  | (millions) |
| 2006 | 92 | 0.84 | 78.05 | 52 | 1.22 | 64.13 | 40 | 0.35 | 13.91 | 0.10 | 1.08 |
| 2007 | 86 | 0.88 | 75.63 | 49 | 1.28 | 63.13 | 37 | 0.34 | 12.50 | 0.10 | 0.33 |
| 2008 | 82 | 0.90 | 73.58 | 46 | 1.34 | 62.08 | 35 | 0.33 | 11.50 | 0.10 | 0.49 |
| 2009 | 77 | 0.93 | 71.48 | 44 | 1.39 | 60.63 | 33 | 0.32 | 10.85 | 0.10 | 0.36 |
| 2010 | 74 | 0.92 | 67.86 | 41 | 1.42 | 57.50 | 33 | 0.31 | 10.36 | 0.12 | 0.51 |
| 2011 | 80 | 0.82 | 66.07 | 38 | 1.45 | 55.56 | 42 | 0.25 | 10.51 | 0.10 | 1.36 |
| 2012 | 93 | 0.70 | 64.94 | 37 | 1.46 | 53.64 | 56 | 0.20 | 11.29 | 0.10 | 2.03 |
| 2013 | 89 | 0.71 | 63.43 | 36 | 1.43 | 51.47 | 53 | 0.22 | 11.96 | 0.11 | 0.39 |
| 2014 | 82 | 0.76 | 62.65 | 36 | 1.41 | 50.07 | 47 | 0.27 | 12.57 | 0.10 | 0.03 |
| 2015 | 76 | 0.82 | 62.73 | 36 | 1.39 | 49.63 | 41 | 0.32 | 13.10 | 0.08 | 0.04 |
| 2016 | 95 | 0.65 | 62.00 | 35 | 1.37 | 47.86 | 60 | 0.23 | 14.14 | 0.12 | 2.58 |
| 2017 | 117 | 0.53 | 61.98 | 35 | 1.32 | 46.26 | 82 | 0.19 | 15.72 | 0.13 | 2.95 |
| 2018 | 114 | 0.53 | 60.04 | 35 | 1.24 | 43.26 | 79 | 0.21 | 16.78 | 0.19 | 0.77 |
| 2019 | 130 | 0.44 | 57.79 | 35 | 1.14 | 39.45 | 96 | 0.19 | 18.35 | 0.24 | 2.70 |
| 2020 | 118 | 0.46 | 54.15 | 34 | 1.02 | 35.03 | 83 | 0.23 | 19.12 | 0.31 | 0.03 |
| 2021 | 104 | 0.47 | 49.18 | 33 | 0.90 | 29.89 | 71 | 0.27 | 19.29 | 0.41 | 0.03 |



Figure 7.1. S. norvegicus in subareas 1 and 2. Total international landings 1908-2021 (in thousand tonnes).


Figure 7.2. S. norvegicus in subareas 1 and 2. Catches (including bycatch) of S. norvegiucs in 2021 from Norwegian logbooks. Due to reporting on the genus level these catches may contain a considerable amount of $S$. mentella.


Figure 7.3a. Illustration of the seasonality in the different Norwegian S. norvegicus fisheries in 2013-2021, also illustrating how the current regulations are working.


Figure 7.3b. Interannual changes in the Norwegian catches by fleet of S. norvegicus fisheries (2003-2021).


Figure 7.4. S. norvegicus. Length frequency of S. norvegicus reported from Norwegian catches in 2019-2021, all gears combined.


Figure 7.5a. Proportion maturity-at-age of S. norvegicus in subareas 1 and 2 derived from Norwegian commercial and survey data (Table E4). The proportions were derived from samples with at least five individuals. Updated for the 2022 assessment, but due to a lack of data in later years only the data up to 2018 was used in the model.


Figure 7.5b. S. norvegicus in subareas 1 and 2. Estimates of maturity-at-age by Gadget. Input data have been proportions of $S$. norvegicus mature both at age and length as collected and classified from Norwegian commercial landings and surveys.


Figure 7.6a. S. norvegicus. Abundance indices disaggregated by length for the winter Norwegian Barents Sea (Division 2.a) bottom trawl survey (BS-NoRu-Q1 (BTr); joint with Russia some of the years since 2000), for 1986-2022 (ref. Table

E2a). Numbers for 2022 are preliminary as Russian data were not available during AFWG 2022. Top: absolute index values, bottom: relative frequencies.


Figure 7.6b. S. norvegicus. Abundance indices by age from the winter Norwegian Barents Sea (Division 2.a) bottom-trawl survey (BS-NoRu-Q1 (BTr); joint with Russia some of the years since 2000), for 1992-2019 (ref. Table E2b). Age readings for 2017 and 2020-2022 not available during AFWG 2021. Top: absolute index, bottom: relative frequencies.


Figure 7.7a. S. norvegicus. Abundance indices disaggregated by length when combining the Norwegian bottom-trawl surveys 1986-2021 in the Barents Sea (winter) and at Svalbard (summer/autumn). Top: absolute index values. Bottom: relative frequencies. Horizontal line indicates the median length in the surveyed population.


Figure 7.7b. S. norvegicus. Abundance indices disaggregated by age. Combined Norwegian bottom-trawl surveys 19922018 in the Barents Sea (winter) and Svalbard survey (summer/autumn). Top: absolute index values, bottom: relative frequencies. Horizontal line indicates median age of the surveyed population. In 2009-2011, 2014-2015, 2017, 20192021 there was insufficient number of age readings to derive numbers-at-age.


Figure 7.8. S. norvegicus. Catch rates (numbers/nm) disaggregated by length for the Barents Sea coastal survey 19982021. Top: absolute catch rates. Bottom: relative values.


Figure 7.9. S. norvegicus in subareas 1 and 2. Comparison of observed and modelled survey indices (total number scaled to sum=100 during the period) for the Barents Sea winter survey in February. Dots: survey indices. Plain lines: survey indices estimated by the model.


Figure 7.10. S. norvegicus in subareas 1 and 2. Estimates of abundance-at-age 3-6 by Gadget for this year's assessment (solid line) and the last assessment (broken line), with data up to 2019 and 2021, respectively. Note that recent year (since 2015) have very little tuning data behind them.


Figure 7.11. S. norvegicus in subareas 1 and 2. Unweighted average fishing mortality of ages $\mathbf{1 5 +}$. Solid line shows this year's assessment (data up to 2021) and the dashed line shows last assessment (data up to 2019).


Figure 7.12. S. norvegicus in subareas 1 and 2. Stock numbers (in millions) and biomass (in $\mathbf{1 0 0 0}$ tonnes) for the total stock ( $3+$; upper panel), and the fishable and mature stock (middle panel), and the immature stock (lower panel), as estimated by Gadget using two surveys as input. Solid line shows this year's assessment (data up to 2021), and the dashed line shows last assessment (data up to 2019).


Figure 7.13. Gadget retrospective trends 2012 to 2019, immature biomass, mature biomass, recruitment-at-age 3, and F(15+).

### 7.5 Additional tables and figures

Table E1. Observed proportion of maturity-at-age 5 through 30 in S. norvegicus in subareas 1 and 2 derived from Norwegian commercial and survey data. The proportions were derived from samples with at least five individuals. Data for years after 2018 was considered insufficient until further age reading and is not presented.

| Year/Age | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.00 | 0.00 | 0.09 | 0.15 | 0.31 | 0.22 | 0.21 | 0.20 | 0.22 | 0.26 | 0.30 | 0.44 | 0.45 | 0.47 | 0.45 | 0.62 | 0.51 | 0.63 | 0.76 | 0.60 | 0.57 | 0.60 | 0.68 | 0.74 | 0.82 | 0.80 |
| 1993 | - | - | 0.00 | 0.00 | 0.10 | 0.29 | 0.54 | 0.47 | 0.53 | 0.67 | 0.80 | 0.75 | 0.78 | 0.82 | 0.91 | 0.85 | 0.82 | 0.87 | 0.75 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1994 | 0.00 | 0.00 | 0.03 | 0.05 | 0.28 | 0.28 | 0.32 | 0.70 | 0.79 | 0.91 | 0.94 | 0.85 | 0.92 | 1.00 | 0.96 | 0.96 | 1.00 | 0.88 | 1.00 | 1.00 | 1.00 | 1.00 | - | 1.00 | 1.00 | - |
| 1995 | 0.00 | 0.00 | 0.00 | 0.05 | 0.02 | 0.22 | 0.25 | 0.48 | 0.61 | 0.64 | 0.68 | 0.80 | 0.87 | 0.88 | 0.76 | 0.89 | 0.90 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 | - | - | - | - |
| 1996 | 0.00 | 0.05 | 0.14 | 0.13 | 0.22 | 0.38 | 0.43 | 0.60 | 0.64 | 0.75 | 0.69 | 0.77 | 0.90 | 0.85 | 0.91 | 0.88 | 0.96 | 0.93 | 1.00 | 0.87 | 0.95 | 0.95 | 1.00 | - | 1.00 | 0.86 |
| 1997 | 0.00 | 0.05 | 0.08 | 0.15 | 0.17 | 0.21 | 0.34 | 0.35 | 0.57 | 0.64 | 0.72 | 0.73 | 0.85 | 0.93 | 0.94 | 1.00 | 1.00 | 0.95 | 0.89 | 0.94 | 0.93 | 0.89 | 1.00 | 1.00 | 1.00 | - |
| 1998 | 0.00 | 0.00 | 0.03 | 0.11 | 0.09 | 0.26 | 0.32 | 0.49 | 0.52 | 0.69 | 0.74 | 0.77 | 0.81 | 0.91 | 0.89 | 0.86 | 1.00 | 1.00 | 0.67 | 0.70 | 1.00 | 1.00 | - | - | 1.00 | 0.88 |
| 1999 | 0.00 | 0.00 | 0.00 | 0.04 | 0.17 | 0.35 | 0.22 | 0.53 | 0.73 | 0.71 | 0.67 | 0.69 | 0.74 | 0.71 | 0.77 | 0.89 | - | 0.83 | - | 1.00 | 0.89 | - | - | - | - | - |
| 2000 | 0.00 | 0.08 | 0.14 | 0.25 | 0.40 | 0.51 | 0.59 | 0.62 | 0.65 | 0.69 | 0.78 | 0.96 | 0.96 | 1.00 | 1.00 | - | - | - | 1.00 | - | - | - | - | - | - | - |
| 2001 | - | 0.00 | 0.06 | 0.14 | 0.28 | 0.32 | 0.40 | 0.52 | 0.53 | 0.60 | 0.76 | 0.74 | 0.81 | 0.85 | 0.60 | 0.70 | 0.56 | - | - | - | - | - | - | - | - | - |
| 2002 | - | 0.00 | 0.05 | 0.07 | 0.23 | 0.44 | 0.41 | 0.63 | 0.74 | 0.93 | 0.77 | 0.89 | 0.90 | 0.94 | 0.96 | 0.92 | 0.95 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | - |
| 2003 | - | - | - | 0.05 | 0.13 | 0.24 | 0.24 | 0.47 | 0.58 | 0.68 | 0.75 | 0.65 | 0.77 | 0.78 | 0.93 | 0.96 | 0.94 | 0.67 | 1.00 | - | 1.00 | - | - | - | - | - |
| 2004 | - | - | 0.03 | 0.07 | 0.13 | 0.43 | 0.21 | 0.51 | 0.46 | 0.63 | 0.64 | 0.86 | 0.82 | 0.96 | 0.92 | 0.95 | 0.89 | 0.88 | 1.00 | 0.86 | 1.00 | - | - | - | - | - |
| 2005 | - | - | - | 0.04 | 0.39 | 0.16 | 0.33 | 0.40 | 0.41 | 0.57 | 0.74 | 0.81 | 0.78 | 0.82 | 0.78 | 0.94 | 0.95 | 0.88 | 0.83 | 1.00 | - | 1.00 | - | - | - | - |
| 2006 | - | - | - | 0.10 | 0.07 | 0.26 | 0.26 | 0.39 | 0.47 | 0.57 | 0.67 | 0.67 | 0.74 | 0.86 | 0.83 | 0.97 | 0.79 | 0.95 | 0.81 | 1.00 | - | 1.00 | - | - | - | - |


| Year/Age | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | - | - | - | 0.08 | 0.30 | 0.26 | 0.20 | 0.66 | 0.68 | 0.70 | 0.88 | 0.86 | 0.89 | 0.99 | 0.98 | 1.00 | 0.96 | 0.94 | 1.00 | 0.92 | 1.00 | 0.83 | 1.00 | 1.00 | 1.00 | - |
| 2008 | - | - | 0.80 | 0.25 | 0.82 | 0.68 | 0.62 | 0.80 | 0.79 | 0.86 | 0.88 | 0.91 | 0.90 | 0.92 | 0.92 | 0.90 | 0.93 | 0.93 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.93 | 1.00 |
| 2009 | - | - | - | - | - | 0.50 | 0.50 | 1.00 | 0.93 | 0.81 | 0.86 | 0.86 | 0.85 | 0.85 | 0.88 | 0.95 | 0.89 | 0.95 | 0.92 | 0.95 | 0.86 | 0.94 | 1.00 | 0.93 | 0.83 | 0.86 |
| 2010 | - | - | - | - | - | - | - | - | 0.78 | 0.77 | 0.87 | 1.00 | 0.64 | 0.93 | 0.91 | 1.00 | 0.95 | 0.90 | 1.00 | 0.73 | 0.80 | 0.83 | 1.00 | 0.60 | 0.60 | - |
| 2011 | - | - | - | - | - | - | - | - | - | - | 0.73 | 0.78 | 0.94 | 0.93 | 0.89 | 0.92 | 0.92 | 0.93 | 0.83 | 0.85 | 1.00 | 1.00 | - | 0.83 | - | - |
| 2012 | - | 0.11 | 0.10 | 0.29 | 0.20 | 0.20 | - | - | - | 0.76 | 0.72 | 0.70 | 0.91 | 0.78 | 0.88 | 0.89 | 0.85 | 0.81 | 0.95 | 0.81 | 0.86 | 1.00 | 0.93 | 1.00 | 1.00 | 1.00 |
| 2013 | - | 0.12 | 0.05 | 0.10 | 0.19 | 0.38 | 0.71 | - | 0.29 | 0.82 | 0.92 | 0.89 | 0.77 | 0.86 | 0.75 | 0.78 | 0.73 | 0.83 | 0.89 | 0.95 | 1.00 | 0.67 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2014 | - | - | 0.02 | 0.08 | 0.21 | 0.43 | 0.41 | 0.53 | 0.33 | 0.58 | 0.69 | 0.71 | 0.80 | 0.92 | 0.92 | 0.95 | 0.63 | 0.96 | 0.90 | 0.84 | 0.95 | 0.83 | 1.00 | - | 0.78 | 0.88 |
| 2015 | - | 0.05 | 0.17 | 0.17 | 0.30 | 0.41 | 0.44 | 0.49 | 0.65 | 0.67 | 0.69 | 0.81 | 0.91 | 0.86 | 0.83 | 0.93 | 0.78 | 0.82 | 1.00 | 0.95 | 0.96 | 0.83 | 0.84 | 1.00 | 0.87 | 0.82 |
| 2016 | - | 0.04 | 0.02 | 0.05 | 0.23 | 0.16 | 0.26 | 0.43 | 0.59 | 0.42 | 0.62 | 0.57 | 0.80 | 0.73 | 0.87 | 0.74 | 0.88 | 0.79 | 0.78 | 0.97 | 0.81 | 0.89 | 0.89 | 0.67 | 1.00 | 0.94 |
| 2017 | - | 0.06 | 0.06 | 0.20 | 0.22 | 0.31 | 0.40 | 0.60 | 0.53 | 0.71 | 0.86 | 0.71 | 0.86 | 0.94 | 0.92 | 0.95 | 1.00 | 0.96 | 0.84 | 0.94 | 0.93 | 0.94 | 0.92 | 0.82 | 0.87 | 1.00 |
| 2018 | - | - | - | - | 0.16 | 0.46 | 0.59 | 0.34 | 0.32 | 0.53 | 0.72 | 0.57 | 0.90 | 0.53 | 0.67 | 0.92 | - | 0.80 | 0.75 | 1.00 | 1.00 | 0.78 | 0.63 | 1.00 | - | - |

Table E2a. S. norvegicus in subareas 1 and 2. Abundance indices (numbers in millions) - on length - from the winter Norwegian Barents Sea (Division 2.a) bottom-trawl survey (BS-NoRu-Q1 (BTr)) from 1986 to 2022. The area coverage was extended from 1993. Indices recalculated from 1994 onwards.

| Length group (cm) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 5.0-9.9 | 10.0-14.9 | 15.0-19.9 | 20.0-24.9 | 25.0-29.9 | 30.0-34.9 | 35.0-39.9 | 40.0-44.9 | > 45.0 | Total |
| 1986 | 3.0 | 11.7 | 26.4 | 34.3 | 17.7 | 21.0 | 12.8 | 4.4 | 2.6 | 133.9 |
| 1987 | 7.7 | 12.7 | 32.8 | 7.7 | 6.4 | 3.4 | 3.8 | 3.8 | 4.2 | 82.5 |
| 1988 | 1.0 | 5.6 | 5.5 | 14.2 | 12.6 | 7.3 | 5.2 | 4.1 | 3.7 | 59.2 |


| Length group (cm) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 5.0-9.9 | 10.0-14.9 | 15.0-19.9 | 20.0-24.9 | 25.0-29.9 | 30.0-34.9 | 35.0-39.9 | 40.0-44.9 | > 45.0 | Total |
| 1989 | 48.7 | 4.9 | 4.3 | 11.8 | 15.9 | 12.2 | 6.6 | 4.8 | 3.0 | 112.2 |
| 1990 | 9.2 | 5.3 | 6.5 | 9.4 | 15.5 | 14.0 | 8.0 | 4.0 | 3.4 | 75.3 |
| 1991 | 4.2 | 13.6 | 8.4 | 19.4 | 18.0 | 16.1 | 14.8 | 6.0 | 4.0 | 104.5 |
| 1992 | 1.8 | 3.9 | 7.7 | 20.6 | 19.7 | 13.7 | 10.5 | 6.6 | 5.8 | 90.3 |
| 1993 | 0.1 | 1.2 | 3.5 | 6.9 | 10.3 | 14.5 | 12.5 | 8.6 | 6.3 | 63.9 |
| 1994 | 0.7 | 7.5 | 10.1 | 12.8 | 10.9 | 17.8 | 10.1 | 4.8 | 2.9 | 77.6 |
| 1995 | 0.4 | 4.7 | 13.5 | 13.1 | 10.4 | 15.4 | 16.2 | 10.6 | 4.6 | 88.9 |
| 1996 | 0.0 | 0.7 | 3.3 | 5.9 | 8.7 | 14.0 | 15.7 | 7.5 | 3.9 | 59.7 |
| 1997 | 0.0 | 0.3 | 1.0 | 2.2 | 5.1 | 20.3 | 28.0 | 8.5 | 3.3 | 68.8 |
| 1998 | 0.1 | 2.4 | 1.3 | 2.6 | 4.5 | 7.4 | 7.5 | 5.1 | 2.2 | 33.0 |
| 1999 | 0.2 | 0.9 | 2.1 | 4.0 | 4.4 | 6.3 | 6.1 | 5.5 | 3.5 | 32.4 |
| 2000 | 0.5 | 1.1 | 1.5 | 4.2 | 4.9 | 5.1 | 3.6 | 1.9 | 1.2 | 23.9 |
| 2001 | 0.1 | 0.4 | 0.4 | 2.5 | 5.8 | 5.4 | 4.5 | 3.2 | 1.7 | 24.1 |
| 2002 | 0.1 | 1.0 | 2.0 | 1.8 | 3.9 | 4.2 | 3.2 | 3.5 | 2.4 | 22.3 |
| 2003 | 0.0 | 0.5 | 1.3 | 1.5 | 4.2 | 4.1 | 2.8 | 3.2 | 3.0 | 20.5 |
| 2004 | 0.7 | 0.2 | 0.4 | 1.0 | 2.8 | 4.4 | 5.4 | 3.9 | 3.0 | 21.8 |
| 2005 | 0.0 | 0.1 | 0.2 | 0.4 | 1.1 | 2.1 | 3.8 | 4.7 | 4.4 | 16.8 |


| Length group (cm) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 5.0-9.9 | 10.0-14.9 | 15.0-19.9 | 20.0-24.9 | 25.0-29.9 | 30.0-34.9 | 35.0-39.9 | 40.0-44.9 | > 45.0 | Total |
| 2006 | 0.0 | 0.0 | 0.0 | 0.2 | 2.5 | 5.5 | 6.3 | 4.2 | 4.3 | 22.9 |
| 2007 | 0.0 | 0.1 | 0.3 | 0.1 | 0.5 | 1.3 | 2.7 | 4.4 | 4.3 | 13.7 |
| 2008 | 1.7 | 2.5 | 0.2 | 0.2 | 0.4 | 0.7 | 2.0 | 2.5 | 4.5 | 14.7 |
| 2009 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.4 | 1.7 | 3.8 | 6.6 | 12.7 |
| 2010 | 0.4 | 2.0 | 1.1 | 0.5 | 0.1 | 0.1 | 0.9 | 1.1 | 4.0 | 10.2 |
| 2011 | 0.3 | 3.2 | 2.1 | 0.3 | 0.4 | 0.1 | 0.3 | 2.3 | 5.3 | 14.4 |
| 2012 | 0.8 | 4.4 | 4.0 | 1.8 | 0.5 | 0.3 | 0.9 | 3.6 | 6.3 | 22.7 |
| 2013 | 0.1 | 7.4 | 4.9 | 4.0 | 1.6 | 0.4 | 0.9 | 0.8 | 3.7 | 23.7 |
| 2014 | 0.1 | 1.0 | 1.5 | 3.0 | 3.3 | 1.0 | 0.5 | 1.4 | 4.1 | 16.0 |
| 2015 | 0.1 | 0.9 | 1.5 | 3.0 | 2.6 | 2.0 | 0.5 | 0.7 | 3.4 | 14.7 |
| 2016 | 0.7 | 1.3 | 1.5 | 2.3 | 4.2 | 3.6 | 3.4 | 1.7 | 5.8 | 24.3 |
| 2017 | 0.3 | 1.3 | 0.9 | 1.1 | 4.5 | 9.1 | 6.7 | 3.0 | 5.0 | 31.7 |
| 2018 | 1.1 | 2.7 | 1.8 | 1.7 | 3.3 | 4.7 | 6.3 | 4.3 | 4.7 | 30.6 |
| 2019 | 0.7 | 3.2 | 1.7 | 1.8 | 2.5 | 3.9 | 9.0 | 9.7 | 9.1 | 41.7 |
| 2020 | 1.0 | 0.6 | 1.5 | 1.0 | 1.9 | 2.4 | 6.5 | 8.8 | 9.9 | 33.6 |
| 2021 | 0.1 | 0.6 | 1.9 | 2.3 | 1.5 | 2.4 | 4.9 | 6.3 | 9.6 | 29.8 |
| $2022^{1}$ | 1.7 | 1.7 | 0.4 | 0.8 | 1.7 | 1.3 | 4.7 | 4.7 | 5.8 | 23.0 |

1 - Provisional figures. Russian data not provided in time for AFWG 2022.

Table E2b. S. norvegicus in subareas 1 and 2. Norwegian bottom-trawl indices (numbers in thousands) - on age - from the annual Winter Norwegian Barents Sea (Division 2.a) bottom trawl survey (BS-NoRu-Q1 (BTr)) from 1986 to 2019. Age readings not available for 2017 and 2020-2022 at the time of AFWG 2022. The area coverage was extended from 1993 onwards. Indices recalculated from 1994 and onwards.

| Year/age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 2509 | 4070 | 6395 | 2375 | 3757 | 10392 | 4299 | 3567 | 11526 | 2276 | 3239 | 3070 | 3666 | 15183 | 76324 |
| 1993 | 996 | 1308 | 1661 | 3005 | 1559 | 7689 | 3346 | 4801 | 2712 | 5480 | 6568 | 2735 | 8801 | 28737 | 79398 |
| 1994 | 0 | 9249 | 2475 | 5998 | 10871 | 6530 | 3523 | 8189 | 4566 | 1639 | 6285 | 1486 | 2964 | 11035 | 74809 |
| 1995 | 3544 | 4554 | 7203 | 9362 | 5598 | 8583 | 3308 | 2305 | 5004 | 7512 | 4602 | 4848 | 5948 | 15455 | 87826 |
| 1996 | 365 | 800 | 1825 | 2917 | 3715 | 8299 | 5343 | 3038 | 6373 | 4653 | 5945 | 3113 | 3720 | 9357 | 59462 |
| 1997 | 154 | 37 | 489 | 1012 | 1588 | 2717 | 3764 | 2925 | 9098 | 6036 | 12131 | 11643 | 2430 | 14607 | 68629 |
| 1998 | 1604 | 1118 | 607 | 550 | 858 | 2233 | 2470 | 2310 | 2157 | 3345 | 4618 | 827 | 2785 | 7320 | 32803 |
| 1999 | 489 | 1079 | 1289 | 2708 | 1220 | 1315 | 2060 | 3177 | 1766 | 3129 | 5342 | 2053 | 2085 | 4828 | 32537 |
| 2000 | 437 | 427 | 588 | 1774 | 2274 | 2559 | 1814 | 2378 | 1850 | 1817 | 2396 | 1838 | 336 | 2089 | 22577 |
| 2001 | 322 | 105 | 280 | 583 | 1346 | 2759 | 3072 | 2603 | 2488 | 2511 | 1886 | 1377 | 1016 | 3552 | 23903 |
| 2002 | 973 | 919 | 796 | 1126 | 640 | 1511 | 2744 | 1694 | 1754 | 2144 | 1090 | 1102 | 2172 | 3492 | 22157 |
| 2003 | 165 | 88 | 773 | 1329 | 523 | 1154 | 2638 | 1391 | 2140 | 1330 | 1890 | 801 | 1165 | 4809 | 20197 |
| 2004 | 0 | 163 | 68 | 250 | 544 | 978 | 1513 | 1069 | 1110 | 2135 | 3150 | 1559 | 2832 | 5541 | 20911 |
| 2005 | 57 | 85 | 86 | 114 | 393 | 532 | 627 | 460 | 689 | 1095 | 1178 | 1713 | 1545 | 8244 | 16818 |
| 2006 | 0 | 0 | 0 | 0 | 26 | 1025 | 1157 | 2641 | 2424 | 1244 | 1888 | 3242 | 1795 | 7480 | 22922 |
| 2007 | 19 | 39 | 256 | 39 | 0 | 320 | 173 | 369 | 293 | 868 | 751 | 809 | 847 | 8941 | 13724 |
| 2008 | 826 | 0 | 0 | 0 | 76 | 97 | 116 | 224 | 477 | 320 | 623 | 885 | 621 | 6744 | 11010 |


| Year/age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 0 | 0 | 0 | 0 | 0 | 0 | 12 | 80 | 176 | 220 | 1168 | 417 | 1018 | 9507 | 12598 |
| 2010 | 0 | 0 | 328 | 1012 | 250 | 0 | 364 | 62 | 0 | 96 | 343 | 264 | 345 | 4955 | 8018 |
| 2011 | 2001 | 1750 | 1283 | 135 | 64 | 0 | 440 | 0 | 103 | 0 | 214 | 119 | 560 | 7110 | 13776 |
| 2012 | 938 | 3955 | 4777 | 547 | 342 | 267 | 391 | 112 | 102 | 86 | 0 | 247 | 506 | 9811 | 22083 |
| 2013 | 1594 | 1773 | 4772 | 2651 | 2504 | 2050 | 1386 | 275 | 0 | 483 | 143 | 166 | 0 | 4925 | 22721 |
| 2014 | 485 | 985 | 724 | 1030 | 2856 | 1906 | 1048 | 532 | 0 | 262 | 228 | 113 | 513 | 5056 | 15737 |
| 2015 | 223 | 438 | 814 | 1034 | 1481 | 1909 | 1947 | 483 | 943 | 484 | 471 | 104 | 53 | 4130 | 14514 |
| 2016 | 338 | 557 | 408 | 390 | 1163 | 2022 | 2567 | 2214 | 1027 | 805 | 2392 | 1324 | 555 | 7162 | 22925 |
| 2017 |  | Age data not available during AFWG 2022 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2018 | 1597 | 1016 | 892 | 354 | 696 | 1784 | 2627 | 1082 | 1596 | 2558 | 2358 | 3461 | 1307 | 7626 | 28953 |
| 2019 | 899 | 1684 | 780 | 2120 | 900 | 1240 | 2821 | 3276 | 5770 | 7289 | 3393 | 2170 | 983 | 5251 | 38577 |

16+ group is considered in the calculation since 2005. Values prior to this date were derived by subtracting the sum of abundance in groups 1-15 to the total abundance, available in Table E1a.

Table E3a. S. norvegicus in subareas 1 and 2. Abundance indices (numbers in thousands) - on length - from the Norwegian Svalbard (Division 2.b) bottom-trawl survey (August-September) from 1985 to 2021. Since 2005 this is part of the Ecosystem survey (Eco-NoRu-Q3 (BTr)).

| Year | Length group (cm) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 5.0-9.9 | 10.0-14.9 | 15.0-19.9 | 20.0-24.9 | 25.0-29.9 | 30.0-34.9 | 35.0-39.9 | 40.0-44.9 | > 45.0 | Total |
| $1985{ }^{1}$ | - | 1307 | 795 | 1728 | 2273 | 1417 | 311 | 142 | 194 | 8167 |
| $1986{ }^{1}$ | 200 | 2961 | 1768 | 547 | 643 | 1520 | 639 | 467 | 196 | 8941 |


| Year | Length group (cm) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 5.0-9.9 | 10.0-14.9 | 15.0-19.9 | 20.0-24.9 | 25.0-29.9 | 30.0-34.9 | 35.0-39.9 | 40.0-44.9 | > 45.0 | Total |
| $1987{ }^{1}$ | 100 | 1343 | 1964 | 1185 | 1367 | 652 | 352 | 29 | 44 | 7036 |
| $1988{ }^{1}$ | 500 | 1001 | 1953 | 1609 | 684 | 358 | 158 | 68 | 95 | 6426 |
| 1989 | 200 | 1629 | 2963 | 2374 | 1320 | 846 | 337 | 323 | 104 | 10096 |
| 1990 | 1700 | 3886 | 4478 | 4047 | 2972 | 1509 | 365 | 140 | 122 | 19219 |
| 1991 | 100 | 5371 | 5821 | 9171 | 8523 | 4499 | 1531 | 982 | 395 | 36393 |
| 1992 | 1700 | 10228 | 8858 | 5330 | 13960 | 12720 | 4547 | 494 | 346 | 58183 |
| 1993 | 200 | 10160 | 9078 | 5855 | 7071 | 4327 | 2088 | 1552 | 948 | 41279 |
| 1994 | 100 | 3340 | 5883 | 4185 | 3922 | 3315 | 1021 | 845 | 423 | 23034 |
| 1995 | 470 | 2000 | 9100 | 5070 | 3060 | 2400 | 1040 | 920 | 780 | 24840 |
| 1996 | 80 | 130 | 1260 | 2480 | 1030 | 480 | 550 | 990 | 400 | 7400 |
| 1997 | 0 | 810 | 1980 | 5470 | 5560 | 2340 | 590 | 190 | 450 | 17390 |
| 1998 | 180 | 2698 | 1741 | 4620 | 4053 | 1761 | 535 | 545 | 241 | 16374 |
| 1999 | 0 | 794 | 7057 | 3698 | 4563 | 2449 | 467 | 619 | 369 | 20016 |
| 2000 | 40 | 360 | 1240 | 1390 | 2010 | 760 | 400 | 160 | 390 | 6750 |
| 2001 | 10 | 110 | 790 | 1470 | 3710 | 4600 | 1880 | 680 | 370 | 13620 |
| 2002 | 0 | 0 | 65 | 415 | 459 | 880 | 621 | 565 | 521 | 3526 |
| 2003 | 87 | 87 | 104 | 84 | 534 | 635 | 459 | 759 | 738 | 3487 |


| Length group (cm) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 5.0-9.9 | 10.0-14.9 | 15.0-19.9 | 20.0-24.9 | 25.0-29.9 | 30.0-34.9 | 35.0-39.9 | 40.0-44.9 | > 45.0 | Total |
| 2004 | 0 | 8 | 9 | 192 | 581 | 667 | 607 | 395 | 213 | 2672 |
| 2005 | 0 | 52 | 0 | 84 | 267 | 608 | 411 | 274 | 283 | 1979 |
| 2006 | 0 | 0 | 75 | 74 | 138 | 437 | 470 | 668 | 1264 | 3126 |
| 2007 | 0 | 29 | 52 | 938 | 1069 | 4268 | 5154 | 892 | 1390 | 13792 |
| 2008 | 8603 | 4255 | 211 | 25 | 50 | 169 | 525 | 180 | 536 | 14554 |
| 2009 | 216 | 1403 | 108 | 108 | 0 | 0 | 197 | 214 | 220 | 2466 |
| 2010 | 868 | 1117 | 1845 | 607 | 0 | 123 | 189 | 0 | 996 | 5745 |
| 2011 | 0 | 0 | 850 | 50 | 0 | 0 | 0 | 159 | 578 | 1637 |
| 2012 | 0 | 111 | 1565 | 2242 | 2217 | 285 | 0 | 0 | 146 | 6566 |
| 2013 | 56 | 489 | 2155 | 3307 | 2738 | 433 | 136 | 34 | 349 | 9697 |
| 2014 | 64 | 0 | 425 | 167 | 296 | 531 | 74 | 0 | 312 | 1869 |
| 2015 | 0 | 0 | 0 | 216 | 198 | 303 | 877 | 18 | 810 | 2422 |
| 2016 | 0 | 0 | 121 | 119 | 813 | 1007 | 754 | 300 | 498 | 3612 |
| 2017 | 838 | 675 | 577 | 93 | 585 | 291 | 476 | 288 | 262 | 4085 |
| 2018 | 826 | 11129 | 5619 | 1000 | 677 | 2741 | 1134 | 127 | 110 | 23363 |
| 2019 | 78 | 90 | 104 | 219 | 68 | 0 | 115 | 131 | 182 | 987 |
| 2020 | 527 | 1193 | 1728 | 1591 | 290 | 368 | 318 | 365 | 264 | 6644 |


| Length group (cm) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 5.0-9.9 | 10.0-14.9 | 15.0-19.9 | 20.0-24.9 | 25.0-29.9 | 30.0-34.9 | 35.0-39.9 | 40.0-44.9 | > 45.0 | Total |
| 2021 | 0 | 184 | 1277 | 1849 | 1074 | 95 | 407 | 20 | 69 | 4975 |

1 - Old trawl equipment (bobbins gear and 80 m sweep length).
Table E3b. S. norvegicus in subareas 1 and 2. Norwegian bottom trawl survey indices-on age-from the Norwegian Svalbard (Division 2.b) bottom trawl survey (August-September) from 1985 to 2019. Since 2005 this is part of the Ecosystem survey (Eco-NoRu-Q3 (BTr)). In 2009-2011, 2014-2015 and 2019-2021, there was insufficient number of age readings to derive numbers-at-age, or age readings were not available at the time of the AFWG 2022.

|  | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | Total |
| 1992 | 284 | 12378 | 5576 | 2279 | 371 | 2064 | 3687 | 5704 | 9215 | 6413 | 1454 | 1387 | 696 | 22 | 51530 |
| 1993 | 32 | 10704 | 5710 | 5142 | 1855 | 1052 | 1314 | 3520 | 2847 | 2757 | 2074 | 1245 | 844 | 119 | 39215 |
| 1994 | 429 | 1150 | 3418 | 2393 | 1723 | 1106 | 1714 | 1256 | 1938 | 1596 | 2039 | 484 | 550 | 319 | 20115 |
| 1995 | 600 | 1600 | 6400 | 5100 | 1800 | 2200 | 1800 | 700 | 700 | 400 | 700 | 500 | 400 | 500 | 23400 |
| 1996 | 40 | 110 | - | 560 | 1050 | 940 | 930 | 400 | 1050 | 280 | 320 | 590 | 160 | 70 | 6500 |
| 1997 | 320 | 490 | - | 480 | 1500 | 6950 | 2720 | 1680 | 800 | 1310 | 550 | 30 | - | 120 | 16950 |
| 1998 | 210 | 1817 | 881 | 202 | 1555 | 2187 | 4551 | 1913 | 1010 | 797 | 49 | 264 | 73 | 187 | 15696 |
| 1999 | 0 | 760 | 2893 | 1339 | 3534 | 1037 | 3905 | 2603 | 762 | 1663 | 481 | 361 | 258 | 152 | 19748 |
| 2000 | 40 | 20 | 400 | 350 | 840 | 480 | 730 | 1670 | 620 | 340 | 510 | 100 | 80 | 70 | 6250 |
| 2001 | 0 | 40 | 50 | 450 | 330 | 790 | 1760 | 1970 | 3300 | 1200 | 1810 | 150 | 660 | 430 | 12940 |
| 2002 | 0 | 0 | - | - | 65 | 160 | 204 | 326 | 364 | 614 | 442 | 328 | 15 | 0 | 2518 |
| 2003 | 0 | 0 | 0 | 0 | 95 | 0 | 283 | 227 | 93 | 296 | 285 | 189 | 228 | 341 | 2035 |


|  | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | Total |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 359 | 144 | 362 | 152 | 343 | 315 | 316 | 220 | 2209 |
| 2005 | 0 | 50 | 0 | 0 | 0 | 73 | 25 | 286 | 106 | 191 | 271 | 167 | 125 | 152 | 1447 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 71 | 0 | 0 | 233 | 106 | 174 | 194 | 305 | 179 | 1261 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 513 | 776 | 399 | 0 | 0 | 292 | 1752 | 1759 | 1349 | 6841 |
| 2008 | 7844 | 0 | 0 | 0 | 0 | 0 | 0 | 37 | 98 | 16 | 18 | 148 | 86 | 164 | 8412 |
| 2009 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 2010 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 2011 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 2012 | 0 | 40 | 123 | 2445 | 2105 | 1205 | 642 | 92 | 35 | 0 | 0 | 0 | 0 | 0 | 6687 |
| 2013 | 0 | 56 | 383 | 1532 | 3963 | 377 | 1910 | 1029 | 214 | 121 | 250 | 0 | 0 | 166 | 10000 |
| 2014 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 2015 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 2016 | 0 | 0 | 124 | 0 | 0 | 0 | 0 | 813 | 455 | 739 | 0 | 483 | 136 | 263 | 3015 |
| 2017 | 356 | 187 | 322 | 97 | 145 | 130 | 193 | 205 | 79 | 292 | 205 | 176 | 278 | 0 | 2667 |
| 2018 | 543 | 0 | 1363 | 4066 | 0 | 367 | 885 | 422 | 0 | 970 | 1625 | 0 | 0 | 0 | 10239 |

Table E4．S．norvegicus in Sub－area 1 and 2．Mean catch rates（numbers／nm）of S．norvegicus from the Norwegian Coastal Surveys（NOcoast－Aco－Q4；Division 2．a）in 1998－2021．

| Length range （cm） | O | ம் | $\begin{aligned} & \underset{7}{\text { I }} \end{aligned}$ |  | $\underset{\sim}{\underset{N}{\mathbf{N}}}$ | $\begin{gathered} \underset{\sim}{\dot{N}} \end{gathered}$ | $\begin{aligned} & \text { む } \\ & \text { ìm } \end{aligned}$ | $\begin{aligned} & \text { ON} \\ & \underset{\sim n}{n} \end{aligned}$ | $\begin{aligned} & \text { 寸 } \\ & \text { í } \end{aligned}$ | $\begin{aligned} & \text { ஷ寸 } \\ & \text { ஷி } \end{aligned}$ | $\begin{aligned} & \text { Ư } \\ & \text { ì } \end{aligned}$ | $\begin{aligned} & \text { On } \\ & \text { ஸin } \end{aligned}$ | : | $\begin{aligned} & \frac{n}{5} \\ & \frac{\widetilde{T}}{1} \\ & \# \end{aligned}$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 0 | 0 | 692 | 6632 | 73075 | 22255 | 22430 | 130161 | 116216 | 23519 | 2547 | 880 | 0 | 89 | 139 | 778 | NA | 43574 |
| 1999 | 0 | 7587 | 77067 | 317802 | 369258 | 165769 | 67222 | 178802 | 163919 | 20445 | 3642 | 1520 | 0 | 103 | 138 | 2144 | NA | 43574 |
| 2000 | 0 | 0 | 1856 | 13048 | 6459 | 13065 | 42990 | 156418 | 171407 | 29117 | 3036 | 331 | 191 | 99 | 144 | 756 | 503 | 43574 |
| 2001 | 0 | 295 | 2031 | 11787 | 12305 | 22408 | 14127 | 74790 | 150763 | 26573 | 1787 | 345 | 191 | 81 | 113 | 460 | 325 | 43574 |
| 2002 | 0 | 0 | 0 | 0 | 2321 | 7588 | 34283 | 1011273 | 754947 | 26769 | 3195 | 513 | 0 | 109 | 172 | 3289 | 332 | 43574 |
| 2003 | 0 | 0 | 2579 | 10118 | 44506 | 72473 | 52479 | 224734 | 228374 | 62121 | 5536 | 481 | 0 | 123 | 160 | 1367 | 1053 | 43574 |
| 2004 | 0 | 937 | 3139 | 5591 | 21042 | 66182 | 34613 | 351154 | 552183 | 41851 | 2666 | 1345 | 0 | 104 | 130 | 1290 | 950 | 43574 |
| 2005 | 0 | 554 | 5209 | 4627 | 30272 | 46072 | 48379 | 189993 | 170639 | 37468 | 1450 | 0 | 0 | 99 | 132 | 833 | 780 | 43574 |
| 2006 | 0 | 0 | 2884 | 496 | 1738 | 3065 | 29933 | 144743 | 256394 | 65959 | 9272 | 0 | 0 | 112 | 112 | 771 | 680 | 43574 |
| 2007 | 0 | 0 | 0 | 0 | 4335 | 7308 | 17338 | 129412 | 177332 | 29042 | 1182 | 0 | 0 | 131 | 140 | 637 | 637 | 43574 |
| 2008 | 0 | 3644 | 4555 | 955 | 3957 | 4679 | 17440 | 362633 | 490611 | 99469 | 11772 | 1630 | 0 | 110 | 139 | 1156 | 850 | 43574 |
| 2009 | 0 | 0 | 6976 | 2285 | 2984 | 4530 | 39275 | 800208 | 945004 | 106479 | 6244 | 663 | 1122 | 114 | 136 | 2947 | 598 | 43574 |
| 2010 | 0 | 39758 | 77542 | 20364 | 8814 | 1378 | 2582 | 66948 | 214182 | 99061 | 7417 | 2454 | 0 | 117 | 136 | 833 | 690 | 43574 |
| 2011 | 0 | 3654 | 67407 | 55725 | 193640 | 35323 | 10043 | 72244 | 296697 | 107318 | 27832 | 286 | 0 | 113 | 104 | 998 | 571 | 43574 |
| 2012 | 0 | 39530 | 59337 | 95227 | 150260 | 89534 | 12686 | 58890 | 356556 | 163645 | 46792 | 4640 | 263 | 98 | 96 | 1191 | 778 | 43574 |
| 2013 | 0 | 5176 | 137751 | 72253 | 540679 | 260689 | 38079 | 34628 | 384207 | 190595 | 21534 | 3528 | 2091 | 93 | 95 | 2231 | 1105 | 43574 |


| Length range （cm） | ¢ | in |  | $\begin{aligned} & \text { ন } \\ & \text { ñ } \end{aligned}$ | $\begin{gathered} \underset{N}{N} \\ \underset{N}{2} \end{gathered}$ | $\begin{aligned} & \text { ণ̀ } \\ & \dot{N} \end{aligned}$ | $\begin{aligned} & \text { H } \\ & \text { íd } \end{aligned}$ | $\begin{aligned} & \text { ON } \\ & \text { in } \end{aligned}$ | \& | $\begin{aligned} & \text { g寸 } \\ & \text { U } \end{aligned}$ | $\begin{aligned} & \text { せ } \\ & \text { ì } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { గ̛̣ } \\ & \text { ஸin } \end{aligned}$ | $\stackrel{0}{1}$ | $\begin{aligned} & \frac{n}{工} \\ & \frac{\pi}{T} \\ & \# \end{aligned}$ |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2014 | 0 | 524 | 28653 | 89876 | 78267 | 144543 | 109523 | 47736 | 302185 | 157358 | 30251 | 2343 | 3361 | 107 | 108 |  | 1717 | 777 | 43574 |
| 2015 | 0 | 5081 | 69615 | 93690 | 193721 | 189891 | 246181 | 77869 | 202765 | 163442 | 41838 | 3335 | 0 | 97 | 103 |  | 1886 | 984 | 43574 |
| 2016 | 0 | 0 | 100206 | 49233 | 177926 | 186202 | 81997 | 49197 | 145043 | 163426 | 41278 | 869 | 567 | 99 | 101 |  | 1648 | 1153 | 43574 |
| 2017 | 0 | 1789 | 51611 | 101305 | 67426 | 140564 | 205389 | 191361 | 182391 | 134508 | 21507 | 1130 | 515 | 110 | 147 |  | 2996 | 1866 | 43574 |
| 2018 | 0 | 509 | 5230 | 16112 | 43173 | 50831 | 52728 | 124778 | 273489 | 200310 | 67433 | 4181 | 988 | 154 | 220 |  | 2182 | 1837 | 43574 |
| 2019 | 0 | 646 | 10371 | 6780 | 31170 | 26133 | 34875 | 145733 | 303319 | 158832 | 48546 | 1234 | 635 | 159 | 182 |  | 1856 | 1363 | 43574 |
| 2020 | 0 | 8763 | 19753 | 7782 | 16762 | 75324 | 104097 | 184328 | 200398 | 113592 | 40320 | 4186 | 475 | 136 | 201 |  | 3338 | 1703 | 43574 |
| 2021 | 2786 | 28669 | 51554 | 12878 | 4767 | 41451 | 78399 | 142549 | 404448 | 238166 | 60729 | 530 | 470 | 127 | 160 |  | 2482 | 1484 | 43574 |



Figure E1. Overview of the Norwegian biological age samples (number individuals, number hauls/sets, number of boats) from the commercial fisheries for S. norvegicus in 2013 representing more than $80 \%$ of the catches and which the input data to the Gadget model are based upon. The colours denote which sampling platform has been used: High Seas Reference fleet, port sampling, Coast guard, Coastal Reference Fleet, or inspectors/observers at sea. The green crosses show the catch in tonnes for the different seasons, areas and gears.


Figure E2. S. norvegicus in subareas 1 and 2. Yield-per-recruit for S. norvegicus, computed from the GADGET assessment model presented at the benchmark assessment in January 2018 (WKREDFISH, ICES 2018a).

## 8 Northeast Arctic Greenland halibut ${ }^{1}$

### 8.1 Status of the fisheries

### 8.1.1 Landings prior to 2022 (Tables 8.1-8.8, Figures 8.1-8.3)

Nominal landings by country for subareas 1 and 2 combined are presented in Table 8.1. Tables 8.2 to 8.4 give the landings for Subarea 1 and divisions 2 .a and $2 . b$ separately, and landings separated by gear type are presented in Table 8.5. For most countries, the landings listed in the tables are similar to those officially reported to ICES. Some of the values in the tables vary slightly from the official statistics and represent those presented to the Working Group by the members. Catch per unit effort is presented in Table 8.6 and total catch from 1935 to now in Table 8.7 and Figure 8.1.

The preliminary estimate of the total landings for 2021 is 28431 t . This is 282 t less than the landings in 2020 and about 5431 t more than the ICES advised maximum catch for 2021 ( 23000 t ). The catches from most countries remained stable, compared to 2020. Combined landings exceeded the quotas set by the Joint Russian-Norwegian Fisheries Commission for 2021 by 1431 t (total TAC 27000 t ). One explanation is the difficulties in bycatch regulation. Also, catches in the report include all landings in ICES 1 and 2, and thus include catches in EU waters in the southern part of ICES 2.

Some fishing for Greenland halibut has taken place in the northern part of Division 4.a during the past 20-30 years, varying between a few tonnes and up to 1670 t in 1995 and 2577 in 1999. From 2005 to 2011 this catch was mostly below 200 t, taken mostly by Norway, France, and the UK. Preliminary numbers show 144 t in 2021, a reduction from 719 t the year before mainly due to that the Norwegian trawl fleets did not have access to British waters in 2021 (Table 8.8, Figures 8.2 and 8.3). Although there is a continuous distribution of this species from the southern part of Division 2a along the continental slope towards the Shetland area, the stock structure is unclear in this area and these landings have therefore not been added to the total from subareas 1 and 2. Recent mark-recapture and genetic investigations indicate that the stock might have a more south and westward distribution than the current ICES definition of the stock boundaries (Albert and Vollen, 2015; Westgaard et al., 2016).

### 8.1.2 ICES advice applicable to 2021-2023

The roll over advice from ICES for 2021 was as follows:
ICES advises that when the precautionary approach is applied, catches in 2020 should be no more than 23000 tonnes. This corresponds to a harvest rate of $\approx 0.036$. All catches are assumed to be landed.

## Last advice:

ICES advises that when the precautionary approach is applied, catches in the year 2022 should be no more than 19094 tonnes and catches in the year 2023 should be no more than 18494 tonnes.

[^11]
### 8.1.2.1 Additional considerations

A benchmark and data workshop process led to an agreed analytic assessment in 2015.
A benchmark meeting (WKBUT; ICES 2013/ACOM:44) was held for the Northeast Arctic (NEA) Greenland halibut in 2013, but the benchmark process was prolonged due to problems with data. A data workshop was conducted in November 2014 (DCWKNGHD ICES CM 2014/ACOM:65), followed by a benchmark by correspondence that ended in 2015. The assessment is reported in the benchmark by correspondence (IBPHALI; ICES CM 2015/ACOM:54) and in the stock annex.

A new benchmark is planned in early 2023.

### 8.1.3 Management

The $38^{\text {th }}$ JRNFC's session in 2009 decided to cancel the ban against targeted Greenland halibut fishery and established the TAC at 15000 t for the next three years (2010-2012). The $40^{\text {th }}$ JRNFC Session in 2011 decided to increase the TAC for 2012 up to 18000 t , and at the $42^{\text {nd }}$ JRNFC Session in 2012, the TAC for 2013 was increased to 19000 t . The $43^{\text {rd }}$ and $44^{\text {th }}$ sessions kept the same TAC for 2014 and 2015. For 2016 and 2017 TAC was set to 22 and 24 thousand tonnes, respectively. The TAC for 2018 was 27 thousand tonnes and the same for 2019, 2020 and 2021.

The TAC for Greenland halibut set by JNRFC applies to catches in ICES areas 1, $2 a$ and $2 b$, except the Jan Mayen EEZ and the part of the EU EEZ which is north of $62^{\circ} \mathrm{N}$.

In 2021 catches of 32 tonnes were taken in the Jan Mayen area (within ICES Subarea 2), where Greenland halibut fisheries are not regulated by TAC.

Norway previously had a quota for Greenland halibut in the EU EEZ which could be fished in ICES areas 2 a and 6 . Thus this TAC was given partly within and partly outside the stock boundary. This area is now in UK EEZ and there was no agreement for quota to Norway in this area for 2021. Norway and UK now have agreement on 600 t quota to Norway in area 2a, 4, 5b, 6 in 2022, with only longline fisheries allowed in area 6 . There is no ICES separate advice for the fishery in this area.

The TAC sat by EU for 2020 applied to "Union waters of 2a and 4; Union and international waters of 5 b and 6 " were allocated to Norway with the footnote "To be taken in Union waters of 2a and 6. In 6, this quantity may only be fished with longlines (GHL/*2A6-C)." Additionally EU had sat another TAC in "International waters of 1 and $2(\mathrm{GHL} / 1 / 2 \mathrm{INT})$ " and a minor quota in "Norwegian waters of 1 and 2 (GHL/1N2AB.)", both with the footnote "Exclusively for bycatches ${ }^{2}$.

EU has sat a TAC of 629 t for 2021 to be taken in Union waters of 2a and 6. In 6, this quantity may only be fished with longlines. EU has sat 1800 t TAC in international waters of ICES 1 and 2, exclusively for bycatches. No directed fisheries are permitted under this ${ }^{3}$.

EU has sat a TAC of 2571 t for 2022 in area6; United Kingdom and Union waters of 4; United Kingdom waters of 2 a and United Kingdom and international waters of 5 b (GHL/2A-C46) ${ }^{3}$.

Further information on regulations is found in the Stock Annex.

[^12]
### 8.1.4 Expected landings in 2022

Catches in 2021 were 28431 t , and exceeded the TAC sat by JRNFC. The total Greenland halibut landings in the Barents Sea and adjacent waters (ICES Subarea 1 and divisions 2a and 2b) in 2022 may thus be higher than the JRNFC TAC of 25000 t . Discards at present are not regarded as a problem.

### 8.2 Status of research

### 8.2.1 Survey results (Tables 8.9-8.13, Figures 8.4-8.14)

Survey indices from the Russian autumn survey (Figures 8.4-8.6), the Norwegian slope survey (Figures 8.4-8.5 and 8.7-8.8), the Joint Norwegian-Russian Ecosystem survey (A5216), Eco-juv and Eco-south indices; Figures 8.9-8.10) and the Joint Norwegian-Russian Winter Survey (Figure 8.11) are given. Length distributions from these surveys are presented in Tables 8.9-8.12 and Figure 8.12. Results from Spanish surveys are presented in Table 8.13 and Figure 8.13. Results from a Polish spring survey is presented in Figure 8.14.

The Russian bottom-trawl surveys in October-December (ICES acronym: RU-BTr-Q4) are important since they usually cover large parts of the total known distribution area of the Greenland halibut within 100-900 m depth. However, it has been considered imprudent to use 2002, 2003 and 2013 data from this survey series. During the 2002 survey, no observations were available from the Exclusive Economic Zone of Norway (NEEZ). In 2003, observations on the main spawning grounds were conducted three weeks later than usual because access to NEEZ was obtained too late. The number of trawl stations was also insufficient due to the same reason. Due to technical problems indices in 2013 were not obtained. Technical and practical changes were made in 2003. In 2017 and 2019 the coverage was insufficient. The 2020 estimate was not considered appropriate to use due to gear-related problems during the survey. A working document with a revision of the Russian index was provided to the 2021 meeting (Russkikh WD12). Revised and recalculated length distributions were not implemented in the 2021 assessment but will be subject to the upcoming benchmark. Length distributions by year for this survey are given in Table 8.9. The biomass indices for this survey increased steeply from 2005 to 2011, but have mainly showed a downward trend since then (Figures 8.4 and 8.5).

Total biomass indices from the Norwegian autumn slope survey ( $\mathrm{NO}-\mathrm{GH}-\mathrm{Btr}$-Q3) showed an upward trend in biomass estimates between 1994 and 2003, then a downward trend until 2008 until it increased again in 2009 but levelled out again in 2011, 2013, and 2015 (Figures 8.4-8.5, and 8.7-8.8). Since then, there has been a downward trend until 2020 when the index was at its lowest since the start of the survey. In 2021 there was an increase in the index but it is still among the lowest estimates in the time series. The length distributions from this survey (Figure 8.12, Tables 8.10 and 8.11 ) show modes that can be followed through the years and indicate new recruitment to the adult stock in 2007. Since then, no such large recruit events are apparent in the length distributions, and since 2009 abundance of fish in adult lengths has been declining s as well. This survey was conducted every year during 1994-2009 but is now run biennially.

The Joint Ecosystem Survey in autumn (A5216; Eco-NoRu-Q3 (Btr)) covers a large part of the Barents Sea down to 500 m and concerning Greenland halibut it can be regarded to be in the areas where mainly juveniles and immature fish are found. Two indices for Greenland halibut are based on the Joint Ecosystem Survey in the Barents Sea and previous juvenile survey, one for juvenile areas (Figure 8.9) denoted Eco-juv index in the northernmost survey area, and another denoted Eco-south index defined by the survey area south from $76.5^{\circ} \mathrm{N}$ and west of Spitsbergen (Figure 8.10). The juvenile index, covering the juvenile area (see section 8.3), indicates a highly
variable recruitment success with years between good year classes. The trend has mainly been downward since around 2007 and the 2015 estimates are the lowest registered so far, followed by a minor peak in 2017. The juvenile index has increased the last two years and is now around average for the time series. The Eco-south index for both females and males showed an increasing trend until 2012, followed by a decrease since then. The 2018 estimate in the Eco-south index was excluded from the 2021-assessment. The abundance estimate in 2018 peaked to extend that can be considered unrealistic for a slow-growing species. Additionally, there are concerns about the quality of the estimate due to the lack of survey coverage in 2018, especially in the area south of $76.5^{\circ} \mathrm{N}$ as defined for the Eco-south index. The male index shows a similar trend except the increase started a year later, in 2016-2018, but is also down in 2019. The general downward trend continues in 2021. Length distributions by year for this survey are given in Table 8.11.

The joint winter survey in the Barents Sea (Eco-NoRu-Q3 (Btr)) has been run from 1986 to the present (jointly with Russia since 2000, except 2006 and 2007). The survey mainly covers depths of $100-500 \mathrm{~m}$ and does not cover the deeper slope areas. Spatially, the survey focuses on the central Barents Sea, and west of Svalbard for some years. The northward coverage is limited by sea ice in some years. It is conducted in February and can thus give information on the stock at a different time of the year, as the other surveys are run in autumn. The biomass index has shown an increasing trend since 2004 with large variations in recent years. This survey is not currently used in the assessment.

The Spanish bottom-trawl survey, (Table 8.13, Figure 8.13) was carried out on a new hired commercial vessel and some changes have been done in the initial standard protocol. The indices for Greenland halibut from earlier Spanish surveys (1997-2005) cannot be standardized with more recent ones (2008 to present, Basterretxea et al., WD13 2013). This means that biomass estimates from the survey are only available for years 2008 and onwards. The Spanish survey has since 2015 only been run in autumn. This survey is not conducted every year. The biomass index from the Spanish survey shows a downward trend since around 2012. This survey is not currently used in the assessment.

Polish bottom-trawl surveys on Greenland halibut were carried out in the Svalbard-Bear Island area (ICES 2b) in October 2006, April 2007, April 2008, June 2009, and March 2011. The main objectives of the survey were to determine the biological structure, distribution, density and standing biomass of Greenland halibut in the survey area (Trella and Janusz, WD6 ICES AFWG 2012). The survey has not been conducted since then. Polish survey index is shown in Figure 8.14, no new data were presented to the meeting. This survey is not currently used in the assessment.

### 8.2.2 Commercial catch-per-unit-effort (Table 8.6, Figure 8.15)

The CPUE series for the stock was subject to the last benchmark and following data workshops (see reports from WKBUT 2013, DCWKNGHD 2014 and IBPHALI 2015, and working documents by Bakanev (WD14 WKBUT 2013) and Nedreaas (WD 2 DCWKNGHD 2014); Figure 8.15). An alternative CPUE series for the Russian fisheries for the years 2004-2015 was presented at the 2016 meeting (Mikhaylov, WD14 ICES AFWG 2016). It shows some discrepancies compared to the previous CPUE series used for the Russian fisheries for the same years. See the Stock Annex for further comments. The CPUE series are not currently used in the assessment.

### 8.2.3 Age readings

Based on the scientific understanding that the species is slower growing and more vulnerable than the previous age readings suggest, the Norwegian age reading methods were changed in
2006. The new Norwegian age readings are not comparable with older data or the Russian age readings.

The report from Workshop on Age Reading of Greenland Halibut (WKARGH) 14-17 February 2011 (ICES CM 2011/ACOM:41) described and evaluated several age reading methods for Greenland halibut.

The different methods can be classified into two groups: A) Those that produce age-length relationships that broadly compare with the traditional methods described by the joint NAFO-ICES workshop in 1996 (ICES CM 1997/G:1); and B) Several recently developed techniques that show much higher longevity and approximately half the growth rate from $40-50 \mathrm{~cm}$ onwards compared to the traditional method.

A second workshop on age reading of Greenland halibut (WKARGH 2) was conducted in August 2016 and worked on further validation on new age reading methods. The workshop recommended that two of the new methods can be used to provide age estimations for stock assessments. Further, recognizing some bias and low precision in methods, the WKARGH2 suggested that an aging error matrix or growth curve with error be provided for use in future stock assessments (WKARGH2 report 2016, ICES CM 2016/SSGIEOM:16).

WKARGH2 recommends regular inter-lab calibration exercises to improve precision (i.e. exchange of digital images between readers for each method and between methods).

AFWG suggests that Russian and Norwegian scientists and age readers meet to work out issues of disagreements on Greenland halibut aging.

### 8.3 Data used in the assessment

In the assessment, the catch data are split into four aggregated fleets by gear and countries. Longline/gillnet fleets include landings from gillnet, longline, and handline. Trawl fleets include landings from bottom trawl, purse-seine (very minor catches, can be bycatch or misreporting) and Danish seine. Catch in tonnes and length distributions per quarter per fleet per sex from 19922020 are used in the assessment. Fleets are split between Norwegian (including $3{ }^{\text {rd }}$ countries) and Russian catches, and selectivities are allowed to vary by sex (logistic for gill fleets, asymmetric dome-shaped for trawl fleets), to account for sexual dimorphism influencing vulnerability to fishing. For each fleet listed below, length distributions and reported catch in tonnes are split by quarter and sex (although length distributions are not available for all quarters for some fleets).

- Russian, trawl and minor gears (split by sex)
- Russian, gillnet and longline (split by sex)
- $\quad$ Norwegian and $3^{\text {rd }}$ countries, trawl and minor gears (split by sex)
- $\quad$ Norwegian and $3^{\text {rd }}$ countries, gillnet and longline (split by sex)

In addition, the model has four surveys, all modelled with asymmetric dome-shaped selectivities (note that in a model context "selectivity" encompasses all aspects of vulnerability to the fishery, including gear effects, vessel effects, area effects etc.). In each case, data are used as length distribution and biomass index. The biomass index was not available to split by sex for all years, so a combined sex index is used. The four survey indices that go into the current assessment are:

- Norway slope (NO-GH-Btr-Q3)- based on the Norwegian Greenland halibut slope survey (yearly 1996-2009, biennially since then). Split by sex.
- EcoJuv - a juvenile index based on data from the northern/northeastern areas of the Joint Ecosystem survey (A5216; Eco-NoRu-Q3 (Btr); 2003-present) and the precursory Norwegian juvenile Greenland halibut survey north and east of Svalbard (1996-2002; Hallfredsson and Vollen, WD 1 ICES IBPhali 2015). Split by sex.
- EcoSouth - an index for the Barents Sea south of $76.5^{\circ} \mathrm{N}$, based on data from the Joint Ecosystem survey (A5216; Eco-NoRu-Q3 (Btr); 2003-present; Hallfredsson and Vollen, ICES AFWG, WD 20, April 2015). Split by sex.
- Russian - Russian bottom-trawl survey in the Barents Sea (1992-2015 and 2017; RU-BTrQ4). Sex aggregated (can be split by sex in future work).

No age data or CPUE indices are used in the tuning.

### 8.4 Methods used in the assessment

A new assessment method with a length-based GADGET model was benchmarked in 2015 (IPHALI 2015) and accepted by ACOM the same year. The model is further described in the IPHALI report and the Stock Annex. Advice for the stock is given biennially and last advice applies for 2022 and 2023. Next advice year is 2023 for the years 2024 and 2025. Thus, no analytical assessment was run this year. For description of last assessment see ICES AFWG 2021 report.

### 8.4.1 Model settings

For last assessment see ICES AFWG 2021 report.

### 8.4.1.1 Estimated parameters:

For last assessment see ICES AFWG 2021 report.

### 8.5 Results of the assessment

For last assessment see ICES AFWG 2021 report.

### 8.5.1 Biological reference points

For last assessment see ICES AFWG 2021 report

### 8.6 Comments to the assessment

For last assessment see ICES AFWG 2021 report.

### 8.6.1 Future work

Further development of the assessment is needed, in consistency with conclusions of the IBPHALI benchmark and report of the external benchmark reviewer.

[^13]
### 8.7 Tables and figures

Table 8.1. Greenland halibut in subareas 1 and 2. Nominal Catch ( $\mathbf{t}$ ) by countries (Subarea 1, divisions 2a, and 2b combined) as officially reported to ICES.

| 㐫 |  |  |  |  |  |  |  |  | $\sum_{\substack{0\\}}$ |  | $\begin{aligned} & \text { त } \\ & \text { n } \\ & 0 \\ & 0 \end{aligned}$ |  |  | $\begin{aligned} & \stackrel{m}{n} \\ & \stackrel{\pi}{n} \\ & \underset{\sim}{z} \end{aligned}$ | $\begin{aligned} & \stackrel{\check{10}}{0} \\ & \stackrel{0}{n} \end{aligned}$ | © |  |  | $\begin{gathered} \bar{\circ} \\ \stackrel{\circ}{\circ} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0 | 0 | 0 | 138 | 2165 | 0 | 0 | 0 | 0 | 0 | 4376 | 0 | 0 | 15181 | 0 | 0 | 23 | 0 | 21883 |
| 1985 | 0 | 0 | 0 | 239 | 4000 | 0 | 0 | 0 | 0 | 0 | 5464 | 0 | 0 | 10237 | 0 | 0 | 5 | 0 | 19945 |
| 1986 | 0 | 0 | 42 | 13 | 2718 | 0 | 0 | 0 | 0 | 0 | 7890 | 0 | 0 | 12200 | 0 | 0 | 10 | 2 | 22875 |
| 1987 | 0 | 0 | 0 | 13 | 2024 | 0 | 0 | 0 | 0 | 0 | 7261 | 0 | 0 | 9733 | 0 | 0 | 61 | 20 | 19112 |
| 1988 | 0 | 0 | 186 | 67 | 744 | 0 | 0 | 0 | 0 | 0 | 9076 | 0 | 0 | 9430 | 0 | 0 | 82 | 2 | 19587 |
| 1989 | 0 | 0 | 67 | 31 | 600 | 0 | 0 | 0 | 0 | 0 | 10622 | 0 | 0 | 8812 | 0 | 0 | 6 | 0 | 20138 |
| 1990 | 0 | 0 | 163 | 49 | 954 | 0 | 0 | 0 | 0 | 0 | 17243 | 0 | 0 | 4764 | 0 | 0 | 10 | 0 | 23183 |
| 1991 | 11 | 2564 | 314 | 119 | 101 | 0 | 0 | 0 | 0 | 0 | 27587 | 0 | 0 | 2490 | 132 | 0 | 0 | 2 | 33320 |
| 1992 | 0 | 0 | 16 | 111 | 13 | 13 | 0 | 0 | 0 | 0 | 7667 | 0 | 31 | 718 | 23 | 0 | 10 | 0 | 8602 |
| 1993 | 2 | 0 | 61 | 80 | 22 | 8 | 56 | 0 | 0 | 30 | 10380 | 0 | 43 | 1235 | 0 | 0 | 16 | 0 | 11933 |
| 1994 | 4 | 0 | 18 | 55 | 296 | 3 | 15 | 5 | 0 | 4 | 8428 | 0 | 36 | 283 | 1 | 0 | 76 | 2 | 9226 |
| 1995 | 0 | 0 | 12 | 174 | 35 | 12 | 25 | 2 | 0 | 0 | 9368 | 0 | 84 | 794 | 1106 | 0 | 115 | 7 | 11734 |
| 1996 | 0 | 0 | 2 | 219 | 81 | 123 | 70 | 0 | 0 | 0 | 11623 | 0 | 79 | 1576 | 200 | 0 | 317 | 57 | 14347 |


|  |  | $\begin{aligned} & \stackrel{\widetilde{0}}{\bar{L}} \\ & \stackrel{4}{山} \\ & \hline \end{aligned}$ |  | $\begin{aligned} & \text { シ } \\ & \text { 튼 } \\ & \text { Non } \end{aligned}$ |  |  |  | $\begin{aligned} & \text { ס} \\ & \text { 들 } \\ & \underline{\underline{N}} \end{aligned}$ | ${\underset{\sim}{\top}}_{\substack{0}}$ |  | $\begin{aligned} & \text { त } \\ & \text { 3 } \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \frac{\pi}{0} \end{aligned}$ | ⿹ㅣ 0 0 능 | $\begin{aligned} & \stackrel{m}{n} \\ & \stackrel{\pi}{n} \\ & \underset{\sim}{\boldsymbol{n}} \end{aligned}$ | $\begin{aligned} & \frac{\pi}{0} \\ & \stackrel{0}{n} \end{aligned}$ | © |  |  | $\begin{gathered} \overline{0} \\ \stackrel{0}{0} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 0 | 0 | 27 | 253 | 56 | 0 | 62 | 2 | 0 | 0 | 7661 | 12 | 50 | 1038 | 157 | 0 | 67 | 25 | 9410 |
| 1998 | 0 | 0 | 57 | 67 | 34 | 0 | 23 | 2 | 0 | 0 | 8435 | 31 | 99 | 2659 | 259 | 0 | 182 | 45 | 11893 |
| 1999 | 0 | 0 | 94 | 0 | 34 | 38 | 7 | 2 | 0 | 0 | 15004 | 8 | 49 | 3823 | 319 | 0 | 94 | 45 | 19517 |
| 2000 | 0 | 0 | 0 | 45 | 15 | 0 | 16 | 1 | 0 | 0 | 9083 | 3 | 37 | 4568 | 375 | 0 | 111 | 43 | 14297 |
| 2001 | 0 | 0 | 0 | 122 | 58 | 0 | 9 | 1 | 0 | 0 | 10896 | 2 | 35 | 4694 | 418 | 0 | 100 | 30 | 16365 |
| 2002 | 0 | 219 | 0 | 7 | 42 | 22 | 4 | 6 | 0 | 0 | 7143 | 5 | 14 | 5584 | 178 | 0 | 41 | 28 | 13293 |
| 2003 | 0 | 0 | 459 | 2 | 18 | 14 | 0 | 1 | 0 | 0 | 8216 | 5 | 19 | 4384 | 230 | 0 | 41 | 58 | 13447 |
| 2004 | 0 | 0 | 0 | 0 | 9 | 0 | 9 | 0 | 0 | 0 | 13939 | 1 | 50 | 4662 | 186 | 0 | 43 | 0 | 18899 |
| 2005 | 0 | 170 | 0 | 32 | 8 | 0 | 0 | 0 | 0 | 0 | 13011 | 0 | 23 | 4883 | 660 | 0 | 29 | 18 | 18834 |
| 2006 | 0 | 0 | 204 | 46 | 8 | 0 | 8 | 0 | 0 | 196 | 11119 | 201 | 26 | 6055 | 29 | 0 | 10 | 2 | 17904 |
| 2007 | 0 | 0 | 203 | 41 | 8 | 198 | 15 | 0 | 0 | 0 | 8230 | 200 | 47 | 6484 | 8 | 0 | 11 | 8 | 15453 |
| 2008 | 0 | 0 | 663 | 42 | 5 | 0 | 28 | 0 | 0 | 0 | 7393 | 201 | 46 | 5294 | 94 | 0 | 16 | 10 | 13792 |
| 2009 | 0 | 0 | 422 | 16 | 19 | 16 | 15 | 2 | 0 | 0 | 8446 | 204 | 237 | 3335 | 210 | 0 | 9 | 60 | 12990 |
| 2010 | 0 | 0 | 272 | 102 | 14 | 15 | 16 | 0 | 0 | 0 | 7700 | 3 | 11 | 6888 | 182 | 0 | 4 | 22 | 15229 |
| 2011 | 0 | 0 | 538 | 46 | 80 | 4 | 7 | 0 | 0 | 234 | 8270 | 169 | 21 | 7053 | 144 | 0 | 36 | 4 | 16606 |
| 2012 | 0 | 0 | 564 | 40 | 40 | 12 | 13 | 0 | 0 | 0 | 9331 | 22 | 1 | 10041 | 190 | 0 | 21 | 14 | 20288 |



* Provisional figures.


## Table 8.2. Greenland halibut in subareas 1 and 2. Nominal catch ( $\mathbf{t}$ ) by countries in Subarea 1 as officially reported to ICES.

| 㐫 |  |  |  |  |  |  | $\begin{aligned} & \text { ס } \\ & \underline{\Pi} \\ & \underline{\underline{0}} \end{aligned}$ | $\underset{\substack{0 \\ \underset{\sim}{0} \\ \hline}}{ }$ |  | $\begin{aligned} & \text { त } \\ & \text { 3 } \\ & \text { 30 } \end{aligned}$ |  | $\overline{5}$ 0 0 0 0 | $\begin{aligned} & \stackrel{\pi}{\omega} \\ & \stackrel{\pi}{\omega} \\ & \underset{\sim}{c} \end{aligned}$ | $\begin{aligned} & . \underline{0} \\ & \text { ion } \\ & \text { in } \end{aligned}$ | © |  |  | $\begin{aligned} & \overline{\mathrm{O}} \\ & \hline \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 593 | 0 | 0 | 81 | 0 | 0 | 17 | 0 | 691 |
| 1985 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 602 | 0 | 0 | 122 | 0 | 0 | 1 | 0 | 725 |
| 1986 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 557 | 0 | 0 | 615 | 0 | 0 | 5 | 1 | 1179 |
| 1987 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 984 | 0 | 0 | 259 | 0 | 0 | 10 | 0 | 1255 |
| 1988 | 0 | 9 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 978 | 0 | 0 | 420 | 0 | 0 | 7 | 0 | 1418 |
| 1989 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2039 | 0 | 0 | 482 | 0 | 0 | 0 | 0 | 2521 |
| 1990 | 0 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1304 | 0 | 0 | 321 | 0 | 0 | 0 | 0 | 1632 |
| 1991 | 164 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2029 | 0 | 0 | 522 | 0 | 0 | 0 | 0 | 2715 |
| 1992 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2349 | 0 | 0 | 467 | 0 | 0 | 0 | 0 | 2816 |
| 1993 | 0 | 32 | 0 | 0 | 0 | 56 | 0 | 0 | 0 | 1754 | 0 | 0 | 867 | 0 | 0 | 0 | 0 | 2709 |
| 1994 | 0 | 17 | 217 | 0 | 0 | 15 | 0 | 0 | 0 | 1165 | 0 | 0 | 175 | 0 | 0 | 0 | 0 | 1589 |
| 1995 | 0 | 12 | 0 | 0 | 0 | 25 | 0 | 0 | 0 | 1352 | 0 | 0 | 270 | 84 | 0 | 0 | 0 | 1743 |
| 1996 | 0 | 2 | 0 | 0 | 0 | 70 | 0 | 0 | 0 | 911 | 0 | 0 | 198 | 0 | 0 | 0 | 0 | 1181 |
| 1997 | 0 | 15 | 0 | 0 | 0 | 62 | 0 | 0 | 0 | 610 | 0 | 0 | 170 | 0 | 0 | 0 | 0 | 857 |
| 1998 | 0 | 47 | 0 | 0 | 0 | 23 | 0 | 0 | 0 | 859 | 0 | 0 | 491 | 0 | 0 | 2 | 0 | 1422 |


| $\begin{aligned} & \text { む } \\ & \text { む } \end{aligned}$ |  |  |  |  |  |  |  |  |  | $\begin{aligned} & \text { 㐅} \\ & \text { 3 } \\ & 0 \\ & \mathbf{z} \end{aligned}$ | $\begin{aligned} & \text { ㄷ } \\ & \frac{C}{10} \\ & \text { Q } \end{aligned}$ | $\overline{5}$ 00 00 0 0 | $\begin{aligned} & \underset{\sim}{n} \\ & \underset{\sim}{\tilde{x}} \\ & \end{aligned}$ | $\begin{aligned} & \text {.듣 } \\ & \text { io } \end{aligned}$ | O\％ |  |  | $\stackrel{\bar{\oplus}}{\stackrel{\rightharpoonup}{0}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 0 | 91 | 0 | 0 | 13 | 7 | 0 | 0 | 0 | 1101 | 0 | 0 | 1203 | 0 | 0 | 0 | 0 | 2415 |
| 2000 | 0 | 0 | 0 | 0 | 0 | 16 | 0 | 0 | 0 | 1021 | 0 | 0 | 1169 | 0 | 0 | 0 | 0 | 2206 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 9 | 0 | 0 | 0 | 925 | 0 | 0 | 951 | 0 | 0 | 2 | 0 | 1887 |
| 2002 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 834 | 0 | 0 | 1167 | 0 | 0 | 0 | 0 | 2004 |
| 2003 | 0 | 48 | 0 | 0 | 2 | 0 | 1 | 0 | 0 | 962 | 1 | 0 | 735 | 0 | 0 | 0.3 | 0 | 1749 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0.3 | 0 | 0 | 0 | 866 | 0 | 0 | 633 | 0 | 0 | 3 | 0 | 1503 |
| 2005 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 572 | 0 | 0 | 595 | 0 | 0 | 3 | 0 | 1171 |
| 2006 | 0 | 17 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 575 | 0 | 0 | 626 | 2 | 0 | 2 | 0 | 1224 |
| 2007 | 0 | 18 | 0 | 1 | 198 | 3 | 0 | 0 | 0 | 514 | 0 | 3 | 438 | 0 | 0 | 4 | 0 | 1179 |
| 2008 | 0 | 13 | 0 | 1 | 0 | 5 | 0 | 0 | 0 | 599 | 0 | 0 | 390 | 0 | 0 | 0 | 0 | 1008 |
| 2009 | 0 | 33 | 0 | 0 | 16 | 5 | 0 | 0 | 0 | 734 | 0 | 0 | 483 | 0 | 0 | 1 | 0 | 1272 |
| 2010 | 0 | 15 | 0 | 0 | 0 | 16 | 0 | 0 | 0 | 659 | 0 | 0 | 708 | 2 | 0 | 0 | 0 | 1399 |
| 2011 | 0 | 63 | 0 | 0 | 0 | 6 | 0 | 0 | 0 | 867 | 0 | 0 | 782 | 0 | 0 | 0 | 0 | 1718 |
| 2012 | 0 | 8 | 5 | 0 | 0 | 7 | 0 | 0 | 0 | 921 | 0 | 0 | 1368 | 1 | 0 | 7 | 0 | 2318 |
| 2013 | 0 | 39 | 1 | 8 | 0 | 100 | 0 | 0 | 0 | 1055 | 4 | 0 | 1442 | 4 | 0 | 8 | 0 | 2661 |
| 2014 | 0 | 143 | 8 | 11 | 19 | 38 | 0 | 0 | 0 | 1271 | 7 | 0 | 1261 | 10 | 0 | 14 | 0 | 2782 |


|  |  |  |  | $\begin{aligned} & \text { U } \\ & \text { 든 } \\ & \text { ( } \end{aligned}$ |  |  |  | $\underset{\text { N }}{\substack{0 \\ J}}$ |  | $\begin{aligned} & \text { 㐅} \\ & \text { 3 } \\ & \text { z} \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \frac{\pi}{0} \\ & \hline \end{aligned}$ | $\overline{0}$ 0 0 0 0 | $\begin{aligned} & \text { N} \\ & \underset{\sim}{u} \\ & \underset{\sim}{z} \end{aligned}$ | $\begin{aligned} & \text { •ㅡㅡㅁ } \\ & \text { in } \end{aligned}$ | O |  |  | $\stackrel{\bar{\oplus}}{\stackrel{\rightharpoonup}{\circ}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 0 | 96 | 14 | 3 | 12 | 47 | 0 | 0 | 5 | 1424 | 5 | 0 | 1681 | 8 | 0 | 4 | 0 | 3299 |
| 2016 | 353 | 84 | 2 | 3 | 3 | 38 | 0 | 0 | 0 | 1265 | 7 | 0 | 1172 | 7 | 0 | 20 | 0 | 2954 |
| 2017 | 519 | 125 | 4 | 4 | 2 | 8 | 0 | 1 | 72 | 1389 | 9 | 1 | 1124 | 13 | 0 | 21 | 0 | 3293 |
| 2018 | 574 | 104 | 9 | 16 | 2 | 20 | 0 | 0 | 199 | 1008 | 4 | 1 | 894 | 2 | 97 | 0 | 0 | 2930 |
| 2019 | 587 | 116 | 27 | 9 | 5 | 5 | 0 | 32 | 347 | 939 | 119 | 0 | 932 | 15 | 49 | 0 | 0 | 3182 |
| 2020 | 578 | 123 | 37 | 7 | 11 | 18 | 0 | 142 | 223 | 1388 | 96 | 17 | 787 | 36 | 1 | 0 | 0 | 3464 |
| 2021* | 382 | 207 | 17 | 1 | 10 | 35 | 0 | 96 | 159 | 1617 | 9 | 14 | 713 | 14 | 11 | 0 | 0 | 3285 |

* Provisional figures.


## Table 8.3. Greenland halibut in subareas 1 and 2. Nominal catch ( $t$ ) by countries in Division $2 a$ as officially reported to ICES.

|  |  |  |  |  |  |  |  |  | $\begin{aligned} & \text { 㐅} \\ & \text { n } \\ & 0 \\ & \text { Z } \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \frac{\pi}{0} \end{aligned}$ | $\overline{0}$ 0 0 0 | $\begin{aligned} & \stackrel{n}{n} \\ & \stackrel{\pi}{n} \\ & \underset{\sim}{c} \end{aligned}$ | $\begin{aligned} & \stackrel{\varrho}{\overline{0}} \\ & \text { in } \end{aligned}$ | O |  |  | $\begin{aligned} & \overline{\mathrm{O}} \\ & \hline \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0 | 0 | 265 | 138 | 0 | 0 | 0 | 0 | 3703 | 0 | 0 | 5459 | 0 | 0 | 1 | 0 | 9566 |
| 1985 | 0 | 0 | 254 | 239 | 0 | 0 | 0 | 0 | 4791 | 0 | 0 | 6894 | 0 | 0 | 2 | 0 | 12180 |
| 1986 | 0 | 6 | 97 | 13 | 0 | 0 | 0 | 0 | 6389 | 0 | 0 | 5553 | 0 | 0 | 5 | 1 | 12064 |
| 1987 | 0 | 0 | 75 | 13 | 0 | 0 | 0 | 0 | 5705 | 0 | 0 | 4739 | 0 | 0 | 44 | 10 | 10586 |




|  | $\begin{aligned} & \text { 苊 } \\ & \stackrel{0}{\overleftarrow{0}} \\ & \text { H } \end{aligned}$ |  |  | ㅡ․ 든 ( |  |  |  |  | $\begin{aligned} & \text { 㐅} \\ & \text { n } \\ & 0 \\ & 2 \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \frac{\pi}{0} \\ & 0 \end{aligned}$ | $\begin{aligned} & \overline{0} \\ & 0 \\ & \text { t } \\ & 0 \\ & 0 \end{aligned}$ |  | $\begin{aligned} & \stackrel{\unrhd}{0} \\ & \text { in } \end{aligned}$ | © |  |  | $\begin{aligned} & \overline{ \pm 0} \\ & \stackrel{0}{\circ} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 0 | 0 | 3 | 31 | 0 | 0 | 0 | 0 | 11216 | 0 | 11 | 1406 | 0 | 0 | 5 | 18 | 12690 |
| 2006 | 0 | 175 | 0 | 38 | 0 | 0 | 7 | 0 | 8897 | 0 | 6 | 950 | 0 | 0 | 6 | 2 | 10081 |
| 2007 | 0 | 162 | 2 | 37 | 0 | 0 | 12 | 0 | 6761 | 0 | 2 | 489 | 1 | 0 | 2 | 8 | 7475 |
| 2008 | 0 | 646 | 4 | 38 | 0 | 0 | 23 | 0 | 5566 | 1 | 1 | 1170 | 0 | 0 | 6 | 10 | 7465 |
| 2009 | 0 | 379 | 0 | 13 | 0 | 0 | 10 | 0 | 6456 | 0 | 9 | 1531 | 0 | 0 | 0 | 60 | 8459 |
| 2010 | 0 | 255 | 0 | 102 | 15 | 0 | 0 | 0 | 6426 | 0 | 0 | 4757 | 0 | 0 | 0 | 22 | 11577 |
| 2011 | 0 | 467 | 0 | 45 | 4 | 0 | 1 | 0 | 6637 | 0 | 0 | 3643 | 2 | 0 | 0 | 4 | 10803 |
| 2012 | 0 | 553 | 0 | 37 | 12 | 0 | 6 | 0 | 7934 | 0 | 0 | 3878 | 0 | 0 | 0 | 14 | 12434 |
| 2013 | 0 | 739 | 0 | 150 | 22 | 0 | 6 | 0 | 8215 | 0 | 2 | 4143 | 0 | 0 | 0 | 75 | 13352 |
| 2014 | 0 | 741 | 0 | 255 | 1 | 0 | 48 | 0 | 8640 | 0 | 0 | 4800 | 0 | 0 | 0 | 184 | 14669 |
| 2015 | 0 | 215 | 2 | 221 | 2 | 0 | 6 | 0 | 8166 | 0 | 1 | 3691 | 0 | 0 | 0 | 79 | 12383 |
| 2016 | 6 | 380 | 6 | 216 | 14 | 0 | 41 | 0 | 10073 | 0 | 6 | 1797 | 7 | 0 | 0 | 19 | 12566 |
| 2017 | 0 | 773 | 0 | 161 | 20 | 0 | 2 | 0 | 10122 | 0 | 7 | 1852 | 1 | 0 | 16 | 0 | 12955 |
| 2018 | 0 | 297 | 1 | 104 | 9 | 0 | 4 | 1 | 11226 | 2 | 5 | 695 | 0 | 6 | 0 | 0 | 12350 |
| 2019 | 0 | 232 | 15 | 94 | 16 | 0 | 3 | 0 | 12122 | 3 | 7 | 2754 | 3 | 11 | 0 | 0 | 15260 |
| 2020 | 0 | 385 | 21 | 34 | 28 | 0 | 1 | 0 | 11437 | 0 | 8 | 2691 | 0 | 3 | 0 | 0 | 14608 |
| 2021* | 0 | 529 | 19 | 123 | 4 | 0 | 5 | 0 | 9647 | 0 | 5 | 842 | 5 | 108 | 0 | 0 | 11287 |

## Provisional figures.

Table 8.4. Greenland halibut in subareas 1 and 2 . Nominal catch ( $\mathbf{t}$ ) by countries in Division 2 b as officially reported to ICES.

|  |  |  |  |  | $\begin{aligned} & \text { ․ } \\ & \text { 듄 } \end{aligned}$ |  | $\begin{aligned} & \text { ס } \\ & \underline{\bar{C}} \\ & \underline{\underline{N}} \end{aligned}$ | $\sum_{\substack{\pi \\ \hline}}$ |  | $\begin{aligned} & \text { 㐅} \\ & \text { 3 } \\ & \text { 30 } \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \text { त } \\ & \hline \mathbf{0} \end{aligned}$ |  | $\begin{aligned} & \stackrel{\pi}{\omega} \\ & \stackrel{\pi}{\omega} \\ & \underset{\sim}{\underset{\sim}{2}} \end{aligned}$ | $\begin{aligned} & \text {.드̃ } \\ & \text { in } \end{aligned}$ | O |  |  | $\begin{aligned} & \overline{\boxed{0}} \\ & \hline- \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0 | 0 | 0 | 1900 | 0 | 0 | 0 | 0 | 0 | 80 | 0 | 0 | 9641 | 0 | 0 | 5 | 0 | 11626 |
| 1985 | 0 | 0 | 0 | 3746 | 0 | 0 | 0 | 0 | 0 | 71 | 0 | 0 | 3221 | 0 | 0 | 2 | 0 | 7040 |
| 1986 | 0 | 0 | 36 | 2620 | 0 | 0 | 0 | 0 | 0 | 944 | 0 | 0 | 6032 | 0 | 0 | 0 | 0 | 9632 |
| 1987 | 0 | 0 | 0 | 1947 | 0 | 0 | 0 | 0 | 0 | 572 | 0 | 0 | 4735 | 0 | 0 | 7 | 10 | 7271 |
| 1988 | 0 | 0 | 0 | 590 | 0 | 0 | 0 | 0 | 0 | 239 | 0 | 0 | 5008 | 0 | 0 | 19 | 0 | 5856 |
| 1989 | 0 | 0 | 0 | 496 | 0 | 0 | 0 | 0 | 0 | 533 | 0 | 0 | 3366 | 0 | 0 | 0 | 0 | 4395 |
| 1990 | 0 | 0 | 23 | 942 | 0 | 0 | 0 | 0 | 0 | 7706 | 0 | 0 | 3197 | 0 | 0 | 9 | 0 | 11877 |
| 1991 | 11 | 1000 | 0 | 80 | 0 | 0 | 0 | 0 | 0 | 14369 | 0 | 0 | 1663 | 132 | 0 | 0 | 1 | 17256 |
| 1992 | 0 | 0 | 0 | 12 | 3 | 0 | 0 | 0 | 0 | 1732 | 0 | 16 | 193 | 23 | 0 | 9 | 0 | 1988 |
| 1993 | 2 | 0 | 0 | 8 | 2 | 0 | 0 | 0 | 30 | 649 | 0 | 26 | 158 | 0 | 0 | 14 | 0 | 889 |
| 1994 | 4 | 0 | 1 | 46 | 8 | 0 | 1 | 0 | 4 | 881 | 0 | 10 | 41 | 1 | 0 | 62 | 2 | 1061 |
| 1995 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 1662 | 0 | 24 | 297 | 1022 | 0 | 32 | 5 | 3047 |
| 1996 | 0 | 0 | 0 | 47 | 0 | 0 | 0 | 0 | 0 | 1204 | 0 | 24 | 912 | 196 | 0 | 39 | 0 | 2422 |
| 1997 | 0 | 0 | 12 | 33 | 0 | 0 | 2 | 0 | 0 | 1349 | 12 | 9 | 534 | 156 | 0 | 46 | 0 | 2153 |
| 1998 | 0 | 0 | 10 | 18 | 0 | 0 | 1 | 0 | 0 | 915 | 31 | 19 | 1638 | 254 | 0 | 106 | 4 | 2996 |


|  |  |  |  |  | $\begin{aligned} & \text { シ } \\ & \text { 든 } \\ & \text { ( } \end{aligned}$ |  | $\begin{aligned} & \text { ס्ट } \\ & \text { त्0 } \\ & \underline{\underline{0}} \end{aligned}$ | － |  | $\begin{aligned} & \text { 㐅 } \\ & \text { そu } \\ & \text { z } \end{aligned}$ | $\begin{aligned} & \text { ס } \\ & \frac{\Gamma}{0} \\ & 0 \end{aligned}$ | ㅈ․ 0 능 0 | $\begin{aligned} & \underset{\sim}{\pi} \\ & \stackrel{\pi}{n} \\ & \underset{\sim}{2} \end{aligned}$ | $\begin{aligned} & \stackrel{.}{\overline{0}} \\ & \text { in } \end{aligned}$ | © |  |  | $\stackrel{\bar{\oplus}}{\stackrel{0}{\circ}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 0 | 0 | 3 | 14 | 0 | 0 | 0 | 0 | 0 | 839 | 8 | 16 | 1886 | 318 | 0 | 31 | 0 | 3115 |
| 2000 | 0 | 0 | 0 | 5 | 2 | 0 | 1 | 0 | 0 | 526 | 3 | 19 | 2709 | 374 | 0 | 46 | 0 | 3685 |
| 2001 | 0 | 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 1231 | 2 | 22 | 3017 | 413 | 0 | 42 | 0 | 4736 |
| 2002 | 0 | 219 | 0 | 30 | 0 | 0 | 6 | 0 | 0 | 432 | 5 | 11 | 3568 | 178 | 0 | 29 | 0 | 4478 |
| 2003 | 0 | 0 | 21 | 13 | 0 | 0 | 0 | 0 | 0 | 541 | 4 | 9 | 1887 | 216 | 0 | 35 | 0 | 2726 |
| 2004 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 1369 | 1 | 26 | 3219 | 182 | 0 | 39 | 0 | 4840 |
| 2005 | 0 | 170 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 1223 | 0 | 12 | 2882 | 660 | 0 | 21 | 0 | 4973 |
| 2006 | 0 | 0 | 12 | 7 | 8 | 0 | 0 | 0 | 196 | 1647 | 201 | 20 | 4479 | 27 | 0 | 2 | 0 | 6600 |
| 2007 | 0 | 0 | 23 | 6 | 3 | 0 | 0 | 0 | 0 | 955 | 200 | 45 | 5557 | 7 | 0 | 5 | 0 | 6801 |
| 2008 | 0 | 0 | 4 | 1 | 3 | 0 | 0 | 0 | 0 | 1228 | 200 | 45 | 3734 | 94 | 0 | 10 | 0 | 5319 |
| 2009 | 0 | 0 | 10 | 19 | 3 | 0 | 2 | 0 | 0 | 1256 | 204 | 228 | 1321 | 210 | 0 | 8 | 0 | 3260 |
| 2010 | 0 | 0 | 2 | 14 | 0 | 0 | 0 | 0 | 0 | 615 | 3 | 11 | 1423 | 180 | 0 | 4 | 0 | 2252 |
| 2011 | 0 | 0 | 8 | 80 | 1 | 0 | 0 | 0 | 234 | 766 | 169 | 21 | 2628 | 142 | 0 | 36 | 0 | 4085 |
| 2012 | 0 | 0 | 2 | 35 | 3 | 0 | 0 | 0 | 0 | 476 | 22 | 1 | 4795 | 189 | 0 | 14 | 0 | 5537 |
| 2013 | 0 | 0 | 5 | 48 | 10 | 0 | 1 | 0 | 0 | 1133 | 26 | 5 | 4725 | 192 | 0 | 9 | 0 | 6154 |
| 2014 | 0 | 0 | 3 | 25 | 3 | 0 | 0 | 0 | 0 | 1321 | 12 | 0 | 4000 | 196 | 0 | 14 | 0 | 5574 |
| 2015 | 0 | 0 | 1 | 17 | 3 | 0 | 0 | 0 | 0 | 1284 | 8 | 0 | 7581 | 151 | 0 | 21 | 0 | 9066 |


|  |  | $\begin{aligned} & . \frac{0}{L} \\ & \text { D} \\ & \text { W } \end{aligned}$ |  |  |  |  | $\begin{aligned} & \text { ס } \\ & \text { त्0 } \\ & \underline{\underline{N}} \end{aligned}$ | $\sum_{\substack{0 \\ \hline}}$ |  | $\begin{aligned} & \text { 㐅 } \\ & \text { 3 } \\ & 0 \\ & 2 \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \text { तo } \end{aligned}$ | $\begin{aligned} & \overline{0} \\ & 0 \\ & \text { 능 } \\ & \end{aligned}$ | $\begin{aligned} & \stackrel{\pi}{\omega} \\ & \stackrel{\pi}{\omega} \\ & \underset{\sim}{c} \end{aligned}$ | $\begin{aligned} & \stackrel{\unrhd}{\bar{n}} \\ & \text { in } \end{aligned}$ | O |  |  | $\begin{aligned} & \bar{\oplus} \\ & \stackrel{0}{0} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 | 2 | 0 | 19 | 1 | 10 | 0 | 0 | 0 | 0 | 1594 | 1 | 13 | 7608 | 183 | 0 | 0 | 0 | 9431 |
| 2017 | 0 | 4 | 19 | 17 | 12 | 3 | 0 | 0 | 0 | 2230 | 17 | 5 | 7737 | 42 | 0 | 46 | 0 | 10132 |
| 2018 | 2 | 0 | 1 | 40 | 30 | 9 | 0 | 6 | 0 | 2477 | 21 | 0 | 10483 | 58 | 31 | 0 | 0 | 13159 |
| 2019 | 0 | 0 | 2 | 2 | 0 | 01 | 0 | 0 | 0 | 1784 | 0 | 1 | 8512 | 68 | 14 | 0 | 0 | 10353 |
| 2020 | 1 | 0 | 6 | 15 | 8 | 2 | 0 | 6 | 3 | 1708 | 1 | 3 | 8788 | 60 | 40 | 0 | 0 | 10641 |
| 2021* | 1 | 0 | 18 | 50 | 13 | 0 | 0 | 0 | 0 | 2744 | 5 | 27 | 10839 | 105 | 57 | 0 | 0 | 13859 |

* Provisional figures.


## Table 8.5. Greenland halibut in subareas 1 and 2. Landings by gear (tonnes). Approximate figures, the total may differ slightly from Table 8.1.

| Year | Gillnet | Longline | Trawl | Danish seine | Other |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 1189 | 336 | 11759 | - | - |
| 1981 | 730 | 459 | 13829 | - | - |
| 1982 | 748 | 679 | 15362 | - | - |
| 1983 | 1648 | 1388 | 19111 | - | - |
| 1984 | 1200 | 1453 | 19230 | - | - |
| 1985 | 1668 | 750 | 17527 | - | - |
| 1986 | 1677 | 497 | 20701 | - | - |
| 1987 | 2239 | 588 | 16285 | - | - |


| Year | Gillnet | Longline | Trawl | Danish seine | Other |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 2815 | 838 | 15934 | - | - |
| 1989 | 1342 | 197 | 18599 | - | - |
| 1990 | 1372 | 1491 | 20325 | - | - |
| 1991 | 1904 | 4552 | 26864 | - | - |
| 1992 | 1679 | 1787 | 5787 | - | - |
| 1993 | 1497 | 2493 | 7889 | - | - |
| 1994 | 1403 | 2392 | 5353 | - | - |
| 1995 | 1500 | 4034 | 5494 | - | - |
| 1996 | 1480 | 4616 | 7977 | - | - |
| 1997 | 998 | 3378 | 5198 | - | - |
| 1998 | 1327 | 7395 | 6664 | - | - |
| 1999 | 2565 | 6804 | 10177 | - | - |
| 2000 | 1707 | 5029 | 7700 | - | - |
| 2001 | 2041 | 6303 | 7968 | - | - |
| 2002 | 1737 | 5309 | 6115 | - | - |
| 2003 | 2046 | 5483 | 6049 | - | - |
| 2004 | 2290 | 7135 | 8778 | 599 | - |
| 2005 | 1842 | 7539 | 9420 | 447 | - |


| Year | Gillnet | Longline | Trawl | Danish seine | Other |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 1503 | 6146 | 10042 | 205 | - |
| 2007 | 997 | 4503 | 9618 | 119 | - |
| 2008 | 901 | 3575 | 9285 | 9 | 8 |
| 2009 | 1409 | 4952 | 6583 | 34 | 18 |
| 2010 | 1449 | 5427 | 8165 | 170 | 10 |
| 2011 | 1583 | 5039 | 9351 | 239 | 15 |
| 2012 | 1929 | 5602 | 12130 | 413 | 5 |
| 2013 | 2398 | 5805 | 13791 | 176 | 0 |
| 2014 | 2647 | 6166 | 13673 | 183 | 0 |
| 2015 | 2508 | 6287 | 15445 | 489 | 18 |
| 2016 | 2646 | 7290 | 14333 | 650 | 304 |
| 2017 | 2677 | 7221 | 15774 | 679 | 29 |
| 2018 | 3021 | 6542 | 17367 | 842 | 20 |
| 2019 | 3323 | 7028 | 17046 | 1119 | 0 |
| 2020 | 2976 | 6989 | 17675 | 1044 | 28 |
| 2021* | 2930 | 7385 | 17203 | 866 | 50 |

* Provisional figures.


## Table 8.6. Greenland halibut in subareas 1 and 2. Catch per unit effort and total effort.

| Year | USSR <br> catch/hour <br> trawling ( t ) $\mathbf{R T}^{\mathbf{1}}$ | PST ${ }^{2}$ | Norway ${ }^{10}$ catch/hour trawling ( t ) |  | Average CPUE |  | Total effort (in '000 hrs trawling) ${ }^{5}$ | CPUE 7+ ${ }^{6}$ | GDR ${ }^{\text {7 }}$ (catch/day tonnage (kg) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | $A^{8}$ | $B^{9}$ | $A^{3}$ | B ${ }^{4}$ |  |  |  |
| 1965 | 0.80 | - | - | - | 0.80 | - | - | - | - |
| 1966 | 0.77 | - | - | - | 0.77 | - | - | - | - |
| 1967 | 0.70 | - | - | - | 0.70 | - | - | - | - |
| 1968 | 0.65 | - | - | - | 0.65 | - | - | - | - |
| 1969 | 0.53 | - | - | - | 0.53 | - | - | - | - |
| 1970 | 0.53 | - | - | - | 0.53 | - | 169 | 0.50 | - |
| 1971 | 0.46 | - | - | - | 0.46 | - | 172 | 0.43 | - |
| 1972 | 0.37 | - | - | - | 0.37 | - | 116 | 0.33 | - |
| 1973 | 0.37 | - | 0.34 | - | 0.36 | - | 83 | 0.36 | - |
| 1974 | 0.40 | - | 0.36 | - | 0.38 | - | 100 | 0.36 | - |
| 1975 | 0.39 | 0.51 | 0.38 | - | 0.39 | 0.45 | 99 | 0.37 | - |
| 1976 | 0.40 | 0.56 | 0.33 | - | 0.37 | 0.45 | 100 | 0.34 | - |
| 1977 | 0.27 | 0.41 | 0.33 | - | 0.30 | 0.37 | 96 | 0.26 | - |
| 1978 | 0.21 | 0.32 | 0.21 | - | 0.21 | 0.27 | 123 | 0.17 | - |
| 1979 | 0.23 | 0.35 | 0.28 | - | 0.26 | 0.32 | 67 | 0.19 | - |
| 1980 | 0.24 | 0.33 | 0.32 | - | 0.28 | 0.33 | 47 | 0.25 | - |


| Year | USSR catch/hour trawling ( t ) |  | Norway ${ }^{10}$ catch/hour trawling ( t ) |  | Average CPUE |  | Total effort (in '000 hrs trawling) ${ }^{5}$ | CPUE 7+ ${ }^{\text { }}$ | GDR ${ }^{\text {7 }}$ (catch/day tonnage (kg) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | RT ${ }^{1}$ | PST ${ }^{2}$ | A $^{8}$ | B ${ }^{9}$ | $A^{3}$ | B4 |  |  |  |
| 1981 | 0.30 | 0.36 | 0.36 | - | 0.33 | 0.36 | 42 | 0.28 | - |
| 1982 | 0.26 | 0.45 | 0.41 | - | 0.34 | 0.43 | 39 | 0.37 | - |
| 1983 | 0.26 | 0.40 | 0.35 | - | 0.31 | 0.38 | 58 | 0.32 | - |
| 1984 | 0.27 | 0.41 | 0.32 | - | 0.30 | 0.37 | 59 | 0.30 | - |
| 1985 | 0.28 | 0.52 | 0.37 | - | 0.33 | 0.45 | 44 | 0.37 | - |
| 1986 | 0.23 | 0.42 | 0.37 | - | 0.30 | 0.40 | 57 | 0.32 | - |
| 1987 | 0.25 | 0.50 | 0.35 | - | 0.30 | 0.43 | 44 | 0.35 | - |
| 1988 | 0.20 | 0.30 | 0.31 | - | 0.26 | 0.31 | 63 | 0.26 | 4.26 |
| 1989 | 0.20 | 0.30 | 0.26 | - | 0.23 | 0.28 | 73 | 0.19 | 2.95 |
| 1990 | - | 0.20 | 0.27 | - | - | 0.24 | 95 | 0.16 | 1.66 |
| 1991 | - | - | 0.24 | - | - | - | 134 | 0.18 | - |
| 1992 | - | - | 0.46 | 0.72 | - | - | 20 | 0.29 | - |
| 1993 | - | - | 0.79 | 1.22 | - | - | 15 | 0.65 | - |
| 1994 | - | - | 0.77 | 1.27 | - | - | 11 | 0.70 | - |
| 1995 | - | - | 1.03 | 1.48 | - | - | - | - | - |
| 1996 | - | - | 1.45 | 1.82 | - | - | - | - | - |
| 1997 | 0.71 | - | 1.23 | 1.60 | - | - | - | - | - |


| Year | USSR <br> catch/hour <br> trawling (t) |  | Norway ${ }^{10}$ <br> catch/hour <br> trawling (t) | Average CPUE | Total effort (in '000 hrs trawling) |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

${ }^{1}$ Side trawlers, 800-1000 hp. From 1983 onwards, stern trawlers (SRTM), 1000 hp . From 1997 based on research fishing.
2 Stern trawlers, up to 2000 HP .
${ }^{3}$ Arithmetic average of CPUE from USSR RT (or SRTM trawlers) and Norwegian trawlers.
${ }^{4}$ Arithmetic average of CPUE from USSR PST and Norwegian trawlers.
${ }^{5}$ For the years 1981-1990, based on average CPUE type B. For 1991-1993, based on the Norwegian CPUE, type A.
${ }^{6}$ Total catch ( $\mathbf{t}$ ) of seven years and older fish divided by total effort.
${ }^{7}$ For the years 1988-1989, frost-trawlers 995 BRT (FAO Code 095). For 1990, factory trawlers S IV, 1943 BRT (FAO Code 090).
${ }^{8}$ Norwegian trawlers, ISSC-code 07, 250-499.9 GRT.
${ }^{9}$ Norwegian factory trawlers, ISSCFV-code 09, 1000-1999.9 GRT
${ }^{10}$ From 1992 based on research fishing. 1992-1993: two weeks in May/June and October; 1994-1995: 10 days in May/June
${ }^{11}$ Based on fishery from April-October only, a period with relatively low CPUE. In previous years fishery was carried out throughout the whole year.
${ }^{12}$ Based on fishery from October-December only, a period with relatively high CPUE.
${ }^{13}$ Based on fishery from October-November only.

Table 8.7. Greenland halibut in subareas 1 and 2. Catch history back to 1935.

| Year | Norway | Russia | Others | Total | Year | Norway | Russia | Others | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1935 | 1534 | $\mathrm{n} / \mathrm{a}$ | - | 1534 | 1979 | 2843 | 10311 | 4088 | 17312 |
| 1936 | 830 | $\mathrm{n} / \mathrm{a}$ | - | 830 | 1980 | 3157 | 7670 | 2457 | 13284 |
| 1937 | 616 | $\mathrm{n} / \mathrm{a}$ | - | 616 | 1981 | 4201 | 9276 | 1541 | 15018 |
| 1938 | 329 | $\mathrm{n} / \mathrm{a}$ | - | 329 | 1982 | 3206 | 12394 | 1189 | 16789 |
| 1939 | 459 | $\mathrm{n} / \mathrm{a}$ | - | 459 | 1983 | 4883 | 15152 | 2112 | 22147 |
| 1940 | 846 | n/a | - | 846 | 1984 | 4376 | 15181 | 2326 | 21883 |
| 1941 | 1663 | n/a | - | 1663 | 1985 | 5464 | 10237 | 4244 | 19945 |


| Year | Norway | Russia | Others | Total | Year | Norway | Russia | Others | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1942 | 955 | $\mathrm{n} / \mathrm{a}$ | - | 955 | 1986 | 7890 | 12200 | 2785 | 22875 |
| 1943 | 824 | $\mathrm{n} / \mathrm{a}$ | - | 824 | 1987 | 7261 | 9733 | 2118 | 19112 |
| 1944 | 678 | $\mathrm{n} / \mathrm{a}$ | - | 678 | 1988 | 9076 | 9430 | 1081 | 19587 |
| 1945 | 1148 | $\mathrm{n} / \mathrm{a}$ | - | 1148 | 1989 | 10622 | 8812 | 704 | 20138 |
| 1946 | 1337 | 25 | - | 1362 | 1990 | 17243 | 4764 | 1176 | 23183 |
| 1947 | 1409 | 28 | - | 1437 | 1991 | 27587 | 2490 | 3243 | 33320 |
| 1948 | 1877 | 110 | - | 1987 | 1992 | 7667 | 718 | 217 | 8602 |
| 1949 | 198 | 177 | - | 375 | 1993 | 10380 | 1235 | 318 | 11933 |
| 1950 | 1853 | 221 | - | 2074 | 1994 | 8428 | 283 | 515 | 9226 |
| 1951 | 2438 | 423 | - | 2861 | 1995 | 9368 | 794 | 1572 | 11734 |
| 1952 | 2576 | 377 | - | 2953 | 1996 | 11623 | 1576 | 1148 | 14347 |
| 1953 | 2208 | 393 | - | 2601 | 1997 | 7661 | 1038 | 711 | 9410 |
| 1954 | 3674 | 416 | - | 4090 | 1998 | 8435 | 2659 | 799 | 11893 |
| 1955 | 3010 | 290 | - | 3300 | 1999 | 15004 | 3823 | 690 | 19517 |
| 1956 | 3493 | 446 | - | 3939 | 2000 | 9083 | 4568 | 646 | 14297 |
| 1957 | 4130 | 505 | - | 4635 | 2001 | 10896 | 4694 | 775 | 16365 |
| 1958 | 2931 | 1261 | - | 4192 | 2002 | 7143 | 5584 | 566 | 13293 |
| 1959 | 4307 | 3632 | - | 7939 | 2003 | 8216 | 4384 | 847 | 13447 |


| Year | Norway | Russia | Others | Total | Year | Norway | Russia | Others | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | 6662 | 4299 | - | 10961 | 2004 | 13939 | 4662 | 298 | 18899 |
| 1961 | 7977 | 3836 | - | 11813 | 2005 | 13011 | 4883 | 940 | 18834 |
| 1962 | 11600 | 1760 | - | 13360 | 2006 | 11119 | 6055 | 730 | 17904 |
| 1963 | 11300 | 3240 | - | 14540 | 2007 | 8230 | 6484 | 739 | 15453 |
| 1964 | 14200 | 26191 | - | 40391 | 2008 | 7393 | 5294 | 1105 | 13792 |
| 1965 | 18000 | 16682 | - | 34751 | 2009 | 8446 | 3335 | 1210 | 12990 |
| 1966 | 16434 | 9768 | 119 | 26321 | 2010 | 7700 | 6888 | 641 | 15229 |
| 1967 | 17528 | 5737 | 1002 | 24267 | 2011 | 8270 | 7053 | 1283 | 16606 |
| 1968 | 22514 | 3397 | 257 | 26168 | 2012 | 9331 | 10041 | 916 | 20288 |
| 1969 | 14856 | 19760 | 9173 | 43789 | 2013 | 10403 | 10310 | 1454 | 22167 |
| 1970 | 15871 | 35578 | 38035 | 89484 | 2014 | 11232 | 10061 | 1732 | 23025 |
| 1971 | 9466 | 54339 | 15229 | 79034 | 2015 | 10874 | 12953 | 921 | 24748 |
| 1972 | 15983 | 16193 | 10872 | 43055 | 2016 | 12932 | 10576 | 1440 | 24948 |
| 1973 | 13989 | 8561 | 7349 | 29938 | 2017 | 13741 | 10714 | 1925 | 26380 |
| 1974 | 8791 | 16958 | 11972 | 37763 | 2018 | 14874 | 12072 | 1598 | 28544 |
| 1975 | 4858 | 20372 | 12914 | 38172 | 2019 | 14813 | 12198 | 1471 | 28482 |
| 1976 | 6005 | 16580 | 13469 | 36074 | 2020 | 14532 | 12266 | 1915 | 28713 |
| 1977 | 4217 | 15045 | 9613 | 28827 | 2021* | 14008 | 12394 | 2029 | 28431 |


| Year | Norway | Russia | Others | Total | Year | Norway | Russia |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1978 | 4082 | 14651 | 5884 | 24617 |  |  |  |

* Provisional figures.

Table 8.8. Greenland halibut in ICES Division 4.a (North Sea). Nominal catch ( $\mathbf{t}$ ) by countries as officially reported to ICES. Not included in the assessment.

|  |  |  |  |  |  | $\begin{aligned} & \text { ס्ट } \\ & \text { त्0 } \\ & \underline{\underline{N}} \end{aligned}$ | $\begin{aligned} & \text { त } \\ & \text { 3 } \\ & \text { 30 } \end{aligned}$ |  | O |  |  |  | $\stackrel{\square}{\square}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1973 | 0 | 0 | 0 | 4 | 0 | 0 | 9 | 8 | 0 | 28 | 0 | 0 | 49 |
| 1974 | 0 | 0 | 0 | 2 | 0 | 0 | 2 | 0 | 0 | 30 | 0 | 0 | 34 |
| 1975 | 0 | 0 | 0 | 1 | 0 | 0 | 4 | 0 | 0 | 12 | 0 | 0 | 17 |
| 1976 | 0 | 0 | 0 | 1 | 0 | 0 | 2 | 0 | 0 | 18 | 0 | 0 | 21 |
| 1977 | 0 | 0 | 0 | 2 | 0 | 0 | 2 | 0 | 0 | 8 | 0 | 0 | 12 |
| 1978 | 0 | 0 | 2 | 30 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 33 |
| 1979 | 0 | 0 | 2 | 16 | 0 | 0 | 2 | 0 | 0 | 1 | 0 | 0 | 21 |
| 1980 | 0 | 177 | 0 | 34 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 216 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 0 | 0 | 0 | 0 | 0 | 7 |
| 1982 | 0 | 0 | 2 | 26 | 0 | 0 | 17 | 0 | 0 | 0 | 0 | 0 | 45 |
| 1983 | 0 | 0 | 1 | 64 | 0 | 0 | 89 | 0 | 0 | 0 | 0 | 0 | 154 |
| 1984 | 0 | 0 | 3 | 50 | 0 | 0 | 32 | 0 | 0 | 0 | 0 | 0 | 85 |
| 1985 | 0 | 1 | 2 | 49 | 0 | 0 | 12 | 0 | 0 | 0 | 0 | 0 | 64 |


|  |  |  | 쁜 픈 |  |  | $\begin{aligned} & \text { ס } \\ & \text { 들 } \\ & \underline{\underline{N}} \end{aligned}$ | $\begin{aligned} & \text { 㐅} \\ & \text { ふ̀ } \\ & \text { 30 } \end{aligned}$ | $\begin{aligned} & \underset{\sim}{n} \\ & \underset{\sim}{\boldsymbol{n}} \end{aligned}$ | © |  |  |  | $\begin{aligned} & \text { ָ̄ } \\ & \stackrel{0}{0} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 0 | 0 | 30 | 2 | 0 | 0 | 34 | 0 | 0 | 0 | 0 | 0 | 66 |
| 1987 | 0 | 28 | 16 | 1 | 0 | 0 | 35 | 0 | 0 | 0 | 0 | 0 | 80 |
| 1988 | 0 | 71 | 62 | 3 | 0 | 0 | 19 | 0 | 0 | 1 | 0 | 0 | 156 |
| 1989 | 0 | 21 | 14 | 1 | 0 | 0 | 197 | 0 | 0 | 5 | 0 | 0 | 238 |
| 1990 | 0 | 10 | 30 | 3 | 0 | 0 | 29 | 0 | 0 | 4 | 0 | 0 | 76 |
| 1991 | 0 | 48 | 291 | 1 | 0 | 0 | 216 | 0 | 0 | 2 | 0 | 0 | 558 |
| 1992 | 1 | 15 | 416 | 3 | 0 | 0 | 626 | 0 | 0 | + | 1 | 0 | 1062 |
| 1993 | 1 | 0 | 78 | 1 | 0 | 0 | 858 | 0 | 0 | 10 | + | 0 | 948 |
| 1994 | + | 103 | 84 | 4 | 0 | 0 | 724 | 0 | 0 | 6 | 0 | 0 | 921 |
| 1995 | + | 706 | 165 | 2 | 0 | 0 | 460 | 0 | 0 | 52 | 283 | 0 | 1668 |
| 1996 | + | 0 | 249 | 1 | 0 | 0 | 1496 | 0 | 0 | 105 | 159 | 0 | 514 |
| 1997 | + | 0 | 316 | 3 | 0 | 0 | 873 | 0 | 0 | 1 | 162 | 0 | 1355 |
| 1998 | + | 0 | 71 | 10 | 0 | 10 | 804 | 0 | 0 | 35 | 435 | 0 | 1365 |
| 1999 | + | 0 |  | 1 | 0 | 18 | 2157 | 0 | 0 | 43 | 358 | 0 | 2577 |
| 2000 | + |  | 41 | 10 | 0 | 19 | 498 | 0 | 0 | 67 | 192 | 0 | 827 |
| 2001 | + |  | 43 | 0 | 0 | 10 | 470 | 0 | 0 | 122 | 202 | 0 | 847 |
| 2002 | + |  | 8 | + | 0 | 2 | 200 | 0 | 0 | 10 | 246 | 0 | 466 |


| $\begin{aligned} & \text { ॠ } \\ & \text { ٪ } \end{aligned}$ |  |  | 凹 |  |  | $\begin{aligned} & \text { ס्ट } \\ & \text { त्0 } \\ & \underline{\underline{N}} \end{aligned}$ | $\begin{aligned} & \text { त } \\ & \text { 3 } \\ & \text { 30 } \end{aligned}$ | $\begin{aligned} & \stackrel{\pi}{n} \\ & \\ & \end{aligned}$ | © |  |  |  | + |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 0 | 0 | 1 | + | + | + | 453 | 0 | 0 | + | 122 | 0 | 576 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 413 | 0 | 0 | 90 | 0 | 0 | 503 |
| 2005 | 0 | 0 | 2 | 0 | 0 | 0 | 58 | 0 | 0 | 4 | 0 | 0 | 64 |
| 2006 | 0 | 0 | 3 | 0 | 0 | 0 | 90 | 0 | 0 | 0 | 7 | 0 | 100 |
| 2007 | 0 | 1 | 0 | 0 | 0 | 0 | 133 | 0 | 0 | 1 | 6 | 0 | 141 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 14 | 0 | 0 | 0 | 22 | 0 | 36 |
| 2009 | 0 | 9 | 22 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 129 | 0 | 165 |
| 2010 | + | 1 | 38 | 0 | 0 | 0 | 10 | 0 | 0 | 0 | 49 | 0 | 98 |
| 2011 | 0 | 1 | 39 | 0 | 0 | 0 | 94 | 0 | 0 | 0 | 44 | 0 | 178 |
| 2012 | 0 | 0 | 14 | 0 | 0 | 0 | 788 | 0 | 0 | 0 | 43 | 0 | 845 |
| 2013 | 0 | 0 | 25 | 0 | 0 | 0 | 122 | 0 | 0 | 0 | 174 | 0 | 321 |
| 2014 | 0 | 2 | 27 | 0 | 0 | 0 | 723 | 0 | 0 |  | 104 | 0 | 856 |
| 2015 | 0 | 0 | 34 | 1 | 0 | 0 | 1151 | 0 | 0 | 0 | 127 | 0 | 1313 |
| 2016 | 0 | 0 | 31 | 0 | 0 | 0 | 983 | 0 | 0 | 0 | 120 | 0 | 1134 |
| 2017 | 0 | 0 | 20 | 0 | 0 | 0 | 753 | 0 | 0 | 0 | 73 | 0 | 846 |
| 2018 | 0 | 0 | 15 | 0 | 0 | 0 | 472 | 0 | 42 | 0 | 0 | 0 | 532 |
| 2019 | 0 | 0 | 21 | 0 | 0 | 0 | 241 | 0 | 14 | 0 | 0 | 1 | 277 |


|  |  |  |  |  |  | $\begin{aligned} & \text { 들 } \\ & \text { त } \\ & \underline{\underline{0}} \end{aligned}$ | $\begin{aligned} & \text { 㐅} \\ & \text { 3 } \\ & \text { 30 } \end{aligned}$ |  | O |  |  |  | ¢ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2020 | 0 | 0 | 10 | 0 | 0 | 0 | 663 | 0 | 45 | 0 | 0 | 1 | 719 |
| 2021* | 0 | 4 | 19 | 0 | 0 | 0 | 0 | 0 | 121 | 0 | 0 | 0 | 144 |

${ }^{*}$ Provisional figures.

Table 8.9. Abundance indices of different length groups in Russian autumn survey.

| Year/Length (cm) | $\leq 30$ | 31-35 | 36-40 | 41-45 | 46-50 | 51-55 | 56-60 | 61-65 | 66-70 | 71-75 | 76-80 | >80 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 4837 | 5078 | 11690 | 21171 | 15167 | 10886 | 7370 | 6549 | 3751 | 1786 | 1128 | 483 | 89896 |
| 1985 | 4003 | 6748 | 16858 | 24897 | 23244 | 15702 | 8376 | 5704 | 3776 | 2054 | 1028 | 698 | 113088 |
| 1986 | 3482 | 6062 | 13765 | 18945 | 15997 | 10369 | 4839 | 3022 | 2534 | 1325 | 440 | 205 | 80985 |
| 1987 | 2010 | 4828 | 7228 | 10490 | 8831 | 5513 | 2123 | 1784 | 1437 | 645 | 481 | 421 | 45791 |
| 1988 | 3374 | 5111 | 9022 | 10147 | 10128 | 5828 | 2265 | 1862 | 1218 | 511 | 361 | 341 | 50168 |
| 1989 | 2030 | 7055 | 13962 | 17252 | 16790 | 10028 | 3789 | 1916 | 1279 | 415 | 200 | 388 | 75104 |
| 1990 | 2762 | 6056 | 12802 | 13061 | 9527 | 9829 | 4967 | 2094 | 589 | 312 | 115 | 119 | 62233 |
| 1991 | 1036 | 5012 | 16237 | 20998 | 17418 | 11728 | 8012 | 4562 | 814 | 181 | 122 | 174 | 86294 |
| 1992 | 184 | 2153 | 17185 | 32399 | 22481 | 12977 | 6229 | 3473 | 1869 | 502 | 182 | 106 | 99740 |
| 1993 | - | 290 | 3593 | 14782 | 21080 | 16013 | 6743 | 3341 | 2031 | 859 | 269 | 164 | 69165 |
| 1994 | 49 | 17 | 1651 | 12582 | 16203 | 12566 | 5391 | 3320 | 2019 | 819 | 188 | 106 | 54911 |
| 1995 | - | 38 | 1245 | 13193 | 20571 | 12445 | 5432 | 2717 | 1587 | 579 | 187 | 82 | 58076 |


| Year/Length (cm) | $\leq 30$ | 31-35 | 36-40 | 41-45 | 46-50 | 51-55 | 56-60 | 61-65 | 66-70 | 71-75 | 76-80 | >80 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996* | - | 11 | 786 | 13012 | 30573 | 18294 | 5730 | 1795 | 773 | 534 | 169 | 12 | 71689 |
| 1997 | 140 | 152 | 1318 | 7744 | 18504 | 17221 | 6932 | 3079 | 1952 | 465 | 195 | 142 | 57844 |
| 1998 | 2449 | 2238 | 2949 | 10847 | 24266 | 19640 | 11112 | 5946 | 2158 | 440 | 172 | 90 | 82307 |
| 1999 | 1070 | 2815 | 4632 | 7886 | 17734 | 18489 | 10158 | 4827 | 2043 | 529 | 196 | 74 | 70453 |
| 2000 | 1274 | 1698 | 5184 | 14996 | 24170 | 20721 | 12805 | 5675 | 3100 | 1228 | 240 | 143 | 91234 |
| 2001 | 1399 | 2887 | 7496 | 18136 | 34752 | 29886 | 13463 | 6759 | 3772 | 1511 | 593 | 369 | 121024 |
| 2002** | 662 | 2033 | 6395 | 13329 | 19810 | 13135 | 7180 | 3406 | 1311 | 381 | 129 | 58 | 67828 |
| 2003*** | 955 | 2396 | 7420 | 13006 | 17160 | 11630 | 7978 | 5332 | 3541 | 985 | 485 | 238 | 71126 |
| 2004 | 1431 | 2705 | 11945 | 16937 | 20155 | 18274 | 12594 | 6948 | 4783 | 2087 | 813 | 536 | 99209 |
| 2005 | 830 | 3970 | 10726 | 17850 | 17547 | 15164 | 9726 | 5859 | 3343 | 1150 | 453 | 545 | 87163 |
| 2006**** | 293 | 1981 | 18471 | 35224 | 36563 | 26335 | 14138 | 7248 | 4943 | 1669 | 668 | 488 | 148021 |
| 2007 | 376 | 1431 | 6937 | 24330 | 26780 | 26086 | 22157 | 15586 | 7480 | 3786 | 932 | 628 | 136510 |
| 2008 | 463 | 4626 | 19991 | 28799 | 30062 | 32159 | 23175 | 11326 | 8368 | 4198 | 1872 | 1089 | 166129 |
| 2009 | 152 | 4919 | 29389 | 48321 | 45833 | 33915 | 24484 | 10227 | 6568 | 3032 | 881 | 616 | 208338 |
| 2010 | 146 | 5097 | 37901 | 66086 | 57863 | 46321 | 25428 | 10058 | 8612 | 3983 | 1587 | 1610 | 264692 |
| 2011 | 456 | 1285 | 22470 | 61115 | 78247 | 64186 | 49620 | 19412 | 11607 | 7226 | 3529 | 874 | 320025 |
| 2012 | 213 | 798 | 12051 | 49062 | 56704 | 52393 | 36362 | 13622 | 7533 | 4213 | 1944 | 1611 | 236506 |

$2013^{* * * * *}$

| Year/Length (cm) | $\leq 30$ | 31-35 | 36-40 | 41-45 | 46-50 | 51-55 | 56-60 | 61-65 | 66-70 | 71-75 | 76-80 | >80 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2014 | 17 | 1697 | 10296 | 34074 | 45287 | 35861 | 22621 | 8613 | 5505 | 2227 | 929 | 427 | 167553 |
| 2015 | 318 | 2099 | 13542 | 35864 | 43551 | 36082 | 21114 | 10924 | 4472 | 1342 | 850 | 339 | 170497 |
| 2016***** |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2017 | 158 | 2198 | 10687 | 32464 | 61577 | 71590 | 40700 | 16830 | 7449 | 3483 | 1206 | 1245 | 249585 |
| 2018***** |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2019 | 144 | 2186 | 13500 | 27129 | 28572 | 22536 | 13943 | 5825 | 3080 | 1654 | 707 | 406 | 119742 |
| 2020***** |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2021***** |  |  |  |  |  |  |  |  |  |  |  |  |  |

* Only half of the standard area was investigated
** No observations in NEEZ
*** Observations in the NEEZ on the main spawning grounds were conducted considerably later than usual
**** Survey was conducted by one vessel with a reduced number of trawls at depths less than 500 m
*****No indices for 2013, 2016, 2018,2020 and 2021

Table 8.10. Abundance indices of different length groups in Norwegian autumn slope survey (in thousands)

| Year | $<30$ | $\mathbf{3 0}$ | $\mathbf{3 1}$ | $\mathbf{3 2}$ | $\mathbf{3 3}$ | $\mathbf{3 4}$ | $\mathbf{3 5}$ | $\mathbf{3 6}$ | $\mathbf{3 7}$ | $\mathbf{3 8}$ | $\mathbf{3 9}$ | $\mathbf{4 0}$ | $\mathbf{4 1}$ | $\mathbf{4 2}$ | $\mathbf{4 3}$ | $\mathbf{4 4}$ | $\mathbf{4 5}$ | $\mathbf{4 6}$ | $\mathbf{4 7}$ | $\mathbf{4 8}$ | $\mathbf{4 9}$ | $\mathbf{5 0}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1994 | 0 | 0 | 0 | 0 | 1 | 15 | 23 | 80 | 197 | 335 | 645 | 1225 | 1611 | 2432 | 3431 | 3511 | 3830 | 3519 | 3940 | 3724 | 2896 | 3020 |
| 1995 | 0 | 0 | 1 | 3 | 6 | 15 | 29 | 86 | 141 | 242 | 472 | 931 | 1210 | 2294 | 3092 | 3840 | 4475 | 4540 | 4633 | 4321 | 3836 | 3856 |
| 1996 | 0 | 2 | 1 | 6 | 6 | 2 | 18 | 49 | 54 | 166 | 321 | 772 | 957 | 1787 | 2912 | 3769 | 4728 | 5199 | 5944 | 5644 | 5224 | 5132 |
| 1997 | 7 | 5 | 11 | 4 | 33 | 27 | 49 | 186 | 250 | 297 | 443 | 862 | 1009 | 1814 | 2888 | 3578 | 5451 | 5402 | 6132 | 5206 | 4125 | 5455 |


| Year | <30 | 30 | 31 | 32 | 33 | 34 | 35 | 36 | 37 | 38 | 39 | 40 | 41 | 42 | 43 | 44 | 45 | 46 | 47 | 48 | 49 | 50 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 7 | 2 | 6 | 15 | 17 | 22 | 51 | 103 | 174 | 219 | 372 | 504 | 727 | 1061 | 1491 | 2103 | 2941 | 3092 | 3609 | 3735 | 3851 | 4850 |
| 1999 | 10 | 4 | 18 | 15 | 20 | 40 | 61 | 75 | 110 | 174 | 202 | 377 | 476 | 862 | 1175 | 1655 | 2397 | 2543 | 3485 | 4214 | 3694 | 5274 |
| 2000 | 2 | 7 | 11 | 30 | 34 | 46 | 128 | 122 | 163 | 264 | 383 | 677 | 739 | 932 | 1183 | 1439 | 2038 | 2030 | 2268 | 2644 | 2846 | 3888 |
| 2001 | 21 | 20 | 35 | 37 | 77 | 147 | 274 | 270 | 440 | 462 | 724 | 986 | 1176 | 1373 | 1630 | 1720 | 2724 | 2655 | 3349 | 3128 | 3973 | 3999 |
| 2002 | 97 | 75 | 107 | 122 | 180 | 267 | 399 | 404 | 723 | 669 | 869 | 1026 | 1097 | 1360 | 1883 | 1870 | 2560 | 2185 | 3322 | 3450 | 3597 | 4032 |
| 2003 | 38 | 27 | 65 | 97 | 172 | 270 | 383 | 692 | 783 | 894 | 1214 | 1100 | 1481 | 1561 | 2082 | 1792 | 2468 | 2104 | 3193 | 3360 | 3506 | 3117 |
| 2004 | 27 | 15 | 47 | 125 | 191 | 402 | 636 | 639 | 951 | 1042 | 1092 | 1206 | 1337 | 1319 | 1398 | 1546 | 2013 | 1967 | 2638 | 2646 | 3337 | 3373 |
| 2005 | 66 | 104 | 285 | 317 | 517 | 765 | 861 | 1220 | 1492 | 1540 | 2053 | 2295 | 2293 | 2588 | 2262 | 2677 | 3041 | 2446 | 2854 | 2095 | 3056 | 2336 |
| 2006 | 12 | 50 | 80 | 158 | 258 | 456 | 849 | 1022 | 1429 | 1579 | 1603 | 1900 | 1823 | 1824 | 2015 | 1974 | 2529 | 2359 | 2350 | 2137 | 2338 | 2175 |
| 2007 | 157 | 96 | 161 | 359 | 766 | 1423 | 2508 | 3142 | 4411 | 5679 | 5346 | 5639 | 5502 | 5038 | 4600 | 3632 | 3667 | 3628 | 3278 | 2571 | 2882 | 2597 |
| 2008 | 378 | 384 | 723 | 1323 | 1763 | 1793 | 2441 | 2911 | 3249 | 3685 | 4229 | 4300 | 4257 | 3568 | 3911 | 3534 | 3020 | 3066 | 2769 | 2582 | 2639 | 2284 |
| 2009 | 31 | 36 | 93 | 349 | 505 | 934 | 1663 | 2660 | 3050 | 3680 | 4138 | 4885 | 5567 | 4148 | 5327 | 4639 | 3688 | 3752 | 3682 | 3410 | 3553 | 3215 |
| 2011 | 0 | 0 | 20 | 36 | 57 | 124 | 288 | 563 | 646 | 1414 | 1454 | 2228 | 2680 | 3174 | 3649 | 3750 | 3532 | 3031 | 3299 | 3991 | 3251 | 2454 |
| 2013 | 17 | 5 | 3 | 1 | 13 | 64 | 103 | 122 | 324 | 582 | 1022 | 1266 | 2138 | 2207 | 3553 | 3748 | 3476 | 4124 | 3717 | 3045 | 3718 | 3052 |
| 2015 | 3 | 24 | 24 | 36 | 131 | 318 | 439 | 721 | 757 | 1043 | 1253 | 1473 | 2602 | 2444 | 3776 | 4459 | 4602 | 4598 | 4371 | 3962 | 4156 | 3694 |
| 2017 | 6 | 20 | 45 | 54 | 63 | 144 | 184 | 328 | 593 | 365 | 928 | 955 | 1267 | 1457 | 1764 | 1983 | 2367 | 2465 | 2651 | 2569 | 2816 | 3011 |
| 2019 | 0 | 0 | 28 | 43 | 128 | 362 | 372 | 569 | 874 | 1322 | 1290 | 1424 | 1667 | 2285 | 2210 | 2168 | 2208 | 2229 | 2434 | 2119 | 2305 | 2405 |
| 2021 | 80 | 67 | 177 | 211 | 375 | 813 | 662 | 1010 | 1103 | 1156 | 1332 | 1680 | 1826 | 2338 | 2439 | 3818 | 3133 | 3597 | 2874 | 3601 | 3688 | 2875 |


| Year | 51 | 52 | 53 | 54 | 55 | 56 | 57 | 58 | 59 | 60 | 61 | 62 | 63 | 64 | 65 | 66 | 67 | 68 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 2545 | 2729 | 2398 | 2092 | 1975 | 1547 | 1488 | 1103 | 920 | 788 | 565 | 702 | 576 | 523 | 577 | 370 | 367 | 386 |
| 1995 | 3165 | 3152 | 2963 | 2647 | 2272 | 1756 | 1586 | 1153 | 970 | 880 | 764 | 690 | 680 | 592 | 525 | 461 | 387 | 334 |
| 1996 | 4106 | 3638 | 3571 | 2752 | 2177 | 1568 | 1443 | 1017 | 867 | 782 | 512 | 449 | 538 | 404 | 391 | 356 | 281 | 248 |
| 1997 | 3644 | 3427 | 3018 | 2302 | 2111 | 1502 | 1131 | 1042 | 617 | 849 | 585 | 576 | 537 | 403 | 446 | 481 | 294 | 230 |
| 1998 | 4211 | 3824 | 3166 | 2988 | 2857 | 1974 | 1714 | 1515 | 981 | 1172 | 783 | 613 | 598 | 668 | 641 | 569 | 479 | 364 |
| 1999 | 4092 | 5196 | 4136 | 3909 | 4122 | 2631 | 2299 | 1787 | 1374 | 1388 | 895 | 1037 | 865 | 886 | 923 | 791 | 807 | 594 |
| 2000 | 3692 | 3681 | 3512 | 3016 | 3197 | 2388 | 2007 | 1545 | 1227 | 1327 | 915 | 1028 | 734 | 630 | 732 | 517 | 509 | 505 |
| 2001 | 3649 | 4512 | 4106 | 3005 | 3358 | 2552 | 2589 | 2147 | 1293 | 1350 | 1099 | 939 | 1187 | 684 | 787 | 612 | 751 | 603 |
| 2002 | 4241 | 3516 | 3966 | 3602 | 3855 | 2837 | 2511 | 2248 | 1672 | 1787 | 1239 | 1237 | 1139 | 808 | 882 | 604 | 679 | 474 |
| 2003 | 4400 | 3465 | 3808 | 3512 | 3907 | 3368 | 3035 | 2319 | 1896 | 1705 | 1612 | 1384 | 1542 | 1130 | 1350 | 972 | 994 | 675 |
| 2004 | 3535 | 4405 | 3614 | 3801 | 3249 | 2751 | 2252 | 1911 | 1493 | 1455 | 1372 | 1360 | 1284 | 1162 | 962 | 763 | 891 | 590 |
| 2005 | 2400 | 2734 | 2413 | 2084 | 2295 | 1882 | 1681 | 1492 | 1458 | 1168 | 1241 | 1057 | 1065 | 984 | 903 | 782 | 865 | 479 |
| 2006 | 2493 | 2125 | 2290 | 2025 | 2189 | 1790 | 1668 | 1542 | 1337 | 1159 | 1188 | 1009 | 925 | 1036 | 807 | 798 | 647 | 678 |
| 2007 | 2109 | 2249 | 2123 | 2142 | 1758 | 1609 | 1581 | 1070 | 1008 | 1044 | 625 | 938 | 672 | 558 | 537 | 526 | 394 | 469 |
| 2008 | 2288 | 2248 | 2229 | 1815 | 1751 | 1514 | 1150 | 1019 | 861 | 668 | 652 | 657 | 508 | 582 | 629 | 523 | 484 | 361 |
| 2009 | 2668 | 2944 | 2850 | 2441 | 2372 | 2233 | 1837 | 1698 | 1503 | 1135 | 845 | 962 | 647 | 858 | 715 | 607 | 653 | 609 |
| 2011 | 2905 | 2746 | 2602 | 2713 | 2387 | 1709 | 1704 | 1529 | 978 | 1179 | 577 | 649 | 554 | 440 | 466 | 315 | 440 | 550 |
| 2013 | 2498 | 2035 | 1905 | 1631 | 1710 | 1573 | 1424 | 1009 | 790 | 671 | 503 | 506 | 400 | 456 | 234 | 266 | 227 | 176 |



| Year | 69 | 70 | 71 | 72 | 73 | 74 | 75 | 76 | 77 | 78 | 79 | >80 | SUM |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | 289 | 254 | 261 | 101 | 140 | 130 | 75 | 52 | 80 | 59 | 47 | 278 | 90260 |
| 2008 | 313 | 258 | 226 | 201 | 138 | 107 | 59 | 62 | 89 | 66 | 76 | 508 | 80851 |
| 2009 | 574 | 541 | 271 | 386 | 219 | 171 | 191 | 112 | 121 | 89 | 100 | 407 | 93764 |
| 2011 | 415 | 409 | 200 | 285 | 235 | 193 | 225 | 204 | 175 | 51 | 87 | 503 | 67066 |
| 2013 | 162 | 173 | 124 | 114 | 109 | 112 | 66 | 72 | 79 | 34 | 43 | 260 | 55662 |
| 2015 | 252 | 265 | 176 | 195 | 186 | 205 | 89 | 78 | 73 | 141 | 53 | 286 | 69236 |
| 2017 | 178 | 185 | 88 | 98 | 77 | 51 | 61 | 50 | 35 | 40 | 46 | 184 | 49195 |
| 2019 | 144 | 117 | 71 | 81 | 50 | 44 | 32 | 31 | 9 | 13 | 12 | 113 | 43056 |
| 2021 | 226 | 188 | 130 | 103 | 154 | 113 | 77 | 58 | 76 | 70 | 27 | 175 | 64668 |

*Biennial surveys since 2009.

Table 8.11. Abundance indices of females of different length groups in Norwegian autumn slope survey (in thousands).

| Year | $<30$ | $\mathbf{3 0}$ | $\mathbf{3 1}$ | $\mathbf{3 2}$ | $\mathbf{3 3}$ | $\mathbf{3 4}$ | $\mathbf{3 5}$ | $\mathbf{3 6}$ | $\mathbf{3 7}$ | 38 | $\mathbf{3 9}$ | $\mathbf{4 0}$ | $\mathbf{4 1}$ | $\mathbf{4 2}$ | $\mathbf{4 3}$ | $\mathbf{4 4}$ | $\mathbf{4 5}$ | $\mathbf{4 6}$ | $\mathbf{4 7}$ | $\mathbf{4 8}$ | $\mathbf{4 9}$ | $\mathbf{5 0}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1994 | 0 | 0 | 0 | 0 | 1 | 15 | 23 | 80 | 196 | 335 | 643 | 1223 | 1611 | 2429 | 3426 | 3503 | 3824 | 3510 | 3934 | 3716 | 2886 | 3018 |
| 1995 | 0 | 0 | 1 | 3 | 6 | 15 | 29 | 86 | 141 | 242 | 472 | 930 | 1210 | 2291 | 3088 | 3837 | 4470 | 4537 | 4629 | 4317 | 3835 | 3855 |
| 1996 | 0 | 0 | 0 | 4 | 0 | 1 | 10 | 26 | 28 | 64 | 123 | 228 | 233 | 424 | 415 | 773 | 937 | 1020 | 1185 | 1151 | 1037 | 1374 |
| 1997 | 6 | 5 | 7 | 4 | 17 | 14 | 36 | 134 | 139 | 146 | 187 | 337 | 331 | 419 | 569 | 685 | 899 | 852 | 1169 | 1058 | 828 | 1226 |
| 1998 | 5 | 0 | 0 | 11 | 4 | 7 | 26 | 41 | 78 | 77 | 156 | 170 | 190 | 274 | 290 | 364 | 413 | 526 | 605 | 665 | 743 | 970 |
| 1999 | 2 | 0 | 1 | 0 | 7 | 14 | 19 | 12 | 41 | 68 | 93 | 137 | 117 | 227 | 285 | 300 | 336 | 313 | 496 | 574 | 533 | 1049 |
| 2000 | 1 | 5 | 6 | 14 | 16 | 16 | 44 | 44 | 65 | 121 | 155 | 201 | 229 | 245 | 268 | 278 | 374 | 311 | 303 | 411 | 410 | 517 |


| Year | <30 | 30 | 31 | 32 | 33 | 34 | 35 | 36 | 37 | 38 | 39 | 40 | 41 | 42 | 43 | 44 | 45 | 46 | 47 | 48 | 49 | 50 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 13 | 6 | 14 | 15 | 38 | 61 | 118 | 123 | 177 | 167 | 293 | 411 | 462 | 355 | 425 | 376 | 544 | 477 | 493 | 379 | 558 | 673 |
| 2002 | 51 | 48 | 58 | 60 | 77 | 109 | 178 | 182 | 290 | 275 | 326 | 319 | 306 | 407 | 500 | 378 | 515 | 331 | 483 | 461 | 501 | 575 |
| 2003 | 25 | 25 | 27 | 43 | 100 | 124 | 182 | 276 | 413 | 429 | 532 | 504 | 512 | 545 | 610 | 450 | 552 | 394 | 539 | 487 | 523 | 406 |
| 2004 | 15 | 3 | 13 | 61 | 83 | 160 | 305 | 278 | 436 | 358 | 434 | 404 | 440 | 384 | 381 | 454 | 413 | 362 | 382 | 309 | 427 | 472 |
| 2005 | 30 | 24 | 110 | 99 | 182 | 258 | 322 | 464 | 565 | 537 | 723 | 758 | 619 | 630 | 452 | 633 | 723 | 467 | 593 | 293 | 500 | 329 |
| 2006 | 4 | 19 | 48 | 81 | 148 | 187 | 327 | 442 | 595 | 674 | 713 | 686 | 648 | 568 | 649 | 482 | 619 | 501 | 503 | 512 | 468 | 452 |
| 2007 | 85 | 67 | 104 | 178 | 371 | 731 | 1321 | 1539 | 2259 | 2654 | 2515 | 2403 | 2454 | 2145 | 1580 | 1242 | 1132 | 988 | 851 | 727 | 640 | 554 |
| 2008 | 216 | 210 | 432 | 698 | 829 | 958 | 1190 | 1372 | 1529 | 1597 | 1720 | 1516 | 1625 | 1069 | 1180 | 928 | 889 | 948 | 834 | 677 | 773 | 615 |
| 2009 | 13 | 19 | 33 | 146 | 210 | 343 | 662 | 1001 | 1263 | 1470 | 1491 | 1814 | 1979 | 1441 | 1752 | 1533 | 1044 | 1195 | 1037 | 988 | 922 | 878 |
| 2011 | 0 | 0 | 8 | 22 | 24 | 31 | 103 | 175 | 195 | 469 | 311 | 538 | 642 | 722 | 623 | 645 | 686 | 664 | 528 | 665 | 751 | 298 |
| 2013 | 0 | 0 | 0 | 0 | 3 | 11 | 49 | 30 | 50 | 186 | 261 | 246 | 521 | 286 | 650 | 509 | 621 | 693 | 626 | 664 | 745 | 576 |
| 2015 | 0 | 7 | 7 | 19 | 67 | 149 | 183 | 304 | 380 | 358 | 391 | 377 | 491 | 387 | 549 | 490 | 682 | 904 | 632 | 689 | 761 | 766 |
| 2017 | 4 | 17 | 16 | 43 | 44 | 79 | 83 | 120 | 267 | 117 | 395 | 312 | 365 | 373 | 288 | 411 | 524 | 444 | 6277 | 453 | 439 | 579 |
| 2019 | 0 | 0 | 16 | 25 | 92 | 119 | 183 | 300 | 360 | 500 | 527 | 498 | 604 | 609 | 512 | 517 | 426 | 558 | 489 | 503 | 541 | 479 |
| 2021 | 41 | 15 | 96 | 105 | 239 | 423 | 355 | 536 | 475 | 484 | 450 | 595 | 551 | 475 | 592 | 450 | 522 | 539 | 450 | 733 | 744 | 591 |

*Biennial surveys since 2009.

| Year | $\mathbf{5 1}$ | $\mathbf{5 2}$ | $\mathbf{5 3}$ | $\mathbf{5 4}$ | $\mathbf{5 5}$ | $\mathbf{5 6}$ | $\mathbf{5 7}$ | $\mathbf{5 8}$ | $\mathbf{6 9}$ | $\mathbf{6 0}$ | $\mathbf{6 1}$ | $\mathbf{6 2}$ | $\mathbf{6 3}$ | $\mathbf{6 4}$ | $\mathbf{6 5}$ | $\mathbf{6 6}$ | $\mathbf{6 7}$ | $\mathbf{6 8}$ | $\mathbf{6 9}$ | $\mathbf{7 0}$ | $\mathbf{7 1}$ | $\mathbf{7 2}$ | $\mathbf{7 3}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1994 | 2535 | 2719 | 2384 | 2088 | 1969 | 1545 | 1482 | 1098 | 917 | 785 | 560 | 700 | 571 | 522 | 573 | 368 | 364 | 385 | 254 | 253 | 151 | 136 | 122 |
| 1995 | 3162 | 3145 | 2958 | 2646 | 2271 | 1752 | 1586 | 1152 | 968 | 875 | 761 | 689 | 680 | 592 | 525 | 461 | 387 | 333 | 339 | 244 | 181 | 179 | 97 |


| Year | 51 | 52 | 53 | 54 | 55 | 56 | 57 | 58 | 69 | 60 | 61 | 62 | 63 | 64 | 65 | 66 | 67 | 68 | 69 | 70 | 71 | 72 | 73 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 1044 | 886 | 895 | 771 | 527 | 547 | 639 | 548 | 508 | 602 | 410 | 401 | 481 | 383 | 387 | 344 | 281 | 230 | 232 | 167 | 118 | 123 | 93 |
| 1997 | 911 | 985 | 824 | 650 | 669 | 590 | 523 | 562 | 346 | 633 | 484 | 501 | 506 | 364 | 433 | 437 | 289 | 225 | 171 | 207 | 216 | 119 | 109 |
| 1998 | 995 | 1043 | 999 | 1056 | 903 | 758 | 754 | 831 | 667 | 907 | 615 | 543 | 569 | 639 | 638 | 567 | 453 | 362 | 308 | 307 | 235 | 222 | 225 |
| 1999 | 830 | 1105 | 928 | 1042 | 1287 | 1019 | 1002 | 955 | 845 | 1106 | 754 | 927 | 816 | 814 | 890 | 780 | 798 | 582 | 478 | 403 | 384 | 317 | 182 |
| 2000 | 590 | 591 | 593 | 663 | 756 | 816 | 704 | 649 | 670 | 839 | 699 | 829 | 620 | 588 | 665 | 487 | 491 | 495 | 328 | 376 | 230 | 210 | 167 |
| 2001 | 479 | 632 | 761 | 643 | 680 | 698 | 962 | 877 | 743 | 936 | 928 | 714 | 1062 | 594 | 772 | 577 | 746 | 598 | 488 | 370 | 279 | 170 | 207 |
| 2002 | 610 | 438 | 638 | 694 | 823 | 672 | 824 | 779 | 780 | 989 | 780 | 1024 | 813 | 705 | 827 | 598 | 656 | 443 | 458 | 383 | 295 | 251 | 183 |
| 2003 | 604 | 582 | 662 | 611 | 968 | 854 | 1111 | 964 | 1057 | 1126 | 1260 | 1165 | 1314 | 1085 | 1278 | 938 | 962 | 670 | 555 | 625 | 462 | 249 | 242 |
| 2004 | 461 | 638 | 570 | 693 | 760 | 937 | 876 | 839 | 966 | 998 | 1202 | 1186 | 1227 | 1116 | 932 | 749 | 885 | 585 | 639 | 420 | 373 | 325 | 461 |
| 2005 | 378 | 411 | 427 | 451 | 597 | 638 | 775 | 718 | 800 | 871 | 935 | 938 | 965 | 904 | 860 | 740 | 860 | 449 | 523 | 465 | 390 | 262 | 192 |
| 2006 | 490 | 458 | 461 | 392 | 537 | 523 | 545 | 678 | 805 | 796 | 893 | 865 | 820 | 927 | 775 | 768 | 637 | 633 | 468 | 499 | 376 | 285 | 178 |
| 2007 | 476 | 499 | 471 | 491 | 469 | 533 | 607 | 549 | 566 | 776 | 494 | 790 | 587 | 534 | 517 | 515 | 394 | 469 | 278 | 254 | 261 | 101 | 133 |
| 2008 | 509 | 481 | 515 | 495 | 443 | 547 | 441 | 543 | 466 | 490 | 530 | 572 | 482 | 539 | 610 | 514 | 483 | 361 | 309 | 252 | 226 | 201 | 138 |
| 2009 | 640 | 665 | 738 | 639 | 733 | 724 | 698 | 783 | 814 | 605 | 653 | 765 | 534 | 776 | 701 | 525 | 616 | 587 | 561 | 526 | 263 | 378 | 219 |
| 2011 | 557 | 468 | 480 | 472 | 466 | 369 | 329 | 469 | 324 | 378 | 341 | 523 | 477 | 348 | 450 | 300 | 415 | 550 | 393 | 409 | 192 | 285 | 235 |
| 2013 | 518 | 381 | 477 | 308 | 375 | 529 | 526 | 304 | 296 | 334 | 324 | 377 | 329 | 390 | 218 | 260 | 227 | 174 | 159 | 173 | 120 | 114 | 109 |
| 2015 | 826 | 770 | 744 | 579 | 811 | 649 | 471 | 494 | 553 | 537 | 470 | 462 | 420 | 450 | 270 | 283 | 339 | 283 | 251 | 265 | 176 | 195 | 186 |
| 2017 | 530 | 438 | 516 | 448 | 392 | 555 | 578 | 498 | 563 | 530 | 473 | 330 | 378 | 371 | 271 | 286 | 243 | 245 | 178 | 185 | 88 | 98 | 77 |


| Year | $\mathbf{5 1}$ | $\mathbf{5 2}$ | $\mathbf{5 3}$ | $\mathbf{5 4}$ | $\mathbf{5 5}$ | $\mathbf{5 6}$ | $\mathbf{5 7}$ | $\mathbf{5 8}$ | $\mathbf{6 9}$ | $\mathbf{6 0}$ | $\mathbf{6 1}$ | $\mathbf{6 2}$ | $\mathbf{6 3}$ | $\mathbf{6 4}$ | $\mathbf{6 5}$ | $\mathbf{6 6}$ | $\mathbf{6 7}$ | $\mathbf{6 8}$ | $\mathbf{6 9}$ | $\mathbf{7 0}$ | $\mathbf{7 1}$ | $\mathbf{7 2}$ | $\mathbf{7 3}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2019 | 401 | 481 | 431 | 494 | 351 | 391 | 324 | 458 | 402 | 367 | 277 | 254 | 260 | 257 | 210 | 218 | 174 | 123 | 143 | 114 | 71 | 81 | 50 |
| 2021 | 623 | 672 | 574 | 541 | 506 | 440 | 555 | 692 | 687 | 603 | 721 | 741 | 557 | 676 | 585 | 382 | 387 | 379 | 226 | 188 | 130 | 103 | 154 |

## *Biennial surveys since 2009

| Year | 74 | 75 | 76 | 77 | 78 | 79 | >80 | SUM |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 74 | 113 | 47 | 39 | 40 | 30 | 95 | 51911 |
| 1995 | 100 | 137 | 56 | 53 | 53 | 34 | 99 | 58202 |
| 1996 | 92 | 61 | 28 | 40 | 39 | 21 | 74 | 18961 |
| 1997 | 111 | 104 | 61 | 29 | 35 | 40 | 185 | 20387 |
| 1998 | 144 | 102 | 64 | 65 | 61 | 43 | 192 | 19839 |
| 1999 | 205 | 223 | 125 | 109 | 140 | 47 | 328 | 22940 |
| 2000 | 153 | 141 | 77 | 96 | 77 | 47 | 233 | 17914 |
| 2001 | 178 | 157 | 85 | 131 | 69 | 49 | 306 | 22069 |
| 2002 | 163 | 131 | 104 | 130 | 48 | 65 | 251 | 21985 |
| 2003 | 170 | 242 | 201 | 128 | 125 | 114 | 356 | 28378 |
| 2004 | 241 | 181 | 135 | 119 | 100 | 109 | 431 | 25728 |
| 2005 | 149 | 156 | 152 | 109 | 82 | 61 | 426 | 24995 |
| 2006 | 259 | 185 | 138 | 136 | 81 | 96 | 491 | 24521 |
| 2007 | 124 | 75 | 52 | 80 | 59 | 47 | 275 | 38016 |
| 2008 | 107 | 59 | 62 | 89 | 66 | 76 | 506 | 32917 |


| Year | 74 | 75 | 76 | 77 | 78 | 79 | >80 | SUM |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 171 | 191 | 104 | 121 | 80 | 100 | 385 | 36529 |
| 2011 | 193 | 225 | 204 | 175 | 51 | 87 | 503 | 18768 |
| 2013 | 112 | 66 | 72 | 79 | 34 | 43 | 260 | 14415 |
| 2015 | 205 | 89 | 78 | 73 | 141 | 53 | 286 | 20002 |
| 2017 | 51 | 61 | 50 | 35 | 40 | 46 | 184 | 20388 |
| 2019 | 44 | 32 | 31 | 9 | 13 | 12 | 113 | 14444 |
| 2021 | 113 | 77 | 58 | 76 | 70 | 27 | 175 | 21179 |

${ }^{*}$ Biennial surveys since 2009.
Table 8.12. Abundance indices (numbers in thousands) from bottom-trawl surveys in the Barents Sea standard area winter (Mehl et al., WD4 AFWG 2019).

| Year | Length group (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Biomass (tonnes) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\leq 14$ | 15-19 | 20-24 | 25-29 | 30-34 | 35-39 | 40-44 | 45-49 | 50-54 | 55-59 | 60-64 | 65-69 | 70-74 | 75-79 | $\geq 80$ | Total |  |
| 1994 | 0 | 0 | 21 | 76 | 148 | 1117 | 3139 | 4740 | 3615 | 1941 | 889 | 541 | 21 | 0 | 0 | 16248 | 19228 |
| 1995 | 298 | 0 | 0 | 0 | 90 | 129 | 2877 | 7182 | 5739 | 2027 | 1622 | 839 | 489 | 86 | 0 | 21378 | 27459 |
| 1996 | 4121 | 0 | 0 | 0 | 62 | 124 | 1214 | 4086 | 4634 | 1871 | 1112 | 638 | 337 | 74 | 12 | 18285 | 20256 |
| $1997{ }^{1}$ | 0 | 68 | 0 | 0 | 55 | 163 | 949 | 4313 | 5629 | 2912 | 1609 | 643 | 300 | 65 | 21 | 16728 | 24214 |
| $1998{ }^{1}$ | 68 | 220 | 945 | 578 | 481 | 487 | 1088 | 4016 | 6591 | 3076 | 1798 | 707 | 326 | 93 | 44 | 20518 | 27248 |
| 1999 | 43 | 84 | 241 | 436 | 566 | 269 | 784 | 1701 | 3097 | 1669 | 1094 | 491 | 89 | 75 | 0 | 10640 | 14681 |
| 2000 | 140 | 184 | 344 | 836 | 1722 | 3857 | 2253 | 1560 | 2144 | 1714 | 1191 | 615 | 249 | 76 | 0 | 16883 | 17246 |
| 2001 | 68 | 49 | 147 | 179 | 737 | 1525 | 3716 | 3271 | 2302 | 2010 | 1088 | 529 | 160 | 50 | 39 | 15871 | 18224 |



| Year | Length group (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Biomass <br> (tonnes) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\leq 14$ | 15-19 | 20-24 | 25-29 | 30-34 | 35-39 | 40-44 | 45-49 | 50-54 | 55-59 | 60-64 | 65-69 | 70-74 | 75-79 | $\geq 80$ | Total |  |
| 2019 | 0 | 0 | 0 | 375 | 272 | 1671 | 3285 | 4034 | 5177 | 4265 | 3570 | 2526 | 1328 | 535 | 137 | 27176 | 45912 |
| $2020^{3}$ | 80 | 91 | 2464 | 442 | 790 | 2272 | 4391 | 5136 | 4929 | 4613 | 3278 | 1803 | 894 | 384 | 250 | 29599 | 43631 |
| $2021{ }^{3}$ | 0 | 154 | 927 | 927 | 2370 | 2976 | 3869 | 4265 | 3516 | 2991 | 2378 | 1649 | 670 | 682 | 238 | 27613 | 37090 |

${ }^{1}$ Indices raised to also represent the Russian EEZ
${ }^{2}$ Not complete coverage in southeast due to restrictions, strata 7 area set to default and strata 13 as in 2005
${ }^{3}$ Indices not raised to also represent uncovered parts of the Russian EEZ.
${ }^{4}$ Indices raised to also represent uncovered parts of the Russian EEZ

Table 8.13. Greenland halibut catch in weight, numbers, and biomass (in tonnes) and abundance (in thousands) estimated from Spanish autumn and spring surveys. NB. Absolute biomass and abundance values must not be compared between spring and autumn surveys due to different gears. The trawl used during spring surveys is considered less efficient on benthic species as Greenland halibut and skates, and better to catch species less associated with bottom.

Autumn survey

| Year | Catch ( Kg ) | Catch (numbers) | Biomass $^{\text {TM }}$ | Abundance ('000) |
| :---: | :---: | :---: | :---: | :---: |
| 1997 | 195056 | 211533 | 344014 | 379444 |
| 1998 | 180974 | 187259 | 351466 | 373149 |
| 1999 | 198781 | 172687 | 436956 | 377792 |
| 2000 | 169389 | 140355 | 340619 | 291265 |
| 2001 | 152681 | 129289 | 283511 | 249219 |
| 2002 | 144335 | 115213 | 256460 | 207466 |
| 2003 | 151952 | 132117 | 283644 | 256327 |
| 2004 | 153859 | 135631 | 320485 | 283965 |
| 2005 | 144573 | 134566 | 317320 | 313459 |
| 2008 | 91573 | 101578 | 129221* | 144561* |
| 2010 | 167862 | 182464 | 191510* | 216731* |
| 2012 | 178607 | 174670 | 336543* | 339697* |
| 2013 | 172762 | 168619 | 264101* | 267548* |
| 2014 | 175553 | 160557 | 321485* | 307679* |
| 2016 | 176015 | 142413 | 247644* | 214778* |
| 2019 | 50880 | 45631 | 209439* | 187830* |

No survey in 2006, 2007, 2009, 2011, 2015, 2017, 2018, 2020 and 2021.
*New swept-area estimation method

Spring survey

| Year | Catch (Kg) | Catch (numbers) | Biomass $^{\text {TM }}$ | Abundance ('000) |
| :--- | :--- | :--- | :--- | :--- |
| 2008 | 96797 | 109515 | 38406 | 38951 |
| 2009 | 200299 | 222018 | 58273 | 65464 |
| 2011 | 136610 | 160566 | 98142 | 117666 |
| $2015^{* *}$ | 111425 | 105385 | 150385 | 155333 |

No survey in 2010, 2012, 2013 and 2014.
**Different from the one used during the 2014 Spanish "autumn" survey.

Table 8.14. Greenland halibut in subareas 1 and 2. The catch scenarios. Weights in tonnes. Assessment 2021 as basis for advice for 2022 and 2023. NB. according to working group forecast, this may diverge slightly from final advice by ACOMTAC for 2021 from EU/UK was not sat at the time of the working group and TAC change is thus relative only to the TAC sat by JRNFC.

Table a Greenland halibut in subareas 1 and 2. Annual catch scenarios for 2022. All weights are in tonnes

| Basis | Total catch (2022) | $H_{R}$ total (2022) | Biomass $45 \mathrm{~cm}+(2023)$ | \% Biomass $45 \mathrm{~cm}+$ change * | \% TAC change | \% Advice change *** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |
| $H \mathrm{C}=0.035$ | 19094 | 0.035 | 535 | -5\% | -29\% | -17\% |
| Other scenarios |  |  |  |  |  |  |
| $H R=0$ | 0 | 0 | 554 | -1\% | -100\% | -100\% |
| $H R=0.025$ | 13873 | 0.025 | 540 | -4\% | -49\% | -40\% |
| $\begin{aligned} & \text { Catch_SQ } \\ & \text { (HR=0.052/0.055) } \end{aligned}$ | 28713 | 0.052/0.055 | 526 | -6\% | 6\% | 25\% |

* Biomass $45 \mathrm{~cm}+2023$ relative to 2022 (561 tonnes).
** Advice in 2022 relative to TAC in 2021. Only TAC sat by JRNFC in 2021 ( 27000 tonnes) was available.
*** Advice value for 2022 relative to the advice value for 2021

Table b Greenland halibut in subareas 1 and 2. Annual catch scenarios for 2023. All weights are in tonnes.

| Basis | Total catch (2023) | $\begin{aligned} & \mathrm{HR}_{\text {total }} \\ & (\mathbf{2 0 2 3}) \end{aligned}$ | Biomass $45 \mathrm{~cm}+$ (2024) | \% Biomass 45 cm+ change * | \% Advice change ** |
| :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |
| $H R=0.035$ | 18494 | 0.035 | 523 | -2\% | -3\% |
| Other scenarios |  |  |  |  |  |
| $H R=0$ | 0 | 0 | 558 | 1\% | 0\% |
| $H R=0.025$ | 13590 | 0.025 | 533 | -1\% | -2\% |
| $\begin{aligned} & \text { Catch_SQ } \\ & \text { (HR=0.052/0.055) } \end{aligned}$ | 28713 | 0.052/0.055 | 505 | -4\% | 0\% |

[^14]

Figure 8.1. NEA Greenland halibut landings. Historical landings (Nedreaas and Smirnov 2003 and AFWG).


Figure 8.2. Spatial distribution of Greenland halibut catches in 2021 according to Norwegian electronic logbooks, in all registered fisheries including bycatch (A), and catches where G. halibut make more than $\mathbf{5 0 \%}$ of the total catches (B).


Figure 8.3. Spatial distribution of catches where Greenland halibut make more than $50 \%$ of the total catches, according to Norwegian electronic logbooks from 2021. Bubble area is proportional to the size of single catches expressed in metric tonnes. The panels show longline (A), gillnet (B) and trawl (C) catches.


Figure 8.4. NEA Greenland halibut. Total biomass estimates from Russian autumn survey and the Norwegian slope survey. Note that the Norwegian survey is run every other year since 2009. Uncertain estimate for 2013 from the Russian survey. Russian data from 1992 and onwards are revised in 2021 (Russkikh WD12). No Russian data for 2016, 2018 and 2020.


Figure 8.5. NEA Greenland halibut. Swept-area estimate of the female biomass based on the data from the Norwegian slope survey in August (every other year since 2009) and the Russian trawl survey in October-December (compared to previous reports, . Russian data from 1992 and onwards are revised in 2021 (Russkikh WD12)). Uncertain estimate for 2013 from the Russian survey.


Figure 8.6. Russian autumn survey; Greenland halibut abundance by sex (Russkikh and Smirnov, WD16 AFWG 2016). Russian data from 1992 and onwards are revised in 2021 (Russkikh WD12). In this figure the 1992, 1996, 2002, 2017 and 2019 indices were not raised to also represent uncovered parts of the standard survey area.


Figure 8.7. Estimated Greenland halibut total abundance in biomass and by number of individuals from the Norwegian slope surveys. The vertical bars show $95 \%$ confidence intervals.


Figure 8.8. Estimated Greenland halibut abundance (upper panel) and biomass (lower panel), by sex, from the Norwegian autumn slope survey.


Figure 8.9. Total juvenile biomass index (EcoJuv) (sex distribution is assumed 50/50 in the juvenile area so in the figure female biomass = male biomass) for Greenland halibut based on the Barents Sea Ecosystem Survey (A5216) (2014 not included due to poor survey coverage in the juvenile area) and the juvenile survey 1996-2002 (for area see Hallfredsson and Vollen, WD20 AFWG 2015).


Figure 8.10. Eco-south biomass index by sex for Greenland halibut in the Barents Sea Ecosystem Survey (A5216) , outside the juvenile area (for area see Hallfredsson and Vollen, WD20 AFWG 2015). The 2018 estimate is not considered reliable mainly due to lack in survey coverage, and was excluded from the 2021 assessment.


Figure 8.11. Joint Norwegian-Russian winter survey in the Barents Sea; Greenland halibut abundance and biomass estimates.


Figure 8.12. Length frequency distribution estimates for the entire area covered by the Norwegian Slope survey during autumn. Note biennial surveys after 2009.


Figure 8.13. Abundance and biomass estimates from Spanish autumn surveys (lower panel) (Muñoz et al., WD7 AFWG 2017), and abundance and biomass estimates from Spanish spring surveys (upper panel) (Muñoz et al., WD10 AFWG 2016). Note that $X$-axis is not continuous.


Figure 8.14. Biomass estimates from Polish spring survey (based on: Janusz et al., WD8 AFWG 2008; Janusz and Trella, WD10 AFWG 2009; Trella and Janusz, WD6 AFWG 2012). No update presented to the 2020 AFWG.


Figure 8.15. Dynamics of indices of the Barents Sea Greenland halibut stock in 1964-2015. Indices are divided by corresponding mean to put them in comparable scale. CPUE series divided in two, 1964-1991 and after 1996. In addition to the standardized CPUE three survey indices are shown; the Russian autumn survey (RUS), the Norwegian autumn survey (NOR) and the EcoSouth index (ECO).

## 9 Northeast Arctic anglerfish ${ }^{1}$

### 9.1 General

Our present knowledge of anglerfish (Lophius spp.) in ICES subareas 1 and 2 is based on two masters' theses (Staalesen, 1995; Dyb, 2003), a report from a Nordic project (Thangstad et al., 2006), working documents to the ICES ASC, WGNSDS, and WGCSE, and more recent catch data collected by the Norwegian Reference Fleet since 2006 (Anon., 2013; Clegg and Williams, 2021). In February 2018, anglerfish in ICES subareas 1 and 2 was subject to a benchmark assessment (WKANGLER 2018). After this benchmark assessment, it was determined that this stock (or rather a stock component and a management unit) is considered a category 3 stock, for which survey or other indices are available that provide reliable indications of trends in stock metrics, such as total mortality, recruitment, and biomass.

### 9.1.1 Species composition

Two European anglerfish species of the genus Lophius are distributed in the Northeast Atlantic: white (or white-bellied) anglerfish (Lophius piscatorius) and black (or black-bellied) anglerfish (Lophius budegassa). L. budegassa are rarely caught in Nordic waters. In Norwegian waters, 1 out of about 2600 anglerfish landed from the Møre coast north of $62^{\circ} \mathrm{N}(2 . a)$ and 1 out of about 1000 from the North Sea were L. budegassa back in 2003 (Dyb, 2003; K. Nedreaas, pers. comm.). In the most recent period (2014-2021), the ratio of L. budegassa in Norwegian waters has been up to 1 out of 200 anglerfish for some years, but usually about 1 out of 1000 .

### 9.1.2 Stock description and management units

The WGNSDS (Northern Shelf Demersal Stocks) considered the stock structure on a wider European scale in 2004, and found no conclusive evidence to indicate an extension of the stock area northwards to include Division 2.a. Anglerfish in 2.a have therefore been treated and described separately by the ICES Celtic Sea Ecoregion Working Group (WGCSE) who is now assessing the anglerfish in the neighbouring areas. Currently, anglerfish on the Northern Shelf are split into Subarea 6 (including 5.b (EC), 12 and 14) and the North Sea (and 2.a (EC)) for management purposes. However, genetic studies have found no evidence of separate stocks over these two regions (including Rockall) and particle-tracking studies have indicated interchange of larvae between the two areas and further towards ICES divisions 2.a, 5.a and 5.b (Hislop et al., 2001). So, previous working groups assessments have been made for the whole Northern Shelf area combined, but exclusive ICES divisions 2.a, 5.a and 5.b. In fact, both microsatellite DNA analysis (O'Sullivan et al., 2006) and particle tracking studies carried out as part of EC 98/096 also suggested that anglerfish from further south (Subarea 7) could also be part of the same stock. Hislop et al. (2001) simulated the dispersal of Lophius eggs and larvae using a particle tracking model. Their results also show the likelihood of Lophius around Iceland (Solmundsson et al., 2007), Faroe Islands (Ofstad, 2013) and Norwegian waters north of $62^{\circ} \mathrm{N}$ (i.e. subareas 1 and 2) are recruited from the area west of Scotland including Rockall. This is also supported by research survey data as a migration east-/north-eastwards with size is seen in the International Bottom Trawl Survey (IBTS) and other survey data (e.g. Dyb, 2003).

[^15]Results from the use of otolith shape analysis in stock identification of anglerfish (L. piscatorius) in the Northeast Atlantic (Cañás et al., 2012) and previous references on L. piscatorius stock identification find no biological evidence to support the current separation of Lophius stocks in the Northeast Atlantic, but find substructures within the area.

Anglerfish were tagged during two IBTS surveys in the North Sea and five one-day trips using a small ( 15 m ) Danish seiner off the Norwegian coast at around $62^{\circ} 40^{\prime} \mathrm{N}$ (Møre; Thangstad et al., 2006; Otte Bjelland, IMR-Norway, pers. comm.). A total of 872 individuals were tagged with conventional Floy dart type tags, 123 in the North Sea $(25-78 \mathrm{~cm})$ and 749 at Møre $(30-102 \mathrm{~cm})$. Some of this is further described in Thangstad et al. (2006). The 2019 AFWG report shows the tagging locations and the hitherto recaptures. There are migrations in all directions, i.e. recaptures from the southern North Sea, at the Shetland/Faroes and northwards to Lofoten. Most of the recaptures were done at Møre where most of the fish were tagged.

In 2000-2001 a total of 1768 trawl caught L. piscatorius was tagged using conventional dart tags and released on inshore fishing grounds at Shetland (Laurenson et al., 2005). Anglerfish between 25 and 83 cm total length were tagged. The overall recapture rate was $4.5 \%$ and times at liberty ranged from 5 to 1078 days. After Laurenson et al. (2005), Dr Laurenson reported to www.fishupdate.com a 104 cm anglerfish caught off the Norwegian coast near Ålesund in 2006. The fish had been tagged and released in the Scalloway Deeps on 13 September 2000 when it was 45 cm long and had hence been at liberty for five years and nine months. This is of particular importance as it may indicate a wider mixing of stocks and validate the growth rate of anglerfish.

WKANGLER (2018) considered that most recruitment in subareas 1 and 2 is from the more southerly stock unit, and this would require further R\&D work in collaboration with ICES 3.a, 4, and 6 looking at egg and larval dispersion and transportation as well as tagging and genetic studies. To address stock structure, mixing rates, and growth estimates, WKANGLER (2018) recommended a tagging program coordinated between all countries harvesting Lophius and to align tagging methods, measurement protocols and outreach to industry. The WK further recommended a shared site for Lophius tagging data and other applicable research projects concerning Lophius. Until the true biological stock structure is better understood, WKANGLER (2018) recommends keeping the anglerfish in subareas 1 and 2 as a separate management unit for the time being.

### 9.1.3 Biology

Sex ratios in Subarea 2 show that females outnumber males ( $>50 \%$ ) above approximately 75 cm , and above 100 cm all fish were females (Thangstad et al., 2006). This is very similar to the sex ratios reported from distant Portuguese and Spanish waters (Duarte et al., 1997) and hence supports a sex growth difference independent of latitude.

Spawning has been documented to occur in ICES Division 2.a in spring, but the present abundance of anglerfish in subareas 1 and 2 seems to be dependent on the influx or migration of juveniles from ICES subareas 4 and 6. Estimates of GSI (gonad-somatic index) for females in Division 2.a indicate that ovaries develop from January to June. The highest values of GSI were found in June when some of the ovaries were $20-30 \%$ of the round weight. Only females bigger than 90 cm had elevated GSI values indicating developing or developed ovaries. Dyb (2003) found that the length at which $50 \%$ of the females were mature (L50) was between $60-65 \mathrm{~cm}$ and that all females above 80 cm were mature.

Some age readings exist for anglerfish in Division 2.a, and comparative analyses of different structures, preparations and methods used for age readings were done by Staalesen (1995) and Dyb (2003). The Norwegian Institute of Marine Research adopted the ICES age reading criteria using the first dorsal fin ray (illicium) as its routine method, but few fish have been aged since
the above-mentioned projects. The material collected and read was, however, considered sufficient for preliminary yield-per-recruit estimations (ICES, 2019). As a very simplified 'rule of thumb' one may divide the fish length by 10 get an approximate age, i.e. a fish of 100 cm is approximately 10 years old and 13 kg while a fish of 70 cm is about 7 years old and 7 kg .

Exploitation using gillnets with 300 mm mesh size will select for males and females in a more equal ratio than 360 mm gillnets (Dyb, 2003). However, a change to lower mesh size will, without additional regulations, not decrease the effort, but rather increase it, at least towards younger fish. A mesh size of 300 mm will catch more anglerfish down to 50 cm , i.e. more immature fish. Preliminary analyses have also shown that the maximum yield-per-recruit will be $22 \%$ less using 300 mm instead of 360 mm gillnets (Staalesen, 1995). A possible sudden increase in catch rates when going from 360 mm to 300 mm would therefore be of short duration. A mesh size of 360 mm is also more in line with the minimum legal catch size of 60 cm , the length at first maturity of females and the utilization of the species' (especially the females') growth potential.

Some basic biological input parameters for the current assessment approaches are shown in Table 9.3. Some of these are further described in WKANGLER (2018).

### 9.1.4 Fishery

In autumn 1992 a direct gillnet fishery for anglerfish (L. piscatorius) started on the continental shelf in ICES Division 2.a off the northwest coast of Norway (Norwegian statistical area 07; Figure 9.1). The anglerfish had previously only been taken as bycatch in trawls and gillnets. Until 2010-2011 there was a geographical expansion of the fishery which was largely due to a northward expansion of the Norwegian gillnet fishery (Figure 9.2). It is not known to what extent this northwards expansion of the fishing area is caused by an expansion of favourable environmental conditions for the anglerfish or the fishers discovering new anglerfish grounds.

Near Iceland, Solmundsson et al. (2007) concluded that changes in the distribution of anglerfish and increased stock size have co-occurred with rising water temperatures that have expanded suitable grounds for the species. Another observed feature of the fisheries is that regional peaks in the landings of anglerfish representing northward migration become visible after multiple years of data collection (Figure 9.2). The recent increase in landings first happened along the coast of western Norway but did the last year expand to all subareas north of $62^{\circ} \mathrm{N}$ as well.

Norway is by far the largest exploiter of the anglerfish in subareas 1 and 2 accounting for 96$99 \%$ of the official landings (Table 9.1). The coastal gillnetting accounts for more than $90 \%$ of the landings (Table 9.2). The landings of anglerfish in subareas 1 and 2 have been about $1 / 4-1 / 3$ of the total landings from the other Northern Shelf areas (3.a, 4, and 6), but was in 2017 only 7\% of the total landings in these areas.

No TAC is given for subareas 1 and 2 of Norwegian waters. Catches of anglerfish in Division 2.a of the former European Union (EC) waters, now UK waters, are taken as a part of the combined EC/UK anglerfish quota for ICES areas 3, 4, and 6, or as part of the Norwegian 'others' quota in EC/UK waters. The Norwegian fishery is regulated through:

- A discard ban on anglerfish regardless of size.
- A prohibition against targeting anglerfish with other fishing gear than 360 mm (stretched mesh) gillnets.
- A minimum catch size of 60 cm in all gillnet fisheries, and maximum permission of $5 \%$ anglerfish (in numbers) below 60 cm when fishing with gillnets.
- $\quad 72$ hours maximum soak time in the gillnet fishery.
- A maximum of 500 gillnets (each net being maximum 27.5 m long) per vessel.
- Closure of the gillnet fishery from 1 March to 20 May. This closure period was expanded to 20 December- 20 May in the areas north of $65^{\circ} \mathrm{N}$ in 2008 and further expanded southwards to $64^{\circ} \mathrm{N}$ since 2009.
- A maximum of $15 \%$ bycatch (in weight) of anglerfish in the trawl- and Danish seine fisheries, and maximum $10 \%$ bycatch (in weight) of anglerfish in the shrimp trawl fishery. When fishing for argentines and Norway pout/sandeel a maximum of $0.5 \%$ bycatch is allowed within a maximum limit of 500 kg anglerfish per trip.
- A maximum of $5 \%$ bycatch (in weight) of anglerfish is allowed to be caught in gillnets targeting other species.


### 9.1.5 Scientific surveys

Anglerfish appear in demersal trawl surveys along the Norwegian shelf, but in very small numbers. The survey design has changed from single species to multispecies during recent years. The procedures for data collection on anglerfish have varied and, at present, no time-series from surveys in Division 2.a yields reliable information on the abundance of anglerfish. On the other hand, surveys in the North Sea and especially the SIAMISS (Scottish Irish Anglerfish Megrim Industry Science Survey; Figure 9.3), seem to be predictive for the recruitment of anglerfish to the ICES subareas 1 and 2 (Northeast Arctic). This is seen with the likely development of the large 2012 year class in the SIAMISS survey (Figure 9.4), which is corroborated with a subsequent decrease in mean catch length in Division 2a in 2017 and an increase in fishing effort at the same time.

The SIAMISS is a dedicated anglerfish survey (see ICES 2021). It covers much of the known distribution of the northern shelf anglerfish (ICES divisions 4a, 6a and 6b), with the exception of the central and southern parts of Subarea 4 and the Skagerrak and Kattegat (Division 3a). The survey began in 2005 and has been more or less carried out on an annual basis (usually in spring, but sometimes in November). The total biomass estimate for the Northern Shelf in 2021, the most recent survey year, was 48355 t . This is a decrease of $19 \%$ compared to 2019 (there is no 2020 estimate due to incomplete survey coverage) and the lowest value since 2013. A large proportion of total population numbers consisted of individuals $<30 \mathrm{~cm}$ in 2021, suggesting strong incoming recruitment (ICES 2021).

In Subarea 4, the International Bottom Trawl Surveys in the North Sea (indices NS-IBTS-Q1 and Q3) show declining mean weights per hour for the recent five years across all length groupings (ICES 2021). The IBTS surveys are currently not used in the assessment of anglerfish in ICES subareas 4 and 6, and in Division 3.a.

### 9.2 Data

### 9.2.1 Landings data

The official landings as reported to ICES for subareas 1 and 2 for each country are shown in Table 9.1. Landings decreased rapidly from 2010 to 2015, to the lowest since 1997, but has since shown an increase until last year. It is worth noting that the recent increase in landings first happened along the coast of western Norway, and then in the following years also subsequently further north in ICES Subarea 2. And likewise, the decrease seen in 2021 happened first in the south, i.e. both along the coast of western Norway and in the southern part of ICES Subarea 2 while the northern areas still showed an increase. Norway has by far the largest reported catches of the anglerfish in subareas 1 and 2, accounting for $96-99 \%$ of the official international landings. The coastal gillnetting accounts for more than $90 \%$ of the landings, of which about $90 \%$ are caught by the special designed large-meshed gillnets ( 360 mm stretched meshes; Table 9.2).

The Norwegian coastal reference fleet (see Appendix figure H1) provides us with length measurements and catch per gillnet days from ICES subareas through 4, from 2007-present and these have been presented for the AFWG in recent years. The catch rates vary spatially and temporally, and the WKANGLER (2018) therefore recommended to model and standardize the catch rates to better represent the general abundance trend of anglerfish in the entire ICES Subarea 2. The available material is shown in Tables 9.4 and 9.5 for the Norwegian statistical coastal areas (Figure 9.1) and total for ICES subareas 1 and 2.

### 9.2.2 Discards

The absence of a TAC in Norwegian waters probably reduces the incentive to underreport landings. Anecdotal evidence from the industry, observer trips and data from the self-sampling fleet (the Norwegian reference fleet; Anon. 2013; Clegg and Williams 2021) suggest that up to 8-9\% of the catch (not marketable) is discarded. This happens when the soaking time is too long, mostly due to bad weather. The average percentage of discarded anglerfish was higher south of $62^{\circ} \mathrm{N}$ (ICES 3 and 4) than north of $62^{\circ} \mathrm{N}$ (ICES 2.a). Average length of discarded anglerfish was on average only $6-7 \mathrm{~cm}$ smaller than the landed anglerfish. This is also confirmed by Berg and Nedreaas (2021) who estimated the annual discards of anglerfish by the Coastal reference fleet in subareas 1 and 2 to vary between 11 and 32 tonnes during 2014-2018 (i.e. 1.5-2.5\% of total gillnet catch) but went up to 178 tonnes (7.2\%) in 2012.

### 9.2.3 Length composition data

Length distributions are available from the directed gillnet fishery during the period 1992-2021, but data are lacking for 1997-2001 (Table 9.3). The length data indicates a drop in mean length of $15-20 \mathrm{~cm}$ occurring during the period without length samples (Figure 9.5). Since then, the mean length increased steadily during the last decade to about 95 cm (about 10 years old and 12 kg ) in 2014-2016, i.e. the same size level as seen during the 1990s. One-third of the anglerfish measured during the 1990s were above 100 cm , this proportion was between $1-6 \%$ for the early 2000 s, $12-17 \%$ in $2006-2013$ and $15 \%$ in 2021. This indicates strong recruitment into Subarea 2 during 1997-2001, which has not been observed again until 2017-2019 when a new drop in mean length is seen, again indicating some recruitment of smaller sized anglerfish to the area (ref. Figure 9.4).

Length distributions of retained anglerfish (L. piscatorius) caught by the reference fleet as target species during 2007-2021 by the specially designed-large-meshed gillnets, and as bycatch in other gillnets or other gears are shown in Appendix figures $\mathrm{H} 2-\mathrm{H} 4$. All subsequent analyses (in the methods and results section) have only used the length distributions from the target fishery since 2007 using the large-meshed gillnets which represent more than $80 \%$ of the international landings in subareas 1 and 2.

### 9.2.4 Catch per unit effort (CPUE) data

The Norwegian coastal reference fleet (see Appendix figure H1) has reported catch per gillnet soaking time (CPUE) from their daily catch operations. For the current modelling and hence standardization of the annual CPUE from subareas 1 and 2, we have used the following data:

- Only catch rates of retained anglerfish from the fishery using special large-meshed anglerfish gillnets (stretched meshes $=360 \mathrm{~mm}$ ).
- Years 2007-2021.
- Discards excluded.
- Adding zero catches where gillnets are used, but anglerfish not present.
- All coastal areas (i.e. ICES 3.a, 4.a, 2.a, and 1) included in the model since it is documented (e.g. WKANGLER 2018) that anglerfish are migrating across the ICES area borders.
- The area $\left(\mathrm{km}^{2}\right)$ of each subarea inside 12 nautical miles (covering most of the anglerfish distribution) is calculated and used as weighing factor when annual CPUEs are estimated for each subarea (Figure 9.6).


### 9.3 Methods and results

### 9.3.1 The length-based-spawning-potential-ratio (LBSPR) approach

The LBSPR method has been developed for data-limited fisheries, where only a few data are available: some representative sample of the size structure of the vulnerable portion of the population (i.e. the catch) and an understanding of the life history of the species (Hordyk et al., 2016). The LBSPR method does not require knowledge of the natural mortality rate ( M ) but instead uses the ratio of natural mortality and the von Bertalanffy growth coefficient ( K ; $\mathrm{M} / \mathrm{K}$ ), which is believed to vary less across stocks and species than M (Prince et al., 2015) although individual estimates of $M$ and $K$ can be used if available. Like any assessment method, the LBSPR model relies on a number of simplifying assumptions. In particular, the model is equilibrium-based, assumes that the length composition data are representative of the exploited population at steady state, and logistic selectivity (see the results section below for more discussion).

The LBSPR model originally developed by Hordyk et al. (2015a; 2015b) used a conventional agestructured equilibrium population model and a size-based selectivity. As a consequence, this approach could not account for "Lee's phenomenon" - the fact that larger specimens-at-age experience greater mortality than its cohort of smaller size because of the size-based selectivity. This is because the age-structured model has a 'regeneration assumption' i.e. it redistributes at each time-step the length-at-age using the same distribution. Hordyk et al. (2016) since developed a length-structured version of the LBSPR model that used growth-type-groups (GTG) to account for the above phenomenon and showed that the new approach reduced bias related to the "Lee's phenomenon" ${ }^{2}$. GTG LBSPR is therefore used for all subsequent analyses.

Some of the life-history parameters for the analysis were originally taken from WKANGLER (2018) but kept the same as in AFWG 2021. Hordyk et al. (2015a; 2015b) showed that the LBSPR approach was sensitive to the input parameters. We, therefore, drew 1000 random samples for each input parameter (i.e. from a bivariate normal distribution for Linf and K, a univariate normal distribution for M, L50, L95 (see Table 9.3)) and rerun the model in order to account for the effect of uncertainty around the input parameters on the results. We will refer to it as the "stochastic LBSPR approach" hereon.

Once the stochastic LBSPR runs were finished, we conducted some simulations through the LBSPR package to calculate some target SPR value. To do this, we used the mean input values from the stochastic LBSPR, the average estimated parameters values (from the stochastic LBSPR approach) and set the "steepness" to a value between 0.7 and 0.9 to perform a YPR analysis and determine the target reference points (which gives the maximum yield). Steepness values between 0.7 and 0.9 were chosen based on a literature search (values close to 1 are also found in the literature but were not included in the test as it seemed unrealistic for the species). The analysis gave target reference points of $\mathrm{SPR}=0.4$ (with $\mathrm{F} / \mathrm{M} \sim 1$ ) and $\mathrm{SPR}=0.25$ (with $\mathrm{F} / \mathrm{M} \sim 2$ ) for steepness values of 0.7 and 0.9 , respectively. What we obtained from the stochastic LBSPR runs instead are relatively stable annual estimates of SPR (between 0.15 and 0.5 (the IQR range)) and F/M

[^16](between 1.5 and 2.5; Figure 9.7). This suggests that—while there is a lot of uncertainty-fishing effort is probably slightly above but close to the effort that would lead to maximum yield.

The relationship between the biomass of reproductively mature individuals (spawning stock) and the resulting offspring added to the population (recruitment), the stock-recruitment relationship, is a fundamental and challenging problem in all population biology. The steepness of this relationship is the fraction of unfished recruitment obtained when the spawning-stock biomass is $20 \%$ of its unfished level. Steepness has become widely used in fishery management, where it is usually treated as a statistical quantity. If one has sufficient life-history information to construct a density-independent population model then one can derive an associated estimate of steepness (Mace and Doonan, 1988; Mangel et al., 2010; 2013).

As mentioned in the introduction, the LBSPR approach is an equilibrium-based method (i.e. assumes that the fishery experiences constant recruitment and F over time) and violation of this assumption can lead to biased SPR estimates. However, some management strategy evaluations conducted by Hordyk et al. (2015) on harvest control rules based on SPR-based size targets showed that while annual assessments of SPR may be imprecise due to the transitory dynamics of a population's size structure, smoothed trends estimated over several years may provide a robust metric for harvest control rules. SPR estimates in our study were relatively stable, thus large recruitment fluctuations may not be an issue.

### 9.3.2 CPUE standardization

Raw CPUE data are seldom proportional to population abundance as many factors (e.g. changes in fish distribution, catch efficiency, effort, etc) potentially affect its value. Therefore, CPUE standardization is an important step that attempts to derive an index that tracks relative population dynamics.

In the data preparation step, we quickly noticed that there was not enough data from ICES Subarea 1 to perform model inference. Therefore, we decided to omit data from this Subarea from the analyses. ICES Subarea 1 is the northern margin of L. piscatorius distribution, and only 3 tonnes were caught in this area in 2019, mostly as bycatch in other fisheries.

Below, we defined some important terms we used for the CPUE standardization:
Standardized effort (gillnet day) = gear count $x$ soaking time (hours)/24 hours
CPUE (per gillnet day) = catch weight/standardized effort
CPUE standardization was performed using the glmmTMB package (Brooks et al., 2017) and the best model was chosen based on AICc and residuals checks using the DHARMa package (Hartig 2021) i.e. the most parsimonious model had the lowest AICc while showing no problematic residuals pattern (i.e. overdispersion, underdispersion, etc). If problematic residual patterns were found, we tried to address the issue by either reconsidering the input data, changing model parameterization, or changing the model distribution assumption.

Using the model investigation/selection steps as in the last assessment (AFWG 2021), the final model was based on the Tweedie distribution. The Tweedie distribution belongs to the exponential family and its variance term is modelled as a power function of the mean $(\mu)$ i.e. $\varphi \mu$ p. This distribution is commonly used for generalized linear models (e.g. Jørgensen 1997) - with the following parameterization (for fixed and random effects):

$$
\begin{aligned}
& \text { CPUE }=\text { year }+ \text { subarea }+ \text { month }+(1 \mid \text { vessel })+(1 \mid \text { subarea_year })+(1 \mid \text { month_year }) \\
& +(1 \mid \text { month_subarea })
\end{aligned}
$$

The expression ( $1 \mid$ vessel) indicates that the vessel effect is considered a random effect and acts on the intercept. The expression (1 I month_year) indicates that the month and year variable was
concatenated into a single variable and considered as a random effect. In essence, this treatment models the interaction effect between year and month, but the approach only considers existing interaction (as opposed to all possible combinations of year and month which would be un-estimable) - which is an advantage in a data-limited situation such as ours.

Additionally, like the last two assessments (AFWG 2020, 2021), data were filtered to keep only vessels that had more than 10 observations (as these rare vessel observations were causing deviations in the residual patters). Using the 10-minimum-observations criteria improved the residual pattern of the model but was not able to eliminate the residuals pattern (Figure 9.8). Such residual pattern started to appear in the last assessment (AWFG 2021) - though much less pronounced, thus not investigated - but was absent from the 2020 assessment (AFWG 2020). Therefore, another type of modelling approach, namely a delta model (or hurdle model), was briefly examined in this study to possibly correct for the observed residual pattern.
A delta model consists of a pair of models: one that models the species occurrence (presence/absence) and another that models the positive values.

```
Presence \(=(1 \mid\) year \()+\) subarea + month \(+(1 \mid\) vessel \()+(1 \mid\) subarea_year \()+\)
(1 I month_year) + (1 I month_subarea)
CPUE_pos = year + subarea + month + (1 | vessel) + (1 | subarea_year) + (1 | month_-
year) + (1 | month_subarea)
```

Anglerfish occurrence was modelled using a binomial model with logit transform and positive CPUE was modelled using a Gamma distribution with log link. All variables were kept the same as in the original Tweedie model except for the year effect in the occurrence model that was converted to a random effect due to some estimability issue. The delta model specification helped improve the residual behaviour but did not completely eliminate the pattern (Figure 9.8).

For all subsequent analysis, we therefore examined the sensitivity of the model results to using the delta model as CPUE standardization approach.

As in the AFWG 2021, the standardized annual CPUE index was created by summing up all predictions based on all possible combination of year (2007-2021), subarea (in ICES Area 2.a), and month (1-12) after weighting the prediction for each subarea by its surface (in $\mathrm{km}^{2}$ within the 12 nautical miles as shown in Figure 9.6) relative to the total surface (sum of all subarea surfaces in the ICES area 2a). In this process, we removed the vessel random effect (assuming it equals 0 , the mean value) as it only affects catch efficiency and does not represent the underlying fish abundance. We note that glmmTMB can handle any missing new levels for random effect variables when making prediction (it assumes it is equal to zero and inflates the prediction error by its associated random effect variance). The standard deviation of the summed prediction (for the original Tweedie model) was directly calculated in glmmTMB by modifying the source code ('glmmTMB.cpp’ file).

A similar approach was taken for the delta model to derive an abundance index except that model predictions were manually calculated while accounting for the covariance in fixed effect estimates. More precisely, fixed effect parameters were resampled 100,000 times based on their estimated mean and covariance for both components of the delta model (while random effect were kept at their MLE). These values were then used to predict the probability of occurrence and CPUE value for all combination of year, subarea, and month (as above) that were then multiplied together to calculate the expected CPUE. The final index was then calculated in a similar approach as in the original Tweedie model by weighted average of the predictions by area.

Figure 9.9 shows that anglerfish population in ICES Subarea 2a might have declined over the last decade (as well as the raw effort) but could be increasing again in more recent years. Nonetheless, there is a lot of year-to-year variability and uncertainty around the point estimates.

### 9.3.3 JABBA

JABBA stands for 'Just Another Bayesian Biomass Assessment' and is open-source modelling software that can be used for biomass dynamic stock assessment applications. It has emerged from the development of a Bayesian State-Space Surplus Production Model framework applied in stock assessments of sharks, tuna, and billfish around the world (Winker et al., 2018). JABBA requires a minimum of two input comma-separated value files (.csv) in the form of catch and abundance indices (and SE; see Appendix table H1). The Catch input file contains the time-series of year and catch by weight, aggregated across fleets for the entire fishery. Missing catch years or catch values are not allowed. JABBA is formulated to accommodate abundance indices from multiple sources (i.e. fleets) in a single CPUE file, which contains all considered abundance indices. The first column of the CPUE input is year, which must match the range of years provided in the Catch file. In contrast to the Catch input, missing abundance index (and SE) values are allowed.

The catch data comes from the different fishing countries' official reporting of annual landings to ICES (see Table 9.1) and the CPUE data (along with its standard deviation) comes from the CPUE standardization process described above with values in 1992-1994 imputed based on expert knowledge. We assumed that the CPUE index from ICES Subarea 2.a calculated using data from the anglerfish targeted fishery is representative of the stock status in ICES areas 1 and together.

In addition to these .csv files, JABBA also requires users to define the prior distribution for the model parameters which will be subsequently updated with data to form the posterior distributions (Figure 9.10). In addition to the base case, 10 additional scenarios were run to examine the sensitivity of the model results to the choice of priors (Table 9.6).

Figure 9.11 shows the trajectory of the population estimates from 1990-2021 based on the tested scenarios (Table 9.7). In general, population abundance has never fallen below $\mathrm{B}_{\mathrm{MSY}}$ (at least the mean trajectory) but fishing mortality fluctuated above and below the Fmsy (Figure 9.12). Figure 9.13 is the Kobe plot from the base model run showing the estimated trajectories of $\mathrm{B} / \mathrm{Bmš}_{\text {m }}$ and F/FmsY along with the credibility intervals of the 2021 estimates of biomass and fishing mortality. The percentage numbers at the top right indicate how much of the 2021 population estimates that fall within the green (not overfished, no overfishing), yellow (overfished, but no overfishing), orange (overfishing, but not overfished), and red (overfished and overfishing) zones, after accounting for all the parameter uncertainty (basically, the area under the oval-shaped density plot that falls into each coloured quadrant). The model estimates that there is roughly a $15 \%$ ( $23 \%$ ) probability that the 2021 population estimate falls within the red zone, $30 \%$ ( $22 \%$ ) in the orange, $0.5 \%(2 \%)$ in the yellow, and $54.5 \%(53 \%)$ in the green zone (in parentheses corresponding percentages from last year's assessment). Finally, retrospective analysis on the base model run was slightly worse than the previous assessment cycle (AFWG 2021) overall (especially for F/FMSY; Figure 9.13, Table 9.7). In general, the error in parameter estimates was consistent (same sign) across peel $2-5$ but changed sign in peel 1 (Table 9.7).

The sensitivity analysis says that MSY could be around or slightly above 2000 tonnes, with a BMSY ~ 30000 tonnes (Figure 9.14). Though the MSY value is quite sensitive to the choice of prior on $r=$ population growth rate, which makes sense if population grows slowly, one cannot fish too hard, i.e. lower MSY.

However, the retrospective analysis (Figure 9.15) also shows that the estimate of MSY could be influenced by the addition of 1 year of data, i.e. the scaling of $\mathrm{F} / \mathrm{F}_{\mathrm{MSY}}$ is not very steady across time, and the figure suggests that it could be a bit lower, maybe between 1500-2000 t . Though the Bmsy still stays around ~30 000 tonnes. So an initial guestimate of MSY would be somewhere between 1500-2000 t . MSY of 1500 t was also the MSY estimate based on the low r scenario.

### 9.4 Management considerations and future investigations

The present abundance of anglerfish in subareas 1 and 2 seems to depend on the influx or migration of juveniles from ICES subareas 4 and 6 . It is therefore expected that an effective discard ban on anglerfish in subareas 4 and 6 will have a positive effect on the abundance north of $62^{\circ} \mathrm{N}$. Reduced mean size of the landed anglerfish in recent years (fishing with the same large-meshed gillnets) indicates a new influx of recruitment to the ICES subareas 1 and 2. Monitoring of the fishery will be important in the near future to protect the young specimens from recruitmentand growth- overfishing.
AFWG has previously recommended that the anglerfish stock component in subareas 1 and 2 is annually monitored and a $20 \%$ reduction in fishing effort per year (also as an uncertainty cap) should be imposed until the decrease in CPUE is stopped. Despite that the decrease in CPUE has stopped for time being, the current exploratory assessment shows that there is nothing to gain in increasing effort. The "2-over-3" rule used on the CPUE time-series, including both an uncertainty cap and a precautionary buffer, suggest a $10 \%$ reduction in effort or catch advice for 2023.

The standardized CPUE analysis shows that anglerfish population in ICES Subarea 2.a has declined over the last decade (as well as the raw effort) with an increase in the most recent year.

The spawning potential ratio, as calculated by the LBSPR method using input biological parameters and the estimated exploitation parameters suggests that-while there is a lot of uncertainty - fishing effort is probably slightly above but close to the effort that would lead to maximum yield.

The relative population stock status is around Bmsy, though fishing intensity could be close or slightly higher than Fmsy. Therefore, effort should not be increased at the risk of the population falling below the biomass and SPR targets. New promising recruitment seen from scientific survey in the northern North Sea is expected to contribute to the fishery north of $62^{\circ} \mathrm{N}$ after 3-4 years if it continues to develop as the observed recruits seen in the 2013-2014 surveys did. The quality of the current assessment was this year further evaluated by analysing another type of CPUE modelling approach (the Delta model), to possibly correct for an observed residual pattern. The AFWG considers the current assessment of sufficient quality to base catch advice on for subareas 1 and 2.

When it comes to reference points, it should be further discussed if and which defined values of F/M, F/Fmsy, SPR and B/Bmsy may be used.

Any potential harvest control should take account of both recruitment- and growth- overfishing. LBSPR provides measures for both, F/M and SPR, with the SPR values being the transient SPR and thus an estimate of current stock status. While maximum sustainable catch is often a key management objective, it may not be the only one. In that case, it may be worth modifying a reference point to reflect other management objectives.

The AFWG supports that ICES subareas 1, 2, 3, 4, and 6 should be investigated together to get a more complete understanding of migrations and distributions.

### 9.5 Tables and figures

Table 9.1. Nominal catch ( t ) of anglerfish in ICES subareas 1 and 2, 2001-2021, as officially reported to ICES.

|  | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DK | 2 | + | - | 1 | - | - | - | - | + | - | - | - | - | - | - | - | - | - | - | - | - |
| Faroes | 1 | 1 | 2 | 5 | 11 | 4 | 7 | 4 | 2 | 1 | + | + | 1 | + | + | 1 | 1 | + | + | 1 | - |
| France | - | - | - | - | - | 1 | - | - | - | - | 1 | 3 | 2 | - | 4 | 2 | 4 | 3 | 8 | 5 | 4 |
| Germany | 65 | 59 | 55 | 70 | 55 | + | + | 0 | + | 82 | 70 | 0 | - | + | + | + | 1 | 1 | 50 | - | - |
| Iceland | - | - | - | - | - | - | - | - | - | - | 7 | - | - | - | - | - | - | - | - | - | - |
| Norway | 3554 | 2000 | 2405 | 2907 | 2650 | 4257 | 4470 | 4007 | 4298 | 5391 | 5031 | 3758 | 2988 | 1655 | 933 | 1355 | 1473 | 1884 | 2750 | 2258 | 2584 |
| Portugal | - | - | - | - | - | - | - | 2 | 6 | 1 | + | - | - | - | - | - | - | - | - | - | - |
| UK | 2 | 11 | 15 | 18 | 19 | 86 | 114 | 138 | 152 | 40 | 3 | 3 | 111 | 2 | 105 | 76 | 5 | 15 | + | 16 | 13 |
| Others |  |  |  |  |  |  |  |  |  |  |  | 1 | 1 | - | - | + | - | + | - | - | - |
| Total | 3624 | 2071 | 2477 | 3001 | 2735 | 4348 | 4591 | 4151 | 4458 | 5515 | 5112 | 3765 | 3103 | 1657 | 1043 | 1435 | 1484 | 1903 | 2809 | 2280 | 2601 |

*Preliminary.

Table 9.2. Anglerfish in ICES subareas 1 and 2. Norwegian landings (tonnes) by fishery in 2008-2021. The coastal area is here defined as the area inside $\mathbf{1 2}$ nautical miles from the baseline.

| Fleet NORWAY | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Coastal gillnet | 3574 | 3934 | 4806 | 4557 | 3521 | 2758 | 1506 | 829 | 1231 | 1320 | 1727 | 2502 | 1939 | 2233 |
| Offshore gillnet | 240 | 171 | 391 | 319 | 115 | 158 | 95 | 52 | 62 | 87 | 68 | 153 | 168 | 229 |


| Danish seine | 75 | 68 | 40 | 26 | 16 | 19 | 11 | 12 | 17 | 23 | 28 | 26 | 35 | 78 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Demersal trawl | 34 | 36 | 48 | 19 | 11 | 8 | 7 | 3 | 5 | 6 | 10 | 5 | 3 | 2 |
| Other gears | 84 | 89 | 106 | 83 | 96 | 45 | 36 | 37 | 40 | 31 | 51 | 64 | 113 | 42 |
| Total | 4007 | 4298 | 5391 | 5031 | 3759 | 2988 | 1655 | 934 | 1355 | 1468 | 1884 | 2750 | 2258 | 2584 |

*Preliminary per 5 April 2022.

Table 9.3. Basic input parameters and parameters for resampling as used for the LBSPR analysis.

| Basic input parameters | Value |
| :---: | :---: |
| von Bertalanffy K parameter (mean) | 0.12 |
| von Bertalanffy Linf parameter (mean) | 146 |
| von Bertalanffy t0 parameter | -0.34 |
| Length-weight parameter a | 0.149 |
| Length-weight parameter b | 2.964 |
| Steepness | 0.8 |
| Maximum age | 25 |
| Length at 50\% maturity (L50; mean) | 82 |
| Length at 95\% maturity (L95; mean) | 100 |
| $\Delta \mathrm{Mat}=\mathrm{L} 95-\mathrm{L} 50$ (mean) | 18 |
| Length at first capture | 40 |
| Length at full selection | 60 |
| M (mean) | 0.2 |


| Basic input parameters | Value |
| :--- | :---: |
| $M / \mathrm{k}$ (mean) | 1.67 |
| Parameters for resampling | Value |
| $\mathrm{N}_{\text {samp }}$ | 1000 |
| CV(M) | 0.15 |
| Cor (Lin_K) | 0.9 |
| CV(K) | 0.3 |
| CV(Linf) | 0.15 |
| $C V(L 50)$ | 0.05 |
| $C V(\Delta M a t)$ | 0.05 |

Table 9.4. Number of coastal reference fleet fishing days with anglerfish, per national stat. subareas (0-7) and total for ICES subareas 1 and 2. Only large-meshed gillnets included.

| Year/ area | $\mathbf{0}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | ICES 1 and 2 |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2007 | 106 | 26 |  | 280 | 412 |
| 2008 | 62 | 37 | 6 | 171 | 276 |
| 2009 | 86 | 35 | 36 | 176 | 333 |
| 2010 | 14 | 41 | 37 | 143 | 235 |
| 2011 | 64 | 19 | 51 | 116 | 250 |
| 2012 | 49 | 12 | 24 | 21 | 106 |
| 2013 | 64 | 20 | 18 | 81 | 183 |
| 2014 | 5 |  | 19 | 107 | 131 |
| 2015 | 109 |  | 5 | 116 | 230 |
| 2016 | 92 |  | 22 | 35 | 149 |
| 2017 | 88 |  |  | 109 | 197 |
| 2018 | 108 |  |  | 89 | 197 |
| 2019 | 86 | 34 |  | 63 | 183 |
| 2020 | 74 | 28 | 52 | 104 | 258 |
| 2021 | 66 |  | 72 | 78 | 216 |

Table 9.5. Number of fishing days with length measured anglerfish (left) and number of length measured fish (right). Only large-meshed gillnets included.

| Year | ICES 1 and 2 |
| ---: | ---: |
| 2007 | 78 |
| 2008 | 43 |
| 2009 | 47 |
| 2010 | 67 |
| 2011 | 78 |
| 2012 | 39 |
| 2013 | 52 |
| 2014 | 29 |
| 2015 | 31 |
| 2016 | 45 |
| 2017 | 74 |
| 2018 | 64 |
| 2019 | 50 |
| 2020 | 83 |
| 2021 | 75 |


| Year | ICES 1 and 2 |
| ---: | ---: |
| 2007 | 2265 |
| 2008 | 1407 |
| 2009 | 2325 |
| 2010 | 2171 |
| 2011 | 2423 |
| 2012 | 995 |
| 2013 | 1305 |
| 2014 | 546 |
| 2015 | 1063 |
| 2016 | 654 |
| 2017 | 1593 |
| 2018 | 1451 |
| 2019 | 1486 |
| 2020 | 2149 |
| 2021 | 1608 |

Table 9.6. Eleven scenarios were run to examine the sensitivity of the model results to the choice of priors.

| Scenario name | K | $\boldsymbol{r}$ | $\boldsymbol{\sigma}_{\mathbf{P}}$ | Initial depletion | B $_{\text {MSY }} /$ K value |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Base | $\mathrm{LN}(1 \mathrm{e} 6,1)$ | $\mathrm{LN}(0.1,1)$ | $\mathrm{IG}(4,0.01)$ | $\mathrm{LN}(0.8,0.5)$ | 0.35 |
| Low_K | $\mathrm{LN}(5 \mathrm{e} 5,1)$ | $\mathrm{LN}(0.1,1)$ | $\mathrm{IG}(4,0.01)$ | $\mathrm{LN}(0.8,0.5)$ | 0.35 |


| High_K | LN(1.5e6,1) | LN(0.1,1) | $\mathrm{IG}(4,0.01)$ | LN(0.8,0.5) | 0.35 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Low_r | LN(1e6,1) | LN(0.05,1) | $\mathrm{IG}(4,0.01)$ | LN(0.8,0.5) | 0.35 |
| High_r | LN(1e6,1) | LN(0.2,1) | IG(4,0.01) | LN(0.8,0.5) | 0.35 |
| Low_sigmaP | LN(1e6,1) | LN(0.1,1) | IG(4,0.005) | LN(0.8,0.5) | 0.35 |
| High_sigmaP | LN(1e6,1) | LN(0.1,1) | $\mathrm{IG}(4,0.02)$ | LN(0.8,0.5) | 0.35 |
| Low_initdep | LN(1e6,1) | LN(0.1,1) | $\mathrm{IG}(4,0.01)$ | LN(0.7,0.5) | 0.35 |
| High_initdep | LN(1e6,1) | LN(0.1,1) | IG(4,0.01) | LN(0.9,0.5) | 0.35 |
| Low_BmsyK | LN(1e6,1) | LN(0.1,1) | IG(4,0.01) | LN(0.8,0.5) | 0.30 |
| High_BmsyK | LN(1e6,1) | LN(0.1,1) | IG(4,0.01) | LN(0.8,0.5) | 0.40 |

*LN stands for lognormal and IG stands for inverse gamma distribution. $\mathrm{B}_{\mathrm{MS} \mathrm{\gamma}} / \mathrm{K}$ value controls for the position of the inflection point of the surplus production curve with respect to $K$ (a value from $o 1$ ).

Table 9.7. Relative error (RE) in parameter estimates between the base run with full dataset (Table 9.6) and the retrospective peels ( 1 to 5 years) and the associated Mohn's rho statistics (i.e. average RE from the 5 peels). Relative error is calculated as: RE = (peel-ref)/ref.

|  | B | F | B/BMsy | F/FMSY | B/B0 | MSY |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| RE_peel1 | -0.031 | 0.032 | 0.063 | -0.025 | 0.063 | -0.046 |
| RE_peel2 | -0.142 | 0.166 | -0.211 | 0.779 | -0.211 | -0.324 |
| RE_peel3 | -0.112 | 0.127 | -0.288 | 0.855 | -0.288 | -0.277 |
| RE_peel4 | -0.122 | 0.139 | -0.084 | 0.295 | -0.084 | -0.173 |
| RE_peel5 | -0.092 | 0.102 | -0.117 | 0.186 | -0.117 | -0.074 |
| Mohn's rho | -0.100 | 0.113 | -0.127 | 0.418 | -0.127 | -0.179 |



Figure 9.1. Map showing the Norwegian statistical coastal areas. Area 03 is part of ICES Subarea 1; areas 04, 05, 00, 06, and 07 are part of ICES Subarea 2; Areas 28 and 08 are part of ICES Subarea 4, and Area 09 corresponds roughly with ICES Subarea 3.


Figure 9.2. Norwegian official landings (in tonnes) of anglerfish (Lophius piscatorius) per statistical area (see Figure 9.1) within ICES areas 1 and 2 during 1992-2021. Norwegian landings from the area south of $62^{\circ} \mathrm{N}$ (ICES 4 and 3 ) are shown for comparison.


Figure 9.3. Excerpt from WGCSE 2021: A) WGCSE 2021 figure 4.16 - Numbers of anglerfish per km2 observed by SIAMISS surveys 2021. B) WGCSE 2021 figure 4.17-Weight of anglerfish (kg) per km2 observed by SIAMISS surveys 2021.


Figure 9.4. Excerpt from WGCSE 2021: Figure 4.8. SIAMISS-Q2 estimates of total numbers (millions) at-length (cm) for subareas 4.a (blue)-c and 6.a (yellow)-b (red) combined, 2012-2021.


Figure 9.5. Anglerfish (Lophius piscatorius) in ICES Subareas 1 and 2. Mean lengths for anglerfish caught in the directed coastal gillnetting in Division 2.a during 1992-2021, dotted lines represent $\pm 2$ SE of the mean. Note that data are lacking for 1997-2001. This illustrates pulses of new recruitment entering Division 2.a from ICES subareas 4 and 6; last time during 2002-2003, and to a lesser extent in 2017-2019.


Figure 9.6. Map showing the area (km2) of each Norwegian statistical subarea inside $\mathbf{1 2}$ nautical miles. The subareas 4, $5,0,6$, and 7 belong to the ICES Division 2.a.


Figure 9.7. Annual estimates of $F / M$ (above) and SPR (below) from the stochastic LBSPR approach using the length composition data from 2006-2021.


Figure 9.8. CPUE model residual diagnostics. Top panel shows the residual pattern in the original Tweedie model using the latest data. Bottom panel shows the results from the new delta model.


Figure 9.9. Standardized CPUE (kg per gillnet day) +/-SD (solid black line with error bars for the original Tweedie model, and solid red line with error bars for the new delta model) and the corresponding standardized effort (dash line) for anglerfish based on the data from the Norwegian coastal reference fleet in ICES Subarea 2a, from vessels targeting anglerfish with large meshed gillnets.


Figure 9.10. Prior and posterior distributions of the JABBA model parameters for the anglerfish assessment.


Figure 9.11. Estimated trajectories for $B / B_{\text {MSY }}$ for the ICES subareas 1 and 2 anglerfish based on 11 JABBA scenarios (the name of scenario and the associated colour is indicated in the figure). The lines show the mean trajectory and the shaded areas denote $95 \%$ credibility intervals.


Figure 9.12. Estimated trajectories for F/F MSY for the ICES subareas 1 and 2 anglerfish based on 11 JABBA scenarios (the name of scenario and the associated colour is indicated in the figure). The lines show the mean trajectory and the shadedareas denote $95 \%$ credibility intervals.

9.13. Kobe plot for the JABBA base case scenario showing the estimated joint trajectories (1990-2021) of $B / B_{\text {msy }}$ and F/F ${ }_{\text {Msr }}$. Different grey shaded areas denote the $50 \%, 80 \%$, and $95 \%$ credibility interval for the terminal assessment year. The probability of terminal year points falling within each quadrant is indicated in the figure legend. The figure on the left shows the results using the original Tweedie model when calculating the abundance index while the figure on the right uses the index derived from the delta model.


Figure 9.14. Sensitivity analysis for the ICES subareas 1 and 2 anglerfish based on 11 JABBA scenarios (the name of scenario and the associated colour is indicated in the figure). The analysis says that MSY could be around 2000 tonnes, with $B_{\text {MSY }} \sim 30000$ tonnes. Note that the MSY value is quite sensitive to the choice of prior on $r=$ population growth rate.


Figure 9.15. Retrospective analysis from the JABBA base case scenario. Different colours illustrate the results from different peels.


Figure 9.16. Catch per unit effort for five boats in the gillnet fishery for anglerfish in Møre and Romsdal (the same area as vessel A in figure 8 is fishing in) in the period October 1992 to October 1994. Boat $1>25 \mathrm{~m}$; Boat 2 ca. 20 m ; Boat 3 ca. 10 m ; Boat 4 and 5 ca. 16 m . Boats 1-4 were fishing with gillnet $\mathbf{3 6 0} \mathrm{mm}$ nesh size, boat 5 with $\mathbf{3 0 0} \mathrm{mm}$ mesh size.

## Appendix figure H 1.



Appendix table H1. Input data to the JABBA assessment in the form of catch and abundance indices of anglerfish ( $L$. piscatorius) in ICES subareas 1 and 2.

| Year | Catch | CPUE (mean) | CPUE (SE) |
| :---: | :---: | :---: | :---: |
| 1990 | 151 |  |  |
| 1991 | 180 |  |  |
| 1992 | 488 | 1.5 | 0.3 |
| 1993 | 3042 | 1 | 0.2 |
| 1994 | 1024 | 0.5 | 0.1 |
| 1995 | 526 |  |  |
| 1996 | 887 |  |  |
| 1997 | 601 |  |  |
| 1998 | 1549 |  |  |
| 1999 | 1743 |  |  |
| 2000 | 2999 |  |  |
| 2001 | 3624 |  |  |
| 2002 | 2071 |  |  |
| 2003 | 2477 |  |  |
| 2004 | 3001 |  |  |
| 2005 | 2735 |  |  |
| 2006 | 4348 |  |  |
| 2007 | 4591 | 0.52 | 0.08 |
| 2008 | 4151 | 0.56 | 0.07 |
| 2009 | 4458 | 0.51 | 0.06 |
| 2010 | 5515 | 0.45 | 0.05 |
| 2011 | 5112 | 0.47 | 0.06 |
| 2012 | 3765 | 0.45 | 0.06 |
| 2013 | 3103 | 0.33 | 0.04 |
| 2014 | 1657 | 0.41 | 0.06 |
| 2015 | 1043 | 0.41 | 0.07 |
| 2016 | 1435 | 0.32 | 0.05 |
| 2017 | 1484 | 0.29 | 0.06 |


| Year | Catch | CPUE (mean) | CPUE (SE) |
| :--- | :--- | :--- | :--- |
| 2018 | 1903 | 0.36 | 0.08 |
| 2019 | 2809 | 0.30 | 0.05 |
| 2020 | 2280 | 0.48 | 0.06 |
| 2021 | 2601 | 0.42 | 0.06 |

Appendix figure H2. Length distributions of anglerfish (L. piscatorius) caught and retained in large-meshed gillnets per year and Norwegian statistical areas. Areas 0,5, 6 and 7 represent ICES Subarea 2. Note the different scale of the $y$-axis in App. figs H2-H4.


Appendix figure H3. Length distributions of anglerfish (L. piscatorius) caught as bycatch and retained in other gillnets per year and Norwegian statistical areas. Note the different scale of the y -axis in App. figs H 2 - H 4 .


Appendix figure H4. Length distributions of anglerfish (L. piscatorius) caught as bycatch and retained in other gears per year and Norwegian statistical areas. Note the different scale of the $\mathbf{y}$ axis in App. figs $\mathrm{H} 2-\mathrm{H} 4$.


## 10 Barents Sea capelin ${ }^{1}$

### 10.1 Regulation of the Barents Sea capelin fishery

Since 1979, the Barents Sea capelin fishery has been regulated by a bilateral fishery management agreement between Russia (former USSR) and Norway. A TAC has been set separately for winter fishery and for autumn fishery. From 1999, no autumn fishery has taken place, except for a small Russian experimental fishery in some years. A minimum landing size of 11 cm has been in force since 1979. AFWG strongly recommends capelin fishery only on mature fish during the period from January to April.

### 10.2 TAC and catch statistics (Table 10.1)

The Joint Russian-Norwegian Fishery Commission set a zero TAC for 2021 and a TAC of 70000 tonnes for 2022. For both years, the quotas were in accordance with the ICES advice. The international historical catch by country and season in the years 1965-2022 is given in Table 10.1. The Norwegian catch in 2022 was 42346 tonnes which was 396 tonnes above the national TAC. Russian catches were 22646 tonnes which was 5404 tonnes below the national TAC.

The age-length distribution of Russian catches in 2022 is summarized in the text table below (age distribution (\%) for each length group and catch at length in numbers).

Norwegian age-length composition of 2022 catches were not ready at the time of AFWG.
The capelin sampling from the Barents Sea in 2021-2022 is summarized below:

| Investigation | No. of trawl hauls | Length measurements | Aged <br> individuals |
| :--- | :--- | :--- | :--- |
| Winter capelin survey 2022 (Norway) | 25 | 2383 | 978 |
| Joint Winter survey 2022 (Norway) | 292 | 10859 | 1059 |
| Joint Winter survey 2022 (Russia) | 97 | 5759 | 200 |
| IESNS 2021 (Russia) | 12 | 22261 | 6221 |
| BESS 2021 (Norway) | 339 | 15255 | 1103 |
| BESS 2021 (Russia) | 195 |  | 202 |

[^17]
### 10.3 Stock assessment

### 10.3.1 Acoustic stock size estimates in 2021 (Table 10.2, Figure 10.1, 10.2 and 10.3)

The geographical survey coverage of the Barents Sea capelin stock during the BESS in 2021 was almost complete (Figure 10.1). However, as last year, an area in the central part of the Barents Sea ("Loophole") was not covered.

The geographical distribution of capelin in 2021 is shown in Figure 10.1, and the position and weighting of the trawl stations is shown in Figure 10.2.

The stock estimate from the area covered by the 2021 survey was 3.998 million tonnes (Table 10.2). About $36 \%$ ( 1.438 million tonnes) of the estimated stock biomass consisted of maturing fish $(>14.0 \mathrm{~cm})$. The mean weight at age in the 2021 survey was the lowest since 2014 for age 2 (Figure 10.3).

As decided during the 2016 assessment meeting, the capelin abundance was estimated using the software StoX (Johnsen et al. 2019), applying agreed settings.
A fixed sampling variance expressed as Coefficient of Variance (CV) of 0.2 per age group has been applied as input for CapTool in the capelin assessment and was also used this year (Tjelmeland 2002; Gjøsæter et al. 2002). The survey design and estimation software now allow for estimation of a direct CV by age group, and for the 2021 survey this was estimated:

- for age group 1: 0.17;
- for age group 2: 0.10;
- and for age group 3: 0.29.

These values are lower than previous years for age group 1 and 2 and similar for age group 3 . Relative sampling error based only on acoustic recordings (Nautical Area Scattering Coefficient (NASC; m² $\mathrm{nmi}^{-2}$ )) was estimated to $9.27 \%$ which is much lower than in the two previous years. Detailed information about previous CV estimates can be found in AFWG WD5, 2018. Future implementation of direct survey CV in the assessment is discussed under future work (10.4.6).

### 10.3.2 Stock assessment in 2021 (Table 10.3-10.5, Figure 10.4)

Probabilistic projections of the maturing stock to the time of spawning on 1 April 2022 were made using the spreadsheet model CapTool (implemented in the @RISK add-on for EXCEL, 50000 simulations were used). The settings were the same as last year. The projection was based on a maturation and predation model with parameters estimated by the model Bifrost and data on cod abundance and size at age in 2022 from the 2021 Arctic Fisheries Working Group (ICES Scientific Reports. 2.52). The revised cod assessment made in September 2021 was used.

The methodology is described in the 2009 WKSHORT report (ICES C.M. 2009/ACOM:34) and the WKARCT 2015 report (ICES C.M. 2015/ACOM:31). The natural mortality M for the months October to December is drawn among a set of M-values estimated for different years based on historical data. The same set of M-values was used in 2021 as in 2020 (ICES 2011/ACOM:05, Annex 12).

The CapTool forecast methodology has been implemented in the R package Bifrost and was run alongside the standard procedure. The results were similar, and it produced the same advice.

With no catch, the estimated median spawning stock size on 1 April 2022 is 479000 tonnes ( $90 \%$ confidence interval: 259 000-916 000 tonnes), and the probability for the spawning stock to be above $\operatorname{Blim}(200000 \mathrm{t})$ is $99 \%$.

With a catch of 70000 tonnes, the probability for the spawning stock in 2022 to be below 200000 t , the Blim value used by ACFM in recent years, is $5 \%$ (Figure 10.4). The median spawning stock size in 2022 will then be 420000 tonnes ( $90 \%$ confidence interval: $200000-833000$ tonnes), and the corresponding median modelled consumption by immature cod in the period JanuaryMarch 2022 will then be 570000 tonnes. Figure 10.4 shows the probabilistic forecast from 1 October 2021 to 1 April 2022 conditional on a quota of 70000 tonnes, while Figure 10.5 shows the probability of SSB $<$ Blim as a function of the catch.

As in previous years, the catch corresponding to $95 \%$ probability of being above Blim was calculated to the nearest 5000 tonnes.

Estimates of stock in number by age group and total biomass for the historical period are shown in Table 10.4. Other data which describe the stock development are shown in Table 10.5. Information about spawning surveys going back to the 1980s are given in Gjøsæter and Prozorkevitch (WD05, 2020). Summary plots are given in Figure 10.6.

### 10.3.3 Recruitment

The coverage of the 0 -group survey in 2018 and 2020 was incomplete, and an estimate of the $0-$ group numbers was made for only half of the survey area. In 2021, the coverage was complete. The 0 -group series was recalculated by WGIBAR in 2022. Table 10.3 shows the number of fish in the various year classes from surveys at age $0-2$, and their "survey mortality" from age one to age two is also shown in Figure 10.7.

The 1-group abundance in 2021 was 220.8 billion which is higher than the long-term average (Figure 10.6). The most recent evaluation of the spawning stock and recruitment time-series was made by Gjøsæter et al. (2016).

Future recruitment conditions: High abundance of young herring (mainly age groups 1 and 2) has been suggested to be a necessary but not a single factor causing recruitment failure in the capelin stock (Hjermann et al., 2010; Gjøsæter et al. 2016). In 2021, very little herring at age 1-4 were recorded in the Barents Sea during the ecosystem survey.

### 10.3.4 Comments to the assessment

### 10.3.4.1 Ecological considerations

The number of young herring in the Barents Sea can be an important factor that affects the capelin recruitment. It is not currently taken into account in the assessment model. The benchmark for capelin stocks in the Barents Sea (WKARCT, ICES C.M. 2015/ACOM:31) noted the need for further study of this effect as well as better monitoring of the young herring abundance.
The amount of other food than capelin for cod and other predators may also have changed in recent years. This may also indirectly have affected the predation pressure on capelin. A more detailed discussion of interactions between capelin and other species is given in the 2016-2022 ICES WGIBAR reports.

The abundance of 2-year-olds observed is the highest in 30 years and the high abundance is corresponded by low length-at-age. This is likely a result of high internal competition for food and reduced growth. This tendency is likely enhanced by a strong 2020-yearclass at least partly competing for the same food. The implication is that the majority of this year class had not reached a length of 14 cm and is not expected to migrate to the coast and spawn before winter of 2023.

### 10.3.5 Further work on survey and assessment methodology

### 10.3.5.1 Survey

On 27 February-13 March 2022, IMR carried out a trawl-acoustic monitoring and stock estimation of spawning capelin (Skaret et al. 2022). The survey is the fourth in a series to evaluate whether such a monitoring can be used in the assessment to improve the advice. The initiative and funding come from the Norwegian industry, and the idea in the long term is that monitoring closer to when fishery and spawning happens, can reduce uncertainty in stock advice. Monitoring during spawning has been attempted before, last time in 2007-2009, and has proven to be methodologically challenging due to unpredictable timing and location of the spawning migration.

The survey was carried out using two fishing vessels 'Vendla' and 'Eros'. A stratified design using zig-zag transects with randomized starting points was used and the effort was allocated based on historical and recent information about capelin distribution. The fishery sonar was used actively during the whole survey to estimate size distribution of capelin schools, migration speed and direction. In addition, target-strength measurements were carried out using submersible TSprobes on both vessels. The coverage of the capelin spawning migration was successful and the estimate of ca. 427000 tonnes with a CV of 0.42 was within the expected range from the predictions made in autumn 2021.

Despite the methodological challenges due to timing and distribution of capelin as well as acoustic target strength, the survey results from all four test years have fallen within the uncertainty range of the autumn prediction. This consistency is promising for the use of the survey in an advisory process. An evaluation of the four-year series will be carried out as part of the benchmark for this stock which is planned to be held in 2022.

### 10.3.5.2 Assessment model

In the present capelin assessment model, the only species interaction in the Barents Sea taken explicitly into account is predation by cod on mature capelin. The model does not take into account possible changes in capelin stock dynamics (e.g. maturation), the current state of the environment and stock status of other fish species and mammals in the Barents Sea. The ICES Working Group of Integrated Assessment of the Barents Sea (WGIBAR) has addressed some of these issues.

Consumption of prespawning capelin by mature cod in winter-spring season and autumn season is still not included in the assessment model. It may have a significant impact on capelin SSB calculations.

Gjøsæter et al. (2015) calculated what the quota advice and spawning stock would have been in the period 1991-2013, given the present assessment model and knowledge of the cod stock. By exchanging that cod forecast with the actual amount of cod from the cod assessment model run later in time and rerunning the model, they showed that considerably smaller annual quotas would have been advised if the amount of cod had been known and the present assessment model had been used when the capelin quota was set. Following this work, a retrospective analysis of the capelin assessment as well as of the assessment performance should be included annually. This is a feature which so far has been missing from the capelin assessment.

There is ongoing work to address specific points related to modelling for the benchmark meeting in 2022. These include implementation of survey CV in the capelin assessment model, incorporating the assessment model in Template Model Builder (R-package), validating both the cod consumption part of the model, and the capelin maturation part and updating consumption parameters to reflect recent state in the Barents Sea. As mentioned above, the CapTool methodology
for half-year predictions has already been implemented in R. Historic CVs of SSB estimates will be calculated back to 2004.

### 10.3.6 Reference points

A Blim (SSBlim) management approach has been suggested for this stock (Gjøsæter et al., 2002). In 2002, the JRNFC agreed to adopt a management strategy based on the rule that, with $95 \%$ probability, at least 200000 tonnes of capelin should be allowed to spawn. Consequently, 200000 tonnes was used as a Blim. Alternative harvest control rules of 80,85 and $90 \%$ probability of SSB > Blim were suggested by JNRFC and evaluated by ICES (WKNEAMP-2, ICES C. M. 2016/ACOM:47). ICES considers these rules not to be precautionary. At its 2016 meeting, JNRFC decided not to change the adopted management strategy.

Table 10.1 Barents Sea CAPELIN. International catch (' 000 t ) as used by the Working Group.

| Year | Winter-Spring |  |  | Summer-Autumn |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Norway | Russia | Others | Total | Norway | Russia | Total |  |
| 1965 | 217 | 7 | 0 | 224 | 0 | 0 | 0 | 224 |
| 1966 | 380 | 9 | 0 | 389 | 0 | 0 | 0 | 389 |
| 1967 | 403 | 6 | 0 | 409 | 0 | 0 | 0 | 409 |
| 1968 | 460 | 15 | 0 | 475 | 62 | 0 | 62 | 537 |
| 1969 | 436 | 1 | 0 | 437 | 243 | 0 | 243 | 680 |
| 1970 | 955 | 8 | 0 | 963 | 346 | 5 | 351 | 1314 |
| 1971 | 1300 | 14 | 0 | 1314 | 71 | 7 | 78 | 1392 |
| 1972 | 1208 | 24 | 0 | 1232 | 347 | 13 | 360 | 1591 |
| 1973 | 1078 | 34 | 0 | 1112 | 213 | 12 | 225 | 1337 |
| 1974 | 749 | 63 | 0 | 812 | 237 | 99 | 336 | 1148 |
| 1975 | 559 | 301 | 43 | 903 | 407 | 131 | 538 | 1441 |
| 1976 | 1252 | 228 | 0 | 1480 | 739 | 368 | 1107 | 2587 |
| 1977 | 1441 | 317 | 2 | 1760 | 722 | 504 | 1226 | 2986 |
| 1978 | 784 | 429 | 25 | 1238 | 360 | 318 | 678 | 1916 |
| 1979 | 539 | 342 | 5 | 886 | 570 | 326 | 896 | 1782 |
| 1980 | 539 | 253 | 9 | 801 | 459 | 388 | 847 | 1648 |
| 1981 | 784 | 429 | 28 | 1241 | 454 | 292 | 746 | 1986 |
| 1982 | 568 | 260 | 5 | 833 | 591 | 336 | 927 | 1760 |
| 1983 | 751 | 373 | 36 | 1160 | 758 | 439 | 1197 | 2357 |
| 1984 | 330 | 257 | 42 | 629 | 481 | 368 | 849 | 1477 |


| Year | Winter-Spring |  |  | Summer-Autumn |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Norway | Russia | Others | Total | Norway | Russia | Total |  |
| 1985 | 340 | 234 | 17 | 591 | 113 | 164 | 277 | 868 |
| 1986 | 72 | 51 | 0 | 123 | 0 | 0 | 0 | 123 |
| 1987-1990 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 528 | 159 | 20 | 707 | 31 | 195 | 226 | 933 |
| 1992 | 620 | 247 | 24 | 891 | 73 | 159 | 232 | 1123 |
| 1993 | 402 | 170 | 14 | 586 | 0 | 0 | 0 | 586 |
| 1994-1996 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 |
| 1998 | 0 | 2 | 0 | 2 | 0 | 1 | 1 | 3 |
| 1999 | 50 | 33 | 0 | 83 | 0 | 22 | 22 | 105 |
| 2000 | 279 | 94 | 8 | 381 | 0 | 29 | 29 | 410 |
| 2001 | 376 | 180 | 8 | 564 | 0 | 14 | 14 | 578 |
| 2002 | 398 | 228 | 17 | 643 | 0 | 16 | 16 | 659 |
| 2003 | 180 | 93 | 9 | 282 | 0 | 0 | 0 | 282 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2007 | 2 | 2 | 0 | 4 | 0 | 0 | 0 | 4 |
| 2008 | 5 | 5 | 0 | 10 | 0 | 2 | 0 | 12 |
| 2009 | 233 | 73 | 0 | 306 | 0 | 1 | 1 | 307 |
| 2010 | 246 | 77 | 0 | 323 | 0 | 0 | 0 | 323 |
| 2011 | 273 | 87 | 0 | 360 | 0 | 0 | 0 | 360 |
| 2012 | 228 | 68 | 0 | 296 | 0 | 0 | 0 | 296 |
| 2013 | 116 | 60 | 0 | 177 | 0 | 0 | 0 | 177 |
| 2014 | 40 | 26 | 0 | 66 | 0 | 0 | 0 | 66 |
| 2015 | 71 | 44 | 0 | 115 | 0 | 0 | 0 | 115 |
| 2016-2017 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 129 | 66 | 0 | 195 | 0 | 0 | 0 | 195 |


| Year | Winter-Spring |  | Summer-Autumn |  | Total |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Norway | Russia | Others | Total | Norway | Russia | Total |  |
| $2019-2021$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 42 | 23 | 0 | 65 |  |  |  |  |

Table 10.2. Barents Sea CAPELIN. Stock size estimation table. Estimated stock size ( $10^{9}$ ) by age and length, and biomass ( 1000 tonnes) from the acoustic survey in August-October 2021. TSN: Total stock number. TSB: Total-stock biomass. MSN: Maturing stock number. MSB: Maturing stock biomass.

| Length (cm) | Age/year class |  |  |  |  | $\begin{aligned} & \text { Sum } \\ & 10^{9} \end{aligned}$ | Biomass ( $10^{3} \mathrm{t}$ ) | Mean weight (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 |  |  |  |
|  | 2020 | 2019 | 2018 | 2017 | 2016 |  |  |  |
| 7.0-7.5 | 1.92 |  |  |  |  | 1.92 | 2.53 | 1.32 |
| 7.5-8.0 | 4.82 |  |  |  |  | 4.82 | 9.07 | 1.88 |
| 8.0-8.5 | 15.46 |  |  |  |  | 15.46 | 34.93 | 2.26 |
| 8.5-9.0 | 26.72 | 1.07 |  |  |  | 27.79 | 73.09 | 2.63 |
| 9.0-9.5 | 53.27 | 2.98 |  |  |  | 56.25 | 170.44 | 3.03 |
| 9.5-10.0 | 60.28 | 6.18 |  |  |  | 66.46 | 227.95 | 3.43 |
| 10.0-10.5 | 32.24 | 14.56 |  |  |  | 46.8 | 187.67 | 4.01 |
| 10.5-11.0 | 15.64 | 44.08 |  |  |  | 59.72 | 284.86 | 4.77 |
| 11.0-11.5 | 4.68 | 39.57 |  |  |  | 44.25 | 241.61 | 5.46 |
| 11.5-12.0 | 2.93 | 40.58 | 0.02 |  |  | 43.53 | 278.59 | 6.4 |
| 12.0-12.5 | 1.41 | 34.22 |  |  |  | 35.63 | 265.09 | 7.44 |
| 12.5-13.0 | 0.93 | 31.6 | 0.17 |  |  | 32.7 | 285.18 | 8.72 |
| 13.0-13.5 | 0.35 | 26.38 | 0.24 |  |  | 26.97 | 273.76 | 10.15 |
| 13.5-14.0 | 0.13 | 18.48 | 0.44 |  |  | 19.04 | 224.8 | 11.81 |
| 14.0-14.5 | 0.07 | 15.84 | 0.34 |  |  | 16.25 | 215.82 | 13.28 |
| 14.5-15.0 |  | 13.36 | 0.53 |  |  | 13.89 | 215.3 | 15.5 |
| 15.0-15.5 |  | 14.24 | 0.23 |  |  | 14.47 | 251.54 | 17.38 |
| 15.5-16.0 |  | 9.74 | 1.51 |  |  | 11.25 | 223.36 | 19.85 |
| 16.0-16.5 |  | 6.27 | 0.68 |  |  | 6.95 | 154.24 | 22.18 |
| 16.5-17.0 |  | 6.74 | 0.32 |  |  | 7.06 | 177.3 | 25.1 |
| 17.0-17.5 |  | 2.774 | 1.03 | 0.01 |  | 3.814 | 105.26 | 27.6 |


| Length (cm) | Age/year class |  |  |  |  | $\begin{aligned} & \text { Sum } \\ & 10^{9} \end{aligned}$ | Biomass ( $10^{\mathbf{3}} \mathrm{t}$ ) | Mean weight (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 |  |  |  |
|  | 2020 | 2019 | 2018 | 2017 | 2016 |  |  |  |
| 17.5-18.0 |  | 1.043 | 0.454 |  |  | 1.497 | 48.24 | 32.23 |
| 18.0-18.5 |  | 0.164 | 0.924 |  |  | 1.089 | 36.55 | 33.58 |
| 18.5-19.0 |  | 0.115 |  |  |  | 0.115 | 4.3 | 37.39 |
| 19.0-19.5 |  | 0.0344 | 0.1013 | 0.0006 |  | 0.1362 | 5.38 | 39.46 |
| 19.5-20.0 |  |  |  | 0.0208 |  | 0.0208 | 0.91 | 43.87 |
| 20.5-21.0 |  |  |  | 0.0002 |  | 0.0002 | 0.01 | 47.88 |
| TSN (10 ${ }^{\text {a }}$ ) | 220.85 | 330.0204 | 6.996 | 0.0316 |  | 557.89 |  |  |
| TSB (10) | 757.71 | 3081.46 | 157.23 | 1.22 |  |  | 3997.62 |  |
| Mean length (cm) | 9.58 | 12.57 | 16.11 | 18.95 |  | 11.43 |  |  |
| Mean weight (g) | 3.43 | 9.34 | 22.47 | 38.66 |  |  |  | 7.17 |
| SSN (109) | 0.07 | 70.3204 | 6.12 | 0.0316 |  | 76.54 |  |  |
| SSB (10 ${ }^{9}$ ) | 0.93 | 1287.85 | 147.96 | 1.22 |  |  | 1437.96 |  |

Table 10.3 Barents Sea CAPELIN. Recruitment and natural mortality table. Larval abundance estimate in June, 0-group indices and acoustic estimate in August-September, total mortality from age 1+ to age 2+.

| Year class |  | Larval abundance $\left(10^{12}\right)$ | 0-group swept-area numbers ( $10^{9}$ ind.) | Acoustic estimate (10ind.) | Mortality survey(1- <br> 2) |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 (Y) | ) $0+(\mathrm{Y})$ | 1(Y+1) | 2(Y+2) | \% |
| 1980 | - | 740 | 402.6 | 147.6 | 63 |
| 1981 | 9.7 | 477 | 528.3 | 200.2 | 62 |
| 1982 | 9.9 | 600 | 514.9 | 186.5 | 64 |
| 1983 | 9.9 | 340 | 154.8 | 48.3 | 69 |
| 1984 | 8.2 | 275 | 38.7 | 4.7 | 88 |
| 1985 | 8.6 | 64 | 6.0 | 1.7 | 72 |
| 1986 | 0.0 | 42 | 37.6 | 28.7 | 24 |
| 1987 | 0.3 | 4 | 21.0 | 17.7 | 16 |
| 1988 | 0.3 | 65 | 189.2 | 177.6 | 6 |
| 1989 | 7.3 | 862 | 700.4 | 580.2 | 17 |


| Year class |  | Larval abundance $\left(10^{12}\right)$ | 0-group swept-area numbers ( $10^{9}$ ind.) | Acoustic estimate ( $10^{9}$ ind.) | Mortality survey(12) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 13.0 | 0116 | 402.1 | 196.3 | 51 |
| 1991 | 3.0 | 169 | 351.3 | 53.4 | 85 |
| 1992 | 7.3 | 2 | 2.2 | 3.4 | -- |
| 1993 | 3.3 | 1 | 19.8 | 8.1 | 59 |
| 1994 | 0.1 | 14 | 7.1 | 11.5 | -- |
| 1995 | 0.0 | 3 | 81.9 | 39.1 | 52 |
| 1996 | 2.4 | 137 | 98.9 | 72.6 | 27 |
| 1997 | 6.9 | 189 | 179.0 | 101.5 | 43 |
| 1998 | 14.1 | 1113 | 156.0 | 110.6 | 29 |
| 1999 | 36.5 | 5288 | 449.2 | 218.7 | 51 |
| 2000 | 19.1 | 1141 | 113.6 | 90.8 | 20 |
| 2001 | 10.7 | 7 90 | 59.7 | 9.6 | 84 |
| 2002 | 22.4 | 4.467 | 82.4 | 24.8 | 70 |
| 2003 | 11.9 | 9 341 | 51.2 | 13.0 | 75 |
| 2004 | 2.5 | 54 | 26.9 | 21.7 | 19 |
| 2005 | 8.8 | 148 | 60.1 | 54.7 | 9 |
| 2006 | 17.1 | 1516 | 221.7 | 231.4 | -- |
| 2007 | - | 480 | 313.0 | 166.4 | 46 |
| 2008 | - | 995 | 124.0 | 127.6 | -- |
| 2009 | - | 673 | 248.2 | 181.1 | 27 |
| 2010 | - | 319 | 209.6 | 156.4 | 25 |
| 2011 | - | 594 | 145.9 | 216.2 | - |
| 2012 | - | 989 | 324.5 | 106.6 | 67 |
| 2013 | - | 316 | 105.1 | 40.5 | 62 |
| 2014 | - | 164 | 39.5 | 8.1 | 79 |
| 2015 | - | 457 | 31.6 | 123.7 | - |
| 2016 | - | 779 | 86.4 | 59.6 | 31 |
| 2017 | - | 214 | 58.6 | 7.0 | 88 |


| Year class | Larval abundance <br> $\left(10^{12}\right)$ | 0-group swept-area <br> numbers (109 ind.) | Acoustic estimate <br> $\left(10^{9}\right.$ ind.) | Mortality survey(1一 <br> 2) |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 2018 | - | 680 | 17.5 | 31.1 | 10 |
| 2019 | - | 1465 | 366.4 | 330.0 |  |
| 2020 | - | 325 |  |  |  |
| 2021 | 972 | 176.8 | 105.2 |  |  |
| Average | 9.0 |  |  |  |  |

*In the brackets - the correction numbers, taking into account not surveyed area.

Table 10.4 Barents Sea CAPELIN. Stock size in numbers by age, total-stock biomass, biomass of the maturing component (MSB) at 1. October.

| Year | Stock in numbers ( $10^{9}$ ) |  |  |  |  |  | Biomass ( $10^{3}$ tonnes) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Total | Total | MSB |
| 1973 | 528 | 375 | 40 | 17 | 0 | 961 | 5144 | 1350 |
| 1974 | 305 | 547 | 173 | 3 | 0 | 1029 | 5733 | 907 |
| 1975 | 190 | 348 | 296 | 86 | 0 | 921 | 7806 | 2916 |
| 1976 | 211 | 233 | 163 | 77 | 12 | 696 | 6417 | 3200 |
| 1977 | 360 | 175 | 99 | 40 | 7 | 681 | 4796 | 2676 |
| 1978 | 84 | 392 | 76 | 9 | 1 | 561 | 4247 | 1402 |
| 1979 | 12 | 333 | 114 | 5 | 0 | 464 | 4162 | 1227 |
| 1980 | 270 | 196 | 155 | 33 | 0 | 654 | 6715 | 3913 |
| 1981 | 403 | 195 | 48 | 14 | 0 | 660 | 3895 | 1551 |
| 1982 | 528 | 148 | 57 | 2 | 0 | 735 | 3779 | 1591 |
| 1983 | 515 | 200 | 38 | 0 | 0 | 754 | 4230 | 1329 |
| 1984 | 155 | 187 | 48 | 3 | 0 | 393 | 2964 | 1208 |
| 1985 | 39 | 48 | 21 | 1 | 0 | 109 | 860 | 285 |
| 1986 | 6 | 5 | 3 | 0 | 0 | 14 | 120 | 65 |
| 1987 | 38 | 2 | 0 | 0 | 0 | 39 | 101 | 17 |
| 1988 | 21 | 29 | 0 | 0 | 0 | 50 | 428 | 200 |
| 1989 | 189 | 18 | 3 | 0 | 0 | 209 | 864 | 175 |
| 1990 | 700 | 178 | 16 | 0 | 0 | 894 | 5831 | 2617 |
| 1991 | 402 | 580 | 33 | 1 | 0 | 1016 | 7287 | 2248 |


| Year | Stock in numbers ( $10^{9}$ ) |  |  |  |  |  | Biomass ( $10^{3}$ tonnes) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Total | Total | MSB |
| 1992 | 351 | 196 | 129 | 1 | 0 | 678 | 5150 | 2228 |
| 1993 | 2 | 53 | 17 | 2 | 2 | 75 | 796 | 330 |
| 1994 | 20 | 3 | 4 | 0 | 0 | 28 | 200 | 94 |
| 1995 | 7 | 8 | 2 | 0 | 0 | 17 | 193 | 118 |
| 1996 | 82 | 12 | 2 | 0 | 0 | 96 | 503 | 248 |
| 1997 | 99 | 39 | 2 | 0 | 0 | 140 | 911 | 312 |
| 1998 | 179 | 73 | 11 | 1 | 0 | 263 | 2056 | 931 |
| 1999 | 156 | 101 | 27 | 1 | 0 | 285 | 2776 | 1718 |
| 2000 | 449 | 111 | 34 | 1 | 0 | 595 | 4273 | 2099 |
| 2001 | 114 | 219 | 31 | 1 | 0 | 364 | 3630 | 2019 |
| 2002 | 60 | 91 | 50 | 1 | 0 | 201 | 2210 | 1290 |
| 2003 | 82 | 10 | 11 | 1 | 0 | 104 | 533 | 280 |
| 2004 | 51 | 25 | 6 | 1 | 0 | 82 | 628 | 294 |
| 2005 | 27 | 13 | 2 | 0 | 0 | 42 | 324 | 174 |
| 2006 | 60 | 22 | 6 | 0 | 0 | 88 | 787 | 437 |
| 2007 | 222 | 55 | 4 | 0 | 0 | 280 | 1882 | 844 |
| 2008 | 313 | 231 | 25 | 2 | 0 | 571 | 4427 | 2468 |
| 2009 | 124 | 166 | 61 | 0 | 0 | 352 | 3756 | 2323 |
| 2010 | 248 | 128 | 61 | 1 | 0 | 438 | 3500 | 2051 |
| 2011 | 209 | 181 | 55 | 8 | 0 | 454 | 3707 | 2115 |
| 2012 | 146 | 156 | 88 | 2 | 0 | 392 | 3586 | 1997 |
| 2013 | 324 | 216 | 59 | 7 | 0 | 610 | 3956 | 1471 |
| 2014 | 105 | 107 | 39 | 2 | 0 | 253 | 1949 | 873 |
| 2015 | 40 | 40 | 13 | 1 | 0 | 94 | 842 | 375 |
| 2016 | 32 | 8 | 3 | 0 | 0 | 43 | 328 | 181 |
| 2017 | 86 | 124 | 17 | 0 | 0 | 227 | 2506 | 1723 |
| 2018 | 59 | 60 | 21 | 0 | 0 | 140 | 1597 | 1056 |
| 2019 | 17 | 9 | 7 | 1 | 0 | 35 | 411 | 302 |


| Year | Stock in numbers (109) |  |  |  | Biomass (103 tonnes) |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Total | Total | MSB |
| 2020 | 366 | 31 | 4 | 1 | 0 | 403 | 1884 | 533 |
| 2021 | 221 | 330 | 7 | 0 | 0 | 558 | 3998 | 1438 |

Table 10.5 Barents Sea CAPELIN. Summary stock and data for prognoses table. Recruitment and total biomass (TSB) are survey estimates back-calculated to 1 August (before autumn fishing season) for 1985 and earlier; for 1986 and later it is the survey estimate. Maturing biomass (MSB) is the survey estimate of fish above length of maturity ( 14.0 cm ). SSB is the median value of the modelled stochastic spawning-stock biomass (after winter/spring fishery). * - indicates a very small spawning stock. "SSB by winter" is acoustic assessment in the winter-spring survey in next year. For most of the years, the survey area was covered partly. Estimates from spawning surveys going back to the 1980s are given in Gjøsæter and Prozorkevitch (WD05, AFWG 2021) and not included here.

| Year |  | tic t) | SSB, assess- <br> ment <br> model, <br> April 1 year+1 <br> $\left(10^{3} \mathrm{t}\right)$ | Recruitment <br> Age 1, <br> survey <br> assessment <br> 1 October <br> $10^{9} \mathrm{sp}$. | Young herring biomass age $1+2$ ( $10^{3}$ tons) source: WGIBAR 2022 | Herring 0 group sweptarea index ( $10^{9}$ ind.p) source: <br> WGIBAR 2022 | Capelin landing $\left(10^{3} t\right)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | TSB | MSB |  |  |  |  |  |
| 1972 | 6600 | 2727 |  | 152 | 2 | 1591 |  |
| 1973 | 5144 | 1350 | 0 33 | 529 | 2 | 1337 |  |
| 1974 | 5733 | 907 | * | 305 | 48 | 1148 |  |
| 1975 | 7806 | 2916 | 6 | 190 | 74 | 1441 |  |
| 1976 | 6417 | 3200 | 0253 | 211 | 39 | 2587 |  |
| 1977 | 4796 | 2676 | 6 22 | 360 | 46 | 2986 |  |
| 1978 | 4247 | 1402 | 2 | 84 | 52 | 1916 |  |
| 1979 | 4162 | 1227 | 7 | 12 | 39 | 1782 |  |
| 1980 | 6715 | 3913 | 3 | 270 | 66 | 0 | 1648 |
| 1981 | 3895 | 1551 | 1316 | 403 | 47 | 0 | 1986 |
| 1982 | 3779 | 1591 | 106 | 528 | 9 | 3 | 1760 |
| 1983 | 4230 | 1329 | 100 | 515 | 12 | 195 | 2357 |
| 1984 | 2964 | 1208 | 109 | 155 | 1467 | 27 | 1477 |
| 1985 | 860 | 285 | * | 39 | 2638 | 20 | 868 |
| 1986 | 120 | 65 | * | 6 | 191 | 0 | 123 |
| 1987 | 101 | 17 | 34 | 38 | 288 | 0 | 0 |
| 1988 | 428 | 200 | * | 21 | 77 | 61 | 0 |



| Year |  |  | SSB, assessment model, April 1 year+1 $\left(10^{3} \mathrm{t}\right)$ | Recruitment <br> Age 1, <br> survey <br> assessment <br> 1 October <br> $10^{9} \mathrm{sp}$. | Young herring biomass age $1+2$ ( $10^{3}$ tons) source: WGIBAR 2022 | Herring 0group sweptarea index ( $10^{9}$ ind.p) source: <br> WGIBAR 2022 | Capelin landing $\left(10^{3} t\right)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 842 | 375 | 82 | 40 | 963 | 83 | 115 |
| 2016 | 328 | 181 | 37 | 32 | 498 | 79 | 0 |
| 2017 | 2506 | 1723 | 3462 | 124 | 1106 | 154 | 0 |
| 2018 | 1597 | 1056 | 6317 | 59 | 2034 | 55 | 195 |
| 2019 | 411 | 302 | 85 | 17 | 389 | 50 | 0 |
| 2020 | 1884 | 533 | 154 | 366 | 359 | 12 | 0 |
| 2021 | 3998 | 1438 | $8 \quad 420$ | 221 | 152 | 209 | 0 |
| 2022 |  |  |  |  |  | 65 |  |



Figure 10.1. Geographical distribution of capelin in August-September 2021.


Figure 10.2. Position of trawl hauls and weighting of the corresponding capelin length distributions applied in the acoustic estimate in 2021. The weighting is proportional to NASC within a $\mathbf{1 0} \mathbf{n m}$ radius.


Figure 10.3 Weight-at-age (grammes) for capelin from the autumn survey.


Figure 10.4. Probabilistic prognosis 1 October 2021-1 April 2022 for Barents Sea capelin maturing stock, with a catch of 70000 tonnes (model CapTool, 50000 simulations).


Figure 10.5. Probability of SSB 2022 < $\mathrm{B}_{\mathrm{lim}}$ as a function of the catch


Recruitment (age 1)



Figure 10.6. Capelin in subareas 1 and 2, excluding Division 2a west of $5^{\circ} \mathrm{W}$ (Barents Sea capelin). Landings, recruitment, and summary of stock assessment (mature and immature stock biomass in tonnes.


Figure 10.7. Capelin survey mortality per year class from age 1-2 (survey data).

## 11 References

Aanes, S. 2002. Precision in age determination of Northeast Arctic cod. Working document in: Report of the Arctic Fisheries Working Group. ICES Headquarters 16-25 April 2002. ICES CM 2002/ACFM:18. 451 pp.

Aanes, S. and Pennington, M. 2003. On estimating the Age composition of the Commercial Catch of Northeast Arctic cod from a Sample of Clusters. ICES Journal of Marine Science, 60: 297-303.

Aanes, S. 2016. Stock Annex: Herring (Clupea harengus) in subareas 1, 2, and 5, and in division 4.a and 14.1 (Norwegian Spring Spawning). WKPELA2016

Aas, C. A.2007. Predation by saithe on juvenile fish (cod and others). Master's thesis, University of Tromsø, 2007 (In Norwegian).

Ajiad, A. M., Aglen, A., Nedreaas, K., and Kvamme, C. 2008. Estimating bycatch at age of Northeast arctic cod from the Norwegian shrimp fishery in the Barents Sea 1984-2006. WD2, AFWG 2008.

Albert, O.T. and Vollen, T., 2015. A major nursery area around the Svalbard archipelago provides recruits for the stocks in both Greenland halibut management areas in the Northeast Atlantic. ICES Journal of Marine Science: Journal du Conseil, 72(3): 872-879.

Anon. 2013. The Norwegian Reference Fleet - a trustful cooperation between fishermen and scientists. Focus on Marine Research 3/2013, Institute of Marine Research, Norway. 12 pp.
Anfinsen, L. 2002. Ressursøkologisk betydning av nise (Phocaena phocaena) i norske farvann. Dr. scient thesis. Institute of fisheries and marine biology, University of Bergen, Autumn 2002. 51pp. (In Norwegian).

Bakanev, S 2013. Assessment of the Barents Sea Greenland halibut stock using the stochastic version of the production model. Working document no 14. in: Report of the Benchmark Workshop on Greenland Halibut Stocks (WKBUT), 26-29 November 2013, Copenhagen, Denmark. ICES CM 2013/ ACOM:44. 367 pp.

Barrett, R.T., T. Anker-Nilssen, G.W. Gabrielsen and G. Chapdelaine. 2002. Food consumption by seabirds in Norwegian waters. ICES Journal of Marine Science 59(1): 43-57.

Basterretxea, M., J. Ruiz, A. Iriondo and E. Mugerza 2013. Spanish Greenland halibut survey 2012. Working document in: Report of the Arctic Fisheries Working Group. ICES Headquarters, Copenhagen, 18-24 April 2013. ICES CM 2013/ACOM:05. 726 pp.

Berg, E., and Albert, O. T. 2003. Cod in fjords and coastal waters of North Norway: distribution and variation in length and maturity at age. ICES Journal of Marine Science, 60: 787-797.

Berg, E., Sarvas, T. H., Harbitz, A., Fevolden, S.E. and Salberg, A.B. 2005. Accuracy and precision in stock separation of north-east Arctic and Norwegian coastal cod by otoliths - comparing readings, image analyses and a genetic method. Marine and Freshwater Research, No. 5610 pp.

Berg, H-S. and Nedreaas, K. 2021. Estimering av utkast i norsk kystfiske med garn ved bruk av Kystreferanseflåten. Estimation of discards in the Norwegian coastal gillnet fisheries based on catch reportings from the Coastal reference fleet. Institute of Marine Research report series: 2021-1, 95 pp. (In Norwegian)
Björnsson, H., and Sigurdsson, T. 2003. Assessment of golden redfish (Sebastes marinus L.) in Icelandic waters. Scienta Marina 67 (Suppl. 1):301-314. Scientia Marina, 67: 301-314.

Blom, G. 2015. Omregningsfaktorer for produkter av torsk (Gadus morhua) nord for $62^{\circ}$ nord i vintersesongen 2015/Conversion factors for products of cod (Gadus morhua) north of $62^{\circ}$ north in the winter season 2015. Directorate of Fisheries, Norway, Report no. 14/17412. 65 pp.
Bogstad B., and Gjøsæter H., 1994. A method for calculating the consumption of capelin by cod. ICES J. mar. Sci. 51:273-280.

Bogstad, B., Howell, D., Åsnes, M. N. (2004). A closed life-cycle model for Northeast Arctic cod. ICES C.M.2004/K:26, 12 pp.

Bogstad, B., Haug, T. and Mehl, S. 2000. Who eats whom in the Barents Sea? NAMMCO Sci. Publ. 2: 98119.

Bogstad, B. and Mehl, S. 1997. Interactions Between Cod (Gadus morhua) and Its Prey Species in the Barents Sea. Forage Fishes in Marine Ecosystems. Proceedings of the International Symposium on the Role of Forage Fishes in Marine Ecosystems. Alaska Sea Grant College Program Report No. 97-01: 591-615. University of Alaska Fairbanks.

Brander, K. 2002. Predicting weight at age. Internal ICES note to assessment working groups. 2003. Software implementation of process models. Working Document No. 2 to the Arctic Fisheries Working Group, San Sebastian, Spain, 23 April- 2 May 2003.

Breivik, O. N., Storvik, G. O., Nedreaas, K. H. 2017. Latent Gaussian models to predict historical bycatch in commercial fishery. Fisheries Research 185: 62-72. doi: 10.1016/j.fishres.2016.09.033

Brooks, M.E, Kristensen, K., van Benthem, K.J., Magnusson, A., Berg, C.W., Nielsen, A., Skaug, H.J., Maechlerm M., and Bolker, B.M. 2017. glmmTMB balances speed and flexibility among packages for zeroinflated generalized linear mixed modeling. The R Journal, 9(2), 378-400.

Bulgakova, T., 2005. 'To recruitment prognosis of NEA cod'. Working document \#20 in: Report of the Arctic Fisheries Working Group', Murmansk, Russia, April 19-28, 2005. ICES C.M. 2005/ACFM:20, 564 pp.

Cañás, L., Stransky, C., Schlickeisen, J., Sampedro, M. P., and Fariña, A. C. 2012. Use of the otolith shape analysis in stock identification of anglerfish (Lophius piscatorius) in the Northeast Atlantic. - ICES Journal of Marine Science, 69: 1-7.

Capizzano, C. W., Mandelman, J. W., Hoffman, W. S., Dean, M. J., Zemeckis, D. R., Benor^t, H. P., Kneebone, J., Jones, E., Stettner, M. J., Buchan, N. J., Langan, J. A., and Sulikowski, J. A. 2016. Estimating and mitigating the discard mortality of Atlantic cod (Gadus morhua) in the Gulf of Maine recreational rod-and-reel fishery. - ICES Journal of Marine Science, 73: 2342-2355.
Charnov, E.L., Gislason, H., and Pope, J.G. 2013. Evolutionary assembly rules for fish life histories. Fish Fish. 14(2): 213-224.

Clark, W. G. 2002. F 35\% Revisited Ten Years Later. North American Journal of Fisheries Management, 22: 251-257.
http://www.tandfonline.com/doi/abs/10.1577/15488675\(2002\)022\<0251\%3AFRTYL\>2.0.CO\%3B2 (Accessed 4 January 2018).

Clegg, T. and Williams, T. 2020. Monitoring bycatches in Norwegian fisheries - Species registered by the Norwegian Reference Fleet 2015-2018. Rapport fra Havforskningen 2020-8. ISSN:1893-4536. https://www.hi.no/hi/nettrapporter/rapport-fra-havforskningen-en-2020-8

Dahle, G., Johansen, T., Westgaard, J-I., Aglen, A., Glover, K. 2018. Genetic management of mixed-stock fisheries "real-time": The case of the largest remaining cod fishery operating in the Atlantic in 20072017. Fisheries Research 205 (2018) pp 77-85.

Dahle, G., et al. 2018. Analysis of coastal cod (Gadus morhua L.) sampled on spawning sites reveals a genetic gradient throughout Norway's coastline. - BMC Genetics 19: 42.

Dingsør, G. E. 2001. Estimation of discards in the commercial trawl fishery for Northeast Arctic cod (Gadus morhua L.) and some effects on assessment. Cand. Scient thesis, University of Bergen, 2001.

Dingsør, G.E. 2005. Estimating abundance indices from the international 0-group fish survey in the Barents Sea. Fisheries Research 72(2-3): 205-218.

Dingsør G.E., Bogstad B., Stiansen J.E., and Subbey S. 2010. How can we assess recruitment models for (age3) NEA cod? / WD 19, AFWG 2010.

Dolgov, A.V., Yaragina, N.A., Orlova, E.L., Bogstad, B., Johannesen, E., Mehl, S. 2007. 20 ${ }^{\text {th }}$ anniversary of the PINRO-IMR cooperation in the investigations of fish feeding in the Barents Sea - results and perspectives. In : Haug,. T., Misund, O.A., Gjøsæter, H., and Røttingen, I. (eds.). Long-term bilateral Rus-sian-Norwegian scientific cooperation as a basis for sustainable management of living marine
resources in the Barents Sea. Proceeding of the $12^{\text {th }}$ Norwegian-Russian Symposium. Tromsø, 21-22 August 2007. P.44-78.

Duarte R, Azevedo M, and Pereda P 1997. Study of the growth of southern black and white monkfish stocks. ICES Journal of Marine Science 54(5): 866-874.

Dyb J.E., 2003. Bestandsstudie av breiflabb (Lophius piscatorius L.) langs kysten av Møre og i Nordsjøen. Cand.scient thesis, University of Bergen. 105 pp. (In Norwegian)

Eikeset A.M., Richtera A., Dunlop E.S., Dieckmann U., and Stenseth N. C. 2013. Economic repercussions of fisheries-induced evolution. PNAS vol. 110, no. 30: 12259-12264.

Eriksen, E., Prozorkevich, D. V. and Dingsør, G. E. 2009. An evaluation of 0-group abundance indices of Barents Sea Fish Stocks. The Open fish Science Journal, 2: 6-14.

Fall J., 2020. NEA cod and haddock indices from the Barents Sea winter survey 2020. Working document no 10 in: Report of the Arctic Fisheries Working Group (AFWG), 15-22 April 2020. ICES CM 2015/ACOM:05. 639 pp.
Fotland, Å., Nedreaas, K. 2020. Adjusted conversion factors for products of cod (Gadus morhua) and consequences for Norwegian catch data from ICES Subareas 1 and 2 during 1992-2018. WD no. 9 to ICES AFWG 2020 (ICES 2020).

Gislason, H., Daan, N., Rice, J. C., and Pope, J. G. 2010. Size, growth, temperature and the natural mortality of marine fish. Fish and Fisheries, 11: 149-158. https://onlinelibrary.wiley.com/doi/abs/10.1111/j.14672979.2009.00350.x (Accessed 3 January 2020).

Gjøsæter, H., Bogstad, B. and Tjelmeland, S. 2002. Assessment methodology for Barents Sea capelin, Mallotus villosus (Müller). ICES Journal of Marine Science, 59: 1086-1095.

Gjøsæter, H., Bogstad, B., Tjelmeland, S., and Subbey, S. 2015. Retrospective evaluation of the Barents Sea capelin management advice Marine Biology Research 11(2):135-143.

Gjøsæter, H., Hallfredsson, E. H., Mikkelsen, N., Bogstad, B., and Pedersen, T. 2016. Predation on early life stages is decisive for year class strength in the Barents Sea capelin (Mallotus villosus) stock. ICES Journal of Marine Science 73(2):182-195.

Golovanov S.E., Sokolov A.M., and Yaragina, N.A. 2007. Revised indices of the Northeast Arctic cod abundance according to the 1982-2006 data from Russian trawl-acoustic survey (TAS). Working Document \#3 for AFWG 2007.

Gulland, J. 1964. The abundance of fish stocks in the Barents Sea. Rapp. P.-v. Réun. Cons. Int. Explor. Mer, 155: 126-137.

Håkon Otterå and Kjell Nedreaas, 2020. Effort and catch-per-unit-effort (CPUE) for Norwegian trawlers fishing cod north of $67^{\circ} \mathrm{N}$ in 2011-2019. Working document no 2 in: Report of the Arctic Fisheries Working Group (AFWG), 15-22 April 2020. ICES CM 2015/ACOM:05. 639 pp.

Hallfredsson, E. and Vollen, T. 2015. Juvenile index for Barents Sea Greenland halibut. Working document no. 1 in Report of the Inter Benchmark Process on Greenland Halibut in ICES areas I and II (IBPHALI). ICES CM 2015 $\backslash \mathrm{ACOM}: 54.37 \mathrm{pp}$.

Hallfredsson, E. and Vollen, T. 2015. Two abundance indices for Greenland halibut based on the Joint Ecosystem Survey in the Barents Sea and previous surveys in the nursery area. Working document no 20. in: Report of the Arctic Fisheries Working Group (AFWG), 23-29 April 2015, Hamburg, Germany. ICES CM 2015/ACOM:05. 639 pp.

Hamel, O.S. 2014. A method for calculating a meta-analytical prior for the natural mortality rate using multiple life history correlates. ICES J. Mar. Sci. 72(1): 62-69.

Hartig, F. 2020. DHARMa: residual diagnostics for hierarchical (multi-level/mixed) regression models. R package version 0.2.7. https://CRAN.R-project.org/package=DHARMa

Heino M, Dieckmann U, Godø OR 2002. Reaction norm analysis of fisheries-induced adaptive change and the case of the Northeast Arctic cod. ICES CM 2002/Y: 14.

Hirst, D., Aanes, S., Storvik, G. and Tvete, I.F. 2004. Estimating catch at age from market sampling data using a Bayesian hierarchical model. Journal of the Royal statistical society. Series C, applied statistics, 53: 1-14.

Hirst, D., Storvik, G., Aldrin, M., Aanes, S. and Huseby, R.B. 2005. Estimating catch-at-age by combining data from different sources. Canadian Journal of Fisheries and Aquatic Sciences 62:1377-1385.

Hirst, D., Storvik, G., Rognebakke, H., Aldrin, M., Aanes, S., and Vølstad, J. H. 2012. A Bayesian modelling framework for the estimation of catch-at-age of commercially harvested fish species. Can. J. Fish. Aquat. Sci. 69: 2064-2076.

Hislop, J. R. G., Gallego, A., Heath, M. R., Kennedy, F. M., Reeves, S. A., and Wright, P. J. 2001. A synthesis of the early life history of the anglerfish, Lophius piscatorius (Linnaeus, 1758) in northern British waters. ICES Journal of Marine Science 58:70-86.

Hjermann, D. Ø., Bogstad, B., Eikeset, A. M., Ottersen, G., Gjøsæter, H., and Stenseth, N. C. 2007. Food web dynamics affect Northeast Arctic cod recruitment. Proceedings of the Royal Society, Series B 274:661669.

Hjermann, D. Ø., Bogstad, B., Dingsør, G. E., Gjøsæter, H., Ottersen, G., Eikeset, A. M., and Stenseth, N. C. 2010. Trophic interactions affecting a key ecosystem component: a multi-stage analysis of the recruitment of the Barents Sea capelin. Canadian Journal of Fisheries and Aquatic Science 67:1363-1375.

Höffle H. and Planque B. (in revision). Natural mortality estimations for beaked redfish (Sebastes mentella) - a long-lived ovoviviparous species of the Northeast Arctic. Fisheries Research11 pp.

Höffle H. and Tranang C. A. 2020. Use of RstoX for recalculating numbers at age of Sebastes mentella from the joint NOR-RUS Barents Sea Ecosystem Survey in summer and autumn. WD18-ICES AFWG2020. xx pp.

Hoenig, J. M. 1983. Empirical use of longevity data to estimate mortality rates. Fisheries Bulletin U.S. 81:898903.

Holt, R.E., Bogstad, B., Durant, J. M., Dolgov, A. V., and Ottersen, G. 2019. Barents Sea cod (Gadus morhua) diet composition: long-term interannual, seasonal, and ontogenetic patterns. ICES Journal of Marine Science 76(6): 1641-1652, doi:10.1093/icesjms/fsz082

Hordyk, A.R., Ono, K., Sainsbury, K.J., Loneragan, N., and Prince, J.D. 2015a. Some explorations of the life history ratios to describe length composition, spawning-per-recruit, and the spawning potential ratio. ICES J. Mar. Sci. 72: 204-216.

Hordyk, A.R., Ono, K., Valencia, S.R., Loneragan, N.R., and Prince, J.D. 2015b. A novel length-based empirical estimation method of spawning potential ratio (SPR), and tests of its performance, for smallscale, data-poor fisheries. ICES J. Mar. Sci. 72: 217-231.

Hordyk, A., Ono, K., Prince, J.D., and Walters, C.J. 2016. A simple length-structured model based on life history ratios and incorporating size-dependent selectivity: application to spawning potential ratios for data-poor stocks. Can. J. Fish. Aquat. Sci. 13: 1-13. doi: 13.1139/cjfas-2015-0422.

Howell, D. 2020. Greenland halibut $\mathrm{HR}_{\text {pa }}$ proposal. Working document no 15. in: Report of the Arctic Fisheries Working Group (AFWG), 16-22 April 2020, Copenhagen, Denmark.

Høie, H., Bernreuther, M., Ågotnes, P., Beußel, F., Koloskova, V., Mjanger, H., Schröder, D., Senneset, H. and Zuykova, N. 2009. Report of Northeast Arctic cod otolith exchange between Russia, Norway and Germany 2008. WD \# 6 ICES Arctic Fisheries Working Group 2009, San Sebastian 21-27th April 2009.

ICES 1997. Report of the ICES/NAFO workshop on Greenland halibut age determination. 26-29 November 1996, Reykjavik, Iceland. ICES CM1997/G:1. 16 pp.

ICES 2001. Report of the Arctic Fisheries Working Group. Bergen, Norway, 24 April - 3 May 2001. ICES CM 2001/ACFM:19. 380 pp.

ICES 2003. Study Group on Biological Reference Points for Northeast Arctic Cod. Svanhovd, Norway 1317 January 2003. ICES CM 2003/ACFM:11.

ICES 2006a. ICES Workshop on Biological Reference Points for North East Arctic Haddock (WKHAD). Svanhovd, Norway, 6-10 March 2006. ICES C.M. 2006/ACFM:19, 102 pp.

ICES 2006b. Report of the Arctic Fisheries Working Group, Copenhagen 19-28 April 2006. ICES C.M. 2006/ACFM:25, 594 pp.

ICES 2007b. Report of the Arctic Fisheries Working Group, Vigo, Spain 18-27 April 2007. ICES C.M. 2007/ACFM:16, 651 pp.

ICES 2009. Report of the Benchmark Workshop on Short-lived Species (WKSHORT). ICES C.M. 2009/ACOM: 34: 1-166.

ICES. 2009. Report of the Workshop on analytical methods for evaluation of extinction risk of stocks in poor condition (WKPOOR1), 18-20 May 2009, Copenhagen, Denmark. ICES CM 2009 \ACOM:29. 21 pp.
ICES 2009. Report of the workshop for the exploration of the dynamics of fish stocks in poor conditions (WKPOOR2). ICES CM, 2009/ACOM:49: 30pp.

ICES 2010. Report of the Arctic Fisheries Working Group, Lisbon/Bergen, $22-28$ April 2010. ICES C.M. 2010/ACOM:05, 664 pp.
ICES 2010. Report of the Benchmark Workshop on Roundfish (WKROUND), 9-16 February 2010, Copenhagen, Denmark. ICES CM 2010/ACOM: 36. 183 pp.

ICES 2011. Report of the Arctic Fisheries Working Group, Hamburg, Germany 28 April - 4 May 2011. ICES C.M. 2011/ACOM:05, 659 pp.

ICES 2011. Report of the Workshop of Implementing the ICES Fmsy framework (WKFRAME2), 10-14 February 2011, ICES, Denmark. ICES C. M. 2011/ACOM:33, 110 pp.
ICES 2011. Report of the Benchmark Workshop on Roundfish and Pelagic Stocks, Lisbon 24-31 January 2011. ICES C.M. 2011/ACOM:38, 418 pp.

ICES. 2011. Report of the Planning Group on Commercial Catches, Discards and Biological Sampling (PGCCDBS), 7-11 February 2011, Vienna, Austria. ICES CM 2011/ACOM:40. 174 pp.
ICES. 2011. Report of the Workshop on Age Reading of Greenland Halibut (WKARGH), 14-17 February 2011, Vigo, Spain. ICES CM 2011/ACOM:41. 39 pp.

ICES. 2012. Report of the Study Group on Recruitment Forecasting (SGRF),15-19 October 2012, Barcelona, Spain. ICES CM 2012/ACOM:24. 36 pp.
ICES. 2012b. ICES Implementation of Advice for Data-limited Stocks in 2012 in its 2012 Advice. ICES CM 2012/ACOM 68.42 pp. https://doi.org/10.17895/ices.pub. 5322
ICES 2012. Report of the benchmark workshop on redfish (WKRED). ICES CM, 2012/ACOM: 48: 289 pp .
ICES 2013. Report of the Arctic Fisheries Working Group, Copenhagen, 18-24 April 2013. ICES C.M. 2013/ACOM:05, 726 pp.
ICES. 2013. Report of the Study Group on Recruitment Forecasting (SGRF), 18-22 November 2013,Lisbon. ICES CM 2013/ACOM:24. 29 pp.

ICES 2013. Report of the Benchmark Workshop on Greenland Halibut Stocks (WKBUT), 26-29 November 2013, Copenhagen, Denmark. ICES CM 2013/ACOM:44

ICES 2014. Report of the Data Compilation Workshop on Northeast Arctic Greenland Halibut and Assessment Methods (DCWKNGHD), 10-12 November 2014, Murmansk, Russia. ICES CM 2014/ACOM:65. 58 pp .
ICES 2014. Report of the Working Group on Harp and Hooded Seals Quebec City 17-21 November 2014. ICES C. M. 2014/ACOM:20, 62 pp.

ICES. 2015. Report of the Benchmark Workshop on Arctic Stocks (WKARCT), 26-30 January 2015, ICES Headquarters, Denmark. ICES CM 2015 $\backslash \mathrm{ACOM}: 31.126$ pp.
ICES 2015. Report of the Arctic Fisheries Working Group, Hamburg, Germany, 23-29 April 2015. ICES C.M. 2015/ACOM:05, 590 pp.

ICES 2015. Report of the Inter Benchmark Process on Greenland Halibut in ICES areas I and II (IBPHALI). By Correspondence, August 2015. ICES CM 2015/ACOM:54, 41 pp.

ICES 2015. Report of the first Workshop on Management Plan Evaluation on Northeast Arctic cod and haddock and Barents Sea capelin (WKNEAMP-1), 5, Series title. ICES CM 2015/ACOM:60, 27 pp.
ICES 2016. Report of the second Workshop on Management Plan Evaluation on Northeast Arctic cod and haddock and Barents Sea capelin (WKNEAMP-2), 25-28 January 2016, Kirkenes, Norway. ICES CM 2016/ACOM:47, 76 pp.

ICES 2016. The Third Report of the Working Group on Integrated Assessments of the Barents Sea (WGIBAR). Murmansk, Russia, 22-25 February 2016. ICES CM 2016/SSGIEA:04, 126 pp.
ICES 2016. Report of the Arctic Fisheries Working Group, ICES HQ, Copenhagen, Denmark. 19-25 April 2016. ICES CM 2016/ACOM:06. 621 pp.

ICES 2016. Final Report of the Working Group on International Deep Pelagic Ecosystem Surveys (WGIDEEPS). ICES CM, ICES CM 2016/SSGIEOM:02: 21pp.

ICES 2016. Report of the Workshop on age reading of Greenland halibut 2 (WKARGH2), 22-26 August, Reykjavik, Iceland. ICES CM 2016/SSGIEOM:16. 36 pp.

ICES 2017. Report of Inter-benchmark protocol on Northeast Arctic cod (IBP ARCTIC COD 2017), Copenhagen, 3-6 April 2017. ICES CM 2017/ACOM:29.

ICES 2017, ICES fisheries management reference points for category 1 and 2 stocks. http://ices.dk/sites/pub/Publication\ Reports/Guidelines\ and\ Policies/12.04.03.01_Reference_points_for_category_1_and_2.pdf
ICES. 2018a. Report of the Benchmark Workshop on Redfish Stocks (WKREDFISH), 29 January-2 February 2018, Copenhagen, Denmark. ICES CM 2018/ACOM:34. 174 pp.
ICES. 2018b. Report of the Workshop on the evaluation of harvest control rules for Sebastes mentella in ICES areas 1 and 2 (WKREBMSE), June-August 2018, by correspondence. ICES CM 2018/ACOM:52. 32 pp.
ICES 2018. NAFO/ICES Pandalus Assessment Group Meeting, 17 to 22 October 2018 NAFO Secretariat, Dartmouth, Canada. ICES CM 2018/ACOM:08

ICES 2018. Report of the Arctic Fisheries Working Group, Ispra, Italy, 18-24 April 2018. ICES C.M. 2018/ACOM:06, 865 pp.

ICES. 2018. Technical Guidelines - ICES reference points for stocks in categories 3 and 4. http://www.ices.dk/sites/pub/Publication Reports/Forms/DispForm.aspx?ID=34082 (Accessed 25 February 2021).
ICES. 2019. Interbenchmark Protocol on assessment model changes for Cod (Gadus morhua) in subareas 1 and 2 (Northeast Arctic) (IBPNEACod2019). ICES Scientific Reports. 1:26. 26 pp . http://doi.org/10.17895/ices.pub. 5278
ICES. 2019. Report of the Arctic Fisheries Working Group (AFWG). ICES Scientific Reports. 1:30. 934 pp . http://doi.org/10.17895/ices.pub. 5292

ICES. 2020. Benchmark Workshop for Demersal Species (WKDEM). ICES Scientific Reports. 2:31. 136 pp . http://doi.org/10.17895/ices.pub. 5548

ICES. 2020. Report of the Arctic Fisheries Working Group (AFWG). ICES Scientific Reports. 2:52. 577 pp. http://doi.org/10.17895/ices.pub. 6050
ICES. 2020c. Tenth Workshop on the Development of Quantitative Assessment Methodologies based on LIFE-history traits, exploitation characteristics, and other relevant parameters for data-limited stocks (WKLIFE X). ICES Scientific Reports. 2:98. 72 pp. http://doi.org/10.17895/ices.pub. 5985

ICES. 2021a. Benchmark Workshop for Barents Sea and Faroese Stocks (WKBARFAR 2021). ICES Scientific Reports. 3:21. 205 pp. https://doi.org/10.17895/ices.pub. 7920
ICES. 2021b. Working Group on the Integrated Assessments of the Barents Sea (WGIBAR). ICES Scientific Reports. 3:xx. (in press).

ICES 2021c. ICES Guidance for completing single-stock advice 2021. ICES Advice guide 2021. Last update 22 April 2021.

ICES 2021d. Arctic Fisheries Working Group (AFWG). ICES Scientific Reports. 3:58. 817pp. https://doi.org/10.17895/ices.pub. 8196
ICES 2022. Workshop on the evaluation of northern Norwegian coastal cod harvest control rules (WKNCCHCR). ICES Scientific Reports. 4:49. 115 pp. https://doi.org/10.17895/ices.pub. 20012459

IUCN. 2001. IUCN red list categories and criteria. Version 3.1. IUCN (World Conservation Union, Gland, Switzerland, and Cambridge UK). (Also available in full text format at: www.iucnredlist.org/static/categories_criteria)

IUCN. 2003. Guidelines for application of IUCN red list criteria at regional levels. IUCN (World Conservation Union, Gland, Switzerland, and Cambridge UK). (Also available in full text format at: www.iucnredlist.org/static/categories_criteria)

Jakobsen, T., Korsbrekke, K., Mehl, S., and Nakken, O. 1997. Norwegian combined acoustic and bottom trawl surveys for demersal fish in the Barents Sea during winter. ICES CM 1997/Y:17.
Janusz, J., Trella, K. and Nermer, T. 2008. Report on Polish fishing activity and survey on redfish(Sebastes mentella) in the NEAFC Regulatory Area (ICES IIa) in 2007. Working document no 8. in: Report of the Arctic Fisheries Working Group (AFWG), 21-29 April 2008, ICES Headquarters, Copenhagen. ICES CM 2008\ACOM:01. 531 pp

Janusz, J. and Trella, K. 2009. Results of the Polish fishing survey of Greenland halibut in the Svalbard Protection Zone (ICES IIb) in April 2008. Working document no 10. in: Report of the Arctic Fisheries Working Group (AFWG), 21-27 April 2009, San-Sebastian, Spain. Diane Lindemann. 579 pp.

Johannesen, E., Johansen, G. O., and Korsbrekke, K. 2016. Seasonal variation in cod feeding and growth in a changing sea. Canadian Journal of Fisheries and Aquatic Sciences 73(2): 235-245.

Johnsen, E., A. Totland, Å. Skålevik, A. J. Holmin, G. E. Dingsør, E. Fuglebakk, and N. O. Handegard. 2019. StoX: An open source software for marine survey analyses. 10:1523-1528.
Johansen, T., Westgaard,J.-I., Seliussen, B.B., Nedreaas, K., Dahle, G., Glover, G.A., Kvalsund, R., and Aglen, A. 2017. "Real-time" genetic monitoring of a commercial fishery on the doorstep of an MPA reveals unique insights into the interaction between coastal and migratory forms of the Atlantic cod. ICES Journal of Marine Science (2017), doi:10.1093/icesjms/fsx22

Jørgensen, B 1997. The theory of dispersion models. Chapman \& Hall. ISBN 978-0412997112.
Jørgensen C., Enberg K., Dunlop E.S,. Arlinghaus R., Boukal D.S., Brander K., Ernande B., Gårdmark A., Johnston F., Matsumura S., Pardoe H., Raab K., Silva A., Vainikka A., Dieckmann U., Heino M., Rijnsdorp A.D. 2008. The role of fisheries-induced evolution - response. Science. 320: 48-50.

Kenchington, T.J. Natural mortality estimators for information-limited fisheries. Fish and Fisheries, 2014, 15.4: 533-562.

Kennedy, J., Gundersen, A.C. and Boje, J.2009. When to count your eggs: Is fecundity in Greenland halibut (Reinhardtius hippoglossoides W.) down-regulated? Fisheries Research, 100(3): 260-265.

Kennedy, J., Gundersen, A.C., Høines, Å.S. and Kjesbu, O.S., 2011. Greenland halibut (Reinhardtius hippoglossoides) spawn annually but successive cohorts of oocytes develop over 2 years, complicating correct assessment of maturity. Canadian Journal of Fisheries and Aquatic Sciences, 68(2): 201-209.
Kennedy, J., Hedeholm, R.B., Gundersen, A.C. and Boje, J., 2014. Estimates of reproductive potential of Greenland halibut (Reinhardtius hippoglossoides) in East Greenland based on an update of maturity status. Fisheries Research, 154: 73-81.

Korsbrekke, K. 1997. Norwegian acoustic survey of Northeast Arctic cod on the spawning grounds off Lofoten. ICES C.M 1997/Y:18.

Kovalev, Yu.A., and Yaragina N.A. 2009. The effects of population density on the rate of growth, maturation, and productivity of the stock of the Northeast Arctic cod. Journal of Ichthyology 49, № 1:56-65.

Kovalev, Y., Prozorkevich, D., and Chetyrkin, A. 2017. Estimation of Ecosystem survey 2016 index in situation of not full area coverage. Working Document No. 12 to the Arctic Fisheries Working Group, Copenhagen, 18-25 April 2017.

Kovalev, Y., and Chetyrkin, A. 2019. What does NEA cod want for prediction - Fsq or TAC constrain? Working Document No. 11 to the Arctic Fisheries Working Group. ICES. 2019. Arctic Fisheries Working Group (AFWG). ICES Scientific Reports. 1:30. 934 pp.

Kuparinen A., Stenseth N. C., Hutchings J. A. 2014. Fundamental population-productivity relationships can be modified through density-dependent feedbacks of life-history evolution. Evol Appl 7(10):121825.

Laurenson CH, Johnson A, Priede IG. 2005. Movements and growth of monkfish Lophius piscatorius tagged at the Shetland Islands, Northeastern Atlantic. Fisheries Research. 2005 Febru-ary 28; 71(2):185-95.

Lorenzen, K. 1996. The relationship between body weight and natural mortality in juvenile and adult fish: a comparison of natural ecosystems and aquaculture. - Journal of fish biology 49: 627-642.

Mace, P. M. and Doonan, I. J. 1988. A generalised bioeco-nomic simulation model for Àsh population dynamics. New Zealand Fishery Assessment Research Document 88/4. Fisheries Research Centre, MAFFish, POB 297: Wellington, NZ.

Mace, P.M. and M.P. Sissenwine. 1993. How much spawning per recruit is enough? In S.J. Smith, J.J. Hunt and D. Rivard [eds.] Risk evaluation and biological reference points for fisheries management. Canadian Special Publications in Fisheries and Aquatic Sciences 120:101-118.

Mangel, M., Brodziak, J., and DiNardo, G. 2010. Reproductive ecology and scientific inference of steepness: a fundamental metric of population dynamics and strategic fisheries management. Fish Fish. 11: 89104. doi:13.1111/j.1467-2979.2009.00345.x

Mangel, M., MacCall, A.D., J. Brodziak, E.J. Dick, R. E. Forrest, R. Pourzand, and S. Ralston 2013. A perspective on steepness, reference points, and stock assessment. Can. J. Fish. Aquat. Sci. 70: 930-940 (2013) dx.doi.org/13.1139/cjfas-2012-0372

Mehl, S., Aglen, A., Alexandrov, D.I., Bogstad, B., Dingsør, G.E., Gjøsæter, H., Johannesen, E., Korsbrekke, K., Murashko, P.A., Prozorkevich, D.V., Smirnov, O., Staby, A., and Wenneck, T. de Lange, 2013. Fish investigations in the Barents Sea winter 2007-2012. IMR-Pinro Joint Report Series 1-2013, 97 pp.

Mehl, S., Aglen, A., Bogstad, B., Dingsør, G.E., Gjøsæter, H., Godiksen, J., Johannesen, E., Korsbrekke, K., Staby, A., Wenneck, T. de Lange, Wienerroither, R., Murashko, P. A., and Russkikh, A. 2014. Fish investigations in the Barents Sea winter 2013-2014. IMR-PINRO Joint Report Series 2-2014, 73 pp.

Mehl, S. Aglen, A., Amelkin, A., Dingsør, G.E., Gjøsæter, H., Godiksen, Staby, A., Wenneck, T. de Lange, and Wienerroither, R. 2015. Fish investigations in the Barents Sea, winter 2015. IMR-PINRO report series 2-2015. 61 pp .

Mehl, S., Aglen, A., Bogstad, B., Dingsør, G.E., Korsbrekke, K., Olsen, E., Staby, A., Wenneck, T. de Lange, Wienerroither, R., Amelkin, A. V., and Russkikh, A. A. 2016. Fish investigations in the Barents Sea winter 2016. IMR-PINRO Joint Report Series 4-2016, 78 pp.
Mehl, S., Aglen, A., Berg, E., Dingsør, G. and Korsbrekke, K. 2014. Akustisk mengdemåling av sei, kysttorsk og hyse, Finnmark - Møre, hausten 2014. [Acoustic abundance of saithe, coastal cod and haddock Finnmark - Møre Autumn 2014]. In Norwegian, legends in English. Toktrapport/Havforskningsinstituttet/ISSN 1503-6294/Nr. 1 - 2014 (38pp).

Mehl, S, Aglen, A., Berg, E. Dingsør, G. and Korsbrekke, K. 2015. Akustisk mengdemåling av sei, kyst-torsk og hyse Finnmark - Møre hausten 2015. Acoustic abundance of saithe, coastal cod and haddock Finnmark - Møre Autumn 2015. In Norwegian, legends in English. Toktrapport/Havforskningsinstituttet/ISSN 1503-6294, Nr. 4-2015. 38pp.

Mehl, S, Aglen, A., Berg, E. Dingsør, G. and Korsbrekke, K. 2016. Akustisk mengdemåling av sei, kyst-torsk og hyse Finnmark - Møre hausten 2016. Acoustic abundance of saithe, coastal cod and haddock Finnmark - Møre Autumn 2016. In Norwegian, legends in English. Toktrapport/Havforskningsinstituttet/ISSN 1503-6294, Nr. 15-2016. 38pp.

Mehl, S., Aglen, A., Bogstad, B., Staby, A., de Lange Wenneck, T. Wienerroither, R., and Russkikh, A.A. 2017a. Fish investigations in the Barents Sea winter 2017. Working Document No. 3 to the Arctic Fisheries Working Group, Copenhagen, 18-25 April 2017.
Mehl, S., Aglen, A. and Johnsen, E. 2017b. Re-estimation of swept area indices with CVs for main demersal fish species in the Barents Sea winter survey 1994 - 2016 applying the Sea2Data StoX software2017. Fisken og Havet No. 10, 2016. Institute of Marine Research, Bergen, Norway. 43 pp.

Mehl S., Aglen A., Johnsen E. and Å. Skålevik. 2018a. Estimation of acoustic indices with CVs for cod and haddock in the Barents Sea winter survey 1994-2017 applying the Sea2Data StoX software. Fisken og Havet, №5, 29 p.

Mehl, S. Berg, E., Korsbrekke, K., Olsen, E., and Staby, A. 2018b. Acoustic abundance of saithe, coastal cod and haddock Finnmark - Møre Autumn 2017. Toktrapport/ Havforskningsinstituttet/ISSN 15036294/Nr. 2-2018.

Mehl, S., de Lange Wenneck, T., Aglen, A., Fuglebakk, E., Gjøsæter, H., Godiksen, J. A., Seim, S. E., and Staby, A. 2019. Fish investigations in the Barents Sea winter 2019. IMR-PINRO Joint Report Series 42019, 77 pp.
Mehl, S., and Yaragina, N. A. 1992. Methods and results in the joint PINRO-IMR stomach sampling program. In: Bogstad, B. and Tjelmeland, S. (eds.), Interrelations between fish populations in the Barents Sea. Proceedings of the fifth PINRO-IMR Symposium. Murmansk, 12-16 August 1991. Institute of Marine Research, Bergen, Norway, 5-15.
Mikhaylov, A. 2016. Long-term HCR-parameters estimation for Greenland halibut based on production model. Working paper no 14. in: Report of the Arctic Fisheries Working Group (AFWG), Dates 19-25 April 2016, ICES HQ, Copenhagen, Denmark. ICES CM 2016/ACOM:06. 621 pp.

Mikhaylov, A. 2019. Update reference point estimation for Greenland halibut based on production model. Working document no. 21 in report of the Arctic Fisheries Working Group, ICES CM 2018/ACOM:06.
Mjanger H. and J.A. Godiksen. 2018. Report of the workshop on age reading of cod (Gadus morhua L.) and haddock (Melanogrammus aeglefinus) between IMR and PINRO May 30 - June 1 2017. Working Document no. 10. ICES Arctic Fisheries Working Group, ICES CM 2018/ACOM:XX.

Mortensen, E. 2007. Er det variasjon i diett og lengde ved alder hos torsk (Gadus morhua L.) nord for $64^{\circ} \mathrm{N}$ ? [in Norwegian]. Master Thesis, University of Tromsø, June 2007.

Muñoz, P. D., Martinez-Escauriaza, R., González, C., and Ramilo, L. 2016. Spanish bottom trawl spring survey "Fletán Ártico 2015" in the Slope of Svalbard (ICES Division Ilb2). Working document no 10. in: Report of the Arctic Fisheries Working Group (AFWG), Dates 19-25 April 2016, ICES HQ, Copenhagen, Denmark. ICES CM 2016/ACOM:06. 621 pp.

Muñoz, P. D., Bagués, A. S., González, C., González-Troncoso, D., Nogueira, A., and Ramilo, L. 2017. Spanish bottom trawl autumn survey "Fletán Ártico 2016" in the Slope of Svalbard (ICES Division IIb2). Working document no 7. in: Report of the Arctic Fisheries Working Group (AFWG), 19-25 April 2017, Copenhagen, Denmark. ICES CM 2017/ACOM:06. 493 pp.
Nedreaas, K. 2014. Review of historic commercial catch-per-unit-of-effort (CPUE) series previously used in stock evaluation of Greenland halibut (Reinhardtius hippoglossoides) in ICES Subareas I and II. Are such CPUE series appropriate to use in future Greenland halibut stock assessments? Working document no 2. in: Report of the Data Compilation Workshop on Northeast Arctic Greenland Halibut and Assessment Methods (DCWKNGHD), 10-12 November 2014, Murmansk, Russia. ICES CM 2014/ACOM:65. 56 pp .
Nedreaas K. and Smirnov O, 2003. Stock characteristics, fisheries and management of Greenland halibut (Reinhardtius hippoglossoides Walbaum) in the northeast Arctic. Proceedings of the 10th NorwegianRussian Symposium Bergen, Norway 27-29 August 2003.

Nedreaas, K. 2017. Conversion factors for products of cod (Gadus morhua) north of $62^{\circ} \mathrm{N}$ in the winter season 2015 - inaccurate current practice. WD no. 15 to ICES AFWG 2017.
Nielsen, A., Berg, C.W., 2014. Estimation of time-varying selectivity in stock assessments using state-space models. Fish. Res. 158:96-101.

Núñez, L.A., Hallfredsson, E.H. and Falk-Petersen, I.-B., 2015. Different maturity scales affect estimations of fecundity, TEP and spawning stock size of Greenland halibut, Reinhardtius hippoglossoides (Walbaum, 1792). Marine Biology Research: 1-10.

Ofstad, L. H. 2013. Anglerfish Lophius piscatorius L. in Faroese waters. Life history, ecological importance and stock status. Dr. scient thesis, University of Tromsø. February 2013. 81 pp.

O'Sullivan M., Wright P. J., Verspoor E., Knox D., Piertney S. 2006. Absence of spatial and temporal genetic differentiation at microsatellite loci in north east Atlantic anglerfish (Lophius piscatorius). Journal of Fish Biology 2006; 69:261.

Pedersen, T., Nilsen, M., Berg, E., and Reigstad M. 2007. Trophic model of a lightly exploited cod-dominated ecosystem. In; Nilsen, M: "Trophic interactions and the importance of macrobenthic invertebrate production in two Arctic fjord systems". A dissertation for PhD, University of Tromsø, Autumn 2007

Pedersen, T. and Pope, J.G. 2003a. Sampling and a mortality model of a Norwegian cod (Gadus morhua L.) fjord population. Fish. Res. 63, 1-20.

Pedersen, T., and Pope, J. 2003b. How may feeding data be integrated into a model for a Norwegian fjord population of cod (Gadus morhua L.)? Scientia Marina, 67(Suppl. 1): 155-169.

Planque, B. 2015. S. mentella assessment - handling the +group.: WD03 - ICES AFWG2015. 8 pp .
Planque, B. 2016. Possible use of the Pelagic and slope surveys in the analytical assessment of Sebastes mentella in ICES areas 1 and 2.: WD05-ICES AFWG2016. 6 pp.

Planque, B., Vollen, T., Höffle, H., Harbitz A., 2018. Use of StoX for estimating numbers@age of Sebastes mentella from the international deep pelagic ecosystem survey in the Norwegian Sea.: WD07 - ICES AFWG2018. 38 pp.

Ponomarenko 1973, 1984
Ponomarenko, I.Ya. and N.A.Yaragina. 1990. Long-term dynamics of the Barents Sea cod feeding on capelin, euphausiids, shrimp and the annual consumption of these objects. Feeding resources and interrelations of fishes in the North Atlantic: Selected papers of PINRO. Murmansk. 1990. p.109-130 (in Russian).

Prince, J.D., Hordyk, A.R., Valencia, S.R., Loneragan, N.R., and Sainsbury, K.J. 2015. Revisiting the concept of Beverton-Holt life-history invariants with the aim of informing data-poor fisheries assessment. ICES J. Mar. Sci. 72: 194-203.

Prince, J., Creech, S., Madduppa, H., and Hordyk, A. 2020. Length based assessment of spawning potential ratio in data-poor fisheries for blue swimming crab (Portunus spp.) in Sri Lanka and Indonesia: Implications for sustainable management. Regional Studies in Marine Science, 36: 101309.

Prozorkevich Dmitry, Johannesen Edda, and Johansen Geir Odd. 2020. Barents Sea ecosystem survey 2019: cod and haddock indices. Working document no 1 in: Report of the Arctic Fisheries Working Group (AFWG), 15-22 April 2020. ICES CM 2015/ACOM:05. 639 pp.

Prozorkevitch, D. and van der Meeren, G. (eds) 2021. Survey report from the joint Norwegian/ Russian ecosystem survey in the Barents Sea and adjacent waters August-October 2019. IMR/PINRO Joint Report Series, 1-2021, In press

Russkikh, A. A., amd Smirnov, O. V. 2016. Results of the Russian trawl survey of Greenland halibut in the Barents Sea and adjacent waters in 2015. Working document no 16. in: Report of the Arctic Fisheries Working Group (AFWG), Dates 19-25 April 2016, ICES HQ, Copenhagen, Denmark. ICES CM 2016/ACOM:06. 621 pp.

Skaret, G., D. Prozorkevich, H. Gjøsæter, and B. Bogstad. 2019. Evaluation of potential sources of error leading to an underestimation of the capelin stock in 2016. Pp. 166-175 in Proceedings of the 18th Rus-sian-Norwegian symposium, Murmansk 5-7 June 2018. Edited by Evgeny Shamray, Geir Huse, Alexander Trofimov, Svein Sundby, Andrey Dolgov, Hein Rune Skjoldal, Konstantin Sokolov, Lis Lindal Jørgensen, Anatoly Filin, Tore Haug and Vladimir Zabavnikov. IMR/PINRO report Series 1-2019.
Skaret, G. et al. 2021. Testing of trawl-acoustic stock estimation of spawning capelin 2021. IMR Survey report (in prep.)

Skjœraasen, J. E., Kennedy, J., Thorsen, A., Fonn, M., Strand, B. N., Mayer, I., and Kjesbu, O. S. 2009. Mechanisms regulating oocyte recruitment and skipped spawning in Northeast Arctic cod (Gadus morhua). Canadian Journal of Fisheries and Aquatic Sciences, 66: 1582-1596.

Smirnov, O. 2011. Results of the Russian survey of Greenland halibut in the Barents Sea and adjacent waters 2009. Working document no 21. in: Report of the Arctic Fisheries Working Group (AFWG), 28 April 4 May 2011, Hamburg, Germany. ICES CM 2011/ACOM:05. 659 pp.

Sokolov A., Russkikh A., Kharlin S., Kovalev Yu. A., and Yaragina N.A. 2018. Results of the Russian trawlacoustic survey on cod and haddock in the Barents Sea and adjacent waters in October-December 2017. Working Document no. 11. ICES Arctic Fisheries Working Group, ICES CM 2018/ACOM:06.

Solmundsson, J, Jonsson, E and Björnsson, H. 2007. Recent changes in the distribution and abundance of monkfish (Lophius piscatorius) in Icelandic waters. ICES CM 2007/K:02. 16pp.

Staalesen, B.I. 1995. Breiflabb (Lophius piscatorius L.) langs norskekysten. Cand.scient thesis, University of Bergen. 88 pp. (In Norwegian, summary in English)

Staby, A., E. Aglen, A., Gjøsæter, H. and J. Fall. 2021. Akustisk mengdemåling av sei og kysttorsk Finnmark - Møre høsten 2020. (Acoustic Abundance of saithe and coastal cod Finnmark-Møre 20120). Institute of Marine Research, Norway. Cruise report 4-2021. 34 pp.

Stiansen et al., 2005. IMR status report on the Barents Sea ecosystem, 2004-2005. WD1, AFWG 2005.
Subbey, S., JE. Stiansen, B. Bogstad, T. Bulgakova and O. Titov, 2008. Evaluating Recruitment Models for (Age 3) NEA Cod. Working document \#27. Report of the Arctic Fisheries Working Group (AFWG). 21-29 April 2008, ICES Headquarters, Copenhagen. ICES CM 2008 $\backslash$ ACOM:01.
Svendsen, E., Skogen, M., Budgell, P., Huse, G., Ådlandsvik, B., Vikebø, F., Stiansen, J.E., Asplin, L., and Sundby, S. 2007. An ecosystem modelling approach to predicting cod recruitment. Deep-Sea Research Part II, 54:2810-2821.

Thangstad, T., Bjelland, O., Nedreaas, KH, Jónsson, E., Laurenson, CH and Ofstad, LH 2006. Anglerfish (Lophius spp.) in Nordic waters. TemaNord 2006:570. © Nordic Council of Ministers, Copenhagen 2006. ISBN 92-893-1416-8. 162 pp.

Then, A. Y., Hoenig, J. M., Hall, N. G., and Hewitt, D. A. 2018. Evaluating the predictive performance of empirical estimators of natural mortality rate using information on over 200 fish species. ICES Journal of Marine Science, 75: 1509-1509. https://doi.org/10.1093/icesjms/fsx199 (Accessed 18 January 2021).

Thygesen, U. H., Albertsen, C. M., Berg, C. W., Kristensen, K., and Nielsen, A. 2017. Validation of ecological state space models using the Laplace approximation Environmental and Ecological Statistics 24 (2): 317-339.

Titov, O., Pedchenko, A. and Karsakov, A., 2005. 'Assessment of Northeast Arctic cod and capelin recruitment from data on ecological situation in the Barents Sea in 2004-2005'. Working document \#16 in: Report of the Arctic Fisheries Working Group', Murmansk, Russia, April 19-28, 2005. ICES C.M. 2005/ACFM:20, 564 pp.

Titov O.V. 2010. Assessment of population recruitment abundance of Northeast Arctic cod considering the environment data. WD 22, AFWG 2010.

Titov O. 2018. Assessment of population recruitment abundance of Northeast Arctic cod considering the environment data. Working document 17. ICES Arctic Fisheries Working Group (AFWG), ICES CM 2018 \ACOM:XX.

Tjelmeland, S. 2002. A model for the uncertainty around the yearly trawl-acoustic estimate of biomass of Barents Sea capelin, Mallotus villosus (Müller). ICES Journal of Marine Science, 59: 1072-1080.

Tjelmeland, S. 2005. Evaluation of long-term optimal exploitation of cod and capelin in the Barents Sea using the Bifrost model. Pp. 112-129 in: Shibanov, V. (ed.). "Ecosystem Dynamics and Optimal Longterm Harvest in Barents Sea Fisheries". Proceedings of the 11 th Russian-Norwegian Symposium, Murmansk, Russia, 15-17 August 2005. IMR/PINRO report series 2/2005, 331 pp.

Tjelmeland, S. and Lindstrøm, U. 2005. An ecosystem element added to the assessment of Norwegian spring spawning herring: implementing predation by minke whales. ICES Journal of Marine Science 62(2):285-294.

Totland, A., and Godø, O. R. 2001. BEAM-an interactive GIS application for acoustic abundance estimation. In Proceedings of the First Symposium on Geographic Information System (GIS) in Fisheries Science. Fishery GIS Research Group. Saitama, Japan (Vol. 52).

Tranang C. A., Vollen T. and Höffle H. 2020. Use of StoX for recalculating numbers at age and numbers at length of Sebastes norvegicus from the Barents Sea NOR-RUS demersal fish cruise in winter.: WD17 ICES AFWG2020. 60 pp .
Trella, K. and J. Janusz 2012. Results of the Polish fishing survey of Greenland halibut (Reinhardtius hippoglossoides) in the Svalbard Protection Zone (ICES IIb) in March 2011. Working document no 6. in: Report of the Arctic Fisheries Working Group 2012 (AFWG), 20-26 April 2012, ICES Headquarters, Copenhagen. ICES CM 2012/ACOM:05. 670pp.

Vasilyev D. 2005 Key aspects of robust fish stock assessment. M: VNIRO Publishing, 2005. 105 p.
Vasilyev D. 2006. Change in catchability caused by year class peculiarities: how stock assessment based on separable cohort models is able to take it into account? (Some illustrations for triple-separable case of the ISVPA model - TISVPA). ICES CM 2006/O:18. 35 pp

Vasilyev D. 2020. NEA cod stock assessment by means of TISVPA. Working document no 12 in : Report of the Arctic Fisheries Working Group (AFWG), 15-22 April 2020. ICES CM 2015/ACOM:05. 639 pp.
Vasilyev D. NEA cod stock assessment by means of TISVPA. Working document no 18 in: Report of the Arctic Fisheries Working Group (AFWG), 14-20 April 2021. ICES CM 2021/ACOM:05. 639 pp

Vølstad, J. H., Korsbrekke, K., Nedreaas, K. H., Nilsen, M., Nilsson, G. N., Pennington, M., Subbey, S., and Wienerroither, R. 2011. Probability-based surveying using self-sampling to estimate catch and effort in Norway's coastal tourist fishery. - ICES Journal of Marine Science, doi: 10.1093/icesjms/fsrXXX

Westgaard, J.-I., Saha, A., Kent, M.P., Hansen, H.H., Knutsen, H., Hauser, L., Cadrin, S.X., Albert, O.T. and Johansen, T., 2016. Genetic population structure in Greenland halibut (Reinhardtius hippoglossoides) and its relevance to fishery management. Canadian Journal of Fisheries and Aquatic Sciences.

Weltersbach, M. S., and Strehlow, H. V. 2013. Dead or alive-estimating post-release mortality of Atlantic cod in the recreational fishery. - ICES Journal of Marine Science, 70: 864-872. doi:10.1093/icesjms/fst038

Winker, H., Carvalho, F., Kapur, M. 2018. JABBA: Just Another Bayesian Biomass Assessment. Fisheries Research 204: 275-288.

WKANGLER 2018. Report of the Benchmark Workshop on Anglerfish Stocks in the ICES Area (WKANGLER), 12-16 February 2018, Copenhagen, Denmark. ICES CM 2018/ACOM:31. 177 pp.

Yaragina, N. A. 2010. Biological parameters of immature, ripening and non-reproductive mature Northeast Arctic cod in 1984-2006. ICES Journal of Marine Science, 67: 2033-2041.

Yaragina N.A. and B. Bogstad. 2017. Historic difference in stock weight and maturity at age in Northeast Arctic cod. Working document 10. In: Report of the Arctic Fisheries Working Group (AFWG), 19-25 April 2017, Copenhagen, Denmark. ICES CM 2017/ACOM:06. 486 pp
Yaragina N.A. Nedreaas K.H., Koloskova V., Mjanger H., Senneset H., Zuykova N. and Ágotnes P. 2009. Fifteen years of annual Norwegian-Russian cod comparative age readings. Marine Biology Research 5(1): 54-65.

Yaragina N.A. and Bogstad B. 2017. Historic difference in stock weight and maturity at age in Northeast Arctic cod. Working Document No. 10 to the Arctic Fisheries Working Group, Copenhagen, 18-25 April 2017.

Yaragina N. A., Kovalev Yu. A., and Chetyrkin A. 2018. Extrapolating predation mortalities back in time: an example from North-east Arctic cod cannibalism, Marine Biology Research: https://doi.org/10.1080/17451000.2017.1396342

Zuykova N.V., Koloskova V.P., Mjanger H., Nedreaas K.H., Senneset H., Yaragina N.A., Ågotnes P. and Aanes S. 2009. Age determination of Northeast Arctic cod otoliths through 50 years of history. Marine Biology Research 5(1): 66-74.
Zuykova N.V., Mjanger H. at all. 2020. Report on the meeting between Norwegian and Russian age reading specialists at Polar Branch of FSBSI "VNIRO" Murmansk, 20-24 May 2019. Working document no 8 in: Report of the Arctic Fisheries Working Group (AFWG), 15-22 April 2020. ICES CM 2015/ACOM:05. 639 pp .

WD 15. 2019. Updated mean ratios between the combined and Norwegian data on weight at age and maturity at age in Northeast Arctic cod. Working document no 8 in: Report of the Arctic Fisheries Working Group (AFWG), 15-22 April 2020. ICES CM 2015/ACOM:05. 639 pp.

## Annex 1: List of participants

| Member | Institute | Country of institute | E-mail |
| :---: | :---: | :---: | :---: |
| Anders Nielsen | DTU Aqua | Denmark | an@aqua.dtu.dk |
| Arved Staby | IMR | Norway | arved.staby@hi.no |
| Bjarte Bogstad | IMR | Norway | bjarte.bogstad@hi.no |
| Brian Stock | IMR | Norway | brian.stock@hi.no |
| Caroline Aas Tranang | IMR | Norway | caroline.aas.tranang@hi.no |
| Daniel Howell | IMR | Norway | daniel.howell@hi.no |
| Edda Johannesen | IMR | Norway | edda.johannesen@hi.no |
| Elena Eriksen | IMR | Norway | elena.eriksen@hi.no |
| Elise Eidset | IMR | Norway | elise.eidset@hi.no |
| Elvar H. Hallfredsson | IMR (Troms $\varnothing$ ) | Norway | elvar.hallfredsson@hi.no |
| Erik Berg | IMR (Troms $\varnothing$ ) | Norway | erik.berg@hi.no |
| Georg Skaret | IMR | Norway | georg.skaret@hi.no |
| Hannes Höffle | IMR | Norway | hannes.hoffle@hi.no |
| Harald Gjøsæter | IMR | Norway | harald.gjoesaeter@hi.no |
| Jane Aanestad Godiksen | IMR | Norway | jane.godiksen@hi.no |
| Johanna Fall | IMR | Norway | johanna.fall@hi.no |
| John Tyler Trochta | IMR | Norway | john.tyler.trochta@hi.no |
| José Miguel Casas Sanchez | IEO | Spain | mikel.casas@ieo.csic.es |
| Kjell Nedreaas | IMR | Norway | kjell.nedreaas@hi.no |
| Kristin Windsland | IMR | Norway | kristin.windsland@hi.no |
| Laura Clain | University of La Rochelle | France | laura.clain@etudiant.univIr.fr |
| Maria Fossheim | IMR | Norway | maria.fossheim@hi.no |
| Matthias Bernreuther | Thünen-Institute of Sea Fisheries | Germany | matthias.bernreuther@thuenen.de |
| Olav Nikolai Breivik | Norwegian Computing Center | Norway | olavbr@nr.no |
| Ross Tallman | DFO | Canada | foss.tallman@dfompo.gc.ca |


| Member | Institute | Country of insti- <br> tute | E-mail |
| :--- | :--- | :--- | :--- |
| Samuel Subbey | IMR | Norway | samuel.subbey@hi.no |
| Sofie Gundersen | IMR | Norway | sofie.gundersen@hi.no |
| Tone Vollen | IMR | Norway | tone.vollen@hi.no |

## Annex 2: Resolutions

2021/2/FRSG02

## Approved November 2021

The Arctic Fisheries Working Group (AFWG), chaired by Daniel Howell, Norway, will meet online 21-27 April 2022 to:
d) Address generic ToRs for Regional and Species Working Groups, for all stocks except the Barents Sea capelin, which will be addressed at a meeting in autumn;
e) For Barents Sea capelin oversee the process of providing intersessional assessment;
f) Conduct reviews as required of time any series computed using the STOX and ECA open-source software for use in assessment in the Barents Sea.
The assessments will be carried out on the basis of the stock annex. The assessments must be available for audit on the first day of the meeting.
Material and data relevant to the meeting must be available to the group on the dates specified in the 2022 ICES data call.

AFWG will report by 6 May 2022 and October 2022 for Barents Sea capelin for the attention of the Advisory Committee.

Only experts appointed by national Delegates or appointed in consultation with the national Delegates of the expert's country can attend this Expert Group.

## Annex 3: Working documents

Working documents can be found in the SharePoint folder here.

| WD_01_BESS_cod_index_2004-2021-updated-with-summaries.xls |
| :--- |
| WD_02_BESS_haddock_index_2004-2021-updated-with-summaries.xlsx |
| WD_03 Revision_of_Coastal_cod_catch_data_1977-1993.docx |
| WD_04 Cod effort and CPUE NOR TRAWL LOG BOOK - 2011-2021-per 4 mar 2022.docx |
| WD_05 Haddock effort and CPUE NOR TRAWL LOG BOOK - 2011-2021-per 4 March 2022.docx |
| WD_06 Anglerfish in ICES 1 and 2_per 260422.doc |
| WD_08_Spanish Cod fisheries 2021.docx Pelagic Redfish fishery 2021.docx |
| WD 09 NEA Haddock Cod Ecosystem survey indices StoX 3.3 2020-2021.docx |
| WD_10_EffectOfErrorln2020SurveylndexNCCN67.docx |
| WD_11 S. norvegicus landings data method revision.pdf |

## Annex 4: Audit reports

## Audit of Golden redfish (AFWG 2022)

Date: 10 June 2022
Auditor: Arved Staby

## General

The Northeast Arctic Haddock assessment and draft advice have been approved by the Working Group.
For single stock summary sheet advice:

- Assessment type: Length based model that uses a range of tuning series (commercial and survey)
- Assessment: analytical
- Forecast: presented
- Assessment model: Age-length based Gadget model
- Data issues: Numbers at age could not be estimated using StoXReca
- Consistency:
- Stock status: SSB still very low (under Blim and Bpa) though signs of improvement. Fishing mortality
much too high, and no significant signs of recruitment the last couple of years
- Management Plan: no direct fishery, keep bycatch at a minimum


## General comments

This was a well documented, well ordered and considered section. It was easy to follow and interpret. All data sets described in the stock annex are available.

## Technical comments

No technical comments

## Conclusions

The assessment has been performed correctly and gives a valid basis for advice.

## Audit of Anglerfish in subareas 1 and 2

Date: 04. May 2022
Auditor: John T. Trochta

## General

The Northeast Arctic Anglerfish assessment update has been approved by the Working Group. This is the first audit conducted for Anglerfish.

## For advice other than single-stock summary fisheries advice

Section: Report chapter (Chapter 9)

## Short description of the assessment

1) Assessment type: Update
2) Assessment: Accepted
3) Forecast: Not presented
4) Assessment model: Combination of three assessment methods for Category 3 stocks: LBSPR, CPUE standardization, and JABBA.
5) Consistency: Last year's assessment accepted. Same methodology from 2021 assessment update used this year, but with 2021 data and additional sensitivity analyses.
 posteriors suggest a slightly higher probability of $\mathbf{F}<\mathrm{F}_{\mathrm{MSY}}$ ( $\sim 60 \%$ probability). A shift in length distributions to smaller fish in surveys suggest abundant recruitment that may contribute to fishery in 3-4 years time.
6) Management plan: There is no management plan for Anglerfish. Catch advice comes in the form of the "2-over-3" rule on CPUE with an uncertainty cap and precautionary buffer, which justifies a $10 \%$ reduction in F for 2023. AFWG previously recommended $20 \%$ reduction in fishing effort per year until survey CPUE stopped declining, and CPUE increased in 2021 after a decade of decline.

## General comments

This was a well documented, well ordered and considered section. Aside from minor issues related to clarity in several places, it was easy to follow and interpret. The additional sensitivity analyses for 2022 were well executed and thorough.

## Technical comments

No technical issues were found.

## Conclusions

The assessment has been performed correctly and consistently with the previous assessment (2021), while exploring and improving upon several issues (residual patterns in the CPUE).

## Audit of Haddock in subareas 1 and 2 (AFWG 2022)

Date: 09. June 2022
Auditor: Hannes Höffle
General
The Northeast Arctic Haddock assessment was not updated for 2022, due to it being a shared stock with Russia. Assessment will be done on a bilateral basis later in the year. The audit addresses the updated data tables approved by the Working Group.

## For single stock summary sheet advice:

1) Assessment type: Age-based analytical assessment that uses catches in the models.
2) Assessment: not assessed
3) Forecast: not presented

Assessment model, data issues, consistency, stock status and management plan are unchanged from AFWG 2021.

## General comments

Most of this section was taken over from AFWG 2021. The section is well documented, ordered and considered, offering little to comment on.

## Technical comments

No technical issues were found.

## Conclusions

The section was updated to the extent that was possible for AFWG 2022. Assessment was not performed.

## Audit of Northeast Arctic saithe (AFWG 2022)

Date: 18 May 2022
Auditors: Matthias Bernreuther \& Brian Stock

## General

The Northeast Arctic saithe assessment and draft advice have been approved by the Working Group.

## For single stock summary sheet advice:

1) Assessment type: update
2) Assessment: analytical
3) Forecast: presented
4) Assessment model: SAM - tuning by one acoustic survey (split in two time series)
5) Data issues: The biological sampling from the fishery has been criticized in the last years as being critically low after the termination of the original Norwegian portsampling program in 2009. However, the biological sampling has improved since 2016 and in 2020-2021 the coverage of the commercial fisheries may be (under these circumstances) considered as adequate.
The lack of reliable recruitment estimates is still a major problem for the short-term catch forecast.
6) Consistency: Last year's assessment was accepted. The assessment, recruitment and forecast models have been applied as specified in the stock annex.
7) Stock status: The SSB has been above $B_{p a}$ since 1996 , declined considerably from 2007 to 2011, then increased again and is presently (2021/2022) estimated to be well above $\mathrm{B}_{\mathrm{pa}}$. The fishing mortality was below $\mathrm{F}_{\mathrm{pa}}$ from 1997 to 2009, started to increase in 2005 and was above $\mathrm{F}_{\mathrm{pa}}$ from 2010 to 2012 , but is presently estimated to be most likely below $\mathrm{F}_{\mathbf{p a}}$. The recruitment has since 2005 been at about the long-term geometric mean level.
8) Management Plan: Agreed 2013 (first time in 2007): $\mathrm{F}_{\mathrm{MP}}=0.32$ and SSB above $\mathrm{B}_{\mathrm{pa}}=220000 \mathrm{t}$. The TAC is based on an average TAC for the coming three years based on $\mathrm{F}_{\mathrm{MP}}$. There is a $15 \%$ constraint on TAC change between years. The plan was evaluated by ICES and was found in agreement with the precautionary approach.

## General comments

This was a well documented, well ordered and considered section. It was easy to follow and interpret. All data sets described in the stock annex are available.

## Technical comments

No technical comments

## Conclusions

The assessment has been performed correctly and gives a valid basis for advice.


[^0]:    ICES INTERNATIONAL COUNCIL FOR THE EXPLORATION OF THE SEA CIEM CONSEIL INTERNATIONAL POUR L'EXPLORATION DE LA MER

[^1]:    ${ }^{1}$ Note: The Workshop on the evaluation of northern Norwegian coastal cod harvest control rules (WKNCCHCR) was rescheduled to take place as part of the work of AFWG 26-27 April 2022 after it had been delayed. See: ICES. 2022. Workshop on the evaluation of northern Norwegian coastal cod harvest control rules (WKNCCHCR). ICES Scientific Reports. 4:49. 115 pp. https://doi.org/10.17895/ices.pub. 20012459

[^2]:    ${ }^{2}$ Note: no autumn assessment for Barents Sea capelin was conducted in 2022 as originally planned.

[^3]:    ${ }^{3}$ Currently part of benchmark process WKCAPELIN 2022, expected to report conclusions in 2023.
    ${ }^{4}$ Currently part of benchmark process WKBNORTH 202, together with NWWG Greenland halibut (ghl.27.561214).

[^4]:    ${ }^{1}$ Cod (Gadus morhua) in subareas 1 and 2, north of $67^{\circ} \mathrm{N}$ (Norwegian Sea, Barents Sea), northern Norwegian coastal cod: cod.27.1-2.coastN; Cod (Gadus morhua) in Subarea 2 between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$ (Norwegian Sea), southern Norwegian coastal cod: cod.27.2.coastS.

[^5]:    ${ }^{1}$ Cod (Gadus morhua) in subareas 1 and 2 (Northeast Arctic); cod.27.1-2.

[^6]:    ${ }^{1}$ Haddock (Melanogrammus aeglefinus) in subareas 1 and 2 (Northeast Arctic); had.27.1-2.

[^7]:    ${ }^{1}$ Saithe (Pollachius virens) in subareas 1 and 2 (Northeast Arctic); pok.27.1-2.

[^8]:    ${ }^{2}$ https://github.com/StoXProject/RstoxFDA/

[^9]:    ${ }^{1}$ Beaked redfish (Sebastes mentella) in subareas 1 and 2 (Northeast Arctic); reb.27.1-2.

[^10]:    ${ }^{1}$ Golden redfish (Sebastes norvegicus) in subareas 1 and 2 (Northeast Arctic); reg.27.1-2.

[^11]:    ${ }^{1}$ Greenland halibut (Reinhardtius hippoglossoides) in subareas 1 and 2 (Northeast Arctic); ghl.27.1-2.

[^12]:    ${ }^{2}$ https://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32020R0123\&from=EN
    ${ }^{3}$ https://eur-lex.europa.eu/legal-content/EN/TXT/HTML/?uri=CELEX:32021R0092\&from=EN
    ${ }^{3}$ https://eur-lex.europa.eu/legal-content/EN/TXT/HTML/?uri=CELEX:32022R0515\&qid=1650982320384\&from=en

[^13]:    A new benchmark on the stock is planned for early 2023. Towards the benchmark work is ongoing on revision of all indices that go into the assessment, update of the Gadget model in new version of the program package, amongst other improvements

[^14]:    * Biomass $45 \mathrm{~cm}+2024$ relative to 2023 (biomass 2023 depends on scenario).
    ** Advice value for 2023 relative to the advice value for same scenario in 2022.

[^15]:    ${ }^{1}$ Anglerfish (Lophius budegassa, Lophius piscatorius) in subareas 1 and 2 (Northeast Arctic); anf.27.1-2.

[^16]:    ${ }^{2}$ https://github.com/AdrianHordyk/LBSPR

[^17]:    ${ }^{1}$ Capelin (Mallotus villosus) in subareas 1 and 2 (Northeast Arctic), excluding Division 2.a west of $5^{\circ} \mathrm{W}$ (Barents Sea capelin); cap.27.1-2.

